1 Review Paper

2 Accounting for training data error in machine

3 learning applied to Earth observations

4 Arthur Elmes^{*1,2}, Hamed Alemohammad³, Ryan Avery⁴, Kelly Caylor^{4,5}, J. Ronald Eastman¹, 5 Lewis Fishgold⁶, Mark A. Friedl⁷, Meha Jain⁸, Divyani Kohli⁹, Juan Carlos Laso Bayas¹⁰, 6 Dalton Lunga¹¹, Jessica L. McCarty¹², Robert Gilmore Pontius Jr¹, Andrew B. Reinmann^{13, 14}, 7 John Rogan¹, Lei Song¹, Hristiana Stoynova^{13, 14}, Su Ye¹, Zhuang-Fang Yi¹⁵, Lyndon Estes^{*1} 8 9 ¹ Graduate School of Geography, Clark University, Worcester, MA 01610, USA; reastman@clarku.edu 10 (J.R.E.); rpontius@clarku.edu (R.G.P.); jrogan@clarku.edu (J.R.); lsong@clarku.edu (L.S.); sye@clarku.edu 11 (S.Y.); lestes@clarku.edu (L.E.) 12 ² School for the Environment, University of Massachusetts Boston, Boston, MA 02125, USA; 13 arthur.elmes@umb.edu 14 ³ Radiant Earth Foundation, San Francisco, CA, 94105, USA; hamed@radiant.earth 15 ⁴ Department of Geography, University of California, Santa Barbara, CA 93013, USA; ravery@ucsb.edu 16 ⁵ Bren School of Environmental Science and Management, University of California, Santa Barbara, CA 17 93013, USA; caylor@ucsb.edu 18 ⁶ Azavea, Inc., Philadelphia, PA 19123, USA; <u>lfishgold@azavea.com</u> 19 ⁷ Department of Earth and Environment, Boston University, Boston, MA 02215; <u>friedl@bu.edu</u> 20 ⁸ School for Environment and Sustainability, University of Michigan, 48109, USA; mehajain@umich.edu 21 ⁹ Faculty of Geo-Information Science & Earth Observation (ITC), University of Twente, 7514 AE Enschede, 22 The Netherlands; d.kohli@utwente.nl 23 Center for Earth Observation and Citizen Science, Ecosystems Services and Management Program, 24 International Institute for Applied Systems Analysis (IIASA), Laxenburg, A-2361, Austria; 25 lasobaya@iiasa.ac.at 26 ¹¹ National Security Emerging Technologies, Oak Ridge National Laboratory, Oak Ridge, TN 37831, USA; 27 lungadd@ornl.gov 28 ¹² Department of Geography and Geospatial Analysis Center, Miami University, Oxford, OH 45056, USA; 29 mccartjl@MiamiOH.edu 30 ¹³ Environmental Science Initiative, Advanced Science Research Center at the Graduate Center of the City 31 University of New York (CUNY), New York, NY 10065, USA; Andrew.Reinmann@asrc.cuny.edu 32 ¹⁴ Department of Geography and Environmental Science, Hunter College, New York, NY 10065, USA; 33 Hristiana.stoynova@gmail.com 34 ¹⁵ Development Seed, Washington, DC 20001, USA; nana@developmentseed.org 35 36 *Correspondence: arthur.elmes@umb.edu, Tel. 1-304-906-7946 (A.E.); lestes@clarku.edu, Tel. 1-202-431-0496 37 (L.E.) 38 This paper is a non-peer-reviewed preprint submitted to EarthArXiv. It has been submitted to 39 MDPI Remote Sensing for peer review.

40 Abstract: Remote sensing, or Earth Observation (EO), is increasingly used to understand earth 41 system dynamics and create continuous and categorical maps of biophysical properties and land 42 cover, especially based on recent advances in machine learning (ML). ML models typically require 43 large, spatially explicit training datasets to make accurate predictions. Training data (TD) are 44 typically generated by digitizing polygons on high spatial resolution imagery, by collecting in situ 45 data, or by using pre-existing datasets. TD are often assumed to accurately represent the truth, but 46 in practice almost always have error, stemming from 1) sample design, and 2) sample collection 47 errors. The latter is particularly relevant for image-interpreted TD, an increasingly commonly used 48 method due to its practicality and the increasing training sample size requirements of modern ML 49 algorithms. TD errors can cause substantial errors in the maps created using ML algorithms, which 50 may impact map use and interpretation. Despite these potential errors and their real-world 51 consequences for map-based decisions, TD error is often not accounted for or reported in EO 52 research. Here we review the current practices for collecting and handling TD. We identify the 53 sources of TD error, and illustrate their impacts using several case studies representing different EO 54 applications (infrastructure mapping, global surface flux estimates, and agricultural monitoring), 55 and provide guidelines for minimizing and accounting for TD errors. To harmonize terminology, 56 we distinguish TD from three other classes of data that should be used to create and assess ML 57 models: training reference data, used to assess the quality of TD during data generation; validation 58 data, used to iteratively improve models; and map reference data, used only for final accuracy 59 assessment. We focus primarily on TD, but our advice is generally applicable to all four classes, and 60 we ground our review in established best practices for map accuracy assessment literature, which 61 are essential to follow for TD error accounting. EO researchers should start by determining the 62 tolerable levels of map error and appropriate error metrics. Next, TD error should be minimized 63 during sample design by choosing a representative spatio-temporal collection strategy, use of 64 spatially and temporally relevant imagery and ancillary data sources during TD creation, and 65 selection of a set of legend definitions supported by the data. Further, TD error can be minimized 66 during the collection of individual samples by use of consensus-based collection strategies, by 67 directly comparing interpreted training observations against expert-generated training reference 68 data to derive TD error metrics, and by providing image interpreters with thorough application-69 specific training. We strongly advise that TD error is incorporated in model outputs, either directly 70 in bias and variance estimates or, at a minimum, by documenting the sources and implications of 71 error. TD should be fully documented and made available via an open TD repository, allowing 72 others to replicate and assess its use. To guide researchers in this process, we propose three tiers of 73 TD error accounting standards. Finally, we advise researchers to clearly communicate the 74 magnitude and impacts of TD error on map outputs, with specific consideration given to the likely 75 map audience.

76 Keywords: training data; machine learning; map accuracy; error propagation

77

78 1. Introduction

79 Recent technological advancements have led to a new era in Earth observation (EO, also known 80 as remote sensing), marked by rapid gains in our ability to map and measure features on the Earth's 81 surface such as land cover and land use (LCLU) [e.g. 1,2], vegetation cover and abundance [3], soil 82 moisture [4], infrastructure [5,6], vegetation phenology [7–9], and land surface temperature [10,11]. 83 The resulting data are used by an expanding set of disciplines to gain new insights into socioeconomic 84 and environmental dynamics, such as community-level poverty rates [12], changes in surface water 85 [13] and forest cover [14], and carbon accounting [15]. As such, EO is increasingly shaping our 86 understanding of how the world works, and how it is changing.

87 These breakthroughs are facilitated by several technological advances, particularly the 88 increasing availability of moderate (5-30 m), high (1-5m, High Resolution, HR), and Very High 89 Resolution (<1 m, VHR) imagery, as well as new machine learning (ML) algorithms that frequently 90 require large, high quality training datasets [16–21]. Large training datasets have been necessary for 91 decades in the production of continental and global maps [1,2,22,23], and in the current data-rich era, 92 the impact of TD quality and quantity on map accuracy is even more relevant, especially for maps 93 generated by data-hungry ML algorithms [24–29]. Errors in these products in turn impact the veracity 94 of any downstream products based on those maps [30]. While progress in algorithmic performance 95 continues apace, standards regarding the collection and use of training data (TD) remain 96 uncoordinated across researchers [31]. Additionally, much of the research and development of big 97 data and ML is occurring in industry and the fields of computer science and (non-spatial) data 98 science, leaving a potential knowledge gap for Earth Observation (EO) scientists [32,33].

99 The measurement and communication of map accuracy is a mature topic in EO and related 100 fields, with a variety of metrics and approaches tailored to different data types, analyses, and user 101 groups [34-42]. This includes substantial work to measure error in map reference data (i.e. the 102 independent sample used to assess map accuracy) and account for its impact on map assessment 103 [31,35,43,44]. However, focus on the quality and impacts of TD error has been less systematic. While 104 several efforts have been made to use and evaluate the impact of different aspects of TD quality 105 (noise, sample design, and size) on classifiers [27,29,45–50], much of this work focuses on exploring 106 these issues for specific algorithms [28,45,50,51]. This previous research shows that the impact of TD 107 error can be substantial but varied, suggesting that a more comprehensive approach to this issue is 108 warranted. Furthermore, while TD and map reference data are often collected using the same 109 approaches [52–54] and generally subject to the same errors, the existing procedures to minimize and 110 account for map reference errors [31,35,43,44] are not necessarily relevant for quantifying the impacts 111 of TD error. The problems associated with TD error can be summarized as follows:

112 113

114

115

123

- 1. The 'big data' era vastly increases the demand for TD
- 2. ML-generated map products rely heavily on human-generated TD, which in most cases contain error, particularly when developed through image interpretation
- 116 117

3. Uncertainty in TD is rarely assessed or reported, and TD are often assumed to have perfect accuracy [27] (which is also common with map reference data [54])

- 118
 4. TD errors may propagate to downstream products in surprising and potentially harmful ways (e.g. leading to bad decisions) and can occur without the map producer and/or map user's knowledge. This problem is particularly relevant in the common case where TD and reference data are collected using the same methods, and/or in cases where map reference data error is not known or accounted for, which is still common [54]
- These problems suggest a pressing need to review the issues surrounding TD quality and how it impacts ML-generated maps, and to recommend a set of best practices and standards for minimizing and accounting for those errors, which are the primary aims of this paper. Although map error can also originate from other sources, such as the specific ML classifier selected or the parameterization approach used [28,55,56], we focus solely on issues of input data quality. As such, this paper complements existing work focused on assessing final map accuracy [34–38,41,42].

130 This paper is organized into four sections. In section 1, we review current practices in the 131 treatment of TD for categorical and continuous map creation. We also cover map accuracy 132 procedures, given that the two processes are often intertwined and affected by many of the same 133 issues [47], and accuracy assessment procedures are needed to assess the impacts of TD error. In 134 section 2, identify the most common sources of TD error and inconsistency. In section 3, we illustrate 135 the impacts of uncertainty in TD generation with case studies that span a range of typical EO 136 applications, including building and road mapping, global surface flux estimates, and mapping 137 agricultural systems. In section 4, we propose guidelines for i) best practice in collecting and using 138 TD, including definition of acceptable accuracy levels, ii) minimizing TD errors associated with 139 training sample design error and collection, iii) characterizing and incorporating TD error in final 140 map outputs, and iv) communicating TD error in scientific and public documentation.

141 1.1 Current Trends in Training Data Collection

A large proportion of remote sensing projects make some use of TD, typically created either using geolocated *in situ* data [43,57], by visually interpreting high and/or very high resolution spatial resolution imagery [23,58,59], or by interpreting the images to be classified/modeled themselves [e.g. 52,53,60,61]. Of these collection methods, image interpretation is increasingly common [62], particularly with the rise in crowdsourcing initiatives [19,63]. As such, mapping is strongly constrained by the creation of TD, which (much like map reference data) are often treated as absolute 'truth', if for no other reason than that their accuracy is assumed to be perfect [27,35,44,64]. However, multiple sources of error are possible and indeed likely in TD, whether collected *in situ* or via imageinterpretation [57].

151 The use of large, data-intensive ML algorithms continues to grow in many fields, including 152 remote sensing. Neural Networks (NN)s represent an increasingly used class of ML algorithms, with 153 more complex NNs such as Convolutional Neural Networks (CNN) producing higher output 154 accuracy [65]. While some forms of ML can function effectively with smaller training datasets, the 155 quality of these data is nevertheless critically important [25,28,48]. Additionally, the increasingly 156 popular large-scale, high-complexity NNs require substantially more TD than traditional statistical 157 models, and like many ML approaches are sensitive to noisy and biased data, producing the logistical 158 difficulty of creating very large, 'clean' training datasets [66-68].

Partially to address this need, several recent efforts have been devoted to producing extremely large training datasets that can be used across a wide range of mapping applications, and to serve as comprehensive benchmarks [69,70]. Similarly, a recent trend has emerged in large-scale mapping projects to employ large teams of TD interpreters, often within citizen science campaigns that rely on web-based data creation tools [19,71–73].

164 1.2 Characterizing Training Data Error

165 Due to different disciplinary lineages, terminology associated with the various datasets used to 166 train and evaluate map algorithms is sometimes contradictory or distinct. Here we harmonize 167 terminology by defining four distinct types of data: training, validation, training reference, and map 168 reference. Training data (TD) refers to a sample of observations, typically consisting of points or 169 polygons, that relate image pixels and/or objects to semantic labels. Validation data are typically a 170 random subset of TD that are withheld and used to fit ML model parameters and internally evaluate 171 performance. Training reference data are expert-defined exemplar observations used to assess TD 172 errors during or after data creation. Map reference data are independent observations used to assess 173 final map accuracy; while these may be collected using many of the same procedures as the other 174 three datasets [54], they have more stringent design protocols and can only be used to assess the final 175 map product, rather than used iteratively in model or map improvement [54]. Map reference data are 176 often referred to as the test set in ML literature [74], but we use the former term to align with the 177 terminology commonly used by the EO community.

178 1.2.1 Map Accuracy Assessment Procedures

Map accuracy assessment practices and standards are well-established in the EO literature [36,37,42,54,75]. We briefly review these procedures here because they are essential for quantifying how TD error impacts map accuracy. Additionally, the growing use of ML algorithms developed outside of EO has brought with it accuracy assessment practices and terminology that often differ nominally or substantively from those developed for EO [e.g., 76,77,78]. Reviewing EO accuracy assessment standards can therefore help to harmonize and improve accuracy assessment practices, while providing necessary context for procedures that can help to account for TD error.

186 The accuracy of a map is assessed by evaluating the agreement between the values of the 187 mapped variables and those of a map reference variable, and summarizing those discrepancies using 188 an accuracy metric [38,77]. A number of different accuracy metrics can be used, which vary 189 depending on whether the variable of interest is categorical or continuous, with each type of variable 190 having its own foundation for error analysis [79–83]. For categorical variables, this foundation is 191 provided by the confusion matrix, in which rows (but sometimes columns) typically list how many 192 mapped values fall within each category and columns (but sometimes rows) the distribution of map 193 reference values for each category. In EO, the most widely used metrics calculated from the confusion 194 matrix are user's accuracy (the complement of commission error), producer's accuracy (the 195 complement of omission error), and overall accuracy (i.e. the complement of proportion error) [37]. 196 A fuller explanation of accuracy metrics and other aspects of the error matrix can be found in existing 197 publications [32,34,45,69,74–76]. Another widely used measure in EO is the Kappa Index of 198 Agreement [77], but Kappa varies with class prevalence [78] and inappropriately corrects for chance

agreement [77], thus its continued use is strongly discouraged [37,77]. There are a number of other categorical accuracy metrics suitable for assessing the accuracy of a binary categorical variable, such as the F1 score [78], and the True Skill Statistic [84], which are described in the supplemental materials.

203 The scatter plot provides the basis for error analysis for continuous variables, wherein deviations 204 between the mapped values plotted on the Y-axis are measured against those of the map reference 205 on the X-axis. Several measures are used to summarize these deviations (see SI). The Root Mean 206 Squared Error (RMSE, a.k.a. Root Mean Square Deviation, RMSD) and Mean Absolute Deviation 207 (MAD) summarize deviations along the 1:1 line. The former has widespread use, but we recommend 208 caution since it combines MAD with variation among the deviations [85-87]. Another widely used 209 measure is the R², or coefficient of determination, but this measures deviation relative to the linear 210 regression line, rather than the 1:1 line [80,85].

Beyond these, there are measures for comparing continuous mapped variables to a binary reference variable, including the Receiver Operating Characteristic (ROC) and the Total Operating Characteristic (TOC) [81,88,89]. The area under this curve (AUC) of an ROC/TOC plot is often used as a single measure of overall accuracy that summarizes numerous thresholds for the continuous variable [89]. There are also metrics for assessing the accuracy of object-based image analysis [OBIA, 90], which we do not cover here (but see the SI) because the choice of measure varies according to mapping objectives [62,91].

218 The creation of the map reference sample is an integral part of the accuracy assessment process 219 and has two major aspects. The first of these is the design of the sample itself (i.e. the placement of 220 sample units), which should be probability-based but can follow several different designs (e.g. simple 221 random, stratified, cluster, systematic) depending on the application and a priori knowledge of the 222 study area [77,92]. The second aspect is the response design, which governs the procedures for 223 assigning values to the map reference samples [77,92]. These include the choice of the sample's spatial 224 and temporal units, the source of the data that the sample extracts from (e.g. high resolution imagery), 225 and the procedure for converting reference data values into map-relevant values [77,92]. For a 226 categorical map in which the reference data source is high resolution imagery, the map reference 227 sample is assigned labels corresponding to the map legend (e.g. a land cover scheme) based on a 228 human supervisor's interpretation of the imagery [77,92].

229 A key aspect of response design is that map reference data should be substantially more accurate 230 than the map being assessed, even though they are always likely to have some uncertainty 231 [27,43,44,77,92]. This uncertainty should be measured and factored into the accuracy assessment 232 [43,92]. However, in practice this accounting is rarely done, while map reference data uncertainty is 233 also rarely examined [31,77,93]. This tendency is illustrated by Ye et al. [53], who reviewed 209 journal 234 articles focused on object-based image analysis, finding that one third gave incomplete information 235 about the sample design and size of their map reference data, let alone any mention of error within 236 the sample. Errors in map reference data can bias the map accuracy assessment [42,90], as well as 237 estimates derived from the confusion matrix, such as land cover class proportions and their standard 238 errors [41]. To correct for such impacts to map accuracy assessment, one can use published accuracy 239 assessment procedures, including variance estimators, that account for map reference error [35,43,44]. 240 These approaches depend on quantifying errors in the map reference data.

241 1.2.2 Current approaches for assessing and accounting for training data error

242 Most of the aforementioned considerations regarding map reference data creation largely 243 apply to TD, particularly since map reference data and TD may often be collected together [e.g. 52], 244 provided the former are kept strictly separate to ensure their independence [77]. Considerations 245 regarding TD may diverge with respect to sample design, as TD often needs to be collected in ways 246 that deviate from probability-based sampling, in order to satisfy algorithm-specific requirements 247 related to, for example, class balance and representativeness or the size of the training sample 248 [28,48]. Another difference is that map TD can be usable even with substantial error [45,47,48]--249 although we show in Section 3 that TD error can propagate substantial map error--whereas map

reference data needs to have the highest possible accuracy and its uncertainty should be quantified,

251 as described above [43,77,92].

