



Peer review status:

Accepted for publication in REMOTE SENSING OF ENVIRONMENT

1 **Testing the accuracy and transferability of remotely sensed biomass models across**
2 **heterogeneous grasslands**

3

4 Jan Schweizer^{1,2*}, Leon T. Hauser^{1,2}, Hamed Gholizadeh³, Anna K. Schweiger⁴, Christian
5 Rossi^{1,2,3}

6

7 ¹Department of Geoinformation, Swiss National Park, Runatsch 124, 7530 Zernez, CH

8 ²Department of Geography, University of Zurich, Winterthurerstrasse 190, 8057 Zurich, CH

9 ³Department of Geography, Oklahoma State University, 318 Social Sciences & Humanities,
10 74078 Stillwater, OK, USA

11 ⁴Department of Land Resources & Environmental Sciences, Montana State University, 334
12 Leon Johnson Hall, 59717 Bozeman, MT, USA

13 *Author to whom correspondence should be addressed

14

15 **Keywords**

16 Model comparison, Empirical models, Physically-based models, Hybrid retrieval, Active
17 Learning, Sentinel-2, LUT inversion, PROSAIL, Gaussian process regression

18 **Abstract**

19 Grassland aboveground biomass provides key insights into ecological processes such as carbon
20 sequestration, animal movement patterns, and agricultural management practices. Different
21 model types have been developed to estimate grassland biomass from satellite imagery.
22 However, differences in model performance across sites with varying management and ecology
23 remain largely understudied. In this study, we compared accuracy and transferability of
24 empirical, physically-based, and hybrid models to estimate grassland biomass from
25 multispectral Sentinel-2 data in an agnostic scenario, i.e., the models were not provided with
26 any site-specific information beyond the spectral data. Based on field data from five study sites
27 in Europe and the United States, we assessed (1) site-level accuracy of biomass estimation
28 models, (2) model transferability between sites (domain shift), (3) the performance of models
29 trained or optimized with data from multiple study sites (domain generalization), and (4) the
30 relationship between epistemic uncertainty and model transferability. Our results showed that
31 (1) all models exhibited comparable performance at the site level, (2) physically-based models
32 showed the highest degree of transferability between sites, (3) no model consistently
33 outperformed all other models when trained or optimized with field data from multiple sites,
34 and (4) epistemic uncertainty was not necessarily a reliable measure of model applicability to
35 unseen data. Our findings demonstrate the challenges associated with grassland biomass models
36 under domain shift. This elucidates limits to agnostic inference in targeting diverse grasslands
37 and highlights that model transferability is an integral part of performance assessment towards
38 scalable satellite-based grassland monitoring systems, especially as the community increasingly
39 deploys models at continental to global scales.

40 **1. Introduction**

41 Grasslands cover up to 40% of Earth’s terrestrial surface (White et al., 2000) and two-thirds of
42 the Earth’s agricultural land area (O’Mara, 2012). With their extensive land coverage,
43 grasslands store around a third of the global terrestrial carbon (Bai and Cotrufo, 2022) and host
44 many endemic species (Hobohm and Bruchmann, 2009). At the same time, grasslands provide
45 essential ecosystem services (Lemaire et al., 2011; Zhao et al., 2020) including global food
46 production (Bengtsson et al., 2019; O’Mara, 2012). Given their ecological, cultural, and
47 economic importance, accurate monitoring of grasslands is imperative to counteract declines in
48 biodiversity and ecosystem services (Bardgett et al., 2021).

49 Within the overall objective of grassland monitoring, the accurate and reliable estimation of
50 aboveground biomass (hereafter referred to as biomass) is critical for quantifying numerous
51 ecological processes and effects of human disturbances. Biomass is an important parameter for
52 Earth System Models (e.g., Lawrence et al., 2019) and for estimating the contribution of
53 grasslands to the global carbon cycle (Erb et al., 2018). Grassland biomass is also a key driver
54 of animal movement and grazing patterns (Bailey et al., 1996; Rempfler et al., 2024; Schweiger
55 et al., 2015b), and provides information about management practices such as mowing (De
56 Vroey et al., 2022), whose timing and frequency are linked to biodiversity (Socher et al., 2012;
57 Van Vooren et al., 2018) and productivity (Zhang et al., 2023).

58 For large-scale grassland biomass estimation, spaceborne remote sensing enables repeated
59 observations across large spatial domains. Optical sensors measuring surface reflectance, such
60 as those onboard the European Space Agency’s (ESA) Sentinel-2 and the National Aeronautics
61 and Space Administration’s (NASA)/United States Geological Survey’s (USGS) Landsat
62 satellites, are widely used to examine vegetation dynamics of grasslands (Reinermann et al.,
63 2020). Open data policies have supported the availability of many years of archived data
64 (Gascon et al., 2017; Masek et al., 2020; Roy et al., 2014). The spectral layout of Sentinel-2 has

65 shown to hold viable information to estimate foliar properties and canopy structure used to infer
66 biomass (de Sá et al., 2021; Guerini Filho et al., 2020; Hauser et al., 2021a; Rossi et al., 2020).
67 However, many grasslands are subject to high levels of spectral complexity arising from several
68 factors, including effects of non-photosynthetically active vegetation (NPV; Xu et al., 2014),
69 co-occurring plant functional types (Dixon et al., 2014), and management regimes affecting
70 grassland phenology and species composition (Ali et al., 2016; Rossi et al., 2024).

71 This complexity led to the emergence of different model types to estimate biomass, including
72 the use of vegetation indices (VIs), empirical, physical, and hybrid models. Each of these model
73 types has its specific trade-offs regarding required field data, model complexity, specificity,
74 and transferability (i.e., domain shift). Vegetation indices (VIs), such as the Normalized
75 Difference Vegetation Index (NDVI; Guerini Filho et al., 2020; Li et al., 2016; Wang et al.,
76 2019), are found to correlate to various vegetation properties at both the leaf and canopy level.
77 They are straightforward to use, but can saturate with high amounts of biomass (Huete et al.,
78 2002; Zeng et al., 2023). Empirical models are statistical models trained and validated using
79 field data. Deriving relationships from the training data, they do not rely on prior knowledge
80 about the relationship between input and output variables, and are generally well-equipped to
81 handle non-linearity and noise often present in remote sensing data (Verrelst et al., 2015).
82 Numerous empirical machine and deep learning models such as Random Forest regression
83 (RFR), Support Vector regression (SVR), Extreme Gradient boosting (XGB), Gaussian process
84 regression (GPR) and Deep Neural Networks (DNN) have been successfully used to estimate
85 grassland biomass and vegetation traits from optical Sentinel-2 and Sentinel-1 synthetic
86 aperture radar (SAR) data (Li et al., 2021; Muro et al., 2022; Raab et al., 2020; Schwieder et
87 al., 2020; Verrelst et al., 2012). Physically-based Radiative Transfer Models (RTMs) simulate
88 the interactions between light and matter at leaf- and canopy-scales reducing reliance on field
89 data and improving domain generalization by leveraging universal physical principles (He et

90 al., 2019; Wang et al., 2023). RTM inversion can be achieved by using a look-up table (LUT)
91 approach (Verrelst et al., 2014), which connects simulated or measured spectra with trait
92 combinations linked to those spectra. For example, the PROSAIL RTM (Jacquemoud et al.,
93 2009) can be inverted to estimate grassland biomass derived by multiplying leaf dry matter
94 content with leaf area index (LAI; see e.g., He et al. (2019)). However, the PROSAIL RTM is
95 based on heavily idealized assumptions, such as uniformly distributed leaf constituents and
96 geometrically homogeneous canopies (Jacquemoud and Baret, 1990; Verhoef et al., 2007), that
97 are never met in reality. Hybrid models combine RTM-simulated canopy spectra with machine
98 learning models for model inversion (Verrelst et al., 2015). Active Learning (AL) is often used
99 in combination with hybrid models to select the most informative training samples (Verrelst et
100 al., 2016) to perform the RTM inversion which is inherently ill-posed (Combal et al., 2003),
101 coming however at the potential cost of model transferability (Berger et al., 2021b; Tagliabue
102 et al., 2022). Hybrid models have been successfully used for estimating vegetation properties
103 in croplands (Berger et al., 2021a, 2020; Ranghetti et al., 2022; Tagliabue et al., 2022; Verrelst
104 et al., 2021; Woche et al., 2022) and forests (Binh et al., 2022; Brown et al., 2019; Hauser et
105 al., 2021a; Yuan et al., 2015). However, the findings from these previous studies cannot be
106 simply extended to grasslands, since grasslands are more chemically and structurally diverse
107 compared to croplands and differ in plant size and canopy characteristics from forests (Habel
108 et al., 2013; Wellstein et al., 2013). Grasslands also tend to violate the assumption of
109 geometrically homogeneous canopies of one-dimensional RTMs (Berger et al., 2018; Rossi et
110 al., 2020), complicating the selection of the most appropriate model for a given application.

111 In this paper, we aim to compare well-established empirical, physically-based, and hybrid
112 models to estimate biomass across grassland sites with different management regimes, altitude
113 and climate, and determine model performance by assessing their local accuracy,
114 transferability, and epistemic uncertainty (= model uncertainty referring to the confidence of a

115 model about its prediction) sensu Martínez-Ferrer et al. (2022). In doing so, we are focusing on
116 an agnostic scenario, meaning that the models were not provided with any site-specific
117 information beyond the spectral data. The rationale behind this scenario is the endeavor to
118 develop accurate grassland biomass models using only widely available remote sensing data,
119 as site-specific ancillary data often represent an operational bottleneck. Our goals are to develop
120 adequate grassland biomass estimation models for each site and to identify key considerations
121 for domain shift and generalization.

122 **2. Methods**

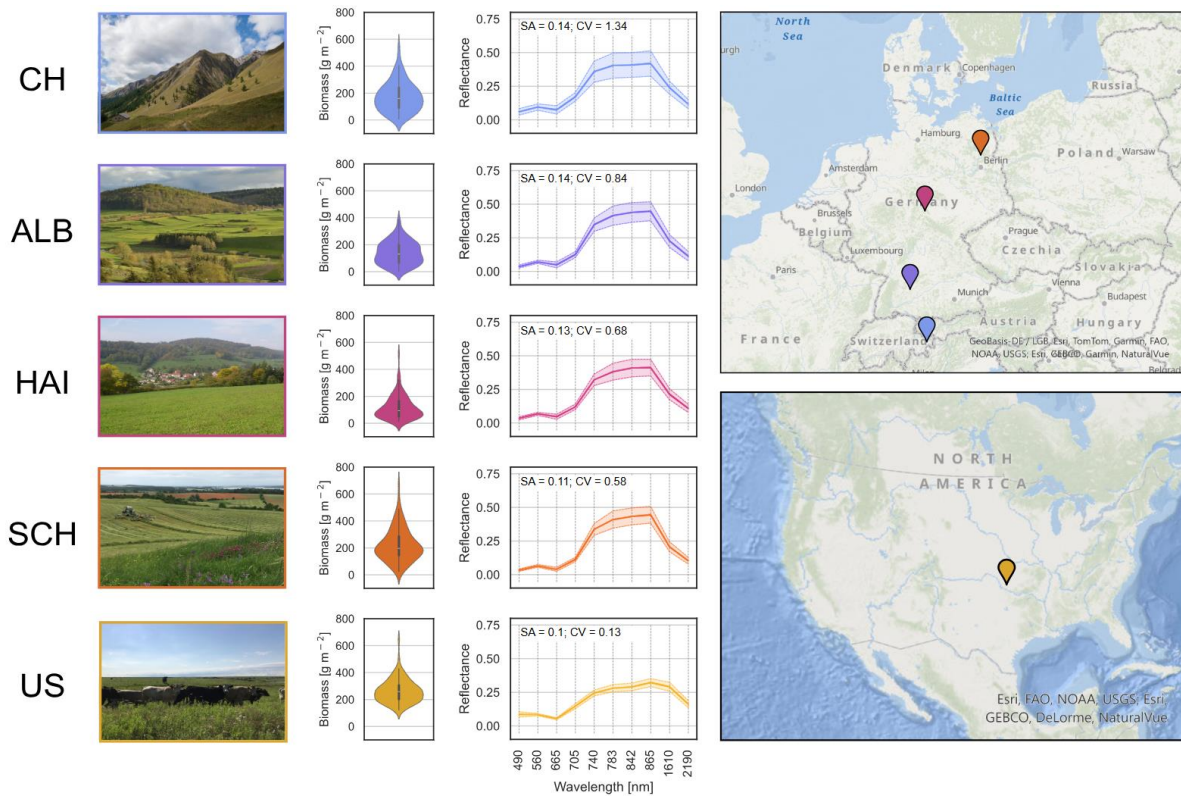
123 2.1. Study sites and data acquisition

124 We used field and remote sensing data from five study sites: one in Switzerland, three in
 125 Germany, and one in the United States differing in environmental characteristics, including
 126 altitude, climate, and management practices (Table 1, Figure 1). These differences make our
 127 compiled dataset particularly valuable for assessing model transferability. The number of
 128 samples per study site ranged from 100 to 429, but we based all our models on 100 samples to
 129 keep them comparable. For sample selection we used Latin hypercube sampling (LHS) as
 130 implemented in the *clhs* package *v0.9.0* (Roudier, 2021) in R *v4.2.1* (R Core Team, 2021).

131 *Table 1: Overview of the characteristics of the five study sites covering a wide range of topographic and climate, and*
 132 *management practices. Remote sensing data were acquired with the Airborne Prism Experiment (APEX) and Sentinel-2*
 133 *between 2010 and 2020. The Köppen-Geiger climate classification is following Beck et al. (2018) for period 1991-2020.*
 134 *Dominant species are provided for Switzerland by the Swiss National Park’s long-term permanent grassland monitoring*
 135 *project, for Germany by Bolliger et al. (2020), and for the United States by Gholizadeh et al. (2022). m.a.s.l.: meters above sea*
 136 *level, MAT: mean annual air temperature, MAP: mean annual precipitation.*

Country	Switzerland		Germany			United States
Site code	CH		ALB	HAI	SCH	US
Site	Swiss National Park	Lower Engadine, Val Müstair	Schwäbische Alb	Hainich-Dün	Schorfheide- Chorin	The Nature Conservancy’s Tallgrass Prairie Preserve (Pawhuska, OK)
Elevation [m.a.s.l.]	1,400 – 2,500		460 – 860	285 – 550	3 – 140	252 - 365
MAT [°C]	1 (at 2,000 m.a.s.l.)		6 – 7	6.5 – 8	8 – 8.5	17.4
MAP [mm/a]	800		700 – 1,000	500 – 800	500 – 600	960
Köppen-Geiger climate class	Cold, no dry season, cold summer (Dfc)		Cold, no dry season, warm summer (Dfb)	Mostly Cfb, small areas Dfb	Temperate, no dry season, warm summer (Cfb)	Temperate, no dry season, hot summer (Cfa)

Dominant species	<i>Erica carnea, Nardus stricta, Carex sempervirens, Festuca rubra</i>		<i>Alopecurus pratensis, Taraxacum sp., Festuca rubra aggr., Bromus erectus</i>	<i>Poa pratensis aggr., Taraxacum sp., Lolium perenne, Alopecurus pratensis</i>	<i>Poa pratensis aggr., Lolium perenne, Poa trivialis, Elymus repens</i>	<i>Schizachyrium scoparium, Andropogon gerardii, Sorghastrum nutans, Panicum virgatum</i>
Management	Strict protection	Grazing, mowing, fertilizing				Grazing, burning
Sensor	Resampled APEX	Sentinel-2				
Years	2010 - 2013	2016 - 2017	2017 - 2020			2022
Number of available samples	407	22	194	185	146	100
Number of selected samples	78	22	100	100	100	100



139

140

Figure 1: Overview of the five study sites. Violin plots show dried aboveground biomass for the 100 selected samples per site.

141

Line plots show the mean spectra (± 1 standard deviation) per site for Sentinel-2 or resampled APEX data. Mean spectral angle

142

in radians (SA; Yuhas et al., 1992) of all unique sample combinations and coefficient of variation (CV) for all spectral bands

143

serve as indicators of site-specific spectral variability. CH: Switzerland, image credits: Swiss National Park/Hans Lozza. ALB:

144

Schwäbische Alb, image credits: Biodiversity Exploratories Information System (BExIS)/Martin Fellendorf. HAI: Hainich-Dün,

145

image credits: BExIS/Steffen Both. SCH: Schorfheide-Chorin, image credits: BExIS/Ulrike Garbe. US: United States, image

146

credits: Nicholas McMillan.

147

2.1.1. Switzerland

148

The Swiss study site (site code CH) encompasses the Lower Engadine and the Val Müstair in

149

the Canton of Grisons in southeast Switzerland. Plots located in the Swiss National Park (SNP),

150

an IUCN (International Union for the Conservation of Nature) category Ia nature reserve

151

(highest protection level – strict nature reserve), are unmanaged. Plots in the Lower Engadine

152

and the Val Müstair adjacent to the SNP are fertilized, mown, and grazed to varying degrees

153

(Rossi et al., 2020).

154 At the CH site, biomass sampling for took place on the day of remote sensing data acquisition
155 in late June to early July of 2010 – 2013 and 2016 – 2017, respectively. Biomass was clipped
156 approximately 1 cm above the ground in 1 m² plots representative for a homogeneous area of 6
157 × 6 m, and dried at 65° for 48 h for 429 samples (Rossi et al., 2020; Schweiger et al., 2017,
158 2015b, 2015a). From 2010 to 2013 (before the launch of Sentinel-2), remote sensing data were
159 acquired with the Airborne Prism Experiment (APEX) imaging spectrometer (Jehle et al., 2010;
160 Schaepman et al., 2015). APEX data were resampled to 2 m pixel size using nearest neighbor
161 interpolation and the parametric geocoding procedure PARGE (Schläpfer and Richter, 2002)
162 and the airborne atmospheric and topographic correction model ATCOR-4 (Richter and
163 Schläpfer, 2002) were used for geometric and atmospheric correction, respectively (Schweiger
164 et al., 2015b). APEX data were resampled to Sentinel-2 bands (Appendix A Section A.1) using
165 the prospectr R package *v0.2.6* (Stevens and Ramirez-Lopez, 2022) in R *v4.2.1*, and can be
166 considered comparable with Sentinel-2 (Helfenstein et al., 2022). In 2016 and 2017, Sentinel-
167 2 Level-1C (top of atmosphere) images were downloaded from the Copernicus Open Access
168 Hub (<https://scihub.copernicus.eu/dhus/>) and processed to Level-2A (surface reflectance) using
169 Sen2Cor *v2.3* (Müller-Wilm et al., 2013) and the SRTM 90 m digital elevation model (Reuter
170 et al., 2007). The 10 m bands (B2, B3, B4, B8) were aggregated to a spatial resolution of 20 m
171 using the arithmetic mean and bidirectional reflectance distribution function (BRDF) correction
172 was applied following Poortinga et al. (2019). No cloud masking was required. For each plot,
173 the spectral reflectance was sampled in the respective Sentinel-2 image by calculating the
174 weighted mean on 20 m resolution around the plot center coordinate.

175 2.1.2. Germany

176 The three German study sites (site codes ALB, HAI, SCH) are part of the *Biodiversity*
177 *Exploratories* (<https://www.biodiversity-exploratories.de/en/>). At each site, 50 grassland plots
178 with different management regimes have been closely monitored since 2009 (Fischer et al.,

179 2010; Hinderling et al., 2023; Ostrowski et al., 2020). Management intensity varies from
180 extensive to moderately intensive, e.g., from no mowing or fertilization to three mowing events
181 per year plus fertilization (Blüthgen et al., 2012; Fischer et al., 2010).

182 Biomass was harvested between late April and mid-July from 2017 to 2020 by clipping biomass
183 approximately 4 cm above the ground on an area of 2 m² in plots representative of a
184 homogeneous area of 50 × 50 m and subsequent drying at 80° for 48 h, resulting in 600 samples
185 (Hinderling et al., 2023). We linearly scaled the dry biomass content to 1 m² for consistency
186 with the other study sites. Sentinel-2 Level-2A data closest to the day of biomass harvest were
187 acquired through the Google Earth Engine (GEE, Gorelick et al., 2017) using the “Harmonized
188 Sentinel-2 MSI: MultiSpectral Instrument, Level-2A” collection. The s2cloudless algorithm
189 was used to mask out clouds and cloud shadows with the cloud probability threshold set to 10%
190 (Zupanc, 2017). Again, the 10 m bands were aggregated to 20 m using the function
191 *reduceResolution* in GEE and BRDF correction was applied. Sampling of the spectral
192 reflectance for each plot follows the protocol for the CH site. An NDVI threshold was applied
193 to prevent the inclusion of plots influenced by artifacts such as remaining cloud shadows or
194 inhomogeneous vegetation cover (Appendix A Section A.2) with 525 samples remaining (Table
195 1).

196 2.1.3. United States

197 The study site in the United States (site code US) is located near Pawhuska, Oklahoma, and
198 falls within The Nature Conservancy’s Tallgrass Prairie Reserve (TGPP), encompassing an area
199 of approximately 160 km² mostly covered by tallgrass prairie with some oak woodland
200 (Hamilton, 2007; The Nature Conservancy, 2023). The TGPP is managed by cattle or bison
201 grazing, and patch burning (Sherrill, 2019). This creates a “*shifting mosaic*” (Fuhlendorf and
202 Engle, 2004) of patches with varying grazing pressure, as bison and cattle tend to primarily

203 graze in recently burned areas with high nutrient availability (Anderson et al., 2006; Fuhlendorf
204 and Engle, 2001).

205 Biomass sampling was conducted between July and August 2022 across 100 plots of 30 × 30
206 m, each containing nine 1 m² quadrats (Gholizadeh et al., 2024). Biomass was clipped in 80 of
207 the 900 1 m² quadrats at approximately 2.54 cm (1 inch) above ground and dried at 65° for 144
208 h. In the remaining quadrats, biomass was determined using the digital obstruction method
209 using the 80 samples for calibration (Limb et al., 2007). For each of the 100 plots, biomass was
210 calculated as the mean biomass across the nine quadrats. Sentinel-2 Level-2A data for each plot
211 were acquired following the same protocol as for the German study sites.

212 2.2. Model types

213 We assessed the accuracy and transferability of three different model types (i.e., empirical,
214 physically-based, and hybrid; Figure 2). Field data were partitioned into training (for the
215 empirical models) or optimization (for the physically-based and hybrid models) and external
216 testing sets using an 80:20% split using LHS (Figure 2A). We used scikit-learn *v1.5.2*
217 (Pedregosa et al., 2012) and xgboost *v2.1.2* (Chen and Guestrin, 2016) in Python 3.13 for model
218 training, optimization, and validation.

219 2.2.1. Empirical models

220 A wide variety of algorithms can be used to train empirical models (Figure 2B). We used
221 Random Forest regression (RFR), Support Vector regression (SVR), Extreme Gradient
222 boosting regression (XGB), and Gaussian process regression (GPR) models to represent both
223 tree-based and kernel-based methods. These algorithms all have been successfully used to
224 estimate grassland biomass (Li et al., 2021; Muro et al., 2022; Raab et al., 2020; Schwieder et
225 al., 2020; Verrelst et al., 2015, 2012). RFR uses an ensemble of decision trees, where each tree
226 is trained on a random subset of samples and features and the final prediction is obtained by

227 averaging the predictions of all trees (Breiman, 2001). SVR was originally introduced by
 228 Vapnik et al. (1997) and uses support vectors to fit hyperplanes in the data within a specified
 229 margin of tolerance. XGB is based on the concept of gradient boosting of regression trees
 230 introduced by Friedman (2001) and incorporates regularization to mitigate overfitting (Chen
 231 and Guestrin, 2016). Lastly, GPR uses a prior belief about the latent function describing the
 232 relationship between input and output variables, and training data to form the posterior
 233 distribution (Rasmussen and Williams, 2006). The advantages of using GPR include automatic
 234 hyperparameter optimization during model training and the provision of epistemic uncertainty
 235 (Verrelst et al., 2013a). To predict a data point, the model returns the mean of the posterior
 236 distribution as the estimated value and the predictive standard deviation (SD) as a measure of
 237 epistemic uncertainty.

238 A 5-fold cross-validation (CV) scheme with negative RMSE for scoring was used to identify
 239 the optimal model parametrization from all possible combinations of parameters listed in Table
 240 2.

241 *Table 2: Parameter values used for cross-validation of empirical Random Forest regression (RF), Support Vector regression*
 242 *(SVR), Extreme Gradient Boosting regression (XGB), and Gaussian process regression (GPR) models. Nomenclature of*
 243 *parameter names for RFR, SVR, and GPR according to Pedregosa et al. (2012), for XGB according to Chen and Guestrin (2016).*

Model	Parameter function	Parameter name	Values
RF	Number of trees	n_estimators	100, 200, 500
	Maximum tree depth	max_depth	None, 5, 10, 15
	Minimum number of samples required to be at a leaf node	min_samples_leaf	1, 2, 5
	Number of features	max_features	'sqrt', 'log2', 10
	Maximum number of leaf nodes	max_leaf_nodes	10, 20, None
SVR	Kernel type	kernel	'rbf', 'linear'
	Kernel coefficient	gamma	'scale', 'auto', 0.01, 0.1, 1
	Regularization parameter	C	0.1, 1, 10, 100
	Epsilon-tube	epsilon	0.01, 0.1, 0.5, 1
XGB	Number of gradient boosted trees	n_estimators	100, 200, 300

	Maximum tree depth	max_depth	3, 5, 7, 9
	Boosting learning rate	learning_rate	0.01, 0.1, 0.2
	Subsample ratio of training instance	subsample	0.8, 1
	Subsample ratio of columns when constructing each tree	subsample_bytree	0.8, 1
	Minimum loss reduction	gamma	0, 0.1, 0.2
GPR	Kernel	kernel	ConstantKernel() * RBF()
	Length scale bounds of RBF kernel	length_scale_bounds	(-100, 100)

244

245 2.2.2. Physically-based model

246 Here we used the PROSAIL RTM (Jacquemoud et al., 2009), which combines the PROSPECT-
247 D (Féret et al., 2017) and 4SAIL (Verhoef et al., 2007) RTMs, to simulate grassland canopy
248 reflectance (Figure 2C). PROSPECT-D simulates leaf level reflectance by considering leaf
249 properties such as chlorophyll content (CHL), leaf mass per area (LMA) and the angle of
250 incoming solar radiation (Féret et al., 2008). Subsequently, 4SAIL computes the bidirectional
251 reflectance at the canopy level, employing canopy properties such as LAI and sun-target-sensor
252 geometry. We used the prosail R package *v1.1.1* (Féret and de Boissieu, 2022) in R *v4.2.1* to
253 create a single LUT for all study sites containing 10,000 simulated canopy reflectance spectra
254 with input parameters selected by means of LHS within their respective value ranges (Table 3)
255 derived from satellite image metadata, prior knowledge, and literature (He et al., 2019; Rossi
256 et al., 2020; Verrelst et al., 2021). The value ranges and size of the LUT were consistent with
257 other studies (Darvishzadeh et al., 2011; Hauser et al., 2021a; Locherer et al., 2015; Punalekar
258 et al., 2018; Rivera et al., 2013; Rossi et al., 2020). We used the psoil parameter as weighting
259 factor for the dry and wet soil spectra with psoil = 0 corresponding to completely wet soil
260 conditions and psoil = 1 corresponding to completely dry soil conditions, respectively. We used
261 the default dry and wet soil spectra of the prosail R package and did not use site-specific soil
262 spectra, as no corresponding reference data were available and their use would have

263 contradicted the agnostic modeling scenario. The simulated reflectance spectra were resampled
 264 to Sentinel-2 bands using the prospectr R package and their corresponding biomass content was
 265 calculated following Quan et al. (2017, Equation 1).

$$266 \quad \text{biomass [g m}^{-2}\text{]} = \text{LMA [g cm}^{-2}\text{]} * \text{LAI} * 10,000 \quad (1)$$

267 *Table 3: Value ranges and distributions of PROSAIL input parameters used in this study.*

Parameter	Variable	Unit	Minimum value	Maximum value	Distribution
Leaf structure parameter	N	[-]	1.5	1.9	uniform
Chlorophyll content	CHL	[$\mu\text{g cm}^{-2}$]	5	75	uniform
Carotenoid content	CAR	[$\mu\text{g cm}^{-2}$]	2	60	uniform
Anthocyanin content	ANT	[$\mu\text{g cm}^{-2}$]	0	2	uniform
Brown pigment content	BROWN	[-]	0	1	uniform
Equivalent water thickness	EWT	[cm]	0.001	0.04	uniform
Leaf mass per area	LMA	[g cm^{-2}]	0.002	0.015	uniform
Angle for incident light at leaf surface	alpha	[$^{\circ}$]	40	40	fixed
Leaf inclination distribution function	TypeLidf	[-]	2	2	fixed
Average leaf angle	LIDFa	[$^{\circ}$]	40	70	uniform
Leaf area index	LAI	[-]	0.1	4	uniform
Hot spot parameter	q	[-]	0.01	0.1	uniform
Sun zenith angle	tts	[$^{\circ}$]	25	75	uniform
Observer zenith angle	tto	[$^{\circ}$]	0	0	fixed
Relative azimuth angle	psi	[$^{\circ}$]	50	180	uniform
Dry/wet soil factor	psoil	[-]	0	1	uniform

268
 269 Two parameters must be determined for the inversion of the generated LUT: the cost function
 270 and the number of spectra with the lowest cost to consider (hereafter referred to as percentage
 271 of solutions). We tested 17 commonly used cost functions listed by Rivera et al. (2013) and
 272 different percentages of solutions, namely 0.01% (= 1 solution), 1%, 2%, 5%, and 10%
 273 (Punalekar et al., 2018; Rivera et al., 2013; Rossi et al., 2020) to optimize parameter choice.
 274 The predicted biomass value was calculated as the mean value of the selected solutions, with

275 the corresponding SD being indicative of epistemic uncertainty (Locherer et al., 2015; Rivera
276 et al., 2013).

277 2.2.3. Hybrid model

278 For the hybrid model, we created a LUT containing 1,000 simulations (Tagliabue et al., 2022;
279 Verrelst et al., 2021) as described above. But in this case, the simulated canopy reflectance
280 spectra were used to train a regression algorithm for RTM inversion (Figure 2D). Here we used
281 a GPR model (see Section 2.2.1 for further details) with the parameters indicated in Table 2 for
282 the regression task. We used AL to avoid biophysically unrealistic variable combinations and
283 redundant information, and select the most informative simulations from the LUT (Verrelst et
284 al., 2016). Using the AL-selected subset of simulations for GPR training can help to mitigate
285 the ill-posedness inherent to RTM inversion as different variable combinations can lead to
286 similar spectra (Combal et al., 2003), and increase computational efficiency (Berger et al.,
287 2021b). AL selects a predefined percentage of simulations from the LUT as an initial training
288 dataset. Subsequently, the GPR model is trained using this initial training set, and its predictive
289 training accuracy is assessed via the root-mean-square error (RMSE) computed with the
290 optimization set. By employing a selection heuristic, such as Euclidean distance-based diversity
291 (EBD), a simulation from the remaining LUT is selected, temporarily added to the training set,
292 and only permanently kept if the updated training set leads to an improved RMSE. This
293 optimization process continues until all simulations are evaluated. Finally, the validation
294 accuracy of the GPR model trained with the optimal training set is determined.

295 In line with previous studies, we used 2% of the data as initial training data (Tagliabue et al.,
296 2022; Wocher et al., 2022). Hybrid models without the use of AL did not lead to meaningful
297 results in the context of this study (Appendix A Section A.3) and were therefore not further
298 analyzed.

299 2.3. Model comparisons

300 We conducted three model comparisons to investigate model accuracy and assess model
301 transferability (Figure 3). First, models were trained (empirical) or optimized (physically-based
302 and hybrid) and validated individually for each study site (hereafter referred to as local models,
303 Figure 3A). This setting corresponds to that of most local to regional scale studies in which
304 field data of a specific area of interest are available. Second, the local models were applied to
305 the field data of the other study sites to assess their transferability (hereafter referred to as
306 transferred models, Figure 3B), simulating the case where a model trained or optimized for one
307 area is applied to another area where no field data are available (domain shift). For the
308 physically-based models, this means that the combination of cost function and percentage of
309 solutions identified during local model optimization is transferred to the other study sites (i.e.,
310 the LUT remains unchanged). Third, models were trained or optimized with field data from
311 four study sites and validated using the field data of the remaining site to examine any
312 improvement in transferability (hereafter referred to as global models, Figure 3C). This mimics
313 the case in which a diverse set of field data are available, for example from a compiled database,
314 and used to make predictions for an area not covered by the database. Model transferability in
315 such a setting is also referred to as the model's ability for domain generalization (Zhou et al.,
316 2022). In addition, we compared epistemic uncertainties with model accuracies for all model
317 comparisons where available – namely empirical GPR, physically-based, and hybrid models.

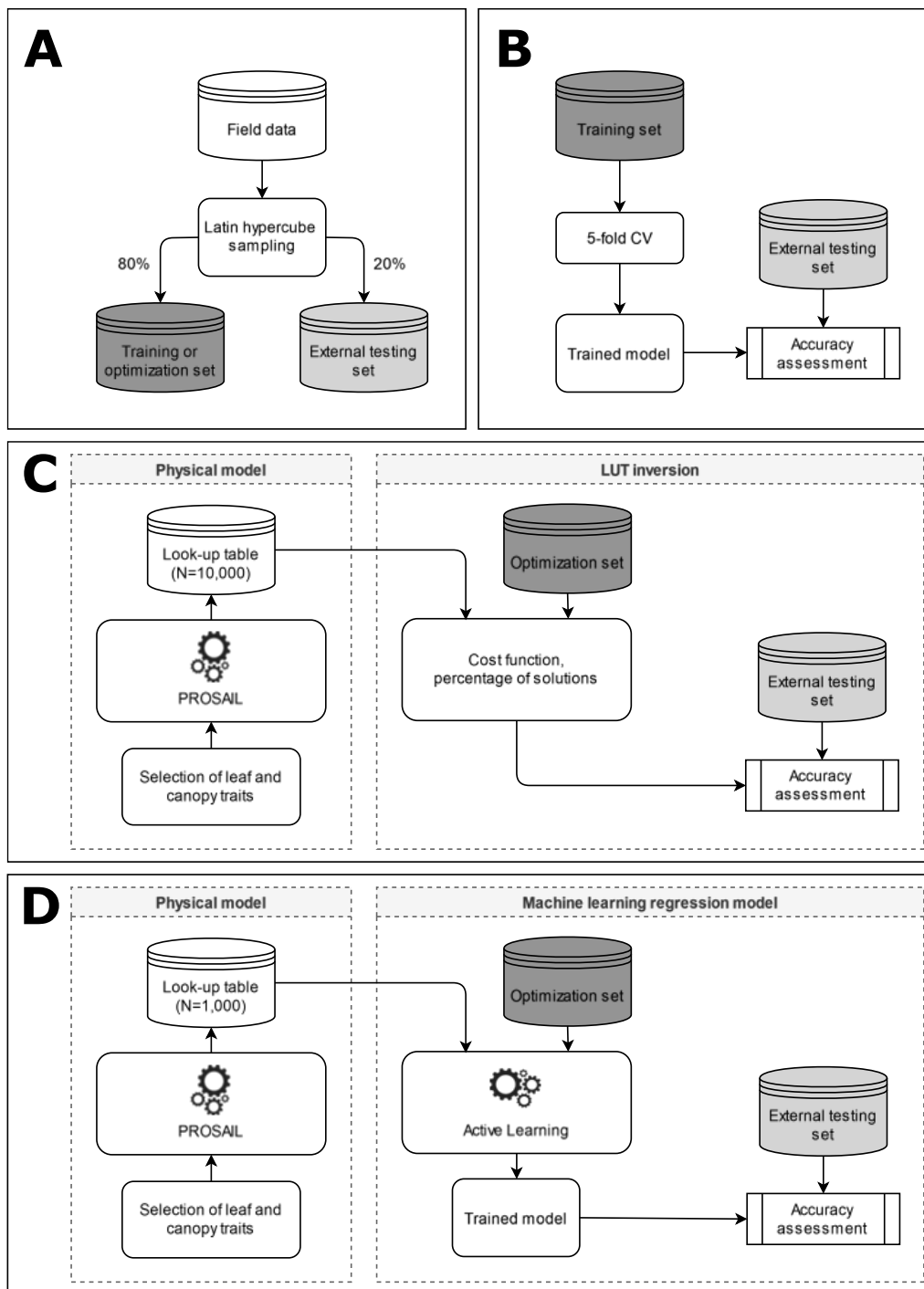
318 Model performance was assessed by coefficient of determination (R^2), relative root-mean-
319 square errors (RRMSE) calculated as the RMSE divided by the mean value of the external
320 testing set (Richter et al., 2012), and mean bias error (MBE) using the external testing set. In
321 all cases, we used the ten Sentinel-2 bands in the visible and near-infrared (B2 – B8A) and the
322 short-wave infrared regions (B11 – B12) as predictor variables with reflectance values being
323 standardized for empirical and hybrid models and normalized, i.e., treating the spectra as
324 probability distributions summing up to 1, for physically-based models (Rivera et al., 2013),

325 respectively. Accordingly, the models should be considered agnostic, meaning that apart from
326 Sentinel-2 spectral data, no site-specific data were used for model training or prediction.

327 We note that due to comparatively small sample size (N=100), partitioning of the data into
328 training or calibration and external testing set was repeated 10 times for the local and transferred
329 models to account for stochastic effects (Muro et al., 2022). Correspondingly, local and
330 transferred model performance of all model types were assessed by calculating both mean and
331 SD for R^2 , RRMSE, and MBE.

332 For the empirical models, the best-performing method was selected to be presented in the results
333 as the focus lied on the comparison of model types and not different empirical models.
334 Comprehensive cross-validation and testing performances for all models including additional
335 performance metrics are enclosed in Appendix A Sections A.4 to A.16.

336 Lastly, the epistemic uncertainty of the local, transferred, and global models was compared to
337 the absolute difference between measured and predicted biomass of the external testing set to
338 test if epistemic uncertainty could potentially be used as an indicator of model transferability
339 for heterogeneous grasslands. Further information about said relation for all models, including
340 the individual repetitions of local and transferred models, is included in Appendix A Section
341 A.17. We tested the relationship between absolute difference and epistemic uncertainty since it
342 has been demonstrated that epistemic uncertainty of GPR models can serve as a *quality*
343 *indicator* to identify reliable and unreliable predictions of transferred models for croplands,
344 even when applied across spatial scales (Verrelst et al., 2013b, 2013a, 2012).

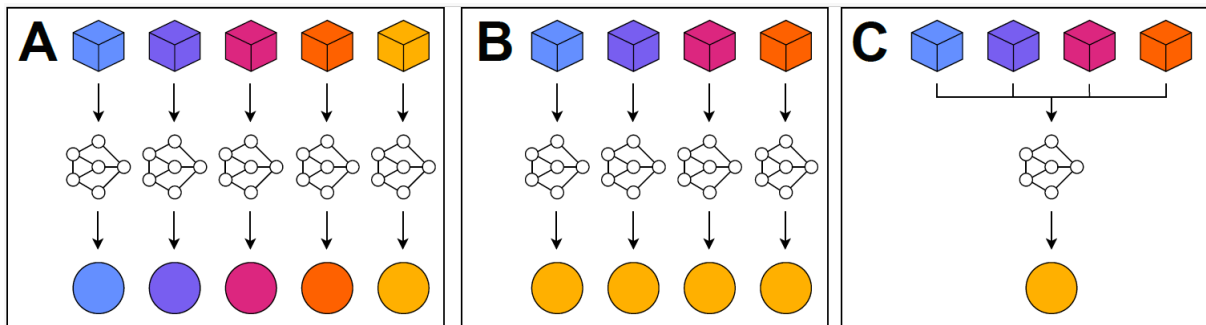


345

346 *Figure 2: Functioning of the models used in this study. A: As a prerequisite for all models, the field data need to be split into*
 347 *training (in the case of empirical models) or optimization set (in case of physically-based and hybrid models; colored in dark*
 348 *grey) and external testing set (colored in light grey) using Latin hypercube sampling. B: Empirical models are data-driven and*
 349 *learn data-specific relationships between predictor variables. The 5-fold cross-validation (CV) was performed with each*
 350 *possible combination of parameters listed in Table 2. C: Physically-based models use a radiative transfer model (RTM) such as*
 351 *PROSAIL to simulate canopy reflectance spectra and a cost function is used to find a predefined number of best matches*
 352 *between each field data point and the simulated spectra, a process commonly referred to as look-up table (LUT) inversion. D:*

353 Hybrid models are trained with a set of RTM-simulated spectra, which are optionally optimized using Active Learning.
 354 Subsequently, a machine learning regression model is used to perform the LUT inversion.

355



356 Figure 3: The three model comparisons conducted in this study. Cubes represent data used for training (of empirical models)
 357 or optimization (of physically-based and hybrid models). Circles represent data used for model validation. Colors represent
 358 data from different study sites. A: Local models with training/optimization and validation data from the same study site. B:
 359 Domain shift: transferred models with training/optimization and validation data from different study site (only one of five
 360 cases shown). C: Domain generalization: global models with training/optimization data from four sites and validation data
 361 from the remaining site (only one of five cases shown).

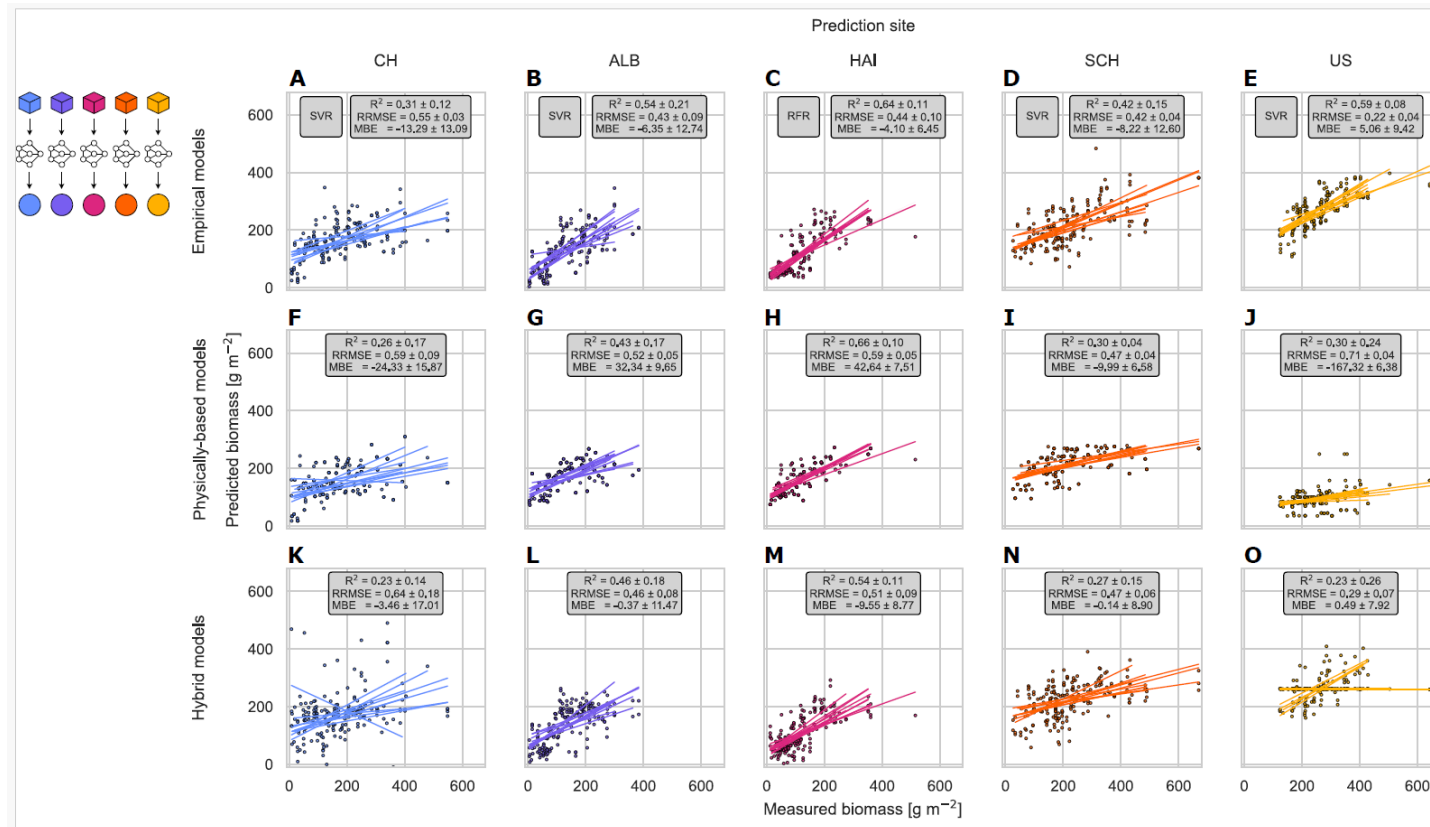
362 **3. Results**

363 3.1. Local models

364 The accuracy of all models varied with the study sites (Figure 4). For empirical models (Figure
365 4A-E), mean R^2 ranged from 0.31 to 0.64 (CH and HAI, respectively) and mean RRMSE from
366 0.22 to 0.55 (US and CH, respectively). Mean MBE indicated a systematic underestimation for
367 all sites between -4.1 and -13.29 g/m^2 (HAI and CH, respectively) except for the US, for which
368 a mean overestimation of 5.06 g/m^2 was observed. SVR outperformed the other empirical
369 models for most sites, only the HAI site was best predicted by RFR.

370 For physically-based models (Figure 4F-J), mean R^2 ranged from 0.26 to 0.66 (CH and HAI,
371 respectively) and mean RRMSE from 0.47 to 0.71 (SCH and US, respectively). For the ALB
372 and HAI sites, a mean MBE of up to 42.64 g/m^2 was observed while on average, biomass was
373 underestimated for CH and SCH. For the US model, a severe underestimation of -167.32 g/m^2
374 was reported. Compared to empirical models, model accuracy slightly decreased for most sites;
375 for the US site it decreased substantially.

376 For hybrid models (Figure 4K-O), mean R^2 ranged from 0.23 to 0.54 (US and HAI,
377 respectively) while mean RRMSE ranged from 0.29 to 0.64 (US and CH, respectively). Except
378 for the US site, a slight underestimation in terms of MBE could be observed. Model accuracy
379 across sites resembled those of empirical models, although they were again slightly lower.



381

382 Figure 4: Scatterplots of measured versus predicted biomass for the local empirical (A-E), physically-based (F-J), and hybrid models (L-O). Textboxes show mean coefficient of determination (R^2),
 383 relative root-mean-square error (RRMSE) and mean bias error (MBE) ± 1 standard deviation across 10 repetitions for each model type. For empirical models, only the best-performing model in terms
 384 of lowest RRMSE is shown with the corresponding model name added in a separate textbox. CH: Switzerland (A, F, K), ALB: Schwäbische Alb (B, G, L), HAI: Hainich-Dün (C, H, M), SCH: Schorfheide-
 385 Chorin (D, I, N), US: United States (E, J, O), RFR: Random Forest regression, SVR: Support Vector regression.

386

387 3.2. Transferred models

388 In general, model accuracy decreased when local models were applied to other sites, although
389 physical models sustained their predictive power the best (Table 4).

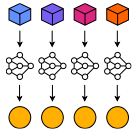
390 Out of all transferred empirical models, the accuracy of the CH models transferred to the SCH
391 site (mean $R^2 = 0.33$, mean RRMSE = 0.49) and models transferred among the three German
392 sites came closest to that of the local models, e.g., the ALB models predicting the HAI site
393 (mean $R^2 = 0.49$, mean RRMSE = 0.54) or the HAI and SCH models predicting the ALB site
394 (mean $R^2 = 0.30$, mean RRMSE = 0.53 and mean $R^2 = 0.49$, mean RRMSE = 0.64,
395 respectively). For the CH site, only a mean R^2 of 0.15 with an associated mean RRMSE of 0.65
396 could be achieved by the transferred empirical model trained at the SCH site. For the US site,
397 all transferred empirical models exhibited a systematic underestimation of the present biomass
398 while the US models themselves overestimated biomass at other sites, e.g., with a mean MBE
399 of 146.09 g/m² when predicting the HAI site.

400 Regarding physically-based models, transferability diverged less strongly between
401 combinations of optimization and prediction sites, with the performance being best for the ALB,
402 HAI, and SCH sites. Overall, variability between the 10 repetitions was comparatively low, as
403 identical combinations of cost function and percentage of solutions were selected for the
404 German sites (Appendix A Section A.5). Best results were achieved for the ALB site with the
405 HAI models having performed similarly to the local models (mean $R^2 = 0.43$, mean RRMSE =
406 0.51). A slight decline in performance could be observed for the HAI site, although the ALB
407 model still achieved a mean R^2 of 0.53 and a mean RRMSE of 0.66, with the decline for the
408 SCH site being more pronounced. For the CH site, the models showed a lower mean R^2 , but
409 only a slightly higher mean RRMSE. For the US site, in contrast, a direct comparison with the
410 local models was difficult; generally, a higher mean R^2 was achieved, but mean RRMSE and

411 systematic underestimation also increased. Moreover, the US models consistently performed
412 worse for all other prediction sites. In comparison with transferred empirical models, the
413 transferability of physically-based models was higher for the CH, ALB, and HAI sites.

414 The transferability of hybrid models also varied among different combinations of optimization
415 and validation sites with best results for combinations of the German study sites such as the
416 ALB models for the HAI site (mean $R^2 = 0.32$, mean RRMSE = 0.63) and vice versa (mean R^2
417 = 0.34, mean RRMSE = 0.52). The ALB models for the SCH site showed comparatively good
418 values for mean R^2 and RRMSE, but with increased systematic underestimation. For the CH
419 and US sites, no satisfactory performance could be achieved. Overall, the transferred hybrid
420 models exhibited similar patterns to the empirical models, although their performance was
421 somewhat lower. Particularly notable was the comparatively good performance among the
422 German study sites and the systematic underestimation of biomass for the US site.

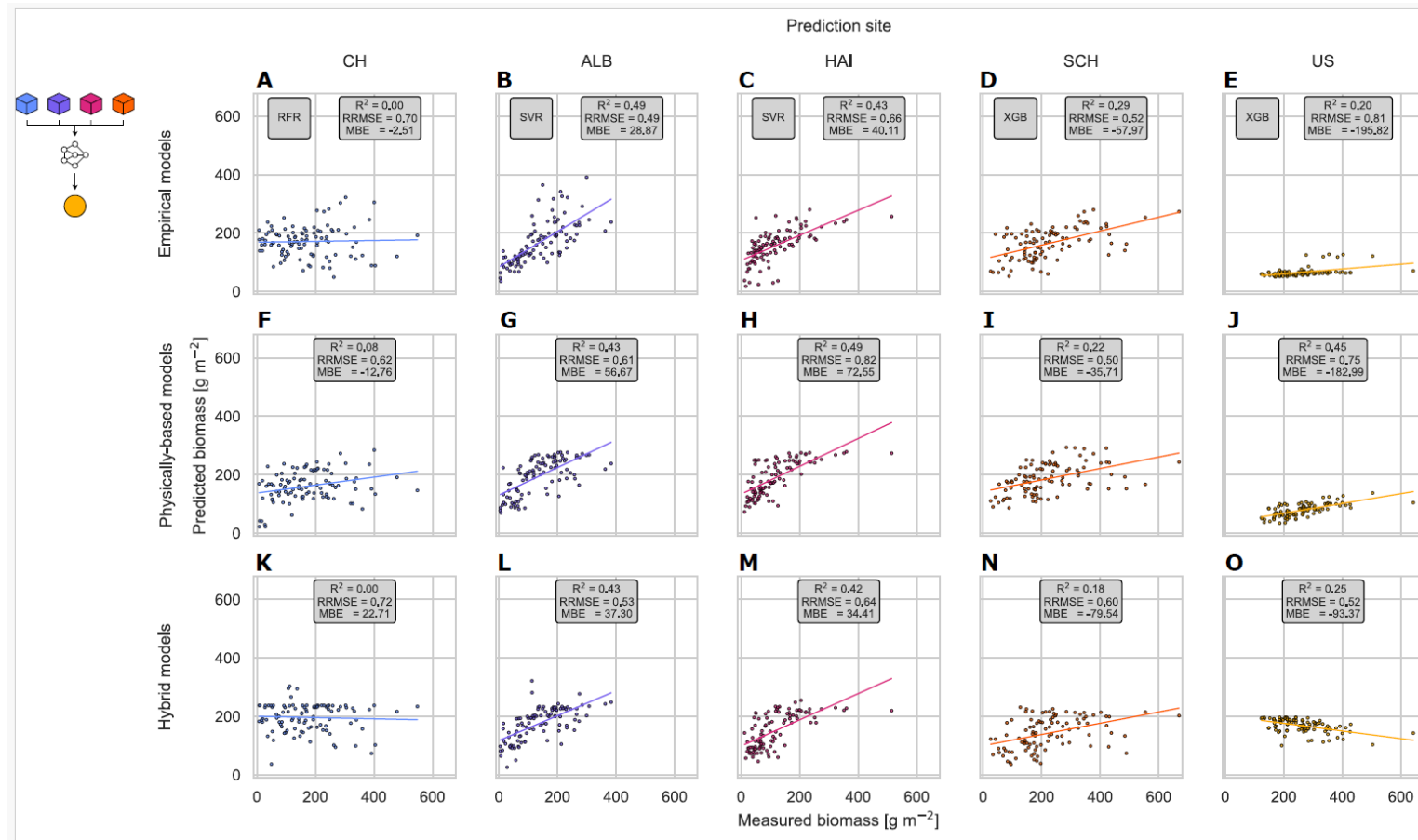
423 Table 4: Mean coefficient of determination (R^2), relative root-mean-square error (RRMSE), and mean bias error (MBE) for ± 1 standard deviation across 10 repetitions for the transferred empirical,
 424 physically-based, and hybrid models. To facilitate an estimation of model transferability, the metrics of the local models were included (in italic). For empirical models, only the best-performing model
 425 in terms of lowest RRMSE is shown with the corresponding model name added in brackets. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States, RFR:
 426 Random Forest regression, SVR: Support Vector regression, XGB: Extreme Gradient Boosting regression, GPR: Gaussian Process regression.

		Prediction site														
		CH			ALB			HAI			SCH			US		
Model type	Training/optimization site	R^2	RRMSE	MBE	R^2	RRMSE	MBE	R^2	RRMSE	MBE	R^2	RRMSE	MBE	R^2	RRMSE	MBE
Empirical	CH	0.31 ± 0.12 <i>(SVR)</i>	0.55 ± 0.03 <i>(SVR)</i>	-13.29 ± 13.09 <i>(SVR)</i>	0.46 ± 0.12 <i>(SVR)</i>	0.70 ± 0.12 <i>(SVR)</i>	67.46 ± 23.94 <i>(SVR)</i>	0.39 ± 0.04 <i>(SVR)</i>	0.92 ± 0.15 <i>(SVR)</i>	77.08 ± 20.47 <i>(SVR)</i>	0.33 ± 0.08 <i>(SVR)</i>	0.49 ± 0.05 <i>(SVR)</i>	-14.42 ± 23.94 <i>(SVR)</i>	0.08 ± 0.05 <i>(GPR)</i>	0.48 ± 0.00 <i>(GPR)</i>	-88.80 ± 0.88 <i>(GPR)</i>
	ALB	0.04 ± 0.02 <i>(GPR)</i>	0.69 ± 0.00 <i>(GPR)</i>	-37.63 ± 0.98 <i>(GPR)</i>	0.54 ± 0.21 <i>(SVR)</i>	0.43 ± 0.09 <i>(SVR)</i>	-6.35 ± 12.74 <i>(SVR)</i>	0.49 ± 0.05 <i>(SVR)</i>	0.54 ± 0.02 <i>(SVR)</i>	-10.15 ± 5.00 <i>(SVR)</i>	0.30 ± 0.03 <i>(XGB)</i>	0.55 ± 0.01 <i>(XGB)</i>	-68.78 ± 3.22 <i>(XGB)</i>	0.04 ± 0.02 <i>(GPR)</i>	0.57 ± 0.00 <i>(GPR)</i>	-123.02 ± 0.86 <i>(GPR)</i>
	HAI	0.00 ± 0.01 <i>(GPR)</i>	0.73 ± 0.00 <i>(GPR)</i>	-62.02 ± 0.58 <i>(GPR)</i>	0.30 ± 0.03 <i>(GPR)</i>	0.53 ± 0.02 <i>(GPR)</i>	-13.74 ± 3.73 <i>(GPR)</i>	0.64 ± 0.11 <i>(RFR)</i>	0.44 ± 0.10 <i>(RFR)</i>	-4.10 ± 6.45 <i>(RFR)</i>	0.16 ± 0.03 <i>(XGB)</i>	0.63 ± 0.02 <i>(XGB)</i>	-83.11 ± 5.78 <i>(XGB)</i>	0.16 ± 0.04 <i>(GPR)</i>	0.65 ± 0.00 <i>(GPR)</i>	-146.93 ± 0.65 <i>(GPR)</i>
	SCH	0.15 ± 0.02 <i>(RFR)</i>	0.65 ± 0.03 <i>(RFR)</i>	22.32 ± 9.96 <i>(RFR)</i>	0.49 ± 0.04 <i>(SVR)</i>	0.64 ± 0.04 <i>(SVR)</i>	64.65 ± 7.68 <i>(SVR)</i>	0.48 ± 0.05 <i>(SVR)</i>	0.76 ± 0.05 <i>(SVR)</i>	62.53 ± 6.07 <i>(SVR)</i>	0.42 ± 0.15 <i>(SVR)</i>	0.42 ± 0.04 <i>(SVR)</i>	-8.22 ± 12.60 <i>(SVR)</i>	0.09 ± 0.03 <i>(GPR)</i>	0.36 ± 0.00 <i>(GPR)</i>	-39.81 ± 1.23 <i>(GPR)</i>
	US	0.01 ± 0.00 <i>(GPR)</i>	0.81 ± 0.00 <i>(GPR)</i>	87.98 ± 1.09 <i>(GPR)</i>	0.03 ± 0.01 <i>(GPR)</i>	1.07 ± 0.01 <i>(GPR)</i>	122.53 ± 1.11 <i>(GPR)</i>	0.02 ± 0.00 <i>(GPR)</i>	1.46 ± 0.01 <i>(GPR)</i>	146.09 ± 1.31 <i>(GPR)</i>	0.01 ± 0.00 <i>(GPR)</i>	0.57 ± 0.00 <i>(GPR)</i>	39.83 ± 1.14 <i>(GPR)</i>	0.59 ± 0.08 <i>(SVR)</i>	0.22 ± 0.04 <i>(SVR)</i>	5.06 ± 9.42 <i>(SVR)</i>
Physically-based	CH	0.26 ± 0.17	0.59 ± 0.09	-24.33 ± 15.87	0.41 ± 0.00	0.53 ± 0.00	34.97 ± 0.37	0.53 ± 0.00	0.69 ± 0.00	51.81 ± 0.23	0.21 ± 0.00	0.51 ± 0.00	-40.07 ± 0.25	0.45 ± 0.00	0.75 ± 0.01	-182.43 ± 1.66
	ALB	0.14 ± 0.00	0.61 ± 0.00	-24.61 ± 0.00	0.43 ± 0.17	0.52 ± 0.05	32.34 ± 9.65	0.53 ± 0.00	0.66 ± 0.00	45.83 ± 0.00	0.20 ± 0.00	0.53 ± 0.00	-45.64 ± 0.00	0.44 ± 0.00	0.75 ± 0.00	-182.09 ± 0.00
	HAI	0.14 ± 0.00	0.61 ± 0.00	-24.61 ± 0.00	0.43 ± 0.00	0.51 ± 0.00	29.33 ± 0.00	0.66 ± 0.10	0.59 ± 0.05	42.64 ± 7.51	0.20 ± 0.00	0.53 ± 0.00	-45.64 ± 0.00	0.44 ± 0.00	0.75 ± 0.00	-182.09 ± 0.00
	SCH	0.12 ± 0.00	0.60 ± 0.00	-3.44 ± 0.00	0.45 ± 0.00	0.63 ± 0.00	60.49 ± 0.00	0.50 ± 0.00	0.85 ± 0.00	77.44 ± 0.00	0.30 ± 0.04	0.47 ± 0.04	-9.99 ± 6.58	0.43 ± 0.00	0.71 ± 0.00	-170.48 ± 0.00
	US	0.13 ± 0.00	0.64 ± 0.04	-18.58 ± 7.28	0.22 ± 0.21	0.68 ± 0.07	19.05 ± 37.63	0.26 ± 0.23	0.84 ± 0.02	41.44 ± 31.11	0.12 ± 0.12	0.60 ± 0.13	-45.40 ± 31.62	0.30 ± 0.24	0.71 ± 0.04	-167.32 ± 6.38
Hybrid	CH	0.23 ± 0.14	0.64 ± 0.18	-3.46 ± 17.01	0.07 ± 0.07	0.88 ± 0.06	55.91 ± 19.27	0.03 ± 0.03	1.13 ± 0.11	70.71 ± 24.06	0.02 ± 0.02	0.64 ± 0.06	-44.35 ± 27.74	0.13 ± 0.10	0.70 ± 0.10	-148.09 ± 28.52
	ALB	0.02 ± 0.02	0.73 ± 0.05	-31.66 ± 21.59	0.46 ± 0.18	0.46 ± 0.08	-0.37 ± 11.47	0.32 ± 0.04	0.63 ± 0.01	0.42 ± 12.18	0.23 ± 0.04	0.58 ± 0.04	-75.77 ± 12.38	0.12 ± 0.10	0.62 ± 0.15	-132.65 ± 46.57
	HAI	0.02 ± 0.03	0.72 ± 0.06	-36.42 ± 18.12	0.34 ± 0.07	0.52 ± 0.03	-19.74 ± 5.34	0.54 ± 0.11	0.51 ± 0.09	-9.55 ± 8.77	0.17 ± 0.07	0.65 ± 0.03	-94.74 ± 5.45	0.14 ± 0.11	0.59 ± 0.04	-120.22 ± 12.54
	SCH	0.03 ± 0.03	0.79 ± 0.07	36.96 ± 48.54	0.41 ± 0.09	0.77 ± 0.06	83.55 ± 6.82	0.25 ± 0.06	0.95 ± 0.04	78.01 ± 6.21	0.27 ± 0.15	0.47 ± 0.06	-0.14 ± 8.90	0.19 ± 0.05	0.45 ± 0.09	-48.41 ± 44.96
	US	0.01 ± 0.01	0.94 ± 0.20	93.61 ± 20.14	0.03 ± 0.02	1.14 ± 0.20	117.33 ± 42.37	0.02 ± 0.03	1.55 ± 0.27	146.89 ± 38.34	0.02 ± 0.03	0.65 ± 0.12	39.24 ± 43.46	0.23 ± 0.26	0.29 ± 0.07	0.49 ± 7.92

427

428 3.3. Global models

429 For empirical models, accuracy was generally low (Figure 5A-E). It was highest for the ALB
430 site, followed by the HAI and SCH sites, but accompanied by a high MBE. For the CH site, the
431 R^2 of the prediction was practically 0.00, while the systematic underestimation for the US site
432 amounted to -195.82 g/m^2 . Comparatively poor results were also produced by the physically-
433 based models in most cases (Figure 5F-J). An exception was the US site, for which an R^2 of
434 0.45 and an RRMSE of 0.75 were achieved. However, the systematic underestimation also
435 tended to be high with an MBE of -183.99 g/m^2 . Regarding hybrid models (Figure 5K-O), best
436 results were obtained for the models predicting the ALB, HAI, and SCH sites, although the
437 MBE indicated substantial over- and underestimations ranging from 37.30 for the ALB model
438 to -79.07 g/m^2 for the SCH model, respectively. No satisfactory performance was achieved for
439 the CH and US sites. Overall, the accuracy of global models was lower than for the local models
440 (with the exception of the physically-based model for the US site, Figure 5K).



441

442 Figure 5: Scatterplots of measured versus predicted biomass for the global empirical (A-E), physically-based (F-J), and hybrid models (L-O). Textboxes show coefficient of determination (R^2), relative
 443 root-mean-square error (RRMSE) and mean bias error (MBE) for each model. For empirical models, only the best-performing model in terms of lowest RRMSE is shown with the corresponding model
 444 name added in a separate textbox. CH: Switzerland (A, F, K), ALB: Schwäbische Alb (B, G, L), HAI: Hainich-Dün (C, H, M), SCH: Schorfheide-Chorin (D, I, N), US: United States (E, J, O), RFR: Random
 445 Forest regression, SVR: Support Vector regression, XGB: Extreme Gradient Boosting regression.

446

447 3.4. Epistemic uncertainty

448 For none of the comparisons (local, transferred, and global), a systematic relationship between
449 epistemic uncertainty and absolute differences between measured and predicted biomass values
450 was observed when considering all predicted data points (10 repetitions of 20 data points for
451 local and transferred models, 100 data points for global models; Tables 5 and 6). Only isolated
452 weak correlations were found, e.g., for local and transferred physically-based models predicting
453 the US site (R^2 up to 0.31) or the global hybrid model predicting the US site ($R^2 = 0.38$).

454 Only for the mean values of epistemic uncertainty and absolute difference, a few patterns could
455 be identified. Regarding the local empirical and hybrid models, a positive correlation between
456 mean epistemic uncertainty and mean absolute difference could be observed. The lowest
457 correlation values were found for the HAI site (for both model types) and the highest values for
458 the SCH and CH sites (for empirical and hybrid models, respectively). The physically-based
459 models displayed a negative correlation; the US site exhibited the lowest mean epistemic
460 uncertainty but the highest mean absolute difference (52.71 and 162.16 g/m^2 , respectively).

461 For the transferred models, a clear pattern between combinations of training or optimization
462 and prediction site was found. The respective empirical and hybrid models of the ALB and HAI
463 sites exhibited low values for both mean epistemic uncertainty and absolute difference for
464 mutual prediction. The highest values for mean epistemic uncertainty and absolute difference
465 were produced by the CH and US models of both model types. For the transferred physically-
466 based models, the lowest values for mean epistemic uncertainty were observed for the US site
467 (38.86 g/m^2 for ALB and HAI models), which did not coincide with the high values for mean
468 absolute difference (182.09 g/m^2 for ALB and HAI models). For the global models, no
469 agreement between mean epistemic uncertainty and mean absolute difference was found for
470 any model type or prediction site.

471 Table 5: Mean difference between predicted and measured grassland biomass and epistemic uncertainty ± 1 standard deviation and coefficient of determination (R^2) for the 10 local (in italic) and
472 local transferred models. For the local models, 200 samples (10 predictions of 20 samples each) and for the transferred models, 1000 samples (10 predictions of 100 samples each) were available. If
473 the percentage of solutions of a given physically-based model was 0.01%, only 1 solution was used to derive the predicted biomass value and no epistemic uncertainty could be calculated for the
474 corresponding predicted samples (see Appendix A Section A.5). The absolute difference was calculated by taking the absolute value after subtracting the measured biomass value from the predicted
475 biomass value for each predicted sample. For the empirical and hybrid models, the epistemic uncertainty for a predicted sample corresponds to the predicted standard deviation of the Gaussian
476 process regression (GPR) models. For the physically-based models, the epistemic uncertainty corresponds to the standard deviation of all selected solutions. CH: Switzerland, ALB: Schwäbische Alb,
477 HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States.