250

252 If the quality of map reference data is often unexamined, TD quality may be even less so. To 253 gain further insight into the level of attention TD receives in EO studies, we reviewed 30 recent, top-254 ranked¹ research papers describing land cover mapping studies (identified via keyword search on 255 Google Scholar) [2,60,61,94–117]. This assessment showed that only 2 papers explicitly assessed the 256 quality of the TD used in classification, while 16 made no mention of TD standards at all. Over 75% 257 of these studies used image interpretation, as opposed to *in situ* data, in either training, accuracy 258 assessment, or both. One-quarter of these papers used unsupervised classifiers in the processing 259 chain to outline training areas, followed by image interpretation to assign labels to the 260 polygons/pixels. Although only a snapshot, this finding suggests that key details regarding the 261 design and collection of TD (and even map reference data) is lacking in the EO literature.

262 Even though TD quality appears to be largely unreported, efforts have been made to examine 263 how TD error can impact ML-based classifications, typically within the context of evaluating specific 264 algorithms. For example, research examining the effectiveness of Random Forests [118] for land cover 265 classification also evaluated their sensitivity to TD error, sample size, and class imbalance [45,48,119], 266 and similar work was done for Support Vector Machines [25,29,49]. Several studies comparing 267 multiple ML algorithms also compared how each reacted variations in TD sample size and/or error 268 [47,56,120,121]. Maxwell et al. [28] touch on a number of these TD quality issues in an even broader 269 review of ML algorithms widely used in EO classification, but excluding newer deep learning 270 approaches.

271 Beyond these examples, several studies have focused more explicitly on how to train ML-272 algorithms for remote sensing classification when TD error is present. Foody et al. [27] conducted 273 tests to examine how two different types of TD labelling error impacted land cover classifications, 274 with a primary interest in SVM. Similarly, Mellor et al.'s [45] study measured uncertainty introduced 275 by TD error in a RandomForest classifier, with specific focus on class imbalance and labelling errors. 276 Swan et al. [46] examined how increasing amounts of error introduced into the TD for a deep learning 277 model impacted its accuracy in identifying building footprints. These studies collectively 278 demonstrate that TD has substantial impact on ML-generated maps. They also reveal that there is no 279 standard, widely accepted practice for assessing TD error, which, similar to map reference data, is 280 generally not reported and thus implicitly treated as error-free [27].

281 2. Sources and Impacts of Training Data Error

282 In the following two sections we describe the common causes of TD error and explore its 283 potential impacts. To describe causes, we divide the sources of TD error into two general classes: 1) 284 errors stemming from the design of the training sample, which include many aspects of both sample 285 and response design as described in existing literature on accuracy assessment (see 1.2.1 above), and 286 2) errors made when collecting the training sample, including the process of digitizing and labeling 287 points or polygons when interpreting imagery, or collecting field measurements. In addressing the 288 impacts of error, we provide a summary of potential problems, and then two concrete case examples 289 for illustrative purposes.

- 290 2.1 Sources of Training Data Error
- 291 2.1.1 Design-related errors

With respect to TD sampling design, errors primarily relate to failures to adequately represent the spatio-temporal-spectral domains of the features of interest in the manner most suited to the specific ML algorithm being used [50]. This causes a disparity between the distribution of TD

¹Based on the Google Scholar search algorithm results. Search performed January, 2019, with terms land cover and land use, including permutations of spelling and punctuation. Twenty-seven articles kept after initial screening for relevance.

compared to the true distribution of the mapped phenomenon in geographic and/or feature space
[25–28]. This problem is highly relevant in ML approaches, which are sensitive to TD quality,
including class balance, labelling accuracy, and class comprehensiveness relative to the study area's
true composition [27].

299 Temporal unrepresentativeness is also a common source of error in the response design of TD, 300 due to the prevalence of image interpretation as a source for TD. In this case, error arises when 301 obsolete imagery is interpreted to collect training points or polygons and their associated labels 302 [36,58]. The problem is illustrated in Figure 1, which contrasts smallholder fields that are clearly 303 visible in a satellite base map (Bing Maps) with ground data collected in 2018. Center pivot fields 304 were installed after the base map imagery was collected, but before ground data collection, creating 305 a temporal mismatch between the base map and the *in situ* data. Labels generated from the base map 306 would therefore introduce substantial error into an ML algorithm classifying more recent imagery. 307 New HR satellites that have more frequent acquisitions [e.g. PlanetScope, 122] can help minimize 308 such temporal gaps for projects that are designed to map present-day conditions (e.g. 2018 land 309 cover), but cannot solve this problem for mapping projects covering earlier time periods (i.e. before 310 2016). The same can be said for aerial and Unmanned Aerial Vehicle acquisitions, which are typically 311 limited in geographic and temporal extent [123]. While hardcopy historical maps can help 312 supplement temporal data gaps, these data sources come with their own problems, such as errors 313 introduced during scanning and co-registration, and unknown production standards and 314 undocumented mapping uncertainties.

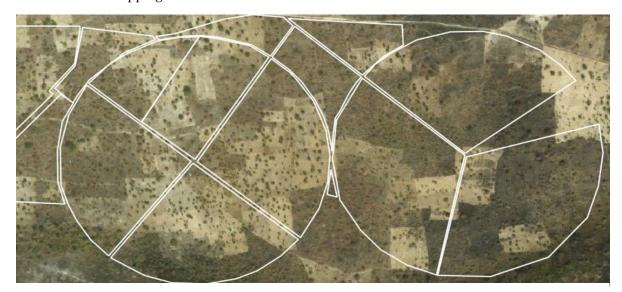


Figure 1. An example of potential training data error that can arise when image interpretation is conducted on older imagery. The underlying imagery is from Bing Maps, which shows smallholder agricultural fields near Kulpawn, Ghana. The white polygons were collected by a team of mappers (hired by Meridia) on the ground using handheld GPS in 2018. The smallholder fields were replaced by larger center-pivot irrigation fields sometime after the imagery in the base map was collected.

315 Spatial alignment can be a substantial source of response design-error when training with HR 316 and VHR commercial satellite imagery. Due to their narrow swath widths, HR/VHR sensors are often 317 tasked, resulting in substantially off-nadir image acquisitions [58]. Due to large view zenith angles 318 and the lack of adequate digital elevation models, side overlapping imagery for stereo 319 photogrammetry, or other relevant control points, HR/VHR imagery often does not meet the same 320 orthorectification standards as coarser resolution, government operated satellites [124–126]. When 321 integrating HR/VHR imagery acquired at different azimuth and elevation angles, features such as 322 building roofs show offsets similar to those caused by topography. These offsets are particularly 323 problematic for a) training repeated mappings of the same features, and/or b) when using an existing 324 vector dataset such as OSM as TD [127-129].

325 TD collected by interpreting HR/VHR imagery is often co-registered with the coarser resolution 326 imagery used as ML model data. This creates a potential spatial resolution conflict because the 327 relationship between image objects and pixel size may be different, where objects delineated as 328 spectrally homogenous areas in HR/VHR imagery may be part of mixed pixels in moderate or coarse 329 resolution model imagery. This mis-match is similar to the concept of H-resolution versus L-330 resolution scene models proposed by Strahler et al. [130]; in H-resolution models, the objects of 331 interest are substantially larger than the pixel size, and vice versa for L-resolution models. The 332 incorporation of mixed pixels may degrade classification model performance, or at least introduce 333 undesired spectral variability within classes [121,131,132]. This situation may be alleviated by 334 displaying both HR/VHR imagery and/or other ancillary datasets as well as coarser model imagery 335 during training data creation [133,134]. However, such practices may not be possible when training 336 data are taken from previous research projects, or when they are to be applied in the context of time 337 series analysis, in which spatial features change over time [e.g. 135].

Similar spatial resolution and scaling issues must be dealt with when combining *in situ* measurements with satellite observations for continuous variables. Field-collected data often cannot practically cover the entire area of a pixel in the model data, especially for moderate or coarse resolution imagery, and can thus induce scaling errors related to the modifiable areal unit problem [136,137]. Spatial representativeness assessments and interpolation methods are used to limit this problem for operational EO science products [138–141], but this issue is likely to be a source of error for most *in situ* TD samples.

Another design-related problem arises from large-scale data collection initiatives that are becoming increasingly common due to the expanding extent of modern EO analyses [e.g. 142]. These efforts, often conducted via crowdsourcing campaigns, typically enlist citizens to collect data a webbased platform [63,e.g. 143–145]. Examples include OpenStreetMap (OSM), Geo-Wiki [63], Collect Earth [146], DIYLandcover [144], and FotoQuest Go [147]. In cases where the resulting data might be purely voluntary [73], the resulting sample may lack spatial representativeness due to uneven geographic contributions [25,148].

352 2.1.2 Collection-related errors

353 There are several common forms of error that occur when collecting both TD and map reference 354 data. The first of these are errors of interpretation [36], which are mistakes created in the process of 355 manual image interpretation. Image interpretation is widely used to generate TD, and often does not 356 yields inconsistent labels between interpreters [31,34,149,150]. Interpreters may lack experience in the 357 task, or be unfamiliar with the context of the study area [e.g. 151]. In an unusually thorough analysis 358 of error in image interpretation, Powell et al. [149] showed that inter-interpreter agreement was on 359 average 86% but ranged from 46 to 92%, depending on land cover. This research, which relied on 360 trained image interpreters, concluded that transitional land cover classes produce substantial 361 interpretation uncertainty, which is particularly problematic since much land cover mapping effort 362 is directed towards change detection. Another image interpretation study that used a crowdsourcing 363 platform found that interpreters' average accuracy in digitizing crop field boundaries in high 364 resolution imagery was ~80%, based on comparisons against training reference data [144]. This result 365 held true whether the interpreters mapped several hundred sites or <50 (Figure 2), indicating that 366 increased interpreter experience does not necessarily eliminate labelling error, even when analysts 367 are highly seasoned [149]. These findings underscore the need to assess uncertainty in TD, as well as

368 map reference data, using predefined
369 training reference data or inter-interpreter
370 comparisons [43,57,149,152,153].

371 Labeling error may also result from 372 inadequate or poorly communicated 373 semantic class definitions [154,155], 374 particularly when identifying land use, as 375 opposed to land cover [156]. This is especially 376 evident in urban environments, which not 377 only exhibit high spatial and spectral 378 heterogeneity, even within HR/VHR imagery 379 [157], but are also are also semantically vague 380 (i.e. hard to define) at the ground level. For 381 example, Figure 3 shows a typical example of 382 TD collection for mapping informal 383 settlements (a.k.a slums), in Nairobi, Kenya, 384 in which several trained interpreters 385 separately delineate the same area [158]. 386 Because slums may defined be bv 387 sociodemographic factors in addition to 388 spatial and spectral properties, TD creation 389 for such areas is prone to error stemming 390 from semantic issues [155]. Complex classes

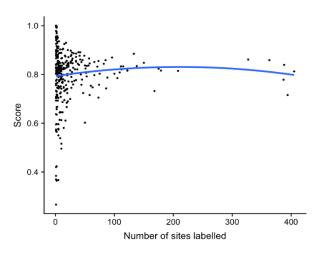


Figure 2: Number of sites mapped per worker versus the average score received at reference sites, where workers' maps were compared to reference maps using a built-in accuracy assessment protocol within a crowdsourcing platform for collect cropland data (Estes et al., 2016).

391 such as slums may exhibit high variability between study areas, as local idiosyncrasies link the 392 definition of slums to different physical, remotely observable characteristics. These characteristics 393 make it hard to develop a generalizable mapping capability for informal settlements. These results 394 further illustrate the importance of consensus mapping for image interpretation, particularly for 395 heterogeneous target classes with vague or regionally idiosyncratic semantic definitions.

396 Categorical mapping projects typically define a crisp set of non-overlapping categories, rather 397 than a fuzzy set [159,160]. However, many human and natural land covers exhibit continuous 398 gradation between classes, implying that crisp map legends will necessarily cause semantic 399 ambiguity for when image pixels in areas that are transitional between land cover types are labelled 400 [161,162]. This problem is particularly acute with moderate and coarse resolution imagery [23]. When 401 scene objects approximate the spatial dimension of the image resolution, local variance is highest, 402 leading to poor classification accuracy [163]. While substantial research has been devoted to the issue 403 of mixed pixels [83,131,132,164–166], crisp categories are still often relied on during the training and 404 testing phases of image classification [167], although less crisp approaches based on fuzzy set theory 405 are available [160,168]. Labelling errors can also arise if analysts are not properly trained on class 406 definitions, or by poor data creation standards, such as failure to capture full metadata while in the 407 field or during digitization, e.g. pertaining to difficult-to-determine cases or potential confusion

408 between spectrally, spatially, or conceptually similar classes [156]. Such inadequacies limit the 409 analysis of TD error, and therefore the ability to account for error propagation.

410 Collection-related errors may be particularly acute in large-scale crowdsourcing campaigns or 411 citizen science initiatives, which are increasingly valued for mapping projects due to their larger size 412 and cheaper acquisition costs [19,63,144,145]. Such datasets are often collected rapidly and entail

413 labeling many observations over a short period of time by participants who are not domain experts

414 [147,169]. In such cases, label quality is a function of interpreter skill, experience, contextual

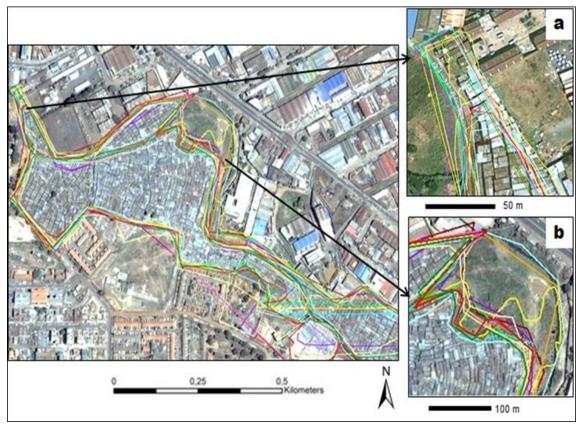


Figure 3: The challenges of mapping slum extent from image interpretation in Nairobi, Kenya. Each colored line indicates a different analyst's delineation of the same slum, illustrating semantic confusion. Adapted with permission from Kohli et al. (2016).

415 knowledge, personal interest, and motivation for involvement in the data collection [19]. Errors can 416 be exacerbated if interpreters are inadequately trained or unfamiliar with the study area, or lack 417 experience with EO data and methods. For example, delineation of different classes of urban land use 418 may be extremely difficult without the benefit of local knowledge [155]. Furthermore, participants 419 may be required to interpret HR/VHR satellite imagery collected over multiple sensors and dates and 420 having varying quality (e.g. cloud cover percentage and atmospheric correction) and view/sun angles 421 [170], which further complicates interpretation. Inadequate or confusing user interfaces may also lead 422 to error [19,155]. Once crowdsourced/citizen science data have been post-processed for noise, they 423 can be highly detailed and spatially extensive [63][66-68]. Nevertheless, quality problems in such 424 datasets can be particularly hard to find and clean, and are thus an important source of TD error that 425 may propagate through ML algorithms into map outputs [54,145,171]. Therefore, these data should 426 be used more cautiously than expert-derived TD.

Errors also arise in *in situ* TD, caused by measurement error, geolocation inaccuracy, incorrect identification of relevant objects (e.g. vegetation species), and other such mistakes [172]. In addition to these factors, some feature types may also be difficult to discern on the ground [27]. Aside from these problems, there are many sources of technologically induced errors, such as defects in the software or hardware of measurement devices, user input error, or calibration errors (e.g. in spectroradiometers or other equipment). However, accounting for quantitative measurement error is more
straightforward than thematic TD creation. Textbook tools to quantify measurement error are widely
available, and *in situ* data collection procedures often include inter-analyst measurement
comparison [173,174].

436 2.2 Impacts of Training Data Error

TD errors carry through to impact the map production process and outcomes. From a design perspective, the size and class composition of TD is particularly impactful on ML algorithms, which are susceptible to overfitting and class imbalance problems [28,70]. Additionally, the assumption of representativeness of training pixels is often overstated, and many TD may in fact not be generalizable to broader scales (discussed by Tuia et al. [148]). TD errors arising from the collection process also impacts maps. Both design and collection-related errors may be particularly hard to discern, or quantify in absolute terms, if the error in the map reference data errors are unknown.

444 Several studies reviewed in Section 1.2.2 provide insight into how much TD error can impact 445 ML-generated land cover maps, focusing on aspects of sample size and balance (design-related 446 errors) and labelling error (collection-related error). This work shows that the impact of each error 447 source varies according to the algorithm used. For example, support vector machines (SVMs) were 448 relatively insensitive to changes in sample size, dropping by only 3-6% under TD size reductions of 449 85-94% [25,175]. RandomForests (RF) also proved robust to, but slightly more affected by, TD sample 450 size, showing accuracy drops of ~4-10+% when TD was reduced by 70-99% [45,48,175]. Sample size 451 also impacts the certainty of RF classification by lowering the mean margin (a measure of certainty 452 related to the number of class votes) by ~50% for sample size reductions of 95% [45]. In contrast to 453 SVM and RF, maps classified with single decision trees are highly affected by TD size, with 13% 454 accuracy loss for TD reductions of 85% [25] all the way up 50-85% loss when TD size reductions of 455 50-70% [48,56]. Other models tested for TD sample size sensitivity include a neural network based on 456 adaptive resonance theory, which had accuracy reductions of ~30 to ~65% when TD samples were 457 halved [56], while a feed-forward neural network lost just 2% accuracy when TD was reduced by 85% 458 [25].

459 Classifiers are also sensitive to class balance within the training data. For example, the accuracy 460 of RF-generated maps declined by ~12 to ~23% and classification confidence fell ~25 to ~50% when 461 TD class balances were highly skewed [45]. Notably, the ranges in these accuracy and confidence 462 declines were attributable to differing TD sample sizes, showing the synergistic effect of sample size 463 and class balance sensitivities. Maxwell et al. [28] provide a more comprehensive review of class 464 imbalance for RF, SVM, artificial neural networks, and k-nearest neighbors (kNN), finding that all 465 models were sensitive to class imbalance, but the impact was most felt in the accuracy for rare classes 466 rather overall map accuracy.