		Prediction site														
		CH			ALB			HAI			SCH			US		
Model type	Training/optimization site	Absolute difference [g m ⁻²]	Epistemic uncertainty [g m ⁻²]	R ²	Absolute difference [g m ⁻²]	Epistemic uncertainty [g m ⁻²]	R ²	Absolute difference [g m ⁻²]	Epistemic uncertainty [g m ⁻²]	R ²	Absolute difference [g m ⁻²]	Epistemic uncertainty [g m ⁻²]	R ²	Absolute difference [g m ⁻²]	Epistemic uncertainty [g m ⁻²]	R ²
Empirical	CH	<i>83.69 ± 74.12</i>	<i>94.67 ± 21.25</i>	<i>0.08</i>	<i>98.15 ± 71.22</i>	<i>108.78 ± 11.78</i>	<i>0.18</i>	<i>119.89 ± 72.50</i>	<i>109.00 ± 10.10</i>	<i>0.19</i>	<i>101.10 ± 82.35</i>	<i>111.90 ± 8.79</i>	<i>0.02</i>	<i>94.41 ± 81.49</i>	<i>116.21 ± 3.83</i>	<i>0.03</i>
	ALB	<i>90.79 ± 77.02</i>	<i>82.00 ± 3.57</i>	<i>0.10</i>	<i>55.08 ± 41.92</i>	<i>64.64 ± 14.89</i>	<i>0.01</i>	<i>53.80 ± 48.66</i>	<i>70.18 ± 12.51</i>	<i>0.05</i>	<i>96.30 ± 90.96</i>	<i>69.07 ± 11.85</i>	<i>0.00</i>	<i>123.80 ± 84.85</i>	<i>82.86 ± 1.06</i>	<i>0.00</i>
	HAI	<i>96.51 ± 83.30</i>	<i>87.96 ± 7.87</i>	<i>0.03</i>	<i>58.07 ± 44.78</i>	<i>65.96 ± 16.17</i>	<i>0.02</i>	<i>41.89 ± 43.50</i>	<i>57.10 ± 17.38</i>	<i>0.03</i>	<i>102.05 ± 95.57</i>	<i>65.59 ± 15.87</i>	<i>0.07</i>	<i>146.93 ± 86.94</i>	<i>89.98 ± 6.35</i>	<i>0.00</i>
	SCH	<i>100.43 ± 66.40</i>	<i>123.63 ± 5.15</i>	<i>0.00</i>	<i>100.29 ± 62.11</i>	<i>111.75 ± 14.43</i>	<i>0.05</i>	<i>105.34 ± 57.31</i>	<i>111.95 ± 13.54</i>	<i>0.13</i>	<i>86.17 ± 76.45</i>	<i>110.94 ± 14.80</i>	<i>0.02</i>	<i>69.60 ± 64.11</i>	<i>123.97 ± 5.05</i>	<i>0.00</i>
	US	<i>120.43 ± 73.51</i>	<i>96.99 ± 7.78</i>	<i>0.00</i>	<i>129.78 ± 73.50</i>	<i>96.99 ± 7.78</i>	<i>0.00</i>	<i>156.39 ± 65.59</i>	<i>96.98 ± 7.78</i>	<i>0.00</i>	<i>104.73 ± 69.81</i>	<i>96.99 ± 7.78</i>	<i>0.00</i>	<i>54.01 ± 48.06</i>	<i>66.35 ± 19.96</i>	<i>0.01</i>
Physically-based	CH	<i>72.56 ± 74.38</i>	<i>72.64 ± 28.22</i>	<i>0.02</i>	<i>64.22 ± 37.31</i>	<i>90.64 ± 25.47</i>	<i>0.00</i>	<i>68.23 ± 41.77</i>	<i>87.73 ± 24.39</i>	<i>0.01</i>	<i>80.90 ± 80.60</i>	<i>95.65 ± 21.07</i>	<i>0.04</i>	<i>182.43 ± 73.30</i>	<i>40.22 ± 11.08</i>	<i>0.20</i>
	ALB	<i>76.45 ± 72.35</i>	<i>71.98 ± 24.64</i>	<i>0.01</i>	<i>63.10 ± 37.78</i>	<i>88.68 ± 25.09</i>	<i>0.00</i>	<i>64.28 ± 41.40</i>	<i>84.10 ± 24.24</i>	<i>0.02</i>	<i>81.28 ± 83.70</i>	<i>92.55 ± 20.37</i>	<i>0.03</i>	<i>182.09 ± 73.19</i>	<i>38.86 ± 10.85</i>	<i>0.21</i>
	HAI	<i>76.45 ± 72.35</i>	<i>71.98 ± 24.64</i>	<i>0.01</i>	<i>60.97 ± 36.65</i>	<i>86.64 ± 25.29</i>	<i>0.00</i>	<i>59.18 ± 34.42</i>	<i>81.59 ± 25.56</i>	<i>0.00</i>	<i>81.28 ± 83.70</i>	<i>92.55 ± 20.37</i>	<i>0.03</i>	<i>182.09 ± 73.19</i>	<i>38.86 ± 10.85</i>	<i>0.21</i>
	SCH	<i>79.70 ± 67.66</i>	<i>95.62 ± 24.64</i>	<i>0.01</i>	<i>76.02 ± 43.42</i>	<i>103.37 ± 27.38</i>	<i>0.04</i>	<i>87.85 ± 44.65</i>	<i>101.54 ± 24.53</i>	<i>0.05</i>	<i>76.14 ± 71.71</i>	<i>109.97 ± 19.35</i>	<i>0.02</i>	<i>170.48 ± 70.77</i>	<i>51.96 ± 15.66</i>	<i>0.17</i>
	US	<i>77.93 ± 69.49</i>	<i>85.86 ± 26.66</i>	<i>0.01</i>	<i>74.45 ± 42.24</i>	<i>100.68 ± 28.68</i>	<i>0.03</i>	<i>83.05 ± 46.07</i>	<i>97.41 ± 27.84</i>	<i>0.07</i>	<i>77.86 ± 71.21</i>	<i>107.35 ± 23.30</i>	<i>0.03</i>	<i>162.16 ± 64.37</i>	<i>52.71 ± 8.87</i>	<i>0.31</i>
Hybrid	CH	<i>85.15 ± 79.45</i>	<i>168.42 ± 64.24</i>	<i>0.02</i>	<i>100.74 ± 69.41</i>	<i>116.83 ± 57.08</i>	<i>0.10</i>	<i>113.54 ± 67.88</i>	<i>84.32 ± 47.26</i>	<i>0.00</i>	<i>109.92 ± 92.14</i>	<i>102.34 ± 55.32</i>	<i>0.02</i>	<i>151.07 ± 108.91</i>	<i>152.27 ± 53.92</i>	<i>0.19</i>
	ALB	<i>96.76 ± 82.75</i>	<i>127.65 ± 34.26</i>	<i>0.03</i>	<i>49.07 ± 43.37</i>	<i>100.59 ± 38.06</i>	<i>0.00</i>	<i>52.49 ± 50.49</i>	<i>76.32 ± 33.13</i>	<i>0.04</i>	<i>96.64 ± 86.68</i>	<i>84.99 ± 36.44</i>	<i>0.06</i>	<i>137.02 ± 92.86</i>	<i>129.28 ± 28.89</i>	<i>0.25</i>
	HAI	<i>94.79 ± 82.74</i>	<i>130.98 ± 27.94</i>	<i>0.03</i>	<i>52.71 ± 49.66</i>	<i>101.47 ± 30.03</i>	<i>0.00</i>	<i>41.62 ± 42.70</i>	<i>83.56 ± 29.03</i>	<i>0.01</i>	<i>106.33 ± 98.54</i>	<i>89.19 ± 31.22</i>	<i>0.07</i>	<i>124.46 ± 93.54</i>	<i>135.26 ± 23.55</i>	<i>0.17</i>
	SCH	<i>111.38 ± 80.29</i>	<i>138.79 ± 41.65</i>	<i>0.00</i>	<i>93.55 ± 52.99</i>	<i>109.98 ± 42.53</i>	<i>0.00</i>	<i>92.91 ± 59.13</i>	<i>87.59 ± 43.11</i>	<i>0.05</i>	<i>79.79 ± 68.74</i>	<i>100.39 ± 44.62</i>	<i>0.08</i>	<i>92.30 ± 77.51</i>	<i>144.42 ± 34.61</i>	<i>0.17</i>
	US	<i>135.55 ± 98.87</i>	<i>137.77 ± 47.69</i>	<i>0.01</i>	<i>135.62 ± 86.92</i>	<i>120.75 ± 38.63</i>	<i>0.04</i>	<i>163.14 ± 82.09</i>	<i>107.63 ± 33.42</i>	<i>0.00</i>	<i>118.59 ± 85.61</i>	<i>112.79 ± 36.61</i>	<i>0.00</i>	<i>58.76 ± 53.06</i>	<i>135.36 ± 38.89</i>	<i>0.04</i>

478 Table 6: Mean difference between predicted and measured grassland biomass and epistemic uncertainty ± 1 standard deviation and coefficient of determination (R^2) for global models. For each
 479 model, 100 samples (1 prediction of 100 samples each) were available. The absolute difference was calculated by taking the absolute value after subtracting the measured biomass value from the
 480 predicted biomass value for each predicted sample. For the empirical and hybrid models, the epistemic uncertainty for a predicted sample corresponds to the predicted standard deviation of the
 481 Gaussian process regression (GPR) models. For the physically-based models, the epistemic uncertainty corresponds to the standard deviation of all selected solutions. CH: Switzerland, ALB:
 482 Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States.

Model type	Prediction site														
	CH			ALB			HAI			SCH			US		
	Absolute difference [g m ⁻²]	Epistemic uncertainty [g m ⁻²]	R ²	Absolute difference [g m ⁻²]	Epistemic uncertainty [g m ⁻²]	R ²	Absolute difference [g m ⁻²]	Epistemic uncertainty [g m ⁻²]	R ²	Absolute difference [g m ⁻²]	Epistemic uncertainty [g m ⁻²]	R ²	Absolute difference [g m ⁻²]	Epistemic uncertainty [g m ⁻²]	R ²
Empirical	89.36 \pm 65.55	112.11 \pm 0.00	0.00	83.81 \pm 55.91	115.07 \pm 0.00	0.00	105.16 \pm 56.70	111.4 \pm 0.00	0.00	92.12 \pm 90.44	107.64 \pm 0.00	0.00	102.31 \pm 81.66	108.8 \pm 0.00	0.00
Physically-based	81.37 \pm 71.47	83.49 \pm 27.37	0.00	74.45 \pm 42.24	100.68 \pm 28.68	0.03	83.05 \pm 46.07	97.41 \pm 27.84	0.07	78.35 \pm 79.36	94.48 \pm 21.91	0.01	182.99 \pm 73.17	39.74 \pm 11.05	0.20
Hybrid	98.72 \pm 75.78	138.74 \pm 25.66	0.01	62.54 \pm 40.14	138.67 \pm 51.71	0.03	57.72 \pm 45.84	88.45 \pm 24.39	0.14	97.38 \pm 92.43	106.99 \pm 34.61	0.11	101.97 \pm 90.03	132.54 \pm 7.63	0.38

483

484

485 **4. Discussion**

486 Remote sensing has gained traction in its application for monitoring ecosystem functioning,
487 with spatial (and temporal) scalability often highlighted as a key advantage. When developing
488 scalable models, model transferability should be an integral part of performance assessment;
489 however, systematic evaluation often lags behind the ambition for scalability within and across
490 ecosystems. In this study, we demonstrated the importance of assessing model transferability
491 when evaluating the performance of remotely sensed grassland biomass estimates and related
492 model selection considerations. Our model comparisons indicated that in most cases physically-
493 based models exhibited the highest transferability when applied to unseen grassland sites.
494 However, no single model consistently outperformed others when trained or optimized with
495 data from multiple sites. These results underscore the challenges in developing scalable models,
496 highlight the importance and possible trade-offs of appropriate model selection, and shed light
497 on the discrepancies between epistemic uncertainty and predictive accuracy.

498 4.1. Accuracy of local models

499 All model types performed similarly when predicting biomass locally with accuracies
500 comparable to those of other studies conducted in alpine and semi-natural grasslands in
501 Switzerland and Germany (Raab et al., 2020; Schweiger et al., 2015a). Generally, model
502 accuracy decreased with increasing biomass range and (to a lesser extent) spectral variability
503 (see also Dehghan-Shoar et al., 2023, Figure 1, and Appendix A Section A.18). Overall,
504 empirical models performed best at the site level, presumably due to their high flexibility,
505 allowing them to incorporate site-specific relationships in the data without necessarily relying
506 on physical principles.

507 For the physical models, we found a large discrepancy between the *in situ* and modeled data
508 for the US site caused by the employed LUT parametrization. Although the models recognized
509 the underlying relationship between biomass and spectral information to a certain extent (mean
510 $R^2 = 0.30$), a high RRMSE of 0.71 resulted probably due to a too narrow range of values for
511 LMA in the LUT since testing an alternative value range with a higher upper bound for LMA
512 resulted in a mean R^2 of 0.51 and mean RRMSE of 0.33 (Appendix A Section A.19). An
513 evaluation of the agreement between the PROSAIL simulations and Sentinel-2 spectra of the
514 different study sites as an indicator of the respective PROSAIL forward mode fit quality can be
515 found in Appendix A Section A.20. However, broadening the trait ranges increased ill-
516 posedness (Combal et al., 2003; Verrelst et al., 2014) and deteriorated model performance for
517 the other sites; therefore, a site-specific parametrization of multiple narrowly defined LUTs
518 would be necessary to obtain comparable results to the empirical models for all sites.

519 The hybrid models performed similarly compared to the empirical models. Selecting training
520 samples with AL led to substantial improvements in performance as stated in AL theory
521 (Verrelst et al., 2016). In our case, the use of AL for selecting the most informative optimization
522 samples was essential because we did not parameterize the LUT *a priori* or excluded
523 biophysically unrealistic simulations as done by others (see, e.g., Campos-Taberner et al.,
524 2018). Moreover, the hybrid models did not display the aforementioned imbalance between R^2
525 and RRMSE for the US site although having been trained on a suboptimally configured LUT,
526 presumably because the employed GPR kernel was able to scale the predictions by the
527 magnitude of the employed constant kernel resulting in a lower RRMSE compared to the
528 physically-based models while keeping R^2 comparatively high (scikit-learn Developers, 2023).

529 The fact that local models in grasslands sometimes underperform compared to croplands is
530 largely due to the specific particularities of grassland systems. Grasslands often contain a
531 complex mixture of photosynthetic and non-photosynthetic vegetation, with litter potentially

532 contributing a substantial fraction of total biomass (Schweiger et al., 2015a). Contrasting
533 absorption and scattering properties of litter compared to green biomass can reduce the
534 sensitivity of reflectance-based predictors to total biomass. High species and trait diversity of
535 grasslands is reflected in heterogeneous canopy architectures that violate the homogeneity
536 assumptions underlying radiative transfer models such as PROSAIL (Rossi et al., 2020). In
537 addition, strong fine-scale spatial heterogeneity driven by grazing by wild and domestic
538 herbivores further confounds the link between plot-scale measurements and satellite-based
539 observations.

540 4.2. Transferability of local models

541 Several factors possibly hampered model transferability. First, the transferability of empirical
542 and hybrid models remained limited to combinations of training or optimization and validation
543 sites sharing similar environmental and management characteristics such as ALB and HAI. This
544 is consistent with the findings of Muro et al. (2022) who, among the German study sites, found
545 the prediction of SCH to be the most challenging due to an elevated level of soil organic content
546 and different management compared to ALB and HAI (Busch et al., 2018). At the other
547 extreme, the US site, which was poorly predicted by transferred models, was sampled during a
548 very dry summer, and the Sentinel-2 spectra showed a substantial increase in reflectance in the
549 short-wave infrared region, likely due to low water content confounding the overall spectra
550 (Jacquemoud et al., 2009; Mesonet, 2024; Appendix A Section A.21). The summer drought
551 may have also led to early senescence, resulting in a potential underestimation of LAI in
552 PROSAIL as the latter is predominantly equipped for modelling green vegetation and struggles
553 to capture NPV (Amin et al., 2021; Delegido et al., 2015; Schiefer et al., 2021; Verrelst et al.,
554 2023). Second, differences in the abundance of C3 and C4 grasses among sites may have further
555 limited model transferability from European sites to the US exhibiting a substantially higher
556 cover fraction of C4 grasses (Kothari and Schweiger, 2022; Shoko et al., 2016; Appendix A

557 Section A.22). Third, the derivation of biomass values in physically-based and hybrid
558 approaches via LUT-based LMA and LAI multiplication resulted in numerous parameter
559 combinations associated with the same biomass values, increasing the prevalent ill-posedness
560 of the model. Fourth, plot definition and field measurements were handled differently among
561 the study sites and field campaigns. For example, the plots being monitored by the Biodiversity
562 Exploratories were selected to be representative of an area of 2,500 m² and biomass was cut for
563 an area of 2 m² (Hinderling et al., 2023), whereas the plots in Switzerland were representative
564 for an area of 36 m² and biomass harvesting was limited to 1 m² (Schweiger et al., 2015b).
565 While such discrepancies between monitoring programs are likely to introduce uncertainties,
566 they are difficult to mitigate when working with already existing field data. Moreover, the *in*
567 *situ* data collection might not necessarily have been optimized for the pixel grid of the utilized
568 remote sensing data, e.g., 20 × 20 m grid for Sentinel-2, leading to the problem of diminished
569 representativeness (Hauser et al., 2021b; Schweiger, 2020) such as in the case of mixed pixels.
570 Even though standardization of field campaigns could potentially benefit the transferability of
571 model trained with remote sensing data, it might be hardly feasible due to ecological or policy
572 constraints and the legitimate interest of existing monitoring programs to ensure temporally
573 consistent measurements.

574 Nevertheless, physically-based models outperformed empirical and hybrid models when
575 predicting novel study sites in an agnostic scenario which is consistent with the claimed
576 transferability of physically-based models and the results of previous studies conducted at leaf
577 level (Féret et al., 2019; Verrelst et al., 2015; Wang et al., 2023). For the hybrid models, we
578 expected a comparatively high transferability due to the physical foundation of the RTM-
579 simulated training data. In addition, previous studies have shown the general ability of machine
580 learning regression models to serve as RTM emulators, i.e., being able to accurately grasp the
581 physical principles of RTMs during model training (Rivera et al., 2015; Verrelst et al., 2017).

582 However, our results showed that the transferability of the hybrid models was weakest in most
583 cases, potentially because the empirical features of hybrid models overrode the physical
584 foundation of the training data by fitting data-specific single-trait relationships between the AL-
585 selected training set and the optimization data. To increase the transferability of hybrid models,
586 it has been suggested to increase the initial training set size (Berger et al., 2021b; Tagliabue et
587 al., 2022), enforcing the inclusion of more general training data prior to the employment of AL.
588 However, no general increase in transferability was apparent in our results even when using
589 different initial training set sizes (Appendix A Section A.23).

590 4.3. Accuracy of global models

591 In this study, global models were defined as models that were trained (empirical) or optimized
592 (physically-based and hybrid) using field data from multiple sites and applied to an unseen
593 prediction site. The reasoning behind this model set-up was that model training or optimization
594 with heterogeneous field data might increase model transferability by allowing the model to
595 learn from a larger pool of diverse field data motivated by previous studies employing similar
596 approaches (Muro et al., 2022) and the emergence of global plant trait products derived from
597 remote sensing data using physically-based or machine learning models (Campos-Taberner et
598 al., 2018; Kovács et al., 2023; Moreno-Martinez et al., 2020) relying on accurate predictions
599 for unsampled areas. The results of this third model comparison were not conclusive as to which
600 model performed best at this task.

601 The global physically-based models achieved lower accuracy than the local models for all sites,
602 indicating a strong influence of the selected combinations of cost function and percentage of
603 solutions, which differed strongly from those of the local models (Appendix A Sections A.5
604 and A.8). This seems to contradict the observed higher transferability of the transferred
605 physically-based models and the eventual conclusion that physically-based models are always
606 the best choice if no validation data is available (see Section 4.2).

607 One major challenge for global empirical and hybrid models was that different confounding
608 relationships between biomass and spectral information are possible for various study sites.
609 This was reflected for the empirical and hybrid models by the increased mean RRMSE of the
610 US site, whose specific conditions could not be adequately learned by the models without
611 additional contextual information (in the case of the global physically-based models, the reason
612 for the high mean RRMSE was most likely the parametrization of the LUT as discussed in 4.1).
613 Nevertheless, comparatively good results were obtained by the empirical models for the ALB
614 site ($R^2 = 0.49$, RRMSE = 0.49, MBE = 28.87 g/m²) and by the hybrid models for the ALB (R^2
615 = 0.43, RRMSE = 0.53, MBE = 37.30 g/m²) and HAI sites ($R^2 = 0.42$, RRMSE = 0.64, MBE =
616 34.41 g/m²) which is consistent with the results of Muro et al. (2022) who reported a lower
617 transferability of models applied to the SCH site due to confounding factors such as soil organic
618 content and management practices (Busch et al., 2018). Identifying the exact factors
619 contributing to the divergent performance is challenging within the multi-site training set-up
620 and limits definitive conclusions to be drawn on model selection and suitability.

621 4.4. Prediction uncertainty as a measure of model applicability to unseen data

622 Value and acceptance of remote sensing products increase with a quantitative specification of
623 uncertainty (Woodcock, 2002); increased attention to uncertainty is also called for in the context
624 of machine learning applications (Meyer and Pebesma, 2020). Motivated by the finding that
625 epistemic uncertainty of GPR models might help to identify reliable and unreliable predictions
626 for croplands (Verrelst et al., 2013b, 2013a, 2012), we tested if epistemic uncertainty could be
627 used as proxy for model transferability across heterogeneous grasslands by comparing the
628 absolute difference between measured and predicted biomass with the epistemic uncertainty
629 associated with each prediction and found low correspondence between the two (as shown in
630 Section 3.4).

631 In the case of GPR, epistemic uncertainty is a direct model output in the form of the predictive
632 SD. According to Rasmussen and Williams (2006), the epistemic uncertainty of GPR models
633 is determined by the similarity between a data point to be predicted and the training data as well
634 as the properties of the optimized kernel, e.g., the magnitude of the constant kernel and the
635 lengthscale of the RBF kernel. Accordingly, it is not surprising that the empirical models for
636 the mutual prediction of the spectrally more similar ALB and HAI sites showed lower mean
637 epistemic uncertainty than the US models, whose training data exhibited different spectral
638 properties, e.g., the high reflectance in the short-wave infrared region for the US site. The fact
639 that the AL-selected training data of the hybrid models did not differ much among the study
640 sites (Appendix A Section A.24) led to a consistently lower epistemic uncertainty for the
641 German sites, for which the spectral properties of the field data were more similar to those of
642 the AL-selected training data. Correspondingly, a high mean epistemic uncertainty resulted for
643 the CH and US sites, since these sites exhibited larger spectral heterogeneity (as shown in
644 Appendix A Section A.18) or contained spectra that were not fully covered by the current
645 parametrization of the PROSAIL LUT, respectively. Thus, a high epistemic uncertainty
646 indicated that the data points to be predicted are not optimally covered by the training data,
647 while a low epistemic uncertainty on the contrary did not necessarily indicate a reliable
648 prediction as suggested by the missing correspondence between low absolute differences and
649 associated epistemic uncertainty. Hence, using the epistemic uncertainty as a measure of model
650 applicability to unseen data might be valid only in the case of comparable relationships between
651 biomass and spectral information – a condition violated in our study. Nevertheless, the GPR
652 uncertainty still serves as valuable information to optimize field sampling efforts as outlined in
653 Verrelst et al. (2012).

654 In the case of physically-based models, the SD of the solutions considered in the LUT inversion
655 process is a comparatively simple approach to express the variability of the candidate solutions

656 and thus to quantify the diversity of the possible solutions. This diversity was relatively
657 consistent per prediction site, e.g., mean epistemic uncertainty was between 38.86 and 51.96
658 g/m^2 for predictions of the US site, suggesting a subordinate influence of the cost functions and
659 percentages of solutions used for different LUT inversions. Rather, the epistemic uncertainty
660 appeared to depend primarily on the spectral properties of the predicted data. For example, the
661 US site had by far the lowest epistemic uncertainty, which was not consistent with the
662 comparatively high mean RRMSE. In fact, a more accurate prediction of the US site would
663 have required a modification of the LUT parameterization (as shown in Appendix A Section
664 A.19), which is obscured by relying solely on the prediction uncertainty. Therefore, in our
665 study, the epistemic uncertainty expressed as diversity of the possible solutions did not allow a
666 direct conclusion on the applicability of the models to unseen data.

667 4.5. Possible improvements

668 Regarding the performance of local models, various opportunities for improvement are
669 conceivable depending on the model type. First, the use of alternative algorithms or cost
670 functions should be considered. For empirical models, alternative algorithms such as Artificial
671 Neural Networks (ANNs) as employed by Ali et al. (2017) and Muro et al. (2022) could be
672 tested to determine further improvements in model accuracy. For the physically-based models,
673 the cost functions discussed by Rivera et al. (2013) were tested but since the choice of the
674 optimal cost functions varied between study sites, the inclusion of additional cost functions
675 such as genetic algorithms (Fang et al., 2003) could benefit the accuracy of the LUT inversion.
676 Second, subsetting the LUT using correlations between PROSAIL input parameters could
677 reduce solution space (Campos-Taberner et al., 2018). For this, *in situ* data of several PROSAIL
678 input parameters are necessary whose correlations could be exploited to prevent unrealistic
679 parameter combinations being included in the LUT (Combal et al., 2003) as done by Schiefer
680 et al. (2021). Ideally, the *in situ* data should be collected for each study site, since the correlation

681 between the parameters is subject to environmental gradients such as climatic and soil factors
682 (Joswig et al., 2021), laying the foundation for a locally informed scenario. In such a scenario,
683 further variables besides the aforementioned LMA (see Section 4.1) could have played a key
684 role in increasing the representativeness of the simulated LUT for site-specific conditions. On
685 the one hand, Berger et al. (2021a) and Verrelst et al. (2023) have shown that NPV can be
686 modeled by coupling PROSPECT-PRO (Féret et al., 2021), the successor model to
687 PROSPECT-D, with 4SAIL by combining carbon-based constituents (CBC) and LAI. The
688 approximation of NPV would allow the local ratio of green and brown vegetation to be better
689 represented. On the other hand, the default spectra for dry and wet soil of the prosail R package
690 could be replaced by locally measured spectra or spectra contained in a spectral library in order
691 to take their spatial variability into account. Using existing soil spectral libraries (Safanelli et
692 al., 2025; Viscarra Rossel et al., 2016), the default spectra used for the PROSAIL simulation
693 could be replaced by spectra tailored to local soil properties. Third, the inclusion of additional
694 predictor variables could improve model accuracy and transferability, for example additional
695 biomass-relevant variables such as vegetation height could lead to improved predictions.
696 However, sampling of additional variables is time- and labor-intensive and deriving them from
697 remote sensing data is challenging, highlighting the need for reliable, accurate, and transferable
698 remote sensing models. For example, the mean average error (MAE) of the estimated canopy
699 height from spaceborne laser altimeters amounts to at least 2 m (Liu et al., 2021) and the use of
700 Sentinel-1 SAR data has led to mixed results, ranging from improving grassland biomass
701 estimation to providing little or no added value when combined with optical data (Muro et al.,
702 2022; Raab et al., 2020; Wang et al., 2019). For hybrid models, it is possible to employ further
703 advanced models such as multi-output GPR models (MOGPs) if additional *in situ* measured
704 PROSAIL parameters are available, which allow to predict multiple output variables while
705 preserving the correlations among the input variables (Liu et al., 2018). The feasibility of
706 employing MOGPs in the context of biophysical variable estimation has already been

707 demonstrated by Pipia et al. (2019) and Caballero et al. (2023) who predicted LAI and
708 vegetation water content (VWC), respectively, by combining Sentinel-1 and Sentinel-2 data.

709 To improve the local models' transferability as well as the global models' accuracy, it is
710 essential to disentangle the site-specific confounded relationships between biomass and spectral
711 information, effectively shifting from agnostic to locally informed biomass models. First,
712 additional predictor variables could help to distinguish these relationships, including vegetation
713 height, climate data, land use intensity as continuous proxy for management information
714 (Blüthgen et al., 2012; Rossi et al., 2024), evapotranspiration, and the quantification of NPV.
715 Second, a stratification based on climatic priors for subsequent LUT optimization would be
716 useful in the context of diverse study sites and resonates with studies making the case for a
717 stronger embedding of remote sensing data with ancillary data (Aguirre-Gutiérrez et al., 2021;
718 Cavender-Bares et al., 2022; Moreno-Martinez et al., 2020; Verrelst et al., 2023). Although a
719 remote sensing-only solution may have been desirable, our results reaffirm, through a new
720 perspective, the trade-off between a LUT parametrization covering the ecological conditions of
721 all sites and a degradation of the LUT inversion accuracy. Such a stratification would allow, in
722 the first step, the data points to be projected into different environmental regimes based on
723 climate data and in the second step, to use LUTs subsetted based on correlations between
724 PROSAIL parameters and with optimized parameter ranges for the physically-based or hybrid
725 models. While coarse-resolution climate data for the first step are freely available over large
726 spatial scales (e.g., ERA5-Land; Muñoz Sabater, 2019), the second step still places high
727 demand on the availability of *in situ* data.

728 **Conclusion**

729 In this study, we compared different models for biomass estimation from multispectral
730 spaceborne remote sensing data across heterogeneous grasslands to assess model assumptions
731 and facilitate model choice for specific applications. In three model comparisons, we
732 investigated the accuracy and transferability of empirical, physically-based, and hybrid models
733 across five study sites regarding (1) their local applicability, (2) their spatial transferability, and
734 (3) the opportunity to compile field data from multiple study sites to increase transferability.
735 Our results showed that:

- 736 1) on the local level, all models performed similarly well in terms of RRMSE and R^2 .
737 Further, we found that in the context of hybrid models, employing AL to identify the
738 most informative training samples was required.
- 739 2) when transferring local models to a different study site, the physically-based models led
740 to the most promising results for most combinations of training (empirical models) or
741 optimization (physically-based and hybrid models) and validation sites. Moreover, we
742 observed a trade-off between LUT specificity and generality, impeding the universal
743 application of a single physically-based model. The transferability of empirical and
744 hybrid models was limited to combinations of sites sharing similar ecological and/or
745 spectral conditions.
- 746 3) when compiling the field data of four study sites to predict the remaining one, no model
747 clearly outperformed the others. Differences in model performance remained
748 challenging to explain, highlighting the need to further explore the possibilities and
749 characterize the trade-offs of developing models applicable on a large spatial scale and
750 across ecological gradients.
- 751 4) common epistemic uncertainty implementations were not necessarily reliable measures
752 of model applicability to unseen data in the case of varying relationships between

753 biomass and spectral information across study sites. While high epistemic uncertainty
754 indicates suboptimal coverage of the data points to be predicted by the training data and
755 can therefore guide future sampling efforts, low epistemic uncertainty does not
756 necessarily indicate high prediction accuracy.

757 5) model transferability needs to be thoroughly tested when developing remote sensing
758 applications whose intended applicability goes beyond the local scale.

759 Possibilities for improving local model accuracy and transferability include the testing of
760 alternative machine and deep learning algorithms and cost functions, hierarchical subsetting of
761 LUTs based on ecological priors, and the use of multi-output models to preserve correlations
762 among predictor variables, incorporating additional predictors such as NPV or climate data, and
763 exploring multi-sensor approaches. To fully exploit the spatio-temporal potential of satellite
764 observations, evaluating and improving model transferability should be a priority; this requires
765 datasets that enable rigorous testing across sites and conditions, along with theory-driven
766 foundations to build models capable of robust prediction when applied to new unseen data.

767 **CRedit authorship contribution statement**

768 **Jan Schweizer:** Conceptualization, Methodology, Software, Investigation, Data Curation,
769 Writing – Original draft, Visualization. **Leon T. Hauser:** Conceptualization, Methodology,
770 Investigation, Data Curation, Writing – Original draft, Supervision, Funding Acquisition.
771 **Hamed Gholizadeh:** Data Curation, Writing – Review & Editing. **Anna-Katharina**
772 **Schweiger:** Data Curation, Writing – Review & Editing. **Christian Rossi:** Conceptualization,
773 Methodology, Investigation, Data Curation, Writing – Original draft, Supervision, Funding
774 Acquisition.

775

776 **Acknowledgements**

777 This work was conducted under the ‘Biodiv-Watch’ project, funded by the ESA Initial Support
778 for Innovation (EISI) [Grant Number 4000138090], and with the financial and operational
779 support of the Swiss National Park (SNP). Hamed Gholizadeh was supported by a NASA NIP
780 award [80NSSC21K0941]. This work is based on data obtained within the DFG Priority
781 Program 1374 ‘Infrastructure-Biodiversity-Exploratories’. We thank the staff of the three
782 exploratories, the BE office and the BExIS team for their work in maintaining the plot and
783 project infrastructure, and Markus Fischer, the late Elisabeth Kalko, Eduard Linsenmair,
784 Dominik Hessenmöller, Jens Nieschulze, Daniel Prati, Ingo Schöning, François Buscot, Ernst-
785 Detlef Schulze and Wolfgang W. Weisser for their role in setting up the Biodiversity
786 Exploratories project. Field work permits were issued by the responsible state environmental
787 offices of Baden-Württemberg, Thüringen, and Brandenburg (according to § 72 BbgNatSchG).
788 The authors thank Antonia Ludwig for her insightful comments on an earlier version of the
789 manuscript and Matteo Delucchi of the consulting team of the Applied Statistics group at the
790 Department of Mathematical Modeling and Machine Learning, University of Zurich.

791 **Data availability**

792 This work is based on data elaborated by the BExIS and Botany core projects of the Biodiversity
793 Exploratories program (DFG Priority Program 1374). The datasets are publicly available in the
794 Biodiversity Exploratories Information System (<http://doi.org/10.17616/R32P9Q>). The
795 datasets are listed in the references section. Field data for Switzerland and Oklahoma are
796 available upon request. Python codes and synthetic mock data are available upon request.