467 The impact of TD labeling errors, also referred to as noise, vary substantially between mapping 468 algorithms. SVMs and closely related derivatives appear least sensitive to mislabeling. SVMs lost just 469 0-5% in the accuracy or kappa of land cover classifications when 20-30% of TD samples were 470 mislabeled either randomly or uniformly across classes [27,49,120]. Relative vector machines (RVMs) 471 were even less sensitive under these conditions [2.5% accuracy loss for 20% mislabelling; 27], and an 472 SVM designed specifically for handling noisy TD (context-sensitive semi-supervised SVM) was even 473 more robust [2.4% reduction in kappa for 28% mislabelling; 49]. However, the impact of TD noise 474 was greater for all three models when mislabeling was confined to specific classes. SVMs lost 9% 475 accuracy and 31% kappa when 20-28% of samples in spectrally similar classes were mislabeled 476 [27,49]. The RVM showed a 6% accuracy loss [27] and specialized SVM an 12% kappa reduction [49] 477 under the same conditions. As with sample size, RF is the next least sensitive to TD noise [45,48]. 478 Mislabeling 25% of TD samples reduced RF accuracy by 3-7% for a binary classifier and 7-10% for a 479 multiclass model, with the ranges in accuracy loss also varying according to TD sample size [45]. 480 Classification certainty was more heavily impacted by label error, dropping by 45-55%, as measured 481 by the mean margin [45]. Other classification models showed larger impacts due to label noise,

including 11-41% kappa declines for a kNN [28% label noise; 49], and 24% [120,176] and 40-43%
accuracy loss for a kernel perceptron and neural network, respectively, that were each trained with
30% of TD labelled incorrectly [56,120,176]. Single decision tree models were most sensitive to label
error, registering 39 to nearly 70% accuracy declines for 30% label noise [56,120,176].

486 The aforementioned work provides substantial information on how TD error can impact the 487 accuracy and certainty of older-generation ML classifiers. Further understanding of the consequences 488 of these errors can be inferred from literature examining the impact of errors in map reference data. 489 Map reference errors can substantially bias areal estimates of land cover classes, as well as the 490 estimation of variance in those classes, particularly when examining land cover change [43,177,178]. 491 While methods exist to incorporate map reference data error into map accuracy assessments and area 492 estimates [35,43,44], and also to account for TD uncertainty in assessing classifier accuracy [45], there 493 has been little work that shows how to address both TD and map reference error.

494 Less information is available regarding how TD error might propagate beyond the map it 495 initially creates. Some insight can be found in a study that examined how error propagates from a 496 primary land cover map into subsequent derived products [30]. This work used a high-quality 497 reference cropland map to quantify the errors in 1 km cropland fractions derived from existing land 498 cover datasets, and measured how these errors propagated in several map-based analyses that drew 499 on cropland fractions for inputs. The results suggest that downstream errors were in some instances 500 (e.g. carbon stock estimates, Figure 4) several fold larger than those in the input cropland maps, 501 whereas in other cases (e.g. evapotranspiration estimates) errors were muted. In either case, the 502 degree to which the error magnifies or reduces in subsequent maps is hard to anticipate, and the high 503 likelihood that it could have the former effect means that any conclusions based on such land cover-504 derived maps must be treated with caution if the error propagation is not quantified. This analysis 505 suggests how TD errors might impact the maps they generate and provides a potential method for 506 quantifying their impacts on map accuracy.

507

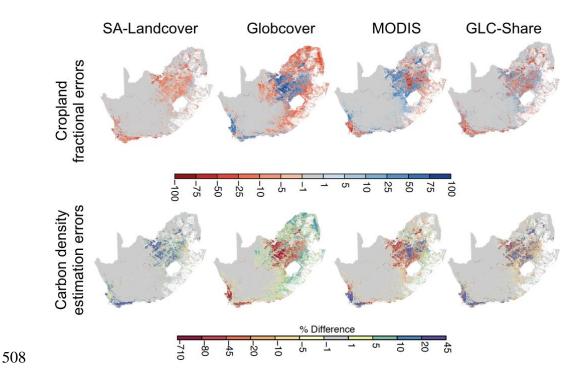


Figure 4: An examination of how error in pixel-wise cropland fractional estimates (expressed as a percentage, top row) can propagate error (expressed as a percentage) in maps that use land cover data as inputs, such as estimates of carbon density (bottom row). Figure adapted from Estes et al., (2018).

509 Another example illustrating the impact of map input errors is seen in the practice of using well-510 known standard datasets, such as the National Land Cover Map [NLCD, 179], to map quantities of 511 interest, such as urban tree canopy biomass. Urban trees play a crucial role but in regional carbon 512 cycles [180–182] but are often omitted from EO studies of carbon dynamics [e.g., MODIS NPP, 183]. 513 As urban lands are expected to triple between 2000 and 2030 [184,185], the need to factor them into 514 carbon accounting is pressing, but remotely mapping urban tree cover is limited by a) spatial 515 resolutions that are too coarse for highly variable urban landscapes and b) TD that are often biased 516 to forested, agricultural, and other rural landscapes. For these reasons, the Landsat-derived NLCD 517 Percent Tree Cover (PTC) product [186], which estimates canopy cover at 30-m resolution across the 518 U.S, provides a practical input for empirical models used to map tree biomass. However, previous 519 studies showed uncertainty of this product in urban areas [186], and a tendency to underestimate 520 urban canopy cover compared to a high resolution dataset. Therefore, to quantify the potential impact 521 of NLCD PTC error on canopy biomass estimates, we compared the accuracy of the NLCD PTC 522 dataset to canopy cover estimates derived from manually digitized VHR Imagery for a suburb of 523 Washington, D.C., USA. We found that NLCD PTC underestimated canopy cover by 15.9%, 524 particularly along forest edges (Figure 5) where it underestimated canopy cover by 27%. This 525 discrepancy is particularly important in heterogeneous urban landscapes, where forest edges 526 comprise a high proportion of total forest area. Scaling field data from forest plots to the entire study

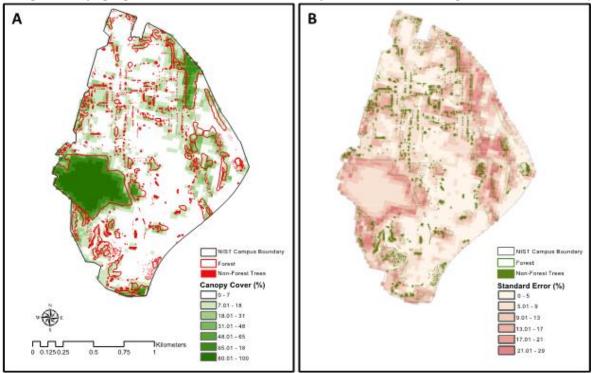


Figure 5: Spatial variations in canopy cover (A) and uncertainty in canopy cover estimates (B) in forested and non-forested areas of the heterogeneous suburban landscape of the National Institute of Standards and Technology campus in Gaithersburg, Maryland. Percent canopy cover at a 30-m resolution from the commonly used National Land Cover Database (NLCD) Percent Canopy Cover product (and its uncertainty) is superimposed over a high-resolution map of forested areas (hollow outlined polygons) and non-forest trees (e.g., street trees; solid polygons) that were manually mapped using <1-m resolution Wayback World Imagery. Note the lower estimates of percent canopy cover along forest edges (A) and the associated higher levels of uncertainty (B) using the NLCD product.

yielded an estimate of 8,164 Mg C stored in aboveground forest biomass, based on our manually
digitized canopy cover map, compared to only 5,960 Mg C based on the NLCD PTC. This finding

529 indicates the significance of these map errors for carbon accounting, as temperate forest carbon

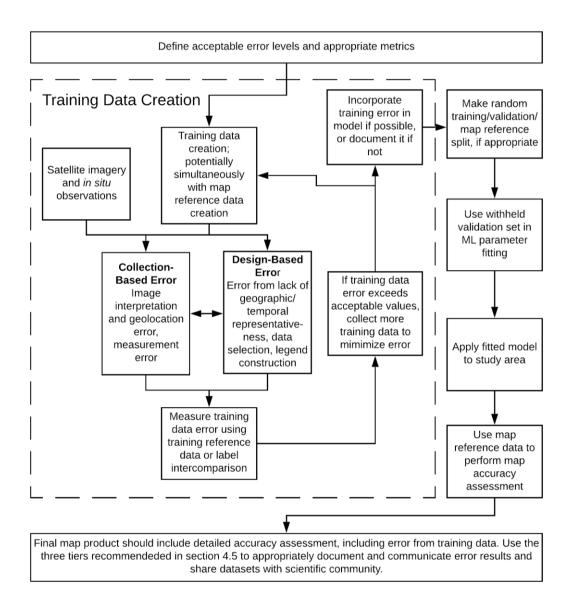


Figure 6: Flow chart of typical workflow for machine learning applications in Earth observation data.

530 storage and rates of sequestration are much larger (64% and 89%, respectively) than in forest interiors

[187]. Quantifying errors in the NLCD is thus important for correcting subsequent estimates trainedon these data.

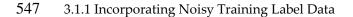
533 These brief examples help illustrate the potential problems of TD error, but the range of potential 534 impacts is as varied as the number of mapping projects underway across academic research, 535 commercial operations, and the public sphere. To represent the growing set of remote sensing 536 applications in which TD error may be encountered, we present a set of case studies below. To help 537 lay a common framework, we show a typical methods sequence for a ML-based remote sensing 538 analysis in Figure 6, which also helps clarify the terminology used in this paper. The figure shows 539 the various sources and implications of error in the modeling and mapping process, beginning with 540 issues in the data sources and sample design, and continuing through model training, validation, and 541 ultimately in map accuracy assessment.

542 3. Case Studies

543 To better illustrate the potential impact of TD error, we provide several case studies across 544 different mapping applications that represent the broad range of ML-based mapping and modeling

545 applications that rely on TD.

546 3.1 Infrastructure Mapping



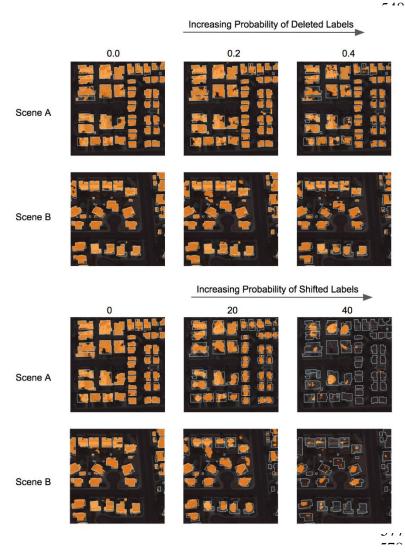


Figure 7: Predictions of the model trained on different noisy datasets. Each row shows a single scene over different noise levels. The top two rows show noisy drops, while the bottom two rows show noisy shifts. The ground truth is outlined in light blue, and the predictions are filled in orange.

304

585 affects the accuracy of the resulting model.

To measure the relationship between label noise and model accuracy, the amount of label noise was varied while holding all other variables constant. To do this, an off-the-shelf dataset (the SpaceNet Vegas buildings data set) was used in place of OSM, into which label errors were systematically introduced. To this relatively large training data set (~30,000 labeled buildings)⁴, missing and imprecisely drawn building errors were systematically introduced, and the resulting model accuracy was measured. The experimental design consisted of two series of six datasets each,

592 with random deletion or shift of buildings at increasing probabilities and magnitudes, respectively.

Automated building footprint detection is an important but difficult mapping task, potentially benefiting a wide range of applications. The following case study illustrates the use of Raster Vision², an open source deep learning framework, to train several models for automated building detection from high resolution imagery³. These models perform best when trained on a large number of correctly labeled examples, usually generated by a paid team of professional labelers. An alternative, less costly approach was conducted which а building in segmentation model was trained using labels extracted from **OpenStreetMap** (OSM). However, the labeled training polygons from OSM generated errors: contain some buildings are missing, and others are poorly aligned with the imagery or This missing details. provides a good test case for experimentation on how noise in the labels

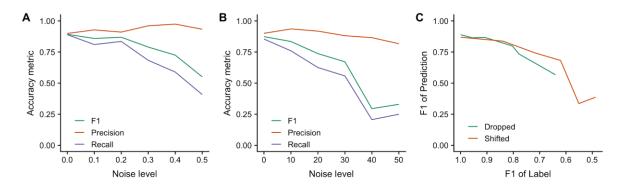
² <u>https://rastervision.io/</u>

³ Additional detail available at: <u>https://www.azavea.com/blog/2019/08/05/noisy-labels-deep-learning/</u>

⁴ <u>https://spacenetchallenge.github.io/datasets/spacenetBuildings-V2summary.html</u>

593 For each dataset, a UNet semantic segmentation model with a ResNet18 backbone was trained using 594 the fastai/PyTorch plugin for Raster Vision⁵. These experiments, including data preparation and 595 visualization, can be replicated using code at <u>https://github.com/azavea/raster-vision-</u> 596 <u>experiments/tree/master/noisy_buildings_semseg</u>.

597 Figure 7 shows the ground truth and predictions for a variety of scenes and noise levels, showing 598 that the quality of the predictions decreases with the noise level. Also, the background and central 599 portions of buildings tend to be predicted correctly, whereas the outer periphery of buildings 600 presents a greater challenge. These results are quantified in Figure 8, which shows F1, precision, and 601 recall values for each of the noise levels below (see Table S1 for terminology description). The 602 precision falls more slowly than recall (and even increases for noisy drops), which is consistent with 603 the pattern of errors observed in the prediction plots. Pixels that are predicted as building tend to be 604 in the central portion of buildings, leading to high precision.



605

Figure 8: The precision, recall, and F1 scores across different noise levels are shown for the cases in which labels are randomly dropped (A) or randomly shifted (B).

606 In panels (A) and (B) of Figure 8, the x-axis shows the noise from randomly dropped and 607 randomly shifted labels, respectively. Panel (C) combines the effects of noisy deletions and noisy 608 shifts on accuracy in a single graph, showing F1 of the labels on the x-axis and F1 of the prediction 609 on the y-axis. The F1 score of the noisy versus ground truth labels is a function of the pixel-wise 610 errors; this metric has the benefit of measuring the effect of noise on error in a way that is comparable 611 across datasets and object classes. For instance, a noisy shift of 10 in a dataset with large buildings 612 might result in a different proportion of erroneous label pixels than in another dataset with small 613 buildings. From this, panel (C) shows that while some of the shifted datasets have a greater level of 614 noise, the prediction F1 scores are similar between the two series when the noise level is similar.

615 These experiments present a small step toward answering the question: how much accuracy is 616 sacrificed by using TD from OSM? Preliminary results indicate, as expected, that accuracy decreases 617 as noise increases and that the model becomes more conservative as the noise level increases, only 618 predicting central portions of buildings. Furthermore, the noisy shift experiments suggest that the 619 relationship between noise level and accuracy is nonlinear. Future work will quantify the functional 620 form of this relationship, and how it varies with the size of the training set. Some preliminary work 621 toward this goal has been described in Rolnick et al. [188], which focuses on image classification of 622 Imagenet-style images.

One limitation of these results is that the magnitude of error in OSM for most areas is unknown, making it difficult to predict the effect of using OSM labels to train models in a generalized, global sense. "Noisy" error in OSM can be estimated by measuring the disparity between OSM labels to clean labels, such as the SpaceNet labels used in this case, providing a local estimate of OSM noise. A more general but less rigorous approach is to roughly estimate the noise level by visually inspecting the labels in OSM, and comparing to Figure 7, which shows examples of the labels at different noise levels.

⁵ <u>https://github.com/azavea/raster-vision-fastai-plugin</u>

630 3.1.2 Detecting Roads from Satellite Imagery

Road networks constitute a critical geographical data layer used to assist national decision makers in resource allocation, infrastructure planning, vaccination campaigns, and disaster response, among others. However, accurate and up-to-date road networks are not available in many developing countries. High resolution satellite imagery, paired with deep learning methods, provides the capacity to detect and map roads at large spatial scales. This important goal, however, is dependent on availability of local high-quality TD.

637 To evaluate the impact of local TD availability on predicted road network accuracy, a study was 638 carried out in Kumasi, Ghana [189]. Two datasets were used to train ML models: 1) the SpaceNet⁶ 639 dataset [190] in Khartoum, Sudan, and Las Vegas, USA, and 2) OSM data in Kumasi, Ghana. The 640 SpaceNet Dataset includes high quality road labels with human expert validation, but unfortunately 641 was not available in Kumasi, Ghana. Therefore, the latter study site relied on OSM data, consisting 642 of crowdsourced labels with no accuracy assessment or expert validation. A series of experiments 643 were carried out to assess the feasibility of using transfer learning, using the Raster Vision Python 644 library for training and evaluation. For all MobileNet V2 models introduced in the following list, the

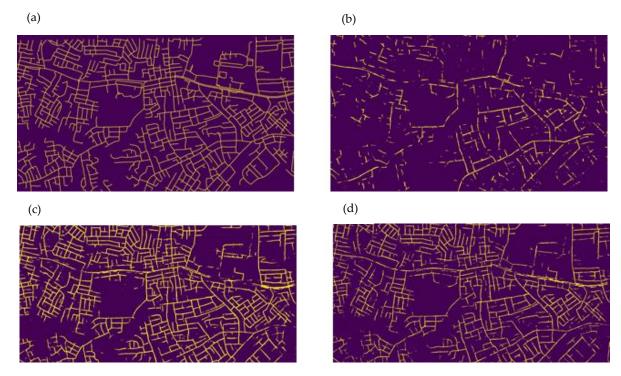


Figure 9: (a) Labels generated by experts for validation. (b) Predictions from the Khartoum Model. (c) Predictions from Kumasi Model. (d) Predictions from Khartoum Model retrained in Kumasi with 10K steps.

- 645 image chip size was set to 300 x 300 pixels, and the training/validation split was 80/20.
- The Las Vegas Model was trained and validated on SpaceNet data in Las Vegas and produced
- 647 very high accuracy predictions. However, when this model was used in Kumasi, it predicted very
- 648 few roads, with only scattered road segments. The Khartoum Model was also trained using SpaceNet
- data in Khartoum. The Kumasi Model used Maxar WorldView-3 imagery and labels from OSM as
- input. OSM was used to test the quality of crowdsourced labels in training a road detection model.The Khartoum Model was then fine-tuned on OSM labels in Kumasi for three different steps of 100K.
- The Khartoum Model was then fine-tuned on OSM labels in Kumasi for three different steps of 100K,
 50K and 10K. All models used the same hyperparameters, to isolate the role of TD on model
- 653 performances.

⁶ <u>https://spacenetchallenge.github.io/</u>

654 To validate the models' performance using an independent dataset, a set of expert labels were 655 generated over a small part of Kumasi. Figure 9 shows the region with human expert data vetting, 656 along with the three model predictions. The Las Vegas model is excluded from this figure as it does 657 not have any meaningful prediction in Kumasi. Quantitative performance metrics were calculated 658 using the human expert labels, which the models had been blind to during training. The results 659 indicate that, as shown by Figure 9, the F1 score for roads was substantially higher for the Kumasi 660 Model (0.6458) than when using the Khartoum model (0.3780). However, by retraining and fine-661 tuning the Khartoum model, the F1 score for roads increased to 0.6135. The full accuracy results for 662 this experiment are presented in Table S2.

Based on these results, it is concluded that: 1) lack of diverse TD significantly limits the geographic applicability of models, as the types, surfaces, and arrangements of roads varies substantially between regions; 2) regional training datasets are essential for the model to learn the feature of roads in that region; and 3) transfer learning from a reasonably similar geography can help train models.

668 3.2 Global Surface Flux Estimates

Fluxes at the land-atmosphere boundary play a key role in regulating water, carbon and energy cycles. These fluxes include latent heat flux (LE), sensible heat flux (H), and gross primary production (GPP). While these fluxes cannot be measured directly from remote sensing observations, other remotely sensed variables can be used to estimate these fluxes. Moreover, these three fluxes are highly coupled, and therefore a coupled model is ideal.

- 674 A fully connected neural network model was developed for this purpose [191], named Water, 675 Energy, and Carbon with Artificial Neural Networks (WECANN). Inputs to WECANN are remotely 676 sensed estimates of precipitation, soil moisture, net radiation, snow water equivalent, air temperature 677 and solar induced fluorescence. The target variables for training the model were derived from 678 outputs of global models. However, this presents the difficulty that the target variables are model 679 outputs that can have substantial error, which will propagate in the WECANN model. To mitigate 680 this problem, three independent estimates of each of the three fluxes (LE, H and GPP) were retrieved 681 from the global models. Then a novel statistical approach, named Triple Collocation (TC, Figure S1, 682 equation S1), was used to combine those estimates to a new dataset for training the WECANN model.
- Triple collocation (TC) is a technique for estimating the unknown error (measured with standard deviations or RMSEs) of three mutually independent measurement systems, without treating any one system as zero-error "truth" [192]. The three measurement systems estimate a variable collocated in space and time, hence called Triple Collocation. Using these probabilities, at each pixel and at each time one of the three estimates of the target variable is randomly selected to generate the TD.
- 688 The results of WECANN model outputs were evaluated against ground measurements from 689 global FLUXNET towers from 2007 to 2015 (Figure 10), using both the coefficient of determination 690 and Root-Mean-Squared-Error (RMSE) to evaluate accuracy. These show that WECANN's 691 correlation was on average 17% higher (range 8-51%) than that of any one of the three individual 692 inputs, while the RMSE was 21% lower (range 4-54%). These differences provide a partial 693 quantification of the error inherent in any one of these training inputs and show that by combining 694 them using the TC technique, we can reduce error in an ML model for predicting the fluxes at global 695 scale. This case study illustrates a means of assessing and accounting for error in training data for 696 cases in which these data are not created specifically for the project, but rather are pre-existing data 697 products with potentially quite different characteristics and potentially unknown error.
- 698

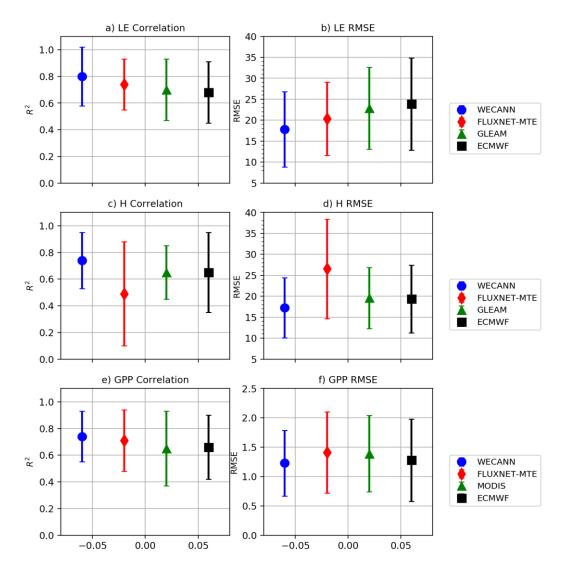


Figure 10: R² and RMSE of the WECANN model output against ground measurements from FLUXNET towers in comparison to the three datasets used to generate the target training data for LE (a, b), H (c, d) and GPP (e, f).

700 Two agricultural cases illustrate how TD error can impact both categorical and quantitative 701 remotely sensed measures. The first relates to cropland mapping and is drawn from an ongoing study 702 focused on mapping smallholder agricultural fields at high spatial resolution (3-4 m) in Ghana. The 703 mapping method is based on "active learning", in which a RandomForest-based [118,193,194] ML 704 algorithm is iteratively trained and validated by a crowdsourcing platform, which enlists human 705 trainers to visually interpret and digitize field boundaries visible within the imagery (PlanetScope 706 visual and near-infrared surface reflectance [122]) being classified [143,144,194]. The crowdsourcing 707 platform incorporates a protocol for assessing the accuracy of training labels, in which each worker 708 is periodically directed to a training reference site where the boundaries are already known but are 709 not visible to the worker. Using these training reference sites, their maps are then scored using a 710 multi-dimensional accuracy assessment algorithm [144], resulting in an average TD accuracy score 711 for each worker that ranges between 0 (complete disagreement with reference) and 1 (perfect 712 agreement). Each label site is mapped by at least five workers, and the resulting worker-specific 713 accuracy scores are used within a Bayesian merging algorithm to combine the five sets of labels into 714 a single consensus label, which is then used to train the RandomForest classifier. Here we use the 715 worker-specific training accuracy scores to assess the impact of label quality on map accuracy, by 716 assessing three variants of two RandomForest-generated maps, one over Central Ghana (~3,400 km²) 717 and one over Northern Ghana (~3,100 km²). The first two maps were trained using labels generated

by the least accurate worker to map each training site, the second two by the most accurate worker to map each site, and the third using the consensus labels. The accuracy of each pair of maps was then assessed against the validation set (reserved consensus labels) using the True Skill Statistic [84] (sensitivity + specificity - 1, with scores ranging from -1 to 1). The results show a substantial difference in accuracy between the maps trained with the least and most accurate workers' labels (Figure 11A), with the former having 7-9% more skill than the latter, while maps based on consensus labels have ~3% more skill than those of the most accurate workers' labels.

725 The second case relates to remotely sensed crop estimates of wheat yields collected in 48 726 smallholder fields in Bihar, India in 2016-17 [195]. Yield data were collected via eight 2x1 m² crop cuts 727 within each field, and PlanetScope-derived green chlorophyll vegetation indices (GCVI) were 728 calculated over each field from imagery collected over four dates during the growing season (January 729 13, February 25, March 12, and April 14, 2017). A RandomForest regression was trained on the yield 730 measured for each field, using the four dates of GCVI values as predictors. To test the effect of TD 731 error on the resulting yield predictions, three types of noise were artificially introduced into the yield 732 data used for training: 1) a systematic 0.5 ton/ha overestimate, and randomly distributed errors 733 sampled from a normal distribution with a mean of 0 ton/ha and 2) standard deviations of 0.5 ton/ha

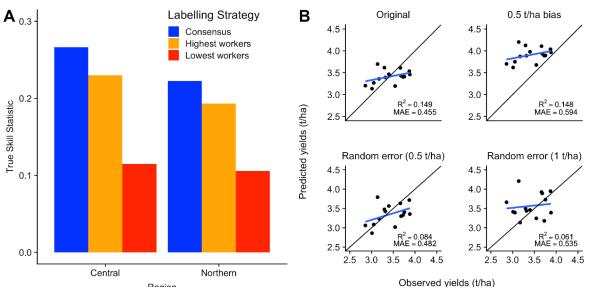


Figure 11. A comparison (A) of the accuracy (based on the True Skill Statistic) of cropland maps over two areas of Ghana when generated by labels of different levels of quality (red = least accurate workers' labels; orange = most accurate workers' labels; blue = "consensus" labels made by merging all workers' labels). (B) Results from a RandomForest model of wheat yields trained on satellite-derived vegetation indices, showing the relationship between predicted yield and independent observed yields, in terms of the fit against the 1:1 line and the regression slope of the relationship (points and regression line represent the mean of a single randomly selected model permutation). The average mean absolute error (MAE) and average regression R²s calculated across all permutations are shown for each model.

and 3) 1 ton/ha. A baseline model fit to unperturbed data was also developed. Each model was trained on three separate randomly selected subsets of 32 perturbed observations, and the predictions were made for the remaining 16 held-out (independent) yield observations, which were not perturbed. This three-fold cross validation process was repeated 50 times, with each permutation using a different random seed to construct the folds, in order to achieve stable error metrics. The model performance was assessed by calculating the averages of the mean absolute error (MAE) of the prediction, and the R² of regressions fit between prediction and observed values (Figure 11B).

The results show that four models, including the baseline, compressed the range of yields, as seen in the shallow slope between observed versus predicted values, but prediction error was 18-31%

higher when training yields had either the high level of random or systematic error within them. The

smaller amount of random noise only added about 6% error to the predictions, suggesting that RandomForest is tolerant to some training error. Note that the average R² of the observed-predicted regression fit was nearly the same for the systematic error case as the baseline, which shows that this metric can be an unreliable measure of performance for quantitative measures, and that it is important to assess fit against the 1:1 line and using a metric such as mean absolute error.

749 4. Guidelines and Recommendations

Our review and case studies show that the impacts of TD error on EO applications can vary, as well as the procedures for assessing those impacts, but several best practices and guidelines can be discerned within this work. Below we synthesize a set of suggested steps for minimizing and accounting for TD error, within the context of undertaking and assessing the accuracy of a typical ML-based mapping project.

755 4.1. Step 1: Define acceptable level of accuracy and choose appropriate metric

756 As a starting point, mapmakers should determine the minimum level of accuracy needed for 757 their application, using the accuracy metric(s) most appropriate for answering their questions [196]. 758 For example, if the goal for creating a categorical map is to obtain an unbiased area estimate for a 759 particular land cover, it is essential to account for the map's commission and omission errors by 760 adjusting the proportional area estimate of the cover type derived from the map by the proportion of 761 that type estimated from the map reference sample [37,77,92]. For a continuous variable in which the 762 absolute accuracy of the mapped variable is most important, then the mean absolute deviation from 763 the 1:1 line is more informative than R^2 [86,87].

764 Error in the map reference data should also be factored in the selected accuracy metrics and the 765 resulting map-derived measures. Several published methods exist for categorical data (see Section 766 1.2.1). For continuous variables, the fit between the mapped and map reference variables can be 767 assessed using Type 2 regression, which allows for error in the dependent (map reference) variable 768 [197], unlike the more commonly used Type 1 regression. Determining map reference data error is 769 critical to determining overall map accuracy. The error in these data effectively determines the upper 770 limit of achievable map accuracy, as it is difficult (but not impossible; see [44]) to know whether a 771 model's predictions are more accurate than its map reference data; if the map reference data are only 772 90% accurate, then the map can be at most 90% accurate. Acceptable accuracy should thus be 773 determined relative to the accuracy of the map reference data, rather than the implicit assumption of 774 100%, which is widely used since map reference data are usually considered perfect [35,36,44,54,64]. 775 Although the aforementioned steps relate primarily to concerns about map accuracy assessment,

Although the aforementioned steps relate primarily to concerns about map accuracy assessment, they are essential to establishing best practices for map training. For one, without undertaking rigorous accuracy assessment as described above, it is not possible to fully assess how TD error impacts map accuracy. Secondly, the processes of map reference data and TD generation are often tightly intertwined and impacted by many of the same sources of error (see Sections 1.2.1-2). The procedures for minimizing and measuring errors in both datasets are thus often the same. Our subsequent recommendations therefore cover both training and map reference datasets, except where we indicate necessary distinctions.

783 4.2. Step 2: Minimize design-related errors

784 The next logical step in a mapping project is to design strategies for independently collecting the 785 training and map reference samples. Although there are numerous factors to consider, there are 786 several general aspects of design that can help minimize potential TD errors.

787 4.2.1 Sample design

788 The first of these relates to sampling design itself, meaning where, when, how many, and what 789 type of samples are placed (e.g. simple random, clustered, stratified, systematic). With respect to the

790 TD, to a certain extent this depends on the requirements of the selected ML algorithm, which can, for

791 example, have differing requirements with respect to geographic dispersal [50] and class balance [e.g. 792 28,45,78]. Geographic representativeness and the degree to which they capture the variability in the 793 feature of interest is an important TD sample design consideration [50,58,144,198]. The road mapping 794 case study shows the errors that can result when maps are trained with samples that do not 795 adequately represent the features in a particular region. TD can in practice be highly localized or 796 relevant for a limited spatial extent or temporal period [155,189]. This problem may become more 797 relevant given the increase in stock or benchmark training libraries, and attempts to transfer pre-798 trained models to other regions, time periods, or scales of observation [70,199]. While such training 799 libraries can be of immense benefit to large extent EO research, if these are to be relied on for training, 800 their representativeness of the features of interest should be assessed, and augmented as needed, as 801 in the Khartoum model case study (Figure 9D). For some widely used ML algorithms, such as 802 RandomForests, the best practice appears to be to train with data collected within the mapping region 803 [e.g. within a particular agroecoregion, 52,200], and to avoid over-generalizing or transferring models 804 to other regions [201]. However, until more published studies are available, it is not clear whether 805 this rule applies to deep learning models. When using citizen science or crowdsourcing approaches 806 to generate these data, representativeness is ensured by directing labellers to the selected TD sites 807 [e.g. 144], rather than having them select regions to map.

808 Samples should also be temporally representative of the imagery that is being classified [58]. 809 That is, relative to the imagery being classified, the TD (and map reference) sample should be 810 collected within a window of time that matches the characteristic rate of change of the feature being 811 mapped. This interval can be estimated by measuring the temporal autocorrelation in the feature of 812 interest [202]. For rapidly changing phenomena, such as deforestation events, snow/ice melt, and 813 vegetation coverage during phenological transition, the sample may need to be captured within a 814 few days or weeks of the acquisition of the imagery being classified, whereas for slower-moving 815 features a sample collected within a few years may be sufficient.

816 In cases where training and reference samples are generated simultaneously, it is essential that 817 TD sample design not undermine the standards required for an independent, probabilistic map 818 reference sample [sensu 64]. Stehman and Foody [54] describe procedures for ensuring the 819 independence of the map reference sample in such cases. Beyond those considerations, it is important 820 to note that the map reference sample's independence is compromised when it is used to iteratively 821 refine the mapping algorithm. This problem can best be understood within the context of cross 822 validation, which is appropriate for ML parameter tuning [e.g. 28]. However, when the number of 823 folds exceed one (as in our yield estimation case study; Figure 11B) then the portions excluded from 824 training lose statistical independence and can no longer serve as the map reference [74]. Map 825 reference data independence may also be undermined when training sites are selected iteratively, in 826 order to increase their representativeness and improve ML performance [52,e.g. 143]. If the gain due 827 to new training sites is assessed against the map reference, then it will also lose independence after 828 the first iteration. Moreover, any error in the map reference sample will be integrated into the final 829 map. Xiong et al. [52] avoided this problem by visually assessing whether their classifier improved 830 map quality after having new TD points added to the initial sample. A more quantitative approach 831 is to divide an initial sample into three splits: one for training, the second for validating algorithm 832 improvements, including those related to the addition of new training sites, and the third as the map 833 reference, used only for final accuracy assessment. This partitioning approach can be implemented 834 in the mapping platform used in the cropland mapping case study [Figure 11A, 193].

835 4.2.2 Training Data Sources

The requirements for temporal representativeness make the source of training imagery a critical consideration for projects that rely on image interpretation. The use of basemap imagery is not recommended for training maps of dynamic features, given their broad range and uneven distribution of image ages [58], unless the age of the imagery being classified can be matched to that of the training imagery. Otherwise, there is substantial potential for introducing error into the mapping algorithm (e.g. Figure 1), and its impact may be hard to assess, particularly if the map 842 reference sample is collected from the basemap. The goal of temporal representativeness must be

balanced with the need to have a sufficiently high spatial resolution for accurate image interpretation,

844 which helps minimize errors during the collection of the sample (see Step 3). Beyond matters of cost,

845 this tradeoff is one reason why HR/VHR basemaps are widely used [58]. New commercial imagery, 846 such as PlanetScope [122], which are collected at high temporal frequency (near-daily) with a spatial

846 such as PlanetScope [122], which are collected at high temporal frequency (near-daily) with a spatial 847 resolution sufficient for many visual interpretation tasks (3-4 m), may be a preferable source of

848 training imagery for developing maps representing the post-2016 period. Finally, in designing an

- 849 image-based sample, it is also important to consider additional characteristics that can influence
- 850 interpreters' judgement, such as atmospheric quality (e.g. clouds, haze), sensor view angle, sun angle,
- 851 spectral band selection, and image contrast stretches [71].

852 4.2.3 Legend design

853 For thematic maps, legend design merits special consideration as it relates to TD, particularly 854 for multi-temporal and/or large area projects that rely on multiple image datasets [58]. As discussed 855 in section 2 above, objects of interest, including land cover types, should be at least twice as large as 856 the pixel resolution of the imagery used in the classification algorithm, assuming a requirement for 857 spectrally pure pixels [130,163,203]. When image spatial resolution is too coarse relative to the scene 858 elements of interest, image interpretation errors are likely due to mixed pixels [121,131,132]. This 859 implies that in designing a legend, researchers should select classes that can be mapped effectively 860 using the coarsest resolution imagery that will be incorporated in the model, and avoid the problem 861 of collecting training samples with mixed pixels [e.g. 52]. This consideration is particularly relevant 862 since HR/VHR imagery is often used to create TD and map reference data, while the mapping 863 algorithm is applied to moderate or coarse resolution imagery [e.g. 52,114,204,205]. Alternatively, 864 researchers may opt to select a classification workflow which explicitly incorporates mixed pixels 865 [91,160,e.g. 168].

Continuous TD, particularly those collected *in situ*, are often point samples, and therefore a sampling protocol should be used to match field measurements and pixel dimensions in order to avoid scaling problems associated with the modifiable areal unit problem [136,137]. Spatial representativeness should be considered as a limiting factor for legend design [50], and to the extent possible, researchers should attempt to use categories that are supported by both the spatial resolution of the model data and the field sampling protocols to be used; we recommend that researchers consult the extensive literature on legend design [22,138–141,206–208].

873 4.3. Step 3: Minimize collection-related errors

There are numerous ways to collect TD for categorical and continuous mapping projects, each with their own sources of error. There are thus many potential approaches for minimizing the associated collection errors, which may be quite specific to a particular variable [e.g. for agricultural area estimates 209]. However, there are several general approaches that can be followed to minimize TD collection errors. Our focus here is primarily on error in image-interpreted TD, which is one of the most common approaches used to training ML mapping algorithms. We also touch on the specific save of model-derived training data.

881 Whenever possible, we recommend using protocols that incorporate training reference data to 882 independently assess TD accuracy, particularly for image-interpreted TD [e.g. 144]. Training 883 reference datasets can be limited in size compared to the ultimate sample size, provided that training 884 reference locations are randomly presented to interpreters during the data creation campaign 885 [144]. Active feedback during training label creation can also help reduce errors on a rolling basis, by 886 providing interpreters information regarding their performance [169].