797 **References**

- 798 Aguirre-Gutiérrez, J., Rifai, S., Shenkin, A., Oliveras, I., Bentley, L.P., Svátek, M., Girardin, C.A.J., Both,
799 S., Riutta, T., Berenguer, E., Kissling, W.D., Bauman, D., Raab, N., Moore, S., Farfan-Rios, W.,
800 Figueiredo, A.E.S., Reis, S.M., Ndong, J.E., Ondo, F.E., N’ssi Bengone, N., Mihindou, V., Moraes
801 de Seixas, M.M., Adu-Bredu, S., Abernethy, K., Asner, G.P., Barlow, J., Burslem, D.F.R.P.,
802 Coomes, D.A., Cernusak, L.A., Dargie, G.C., Enquist, B.J., Ewers, R.M., Ferreira, J., Jeffery, K.J.,
803 Joly, C.A., Lewis, S.L., Marimon-Junior, B.H., Martin, R.E., Morandi, P.S., Phillips, O.L., Quesada,
804 C.A., Salinas, N., Schwantes Marimon, B., Silman, M., Teh, Y.A., White, L.J.T., Malhi, Y., 2021.
805 Pantropical modelling of canopy functional traits using Sentinel-2 remote sensing data. *Remote*
806 *Sens Environ* 252, 112122. <https://doi.org/10.1016/j.rse.2020.112122>
- 807 Ali, I., Cawkwell, F., Dwyer, E., Barrett, B., Green, S., 2016. Satellite remote sensing of grasslands:
808 from observation to management. *Journal of Plant Ecology* 9, 649–671.
809 <https://doi.org/10.1093/jpe/rtw005>
- 810 Ali, I., Cawkwell, F., Dwyer, E., Green, S., 2017. Modeling Managed Grassland Biomass Estimation by
811 Using Multitemporal Remote Sensing Data—A Machine Learning Approach. *IEEE J Sel Top Appl*
812 *Earth Obs Remote Sens* 10, 3254–3264. <https://doi.org/10.1109/JSTARS.2016.2561618>
- 813 Amin, E., Verrelst, J., Rivera-Caicedo, J.P., Pipia, L., Ruiz-Verdú, A., Moreno, J., 2021. Prototyping
814 Sentinel-2 green LAI and brown LAI products for cropland monitoring. *Remote Sens Environ*
815 255, 112168. <https://doi.org/10.1016/j.rse.2020.112168>
- 816 Anderson, R.H., Fuhlendorf, S.D., Engle, D.M., 2006. Soil Nitrogen Availability in Tallgrass Prairie
817 Under the Fire–Grazing Interaction. *Rangel Ecol Manag* 59, 625–631.
818 <https://doi.org/10.2111/05-088R2.1>
- 819 Bailey, D.W., Gross, J.E., Laca, E.A., Rittenhouse, L.R., Coughenour, M.B., Swift, D.M., Sims, P.L., 1996.
820 Mechanisms That Result in Large Herbivore Grazing Distribution Patterns. *Journal of Range*
821 *Management* 49, 386. <https://doi.org/10.2307/4002919>
- 822 Bai, Y., Cotrufo, M.F., 2022. Grassland soil carbon sequestration: Current understanding, challenges,
823 and solutions. *Science (1979)* 377, 603–608. <https://doi.org/10.1126/science.abo2380>
- 824 Bardgett, R.D., Bullock, J.M., Lavorel, S., Manning, P., Schaffner, U., Ostle, N., Chomel, M., Durigan,
825 G., L. Fry, E., Johnson, D., Lavalley, J.M., Le Provost, G., Luo, S., Png, K., Sankaran, M., Hou, X.,
826 Zhou, H., Ma, L., Ren, W., Li, X., Ding, Y., Li, Y., Shi, H., 2021. Combatting global grassland
827 degradation. *Nat Rev Earth Environ* 2, 720–735. <https://doi.org/10.1038/s43017-021-00207-2>
- 828 Beck, H.E., Zimmermann, N.E., McVicar, T.R., Vergopolan, N., Berg, A., Wood, E.F., 2018. Present and
829 future Köppen-Geiger climate classification maps at 1-km resolution. *Sci Data* 5, 180214.
830 <https://doi.org/10.1038/sdata.2018.214>
- 831 Bengtsson, J., Bullock, J.M., Egoh, B., Everson, C., Everson, T., O’Connor, T., O’Farrell, P.J., Smith, H.G.,
832 Lindborg, R., 2019. Grasslands—more important for ecosystem services than you might think.
833 *Ecosphere* 10, e02582. <https://doi.org/10.1002/ecs2.2582>
- 834 Berger, K., Atzberger, C., Danner, M., D’Urso, G., Mauser, W., Vuolo, F., Hank, T., 2018. Evaluation of
835 the PROSAIL Model Capabilities for Future Hyperspectral Model Environments: A Review Study.
836 *Remote Sens (Basel)* 10, 85. <https://doi.org/10.3390/rs10010085>

- 837 Berger, K., Hank, T., Halabuk, A., Rivera-Caicedo, J.P., Woche, M., Mojses, M., Gerhátová, K.,
838 Tagliabue, G., Dolz, M.M., Venteo, A.B.P., Verrelst, J., 2021a. Assessing Non-Photosynthetic
839 Cropland Biomass from Spaceborne Hyperspectral Imagery. *Remote Sens (Basel)* 13, 4711.
840 <https://doi.org/10.3390/rs13224711>
- 841 Berger, K., Rivera Caicedo, J.P., Martino, L., Woche, M., Hank, T., Verrelst, J., 2021b. A Survey of
842 Active Learning for Quantifying Vegetation Traits from Terrestrial Earth Observation Data.
843 *Remote Sens (Basel)* 13, 287. <https://doi.org/10.3390/rs13020287>
- 844 Berger, K., Verrelst, J., Féret, J.-B., Hank, T., Woche, M., Mauser, W., Camps-Valls, G., 2020. Retrieval
845 of aboveground crop nitrogen content with a hybrid machine learning method. *International*
846 *Journal of Applied Earth Observation and Geoinformation* 92, 102174.
847 <https://doi.org/10.1016/j.jag.2020.102174>
- 848 Binh, N.A., Hauser, L.T., Viet Hoa, P., Thi Phuong Thao, G., An, N.N., Nhut, H.S., Phuong, T.A., Verrelst,
849 J., 2022. Quantifying mangrove leaf area index from Sentinel-2 imagery using hybrid models and
850 active learning. *Int J Remote Sens* 43, 5636–5657.
851 <https://doi.org/10.1080/01431161.2021.2024912>
- 852 Blüthgen, N., Dormann, C.F., Prati, D., Klaus, V.H., Kleinebecker, T., Hölzel, N., Alt, F., Boch, S., Gockel,
853 S., Hemp, A., Müller, J., Nieschulze, J., Renner, S.C., Schöning, I., Schumacher, U., Socher, S.A.,
854 Wells, K., Birkhofer, K., Buscot, F., Oelmann, Y., Rothenwöhrer, C., Scherber, C., Tschardt, T.,
855 Weiner, C.N., Fischer, M., Kalko, E.K.V., Linsenmair, K.E., Schulze, E.-D., Weisser, W.W., 2012. A
856 quantitative index of land-use intensity in grasslands: Integrating mowing, grazing and
857 fertilization. *Basic Appl Ecol* 13, 207–220. <https://doi.org/10.1016/j.baae.2012.04.001>
- 858 Bolliger, R., Prati, D., Fischer, M., 2020. Vegetation Records for Grassland EPs, 2008 - 2020. Version 2.
859 Biodiversity Exploratories Information System. Dataset. [https://www.bexis.uni-](https://www.bexis.uni-jena.de/ddm/data/Showdata/27386)
860 [jena.de/ddm/data/Showdata/27386](https://www.bexis.uni-jena.de/ddm/data/Showdata/27386). Dataset ID = 27386.
- 861 Breiman, L., 2001. Random forests. *Mach Learn* 45, 5–32.
862 <https://doi.org/https://doi.org/10.1023/A:1010933404324>
- 863 Brown, L.A., Ogotu, B.O., Dash, J., 2019. Estimating Forest Leaf Area Index and Canopy Chlorophyll
864 Content with Sentinel-2: An Evaluation of Two Hybrid Retrieval Algorithms. *Remote Sens (Basel)*
865 11, 1752. <https://doi.org/10.3390/rs11151752>
- 866 Busch, V., Klaus, V.H., Penone, C., Schäfer, D., Boch, S., Prati, D., Müller, J., Socher, S.A., Niinemets,
867 Ü., Peñuelas, J., Hölzel, N., Fischer, M., Kleinebecker, T., 2018. Nutrient stoichiometry and land
868 use rather than species richness determine plant functional diversity. *Ecol Evol* 8, 601–616.
869 <https://doi.org/10.1002/ece3.3609>
- 870 Caballero, G., Pezzola, A., Winschel, C., Sanchez Angonova, P., Casella, A., Orden, L., Salinero-
871 Delgado, M., Reyes-Muñoz, P., Berger, K., Delegido, J., Verrelst, J., 2023. Synergy of Sentinel-1
872 and Sentinel-2 Time Series for Cloud-Free Vegetation Water Content Mapping with Multi-
873 Output Gaussian Processes. *Remote Sens (Basel)* 15, 1822. <https://doi.org/10.3390/rs15071822>
- 874 Campos-Taberner, M., Moreno-Martínez, Á., García-Haro, F., Camps-Valls, G., Robinson, N., Kattge,
875 J., Running, S., 2018. Global Estimation of Biophysical Variables from Google Earth Engine
876 Platform. *Remote Sens (Basel)* 10, 1167. <https://doi.org/10.3390/rs10081167>
- 877 Cavender-Bares, J., Schneider, F.D., Santos, M.J., Armstrong, A., Carnaval, A., Dahlin, K.M., Fatoyinbo,
878 L., Hurtt, G.C., Schimel, D., Townsend, P.A., Ustin, S.L., Wang, Z., Wilson, A.M., 2022. Integrating

879 remote sensing with ecology and evolution to advance biodiversity conservation. *Nat Ecol Evol*
880 6, 506–519. <https://doi.org/10.1038/s41559-022-01702-5>

881 Chen, T., Guestrin, C., 2016. XGBoost: A scalable tree boosting system. *Proceedings of the ACM*
882 *SIGKDD International Conference on Knowledge Discovery and Data Mining 13-17-August-2016*,
883 785–794. <https://doi.org/10.1145/2939672.2939785>

884 Combal, B., Baret, F., Weiss, M., Trubuil, A., Macé, D., Pragnère, A., Myneni, R., Knyazikhin, Y., Wang,
885 L., 2003. Retrieval of canopy biophysical variables from bidirectional reflectance. *Remote Sens*
886 *Environ* 84, 1–15. [https://doi.org/10.1016/S0034-4257\(02\)00035-4](https://doi.org/10.1016/S0034-4257(02)00035-4)

887 Darvishzadeh, R., Atzberger, C., Skidmore, A., Schlerf, M., 2011. Mapping grassland leaf area index
888 with airborne hyperspectral imagery: A comparison study of statistical approaches and
889 inversion of radiative transfer models. *ISPRS Journal of Photogrammetry and Remote Sensing*
890 66, 894–906. <https://doi.org/10.1016/J.ISPRSJPRS.2011.09.013>

891 Dehghan-Shoar, M.H., Orsi, A.A., Pullanagari, R.R., Yule, I.J., 2023. A hybrid model to predict nitrogen
892 concentration in heterogeneous grassland using field spectroscopy. *Remote Sens Environ* 285,
893 113385. <https://doi.org/10.1016/j.rse.2022.113385>

894 Delegido, J., Verrelst, J., Rivera, J.P., Ruiz-Verdú, A., Moreno, J., 2015. Brown and green LAI mapping
895 through spectral indices. *International Journal of Applied Earth Observation and*
896 *Geoinformation* 35, 350–358. <https://doi.org/10.1016/j.jag.2014.10.001>

897 de Sá, N.C., Baratchi, M., Hauser, L.T., van Bodegom, P., 2021. Exploring the Impact of Noise on
898 Hybrid Inversion of PROSAIL RTM on Sentinel-2 Data. *Remote Sens (Basel)* 13, 648.
899 <https://doi.org/10.3390/rs13040648>

900 De Vroey, M., de Vendictis, L., Zavagli, M., Bontemps, S., Heymans, D., Radoux, J., Koetz, B.,
901 Defourny, P., 2022. Mowing detection using Sentinel-1 and Sentinel-2 time series for large scale
902 grassland monitoring. *Remote Sens Environ* 280, 113145.
903 <https://doi.org/10.1016/j.rse.2022.113145>

904 Dixon, A.P., Faber-Langendoen, D., Josse, C., Morrison, J., Loucks, C.J., 2014. Distribution mapping of
905 world grassland types. *J Biogeogr* 41, 2003–2019. <https://doi.org/10.1111/jbi.12381>

906 Erb, K.-H., Kastner, T., Plutzer, C., Bais, A.L.S., Carvalhais, N., Fetzel, T., Gingrich, S., Haberl, H., Lauk,
907 C., Niedertscheider, M., Pongratz, J., Thurner, M., Luysaert, S., 2018. Unexpectedly large
908 impact of forest management and grazing on global vegetation biomass. *Nature* 553, 73–76.
909 <https://doi.org/10.1038/nature25138>

910 Fang, H., Liang, S., Kuusk, A., 2003. Retrieving leaf area index using a genetic algorithm with a canopy
911 radiative transfer model. *Remote Sens Environ* 85, 257–270. [https://doi.org/10.1016/S0034-4257\(03\)00005-1](https://doi.org/10.1016/S0034-4257(03)00005-1)

913 Féret, J.-B., Berger, K., de Boissieu, F., Malenovsky, Z., 2021. PROSPECT-PRO for estimating content of
914 nitrogen-containing leaf proteins and other carbon-based constituents. *Remote Sens Environ*
915 252, 112173. <https://doi.org/10.1016/j.rse.2020.112173>

916 Féret, J.-B., François, C., Asner, G.P., Gitelson, A.A., Martin, R.E., Bidet, L.P.R., Ustin, S.L., le Maire, G.,
917 Jacquemoud, S., 2008. PROSPECT-4 and 5: Advances in the leaf optical properties model
918 separating photosynthetic pigments. *Remote Sens Environ* 112, 3030–3043.
919 <https://doi.org/10.1016/j.rse.2008.02.012>

- 920 Féret, J.-B., Gitelson, A.A., Noble, S.D., Jacquemoud, S., 2017. PROSPECT-D: Towards modeling leaf
921 optical properties through a complete lifecycle. *Remote Sens Environ* 193, 204–215.
922 <https://doi.org/10.1016/j.rse.2017.03.004>
- 923 Féret, J.-B., le Maire, G., Jay, S., Berveiller, D., Bendoula, R., Hmimina, G., Cheraiet, A., Oliveira, J.C.,
924 Ponzoni, F.J., Solanki, T., de Boissieu, F., Chave, J., Nouvellon, Y., Porcar-Castell, A., Proisy, C.,
925 Soudani, K., Gastellu-Etchegorry, J.-P., Lefèvre-Fonollosa, M.-J., 2019. Estimating leaf mass per
926 area and equivalent water thickness based on leaf optical properties: Potential and limitations
927 of physical modeling and machine learning. *Remote Sens Environ* 231, 110959.
928 <https://doi.org/10.1016/j.rse.2018.11.002>
- 929 Féret, J., de Boissieu, F., 2022. prosail: PROSAIL leaf and canopy radiative transfer model and
930 inversion routines. R package version 1.1.1 [WWW Document]. URL
931 <https://gitlab.com/jbferet/prosail>
- 932 Fischer, M., Bossdorf, O., Gockel, S., Hänsel, F., Hemp, A., Hessenmöller, D., Korte, G., Nieschulze, J.,
933 Pfeiffer, S., Prati, D., Renner, S., Schöning, I., Schumacher, U., Wells, K., Buscot, F., Kalko, E.K.V.,
934 Linsenmair, K.E., Schulze, E.-D., Weisser, W.W., 2010. Implementing large-scale and long-term
935 functional biodiversity research: The Biodiversity Exploratories. *Basic Appl Ecol* 11, 473–485.
936 <https://doi.org/10.1016/j.baae.2010.07.009>
- 937 Friedman, J.H., 2001. Greedy function approximation: A gradient boosting machine. *Ann Stat* 29,
938 1189–1232. <https://doi.org/10.1214/AOS/1013203451>
- 939 Fuhlendorf, S.D., Engle, D.M., 2004. Application of the fire-grazing interaction to restore a shifting
940 mosaic on tallgrass prairie. *Journal of Applied Ecology* 41, 604–614.
941 <https://doi.org/10.1111/j.0021-8901.2004.00937.x>
- 942 Fuhlendorf, S.D., Engle, D.M., 2001. Restoring Heterogeneity on Rangelands: Ecosystem Management
943 Based on Evolutionary Grazing Patterns. *Bioscience* 51, 625–632.
944 [https://doi.org/https://doi.org/10.1641/0006-3568\(2001\)051\[0625:RHOREM\]2.0.CO;2](https://doi.org/https://doi.org/10.1641/0006-3568(2001)051[0625:RHOREM]2.0.CO;2)
- 945 Gascon, F., Bouzinac, C., Thépaut, O., Jung, M., Francesconi, B., Louis, J., Lonjou, V., Lafrance, B.,
946 Massera, S., Gaudel-Vacaresse, A., Languille, F., Alhammoud, B., Viallefont, F., Pflug, B., Bieniarz,
947 J., Clerc, S., Pessiot, L., Trémas, T., Cadau, E., De Bonis, R., Isola, C., Martimort, P., Fernandez, V.,
948 2017. Copernicus Sentinel-2A Calibration and Products Validation Status. *Remote Sens (Basel)* 9,
949 584. <https://doi.org/10.3390/rs9060584>
- 950 Gholizadeh, H., Friedman, M.S., McMillan, N.A., Hammond, W.M., Hassani, K., Sams, A. V., Charles,
951 M.D., Garrett, D.R., Joshi, O., Hamilton, R.G., Fuhlendorf, S.D., Trowbridge, A.M., Adams, H.D.,
952 2022. Mapping invasive alien species in grassland ecosystems using airborne imaging
953 spectroscopy and remotely observable vegetation functional traits. *Remote Sens Environ* 271,
954 112887. <https://doi.org/10.1016/j.rse.2022.112887>
- 955 Gholizadeh, H., Rakotoarivony, M.N.A., Hassani, K., Johnson, K.G., Hamilton, R.G., Fuhlendorf, S.D.,
956 Schneider, F.D., Bachelot, B., 2024. Advancing our understanding of plant diversity-biological
957 invasion relationships using imaging spectroscopy. *Remote Sens Environ* 304, 114028.
958 <https://doi.org/10.1016/j.rse.2024.114028>
- 959 Gorelick, N., Hancher, M., Dixon, M., Ilyushchenko, S., Thau, D., Moore, R., 2017. Google Earth
960 Engine: Planetary-scale geospatial analysis for everyone. *Remote Sens Environ* 202, 18–27.
961 <https://doi.org/10.1016/j.rse.2017.06.031>

- 962 Guerini Filho, M., Kuplich, T.M., Quadros, F.L.F. De, 2020. Estimating natural grassland biomass by
 963 vegetation indices using Sentinel 2 remote sensing data. *Int J Remote Sens* 41, 2861–2876.
 964 <https://doi.org/10.1080/01431161.2019.1697004>
- 965 Habel, J.C., Dengler, J., Janišová, M., Török, P., Wellstein, C., Wiezik, M., 2013. European grassland
 966 ecosystems: threatened hotspots of biodiversity. *Biodivers Conserv* 22, 2131–2138.
 967 <https://doi.org/10.1007/s10531-013-0537-x>
- 968 Hamilton, R.G., 2007. Restoring heterogeneity on the Tallgrass Prairie Preserve: applying the fire-
 969 grazing interaction model, in: Masters, R.E., Galley, K.E.M. (Eds.), *Proceedings of the 23rd Tall*
 970 *Timbers Fire Ecology Conference: Fire in Grassland and Shrubland Ecosystems*. Tall Timbers
 971 Research Station, Tallahassee, FL, pp. 163–169.
- 972 Hauser, L.T., Féret, J.-B., An Binh, N., van der Windt, N., Sil, Â.F., Timmermans, J., Soudzilovskaia,
 973 N.A., van Bodegom, P.M., 2021a. Towards scalable estimation of plant functional diversity from
 974 Sentinel-2: In-situ validation in a heterogeneous (semi-)natural landscape. *Remote Sens Environ*
 975 262, 112505. <https://doi.org/10.1016/j.rse.2021.112505>
- 976 Hauser, L.T., Timmermans, J., van der Windt, N., Sil, Â.F., César de Sá, N., Soudzilovskaia, N.A., van
 977 Bodegom, P.M., 2021b. Explaining discrepancies between spectral and in-situ plant diversity in
 978 multispectral satellite earth observation. *Remote Sens Environ* 265, 112684.
 979 <https://doi.org/10.1016/j.rse.2021.112684>
- 980 Helfenstein, I.S., Schneider, F.D., Schaepman, M.E., Morsdorf, F., 2022. Assessing biodiversity from
 981 space: Impact of spatial and spectral resolution on trait-based functional diversity. *Remote Sens*
 982 *Environ* 275, 113024. <https://doi.org/10.1016/j.rse.2022.113024>
- 983 He, L., Li, A., Yin, G., Nan, X., Bian, J., 2019. Retrieval of Grassland Aboveground Biomass through
 984 Inversion of the PROSAIL Model with MODIS Imagery. *Remote Sensing* 2019, Vol. 11, Page 1597
 985 11, 1597. <https://doi.org/10.3390/RS11131597>
- 986 Hinderling, J., Penone, C., Fischer, M., Prati, D., Hoelzel, N., 2023. Biomass data for grassland EPs,
 987 2009 - 2022. Version 8. Biodiversity Exploratories Information System. Dataset.
 988 <https://www.bexis.uni-jena.de/ddm/data/Showdata/31448>. Dataset ID = 31448.
- 989 Hobohm, C., Bruchmann, I., 2009. Endemische Gefäßpflanzen und ihre Habitate in Europa - Plädoyer
 990 für den Schutz der Grasland-Ökosysteme. *Berichte der Reinhold-Tüxen-Gesellschaft* 21, 142–
 991 161.
- 992 Huete, A., Didan, K., Miura, T., Rodriguez, E.P., Gao, X., Ferreira, L.G., 2002. Overview of the
 993 radiometric and biophysical performance of the MODIS vegetation indices. *Remote Sens*
 994 *Environ* 83, 195–213. [https://doi.org/10.1016/S0034-4257\(02\)00096-2](https://doi.org/10.1016/S0034-4257(02)00096-2)
- 995 Jacquemoud, S., Baret, F., 1990. PROSPECT: A model of leaf optical properties spectra. *Remote Sens*
 996 *Environ* 34, 75–91. [https://doi.org/10.1016/0034-4257\(90\)90100-Z](https://doi.org/10.1016/0034-4257(90)90100-Z)
- 997 Jacquemoud, S., Verhoef, W., Baret, F., Bacour, C., Zarco-Tejada, P.J., Asner, G.P., François, C., Ustin,
 998 S.L., 2009. PROSPECT + SAIL models: A review of use for vegetation characterization. *Remote*
 999 *Sens Environ* 113, S56–S66. <https://doi.org/10.1016/J.RSE.2008.01.026>
- 1000 Jehle, M., Hueni, A., Damm, A., D’Odorico, P., Weyermann, J., Kneubihler, M., Schlapfer, D.,
 1001 Schaepman, M.E., Meuleman, K., 2010. APEX - current status, performance and validation
 1002 concept, in: *2010 IEEE Sensors*. IEEE, pp. 533–537.
 1003 <https://doi.org/10.1109/ICSENS.2010.5690122>

- 1004 Joswig, J.S., Wirth, C., Schuman, M.C., Kattge, J., Reu, B., Wright, I.J., Sippel, S.D., Rüger, N., Richter,
1005 R., Schaepman, M.E., van Bodegom, P.M., Cornelissen, J.H.C., Díaz, S., Hattingh, W.N., Kramer,
1006 K., Lens, F., Niinemets, Ü., Reich, P.B., Reichstein, M., Römermann, C., Schrodtt, F., Anand, M.,
1007 Bahn, M., Byun, C., Campetella, G., Cerabolini, B.E.L., Craine, J.M., Gonzalez-Melo, A., Gutiérrez,
1008 A.G., He, T., Higuchi, P., Jactel, H., Kraft, N.J.B., Minden, V., Onipchenko, V., Peñuelas, J., Pillar,
1009 V.D., Sosinski, Ê., Soudzilovskaia, N.A., Weiher, E., Mahecha, M.D., 2021. Climatic and soil
1010 factors explain the two-dimensional spectrum of global plant trait variation. *Nat Ecol Evol* 6, 36–
1011 50. <https://doi.org/10.1038/s41559-021-01616-8>
- 1012 Kothari, S., Schweiger, A.K., 2022. Plant spectra as integrative measures of plant phenotypes. *Journal*
1013 *of Ecology* 110, 2536–2554. <https://doi.org/10.1111/1365-2745.13972>
- 1014 Kovács, D.D., Reyes-Muñoz, P., Salinero-Delgado, M., Mészáros, V.I., Berger, K., Verrelst, J., 2023.
1015 Cloud-Free Global Maps of Essential Vegetation Traits Processed from the TOA Sentinel-3
1016 Catalogue in Google Earth Engine. *Remote Sens (Basel)* 15, 3404.
1017 <https://doi.org/10.3390/rs15133404>
- 1018 Lawrence, D.M., Fisher, R.A., Koven, C.D., Oleson, K.W., Swenson, S.C., Bonan, G., Collier, N., Ghimire,
1019 B., van Kampenhout, L., Kennedy, D., Kluzek, E., Lawrence, P.J., Li, F., Li, H., Lombardozzi, D.,
1020 Riley, W.J., Sacks, W.J., Shi, M., Vertenstein, M., Wieder, W.R., Xu, C., Ali, A.A., Badger, A.M.,
1021 Bisht, G., van den Broeke, M., Brunke, M.A., Burns, S.P., Buzan, J., Clark, M., Craig, A., Dahlin, K.,
1022 Drewniak, B., Fisher, J.B., Flanner, M., Fox, A.M., Gentine, P., Hoffman, F., Keppel-Aleks, G.,
1023 Knox, R., Kumar, S., Lenaerts, J., Leung, L.R., Lipscomb, W.H., Lu, Y., Pandey, A., Pelletier, J.D.,
1024 Perket, J., Randerson, J.T., Ricciuto, D.M., Sanderson, B.M., Slater, A., Subin, Z.M., Tang, J.,
1025 Thomas, R.Q., Val Martin, M., Zeng, X., 2019. The Community Land Model Version 5:
1026 Description of New Features, Benchmarking, and Impact of Forcing Uncertainty. *J Adv Model*
1027 *Earth Syst* 11, 4245–4287. <https://doi.org/10.1029/2018MS001583>
- 1028 Lemaire, G., Hodgson, J., Chabbi, A. (Eds.), 2011. *Grassland productivity and ecosystem services*. CAB
1029 International, Wallingford.
- 1030 Li, C., Zhou, L., Xu, W., 2021. Estimating Aboveground Biomass Using Sentinel-2 MSI Data and
1031 Ensemble Algorithms for Grassland in the Shengjin Lake Wetland, China. *Remote Sensing* 2021,
1032 Vol. 13, Page 1595 13, 1595. <https://doi.org/10.3390/RS13081595>
- 1033 Li, F., Zeng, Y., Luo, J., Ma, R., Wu, B., 2016. Modeling grassland aboveground biomass using a pure
1034 vegetation index. *Ecol Indic* 62, 279–288. <https://doi.org/10.1016/j.ecolind.2015.11.005>
- 1035 Limb, R.F., Hickman, K.R., Engle, D.M., Norland, J.E., Fuhlendorf, S.D., 2007. Digital Photography:
1036 Reduced Investigator Variation in Visual Obstruction Measurements for Southern Tallgrass
1037 Prairie. *Rangel Ecol Manag* 60, 548–552. [https://doi.org/https://doi.org/10.2111/1551-5028\(2007\)60\[548:DPRIVI\]2.0.CO;2](https://doi.org/https://doi.org/10.2111/1551-5028(2007)60[548:DPRIVI]2.0.CO;2)
- 1039 Liu, A., Cheng, X., Chen, Z., 2021. Performance evaluation of GEDI and ICESat-2 laser altimeter data
1040 for terrain and canopy height retrievals. *Remote Sens Environ* 264, 112571.
1041 <https://doi.org/10.1016/j.rse.2021.112571>
- 1042 Liu, H., Cai, J., Ong, Y.-S., 2018. Remarks on multi-output Gaussian process regression. *Knowl Based*
1043 *Syst* 144, 102–121. <https://doi.org/10.1016/j.knosys.2017.12.034>
- 1044 Locherer, M., Hank, T., Danner, M., Mauser, W., 2015. Retrieval of Seasonal Leaf Area Index from
1045 Simulated EnMAP Data through Optimized LUT-Based Inversion of the PROSAIL Model. *Remote*
1046 *Sens (Basel)* 7, 10321–10346. <https://doi.org/10.3390/rs70810321>

- 1047 Martínez-Ferrer, L., Moreno-Martínez, Á., Campos-Taberner, M., García-Haro, F.J., Muñoz-Marí, J.,
 1048 Running, S.W., Kimball, J., Clinton, N., Camps-Valls, G., 2022. Quantifying uncertainty in high
 1049 resolution biophysical variable retrieval with machine learning. *Remote Sens Environ* 280,
 1050 113199. <https://doi.org/10.1016/J.RSE.2022.113199>
- 1051 Masek, J.G., Wulder, M.A., Markham, B., McCorkel, J., Crawford, C.J., Storey, J., Jenstrom, D.T., 2020.
 1052 Landsat 9: Empowering open science and applications through continuity. *Remote Sens Environ*
 1053 248, 111968. <https://doi.org/10.1016/j.rse.2020.111968>
- 1054 Mesonet, 2024. Monthly Rainfall Table: Norman (NRMN) Monthly Actual Oklahoma Mesonet Rainfall
 1055 Table (in inches) [WWW Document]. URL [https://www.mesonet.org/weather/rainfall/monthly-](https://www.mesonet.org/weather/rainfall/monthly-rainfall-table?ref=1210)
 1056 [rainfall-table?ref=1210](https://www.mesonet.org/weather/rainfall/monthly-rainfall-table?ref=1210) (accessed 1.4.24).
- 1057 Meyer, H., Pebesma, E., 2020. Predicting into unknown space? Estimating the area of applicability of
 1058 spatial prediction models. *Methods Ecol Evol* 12, 1620–1633. [https://doi.org/10.1111/2041-](https://doi.org/10.1111/2041-210X.13650)
 1059 [210X.13650](https://doi.org/10.1111/2041-210X.13650)
- 1060 Moreno-Martinez, A., Camps-Valls, G., Kattge, J., Robinson, N., Reichstein, M., van Bodegom, P.,
 1061 Kramer, K., Cornelissen, J.H.C., Reich, P., Bahn, M., Niinemets, Ü, Ulo, Peñuelas, J., Craine, J.,
 1062 Cerabolini, B.E.L., Minden, V., Laughlin, D.C., Sack, L., Allred, B., Baraloto, C., Byun, C.,
 1063 Soudzilovskaia, N.A., Running, S.W., 2020. A Methodology to Derive Global Maps of Leaf Traits
 1064 Using Remote Sensing and Climate Data. *Remote Sens Environ* 218, 69–88.
 1065 <https://doi.org/10.1016/j.rse.2018.09.006>
- 1066 Müller-Wilm, U., Louis, J., Richter, R., Gascon, F., Niezette, M., 2013. Sentinel-2 level 2A prototype
 1067 processor: architecture, algorithms and first results, in: *Proceedings of the ESA Living Planet*
 1068 *Symposium*. Edinburgh, UK, pp. 9–13.
- 1069 Muñoz Sabater, J., 2019. ERA5-Land hourly data from 1950 to present. Copernicus Climate Change
 1070 Service (C3S) Climate Data Store. <https://doi.org/10.24381/cds.e2161bac>
- 1071 Muro, J., Linstädter, A., Magdon, P., Wöllauer, S., Männer, F.A., Schwarz, L.-M., Ghazaryan, G.,
 1072 Schultz, J., Malenovský, Z., Dubovyk, O., 2022. Predicting plant biomass and species richness in
 1073 temperate grasslands across regions, time, and land management with remote sensing and
 1074 deep learning. *Remote Sens Environ* 282, 113262. <https://doi.org/10.1016/j.rse.2022.113262>
- 1075 O'Mara, F.P., 2012. The role of grasslands in food security and climate change. *Ann Bot* 110, 1263–
 1076 1270. <https://doi.org/10.1093/aob/mcs209>
- 1077 Ostrowski, A., Nieschulze, J., Schulze, E.D., König-Ries, B., 2020. Coordinates and Inventory Overview
 1078 of all Grid Plots (GPs). Version 6. Biodiversity Exploratories Information System. Dataset.
 1079 <https://www.bexis.uni-jena.de>. Dataset ID = 20907.
- 1080 Pedregosa, F., Varoquaux, G., Gramfort, A., Michel, V., Thirion, B., Grisel, O., Blondel, M., Müller, A.,
 1081 Nothman, J., Louppe, G., Prettenhofer, P., Weiss, R., Dubourg, V., Vanderplas, J., Passos, A.,
 1082 Cournapeau, D., Brucher, M., Perrot, M., Duchesnay, É., 2012. Scikit-learn: Machine Learning in
 1083 Python. *Journal of Machine Learning Research* 12, 2825–2830.
- 1084 Pipia, L., Muñoz-Marí, J., Amin, E., Belda, S., Camps-Valls, G., Verrelst, J., 2019. Fusing optical and SAR
 1085 time series for LAI gap filling with multioutput Gaussian processes. *Remote Sens Environ* 235,
 1086 111452. <https://doi.org/10.1016/j.rse.2019.111452>
- 1087 Poortinga, A., Tenneson, K., Shapiro, A., Nquyen, Q., San Aung, K., Chishtie, F., Saah, D., 2019.
 1088 Mapping Plantations in Myanmar by Fusing Landsat-8, Sentinel-2 and Sentinel-1 Data along