If comparison against training reference data is not possible, then consensus methods for generating TD may be the next best alternative. Consensus among several domain experts may also be the best and most practical measure for collecting both training reference data and map reference data [31,54]. In the case of image-interpreted samples, consensus approaches should employ multiple

891 interpreters to label the same site. For continuous variables, several independent or repeated *in situ*

892 measurements should be made and averaged. For modeled variables where the error is unknown, as 893 in the surface flux case study, training based on the outputs of multiple independent models is 894 recommended. The agricultural case study shows how multiple mappings can be used to quantify 895 label uncertainty (Figure 12A) and minimize the amount of labeling error, which manifests through 896 improved map accuracy (Figure 11A). The surface flux case study demonstrates these same benefits 897 across several continuous variables (Figure 10). The number of separate measures or interpreters to 898 use will vary depending on the application. Some guidance for image-interpreted tasks comes from 899 the land cover accuracy assessment literature, where consensus between at least 3 interpreters is 900 recommended to allow for majority voting [31,43], but more complex land covers may need up to 7 901 interpreters [43]. In the cropland mapping case study, 5 interpreters labelled each consensus training 902 sample. For the continuous surface flux example, 3 separate modeled inputs were used.

903 Further steps can be taken to minimize TD collection errors arising from image interpretation. 904 Interpreters should be given thorough training regarding the task [31], which may include instruction 905 on remote sensing principles as well as local or regional contextual information. Local domain 906 expertise is particularly helpful for consistent identification of idiosyncratic land covers [158]. 907 Interpreter education is particularly important for crowdsourcing or citizen science data collection 908 campaigns, as participants typically lack formal experience in image interpretation [145,210].

909 As described in Step 2 above, image interpretation is inadvisable when the available imagery 910 does not support the legend categories in terms of spatial, spectral, temporal, or radiometric 911 resolution [211-213]. Researchers must be especially cautious in the similar but potentially more 912 hazardous case that HR/VHR imagery is used to create training samples that are then used with 913 coarser resolution imagery when ingested into the ML model. Assuming that researchers correctly 914 specify their data selection and legend design when using higher spatial resolution imagery to create 915 TD, image interpretation errors due to insufficient resolution should be minimized; however, special 916 care should be given to borderline classes, or classes exhibiting a high degree of spatial and/or 917 spectral variability due to land cover mixtures within the pixel [121,131,132,148,214]. In such cases, 918 we recommend that training polygons be created near the center of scene objects, where pixel mixing 919 is likely to be minimized [e.g. 52].

920 Another important error-minimizing approach relates to cases where TD comes from a process 921 model, as in the surface flux example outlined above. Process models are also increasingly used to 922 train crop yield mapping models, due to the difficulty of obtaining sufficiently large and reliable 923 field-scale yield data for training [215]. To circumvent this challenge, the Scalable Yield Mapping 924 (SCYM) method [216,217] uses a mechanistic crop model to simulate yields under various 925 environmental and management conditions. The model's outputs then become inputs for training an 926 empirical mapping model (typically ML), in which the simulated yield is the dependent variable and 927 a subset of remotely retrievable model variables serve as predictors. TD errors in such cases can be 928 minimized by rigorously calibrating the process model (itself a challenging task) using best practices 929 from the relevant modeling literature [e.g. 218]. Alternatively, if modeled TD are necessary but 930 careful calibration is not possible (e.g. because the data are pre-existing), then a merging approach 931 such as Triple Collocation (Section 4.2) can help reduce training error.

932 4.4. Step 4. Assess error in training data error

The best way to assess both TD (and map reference data) error is to measure it directly. For continuous variables, calculating measurement error should be possible in many cases, even for model-generated data, in which the variance can be calculated from simulation treatments [e.g. 218]. For categorical mapping, label error can be measured using an internal accuracy assessment protocol that makes use of predefined training reference data (e.g. Estes et al., [144]).

However, it can be challenging to produce training reference data, and indeed in some cases the true category is not clear, whether looking at an image or standing on site. In these cases, or when a direct TD error measurement protocol is not available, we recommend that researchers calculate uncertainty estimates based on repeated measures or multiple interpreter approaches [e.g. the crowd standard deviation; 145] described in Step 3 above (and see Figure 12); this is useful for both training 943 and map reference data. We also recommend that additional measures relating to data collection 944 speed, precision, and consistency be collected for individual data creators, as these can generate 945 further insight into relative TD errors. This recommendation is based on experience in crowdsourced 946 data creation [144,145], but it is applicable to any type of data collection, and could greatly bolster the 947 understanding and quantification of error propagation.

948 If it is not possible to either directly quantify or TD error or relative uncertainty, then mapmakers

949 should at a minimum clearly document the data creation methods, and detail likely sources of error

950 and potential uncertainties.

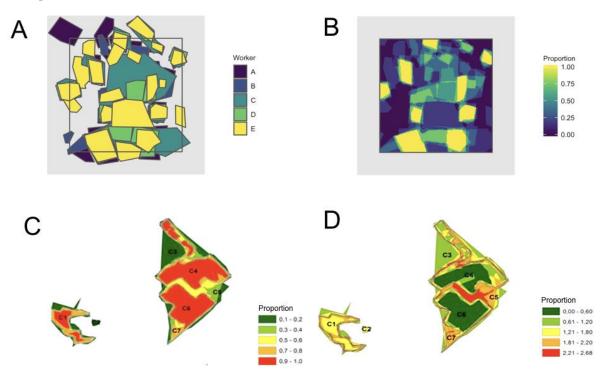


Figure 12: Two examples of consensus-based mapping approaches and their potential use for assessing training (or reference) data uncertainty. Panel A shows a collection of crop field boundary polygons drawn by five independent workers around crop fields visible in PlanetScope imagery collected over Ghana. These labels can be converted into a heat map (B) showing the overall agreement, the inverse of uncertainty. Similarly, 19 independent experts were asked to delineate slum settlements in image subset from Cape Town, South Africa. The polygons are converted into overall agreement and the uncertainty is modeled using random sets (C) shows the covering function, which is then used to calculate standard deviation of random set (D). Both these metrics indicate the variability as well as stability in boundaries delineated by different experts. Adapted with permission from Kohli et al. (2016).

951 4.5. Step 5. Evaluate and communicate the impact of training data error

952 4.5.1 TD Treatment Tiers

953 Due to the wide range of remote sensing research currently underway, a wide variety of TD and 954 classification algorithms are in use. Therefore, it is not possible to specify a single protocol for 955 treatment of TD error. Instead, we outline three tiers that represent different levels of accounting for 956 the impact of TD errors on resulting map products. These three tiers presuppose that mapmakers 957 follow best practices for map accuracy assessment, which includes selecting the most appropriate, 958 literature-recommended accuracy measure(s), quantifying map reference sample error, and 959 accounting for the impact of map reference data error on the accuracy measures (per Step 1). If these 960 best practices are followed, TD error impacts will already be implicitly accounted for within the

961 accuracy measures, and the selected TD accounting tier will be governed by the purposes of the 962 mapping application.

963 Tier 1

964 The optimal TD accuracy assessment, termed Tier 1, involves quantifying TD error using gold 965 standard training reference data (Step 4), and then using that information to quantify varying aspects 966 of TD sample (e.g. class balance, sample size) and collection error (label or measurement error) on 967 model uncertainty and map accuracy, as several initial studies have demonstrated (see Sections 1.2.2 968 and 2.2). For example, the impact of TD error on the certainty of RandomForest classifications can be 969 assessed using measures derived from the margin function [45]. The impact of TD error on map 970 accuracy should also be assessed by training models with TD adjusted to reflect the range of 971 measured TD error, as in our cropland mapping case study, and with respect to variations in TD 972 sample size and class balance [27,e.g. 45,143]. This approach can be used to inform a mapmaker how 973 much map improvement can be obtained by improving TD quality. As such, such tests should be 974 performed against the validation sample to preserve the independence of map reference data.

975 We recommend that developers of benchmark TD libraries adhere to these guidelines. 976 Undertaking such evaluations can provide users important information about appropriate usage of 977 these data for different ML models and application within different mapping geographies. A rigorous 978 quantification of error in the samples themselves is particularly important, since such data are also 979 likely to be used as map reference data. Ideally, this tier should also be followed by the makers of 980 map products intended for widespread public use, along with the release of TD and map reference 981 data that were used [54]. This step would allow users full insight into the quality and usability of the 982 map for their own purposes.

Published TD (and map reference data) should be documented with standard metadata, as
shown in Table S3, including the relevant error metric associated with each observation. The
SpatioTemporal Asset Catalog (STAC⁷) provides a framework for standardization of metadata for
EO data and is increasingly seen as an international standard for geospatial data.

987 Tier 2

988 If it is not possible to directly measure and quantify TD error, the next best approach to account 989 for TD error is to introduce a plausible range of simulated error into the TD and evaluate its impact 990 on model uncertainty and map accuracy after training separate models with the perturbed datasets 991 [e.g. 45]. If multiple workers are tasked with collecting TD for the same site, then the variance in their 992 data can be calculated [e.g. 145] and derive the uncertainty bounds (e.g. Figure 12). This approach is 993 demonstrated in the building mapping case study (section 4.1.1), which illustrates the sensitivity of 994 key accuracy metrics to two different kinds of simulated labelling errors. The wheat yield case study 995 (see section 4.3) provides an example of this approach for a continuous variable.

996 This tier may also provide an acceptable standard for both benchmark datasets and publicly 997 released map products, particularly where absolute error quantification is less important, as well as 998 for publicly released map products. TD and map reference data should also be made openly available 999 with standard metadata, as described above, including the uncertainty metric for each observation. 1000 If it is not possible to publish them (e.g. because of privacy concerns), then researchers should 1001 accompany documentation that summarizes these data and their uncertainty.

1002 Tier 3

1003 If the TD error quantification in Tiers 1 or 2 are not possible, then mapmakers should publish 1004 their TD and map reference data [e.g., 52] with accompanying metadata that includes descriptions of 1005 potential errors and uncertainties. If data cannot be made open, then researchers should at a 1006 minimum publish full descriptions of the potential error in the data. Adherence to this tier, at least

⁷ <u>https://stacspec.org/</u>

1007 the reporting component, should be the minimal standard practice in peer-reviewed, map-based 1008 scientific research.

1009 4.5.1 Communicating error

1010 Finally, uncertainty in ML-generated maps associated with both TD and map reference error 1011 should be faithfully reported within the maps and accompanying documents. Incomplete error 1012 reporting serves to limit the scientific validity and usefulness of these products [54]. Given that ML-1013 generated maps are increasingly used by the public and policy domains, we advise makers of widely 1014 used maps to communicate these uncertainties and their consequences in a manner that is clear and 1015 understandable for broad, including non-specialist, audiences, so that users can understand the map 1016 and its limitations. In general, we recommend including the error on or as close to the actual map 1017 whenever possible, whether by means of metrics, the error matrix, and/or by using cartographic 1018 techniques for representing uncertainty. Examples of effective cartographic techniques for conveying 1019 uncertainty include selection of appropriate, intuitive, and color-blind friendly color schemes for 1020 classes and symbols, varying color value and saturation and font/line weight to indicate levels of 1021 uncertainty, use of crisp versus blurred boundaries and symbols to indicate the range of uncertainty, 1022 or display of consensus maps or side-by-side juxtaposition in cases of multiple, mutually exclusive 1023 predictions for the same place and time (e.g. representing differently specified models) [39,40]. Maps 1024 of consensus in training labels can provide valuable uncertainty information to users, such as shown 1025 in Figure 12A-B.

1026 4.5.2 Towards an Open Training Data Repository

1027 For the scientific community, the ideal standard of openness and replicability is to provide a 1028 complete description of TD collection practices, appropriate accuracy metrics, and perhaps most 1029 importantly of all, the raw data. Ideally, we recommend the creation of a centralized, open source 1030 database of all available and relevant TD, using the details collected in the proposed template (Table 1031 S2), and recorded using the STAC framework. This type of open repository, taking inspiration from 1032 similar large-scale databases for computer vision [ImageNet, 219,SIFT10M Dataset, 220], and remote 1033 sensing [DeepSat, 221,UC Merced Land Use Dataset, 222], should contain full training metadata, 1034 citations to the peer-reviewed literature, as well as links to downloadable versions of TD collection 1035 protocols. Following the philosophy of free and open source software, we strongly recommend that 1036 researchers embrace open source data, which is the only way by which a study can be truly 1037 reproduced.

1038 5. Conclusions

1039 Current practices in EO research are generally inattentive to the need to evaluate and 1040 communicate the impact of TD error on ML-generated maps. This oversight undermines the goals of 1041 scientific reproducibility and may compromise the insights drawn from the resulting maps. 1042 Improving these practices is important given the increasing use of TD-intensive ML algorithms, 1043 which is the goal of our review and its resulting recommendations.

1044 To resolve terminological differences introduced arising from the influence of non-EO 1045 disciplines, and to help contextualize training considerations relative to established map accuracy 1046 assessment practice, we distinguish between four types of "truth" data used in ML-based mapping 1047 projects (training, validation, training reference, and map reference data), and define the appropriate 1048 role for each (Section 1.2). We identify causes of error in TD as well as map reference data, 1049 distinguishing where these vary (Section 2.1), and then assess the impacts of TD error (Section 2.2), 1050 using a set of cases studies to illustrate the consequences TD error across a range of ML-based 1051 mapping applications (Section 3).

We then provide a set of guidelines for minimizing error arising from the design and collection
of TD samples, and recommendations for measuring and accounting for the impact of these errors
(Section 4). Many of these guidelines and procedures also relate to map reference data generation,

1055 and we ground our recommendations in the existing best practices for map accuracy assignment 1056 (Sections 1.2.1 and 4.1). We conclude by defining three tiers of TD error accounting and reporting 1057 standards, which are designed for different kinds of ML-based mapping projects. The highest tiers 1058 should be adopted when creating open training libraries and public map products. Both kinds of 1059 datasets are increasingly being developed to meet the growing demand for EO-derived maps, and to 1060 deal with the growing complexity of EO mapping models. With respect to the latter, there is a 1061 pressing need to rigorously evaluate the training requirements and relative performance of deep 1062 learning models as they become more widely used for EO [33]. This need may be particularly great 1063 for continuous variable applications in Earth System Sciences, such as for hydrological research [223]. 1064 Training datasets that adhere to our suggested accounting standards will help to facilitate such model 1065 comparisons. If adopted within the peer-reviewed literature, these standards may also improve 1066 confidence in scientific findings drawn from map-based research, which can be confounded by 1067 poorly quantified map errors [30,54].

1068

1069 **Author Contributions:** This article synthesizes the ideas of the 20 authors resulting from a workshop focused 1070 on issues of error in training data for Machine Learning approaches in Earth Observation research.

- 1071 Conceptualization, L.E. and A.E.; formal analysis and investigation, L.E., H.A., R.A., J.R.E, L.F., D.K., D.L., A.R.,
- 1072 L.S., S.Y., Z.Y.; writing original draft preparation, A.E., L.E., H.A., L.F., D.K., D.L., R.G.P., A.R., Z.Y.; writing –
- 1073 review and editing, A.E., L.E., K.C., H.A., R.A., L.F., M.F., M.J., D.K., J.C.L.B., J.M., J.R.; visualization, H.A., R.A.,
- 1074 L.F. D.K, A.R., H.S.; supervision, L.E.; project administration, L.E. and A.E.; funding acquisition, L.E., A. R. All
- 1075 authors have read and agreed to the published version of the manuscript.

1076 Acknowledgements

1077 This work represents a synthesis of findings from a workshop held at Clark University on January 8-9, 2019. The 1078 workshop and subsequent paper writing and development was supported by a grant from Omidyar Network's 1079 Property Rights Initiative, now PlaceFund. Additional support for developing methods and data presented here 1080 was provided by NASA (80NSSC18K0158), the National Science Foundation (SES-1801251), National Institute 1081 of Standards and Technology (2017-67003-26615), National Institute of Standards and Technology Summer 1082 Undergraduate Research Fellowship Program, and New York State Department of Environmental Conservation 1083 (DEC01-T00640GG-3350000). We thank Victoria Gammino for helpful input and advice, and David Allen, Ayo 1084 Deas, Lucy Hutyra, Clare Kohler, Barry Logan, Jaret Reblin, Ian Smith for assistance with fieldwork and data

- 1085 compilation.
- 1086

1087 References

- Chen, J.; Chen, J.; Liao, A.; Cao, X.; Chen, L.; Chen, X.; He, C.; Han, G.; Peng, S.; Lu, M.; et
 al. Global Land Cover Mapping at 30 m Resolution: A POK-Based Operational Approach.
 ISPRS J. Photogramm. Remote Sens. 2015, *103*, 7–27.
- Friedl, M.A.; Sulla-Menashe, D.; Tan, B.; Schneider, A.; Ramankutty, N.; Sibley, A.; Huang,
 X. MODIS Collection 5 global land cover: Algorithm refinements and characterization of new
 datasets. *Remote Sens. Environ.* 2010, *114*, 168–182.
- Song, X.-P.; Hansen, M.C.; Stehman, S.V.; Potapov, P.V.; Tyukavina, A.; Vermote, E.F.;
 Townshend, J.R. Global land change from 1982 to 2016. *Nature* 2018, *560*, 639–643.
- Mohanty, B.P.; Cosh, M.H.; Lakshmi, V.; Montzka, C. Soil Moisture Remote Sensing: Stateof-the-Science. *Vadose Zone J.* 2017, *16*.
- 1098 5. Daudt, R.C.; Le Saux, B.; Boulch, A.; Gousseau, Y. Guided Anisotropic Diffusion and
 1099 Iterative Learning for Weakly Supervised Change Detection. *arXiv* [cs.CV] 2019.
- Hecht, R.; Meinel, G.; Buchroithner, M. Automatic identification of building types based on topographic databases a comparison of different data sources. *International Journal of Cartography* 2015, *1*, 18–31.
- Thang, X.; Jayavelu, S.; Liu, L.; Friedl, M.A.; Henebry, G.M.; Liu, Y.; Schaaf, C.B.;
 Richardson, A.D.; Gray, J. Evaluation of land surface phenology from VIIRS data using time
 series of PhenoCam imagery. *Agric. For. Meteorol.* 2018, 256-257, 137–149.
- Tan, B.; Morisette, J.T.; Wolfe, R.E.; Gao, F.; Ederer, G.A.; Nightingale, J.; Pedelty, J.A. An
 Enhanced TIMESAT Algorithm for Estimating Vegetation Phenology Metrics From MODIS
 Data. *IEEE Journal of Selected Topics in Applied Earth Observations and Remote Sensing* 2011, 4, 361–371.
- 1110
 9. Zhang, X.; Friedl, M.A.; Schaaf, C.B. Global vegetation phenology from Moderate Resolution Imaging Spectroradiometer (MODIS): Evaluation of global patterns and comparison with in situ measurements: GLOBAL PHENOLOGY FROM MODIS. J. Geophys. Res. 2006, 111, 981.
- 1114 10. Wan, Z. New refinements and validation of the MODIS Land-Surface Temperature/Emissivity
 products. *Remote Sens. Environ.* 2008, *112*, 59–74.
- 1116
 11. Jiménez-Muñoz, J.C.; Sobrino, J.A.; Skoković, D.; Mattar, C.; Cristóbal, J. Land Surface
 1117
 1118 Temperature Retrieval Methods From Landsat-8 Thermal Infrared Sensor Data. *IEEE*1118 *Geoscience and Remote Sensing Letters* 2014, *11*, 1840–1843.
- 1119
 12. Jean, N.; Burke, M.; Xie, M.; Davis, W.M.; Lobell, D.B.; Ermon, S. Combining satellite imagery and machine learning to predict poverty. *Science* 2016, *353*, 790–794.
- 1121 13. Pekel, J.-F.; Cottam, A.; Gorelick, N.; Belward, A.S. High-resolution mapping of global surface water and its long-term changes. *Nature* 2016, *540*, 418–422.
- 1123 14. Hansen, M.C.; Potapov, P.; Tyukavina, A. Comment on "Tropical forests are a net carbon source based on aboveground measurements of gain and loss." *Science* 2019, *363*.
- 1125 15. Gutierrez-Velez, V.H.; Pontius, R.G. Influence of carbon mapping and land change modelling
 1126 on the prediction of carbon emissions from deforestation. *Environ. Conserv.* 2012, *39*, 325–
 1127 336.
- 1128 16. Deng, J.; Dong, W.; Socher, R.; Li, L.-J.; Li, K.; Fei-Fei, L. Imagenet: A large-scale
 1129 hierarchical image database. In Proceedings of the 2009 IEEE conference on computer vision
 1130 and pattern recognition; Ieee, 2009; pp. 248–255.
- 1131
 17. Helber, P.; Bischke, B.; Dengel, A.; Borth, D. EuroSAT: A Novel Dataset and Deep Learning
 Benchmark for Land Use and Land Cover Classification. *IEEE Journal of Selected Topics in*Applied Earth Observations and Remote Sensing 2019, 1–10.
- 1134
 18. Liu, Q.; Hang, R.; Song, H.; Li, Z. Learning Multiscale Deep Features for High-Resolution
 1135
 Satellite Image Scene Classification. *IEEE Trans. Geosci. Remote Sens.* 2018, 56, 117–126.
- 1136
 19. Laso Bayas, J.C.; Lesiv, M.; Waldner, F.; Schucknecht, A.; Duerauer, M.; See, L.; Fritz, S.;
 1137
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 1138
 <li
- 1139 20. Lary, D.J.; Zewdie, G.K.; Liu, X.; Wu, D.; Levetin, E.; Allee, R.J.; Malakar, N.; Walker, A.;
 1140 Mussa, H.; Mannino, A.; et al. Machine Learning Applications for Earth Observation. In *Earth*
- 1141 *Observation Open Science and Innovation*; Mathieu, P.-P., Aubrecht, C., Eds.; Springer

- 1142 International Publishing: Cham, 2018; pp. 165–218 ISBN 9783319656335.
- 1143
 21. Lary, D.J.; Alavi, A.H.; Gandomi, A.H.; Walker, A.L. Machine learning in geosciences and remote sensing. *Geoscience Frontiers* 2016, *7*, 3–10.
- Loveland, T.R.; Reed, B.C.; Brown, J.F.; Ohlen, D.O.; Zhu, Z.; Yang, L.; Merchant, J.W.
 Development of a global land cover characteristics database and IGBP DISCover from 1 km
 AVHRR data. *Int. J. Remote Sens.* 2000, *21*, 1303–1330.
- Sulla-Menashe, D.; Gray, J.M.; Abercrombie, S.P.; Friedl, M.A. Hierarchical mapping of annual global land cover 2001 to present: The MODIS Collection 6 Land Cover product. *Remote Sens. Environ.* 2019, 222, 183–194.
- 1151 24. Fortier, J.; Rogan, J.; Woodcock, C.E.; Runfola, D.M. Utilizing Temporally Invariant
 1152 Calibration Sites to Classify Multiple Dates and Types of Satellite Imagery. *Photogrammetric*1153 *Engineering & Remote Sensing* 2011, 77, 181–189.
- 1154 25. Foody, G.M.; Mathur, A. Toward intelligent training of supervised image classifications:
 1155 directing training data acquisition for SVM classification. *Remote Sens. Environ.* 2004, 93, 107–117.
- 1157 26. Graves, S.J.; Asner, G.P.; Martin, R.E.; Anderson, C.B.; Colgan, M.S.; Kalantari, L.; Bohlman,
 1158 S.A. Tree Species Abundance Predictions in a Tropical Agricultural Landscape with a
 1159 Supervised Classification Model and Imbalanced Data. *Remote Sensing* 2016, *8*, 161.
- Foody, G.; Pal, M.; Rocchini, D.; Garzon-Lopez, C. The sensitivity of mapping methods to
 reference data quality: Training supervised image classifications with imperfect reference data.
 International Journal of ... 2016.
- 1163 28. Maxwell, A.E.; Warner, T.A.; Fang, F. Implementation of machine-learning classification in remote sensing: an applied review. *Int. J. Remote Sens.* 2018, *39*, 2784–2817.
- Huang, C.; Davis, L.S.; Townshend, J.R.G. An assessment of support vector machines for land cover classification. *Int. J. Remote Sens.* 2002, *23*, 725–749.
- 30. Estes, L.; Chen, P.; Debats, S.; Evans, T.; Ferreira, S.; Kuemmerle, T.; Ragazzo, G.; Sheffield,
 J.; Wolf, A.; Wood, E.; et al. A large-area, spatially continuous assessment of land cover map
 error and its impact on downstream analyses. *Glob. Chang. Biol.* 2018, 24, 322–337.
- 1170
 31. Pengra, B.W.; Stehman, S.V.; Horton, J.A.; Dockter, D.J.; Schroeder, T.A.; Yang, Z.; Cohen,
 W.B.; Healey, S.P.; Loveland, T.R. Quality control and assessment of interpreter consistency
 of annual land cover reference data in an operational national monitoring program. *Remote Sens. Environ.* 2019, 111261.
- 32. Zhu, X.X.; Tuia, D.; Mou, L.; Xia, G.; Zhang, L.; Xu, F.; Fraundorfer, F. Deep Learning in Remote Sensing: A Comprehensive Review and List of Resources. *IEEE Geoscience and Remote Sensing Magazine* 2017, 5, 8–36.
- 33. Ma, L.; Liu, Y.; Zhang, X.; Ye, Y.; Yin, G.; Johnson, B.A. Deep learning in remote sensing applications: A meta-analysis and review. *ISPRS J. Photogramm. Remote Sens.* 2019, *152*, 166–177.
- 1180 34. Foody, G.M. Status of land cover classification accuracy assessment. *Remote Sens. Environ.*1181 2002, 80, 185–201.
- 1182 35. Foody, G.M. Assessing the accuracy of land cover change with imperfect ground reference
 1183 data. *Remote Sensing of Environment* 2010, *114*, 2271–2285.
- 36. Olofsson, P.; Foody, G.M.; Herold, M.; Stehman, S.V.; Woodcock, C.E.; Wulder, M.A. Good
 practices for estimating area and assessing accuracy of land change. *Remote Sens. Environ.*2014, 148, 42–57.
- 37. Pontius, R.G.; Millones, M. Death to Kappa: birth of quantity disagreement and allocation
 disagreement for accuracy assessment. *Int. J. Remote Sens.* 2011, *32*, 4407–4429.
- 1189 38. Congalton, R.G.; Green, K. Assessing the accuracy of remotely sensed data: principles and practices; CRC press, 2008;.
- Monmonier, M. Cartography: uncertainty, interventions, and dynamic display. *Prog. Hum. Geogr.* 2006, *30*, 373–381.
- 1193 40. MacEachren, A.M. Visualizing Uncertain Information. 1 1992, 10–19.
- 41. Goodchild, M.F.; Gopal, S. *The Accuracy Of Spatial Databases*; CRC Press, 1989; ISBN 9780203490235.
- 1196 42. Congalton, R.G. A review of assessing the accuracy of classifications of remotely sensed data.
 1197 *Remote Sens. Environ.* 1991, *37*, 35–46.

- 43. McRoberts, R.E.; Stehman, S.V.; Liknes, G.C.; Næsset, E.; Sannier, C.; Walters, B.F. The
 effects of imperfect reference data on remote sensing-assisted estimators of land cover class
 proportions. *ISPRS J. Photogramm. Remote Sens.* 2018, *142*, 292–300.
- 44. Carlotto, M.J. Effect of errors in ground truth on classification accuracy. *Int. J. Remote Sens.*2009, *30*, 4831–4849.
- 45. Mellor, A.; Boukir, S.; Haywood, A.; Jones, S. Exploring issues of training data imbalance and
 mislabelling on random forest performance for large area land cover classification using the
 ensemble margin. *ISPRS J. Photogramm. Remote Sens.* 2015, 105, 155–168.
- 46. Swan, B.; Laverdiere, M.; Yang, H.L. How Good is Good Enough?: Quantifying the Effects of Training Set Quality. In Proceedings of the Proceedings of the 2Nd ACM SIGSPATIAL
 International Workshop on AI for Geographic Knowledge Discovery; ACM: New York, NY, USA, 2018; pp. 47–51.
- 47. Ghimire, B.; Rogan, J.; Galiano, V.R.; Panday, P.; Neeti, N. An Evaluation of Bagging,
 Boosting, and Random Forests for Land-Cover Classification in Cape Cod, Massachusetts,
 USA. *GISci. Remote Sens.* 2012, 49, 623–643.
- 48. Rodriguez-Galiano, V.F.; Ghimire, B.; Rogan, J.; Chica-Olmo, M.; Rigol-Sanchez, J.P. An
 assessment of the effectiveness of a random forest classifier for land-cover classification. *ISPRS J. Photogramm. Remote Sens.* 2012, 67, 93–104.
- 49. Bruzzone, L.; Persello, C. A Novel Context-Sensitive Semisupervised SVM Classifier Robust to Mislabeled Training Samples. *IEEE Trans. Geosci. Remote Sens.* 2009, 47, 2142–2154.
- 1218 50. Cracknell, M.J.; Reading, A.M. Geological mapping using remote sensing data: A comparison of five machine learning algorithms, their response to variations in the spatial distribution of training data and the use of explicit spatial information. *Comput. Geosci.* 2014, *63*, 22–33.
- 1221 51. Mellor, A.; Boukir, S. Exploring diversity in ensemble classification: Applications in large 1222 area land cover mapping. *ISPRS J. Photogramm. Remote Sens.* **2017**, *129*, 151–161.
- 1223 52. Xiong, J.; Thenkabail, P.S.; Tilton, J.C.; Gumma, M.K.; Teluguntla, P.; Oliphant, A.;
 1224 Congalton, R.G.; Yadav, K.; Gorelick, N. Nominal 30-m Cropland Extent Map of Continental
 1225 Africa by Integrating Pixel-Based and Object-Based Algorithms Using Sentinel-2 and Landsat1226 8 Data on Google Earth Engine. *Remote Sensing* 2017, *9*, 1065.
- 53. Bey, A.; Jetimane, J.; Lisboa, S.N.; Ribeiro, N.; Sitoe, A.; Meyfroidt, P. Mapping smallholder
 and large-scale cropland dynamics with a flexible classification system and pixel-based
 composites in an emerging frontier of Mozambique. *Remote Sens. Environ.* 2020, 239, 111611.
- 1230 54. Stehman, S.V.; Foody, G.M. Key issues in rigorous accuracy assessment of land cover 1231 products. *Remote Sens. Environ.* **2019**, *231*, 111199.
- 1232 55. Zhang, C.; Xie, Z. Object-based Vegetation Mapping in the Kissimmee River Watershed
 1233 Using HyMap Data and Machine Learning Techniques. *Wetlands* 2013, *33*, 233–244.
- 1234 56. Rogan, J.; Franklin, J.; Stow, D.; Miller, J.; Woodcock, C.; Roberts, D. Mapping land-cover
 1235 modifications over large areas: A comparison of machine learning algorithms. *Remote Sens.*1236 *Environ.* 2008, *112*, 2272–2283.
- 1237 57. Copass, C.; Antonova, N.; Kennedy, R. Comparison of Office and Field Techniques for
 1238 Validating Landscape Change Classification in Pacific Northwest National Parks. *Remote*1239 Sensing 2018, 11, 3.
- 1240 58. Lesiv, M.; See, L.; Laso Bayas, J.C.; Sturn, T.; Schepaschenko, D.; Karner, M.; Moorthy, I.;
 1241 McCallum, I.; Fritz, S. Characterizing the Spatial and Temporal Availability of Very High
 1242 Resolution Satellite Imagery in Google Earth and Microsoft Bing Maps as a Source of
 1243 Reference Data. *Land* 2018, 7, 118.
- Biradar, C.M.; Thenkabail, P.S.; Noojipady, P.; Li, Y.; Dheeravath, V.; Turral, H.; Velpuri,
 M.; Gumma, M.K.; Gangalakunta, O.R.P.; Cai, X.L.; et al. A global map of rainfed cropland
 areas (GMRCA) at the end of last millennium using remote sensing. *Int. J. Appl. Earth Obs. Geoinf.* 2009, 11, 114–129.
- Mallinis, G.; Emmanoloudis, D.; Giannakopoulos, V.; Maris, F.; Koutsias, N. Mapping and
 interpreting historical land cover/land use changes in a Natura 2000 site using earth
 observational data: The case of Nestos delta, Greece. *Appl. Geogr.* 2011, *31*, 312–320.
- 1251 61. Jawak, S.D.; Luis, A.J. Improved land cover mapping using high resolution multiangle 8-band
 WorldView-2 satellite remote sensing data. *JARS* 2013, 7, 073573.
- 1253 62. Ye, S.; Pontius, R.G.; Rakshit, R. A review of accuracy assessment for object-based image

- analysis: From per-pixel to per-polygon approaches. *ISPRS J. Photogramm. Remote Sens.* **2018**, *141*, 137–147.
- Fritz, S.; See, L.; Perger, C.; McCallum, I.; Schill, C.; Schepaschenko, D.; Duerauer, M.;
 Karner, M.; Dresel, C.; Laso-Bayas, J.-C.; et al. A global dataset of crowdsourced land cover and land use reference data. *Sci Data* 2017, *4*, 170075.
- 1259 64. Stehman, S.V. Sampling designs for accuracy assessment of land cover. *Int. J. Remote Sens.*1260 2009, 30, 5243–5272.
- 1261 65. Brodrick, P.G.; Davies, A.B.; Asner, G.P. Uncovering Ecological Patterns with Convolutional
 1262 Neural Networks. *Trends Ecol. Evol.* 2019, *34*, 734–745.
- 1263 66. Xiao, T.; Xia, T.; Yang, Y.; Huang, C.; Wang, X. Learning from massive noisy labeled data
 1264 for image classification. In Proceedings of the Proceedings of the IEEE conference on
 1265 computer vision and pattern recognition; cv-foundation.org, 2015; pp. 2691–2699.
- 1266 67. Frénay, B.; Verleysen, M. Classification in the presence of label noise: a survey. *IEEE Trans* 1267 *Neural Netw Learn Syst* 2014, 25, 845–869.
- 1268 68. Brodley, C.E.; Friedl, M.A. Identifying Mislabeled Training Data. 1 1999, 11, 131–167.
- 1269 69. Van Etten, A.; Lindenbaum, D.; Bacastow, T.M. SpaceNet: A Remote Sensing Dataset and
 1270 Challenge Series. *arXiv* [*cs.CV*] 2018.
- 1271 70. Sumbul, G.; Charfuelan, M.; Demir, B.; Markl, V. BigEarthNet: A Large-Scale Benchmark
 1272 Archive For Remote Sensing Image Understanding. *arXiv* [cs.CV] 2019.
- 1273 71. Lesiv, M.; Laso Bayas, J.C.; See, L.; Duerauer, M.; Dahlia, D.; Durando, N.; Hazarika, R.;
 1274 Kumar Sahariah, P.; Vakolyuk, M. 'yana; Blyshchyk, V.; et al. Estimating the global
 1275 distribution of field size using crowdsourcing. *Glob. Chang. Biol.* 2019, *25*, 174–186.
- Fritz, S.; McCallum, I.; Schill, C.; Perger, C.; See, L.; Schepaschenko, D.; van der Velde, M.;
 Kraxner, F.; Obersteiner, M. Geo-Wiki: An Online Platform for Improving Global Land
 Cover. *Environmental Modelling & Software* 2012, *31*, 110–123.
- 1279 73. Goodchild, M.F. Citizens as sensors: the world of volunteered geography. *GeoJournal* 2007, 69, 211–221.
- 1281 74. Kohavi, R.; Others A study of cross-validation and bootstrap for accuracy estimation and
 1282 model selection. In Proceedings of the Ijcai; Montreal, Canada, 1995; Vol. 14, pp. 1137–1145.
- 1283
 1284
 1284
 1284
 1285
 75. Olofsson, P.; Foody, G.M.; Stehman, S.V.; Woodcock, C.E. Making better use of accuracy data in land change studies: Estimating accuracy and area and quantifying uncertainty using stratified estimation. *Remote Sens. Environ.* 2013, *129*, 122–131.
- 1286 76. Catal, C. Performance evaluation metrics for software fault prediction studies. *Acta* 1287 *Polytechnica Hungarica* 2012, *9*, 193–206.
- 1288 77. Stehman, S.V.; Foody, G.M. Key issues in rigorous accuracy assessment of land cover
 1289 products. *Remote Sens. Environ.* 2019, 231, 111199.
- 1290 78. Jeni, L.A.; Cohn, J.F.; De La Torre, F. Facing Imbalanced Data--Recommendations for the Use
 1291 of Performance Metrics. In Proceedings of the 2013 Humaine Association Conference on
 1292 Affective Computing and Intelligent Interaction; 2013; pp. 245–251.
- 1293 79. Kuzera, K.; Pontius, R.G. Importance of Matrix Construction for Multiple-Resolution
 1294 Categorical Map Comparison. *GISci. Remote Sens.* 2008, 45, 249–274.
- 80. Pontius, R.G.; Thontteh, O.; Chen, H. Components of information for multiple resolution
 comparison between maps that share a real variable. *Environ. Ecol. Stat.* 2008, *15*, 111–142.
- 1297 81. Pontius, R.G.; Parmentier, B. Recommendations for using the relative operating characteristic
 (ROC). *Landsc. Ecol.* 2014, 29, 367–382.
- 1299 82. Pontius, R.G. Component intensities to relate difference by category with difference overall.
 1300 *Int. J. Appl. Earth Obs. Geoinf.* 2019, 77, 94–99.
- 1301
 83. Pontius, R.G., Jr.; Connors, J. Range of Categorical Associations for Comparison of Maps with
 1302
 Mixed Pixels. *Photogrammetric Engineering & Remote Sensing* 2009, 75, 963–969.
- 1303 84. Allouche, O.; Tsoar, A.; Kadmon, R. Assessing the Accuracy of Species Distribution Models:
 1304 Prevalence, Kappa and the True Skill Statistic (TSS). J. Appl. Ecol. 2006, 43, 1223–1232.
- 1305
 1305
 85. Willmott, C.J.; Matsuura, K. On the use of dimensioned measures of error to evaluate the performance of spatial interpolators. *Int. J. Geogr. Inf. Sci.* 2006, *20*, 89–102.
- 1307 86. Willmott, C.J.; Matsuura, K.; Robeson, S.M. Ambiguities inherent in sums-of-squares-based 1308 error statistics. *Atmos. Environ.* **2009**, *43*, 749–752.
- 1309 87. Willmott, C.J.; Matsuura, K. Advantages of the mean absolute error (MAE) over the root mean