- 1089 with Systematic Error Quantification. *Remote Sens (Basel)* 11, 831.
 1090 <https://doi.org/10.3390/rs11070831>
- 1091 Punalekar, S.M., Verhoef, A., Quaife, T.L., Humphries, D., Bermingham, L., Reynolds, C.K., 2018.
 1092 Application of Sentinel-2A data for pasture biomass monitoring using a physically based
 1093 radiative transfer model. *Remote Sens Environ* 218, 207–220.
 1094 <https://doi.org/10.1016/j.rse.2018.09.028>
- 1095 Quan, X., He, B., Yebra, M., Yin, C., Liao, Z., Zhang, X., Li, X., 2017. A radiative transfer model-based
 1096 method for the estimation of grassland aboveground biomass. *International Journal of Applied*
 1097 *Earth Observation and Geoinformation* 54, 159–168.
 1098 <https://doi.org/10.1016/J.JAG.2016.10.002>
- 1099 Raab, C., Riesch, F., Tonn, B., Barrett, B., Meißner, M., Balkenhol, N., Isselstein, J., 2020. Target-
 1100 oriented habitat and wildlife management: estimating forage quantity and quality of semi-
 1101 natural grasslands with Sentinel-1 and Sentinel-2 data. *Remote Sens Ecol Conserv* 6, 381–398.
 1102 <https://doi.org/10.1002/rse2.149>
- 1103 Ranghetti, M., Boschetti, M., Ranghetti, L., Tagliabue, G., Panigada, C., Gianinetto, M., Verrelst, J.,
 1104 Candiani, G., 2022. Assessment of maize nitrogen uptake from PRISMA hyperspectral data
 1105 through hybrid modelling. *Eur J Remote Sens.* <https://doi.org/10.1080/22797254.2022.2117650>
- 1106 Rasmussen, C.E., Williams, C.K.I., 2006. *Gaussian Processes for Machine Learning*. MIT Press,
 1107 Cambridge, MA.
- 1108 R Core Team, 2021. *R: A Language and Environment for Statistical Computing*. Vienna.
- 1109 Reiner mann, S., Asam, S., Kuenzer, C., 2020. Remote Sensing of Grassland Production and
 1110 Management—A Review. *Remote Sens (Basel)* 12, 1949. <https://doi.org/10.3390/rs12121949>
- 1111 Rempfler, T., Rossi, C., Schweizer, J., Peters, W., Signer, C., Filli, F., Jenny, H., Hackländer, K.,
 1112 Buchmann, S., Anderwald, P., 2024. Remote sensing reveals the role of forage quality and
 1113 quantity for summer habitat use in red deer. *Mov Ecol* 12, 80. [https://doi.org/10.1186/s40462-](https://doi.org/10.1186/s40462-024-00521-6)
 1114 [024-00521-6](https://doi.org/10.1186/s40462-024-00521-6)
- 1115 Reuter, H.I., Nelson, A., Jarvis, A., 2007. An evaluation of void-filling interpolation methods for SRTM
 1116 data. *International Journal of Geographical Information Science* 21, 983–1008.
 1117 <https://doi.org/10.1080/13658810601169899>
- 1118 Richter, K., Atzberger, C., Hank, T.B., Mauser, W., 2012. Derivation of biophysical variables from Earth
 1119 observation data: validation and statistical measures. *J Appl Remote Sens* 6, 063557–1.
 1120 <https://doi.org/10.1117/1.JRS.6.063557>
- 1121 Richter, R., Schläpfer, D., 2002. Geo-atmospheric processing of airborne imaging spectrometry data.
 1122 Part 2: Atmospheric/topographic correction. *Int J Remote Sens* 23, 2631–2649.
 1123 <https://doi.org/10.1080/01431160110115834>
- 1124 Rivera, J., Verrelst, J., Gómez-Dans, J., Muñoz-Marí, J., Moreno, J., Camps-Valls, G., 2015. An Emulator
 1125 Toolbox to Approximate Radiative Transfer Models with Statistical Learning. *Remote Sens*
 1126 *(Basel)* 7, 9347–9370. <https://doi.org/10.3390/rs70709347>
- 1127 Rivera, J., Verrelst, J., Leonenko, G., Moreno, J., 2013. Multiple Cost Functions and Regularization
 1128 Options for Improved Retrieval of Leaf Chlorophyll Content and LAI through Inversion of the
 1129 PROSAIL Model. *Remote Sens (Basel)* 5, 3280–3304. <https://doi.org/10.3390/rs5073280>

- 1130 Rossi, C., Kneubühler, M., Schütz, M., Schaepman, M.E., Haller, R.M., Risch, A.C., 2020. From local to
 1131 regional: Functional diversity in differently managed alpine grasslands. *Remote Sens Environ*
 1132 236, 111415. <https://doi.org/10.1016/j.rse.2019.111415>
- 1133 Rossi, C., McMillan, N.A., Schweizer, J.M., Gholizadeh, H., Groen, M., Ioannidis, N., Hauser, L.T., 2024.
 1134 Parcel level temporal variance of remotely sensed spectral reflectance predicts plant diversity.
 1135 *Environmental Research Letters* 19, 074023. <https://doi.org/10.1088/1748-9326/ad545a>
- 1136 Roudier, P., 2021. Introduction to conditioned Latin hypercube sampling with the clhs package
 1137 [WWW Document]. URL <https://CRAN.R-project.org/package=clhs>
- 1138 Roy, D.P., Wulder, M.A., Loveland, T.R., C.E., W., Allen, R.G., Anderson, M.C., Helder, D., Irons, J.R.,
 1139 Johnson, D.M., Kennedy, R., Scambos, T.A., Schaaf, C.B., Schott, J.R., Sheng, Y., Vermote, E.F.,
 1140 Belward, A.S., Bindschadler, R., Cohen, W.B., Gao, F., Hipple, J.D., Hostert, P., Huntington, J.,
 1141 Justice, C.O., Kilic, A., Kovalskyy, V., Lee, Z.P., Lymburner, L., Masek, J.G., McCorkel, J., Shuai, Y.,
 1142 Trezza, R., Vogelmann, J., Wynne, R.H., Zhu, Z., 2014. Landsat-8: Science and product vision for
 1143 terrestrial global change research. *Remote Sens Environ* 145, 154–172.
 1144 <https://doi.org/10.1016/j.rse.2014.02.001>
- 1145 Safanelli, J.L., Hengl, T., Parente, L.L., Minarik, R., Bloom, D.E., Todd-Brown, K., Gholizadeh, A.,
 1146 Mendes, W. de S., Sanderman, J., 2025. Open Soil Spectral Library (OSSL): Building reproducible
 1147 soil calibration models through open development and community engagement. *PLoS One* 20,
 1148 e0296545. <https://doi.org/10.1371/journal.pone.0296545>
- 1149 Schaepman, M.E., Jehle, M., Hueni, A., D’Odorico, P., Damm, A., Weyermann, J., Schneider, F.D.,
 1150 Laurent, V., Popp, C., Seidel, F.C., Lenhard, K., Gege, P., Küchler, C., Brazile, J., Kohler, P., De Vos,
 1151 L., Meuleman, K., Meynart, R., Schläpfer, D., Kneubühler, M., Itten, K.I., 2015. Advanced
 1152 radiometry measurements and Earth science applications with the Airborne Prism Experiment
 1153 (APEX). *Remote Sens Environ* 158, 207–219. <https://doi.org/10.1016/j.rse.2014.11.014>
- 1154 Schiefer, F., Schmidlein, S., Kattenborn, T., 2021. The retrieval of plant functional traits from canopy
 1155 spectra through RTM-inversions and statistical models are both critically affected by plant
 1156 phenology. *Ecol Indic* 121, 107062. <https://doi.org/10.1016/j.ecolind.2020.107062>
- 1157 Schläpfer, D., Richter, R., 2002. Geo-atmospheric processing of airborne imaging spectrometry data.
 1158 Part 1: Parametric orthorectification. *Int J Remote Sens* 23, 2609–2630.
 1159 <https://doi.org/10.1080/01431160110115825>
- 1160 Schweiger, A.K., 2020. Spectral Field Campaigns: Planning and Data Collection, in: *Remote Sensing of*
 1161 *Plant Biodiversity*. Springer International Publishing, Cham, pp. 385–423.
 1162 https://doi.org/10.1007/978-3-030-33157-3_15
- 1163 Schweiger, A.K., Risch, A.C., Damm, A., Kneubühler, M., Haller, R., Schaepman, M.E., Schütz, M.,
 1164 2015a. Using imaging spectroscopy to predict above-ground plant biomass in alpine grasslands
 1165 grazed by large ungulates. *Journal of Vegetation Science* 26, 175–190.
 1166 <https://doi.org/10.1111/jvs.12214>
- 1167 Schweiger, A.K., Schütz, M., Anderwald, P., Schaepman, M.E., Kneubühler, M., Haller, R., Risch, A.C.,
 1168 2015b. Foraging ecology of three sympatric ungulate species – Behavioural and resource maps
 1169 indicate differences between chamois, ibex and red deer. *Mov Ecol* 3, 6.
 1170 <https://doi.org/10.1186/s40462-015-0033-x>

- 1171 Schweiger, A.K., Schütz, M., Risch, A.C., Kneubühler, M., Haller, R., Schaepman, M.E., 2017. How to
 1172 predict plant functional types using imaging spectroscopy: linking vegetation community traits,
 1173 plant functional types and spectral response. *Methods Ecol Evol* 8, 86–95.
 1174 <https://doi.org/10.1111/2041-210X.12642>
- 1175 Schwieder, M., Buddeberg, M., Kowalski, K., Pfoch, K., Bartsch, J., Bach, H., Pickert, J., Hostert, P.,
 1176 2020. Estimating Grassland Parameters from Sentinel-2: A Model Comparison Study. *PFG –*
 1177 *Journal of Photogrammetry, Remote Sensing and Geoinformation Science* 88, 379–390.
 1178 <https://doi.org/10.1007/s41064-020-00120-1>
- 1179 scikit-learn Developers, 2023. ConstantKernel [WWW Document]. URL [https://scikit-](https://scikit-learn.org/stable/modules/generated/sklearn.gaussian_process.kernels.ConstantKernel.html)
 1180 [learn.org/stable/modules/generated/sklearn.gaussian_process.kernels.ConstantKernel.html](https://scikit-learn.org/stable/modules/generated/sklearn.gaussian_process.kernels.ConstantKernel.html)
 1181 (accessed 10.12.23).
- 1182 Sherrill, C.W., 2019. Analyzing sericea lespedeza (*lespedeza cuneata*) management practices and the
 1183 importance of forbs in the diet of cattle and bison on tallgrass prairie (Master’s Thesis).
 1184 Oklahoma State University, Stillwater, OK.
- 1185 Shoko, C., Mutanga, O., Dube, T., 2016. Progress in the remote sensing of C3 and C4 grass species
 1186 aboveground biomass over time and space. *ISPRS Journal of Photogrammetry and Remote*
 1187 *Sensing* 120, 13–24. <https://doi.org/10.1016/J.ISPRSJPRS.2016.08.001>
- 1188 Socher, S.A., Prati, D., Boch, S., Müller, J., Klaus, V.H., Hölzel, N., Fischer, M., 2012. Direct and
 1189 productivity-mediated indirect effects of fertilization, mowing and grazing on grassland species
 1190 richness. *Journal of Ecology* 100, 1391–1399. [https://doi.org/10.1111/j.1365-](https://doi.org/10.1111/j.1365-2745.2012.02020.x)
 1191 [2745.2012.02020.x](https://doi.org/10.1111/j.1365-2745.2012.02020.x)
- 1192 Stevens, A., Ramirez-Lopez, L., 2022. An introduction to the prospectr package [WWW Document].
 1193 URL <https://CRAN.R-project.org/package=prospectr>
- 1194 Tagliabue, G., Boschetti, M., Bramati, G., Candiani, G., Colombo, R., Nutini, F., Pompilio, L., Rivera-
 1195 Caicedo, J.P., Rossi, M., Rossini, M., Verrelst, J., Panigada, C., 2022. Hybrid retrieval of crop traits
 1196 from multi-temporal PRISMA hyperspectral imagery. *ISPRS Journal of Photogrammetry and*
 1197 *Remote Sensing* 187, 362–377. <https://doi.org/10.1016/j.isprs.2022.03.014>
- 1198 The Nature Conservancy, 2023. Joseph H. Williams Tallgrass Prairie Preserve [WWW Document]. URL
 1199 [https://www.nature.org/en-us/get-involved/how-to-help/places-we-protect/tallgrass-prairie-](https://www.nature.org/en-us/get-involved/how-to-help/places-we-protect/tallgrass-prairie-preserve/)
 1200 [preserve/](https://www.nature.org/en-us/get-involved/how-to-help/places-we-protect/tallgrass-prairie-preserve/) (accessed 3.29.23).
- 1201 Van Vooren, L., Reubens, B., Broekx, S., Reheul, D., Verheyen, K., 2018. Assessing the impact of
 1202 grassland management extensification in temperate areas on multiple ecosystem services and
 1203 biodiversity. *Agric Ecosyst Environ* 267, 201–212. <https://doi.org/10.1016/j.agee.2018.08.016>
- 1204 Vapnik, V., Golowich, S.E., Smola, A., 1997. Support vector method for function approximation,
 1205 regression estimation, and signal processing. *Adv Neural Inf Process Syst* 281–287.
- 1206 Verhoef, W., Jia, L., Xiao, Q., Su, Z., 2007. Unified Optical-Thermal Four-Stream Radiative Transfer
 1207 Theory for Homogeneous Vegetation Canopies. *IEEE Transactions on Geoscience and Remote*
 1208 *Sensing* 45, 1808–1822. <https://doi.org/10.1109/TGRS.2007.895844>
- 1209 Verrelst, J., Alonso, L., Rivera Caicedo, J.P., Moreno, J., Camps-Valls, G., 2013a. Gaussian Process
 1210 Retrieval of Chlorophyll Content From Imaging Spectroscopy Data. *IEEE J Sel Top Appl Earth Obs*
 1211 *Remote Sens* 6, 867–874. <https://doi.org/10.1109/JSTARS.2012.2222356>

- 1212 Verrelst, J., Camps-Valls, G., Muñoz-Marí, J., Rivera, J.P., Veroustraete, F., Clevers, J.G.P.W., Moreno,
1213 J., 2015. Optical remote sensing and the retrieval of terrestrial vegetation bio-geophysical
1214 properties – A review. *ISPRS Journal of Photogrammetry and Remote Sensing* 108, 273–290.
1215 <https://doi.org/10.1016/j.isprsjprs.2015.05.005>
- 1216 Verrelst, J., Dethier, S., Rivera, J.P., Munoz-Mari, J., Camps-Valls, G., Moreno, J., 2016. Active Learning
1217 Methods for Efficient Hybrid Biophysical Variable Retrieval. *IEEE Geoscience and Remote
1218 Sensing Letters* 13, 1012–1016. <https://doi.org/10.1109/LGRS.2016.2560799>
- 1219 Verrelst, J., Halabuk, A., Atzberger, C., Hank, T., Steinhauser, S., Berger, K., 2023. A comprehensive
1220 survey on quantifying non-photosynthetic vegetation cover and biomass from imaging
1221 spectroscopy. *Ecol Indic* 155, 110911. <https://doi.org/10.1016/j.ecolind.2023.110911>
- 1222 Verrelst, J., Muñoz, J., Alonso, L., Delegido, J., Rivera, J.P., Camps-Valls, G., Moreno, J., 2012. Machine
1223 learning regression algorithms for biophysical parameter retrieval: Opportunities for Sentinel-2
1224 and -3. *Remote Sens Environ* 118, 127–139. <https://doi.org/10.1016/j.rse.2011.11.002>
- 1225 Verrelst, J., Rivera Caicedo, J., Muñoz-Marí, J., Camps-Valls, G., Moreno, J., 2017. SCOPE-Based
1226 Emulators for Fast Generation of Synthetic Canopy Reflectance and Sun-Induced Fluorescence
1227 Spectra. *Remote Sens (Basel)* 9, 927. <https://doi.org/10.3390/rs9090927>
- 1228 Verrelst, J., Rivera-Caicedo, J.P., Reyes-Muñoz, P., Morata, M., Amin, E., Tagliabue, G., Panigada, C.,
1229 Hank, T., Berger, K., 2021. Mapping landscape canopy nitrogen content from space using
1230 PRISMA data. *ISPRS Journal of Photogrammetry and Remote Sensing* 178, 382–395.
1231 <https://doi.org/10.1016/j.isprsjprs.2021.06.017>
- 1232 Verrelst, J., Rivera, J.P., Leonenko, G., Alonso, L., Moreno, J., 2014. Optimizing LUT-Based RTM
1233 Inversion for Semiautomatic Mapping of Crop Biophysical Parameters from Sentinel-2 and -3
1234 Data: Role of Cost Functions. *IEEE Transactions on Geoscience and Remote Sensing* 52, 257–
1235 269. <https://doi.org/10.1109/TGRS.2013.2238242>
- 1236 Verrelst, J., Rivera, J.P., Moreno, J., Camps-Valls, G., 2013b. Gaussian processes uncertainty estimates
1237 in experimental Sentinel-2 LAI and leaf chlorophyll content retrieval. *ISPRS Journal of
1238 Photogrammetry and Remote Sensing* 86, 157–167.
1239 <https://doi.org/10.1016/j.isprsjprs.2013.09.012>
- 1240 Viscarra Rossel, R.A., Behrens, T., Ben-Dor, E., Brown, D.J., Demattê, J.A.M., Shepherd, K.D., Shi, Z.,
1241 Stenberg, B., Stevens, A., Adamchuk, V., Aichi, H., Barthès, B.G., Bartholomeus, H.M., Bayer,
1242 A.D., Bernoux, M., Böttcher, K., Brodský, L., Du, C.W., Chappell, A., Fouad, Y., Genot, V., Gomez,
1243 C., Grunwald, S., Gubler, A., Guerrero, C., Hedley, C.B., Knadel, M., Morrás, H.J.M., Nocita, M.,
1244 Ramirez-Lopez, L., Roudier, P., Campos, E.M.R., Sanborn, P., Sellitto, V.M., Sudduth, K.A.,
1245 Rawlins, B.G., Walter, C., Winowiecki, L.A., Hong, S.Y., Ji, W., 2016. A global spectral library to
1246 characterize the world's soil. *Earth Sci Rev* 155, 198–230.
1247 <https://doi.org/10.1016/j.earscirev.2016.01.012>
- 1248 Wang, G., Liu, S., Liu, T., Fu, Z., Yu, J., Xue, B., 2019. Modelling above-ground biomass based on
1249 vegetation indexes: a modified approach for biomass estimation in semi-arid grasslands. *Int J
1250 Remote Sens* 40, 3835–3854. <https://doi.org/10.1080/01431161.2018.1553319>
- 1251 Wang, Z., Féret, J.-B., Liu, N., Sun, Z., Yang, L., Geng, S., Zhang, H., Chlus, A., Kruger, E.L., Townsend,
1252 P.A., 2023. Generality of leaf spectroscopic models for predicting key foliar functional traits
1253 across continents: A comparison between physically- and empirically-based approaches.
1254 *Remote Sens Environ* 293, 113614. <https://doi.org/10.1016/j.rse.2023.113614>

- 1255 Wellstein, C., Chelli, S., Campetella, G., Bartha, S., Galiè, M., Spada, F., Canullo, R., 2013. Intraspecific
 1256 phenotypic variability of plant functional traits in contrasting mountain grasslands habitats.
 1257 *Biodivers Conserv* 22, 2353–2374. <https://doi.org/10.1007/s10531-013-0484-6>
- 1258 White, R., Murray, S., Rohweder, M., 2000. *Pilot Analysis of Global Ecosystems: Grassland*
 1259 *Ecosystems*. WRI Publications, Hapden Station, Baltimore, MD.
- 1260 Woche, M., Berger, K., Verrelst, J., Hank, T., 2022. Retrieval of carbon content and biomass from
 1261 hyperspectral imagery over cultivated areas. *ISPRS Journal of Photogrammetry and Remote*
 1262 *Sensing* 193, 104–114. <https://doi.org/10.1016/j.isprsjprs.2022.09.003>
- 1263 Woodcock, C.E., 2002. Uncertainty in Remote Sensing, in: Foody, G.M., Atkinson, P.M. (Eds.),
 1264 *Uncertainty in Remote Sensing and GIS*. Wiley, Chichester, pp. 19–24.
 1265 <https://doi.org/10.1002/0470035269.ch2>
- 1266 Xu, D., Guo, X., Li, Z., Yang, X., Yin, H., 2014. Measuring the dead component of mixed grassland with
 1267 Landsat imagery. *Remote Sens Environ* 142, 33–43. <https://doi.org/10.1016/j.rse.2013.11.017>
- 1268 Yuan, H., Ma, R., Atzberger, C., Li, F., Loisel, S., Luo, J., 2015. Estimating Forest fAPAR from
 1269 Multispectral Landsat-8 Data Using the Invertible Forest Reflectance Model INFORM. *Remote*
 1270 *Sens (Basel)* 7, 7425–7446. <https://doi.org/10.3390/rs70607425>
- 1271 Yuhas, R., Goetz, A., Boardman, J., 1992. Discrimination among semi-arid landscape endmembers
 1272 using the spectral angle mapper (SAM) algorithm, in: *Summaries of the Third Annual JPL*
 1273 *Airborne Geoscience Workshop. Volume 1: AVIRIS Workshop*. JPL, pp. 147–149.
- 1274 Zeng, Y., Hao, D., Park, T., Zhu, P., Huete, A., Myneni, R., Knyazikhin, Y., Qi, J., Nemani, R.R., Li, F.,
 1275 Huang, J., Gao, Y., Li, B., Ji, F., Köhler, P., Frankenberg, C., Berry, J.A., Chen, M., 2023. Structural
 1276 complexity biases vegetation greenness measures. *Nat Ecol Evol* 7, 1790–1798.
 1277 <https://doi.org/10.1038/s41559-023-02187-6>
- 1278 Zhang, F., Bennett, J.A., Zhang, B., Zhao, T., Bai, K., Zhao, M., Han, G., 2023. Responses of grassland
 1279 productivity to mowing intensity and precipitation variability in a temperate steppe. *Oecologia*
 1280 201, 259–268. <https://doi.org/10.1007/s00442-022-05305-6>
- 1281 Zhao, Y., Liu, Z., Wu, J., 2020. Grassland ecosystem services: a systematic review of research
 1282 advances and future directions. *Landsc Ecol* 35, 793–814. [https://doi.org/10.1007/s10980-020-](https://doi.org/10.1007/s10980-020-00980-3)
 1283 [00980-3](https://doi.org/10.1007/s10980-020-00980-3)
- 1284 Zhou, K., Liu, Z., Qiao, Y., Xiang, T., Loy, C.C., 2022. Domain Generalization: A Survey. *IEEE Trans*
 1285 *Pattern Anal Mach Intell* 1–20. <https://doi.org/10.1109/TPAMI.2022.3195549>
- 1286 Zupanc, A., 2017. Improving Cloud Detection with Machine Learning [WWW Document]. URL
 1287 [https://medium.com/sentinel-hub/improving-cloud-detection-with-machine-learning-](https://medium.com/sentinel-hub/improving-cloud-detection-with-machine-learning-c09dc5d7cf13)
 1288 [c09dc5d7cf13](https://medium.com/sentinel-hub/improving-cloud-detection-with-machine-learning-c09dc5d7cf13) (accessed 9.28.21).
- 1289

1 **Appendix A**

2 **Section A.1: Pre-processing of APEX data**

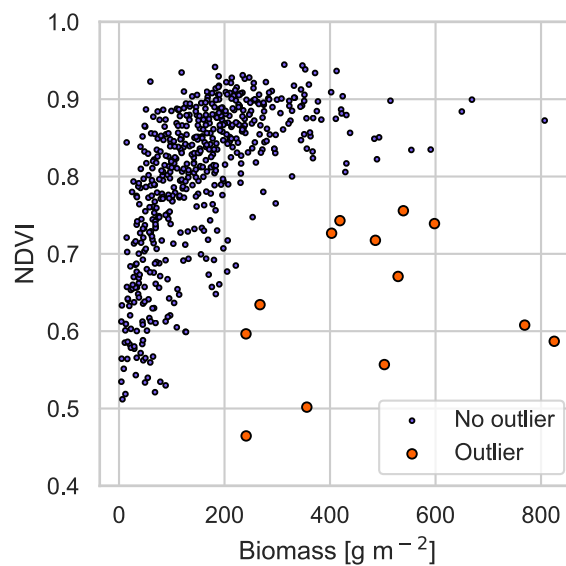
3 Remote sensing data for the 365 plots inside the Swiss National Park (SNP) were acquired on
4 24 June 2010, 26 June 2011, 29 June 2012, and 12 July 2013 with the Airborne Prism
5 Experiment (APEX) imaging spectrometer (Jehle et al., 2010; Schaepman et al., 2015). Of 334
6 spectral bands in the wavelength region from 380 to 2500 nm, 285 (2011), 301 (2011), 299
7 (2012), and 284 bands (2013) remained after noise removal. APEX data were resampled to 2
8 m pixel size using nearest neighbor interpolation and the parametric geocoding procedure
9 PARGE (Schläpfer and Richter, 2002) and the airborne atmospheric and topographic correction
10 model ATCOR-4 (Richter and Schläpfer, 2002) were used for geometric and atmospheric
11 correction, respectively (Schweiger et al., 2015). We considered the 33 plots in 2010 and 9 plots
12 in 2011 covered by two flight strips as individual samples, resulting in 407 samples. APEX
13 spectral reflectance was resampled to the Sentinel-2 bands (ESA, 2015) using the prospectR R
14 package *v0.2.6* (Stevens and Ramirez-Lopez, 2022) individually for each year because of
15 changing APEX spectral response functions.

16 **Section A.2: Outliers of ALB/HAI/SCH sites**

17 The following condition (Equation A.1) was used to determine if a sample from the
18 ALB/HAI/SCH sites was discarded due to an unusual biomass-Normalized Difference
19 Vegetation Index (NDVI) relation while accounting for the saturation effect of NDVI (Huete et
20 al., 1997; Van Der Meer et al., 2001):

21
$$NDVI < \frac{biomass [g m^{-2}]}{400} \text{ AND } NDVI < 0.8 \text{ (A.1)}$$

22 The condition was met for 13 samples of 12 different plots (Figure A.1), for which we
23 removed all samples from 2017-2020.



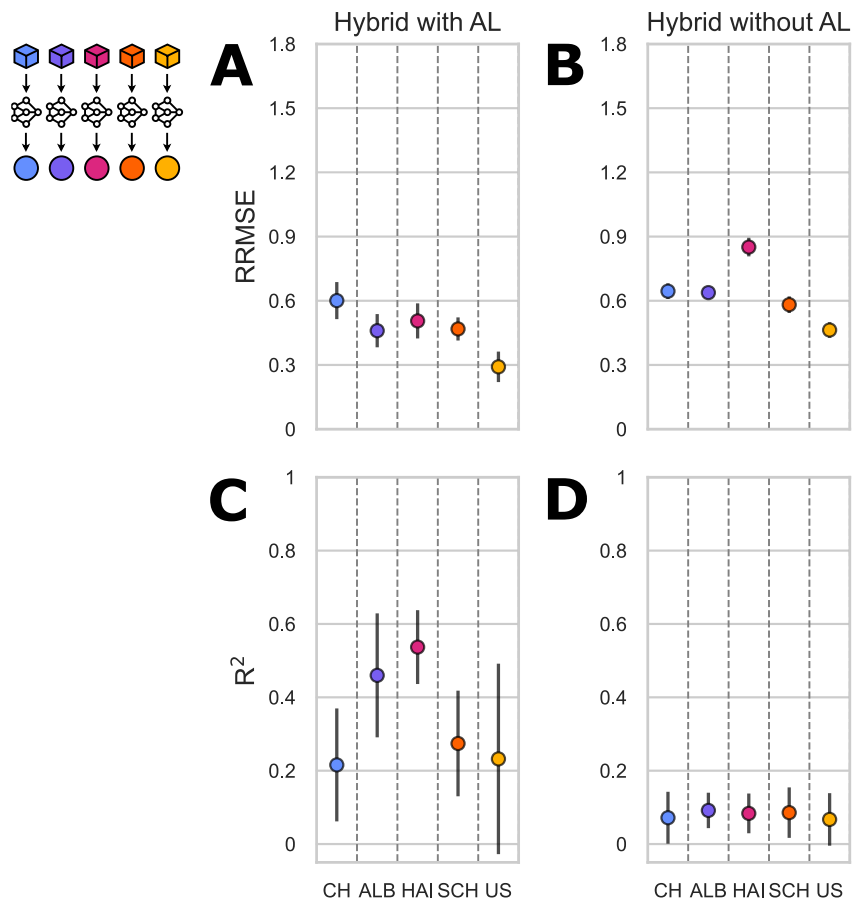
24

25 *Figure A.1: Biomass content of the ALB, HAI, and SCH samples before sample selection plotted against the Normalized*
26 *Difference Vegetation Index (NDVI). ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin.*

27

28 **Section A.3: Local performance of hybrid models without the use of AL**

29 Look-up tables (LUTs) generated with radiative transfer models (RTMs) such as PROSAIL are
30 subject to inherent ill-posedness as the inversion does not necessarily lead to a unique solution
31 (Combal et al., 2003). Additionally, such a LUT might include spectra simulated based on
32 ecologically meaningless parameter combinations. Hence, prior knowledge can be used to
33 subset the LUT based on correlations between plant traits (Schiefer et al., 2021). In this study,
34 we used prior knowledge to define the value ranges and distributions of the selectable
35 parameters, but we did not use prior information to further subset the LUT since covariances
36 between traits might be site-specific and generating different LUTs for the study sites would
37 have prevented any meaningful analysis of transferability. Instead, Active Learning (AL) was
38 used to select the most informative spectra from the LUT (Verrelst et al., 2016).
39 Correspondingly, not using AL leads to models being trained on the full LUT containing
40 redundant and/or unrealistic information and drastically diminished model performance as
41 illustrated in Figure A.2. Furthermore, without AL, model training is also independent of the
42 optimization data, since no optimization of the training data takes place. Therefore, the
43 inclusion of hybrid models without the use of AL for the systematic model comparison was
44 waived.



45

46 *Figure A.2: Mean relative root-mean-square error (RRMSE, A, B) and coefficient of determination (R^2 , C, D) ± 1 standard*

47 *deviation for the local hybrid models with (A, C) and without (B, D) the use of Active Learning (AL) with an initial training set*

48 *size of 2%. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States.*

49 Section A.4: Cross-validation performance of local empirical models

50 Table A.1: Cross-validation performance of local empirical models. The parameters correspond to the best-performing
51 combination of parameters tested during the fivefold cross-validation. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-
52 Dün, SCH: Schorfheide-Chorin, US: United States, RFR: Random Forest regression, SVR: Support Vector regression, XGB:
53 Extreme Gradient Boosting regression, GPR: Gaussian Process regression, R^2 : coefficient of determination, RMSE: root-mean-
54 square error, RRMSE: relative root-mean-square error. Nomenclature of parameter names for RFR, SVR, and GPR according
55 to Pedregosa et al. (2012), for XGB according to Chen and Guestrin (2016).

Site	Model	Seed	Parameters	R^2	RMSE [g m ⁻²]	RRMSE
CH	RFR	1	'max_depth': None, 'max_features': 'sqrt', 'max_leaf_nodes': None, 'min_samples_leaf': 1, 'n_estimators': 100	0.26 2	97.383	0.567
CH	RFR	2	'max_depth': 5, 'max_features': 'sqrt', 'max_leaf_nodes': None, 'min_samples_leaf': 1, 'n_estimators': 100	0.22	97.707	0.559
CH	RFR	3	'max_depth': None, 'max_features': 10, 'max_leaf_nodes': 10, 'min_samples_leaf': 2, 'n_estimators': 100	0.18 4	104.372	0.595
CH	RFR	4	'max_depth': None, 'max_features': 10, 'max_leaf_nodes': None, 'min_samples_leaf': 2, 'n_estimators': 100	0.25	98.643	0.561
CH	RFR	5	'max_depth': 10, 'max_features': 10, 'max_leaf_nodes': None, 'min_samples_leaf': 1, 'n_estimators': 200	0.23 9	97.227	0.553
CH	RFR	6	'max_depth': 10, 'max_features': 'sqrt', 'max_leaf_nodes': None, 'min_samples_leaf': 1, 'n_estimators': 100	0.22 1	94.349	0.55
CH	RFR	7	'max_depth': None, 'max_features': 10, 'max_leaf_nodes': 10, 'min_samples_leaf': 2, 'n_estimators': 100	0.25 4	100.899	0.571
CH	RFR	8	'max_depth': 5, 'max_features': 'sqrt', 'max_leaf_nodes': 20, 'min_samples_leaf': 1, 'n_estimators': 200	0.21 7	99.517	0.578
CH	RFR	9	'max_depth': 5, 'max_features': 'sqrt', 'max_leaf_nodes': None, 'min_samples_leaf': 5, 'n_estimators': 200	0.17 2	104.461	0.593
CH	RFR	10	'max_depth': None, 'max_features': 'sqrt', 'max_leaf_nodes': None, 'min_samples_leaf': 2, 'n_estimators': 500	0.24 3	97.024	0.548
CH	SVR	1	'svr_C': 10, 'svr_epsilon': 0.01, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.32 3	95.135	0.555
CH	SVR	2	'svr_C': 1, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.22	96.989	0.554
CH	SVR	3	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.18 2	96.688	0.558
CH	SVR	4	'svr_C': 1, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.22	98.606	0.566
CH	SVR	5	'svr_C': 100, 'svr_epsilon': 0.01, 'svr_gamma': 'auto', 'svr_kernel': 'rbf'	0.29 4	93.806	0.527
CH	SVR	6	'svr_C': 10, 'svr_epsilon': 0.01, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.22 8	96.36	0.563
CH	SVR	7	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.29 6	93.581	0.531
CH	SVR	8	'svr_C': 1, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.25 9	96.224	0.555
CH	SVR	9	'svr_C': 10, 'svr_epsilon': 0.1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.22 7	96.138	0.549
CH	SVR	10	'svr_C': 1, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.20 5	97.674	0.555
CH	XGB	1	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 7, 'xgb_n_estimators': 300, 'xgb_subsample': 1.0	0.19 8	96.853	0.574
CH	XGB	2	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0, 'xgb_learning_rate': 0.1, 'xgb_max_depth': 3, 'xgb_n_estimators': 100, 'xgb_subsample': 0.8	0.21 7	99.454	0.567
CH	XGB	3	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0, 'xgb_learning_rate': 0.1, 'xgb_max_depth': 9, 'xgb_n_estimators': 100, 'xgb_subsample': 1.0	0.18 1	102.339	0.603
CH	XGB	4	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0.2, 'xgb_learning_rate': 0.2, 'xgb_max_depth': 7, 'xgb_n_estimators': 100, 'xgb_subsample': 0.8	0.24 3	98.511	0.562
CH	XGB	5	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0.2, 'xgb_learning_rate': 0.2, 'xgb_max_depth': 7, 'xgb_n_estimators': 200, 'xgb_subsample': 0.8	0.22 5	90.404	0.523
CH	XGB	6	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0.1, 'xgb_learning_rate': 0.1, 'xgb_max_depth': 7, 'xgb_n_estimators': 300, 'xgb_subsample': 0.8	0.26 6	90.79	0.529
CH	XGB	7	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 3, 'xgb_n_estimators': 300, 'xgb_subsample': 1.0	0.16	98.689	0.579
CH	XGB	8	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0.1, 'xgb_learning_rate': 0.1, 'xgb_max_depth': 7, 'xgb_n_estimators': 300, 'xgb_subsample': 1.0	0.22 9	93.968	0.555
CH	XGB	9	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0, 'xgb_learning_rate': 0.1, 'xgb_max_depth': 7, 'xgb_n_estimators': 100, 'xgb_subsample': 0.8	0.15 7	101.16	0.583
CH	XGB	10	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0.1, 'xgb_learning_rate': 0.1, 'xgb_max_depth': 5, 'xgb_n_estimators': 300, 'xgb_subsample': 0.8	0.24 9	97.957	0.562
CH	GPR	1	1.04**2 * RBF(length_scale=0.555)	0.03	112.276	0.669
CH	GPR	2	1.09**2 * RBF(length_scale=0.614)	0.03 4	113.143	0.66
CH	GPR	3	1.11**2 * RBF(length_scale=0.717)	0.04 6	116.921	0.679
CH	GPR	4	1.05**2 * RBF(length_scale=0.542)	0.09 1	117.013	0.691

CH	GPR	5	1.05**2 * RBF(length_scale=0.663)	0.07 1	111.359	0.644
CH	GPR	6	1.07**2 * RBF(length_scale=0.643)	0.08 9	110.957	0.655
CH	GPR	7	1.06**2 * RBF(length_scale=0.699)	0.07 5	111.392	0.646
CH	GPR	8	1.06**2 * RBF(length_scale=0.677)	0.05 4	107.571	0.631
CH	GPR	9	1.07**2 * RBF(length_scale=0.607)	0.06 3	119.623	0.693
CH	GPR	10	1.06**2 * RBF(length_scale=0.603)	0.04 3	114.67	0.672
AL B	RFR	1	'max_depth': None, 'max_features': 'sqrt', 'max_leaf_nodes': None, 'min_samples_leaf': 5, 'n_estimators': 100	0.49 2	61.884	0.447
AL B	RFR	2	'max_depth': None, 'max_features': 'sqrt', 'max_leaf_nodes': None, 'min_samples_leaf': 5, 'n_estimators': 500	0.38 9	62.498	0.463
AL B	RFR	3	'max_depth': None, 'max_features': 'sqrt', 'max_leaf_nodes': 10, 'min_samples_leaf': 5, 'n_estimators': 100	0.47	60.956	0.444
AL B	RFR	4	'max_depth': 5, 'max_features': 10, 'max_leaf_nodes': 10, 'min_samples_leaf': 5, 'n_estimators': 200	0.47 3	59.714	0.439
AL B	RFR	5	'max_depth': 5, 'max_features': 10, 'max_leaf_nodes': 10, 'min_samples_leaf': 1, 'n_estimators': 100	0.49 1	57.978	0.418
AL B	RFR	6	'max_depth': None, 'max_features': 'sqrt', 'max_leaf_nodes': 10, 'min_samples_leaf': 2, 'n_estimators': 100	0.59 8	54.306	0.388
AL B	RFR	7	'max_depth': 5, 'max_features': 'sqrt', 'max_leaf_nodes': 20, 'min_samples_leaf': 1, 'n_estimators': 100	0.59 9	54.284	0.395
AL B	RFR	8	'max_depth': None, 'max_features': 10, 'max_leaf_nodes': 10, 'min_samples_leaf': 5, 'n_estimators': 200	0.48 7	61.801	0.451
AL B	RFR	9	'max_depth': 5, 'max_features': 10, 'max_leaf_nodes': 10, 'min_samples_leaf': 5, 'n_estimators': 200	0.51 1	60.256	0.437
AL B	RFR	10	'max_depth': None, 'max_features': 10, 'max_leaf_nodes': None, 'min_samples_leaf': 5, 'n_estimators': 100	0.48 4	61.236	0.445
AL B	SVR	1	'svr_C': 10, 'svr_epsilon': 0.01, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.52 7	60.243	0.435
AL B	SVR	2	'svr_C': 10, 'svr_epsilon': 0.5, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.46 5	58.95	0.435
AL B	SVR	3	'svr_C': 10, 'svr_epsilon': 0.01, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.48 8	59.269	0.431
AL B	SVR	4	'svr_C': 10, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.50 2	59.348	0.434
AL B	SVR	5	'svr_C': 10, 'svr_epsilon': 0.01, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.57 2	54.454	0.393
AL B	SVR	6	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 1, 'svr_kernel': 'rbf'	0.6	56.361	0.406
AL B	SVR	7	'svr_C': 10, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.62 4	55.656	0.406
AL B	SVR	8	'svr_C': 10, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.51 2	63.062	0.458
AL B	SVR	9	'svr_C': 100, 'svr_epsilon': 0.01, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.57 1	59.17	0.428
AL B	SVR	10	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'rbf'	0.51 7	59.66	0.436
AL B	XGB	1	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0.1, 'xgb_learning_rate': 0.2, 'xgb_max_depth': 5, 'xgb_n_estimators': 100, 'xgb_subsample': 0.8	0.49 4	63.028	0.454
AL B	XGB	2	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0.2, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 3, 'xgb_n_estimators': 200, 'xgb_subsample': 0.8	0.36 6	63.793	0.47
AL B	XGB	3	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0.1, 'xgb_learning_rate': 0.1, 'xgb_max_depth': 3, 'xgb_n_estimators': 100, 'xgb_subsample': 1.0	0.52 5	60.109	0.44
AL B	XGB	4	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 5, 'xgb_n_estimators': 200, 'xgb_subsample': 0.8	0.50 9	57.863	0.424
AL B	XGB	5	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 3, 'xgb_n_estimators': 300, 'xgb_subsample': 1.0	0.50 6	56.811	0.411
AL B	XGB	6	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0.2, 'xgb_learning_rate': 0.1, 'xgb_max_depth': 3, 'xgb_n_estimators': 100, 'xgb_subsample': 0.8	0.62	51.448	0.367
AL B	XGB	7	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 3, 'xgb_n_estimators': 300, 'xgb_subsample': 0.8	0.62 7	51.687	0.375
AL B	XGB	8	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 3, 'xgb_n_estimators': 300, 'xgb_subsample': 0.8	0.50 2	61.415	0.448
AL B	XGB	9	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0.1, 'xgb_learning_rate': 0.2, 'xgb_max_depth': 5, 'xgb_n_estimators': 100, 'xgb_subsample': 0.8	0.56	58.001	0.416
AL B	XGB	10	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 3, 'xgb_n_estimators': 200, 'xgb_subsample': 0.8	0.49 9	60.569	0.44
AL B	GPR	1	0.969**2 * RBF(length_scale=0.582)	0.36 4	70.641	0.51
AL B	GPR	2	0.988**2 * RBF(length_scale=0.571)	0.22 9	73.522	0.554
AL B	GPR	3	0.971**2 * RBF(length_scale=0.596)	0.31 2	73.961	0.551
AL B	GPR	4	1**2 * RBF(length_scale=0.602)	0.26 3	70.827	0.523
AL B	GPR	5	0.963**2 * RBF(length_scale=0.637)	0.34 6	69.188	0.506
AL B	GPR	6	0.954**2 * RBF(length_scale=0.728)	0.57 7	57.247	0.409
AL B	GPR	7	0.988**2 * RBF(length_scale=0.69)	0.41 8	66.398	0.487
AL B	GPR	8	0.96**2 * RBF(length_scale=0.603)	0.32 7	70.746	0.512
AL B	GPR	9	0.981**2 * RBF(length_scale=0.68)	0.50 3	61.572	0.445

AL B	GPR	10	0.978**2 * RBF(length_scale=0.592)	0.43 7	67.059	0.49
HA I	RFR	1	'max_depth': 10, 'max_features': 'sqrt', 'max_leaf_nodes': 20, 'min_samples_leaf': 2, 'n_estimators': 500	0.60 1	58.344	0.484
HA I	RFR	2	'max_depth': 5, 'max_features': 'sqrt', 'max_leaf_nodes': None, 'min_samples_leaf': 2, 'n_estimators': 200	0.56 7	58.231	0.481
HA I	RFR	3	'max_depth': 5, 'max_features': 'sqrt', 'max_leaf_nodes': 20, 'min_samples_leaf': 1, 'n_estimators': 200	0.49 4	60.211	0.51
HA I	RFR	4	'max_depth': 5, 'max_features': 'sqrt', 'max_leaf_nodes': None, 'min_samples_leaf': 5, 'n_estimators': 200	0.62 6	55.599	0.471
HA I	RFR	5	'max_depth': None, 'max_features': 'sqrt', 'max_leaf_nodes': 20, 'min_samples_leaf': 1, 'n_estimators': 100	0.56 3	57.184	0.48
HA I	RFR	6	'max_depth': 10, 'max_features': 'sqrt', 'max_leaf_nodes': None, 'min_samples_leaf': 1, 'n_estimators': 100	0.67 7	50.357	0.443
HA I	RFR	7	'max_depth': None, 'max_features': 'sqrt', 'max_leaf_nodes': None, 'min_samples_leaf': 5, 'n_estimators': 100	0.54 6	59.045	0.504
HA I	RFR	8	'max_depth': 5, 'max_features': 'sqrt', 'max_leaf_nodes': None, 'min_samples_leaf': 1, 'n_estimators': 500	0.59 3	56.409	0.471
HA I	RFR	9	'max_depth': 15, 'max_features': 'sqrt', 'max_leaf_nodes': None, 'min_samples_leaf': 1, 'n_estimators': 100	0.55 5	60.494	0.516
HA I	RFR	10	'max_depth': None, 'max_features': 'sqrt', 'max_leaf_nodes': None, 'min_samples_leaf': 5, 'n_estimators': 100	0.54 4	59.877	0.509
HA I	SVR	1	'svr_C': 100, 'svr_epsilon': 0.1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.64	55.258	0.461
HA I	SVR	2	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.59 1	55.823	0.463
HA I	SVR	3	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.63 7	52.765	0.447
HA I	SVR	4	'svr_C': 100, 'svr_epsilon': 0.1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.67	51.742	0.438
HA I	SVR	5	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.57 2	55.933	0.469
HA I	SVR	6	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.64 6	50.297	0.441
HA I	SVR	7	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.62 4	53.9	0.458
HA I	SVR	8	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.68 7	50.765	0.423
HA I	SVR	9	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.57 9	59.724	0.506
HA I	SVR	10	'svr_C': 100, 'svr_epsilon': 0.5, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.63 1	53.207	0.447
HA I	XGB	1	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0.2, 'xgb_learning_rate': 0.2, 'xgb_max_depth': 3, 'xgb_n_estimators': 300, 'xgb_subsample': 1.0	0.59 7	59.358	0.491
HA I	XGB	2	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0, 'xgb_learning_rate': 0.2, 'xgb_max_depth': 3, 'xgb_n_estimators': 100, 'xgb_subsample': 0.8	0.57 1	58.098	0.48
HA I	XGB	3	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 3, 'xgb_n_estimators': 300, 'xgb_subsample': 0.8	0.47 6	61.115	0.517
HA I	XGB	4	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 3, 'xgb_n_estimators': 300, 'xgb_subsample': 1.0	0.58 7	59.375	0.501
HA I	XGB	5	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0.1, 'xgb_learning_rate': 0.1, 'xgb_max_depth': 3, 'xgb_n_estimators': 100, 'xgb_subsample': 0.8	0.50 8	64.257	0.54
HA I	XGB	6	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 3, 'xgb_n_estimators': 300, 'xgb_subsample': 1.0	0.62	54.179	0.475
HA I	XGB	7	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0.1, 'xgb_learning_rate': 0.2, 'xgb_max_depth': 9, 'xgb_n_estimators': 100, 'xgb_subsample': 0.8	0.48 3	64.233	0.548
HA I	XGB	8	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0, 'xgb_learning_rate': 0.1, 'xgb_max_depth': 3, 'xgb_n_estimators': 100, 'xgb_subsample': 0.8	0.61 1	56.071	0.467
HA I	XGB	9	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 3, 'xgb_n_estimators': 200, 'xgb_subsample': 0.8	0.48 4	64.76	0.548
HA I	XGB	10	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 3, 'xgb_n_estimators': 200, 'xgb_subsample': 0.8	0.46 4	64.489	0.545
HA I	GPR	1	1.13**2 * RBF(length_scale=0.956)	0.49 1	68.263	0.569
HA I	GPR	2	1.05**2 * RBF(length_scale=0.939)	0.56 4	61.494	0.498
HA I	GPR	3	1.04**2 * RBF(length_scale=0.817)	0.44 3	65.875	0.557
HA I	GPR	4	1.01**2 * RBF(length_scale=0.866)	0.54 1	62.848	0.531
HA I	GPR	5	1.04**2 * RBF(length_scale=0.87)	0.48 9	61.989	0.516
HA I	GPR	6	0.944**2 * RBF(length_scale=0.82)	0.47 7	59.248	0.52
HA I	GPR	7	1.06**2 * RBF(length_scale=0.87)	0.46 9	66.942	0.564
HA I	GPR	8	1.03**2 * RBF(length_scale=0.835)	0.47 9	64.916	0.544
HA I	GPR	9	1.09**2 * RBF(length_scale=0.963)	0.51 2	62.062	0.53
HA I	GPR	10	1.03**2 * RBF(length_scale=0.853)	0.44 9	65.897	0.559
SC H	RFR	1	'max_depth': None, 'max_features': 'sqrt', 'max_leaf_nodes': 10, 'min_samples_leaf': 2, 'n_estimators': 200	0.35 4	95.411	0.432
SC H	RFR	2	'max_depth': 5, 'max_features': 'sqrt', 'max_leaf_nodes': 10, 'min_samples_leaf': 5, 'n_estimators': 100	0.32 8	94.618	0.427
SC H	RFR	3	'max_depth': 5, 'max_features': 'sqrt', 'max_leaf_nodes': 10, 'min_samples_leaf': 2, 'n_estimators': 100	0.43 4	90.285	0.402
SC H	RFR	4	'max_depth': 5, 'max_features': 'sqrt', 'max_leaf_nodes': None, 'min_samples_leaf': 1, 'n_estimators': 100	0.33 1	98.242	0.435

SC H	RFR	5	'max_depth': None, 'max_features': 'sqrt', 'max_leaf_nodes': 10, 'min_samples_leaf': 1, 'n_estimators': 100	0.34 3	97.047	0.445
SC H	RFR	6	'max_depth': 5, 'max_features': 10, 'max_leaf_nodes': 10, 'min_samples_leaf': 5, 'n_estimators': 100	0.39	94.588	0.421
SC H	RFR	7	'max_depth': None, 'max_features': 'sqrt', 'max_leaf_nodes': 10, 'min_samples_leaf': 2, 'n_estimators': 200	0.36 1	101.901	0.456
SC H	RFR	8	'max_depth': 10, 'max_features': 'sqrt', 'max_leaf_nodes': None, 'min_samples_leaf': 2, 'n_estimators': 100	0.40 8	94.495	0.423
SC H	RFR	9	'max_depth': 10, 'max_features': 'sqrt', 'max_leaf_nodes': None, 'min_samples_leaf': 2, 'n_estimators': 500	0.32 3	100.8	0.448
SC H	RFR	10	'max_depth': 5, 'max_features': 'sqrt', 'max_leaf_nodes': 20, 'min_samples_leaf': 2, 'n_estimators': 100	0.33 3	101.181	0.454
SC H	SVR	1	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'rbf'	0.35 4	93.693	0.423
SC H	SVR	2	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.36 1	93.28	0.418
SC H	SVR	3	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'rbf'	0.42 1	93.033	0.413
SC H	SVR	4	'svr_C': 100, 'svr_epsilon': 0.01, 'svr_gamma': 'scale', 'svr_kernel': 'rbf'	0.42 5	91.405	0.404
SC H	SVR	5	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'rbf'	0.36 8	92.754	0.421
SC H	SVR	6	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.42 8	92.56	0.41
SC H	SVR	7	'svr_C': 100, 'svr_epsilon': 0.01, 'svr_gamma': 'scale', 'svr_kernel': 'rbf'	0.36 2	100.513	0.445
SC H	SVR	8	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.47 4	94.222	0.425
SC H	SVR	9	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'rbf'	0.40 8	95.513	0.426
SC H	SVR	10	'svr_C': 100, 'svr_epsilon': 0.01, 'svr_gamma': 'scale', 'svr_kernel': 'rbf'	0.45 4	92.601	0.416
SC H	XGB	1	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 3, 'xgb_n_estimators': 300, 'xgb_subsample': 1.0	0.31 3	97.182	0.439
SC H	XGB	2	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 3, 'xgb_n_estimators': 200, 'xgb_subsample': 0.8	0.27 1	99.375	0.451
SC H	XGB	3	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0.1, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 7, 'xgb_n_estimators': 200, 'xgb_subsample': 1.0	0.42 5	92.165	0.41
SC H	XGB	4	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 3, 'xgb_n_estimators': 300, 'xgb_subsample': 0.8	0.30 2	98.417	0.436
SC H	XGB	5	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 3, 'xgb_n_estimators': 200, 'xgb_subsample': 0.8	0.30 3	97.967	0.448
SC H	XGB	6	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 3, 'xgb_n_estimators': 200, 'xgb_subsample': 1.0	0.41	94.208	0.42
SC H	XGB	7	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 3, 'xgb_n_estimators': 300, 'xgb_subsample': 0.8	0.36 3	105.444	0.473
SC H	XGB	8	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 3, 'xgb_n_estimators': 300, 'xgb_subsample': 0.8	0.39 2	94.985	0.422
SC H	XGB	9	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 3, 'xgb_n_estimators': 200, 'xgb_subsample': 0.8	0.29 1	103.28	0.459
SC H	XGB	10	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 3, 'xgb_n_estimators': 300, 'xgb_subsample': 0.8	0.31 4	102.488	0.461
SC H	GPR	1	1.03**2 * RBF(length_scale=0.542)	0.05 5	112.547	0.511
SC H	GPR	2	1**2 * RBF(length_scale=0.0912)	0	116.645	0.531
SC H	GPR	3	1.05**2 * RBF(length_scale=0.664)	0.02 8	116.241	0.518
SC H	GPR	4	1.05**2 * RBF(length_scale=0.624)	0.06 7	119.988	0.532
SC H	GPR	5	1.04**2 * RBF(length_scale=0.653)	0.14 6	110.328	0.505
SC H	GPR	6	1.04**2 * RBF(length_scale=0.618)	0.00 9	121.548	0.542
SC H	GPR	7	1.07**2 * RBF(length_scale=0.817)	0.14 9	119.432	0.537
SC H	GPR	8	1.04**2 * RBF(length_scale=0.679)	0.22 1	106.116	0.476
SC H	GPR	9	1.06**2 * RBF(length_scale=0.831)	0.24 5	105.753	0.473
SC H	GPR	10	1.05**2 * RBF(length_scale=0.671)	0.05 1	120.172	0.539
US	RFR	1	'max_depth': None, 'max_features': 10, 'max_leaf_nodes': 10, 'min_samples_leaf': 1, 'n_estimators': 100	0.43 3	68.608	0.267
US	RFR	2	'max_depth': 5, 'max_features': 10, 'max_leaf_nodes': 10, 'min_samples_leaf': 2, 'n_estimators': 200	0.44 4	63.617	0.25
US	RFR	3	'max_depth': None, 'max_features': 10, 'max_leaf_nodes': None, 'min_samples_leaf': 5, 'n_estimators': 500	0.37 9	67.218	0.259
US	RFR	4	'max_depth': 5, 'max_features': 10, 'max_leaf_nodes': 10, 'min_samples_leaf': 2, 'n_estimators': 200	0.46 1	70.669	0.281
US	RFR	5	'max_depth': 5, 'max_features': 10, 'max_leaf_nodes': 10, 'min_samples_leaf': 2, 'n_estimators': 100	0.45 1	66.11	0.254
US	RFR	6	'max_depth': None, 'max_features': 'sqrt', 'max_leaf_nodes': 10, 'min_samples_leaf': 2, 'n_estimators': 200	0.42 2	64.712	0.261
US	RFR	7	'max_depth': None, 'max_features': 10, 'max_leaf_nodes': 20, 'min_samples_leaf': 1, 'n_estimators': 100	0.37 7	74.124	0.29
US	RFR	8	'max_depth': None, 'max_features': 10, 'max_leaf_nodes': None, 'min_samples_leaf': 5, 'n_estimators': 100	0.39 5	67.47	0.26
US	RFR	9	'max_depth': None, 'max_features': 10, 'max_leaf_nodes': 10, 'min_samples_leaf': 1, 'n_estimators': 100	0.41 7	68.961	0.266

US	RFR	10	'max_depth': None, 'max_features': 10, 'max_leaf_nodes': None, 'min_samples_leaf': 5, 'n_estimators': 100	0.308	71.819	0.279
US	SVR	1	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.494	62.487	0.24
US	SVR	2	'svr_C': 100, 'svr_epsilon': 0.01, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.602	50.272	0.197
US	SVR	3	'svr_C': 100, 'svr_epsilon': 0.01, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.496	61.333	0.235
US	SVR	4	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.531	60.381	0.236
US	SVR	5	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.547	60.469	0.232
US	SVR	6	'svr_C': 100, 'svr_epsilon': 0.5, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.569	52.136	0.207
US	SVR	7	'svr_C': 100, 'svr_epsilon': 0.1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.562	60.417	0.236
US	SVR	8	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.432	64.762	0.248
US	SVR	9	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.455	64.825	0.247
US	SVR	10	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'linear'	0.449	63.514	0.244
US	XGB	1	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 5, 'xgb_n_estimators': 200, 'xgb_subsample': 0.8	0.411	69.243	0.268
US	XGB	2	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0, 'xgb_learning_rate': 0.2, 'xgb_max_depth': 3, 'xgb_n_estimators': 100, 'xgb_subsample': 0.8	0.398	66.15	0.258
US	XGB	3	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0, 'xgb_learning_rate': 0.1, 'xgb_max_depth': 3, 'xgb_n_estimators': 300, 'xgb_subsample': 0.8	0.375	67.86	0.258
US	XGB	4	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 5, 'xgb_n_estimators': 200, 'xgb_subsample': 0.8	0.376	76.427	0.302
US	XGB	5	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 3, 'xgb_n_estimators': 200, 'xgb_subsample': 1.0	0.502	64.874	0.252
US	XGB	6	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0.2, 'xgb_learning_rate': 0.1, 'xgb_max_depth': 3, 'xgb_n_estimators': 300, 'xgb_subsample': 0.8	0.403	66.616	0.267
US	XGB	7	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0.2, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 7, 'xgb_n_estimators': 200, 'xgb_subsample': 0.8	0.35	74.251	0.291
US	XGB	8	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 3, 'xgb_n_estimators': 300, 'xgb_subsample': 1.0	0.364	67.505	0.258
US	XGB	9	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0.1, 'xgb_learning_rate': 0.2, 'xgb_max_depth': 7, 'xgb_n_estimators': 100, 'xgb_subsample': 1.0	0.355	68.738	0.261
US	XGB	10	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 5, 'xgb_n_estimators': 200, 'xgb_subsample': 1.0	0.384	67.266	0.263
US	GPR	1	1.2**2 * RBF(length_scale=0.816)	0.255	83.361	0.323
US	GPR	2	1.06**2 * RBF(length_scale=0.696)	0.294	68.962	0.269
US	GPR	3	1.14**2 * RBF(length_scale=0.79)	0.21	80.838	0.31
US	GPR	4	1.12**2 * RBF(length_scale=0.725)	0.147	88.532	0.347
US	GPR	5	1.21**2 * RBF(length_scale=0.898)	0.207	85.777	0.33
US	GPR	6	1.06**2 * RBF(length_scale=0.661)	0.248	71.481	0.283
US	GPR	7	1.11**2 * RBF(length_scale=0.646)	0.182	87.213	0.341
US	GPR	8	1.1**2 * RBF(length_scale=0.744)	0.213	82.567	0.314
US	GPR	9	1.19**2 * RBF(length_scale=0.814)	0.167	82.955	0.317
US	GPR	10	1.1**2 * RBF(length_scale=0.718)	0.156	80.403	0.313

57 **Section A.5: Cross-validation performance of local physically-based models**

58 *Table A.2: Cross-validation performance of local physically-based models. The parameters correspond to the best-performing*
 59 *combination of parameters tested during the fivefold cross-validation. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-*
 60 *Dün, SCH: Schorfheide-Chorin, US: United States, R²: coefficient of determination, RMSE: root-mean-square error, RRMSE:*
 61 *relative root-mean-square error.*

Site	Seed	Cost function	Percentage of solutions	R ²	RMSE [g m ⁻²]	RRMSE
CH	1	min_contrast_1	1	0.221	95.683	0.556
CH	2	min_contrast_1	1	0.214	96.421	0.558
CH	3	min_contrast_1	1	0.149	107.756	0.617
CH	4	min_contrast_1	1	0.142	104.529	0.602
CH	5	min_contrast_1	1	0.207	102.279	0.589
CH	6	min_contrast_1	1	0.211	96.402	0.558
CH	7	min_contrast_1	1	0.208	101.18	0.583
CH	8	min_contrast_1	1	0.192	98.848	0.573
CH	9	min_contrast_1	1	0.23	101.751	0.583
CH	10	min_contrast_2	1	0.263	97.142	0.558
ALB	1	neyman_chi_square_divergence	1	0.419	72.6	0.521
ALB	2	neyman_chi_square_divergence	1	0.392	70.695	0.512
ALB	3	neyman_chi_square_divergence	1	0.401	73.617	0.525
ALB	4	neyman_chi_square_divergence	1	0.399	71.542	0.518
ALB	5	neyman_chi_square_divergence	1	0.491	67.975	0.491
ALB	6	neyman_chi_square_divergence	1	0.519	67.184	0.483
ALB	7	neyman_chi_square_divergence	1	0.488	68.575	0.496
ALB	8	neyman_chi_square_divergence	1	0.404	71.385	0.511
ALB	9	neyman_chi_square_divergence	1	0.424	70.498	0.505
ALB	10	neyman_chi_square_divergence	1	0.406	70.858	0.508
HAI	1	neyman_chi_square_divergence	1	0.535	76.289	0.653
HAI	2	neyman_chi_square_divergence	1	0.502	79.265	0.68
HAI	3	neyman_chi_square_divergence	1	0.488	77.544	0.67
HAI	4	neyman_chi_square_divergence	1	0.521	79.296	0.682
HAI	5	neyman_chi_square_divergence	1	0.471	79.608	0.683
HAI	6	neyman_chi_square_divergence	1	0.56	74.294	0.65
HAI	7	neyman_chi_square_divergence	1	0.496	78.714	0.679
HAI	8	neyman_chi_square_divergence	1	0.494	78.26	0.673
HAI	9	neyman_chi_square_divergence	1	0.513	79.514	0.678
HAI	10	neyman_chi_square_divergence	1	0.465	80.354	0.691
SCH	1	laplace_distribution	10	0.232	100.658	0.459
SCH	2	laplace_distribution	10	0.258	99.444	0.451
SCH	3	laplace_distribution	10	0.253	105.421	0.475
SCH	4	laplace_distribution	10	0.252	105.525	0.475
SCH	5	laplace_distribution	10	0.25	99.303	0.451
SCH	6	laplace_distribution	10	0.265	104.778	0.471

SCH	7	laplace_distribution	10	0.237	107.39	0.481
SCH	8	laplace_distribution	10	0.249	105.4	0.473
SCH	9	laplace_distribution	10	0.234	107.052	0.48
SCH	10	laplace_distribution	10	0.235	106.999	0.481
US	1	pearson_chi_square	0.01	0.11	179.487	0.684
US	2	pearson_chi_square	0.01	0.102	173.848	0.67
US	3	min_contrast_4	10	0.411	182.307	0.695
US	4	min_contrast_4	10	0.394	182.016	0.694
US	5	pearson_chi_square	0.01	0.119	176.744	0.674
US	6	pearson_chi_square	0.01	0.101	174.832	0.673
US	7	min_contrast_4	10	0.419	180.943	0.689
US	8	min_contrast_4	10	0.417	181.294	0.692
US	9	min_contrast_4	10	0.434	181.748	0.693
US	10	pearson_chi_square	0.01	0.117	175.263	0.671

62

63

64 **Section A.6: Cross-validation performance of local hybrid models**

65 *Table A.3: Cross-validation performance of local hybrid models. The parameters correspond to the best-performing*
 66 *combination of parameters tested during the fivefold cross-validation using Active Learning with an initial training set size of*
 67 *2%. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States, GPR: Gaussian*
 68 *Process regression, R²: coefficient of determination, RMSE: root-mean-square error, RRMSE: relative root-mean-square error.*
 69 *Nomenclature of parameter names for GPR according to Pedregosa et al. (2012).*

Site	Model	Seed	Parameters	R ²	RMSE [g m ⁻²]	RRMSE
CH	GPR	1	2.21**2 * RBF(length_scale=1.42)	0.499	75.556	0.439
CH	GPR	2	1.38**2 * RBF(length_scale=1.19)	0.452	79.347	0.459
CH	GPR	3	2.3**2 * RBF(length_scale=1.31)	0.352	91.03	0.521
CH	GPR	4	1.65**2 * RBF(length_scale=1.54)	0.391	86.625	0.499
CH	GPR	5	1.4**2 * RBF(length_scale=1.12)	0.443	83.088	0.478
CH	GPR	6	1.35**2 * RBF(length_scale=0.905)	0.294	90.317	0.523
CH	GPR	7	2.17**2 * RBF(length_scale=1.44)	0.454	83.176	0.479
CH	GPR	8	1.86**2 * RBF(length_scale=1.39)	0.528	75.341	0.437
CH	GPR	9	2.01**2 * RBF(length_scale=1.55)	0.373	88.788	0.509
CH	GPR	10	1.88**2 * RBF(length_scale=1.51)	0.416	85.856	0.493
ALB	GPR	1	1.48**2 * RBF(length_scale=1.53)	0.598	54.526	0.391
ALB	GPR	2	1.55**2 * RBF(length_scale=1.18)	0.522	57.823	0.418
ALB	GPR	3	1.71**2 * RBF(length_scale=1.43)	0.59	55.424	0.396
ALB	GPR	4	1.23**2 * RBF(length_scale=1.06)	0.553	55.795	0.404
ALB	GPR	5	1.22**2 * RBF(length_scale=1.05)	0.651	49.486	0.357
ALB	GPR	6	1.24**2 * RBF(length_scale=0.967)	0.662	50.024	0.36
ALB	GPR	7	1.22**2 * RBF(length_scale=0.924)	0.585	54.581	0.395
ALB	GPR	8	1.46**2 * RBF(length_scale=1.77)	0.607	53.931	0.386
ALB	GPR	9	1.38**2 * RBF(length_scale=1.05)	0.605	54.212	0.388
ALB	GPR	10	1.39**2 * RBF(length_scale=1.13)	0.557	57.082	0.409
HAI	GPR	1	1.11**2 * RBF(length_scale=0.986)	0.626	55.158	0.472
HAI	GPR	2	1.61**2 * RBF(length_scale=1.39)	0.567	57.636	0.494
HAI	GPR	3	1.56**2 * RBF(length_scale=1.17)	0.558	58.089	0.502
HAI	GPR	4	1.23**2 * RBF(length_scale=1.22)	0.599	56.084	0.483
HAI	GPR	5	1.18**2 * RBF(length_scale=0.916)	0.558	58.565	0.503
HAI	GPR	6	1.34**2 * RBF(length_scale=0.889)	0.669	46.253	0.405
HAI	GPR	7	1.36**2 * RBF(length_scale=0.949)	0.551	59.404	0.512
HAI	GPR	8	1.22**2 * RBF(length_scale=1.2)	0.593	55.635	0.478
HAI	GPR	9	1.35**2 * RBF(length_scale=0.994)	0.582	58.676	0.5
HAI	GPR	10	1.07**2 * RBF(length_scale=0.729)	0.439	64.907	0.558
SCH	GPR	1	1.43**2 * RBF(length_scale=1.66)	0.42	87.105	0.397
SCH	GPR	2	2.41**2 * RBF(length_scale=1.88)	0.413	87.669	0.398
SCH	GPR	3	1.1**2 * RBF(length_scale=0.739)	0.265	103.892	0.468
SCH	GPR	4	1.67**2 * RBF(length_scale=1.46)	0.42	92.083	0.414
SCH	GPR	5	1.28**2 * RBF(length_scale=1.18)	0.419	88.073	0.4

SCH	GPR	6	1.22**2 * RBF(length_scale=0.963)	0.358	97.48	0.438
SCH	GPR	7	2.21**2 * RBF(length_scale=1.73)	0.347	98.593	0.442
SCH	GPR	8	1.87**2 * RBF(length_scale=1.72)	0.472	87.585	0.393
SCH	GPR	9	1.13**2 * RBF(length_scale=0.843)	0.313	101.202	0.454
SCH	GPR	10	1.5**2 * RBF(length_scale=1.22)	0.422	95.264	0.428
US	GPR	1	1**2 * RBF(length_scale=0.15)	0	87.721	0.334
US	GPR	2	1**2 * RBF(length_scale=0.0601)	0	78.583	0.303
US	GPR	3	1**2 * RBF(length_scale=0.107)	0	87.481	0.334
US	GPR	4	2.88**2 * RBF(length_scale=1.72)	0.522	62.153	0.237
US	GPR	5	1.73**2 * RBF(length_scale=1.58)	0.523	60.741	0.232
US	GPR	6	1**2 * RBF(length_scale=0.215)	0	78.953	0.304
US	GPR	7	2.27**2 * RBF(length_scale=1.37)	0.652	53.078	0.202
US	GPR	8	1**2 * RBF(length_scale=0.371)	0	87.852	0.335
US	GPR	9	1.67**2 * RBF(length_scale=1.35)	0.652	52.129	0.199
US	GPR	10	1**2 * RBF(length_scale=0.119)	0	85.523	0.328