- 1310 square error (RMSE) in assessing average model performance. *Clim. Res.* 2005, *30*, 79–82.
- 1311 88. Pontius, R.G., Jr; Si, K. The total operating characteristic to measure diagnostic ability for
 1312 multiple thresholds. *Int. J. Geogr. Inf. Sci.* 2014, 28, 570–583.
- 1313 89. Fielding, A.H.; Bell, J.F. A review of methods for the assessment of prediction errors in conservation presence/absence models. *Environ. Conserv.* 1997, 24, 38–49.
- 1315 90. Blaschke, T. Object based image analysis for remote sensing. *ISPRS J. Photogramm. Remote*1316 Sens. 2010, 65, 2–16.
- 1317 91. Costa, H.; Foody, G.M.; Boyd, D.S. Supervised methods of image segmentation accuracy
 1318 assessment in land cover mapping. *Remote Sens. Environ.* 2018, 205, 338–351.
- 1319 92. Olofsson, P.; Foody, G.M.; Herold, M.; Stehman, S.V.; Woodcock, C.E.; Wulder, M.A. Good
 1320 Practices for Estimating Area and Assessing Accuracy of Land Change. *Remote Sens. Environ.*1321 2014, 148, 42–57.
- 1322 93. Foody, G.M. Assessing the accuracy of land cover change with imperfect ground reference
 1323 data. *Remote Sens. Environ.* 2010, *114*, 2271–2285.
- 1324 94. Zhong, B.; Ma, P.; Nie, A.; Yang, A.; Yao, Y.; Lü, W.; Zhang, H.; Liu, Q. Land cover mapping using time series HJ-1/CCD data. *Sci. China Earth Sci.* 2014, *57*, 1790–1799.
- Pacifici, F.; Chini, M.; Emery, W.J. A neural network approach using multi-scale textural
 metrics from very high-resolution panchromatic imagery for urban land-use classification. *Remote Sens. Environ.* 2009, 113, 1276–1292.
- 1329 96. Abbas, I.I.; Muazu, K.M.; Ukoje, J.A.; Others Mapping land use-land cover and change
 1330 detection in Kafur local government, Katsina, Nigeria (1995-2008) using remote sensing and
 1331 GIS. *Research journal of environmental and Earth Sciences* 2010, 2, 6–12.
- 1332 97. Sano, E.E.; Rosa, R.; Brito, J.L.S.; Ferreira, L.G. Land cover mapping of the tropical savanna region in Brazil. *Environ. Monit. Assess.* 2010, *166*, 113–124.
- 1334 98. Hu, T.; Yang, J.; Li, X.; Gong, P. Mapping Urban Land Use by Using Landsat Images and
 1335 Open Social Data. *Remote Sensing* 2016, *8*, 151.
- 1336 99. Galletti, C.S.; Myint, S.W. Land-Use Mapping in a Mixed Urban-Agricultural Arid Landscape
 1337 Using Object-Based Image Analysis: A Case Study from Maricopa, Arizona. *Remote Sensing*1338 2014, 6, 6089–6110.
- 100. Hu, Q.; Wu, W.; Xia, T.; Yu, Q.; Yang, P.; Li, Z.; Song, Q. Exploring the Use of Google
 Earth Imagery and Object-Based Methods in Land Use/Cover Mapping. *Remote Sensing* 2013,
 5, 6026–6042.
- 1342 101. Al-Bakri, J.T.; Ajlouni, M.; Abu-Zanat, M. Incorporating Land Use Mapping and
 1343 Participation in Jordan: An Approach to Sustainable Management of Two Mountainous Areas.
 1344 *Mt. Res. Dev.* 2008, 28, 49–57.
- 1345 102. Liu, J.; Kuang, W.; Zhang, Z.; Xu, X.; Qin, Y.; Ning, J.; Zhou, W.; Zhang, S.; Li, R.; Yan,
 1346 C.; et al. Spatiotemporal characteristics, patterns, and causes of land-use changes in China
 1347 since the late 1980s. J. Geogr. Sci. 2014, 24, 195–210.
- 1348 103. Yadav, P.K.; Kapoor, M.; Sarma, K. Land Use Land Cover Mapping, Change Detection
 1349 and Conflict Analysis of Nagzira-Navegaon Corridor, Central India Using Geospatial
 1350 Technology. *International Journal of Remote Sensing and GIS* 2012, 1.
- 1351 104. d. C. Freitas, C.; d. S. Soler, L.; Sant'Anna, S.J.S.; Dutra, L.V.; dos Santos, J.R.; Mura,
 1352 J.C.; Correia, A.H. Land Use and Land Cover Mapping in the Brazilian Amazon Using
 1353 Polarimetric Airborne P-Band SAR Data. *IEEE Trans. Geosci. Remote Sens.* 2008, 46, 2956–
 1354 2970.
- 1355 105. Dewan, A.M.; Yamaguchi, Y. Land use and land cover change in Greater Dhaka,
 1356 Bangladesh: Using remote sensing to promote sustainable urbanization. *Appl. Geogr.* 2009, 29,
 1357 390–401.
- 1358 106. Castañeda, C.; Ducrot, D. Land cover mapping of wetland areas in an agricultural landscape 1359 using SAR and Landsat imagery. *J. Environ. Manage.* **2009**, *90*, 2270–2277.
- 1360 107. Griffiths, P.; van der Linden, S.; Kuemmerle, T.; Hostert, P. A Pixel-Based Landsat
 1361 Compositing Algorithm for Large Area Land Cover Mapping. *IEEE Journal of Selected Topics*1362 *in Applied Earth Observations and Remote Sensing* 2013, *6*, 2088–2101.
- 1363 108. Ge, Y. Sub-pixel land-cover mapping with improved fraction images upon multiple-point 1364 simulation. *Int. J. Appl. Earth Obs. Geoinf.* **2013**, *22*, 115–126.
- 1365 109. Gong, P.; Wang, J.; Yu, L.; Zhao, Y.; Zhao, Y.; Liang, L.; Niu, Z.; Huang, X.; Fu, H.; Liu,

1366 1367		S.; et al. Finer resolution observation and monitoring of global land cover: first mapping results with Landsat TM and ETM+ data. <i>Int. J. Remote Sens.</i> 2013 , <i>34</i> , 2607–2654.
1368	110.	
1369		of TM and Google earth derived imagery in Shrivan-Darasi watershed (Northwest of Iran).
1370	111.	Deng, J.S.; Wang, K.; Hong, Y.; Qi, J.G. Spatio-temporal dynamics and evolution of land
1371		use change and landscape pattern in response to rapid urbanization. Landsc. Urban Plan. 2009,
1372		92, 187–198.
1373	112.	Otukei, J.R.; Blaschke, T. Land cover change assessment using decision trees, support
1374		vector machines and maximum likelihood classification algorithms. Int. J. Appl. Earth Obs.
1375		Geoinf. 2010, 12, S27–S31.
1376	113.	
1377		Hybrid object-based approach for land use/land cover mapping using high spatial resolution
1378		imagery. Int. J. Geogr. Inf. Sci. 2011, 25, 1025–1043.
1379	114.	· · · ·
1380		incorporating remote sensing and GIS inputs. Appl. Geogr. 2011, 31, 533–544.
1381	115.	
1382		multi-source information based on the Dempster–Shafer theory. Int. J. Geogr. Inf. Sci. 2012,
1383		26, 169–191.
1384	116.	
1385		cover at 250 m using MODIS time series data: A case study in the Dry Chaco ecoregion of
1386		South America. <i>Remote Sens. Environ.</i> 2010 , <i>114</i> , 2816–2832.
1387	117.	
1388		use/cover changes in the eastern Mediterranean. Int. J. Appl. Earth Obs. Geoinf. 2009, 11, 46–
1389		53.
1390	118.	Breiman, L. Random Forests. <i>Mach. Learn.</i> 2001 , <i>45</i> , 5–32.
1391	119.	
1392		for stratified designs and unbalanced prevalence in Random Forest models of tree species
1392		distributions in Nevada. <i>Ecol. Modell.</i> 2012 , 233, 1–10.
1393	120.	
1395		J.O.; Feng, M.; Narasimhan, R.; Kim, D.; et al. Global characterization and monitoring of
1396		forest cover using Landsat data: opportunities and challenges. <i>International Journal of Digital</i>
1397		Earth 2012, 5, 373–397.
1398 1399	121.	
		algorithms for the land-cover classification using limited training data points. <i>ISPRS J.</i>
1400		Photogramm. Remote Sens. 2012, 70, 78–87.
1401	122.	Planet Team Planet Application Program Interface: In Space for Life on Earth. San
1402		Francisco, CA 2017.
1403	123.	
1404		Ben Dor, E.; Helman, D.; Estes, L.; Ciraolo, G.; et al. On the Use of Unmanned Aerial
1405		Systems for Environmental Monitoring. <i>Remote Sensing</i> 2018 , <i>10</i> , 641.
1406	124.	
1407		Proceedings of ISPRS Joint Workshop "High Resolution Mapping from Space" 2001;
1408		pdfs.semanticscholar.org, 2001; pp. 19–21.
1409	125.	1
1410		satellite data exploiting the geometric accuracy of TerraSAR-X data. <i>ISPRS J. Photogramm</i> .
1411		Remote Sens. 2011, 66, 124–132.
1412	126.	
1413		orthorectification process from GeoEye-1 and WorldView-2 panchromatic images. Int. J. Appl.
1414		Earth Obs. Geoinf. 2013, 21, 427–435.
1415	127.	
1416		Proceedings of the Proceedings of the 26th International Conference on World Wide Web
1417		Companion; International World Wide Web Conferences Steering Committee: Republic and
1418		Canton of Geneva, Switzerland, 2017; pp. 771–772.
1419	128.	
1420		Image Segmentation From Online Maps. IEEE Trans. Geosci. Remote Sens. 2017, 55, 6054-
1421		6068.

1422 129. Audebert, N.; Le Saux, B.; Lefèvre, S. Joint learning from earth observation and 1423 openstreetmap data to get faster better semantic maps. In Proceedings of the Proceedings of the 1424 IEEE Conference on Computer Vision and Pattern Recognition Workshops; 2017; pp. 67–75. 1425 Strahler, A.H.; Woodcock, C.E.; Smith, J.A. On the nature of models in remote sensing. 130. 1426 Remote Sens. Environ. 1986, 20, 121–139. 1427 Foody, G.M. Relating the land-cover composition of mixed pixels to artificial neural 131. 1428 network classification output. Photogramm. Eng. Remote Sens. 1996, 62, 491-498. 1429 132. Moody, A.; Gopal, S.; Strahler, A.H. Artificial neural network response to mixed pixels in 1430 coarse-resolution satellite data. Remote Sens. Environ. 1996, 58, 329-343. 1431 De Fries, R.S.; Hansen, M.; Townshend, J.R.G.; Sohlberg, R. Global land cover 133. 1432 classifications at 8 km spatial resolution: The use of training data derived from Landsat 1433 imagery in decision tree classifiers. Int. J. Remote Sens. 1998, 19, 3141-3168. 1434 134. Hansen, M.C.; Potapov, P.V.; Moore, R.; Hancher, M.; Turubanova, S.A.; Tyukavina, A.; 1435 Thau, D.; Stehman, S.V.; Goetz, S.J.; Loveland, T.R.; et al. High-resolution global maps of 1436 21st-century forest cover change. Science 2013, 342, 850-853. 1437 135. Kennedy, R.E.; Yang, Z.; Cohen, W.B. Detecting trends in forest disturbance and recovery 1438 using yearly Landsat time series: 1. LandTrendr — Temporal segmentation algorithms. *Remote* 1439 Sens. Environ. 2010, 114, 2897–2910. 1440 136. Oppenshaw, S.; Taylor, P. A million or so correlation coefficients. Statistical methods in 1441 the spatial sciences. Pion, London 1979. 1442 Jelinski, D.E.; Wu, J. The modifiable areal unit problem and implications for landscape 137. 1443 ecology. Landsc. Ecol. 1996, 11, 129-140. 1444 Weiss, M.; de Beaufort, L.; Baret, F.; Allard, D.; Bruguier, N.; Marloie, O. Mapping leaf 138. 1445 area index measurements at different scales for the validation of large swath satellite sensors: 1446 first results of the VALERI project. In Proceedings of the 8th International symposium in 1447 physical measurements and remote sensing, Aussois (France); w3.avignon.inra.fr, 2001; pp. 1448 125-130. 1449 139. Tian, Y.; Woodcock, C.E.; Wang, Y.; Privette, J.L.; Shabanov, N.V.; Zhou, L.; Zhang, Y.; 1450 Buermann, W.; Dong, J.; Veikkanen, B.; et al. Multiscale analysis and validation of the 1451 MODIS LAI product: I. Uncertainty assessment. Remote Sens. Environ. 2002, 83, 414–430. 1452 Masuoka, E.; Roy, D.; Wolfe, R.; Morisette, J.; Sinno, S.; Teague, M.; Saleous, N.; 140. 1453 Devadiga, S.; Justice, C.O.; Nickeson, J. MODIS Land Data Products: Generation, Quality 1454 Assurance and Validation. In Land Remote Sensing and Global Environmental Change: 1455 NASA's Earth Observing System and the Science of ASTER and MODIS; Ramachandran, B., 1456 Justice, C.O., Abrams, M.J., Eds.; Springer New York: New York, NY, 2011; pp. 509-531 1457 ISBN 9781441967497. 1458 Cohen, W.B.; Justice, C.O. Validating MODIS terrestrial ecology products: linking in situ 141. 1459 and satellite measurements. Remote Sens. Environ. 1999, 70, 1-3. 1460 Fritz, S.; See, L.; McCallum, I.; You, L.; Bun, A.; Moltchanova, E.; Duerauer, M.; 142. 1461 Albrecht, F.; Schill, C.; Perger, C.; et al. Mapping global cropland and field size. Glob. Chang. 1462 Biol. 2015, 21, 1980–1992. 1463 Debats, S.R.; Estes, L.D.; Thompson, D.R.; Caylor, K.K. Integrating active learning and 143. 1464 crowdsourcing into large-scale supervised landcover mapping algorithms; PeerJ Preprints, 1465 2017:. 1466 144. Estes, L.D.; McRitchie, D.; Choi, J.; Debats, S.; Evans, T.; Guthe, W.; Luo, D.; Ragazzo, 1467 G.; Zempleni, R.; Caylor, K.K. A Platform for Crowdsourcing the Creation of Representative, 1468 Accurate Landcover Maps. Environmental Modelling & Software 2016, 80, 41–53. 1469 145. Waldner, F.; Schucknecht, A.; Lesiv, M.; Gallego, J.; See, L.; Pérez-Hoyos, A.; 1470 d'Andrimont, R.; de Maet, T.; Bayas, J.C.L.; Fritz, S.; et al. Conflation of expert and crowd 1471 reference data to validate global binary thematic maps. Remote Sens. Environ. 2019, 221, 235-1472 246. 1473 146. Bey, A.; Sánchez-Paus Díaz, A.; Maniatis, D.; Marchi, G.; Mollicone, D.; Ricci, S.; Bastin, 1474 J.-F.; Moore, R.; Federici, S.; Rezende, M.; et al. Collect Earth: Land Use and Land Cover 1475 Assessment through Augmented Visual Interpretation. Remote Sensing 2016, 8, 807. 1476 147. Fritz, S.; Sturn, T.; Karner, M.; Moorthy, I.; See, L.; Laso Bayas, J.C.; Fraisl, D. FotoQuest 1477 Go: A Citizen Science Approach to the Collection of In-Situ Land Cover and Land Use Data