71 **Section A.7: Cross-validation performance of global empirical models**

72 *Table A.4: Cross-validation performance of global empirical models. The parameters correspond to the best-performing*
 73 *combination of parameters tested during the fivefold cross-validation. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-*
 74 *Dün, SCH: Schorfheide-Chorin, US: United States, RFR: Random Forest regression, SVR: Support Vector regression, XGB:*
 75 *Extreme Gradient Boosting regression, GPR: Gaussian Process regression, R²: coefficient of determination, RMSE: root-mean-*
 76 *square error, RRMSE: relative root-mean-square error. Nomenclature of parameter names for RFR, SVR, and GPR according*
 77 *to Pedregosa et al. (2012), for XGB according to Chen and Guestrin (2016).*

Site	Model	Parameters	R ²	RMSE [g m ⁻²]	RRMSE
All except for CH	RFR	'max_depth': None, 'max_features': 'sqrt', 'max_leaf_nodes': None, 'min_samples_leaf': 2, 'n_estimators': 200	0.506	79.86	0.433
All except for CH	SVR	'svr_C': 100, 'svr_epsilon': 0.1, 'svr_gamma': 'scale', 'svr_kernel': 'rbf'	0.522	78.95	0.427
All except for CH	XGB	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0.1, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 3, 'xgb_n_estimators': 300, 'xgb_subsample': 0.8	0.5	80.664	0.437
All except for CH	GPR	1**2 * RBF(length_scale=0.0279)	0.398	88.272	0.478
All except for ALB	RFR	'max_depth': 10, 'max_features': 'sqrt', 'max_leaf_nodes': None, 'min_samples_leaf': 1, 'n_estimators': 100	0.399	90.335	0.467
All except for ALB	SVR	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'rbf'	0.379	92.708	0.479
All except for ALB	XGB	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0.2, 'xgb_learning_rate': 0.1, 'xgb_max_depth': 9, 'xgb_n_estimators': 100, 'xgb_subsample': 0.8	0.377	93.213	0.482
All except for ALB	GPR	1**2 * RBF(length_scale=0.0285)	0.083	109.705	0.567
All except for HAI	RFR	'max_depth': 10, 'max_features': 'sqrt', 'max_leaf_nodes': None, 'min_samples_leaf': 2, 'n_estimators': 500	0.328	91.829	0.462
All except for HAI	SVR	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'rbf'	0.359	90.446	0.454
All except for HAI	XGB	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 3, 'xgb_n_estimators': 300, 'xgb_subsample': 0.8	0.294	94.275	0.474
All except for HAI	GPR	1**2 * RBF(length_scale=0.0297)	0.001	111.558	0.562
All except for SCH	RFR	'max_depth': 10, 'max_features': 10, 'max_leaf_nodes': None, 'min_samples_leaf': 1, 'n_estimators': 500	0.49	77.84	0.449
All except for SCH	SVR	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'scale', 'svr_kernel': 'rbf'	0.439	82.674	0.477
All except for SCH	XGB	'xgb_colsample_bytree': 1.0, 'xgb_gamma': 0.2, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 7, 'xgb_n_estimators': 300, 'xgb_subsample': 0.8	0.463	79.584	0.459
All except for SCH	GPR	1**2 * RBF(length_scale=0.0273)	0.298	90.679	0.525
All except for US	RFR	'max_depth': 15, 'max_features': 'sqrt', 'max_leaf_nodes': None, 'min_samples_leaf': 2, 'n_estimators': 200	0.349	89.073	0.547
All except for US	SVR	'svr_C': 100, 'svr_epsilon': 1, 'svr_gamma': 'auto', 'svr_kernel': 'rbf'	0.366	89.059	0.546
All except for US	XGB	'xgb_colsample_bytree': 0.8, 'xgb_gamma': 0, 'xgb_learning_rate': 0.01, 'xgb_max_depth': 3, 'xgb_n_estimators': 300, 'xgb_subsample': 0.8	0.324	90.521	0.556
All except for US	GPR	1**2 * RBF(length_scale=0.0292)	0.161	101.078	0.62

78

79 **Section A.8: Cross-validation performance of global physically-based models**

80 *Table A.5: Cross-validation performance of global physically-based models. The parameters correspond to the best-performing*
 81 *combination of parameters tested during the fivefold cross-validation. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-*
 82 *Dün, SCH: Schorfheide-Chorin, US: United States, R²: coefficient of determination, RMSE: root-mean-square error, RRMSE:*
 83 *relative root-mean-square error.*

Site	Cost function	Percentage of solutions	R ²	RMSE [g m ⁻²]	RRMSE
All except for CH	laplace_distribution	1	0.034	120.649	0.653
All except for ALB	min_contrast_4	10	0.025	126.039	0.652
All except for HAI	min_contrast_4	10	0.025	124.348	0.625
All except for SCH	laplace_distribution	1	0.009	119.915	0.695
All except for US	min_contrast_1	1	0.266	93.696	0.576

84

85

86 Section A.9: Cross-validation performance of global hybrid models

87 Table A.6: Cross-validation performance of global hybrid models. The parameters correspond to the best-performing
88 combination of parameters tested during the fivefold cross-validation using Active Learning with an initial training set size of
89 2%. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States, GPR: Gaussian
90 Process regression, R^2 : coefficient of determination, RMSE: root-mean-square error, RRMSE: relative root-mean-square error.
91 Nomenclature of parameter names for GPR according to Pedregosa et al. (2012).

Site	Model	Parameters	R^2	RMSE [g m^{-3}]	RRMSE
All except for CH	GPR	$1.39^{**2} * \text{RBF}(\text{length_scale}=1.29)$	0.427	85.269	0.462
All except for ALB	GPR	$1.82^{**2} * \text{RBF}(\text{length_scale}=1.41)$	0.31	95.682	0.495
All except for HAI	GPR	$1.26^{**2} * \text{RBF}(\text{length_scale}=1.11)$	0.237	97.987	0.492
All except for SCH	GPR	$1.63^{**2} * \text{RBF}(\text{length_scale}=1.26)$	0.368	86.338	0.5
All except for US	GPR	$1.16^{**2} * \text{RBF}(\text{length_scale}=1.09)$	0.332	89.463	0.55

92

93

94 **Section A.10: Testing performance of local and transferred empirical models: RFR**

95 *Table A.7: Testing performance of local and transferred empirical Random Forest regression (RFR) models. The parameters*
 96 *correspond to the best-performing combination of parameters tested during the fivefold cross-validation. CH: Switzerland,*
 97 *ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States, R²: coefficient of determination, RMSE:*
 98 *root-mean-square error, sRMSE: systematic component of RMSE, uRMSE: unsystematic component of RMSE, RRMSSE: relative*
 99 *root-mean-square error, MBE: mean bias error.*

		Training site				
Prediction site	Metric	CH	ALB	HAI	SCH	US
CH	R ²	0.30 ± 0.14	0.15 ± 0.02	0.01 ± 0.01	0.15 ± 0.02	0.01 ± 0.01
	RMSE	97.67 ± 14.48	120.81 ± 2.34	142.25 ± 6.22	112.16 ± 5.87	231.97 ± 32.68
	sRMSE	113.35 ± 9.02	127.81 ± 2.37	140.14 ± 6.24	112.57 ± 1.81	219.95 ± 26.43
	uRMSE	59.82 ± 8.76	41.61 ± 3.19	23.47 ± 6.62	29.93 ± 6.00	71.03 ± 27.47
	RRMSE	0.56 ± 0.08	0.70 ± 0.01	0.82 ± 0.04	0.65 ± 0.03	1.34 ± 0.19
	MBE	-1.62 ± 14.43	-65.16 ± 4.60	-86.25 ± 10.08	22.32 ± 9.96	189.73 ± 30.65
ALB	R ²	0.28 ± 0.08	0.51 ± 0.17	0.34 ± 0.02	0.49 ± 0.03	0.30 ± 0.02
	RMSE	126.00 ± 13.26	61.60 ± 10.74	73.85 ± 1.20	104.79 ± 8.47	260.84 ± 43.11
	sRMSE	132.69 ± 11.92	85.05 ± 4.28	87.37 ± 0.61	116.88 ± 4.64	260.01 ± 38.75
	uRMSE	40.51 ± 7.35	55.77 ± 15.19	46.59 ± 2.52	50.43 ± 9.42	46.06 ± 12.45
	RRMSE	0.91 ± 0.10	0.44 ± 0.07	0.53 ± 0.01	0.75 ± 0.06	1.87 ± 0.31
	MBE	101.83 ± 15.73	3.44 ± 12.53	-22.47 ± 2.40	80.84 ± 6.72	245.55 ± 41.03
HAI	R ²	0.32 ± 0.07	0.45 ± 0.02	0.64 ± 0.11	0.43 ± 0.05	0.22 ± 0.05
	RMSE	138.66 ± 14.88	63.99 ± 1.42	50.73 ± 12.60	101.03 ± 4.03	293.26 ± 41.64
	sRMSE	145.82 ± 14.20	85.35 ± 0.16	82.50 ± 9.68	114.20 ± 2.13	292.95 ± 39.55
	uRMSE	44.38 ± 6.74	56.44 ± 1.60	64.28 ± 5.88	52.93 ± 4.83	32.06 ± 12.20
	RRMSE	1.20 ± 0.13	0.55 ± 0.01	0.44 ± 0.10	0.87 ± 0.03	2.53 ± 0.36
	MBE	117.78 ± 17.85	4.41 ± 2.84	-4.10 ± 6.45	76.01 ± 3.20	279.96 ± 41.41
SCH	R ²	0.19 ± 0.06	0.31 ± 0.02	0.16 ± 0.02	0.32 ± 0.09	0.09 ± 0.01
	RMSE	114.93 ± 5.97	123.22 ± 1.85	139.90 ± 1.74	103.30 ± 10.26	227.35 ± 45.71
	sRMSE	124.93 ± 5.18	138.81 ± 1.69	147.01 ± 0.98	120.49 ± 11.52	226.85 ± 41.93
	uRMSE	48.12 ± 8.52	63.90 ± 1.65	44.88 ± 4.71	60.01 ± 16.54	39.01 ± 12.37
	RRMSE	0.52 ± 0.03	0.56 ± 0.01	0.63 ± 0.01	0.46 ± 0.04	1.02 ± 0.21
	MBE	33.07 ± 20.17	-71.83 ± 3.28	-86.43 ± 1.67	0.70 ± 11.84	191.27 ± 49.71
US	R ²	0.16 ± 0.14	0.32 ± 0.08	0.17 ± 0.16	0.28 ± 0.09	0.47 ± 0.16
	RMSE	173.22 ± 9.53	225.50 ± 3.77	226.10 ± 2.65	173.05 ± 8.70	64.73 ± 11.41
	sRMSE	175.60 ± 9.71	227.59 ± 3.44	226.95 ± 3.07	176.04 ± 8.23	84.35 ± 12.85
	uRMSE	26.53 ± 12.43	30.41 ± 4.20	20.87 ± 6.78	31.94 ± 3.65	51.43 ± 17.73
	RRMSE	0.66 ± 0.04	0.86 ± 0.01	0.86 ± 0.01	0.66 ± 0.03	0.25 ± 0.04
	MBE	-153.32 ± 11.21	-211.03 ± 3.71	-210.25 ± 3.37	-153.95 ± 9.37	0.12 ± 11.91

100

101

102 **Section A.11: Testing performance of local and transferred empirical models: SVR**

103 *Table A.8: Testing performance of local and transferred empirical Support Vector regression (SVR) models. The parameters*
 104 *correspond to the best-performing combination of parameters tested during the fivefold cross-validation. CH: Switzerland,*
 105 *ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States, R²: coefficient of determination, RMSE:*
 106 *root-mean-square error, sRMSE: systematic component of RMSE, uRMSE: unsystematic component of RMSE, RRMSE: relative*
 107 *root-mean-square error, MBE: mean bias error.*

		Training site				
Prediction site	Metric	CH	ALB	HAI	SCH	US
CH	R ²	0.31 ± 0.12	0.26 ± 0.13	0.01 ± 0.01	0.11 ± 0.16	0.06 ± 0.04
	RMSE	96.29 ± 6.44	149.67 ± 19.71	195.16 ± 16.16	132.08 ± 27.73	293.03 ± 57.49
	sRMSE	113.92 ± 9.04	158.61 ± 23.47	185.72 ± 17.25	132.78 ± 33.06	276.07 ± 48.37
	uRMSE	58.90 ± 16.63	53.39 ± 16.13	57.92 ± 14.39	33.89 ± 17.02	94.87 ± 42.37
	RRMSE	0.55 ± 0.03	0.86 ± 0.11	1.12 ± 0.09	0.76 ± 0.16	1.69 ± 0.33
	MBE	-13.29 ± 13.09	-111.13 ± 36.03	-148.94 ± 20.88	-33.73 ± 73.95	251.96 ± 53.69
ALB	R ²	0.46 ± 0.12	0.54 ± 0.21	0.47 ± 0.02	0.49 ± 0.04	0.25 ± 0.16
	RMSE	97.08 ± 16.87	59.48 ± 13.07	74.84 ± 3.45	89.75 ± 5.74	280.34 ± 30.59
	sRMSE	109.51 ± 15.79	85.21 ± 4.86	93.83 ± 2.46	106.39 ± 4.62	280.43 ± 29.37
	uRMSE	55.63 ± 2.48	59.73 ± 12.97	56.52 ± 1.47	57.02 ± 0.99	38.65 ± 18.03
	RRMSE	0.70 ± 0.12	0.43 ± 0.09	0.54 ± 0.02	0.64 ± 0.04	2.01 ± 0.22
	MBE	67.46 ± 23.94	-6.35 ± 12.74	-40.95 ± 5.37	64.65 ± 7.68	267.23 ± 30.84
HAI	R ²	0.39 ± 0.04	0.49 ± 0.05	0.65 ± 0.08	0.48 ± 0.05	0.33 ± 0.15
	RMSE	107.06 ± 17.05	62.67 ± 2.62	52.62 ± 11.00	88.42 ± 5.75	281.21 ± 27.60
	sRMSE	115.94 ± 14.23	85.91 ± 0.48	83.64 ± 10.87	105.76 ± 3.62	283.83 ± 26.66
	uRMSE	51.46 ± 5.24	58.60 ± 3.47	64.67 ± 6.39	57.80 ± 2.59	42.84 ± 10.99
	RRMSE	0.92 ± 0.15	0.54 ± 0.02	0.45 ± 0.08	0.76 ± 0.05	2.42 ± 0.24
	MBE	77.08 ± 20.47	-10.15 ± 5.00	-13.82 ± 9.06	62.53 ± 6.07	270.58 ± 27.97
SCH	R ²	0.33 ± 0.08	0.35 ± 0.06	0.24 ± 0.02	0.42 ± 0.15	0.18 ± 0.07
	RMSE	108.79 ± 10.83	129.53 ± 3.18	139.75 ± 2.47	94.39 ± 8.53	224.30 ± 24.65
	sRMSE	122.02 ± 2.28	146.39 ± 2.76	150.82 ± 1.61	120.78 ± 11.39	226.29 ± 23.89
	uRMSE	60.96 ± 4.19	67.90 ± 6.16	56.62 ± 2.66	73.64 ± 17.66	30.70 ± 17.05
	RRMSE	0.49 ± 0.05	0.58 ± 0.01	0.63 ± 0.01	0.42 ± 0.04	1.01 ± 0.11
	MBE	-14.42 ± 23.94	-85.56 ± 4.82	-92.86 ± 2.62	-8.22 ± 12.60	191.96 ± 28.19
US	R ²	0.30 ± 0.03	0.27 ± 0.10	0.32 ± 0.05	0.23 ± 0.05	0.59 ± 0.08
	RMSE	212.61 ± 50.25	259.17 ± 42.80	272.06 ± 35.13	206.50 ± 124.22	56.89 ± 12.09
	sRMSE	216.26 ± 49.34	261.03 ± 44.64	275.78 ± 35.18	201.94 ± 127.46	84.09 ± 12.92
	uRMSE	37.24 ± 10.22	38.86 ± 11.65	44.91 ± 4.87	36.01 ± 10.46	61.61 ± 7.79
	RRMSE	0.81 ± 0.19	0.99 ± 0.16	1.04 ± 0.13	0.79 ± 0.47	0.22 ± 0.04
	MBE	-197.74 ± 53.27	-245.99 ± 48.23	-262.03 ± 37.05	-173.23 ± 140.11	5.06 ± 9.42

108

109

110 **Section A.12: Testing performance of local and transferred empirical models: XGB**

111 *Table A.9: Testing performance of local and transferred empirical Extreme Gradient Boosting (XGB) regression models. The*
 112 *parameters correspond to the best-performing combination of parameters tested during the fivefold cross-validation. CH:*
 113 *Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States, R²: coefficient of*
 114 *determination, RMSE: root-mean-square error, sRMSE: systematic component of RMSE, uRMSE: unsystematic component of*
 115 *RMSE, RRMSE: relative root-mean-square error, MBE: mean bias error.*

		Training site				
Prediction site	Metric	CH	ALB	HAI	SCH	US
CH	R ²	0.21 ± 0.12	0.14 ± 0.04	0.03 ± 0.02	0.12 ± 0.02	0.01 ± 0.01
	RMSE	108.91 ± 14.41	120.79 ± 3.77	135.00 ± 7.13	118.44 ± 9.55	245.19 ± 28.16
	sRMSE	114.75 ± 9.07	126.57 ± 3.28	122.35 ± 8.47	115.46 ± 4.44	226.56 ± 20.67
	uRMSE	45.54 ± 17.66	36.60 ± 9.39	55.24 ± 13.55	32.39 ± 14.39	92.17 ± 25.75
	RRMSE	0.62 ± 0.08	0.70 ± 0.02	0.78 ± 0.04	0.68 ± 0.05	1.41 ± 0.16
	MBE	-4.43 ± 22.19	-62.52 ± 6.61	-50.27 ± 19.49	32.21 ± 15.07	197.58 ± 23.88
ALB	R ²	0.26 ± 0.08	0.52 ± 0.14	0.27 ± 0.04	0.43 ± 0.05	0.29 ± 0.04
	RMSE	160.89 ± 25.32	62.46 ± 9.74	80.56 ± 5.19	110.64 ± 11.96	276.58 ± 47.77
	sRMSE	160.00 ± 25.02	85.18 ± 4.05	85.99 ± 1.05	119.77 ± 6.70	271.73 ± 40.69
	uRMSE	30.33 ± 12.62	55.91 ± 12.22	31.39 ± 11.26	44.40 ± 11.66	48.92 ± 35.67
	RRMSE	1.16 ± 0.18	0.45 ± 0.07	0.58 ± 0.04	0.79 ± 0.09	1.99 ± 0.34
	MBE	134.91 ± 29.78	5.48 ± 12.68	-15.09 ± 5.91	84.81 ± 9.27	257.94 ± 42.81
HAI	R ²	0.28 ± 0.10	0.42 ± 0.03	0.60 ± 0.12	0.39 ± 0.07	0.21 ± 0.07
	RMSE	171.69 ± 29.31	66.22 ± 2.31	54.35 ± 12.33	104.55 ± 5.87	309.70 ± 49.19
	sRMSE	170.75 ± 27.39	85.31 ± 0.16	82.40 ± 9.71	115.71 ± 2.53	305.52 ± 44.89
	uRMSE	29.35 ± 16.82	53.67 ± 2.76	60.99 ± 7.58	48.62 ± 8.13	48.84 ± 29.02
	RRMSE	1.48 ± 0.25	0.57 ± 0.02	0.47 ± 0.09	0.90 ± 0.05	2.67 ± 0.42
	MBE	146.87 ± 32.43	1.04 ± 4.03	-1.05 ± 6.02	78.21 ± 3.66	293.04 ± 46.78
SCH	R ²	0.15 ± 0.05	0.30 ± 0.03	0.16 ± 0.03	0.28 ± 0.11	0.08 ± 0.02
	RMSE	143.38 ± 24.47	122.21 ± 1.83	138.84 ± 3.83	105.79 ± 10.74	245.25 ± 54.24
	sRMSE	144.35 ± 18.33	137.27 ± 1.59	145.15 ± 3.27	120.40 ± 11.45	240.77 ± 50.72
	uRMSE	41.01 ± 15.05	62.41 ± 3.65	42.00 ± 4.92	54.61 ± 18.41	42.26 ± 27.55
	RRMSE	0.65 ± 0.11	0.55 ± 0.01	0.63 ± 0.02	0.48 ± 0.04	1.11 ± 0.24
	MBE	73.97 ± 39.45	-68.78 ± 3.22	-83.11 ± 5.78	1.31 ± 10.30	207.36 ± 58.12
US	R ²	0.11 ± 0.07	0.01 ± 0.01	0.09 ± 0.11	0.20 ± 0.10	0.48 ± 0.14
	RMSE	188.78 ± 15.31	230.54 ± 10.24	218.55 ± 8.05	161.57 ± 10.24	70.77 ± 16.44
	sRMSE	190.10 ± 14.80	229.73 ± 9.88	218.53 ± 8.71	164.25 ± 9.59	85.74 ± 12.11
	uRMSE	28.13 ± 7.24	18.67 ± 5.59	25.85 ± 6.50	27.59 ± 10.06	57.22 ± 15.89
	RRMSE	0.72 ± 0.06	0.88 ± 0.04	0.84 ± 0.03	0.62 ± 0.04	0.27 ± 0.06
	MBE	-169.64 ± 16.38	-213.05 ± 10.63	-201.01 ± 9.52	-140.22 ± 11.11	8.79 ± 16.87

116

117

118 **Section A.13: Testing performance of local and transferred empirical models: GPR**

119 *Table A.10: Testing performance of local and transferred empirical Gaussian Process regression (GPR) models. The parameters*
 120 *correspond to the best-performing combination of parameters tested during the fivefold cross-validation. CH: Switzerland,*
 121 *ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States, R²: coefficient of determination, RMSE:*
 122 *root-mean-square error, sRMSE: systematic component of RMSE, uRMSE: unsystematic component of RMSE, RRMSE: relative*
 123 *root-mean-square error, MBE: mean bias error.*

		Training site				
Prediction site	Metric	CH	ALB	HAI	SCH	US
CH	R ²	0.08 ± 0.05	0.04 ± 0.02	0.00 ± 0.01	0.01 ± 0.01	0.01 ± 0.00
	RMSE	111.15 ± 11.89	119.06 ± 0.86	127.48 ± 0.77	120.40 ± 0.54	141.09 ± 0.68
	sRMSE	112.90 ± 8.85	116.71 ± 0.31	126.56 ± 0.30	119.92 ± 0.44	141.15 ± 0.67
	uRMSE	29.83 ± 7.91	23.29 ± 3.44	14.59 ± 4.63	10.04 ± 6.02	4.10 ± 0.51
	RRMSE	0.64 ± 0.06	0.69 ± 0.00	0.73 ± 0.00	0.69 ± 0.00	0.81 ± 0.00
	MBE	-2.56 ± 6.57	-37.63 ± 0.98	-62.02 ± 0.58	47.10 ± 1.32	87.98 ± 1.09
ALB	R ²	0.08 ± 0.05	0.40 ± 0.17	0.30 ± 0.03	0.16 ± 0.06	0.03 ± 0.01
	RMSE	120.96 ± 8.61	68.44 ± 10.35	73.29 ± 2.35	117.93 ± 3.14	149.15 ± 0.93
	sRMSE	106.10 ± 4.70	84.25 ± 4.60	85.62 ± 0.56	121.40 ± 3.34	149.21 ± 0.91
	uRMSE	57.35 ± 11.71	49.63 ± 11.75	44.02 ± 4.02	27.76 ± 8.15	4.34 ± 1.28
	RRMSE	0.87 ± 0.06	0.49 ± 0.07	0.53 ± 0.02	0.85 ± 0.02	1.07 ± 0.01
	MBE	62.85 ± 7.90	3.40 ± 3.12	-13.74 ± 3.73	87.03 ± 4.69	122.53 ± 1.11
HAI	R ²	0.10 ± 0.02	0.31 ± 0.02	0.54 ± 0.17	0.18 ± 0.08	0.02 ± 0.00
	RMSE	139.71 ± 10.59	72.53 ± 1.41	58.39 ± 15.42	119.74 ± 6.59	169.59 ± 1.21
	sRMSE	124.83 ± 6.69	85.96 ± 0.43	83.06 ± 9.53	124.76 ± 4.28	169.53 ± 1.13
	uRMSE	62.36 ± 10.70	46.09 ± 1.54	57.05 ± 9.37	32.59 ± 11.80	4.80 ± 2.47
	RRMSE	1.20 ± 0.09	0.63 ± 0.01	0.50 ± 0.12	1.03 ± 0.06	1.46 ± 0.01
	MBE	90.08 ± 9.29	9.95 ± 3.68	-3.37 ± 11.12	90.80 ± 5.68	146.09 ± 1.31
SCH	R ²	0.03 ± 0.04	0.13 ± 0.01	0.16 ± 0.04	0.16 ± 0.11	0.01 ± 0.00
	RMSE	130.30 ± 5.04	132.47 ± 1.17	139.77 ± 3.58	114.35 ± 13.95	125.86 ± 0.39
	sRMSE	122.19 ± 1.57	138.90 ± 1.49	147.30 ± 2.73	120.78 ± 11.53	125.90 ± 0.36
	uRMSE	41.31 ± 19.05	41.74 ± 2.03	46.04 ± 6.08	42.75 ± 15.76	3.52 ± 1.53
	RRMSE	0.59 ± 0.02	0.60 ± 0.01	0.63 ± 0.02	0.51 ± 0.05	0.57 ± 0.00
	MBE	-24.84 ± 7.79	-71.68 ± 2.96	-86.85 ± 4.66	0.83 ± 12.62	39.83 ± 1.14
US	R ²	0.08 ± 0.05	0.04 ± 0.02	0.16 ± 0.04	0.09 ± 0.03	0.35 ± 0.07
	RMSE	124.71 ± 1.27	150.08 ± 0.78	170.72 ± 0.70	94.63 ± 0.49	71.05 ± 13.37
	sRMSE	123.64 ± 0.65	150.07 ± 0.70	170.33 ± 0.57	94.82 ± 0.49	84.52 ± 12.87
	uRMSE	14.83 ± 6.96	4.60 ± 1.81	11.10 ± 3.36	5.57 ± 2.35	45.13 ± 6.76
	RRMSE	0.48 ± 0.00	0.57 ± 0.00	0.65 ± 0.00	0.36 ± 0.00	0.27 ± 0.05
	MBE	-88.80 ± 0.88	-123.02 ± 0.86	-146.93 ± 0.65	-39.81 ± 1.23	2.37 ± 12.21

124

125 **Section A.14: Testing performance of local and transferred physically-based models**

126 *Table A.11: Testing performance of local and transferred physically-based models. The parameters correspond to the best-*
 127 *performing combination of parameters tested during the fivefold cross-validation. CH: Switzerland, ALB: Schwäbische Alb,*
 128 *HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States, R²: coefficient of determination, RMSE: root-mean-square error,*
 129 *sRMSE: systematic component of RMSE, uRMSE: unsystematic component of RMSE, RRMSE: relative root-mean-square error,*
 130 *MBE: mean bias error.*

		Calibration site				
Prediction site	Metric	CH	ALB	HAI	SCH	US
CH	R ²	0.26 ± 0.17	0.14 ± 0.00	0.14 ± 0.00	0.12 ± 0.00	0.13 ± 0.00
	RMSE	102.57 ± 16.64	105.26 ± 0.00	105.26 ± 0.00	104.55 ± 0.00	111.23 ± 6.82
	sRMSE	115.97 ± 11.28	112.62 ± 0.00	112.62 ± 0.00	109.99 ± 0.00	111.71 ± 1.21
	uRMSE	56.88 ± 12.05	40.06 ± 0.00	40.06 ± 0.00	34.17 ± 0.00	35.29 ± 0.87
	RRMSE	0.59 ± 0.09	0.61 ± 0.00	0.61 ± 0.00	0.60 ± 0.00	0.64 ± 0.04
	MBE	-24.33 ± 15.87	-24.61 ± 0.00	-24.61 ± 0.00	-3.44 ± 0.00	-18.58 ± 7.28
ALB	R ²	0.41 ± 0.00	0.43 ± 0.17	0.43 ± 0.00	0.45 ± 0.00	0.22 ± 0.21
	RMSE	74.27 ± 0.25	73.17 ± 7.39	71.13 ± 0.00	87.55 ± 0.00	95.30 ± 9.71
	sRMSE	91.30 ± 0.15	90.63 ± 4.41	89.28 ± 0.00	103.76 ± 0.00	94.25 ± 7.34
	uRMSE	53.10 ± 0.09	51.44 ± 13.34	53.95 ± 0.00	55.69 ± 0.00	56.83 ± 2.11
	RRMSE	0.53 ± 0.00	0.52 ± 0.05	0.51 ± 0.00	0.63 ± 0.00	0.68 ± 0.07
	MBE	34.97 ± 0.37	32.34 ± 9.65	29.33 ± 0.00	60.49 ± 0.00	19.05 ± 37.63
HAI	R ²	0.53 ± 0.00	0.53 ± 0.00	0.66 ± 0.10	0.50 ± 0.00	0.26 ± 0.23
	RMSE	80.00 ± 0.20	76.46 ± 0.00	68.05 ± 7.50	98.54 ± 0.00	97.42 ± 2.45
	sRMSE	99.66 ± 0.12	96.69 ± 0.00	93.07 ± 7.16	115.10 ± 0.00	99.09 ± 12.78
	uRMSE	59.42 ± 0.06	59.18 ± 0.00	63.13 ± 6.46	59.47 ± 0.00	54.68 ± 4.44
	RRMSE	0.69 ± 0.00	0.66 ± 0.00	0.59 ± 0.05	0.85 ± 0.00	0.84 ± 0.02
	MBE	51.81 ± 0.23	45.83 ± 0.00	42.64 ± 7.51	77.44 ± 0.00	41.44 ± 31.11
SCH	R ²	0.21 ± 0.00	0.20 ± 0.00	0.20 ± 0.00	0.30 ± 0.04	0.12 ± 0.12
	RMSE	114.19 ± 0.09	116.67 ± 0.00	116.67 ± 0.00	103.95 ± 11.54	133.47 ± 27.96
	sRMSE	125.42 ± 0.08	127.32 ± 0.00	127.32 ± 0.00	120.51 ± 11.65	130.78 ± 11.16
	uRMSE	51.87 ± 0.01	50.98 ± 0.00	50.98 ± 0.00	60.75 ± 5.32	66.61 ± 10.26
	RRMSE	0.51 ± 0.00	0.53 ± 0.00	0.53 ± 0.00	0.47 ± 0.04	0.60 ± 0.13
	MBE	-40.07 ± 0.25	-45.64 ± 0.00	-45.64 ± 0.00	-9.99 ± 6.58	-45.40 ± 31.62
US	R ²	0.45 ± 0.00	0.44 ± 0.00	0.44 ± 0.00	0.43 ± 0.00	0.30 ± 0.24
	RMSE	196.60 ± 1.42	196.25 ± 0.00	196.25 ± 0.00	184.59 ± 0.00	185.01 ± 11.81
	sRMSE	201.31 ± 1.51	201.00 ± 0.00	201.00 ± 0.00	190.55 ± 0.00	187.33 ± 10.73
	uRMSE	43.27 ± 0.53	43.42 ± 0.00	43.42 ± 0.00	47.29 ± 0.00	36.63 ± 5.86
	RRMSE	0.75 ± 0.01	0.75 ± 0.00	0.75 ± 0.00	0.71 ± 0.00	0.71 ± 0.04
	MBE	-182.43 ± 1.66	-182.09 ± 0.00	-182.09 ± 0.00	-170.48 ± 0.00	-167.32 ± 6.38

131

132 **Section A.15: Testing performance of local and transferred hybrid models**

133 *Table A.12: Testing performance of local and transferred hybrid models. The parameters correspond to the best-performing*
 134 *combination of parameters tested during the fivefold cross-validation using Active Learning with an initial training set size of*
 135 *2%. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States, R²: coefficient of*
 136 *determination, RMSE: root-mean-square error, sRMSE: systematic component of RMSE, uRMSE: unsystematic component of*
 137 *RMSE, RRMSE: relative root-mean-square error, MBE: mean bias error.*