- 1478 for Calibration and Validation.; pure.iiasa.ac.at, 2019.
- 1479 148. Tuia, D.; Pasolli, E.; Emery, W.J. Using active learning to adapt remote sensing image classifiers. *Remote Sensing of Environment* 2011, *115*, 2232–2242.
- 1481 149. Powell, R.L.; Matzke, N.; de Souza, C.; Clark, M.; Numata, I.; Hess, L.L.; Roberts, D.A.
 1482 Sources of error in accuracy assessment of thematic land-cover maps in the Brazilian Amazon.
 1483 *Remote Sens. Environ.* 2004, *90*, 221–234.
- 1484 150. Van Coillie, F.M.B.; Gardin, S.; Anseel, F.; Duyck, W.; Verbeke, L.P.C.; De Wulf, R.R.
 1485 Variability of operator performance in remote-sensing image interpretation: the importance of human and external factors. *Int. J. Remote Sens.* 2014, *35*, 754–778.
- 1487 151. Johnson, B.A.; Iizuka, K. Integrating OpenStreetMap crowdsourced data and Landsat time1488 series imagery for rapid land use/land cover (LULC) mapping: Case study of the Laguna de
 1489 Bay area of the Philippines. *Appl. Geogr.* 2016, 67, 140–149.
- 1490
 152. Neigh, C.S.R.; Carroll, M.L.; Wooten, M.R.; McCarty, J.L.; Powell, B.F.; Husak, G.J.;
 1491
 1492
 1493
 152. Neigh, C.S.R.; Carroll, M.L.; Wooten, M.R.; McCarty, J.L.; Powell, B.F.; Husak, G.J.;
 1491
 1492
 1493
 1493
 1493
 1493
 1494
 1494
 1494
 1495
 1495
 1495
 1496
 1496
 1497
 1497
 1498
 1498
 1498
 1498
 1499
 1499
 1490
 1490
 1490
 1491
 1491
 1492
 1492
 1493
 1493
 1494
 1494
 1494
 1495
 1495
 1495
 1496
 1496
 1497
 1497
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 1498
 149
- 1494 153. Clark, M.L.; Aide, T.M.; Riner, G. Land change for all municipalities in Latin America and
 the Caribbean assessed from 250-m MODIS imagery (2001–2010). *Remote Sens. Environ.*1496 2012, 126, 84–103.
- 1497 154. Comber, A.; Fisher, P. What is land cover? *Environment and Planning* 2005.
- 1498 155. Kohli, D.; Sliuzas, R.; Kerle, N.; Stein, A. An ontology of slums for image-based
- 1499 classification. *Comput. Environ. Urban Syst.* **2012**, *36*, 154–163.
- 1500 156. Verburg, P.H.; Neumann, K.; Nol, L. Challenges in using land use and land cover data for
 1501 global change studies. *Glob. Chang. Biol.* 2011, *17*, 974–989.
- 1502 157. Weng, Q. Remote sensing of impervious surfaces in the urban areas: Requirements, methods, and trends. *Remote Sens. Environ.* 2012, *117*, 34–49.
- 1504 158. Kohli, D.; Stein, A.; Sliuzas, R. Uncertainty analysis for image interpretations of urban
 1505 slums. *Comput. Environ. Urban Syst.* 2016, 60, 37–49.
- 1506 159. Rocchini, D. While Boolean sets non-gently rip: A theoretical framework on fuzzy sets for
 1507 mapping landscape patterns. *Ecol. Complex.* 2010, 7, 125–129.
- 1508 160. Woodcock, C.E.; Gopal, S. Fuzzy set theory and thematic maps: accuracy assessment and 1509 area estimation. *Int. J. Geogr. Inf. Sci.* **2000**, *14*, 153–172.
- 161. Rocchini, D.; Foody, G.M.; Nagendra, H.; Ricotta, C.; Anand, M.; He, K.S.; Amici, V.;
 1511 Kleinschmit, B.; Förster, M.; Schmidtlein, S.; et al. Uncertainty in ecosystem mapping by
 1512 remote sensing. *Comput. Geosci.* 2013, *50*, 128–135.
- 1513 162. Zhang, J.; Foody, G.M. A fuzzy classification of sub-urban land cover from remotely
 1514 sensed imagery. *Int. J. Remote Sens.* 1998, *19*, 2721–2738.
- 1515 163. Woodcock, C.E.; Strahler, A.H. The factor of scale in remote sensing. *Remote Sens.*1516 *Environ.* 1987, 21, 311–332.
- 1517 164. Cracknell, A.P. Review article Synergy in remote sensing-what's in a pixel? *Int. J. Remote Sens.* 1998, *19*, 2025–2047.
- 1519 165. Pontius, R.G.; Cheuk, M.L. A generalized cross-tabulation matrix to compare soft-1520 classified maps at multiple resolutions. *Int. J. Geogr. Inf. Sci.* **2006**, *20*, 1–30.
- 1521 166. Silván-Cárdenas, J.L.; Wang, L. Sub-pixel confusion–uncertainty matrix for assessing soft
 1522 classifications. *Remote Sens. Environ.* 2008, *112*, 1081–1095.
- 1523 167. Foody, G.M. The continuum of classification fuzziness in thematic mapping. *Photogramm.* 1524 *Eng. Remote Sens.* 1999, 65, 443–452.
- 1525 168. Foody, G.M. Fully fuzzy supervised classification of land cover from remotely sensed 1526 imagery with an artificial neural network. *Neural Comput. Appl.* **1997**, *5*, 238–247.
- 1527 169. Laso Bayas, J.C.; See, L.; Fritz, S.; Sturn, T.; Perger, C.; Dürauer, M.; Karner, M.;
- 1528Moorthy, I.; Schepaschenko, D.; Domian, D.; et al. Crowdsourcing In-Situ Data on Land1529Cover and Land Use Using Gamification and Mobile Technology. *Remote Sensing* 2016, 8,1530905.
- 1531 170. Tewkesbury, A.P.; Comber, A.J.; Tate, N.J.; Lamb, A.; Fisher, P.F. A critical synthesis of
 remotely sensed optical image change detection techniques. *Remote Sens. Environ.* 2015, *160*,
 1533 1–14.

1534	171.	Stehman, S.V.; Fonte, C.C.; Foody, G.M.; See, L. Using volunteered geographic
1535		information (VGI) in design-based statistical inference for area estimation and accuracy
1536		assessment of land cover. Remote Sens. Environ. 2018, 212, 47–59.
1537	172.	
1538	1/2.	inventory mapping: Some implications for boreal forest management. For. Ecol. Manage.
1539		2007 , 252, 208–221.
	172	
1540	173.	
1541	174.	1
1542		Measurement." In Evaluation of measurement data Guide to the expression of uncertainty in
1543		measurement; JCGM, 2008; pp. 1–10.
1544	175.	
1545		Support Vector Machine Classifiers for Land Cover Classification Using Sentinel-2 Imagery.
1546		Sensors 2017, 18.
1547	176.	Song, K. Tackling Uncertainties and Errors in the Satellite Monitoring of Forest Cover
1548		Change. 2010.
1549	177.	e
1550		change estimation. <i>Int. J. Remote Sens.</i> 2009 , <i>30</i> , 3275–3281.
1551	178.	
1552	170.	change as a function of its abundance. <i>Remote Sens. Lett.</i> 2013 , <i>4</i> , 783–792.
1552	170	6
	179.	
1554	100	(NLCD).
1555	180.	
1556		temperature response to changes in urban albedos and associated CO2 offsets. Environ. Res.
1557		<i>Lett.</i> 2010 , <i>5</i> , 014005.
1558	181.	
1559		cover change and expansion of urban lands: A case study in the Seattle metropolitan region.
1560		<i>Landsc. Urban Plan.</i> 2011 , <i>103</i> , 83–93.
1561	182.	Reinmann, A.B.; Hutyra, L.R.; Trlica, A.; Olofsson, P. Assessing the global warming
1562		potential of human settlement expansion in a mesic temperate landscape from 2005 to 2050.
1563		Sci. Total Environ. 2016, 545-546, 512–524.
1564	183.	
1565	100.	Accounting for urban biogenic fluxes in regional carbon budgets. <i>Sci. Total Environ.</i> 2017,
1566		592, 366–372.
1567	184.	
1568	104.	
		direct impacts on biodiversity and carbon pools. <i>Proc. Natl. Acad. Sci. U. S. A.</i> 2012 , <i>109</i> ,
1569	105	16083–16088.
1570	185.	
1571		expansion: Estimates and projections for all countries, 2000–2050. Prog. Plann. 2011, 75, 53–
1572		107.
1573	186.	Coulston, J.W.; Moisen, G.G.; Wilson, B.T.; Finco, M.V.; Cohen, W.B.; Brewer, C.K.
1574		Modeling percent tree canopy cover: a pilot study. Photogrammetric Engineering & Remote
1575		Sensing 78 (7): 715727 2012 , 78, 715–727.
1576	187.	Reinmann, A.B.; Hutyra, L.R. Edge effects enhance carbon uptake and its vulnerability to
1577		climate change in temperate broadleaf forests. Proc. Natl. Acad. Sci. U. S. A. 2017, 114, 107-
1578		112.
1579	188.	
1580	100.	Noise. <i>arXiv</i> [cs.LG] 2017.
1581	189.	
	109.	
1582		Developing World. In Proceedings of the Proceedings of the IEEE Conference on Computer
1583	100	Vision and Pattern Recognition Workshops; openaccess.thecvf.com, 2019; pp. 83–89.
1584	190.	
1585	191.	
1586		D.; Prigent, C.; Gentine, P. Water, Energy, and Carbon with Artificial Neural Networks
1587		(WECANN): A statistically-based estimate of global surface turbulent fluxes and gross
1588		primary productivity using solar-induced fluorescence. <i>Biogeosciences</i> 2017, 14, 4101–4124.
1589	192.	McColl, K.A.; Vogelzang, J.; Konings, A.G.; Entekhabi, D.; Piles, M.; Stoffelen, A.

- 1590 Extended triple collocation: Estimating errors and correlation coefficients with respect to an unknown target. Geophys. Res. Lett. 2014, 41, 6229-6236. 1591 1592 193. Lyndon D. Estes, Su Ye, Lei Song, Ron Eastman, Sitian Xiong, Tammy Woodard, Boka 1593 Luo, Dennis McRitchie, Ryan Avery, Kelly Caylor, Stephanie, Debats. Improving cropland 1594 maps through tight integration of human and machine intelligence. In preparation. 194. 1595 Debats, S.R.; Luo, D.; Estes, L.D.; Fuchs, T.J.; Caylor, K.K. A Generalized Computer 1596 Vision Approach to Mapping Crop Fields in Heterogeneous Agricultural Landscapes. Remote 1597 Sens. Environ. 2016, 179, 210-221. 1598 195. Jain, M.; Balwinder-Singh; Rao, P.; Srivastava, A.K.; Poonia, S.; Blesh, J.; Azzari, G.; 1599 McDonald, A.J.; Lobell, D.B. The impact of agricultural interventions can be doubled by using 1600 satellite data. Nature Sustainability 2019, 2, 931-934. 1601 Pontius, R.G. Criteria to Confirm Models that Simulate Deforestation and Carbon 196. 1602 Disturbance. Land 2018, 7, 105. 1603 197. Schennach, S.M. Recent Advances in the Measurement Error Literature. Annu. Rev. 1604 Econom. 2016, 8, 341-377. 1605 198. Waldner, F.; De Abelleyra, D.; Verón, S.R.; Zhang, M.; Wu, B.; Plotnikov, D.; Bartalev, S.; 1606 Lavreniuk, M.; Skakun, S.; Kussul, N.; et al. Towards a set of agrosystem-specific cropland 1607 mapping methods to address the global cropland diversity. Int. J. Remote Sens. 2016, 37, 1608 3196-3231. Castelluccio, M.; Poggi, G.; Sansone, C.; Verdoliva, L. Land Use Classification in Remote 1609 199. 1610 Sensing Images by Convolutional Neural Networks. arXiv [cs.CV] 2015. 1611 200. Azevedo, T., Sr.; Souza, C.M., Jr.; Shimbo, J.; Alencar, A. MapBiomas initiative: Mapping 1612 annual land cover and land use changes in Brazil from 1985 to 2017.; adsabs.harvard.edu, 1613 2018: Vol. 2018. 1614 201. Brown, J.F.; Tollerud, H.J.; Barber, C.P.; Zhou, Q.; Dwyer, J.L.; Vogelmann, J.E.; 1615 Loveland, T.R.; Woodcock, C.E.; Stehman, S.V.; Zhu, Z.; et al. Lessons learned implementing 1616 an operational continuous United States national land change monitoring capability: The Land 1617 Change Monitoring, Assessment, and Projection (LCMAP) approach. Remote Sens. Environ. 1618 2019, 111356. 1619 202. Estes, L.; Elsen, P.R.; Treuer, T.; Ahmed, L.; Caylor, K.; Chang, J.; Choi, J.J.; Ellis, E.C. 1620 The spatial and temporal domains of modern ecology. Nat Ecol Evol 2018, 2, 819–826. 1621 203. Jensen, J.R.; Cowen, D.C. Remote sensing of urban/suburban infrastructure and socio-1622 economic attributes. Photogramm. Eng. Remote Sens. 1999, 65, 611-622. 1623 204. Dorais, A.; Cardille, J. Strategies for Incorporating High-Resolution Google Earth 1624 Databases to Guide and Validate Classifications: Understanding Deforestation in Borneo. 1625 Remote Sensing 2011, 3, 1157–1176. 1626 205. Sexton, J.O.; Urban, D.L.; Donohue, M.J.; Song, C. Long-term land cover dynamics by 1627 multi-temporal classification across the Landsat-5 record. Remote Sens. Environ. 2013, 128, 1628 246 - 258.1629 206. Reis, M.S.; Escada, M.I.S.; Dutra, L.V.; Sant'Anna, S.J.S.; Vogt, N.D. Towards a 1630 Reproducible LULC Hierarchical Class Legend for Use in the Southwest of Pará State, Brazil: 1631 A Comparison with Remote Sensing Data-Driven Hierarchies. Land 2018, 7, 65. 1632 207. Anderson, J.R. A Land Use and Land Cover Classification System for Use with Remote 1633 Sensor Data; U.S. Government Printing Office, 1976;. 1634 208. Herold, M.; Woodcock, C.E.; Antonio di Gregorio; Mayaux, P.; Belward, A.S.; Latham, J.; 1635 Schmullius, C.C. A joint initiative for harmonization and validation of land cover datasets. 1636 IEEE Trans. Geosci. Remote Sens. 2006, 44, 1719–1727. 1637 209. Carletto, C.; Gourlay, S.; Winters, P. From Guesstimates to GPStimates: Land Area 1638 Measurement and Implications for Agricultural Analysis. J. Afr. Econ. 2015, 24, 593–628. 1639 See, L.; Comber, A.; Salk, C.; Fritz, S.; van der Velde, M.; Perger, C.; Schill, C.; 210. 1640 McCallum, I.; Kraxner, F.; Obersteiner, M. Comparing the quality of crowdsourced data 1641 contributed by expert and non-experts. PLoS One 2013, 8, e69958. 1642 211. Phinn, S.R. A framework for selecting appropriate remotely sensed data dimensions for 1643 environmental monitoring and management. Int. J. Remote Sens. 1998, 19, 3457-3463. 1644 Phinn, S.R.; Menges, C.; Hill, G.J.E.; Stanford, M. Optimizing Remotely Sensed Solutions 212.
- 1645 for Monitoring, Modeling, and Managing Coastal Environments. *Remote Sens. Environ.* 2000,

- 1646 *73*, 117–132.
- 1647 213. Lu, D.; Weng, Q. A survey of image classification methods and techniques for improving classification performance. *International Journal of Remote Sensing* **2007**, *28*, 823–870.
- 1649 214. Cingolani, A.M.; Renison, D.; Zak, M.R.; Cabido, M.R. Mapping vegetation in a
 1650 heterogeneous mountain rangeland using landsat data: an alternative method to define and
 1651 classify land-cover units. *Remote Sens. Environ.* 2004, *92*, 84–97.
- Burke, M.; Lobell, D.B. Satellite-based assessment of yield variation and its determinants in
 smallholder African systems. *Proc. Natl. Acad. Sci. U. S. A.* 2017.
- 1654 216. Jin, Z.; Azzari, G.; You, C.; Di Tommaso, S.; Aston, S.; Burke, M.; Lobell, D.B.
 1655 Smallholder maize area and yield mapping at national scales with Google Earth Engine.
 1656 *Remote Sens. Environ.* 2019, 228, 115–128.
- 1657 217. Lobell, D.B.; Thau, D.; Seifert, C.; Engle, E.; Little, B. A Scalable Satellite-Based Crop
 1658 Yield Mapper. *Remote Sens. Environ.* 2015, *164*, 324–333.
- 1659 218. Grassini, P.; van Bussel, L.G.J.; Van Wart, J.; Wolf, J.; Claessens, L.; Yang, H.; Boogaard,
 1660 H.; de Groot, H.; van Ittersum, M.K.; Cassman, K.G. How Good Is Good Enough? Data
 1661 Requirements for Reliable Crop Yield Simulations and Yield-Gap Analysis. *Field Crops Res.*1662 2015, 177, 49–63.
- 1663 219. Russakovsky, O.; Deng, J.; Su, H.; Krause, J.; Satheesh, S.; Ma, S.; Huang, Z.; Karpathy,
 1664 A.; Khosla, A.; Bernstein, M.; et al. ImageNet Large Scale Visual Recognition Challenge. *Int.*1665 J. Comput. Vis. 2015, 115, 211–252.
- 1666
 1667
 1667
 1667
 1668
 1668
 1669
 169
 162–177 ISBN 9783319168647.
- 1670 221. Basu, S.; Ganguly, S.; Mukhopadhyay, S.; DiBiano, R.; Karki, M.; Nemani, R. DeepSat: A
 1671 Learning Framework for Satellite Imagery. In Proceedings of the Proceedings of the 23rd
 1672 SIGSPATIAL International Conference on Advances in Geographic Information Systems;
 1673 ACM: New York, NY, USA, 2015; pp. 37:1–37:10.
- Yang, Y.; Newsam, S. Bag-of-visual-words and spatial extensions for land-use
 classification. *Proceedings of the 18th SIGSPATIAL international* 2010.
- 1676 223. Shen, C. A Transdisciplinary Review of Deep Learning Research and Its Relevance for
 1677 Water Resources Scientists. *Water Resour. Res.* 2018, 54, 8558–8593.
- 1678 224. Stehman, S.V.; Czaplewski, R.L. Design and Analysis for Thematic Map Accuracy
- Assessment: Fundamental Principles. *Remote Sens. Environ.* **1998**, 64, 331–344.
- 1680 225. Stehman, S.V. Practical Implications of Design-Based Sampling Inference for Thematic
 1681 Map Accuracy Assessment. *Remote Sens. Environ.* 2000, 72, 35–45.
- 1682 226. Aldwaik, S.Z.; Pontius, R.G., Jr. Intensity analysis to unify measurements of size and
 1683 stationarity of land changes by interval, category, and transition. *Landsc. Urban Plan.* 2012,
 1684 106, 103–114.
- 1685 227. Pontius, R.G.; Gao, Y.; Giner, N.M.; Kohyama, T.; Osaki, M.; Hirose, K. Design and
 1686 Interpretation of Intensity Analysis Illustrated by Land Change in Central Kalimantan,
 1687 Indonesia. *Land* 2013, 2, 351–369.
- 1688 228. Foody, G.M. Harshness in image classification accuracy assessment. *Int. J. Remote Sens.*1689 2008, 29, 3137–3158.
- 1690 229. Cohen, J. A Coefficient of Agreement for Nominal Scales. *Educ. Psychol. Meas.* 1960, 20, 37–46.
- 1692 230. Zhang, Q.-M.; Shang, M.-S.; Zeng, W.; Chen, Y.; Lü, L. Empirical comparison of local
 1693 structural similarity indices for collaborative-filtering-based recommender systems. *Phys.*1694 *Procedia* 2010, *3*, 1887–1896.
- 1695 231. Marçal, A.R.S.; Rodrigues, A.S. A method for multi-spectral image segmentation
 1696 evaluation based on synthetic images. *Comput. Geosci.* 2009, 35, 1574–1581.
- 1697 232. Rahman, M.A.; Wang, Y. Optimizing Intersection-Over-Union in Deep Neural Networks
 1698 for Image Segmentation. In Proceedings of the Advances in Visual Computing; Springer
 1699 International Publishing, 2016; pp. 234–244.
- Shi, R.; Ngan, K.N.; Li, S. Jaccard index compensation for object segmentation evaluation.
 In Proceedings of the 2014 IEEE International Conference on Image Processing (ICIP); 2014;

1702 pp. 4457–4461.

© 2020 by the authors.

1703