		Calibration site				
Prediction site	Metric	CH	ALB	HAI	SCH	US
CH	R ²	0.23 ± 0.14	0.02 ± 0.02	0.02 ± 0.03	0.03 ± 0.03	0.01 ± 0.01
	RMSE	112.22 ± 31.13	127.07 ± 7.94	125.43 ± 9.89	136.69 ± 13.00	164.02 ± 35.27
	sRMSE	113.87 ± 9.18	116.51 ± 7.64	117.34 ± 6.45	125.35 ± 13.19	146.02 ± 14.56
	uRMSE	54.73 ± 39.46	49.79 ± 13.76	44.33 ± 15.06	53.38 ± 17.36	58.53 ± 64.41
	RRMSE	0.64 ± 0.18	0.73 ± 0.05	0.72 ± 0.06	0.79 ± 0.07	0.94 ± 0.20
	MBE	-3.46 ± 17.01	-31.66 ± 21.59	-36.42 ± 18.12	36.96 ± 48.54	93.61 ± 20.14
ALB	R ²	0.07 ± 0.07	0.46 ± 0.18	0.34 ± 0.07	0.41 ± 0.09	0.03 ± 0.02
	RMSE	122.05 ± 8.44	64.39 ± 11.93	72.30 ± 4.12	107.21 ± 8.09	158.60 ± 28.14
	sRMSE	103.03 ± 10.46	84.85 ± 4.76	86.83 ± 1.25	118.81 ± 4.89	147.49 ± 33.86
	uRMSE	61.51 ± 21.43	54.13 ± 12.30	47.60 ± 5.58	50.35 ± 6.74	42.55 ± 41.87
	RRMSE	0.88 ± 0.06	0.46 ± 0.08	0.52 ± 0.03	0.77 ± 0.06	1.14 ± 0.20
	MBE	55.91 ± 19.27	-0.37 ± 11.47	-19.74 ± 5.34	83.55 ± 6.82	117.33 ± 42.37
HAI	R ²	0.03 ± 0.03	0.32 ± 0.04	0.54 ± 0.11	0.25 ± 0.06	0.02 ± 0.03
	RMSE	131.67 ± 12.72	72.81 ± 1.65	58.51 ± 11.47	110.02 ± 5.09	179.92 ± 31.30
	sRMSE	112.79 ± 15.17	86.15 ± 0.83	83.22 ± 9.76	115.72 ± 4.17	171.60 ± 35.17
	uRMSE	65.35 ± 16.62	45.90 ± 3.40	58.49 ± 6.67	35.08 ± 7.02	40.01 ± 41.06
	RRMSE	1.13 ± 0.11	0.63 ± 0.01	0.51 ± 0.09	0.95 ± 0.04	1.55 ± 0.27
	MBE	70.71 ± 24.06	0.42 ± 12.18	-9.55 ± 8.77	78.01 ± 6.21	146.89 ± 38.34
SCH	R ²	0.02 ± 0.02	0.23 ± 0.04	0.17 ± 0.07	0.27 ± 0.15	0.02 ± 0.03
	RMSE	142.90 ± 12.30	129.56 ± 8.21	144.81 ± 6.85	104.35 ± 14.22	144.01 ± 25.57
	sRMSE	129.70 ± 11.57	141.33 ± 6.42	152.09 ± 3.38	120.30 ± 11.46	132.76 ± 16.64
	uRMSE	58.91 ± 12.09	56.03 ± 4.71	44.91 ± 12.62	56.42 ± 18.15	46.69 ± 44.79
	RRMSE	0.64 ± 0.06	0.58 ± 0.04	0.65 ± 0.03	0.47 ± 0.06	0.65 ± 0.12
	MBE	-44.35 ± 27.74	-75.77 ± 12.38	-94.74 ± 5.45	-0.14 ± 8.90	39.24 ± 43.46
US	R ²	0.13 ± 0.10	0.12 ± 0.10	0.14 ± 0.11	0.19 ± 0.05	0.23 ± 0.26
	RMSE	184.27 ± 27.00	161.10 ± 38.01	155.40 ± 9.64	118.22 ± 23.46	76.40 ± 20.77
	sRMSE	171.80 ± 24.72	159.96 ± 38.99	148.05 ± 10.26	106.73 ± 19.36	84.04 ± 13.67
	uRMSE	63.41 ± 27.27	33.24 ± 13.22	45.41 ± 12.44	52.36 ± 23.27	22.45 ± 24.84
	RRMSE	0.70 ± 0.10	0.62 ± 0.15	0.59 ± 0.04	0.45 ± 0.09	0.29 ± 0.07
	MBE	-148.09 ± 28.52	-132.65 ± 46.57	-120.22 ± 12.54	-48.41 ± 44.96	0.49 ± 7.92

138

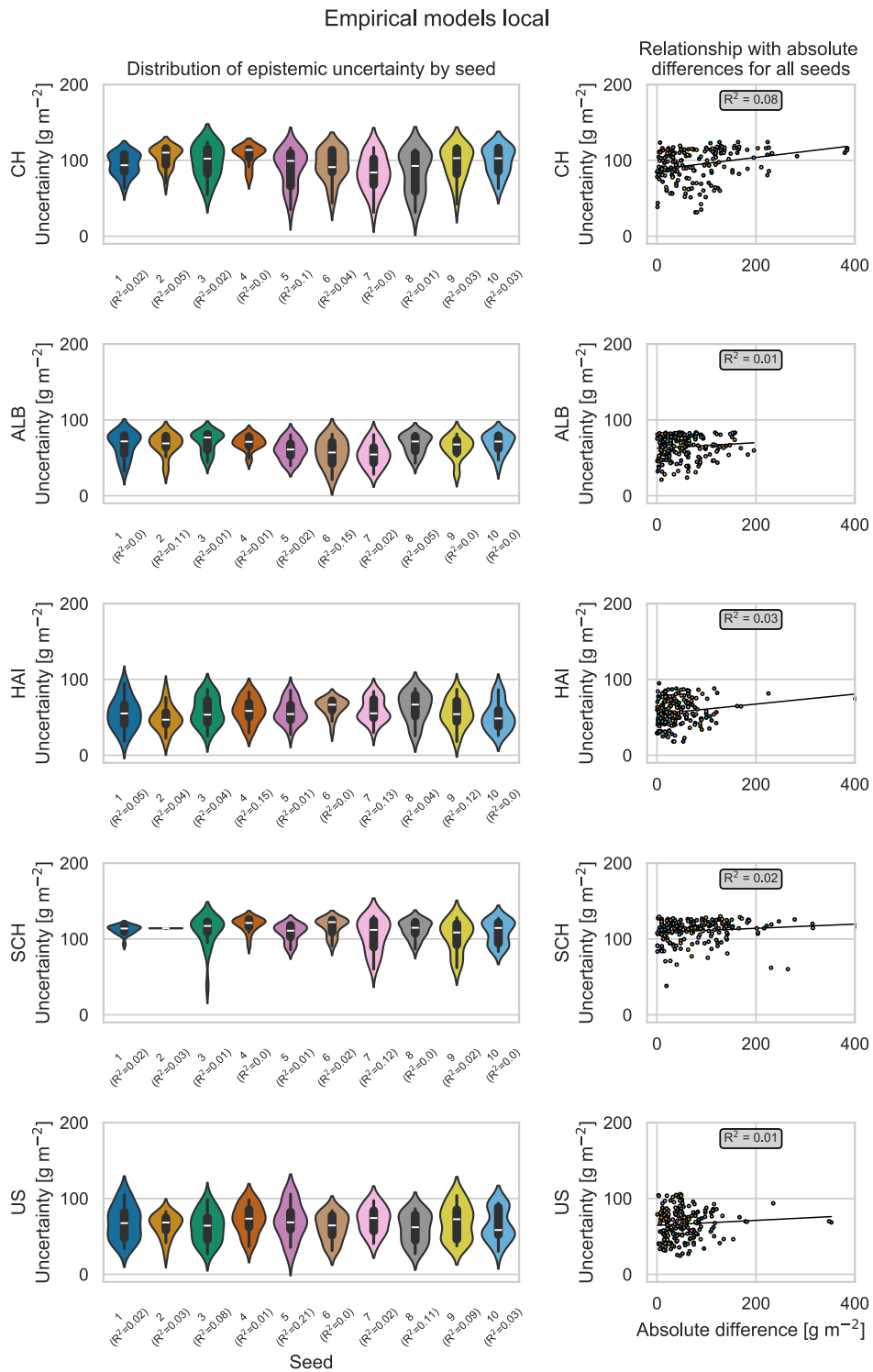
139 **Section A.16: Testing performance of global models**

140 *Table A.13: Testing performance of global models. RFR: Random Forest regression, SVR: Support Vector regression, XGB:*
 141 *Extreme Gradient Boosting regression, GPR: Gaussian process regression, CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-*
 142 *Dün, SCH: Schorfheide-Chorin, US: United States, R²: coefficient of determination, RMSE: root-mean-square error, sRMSE:*
 143 *systematic component of RMSE, uRMSE: unsystematic component of RMSE, RRMSE: relative root-mean-square error, MBE:*
 144 *mean bias error.*

		Model type					
Prediction site	Metric	Empirical RFR	Empirical SVR	Empirical XGB	Empirical GPR	Physically-based	Hybrid
CH	R ²	0	0.02	0	0	0.08	0
	RMSE	120.76	130.09	127.31	110.83	108.31	124.45
	sRMSE	110.28	111.77	110.95	110.83	110.73	112.64
	uRMSE	49.22	66.56	62.43	0	23.04	52.92
	RRMSE	0.7	0.75	0.73	0.64	0.62	0.72
	MBE	-2.51	-17.25	11.75	11	-12.76	22.71
ALB	R ²	0.41	0.49	0.42	0	0.43	0.43
	RMSE	82.9	68.51	83.38	100.75	85.6	74.31
	sRMSE	96.33	89.08	94.89	100.74	101.59	92.2
	uRMSE	49.06	56.93	45.28	1.41	54.72	54.58
	RRMSE	0.6	0.49	0.6	0.72	0.61	0.53
	MBE	46.55	28.87	43.51	54.13	56.67	37.3
HAI	R ²	0.44	0.43	0.34	0	0.49	0.42
	RMSE	85.81	76.13	93.5	119.47	94.98	73.71
	sRMSE	102.59	94.17	105.47	119.47	111.87	91.89
	uRMSE	56.22	55.43	48.81	0	59.12	54.87
	RRMSE	0.74	0.66	0.81	1.03	0.82	0.64
	MBE	57.15	40.11	62.07	83.08	72.55	34.41
SCH	R ²	0.28	0.27	0.29	0	0.22	0.18
	RMSE	117.09	127.69	116.49	129.1	111.52	134.26
	sRMSE	132.45	141.13	132.16	129.1	124.09	143.04
	uRMSE	61.9	60.12	62.42	0	54.43	49.34
	RRMSE	0.53	0.58	0.52	0.58	0.5	0.6
	MBE	-58.6	-76.21	-57.97	-49.32	-35.71	-79.54
US	R ²	0.28	0.05	0.2	0	0.45	0.25
	RMSE	211.53	224.32	211.52	130.9	197.08	136.03
	sRMSE	214.28	221.48	213.6	0	201.81	127.12
	uRMSE	34.21	35.59	29.74	0	43.45	48.43
	RRMSE	0.81	0.86	0.81	0.5	0.75	0.52
	MBE	-196.59	-204.11	-195.82	-98.89	-182.99	-93.37

145

146



148

149 Figure A.3: Epistemic uncertainty of local empirical Gaussian process regression models. Absolute differences correspond to

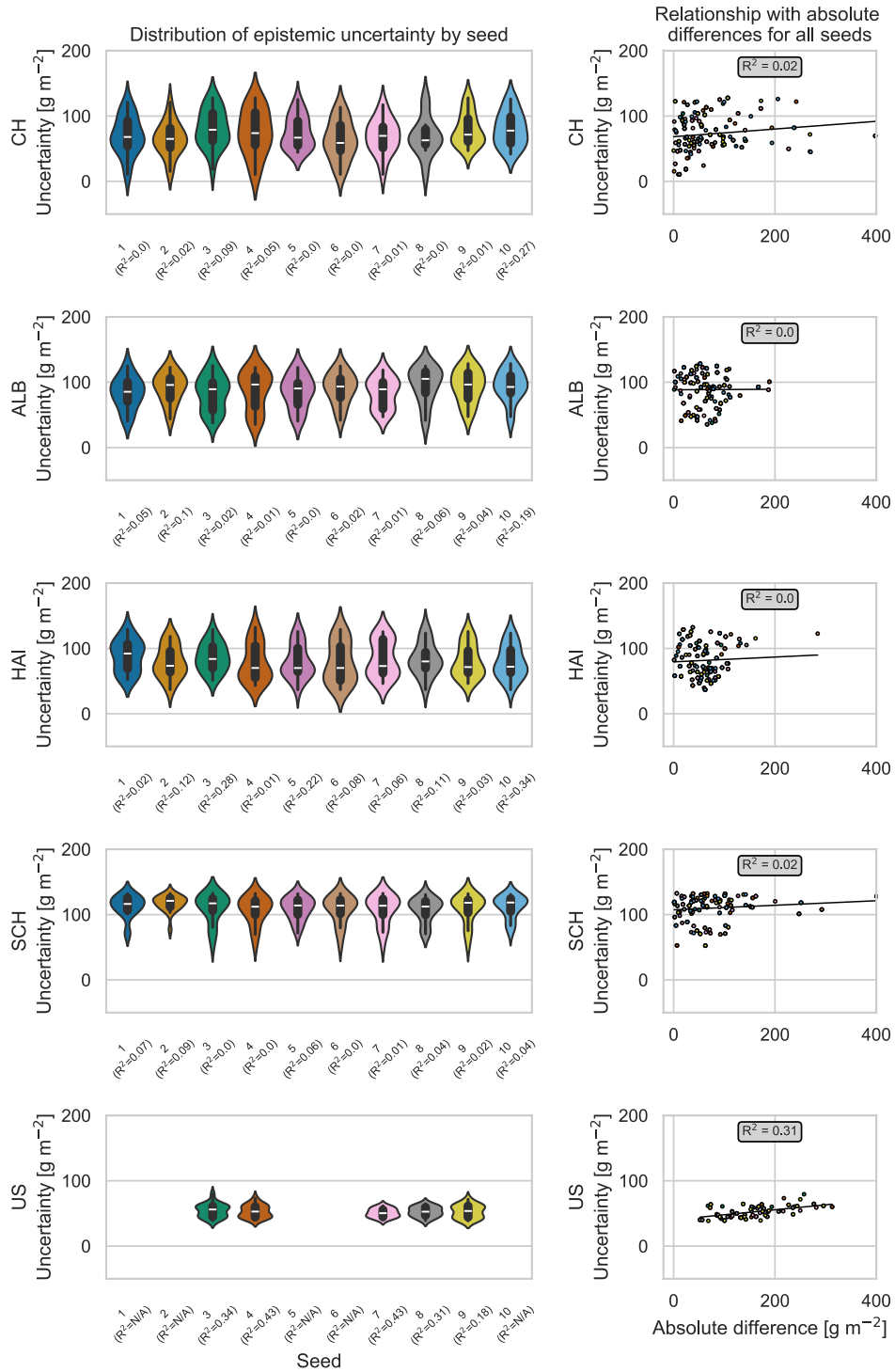
150 the absolute difference between measured and predicted biomass value. In case a model predicted a straight line (e.g., the

151 prediction is constant), the calculation of the coefficient of determination (R^2) was not possible. The corresponding

152 predictions were omitted for the calculation of the overall R^2 showed in the textbox in the right column. CH: Switzerland,

153 ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States.

Physically-based models local



154

155

156

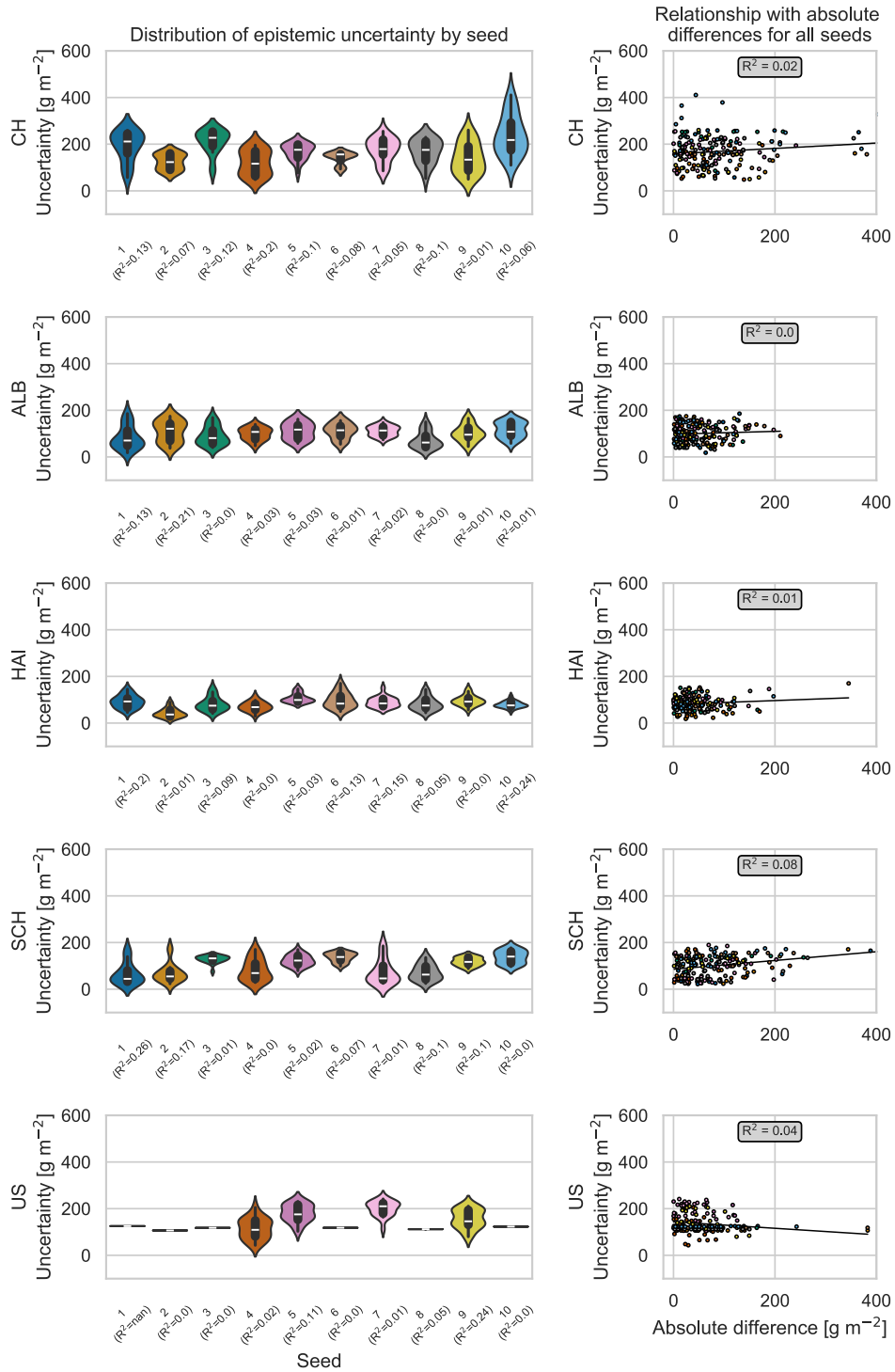
157

158

159

Figure A.4: Epistemic uncertainty of local physically-based models. Absolute differences correspond to the absolute difference between measured and predicted biomass value. If the percentage of solutions was equal to 0.01% (e.g., only 1 sample), no standard deviation and R² could be calculated. The corresponding predictions were omitted for the calculation of the overall R² showed in the textbox in the right column. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States.

Hybrid models local



160

161

162

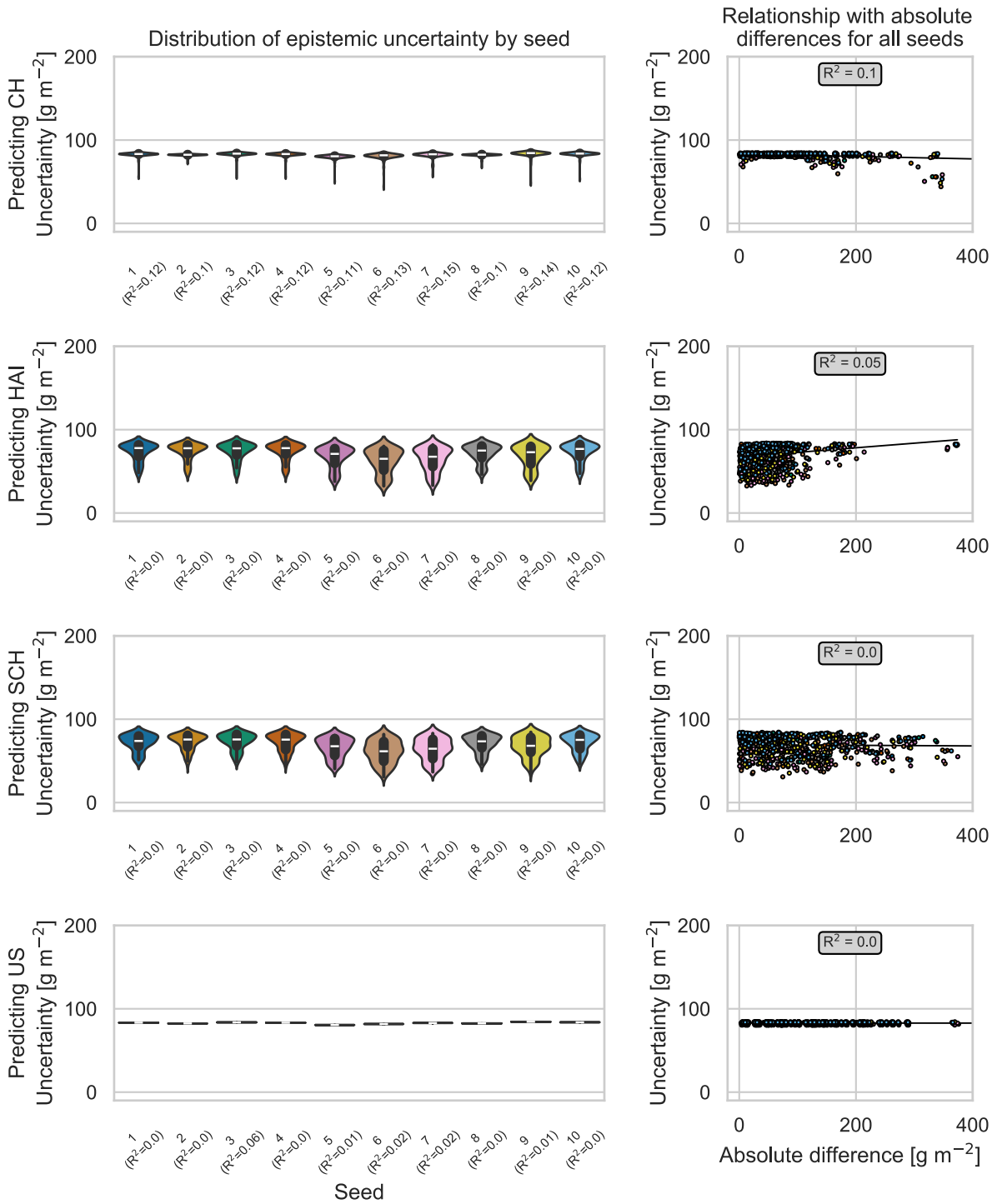
163

164

165

Figure A.5: Epistemic uncertainty of local hybrid models. Absolute differences correspond to the absolute difference between measured and predicted biomass value. In case a model predicted a straight line (e.g., the prediction is constant), the calculation of the coefficient of determination (R^2) was not possible. The corresponding predictions were omitted for the calculation of the overall R^2 showed in the textbox in the right column. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States.

Empirical models transferred ALB model



172

173

174

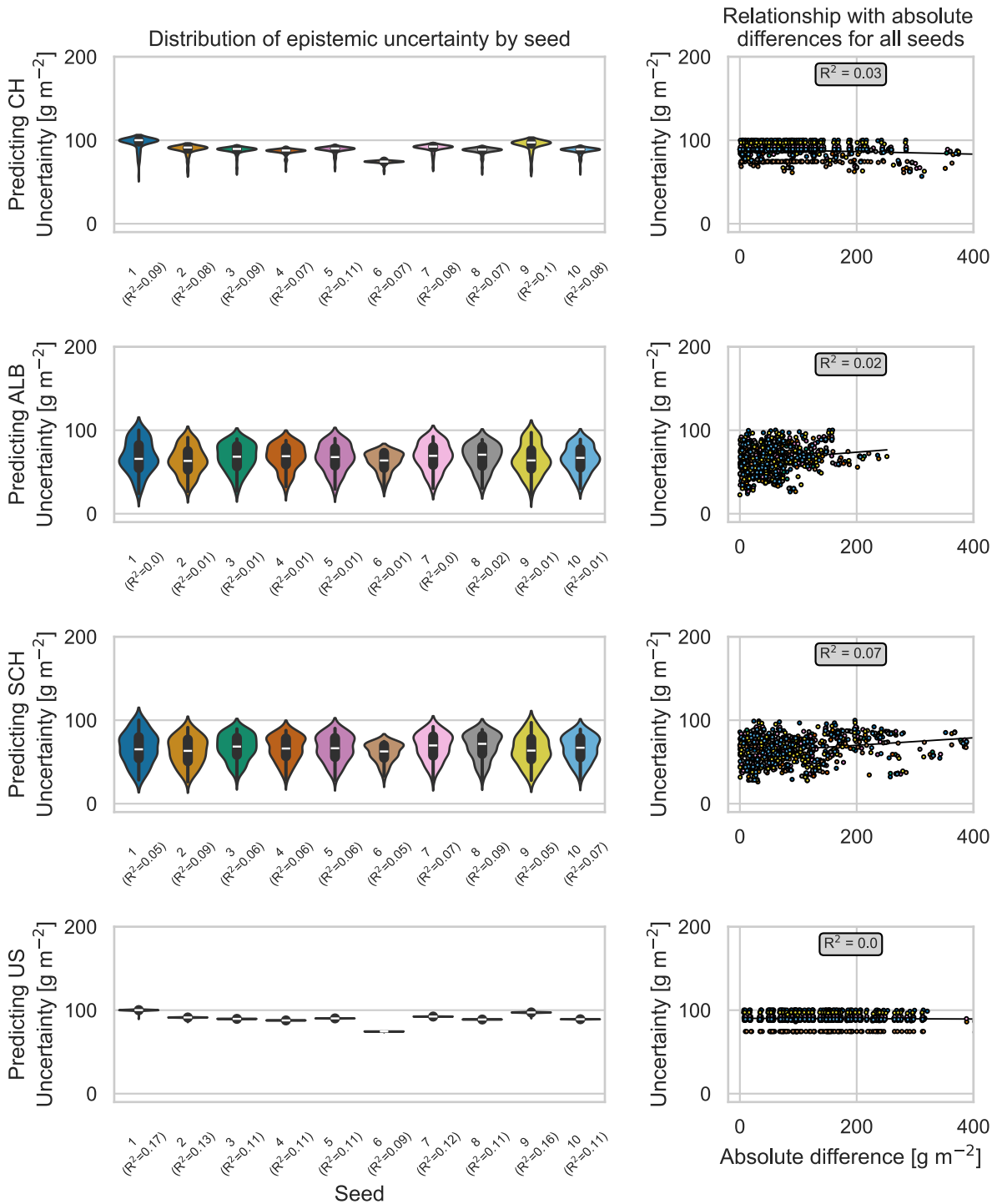
175

176

177

Figure A.7: Epistemic uncertainty of transferred empirical Gaussian process regression models for the ALB site. Absolute differences correspond to the absolute difference between measured and predicted biomass value. In case a model predicted a straight line (e.g., the prediction is constant), the calculation of the coefficient of determination (R^2) was not possible. The corresponding predictions were omitted for the calculation of the overall R^2 showed in the textbox in the right column. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States.

Empirical models transferred HAI model



178

179

180

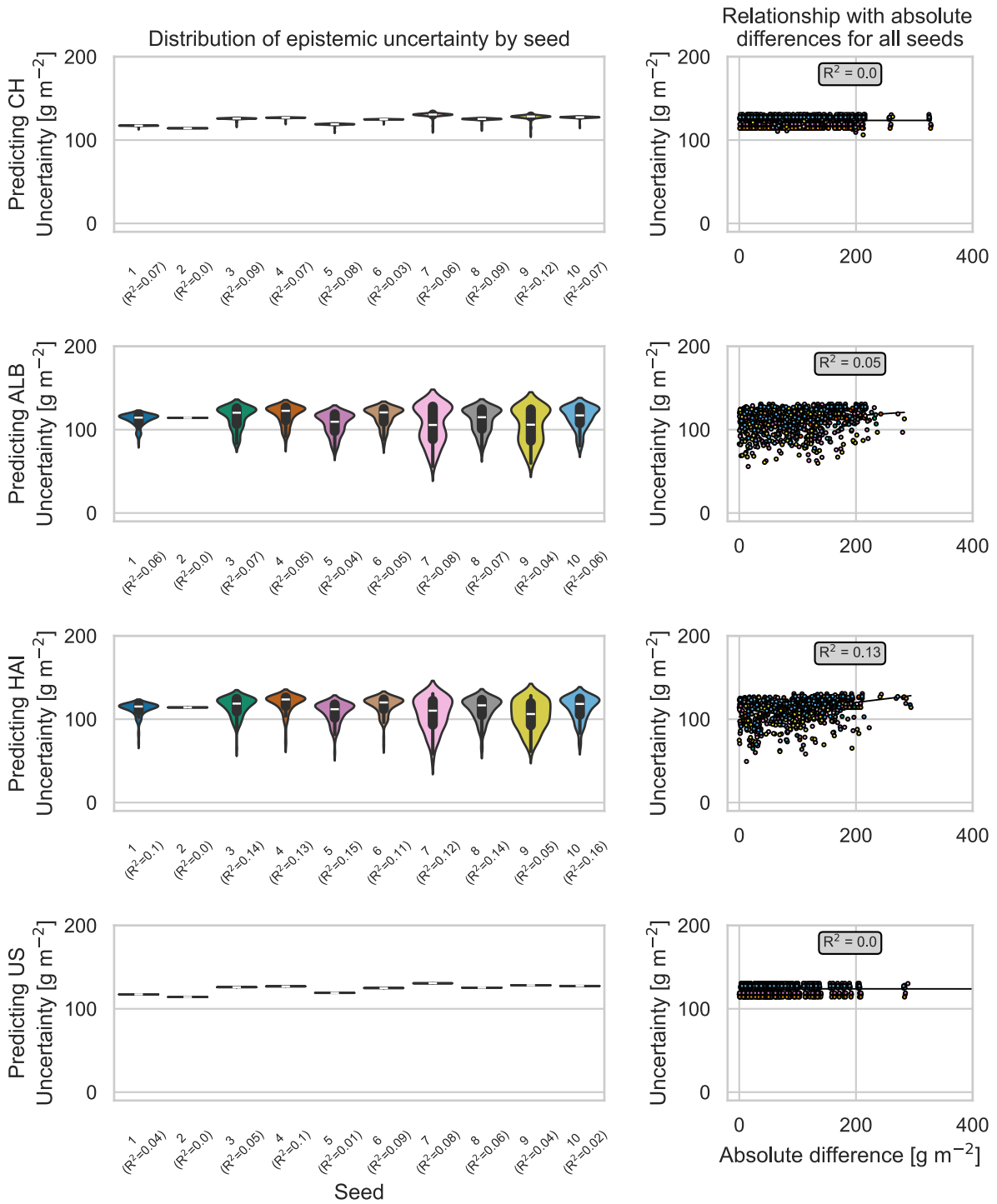
181

182

183

Figure A.8: Epistemic uncertainty of transferred empirical Gaussian process regression models for the HAI site. Absolute differences correspond to the absolute difference between measured and predicted biomass value. In case a model predicted a straight line (e.g., the prediction is constant), the calculation of the coefficient of determination (R^2) was not possible. The corresponding predictions were omitted for the calculation of the overall R^2 showed in the textbox in the right column. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States.

Empirical models transferred SCH model



184

185

186

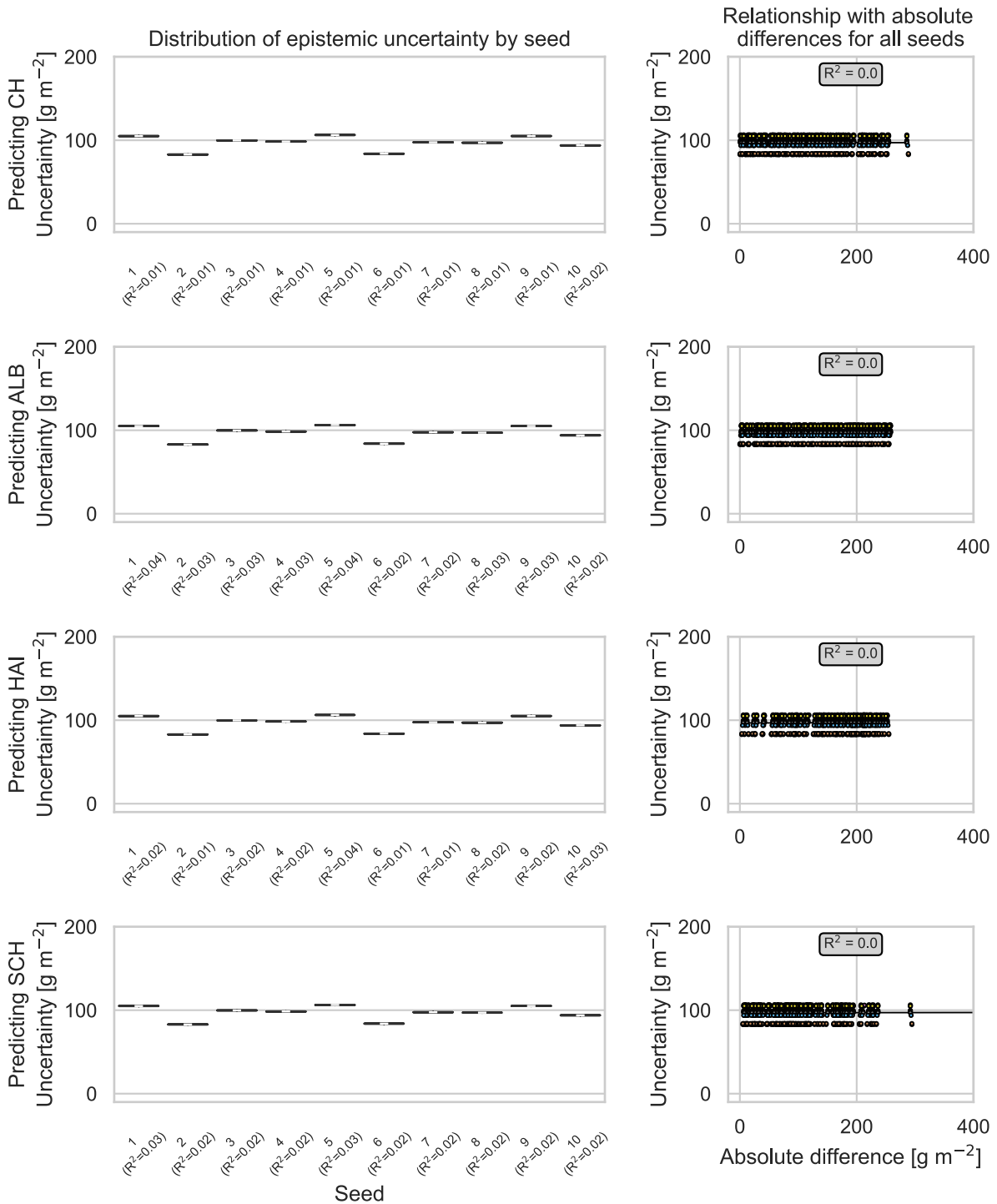
187

188

189

Figure A.9: Epistemic uncertainty of transferred empirical Gaussian process regression models for the SCH site. Absolute differences correspond to the absolute difference between measured and predicted biomass value. In case a model predicted a straight line (e.g., the prediction is constant), the calculation of the coefficient of determination (R^2) was not possible. The corresponding predictions were omitted for the calculation of the overall R^2 showed in the textbox in the right column. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States.

Empirical models transferred US model



190

191

192

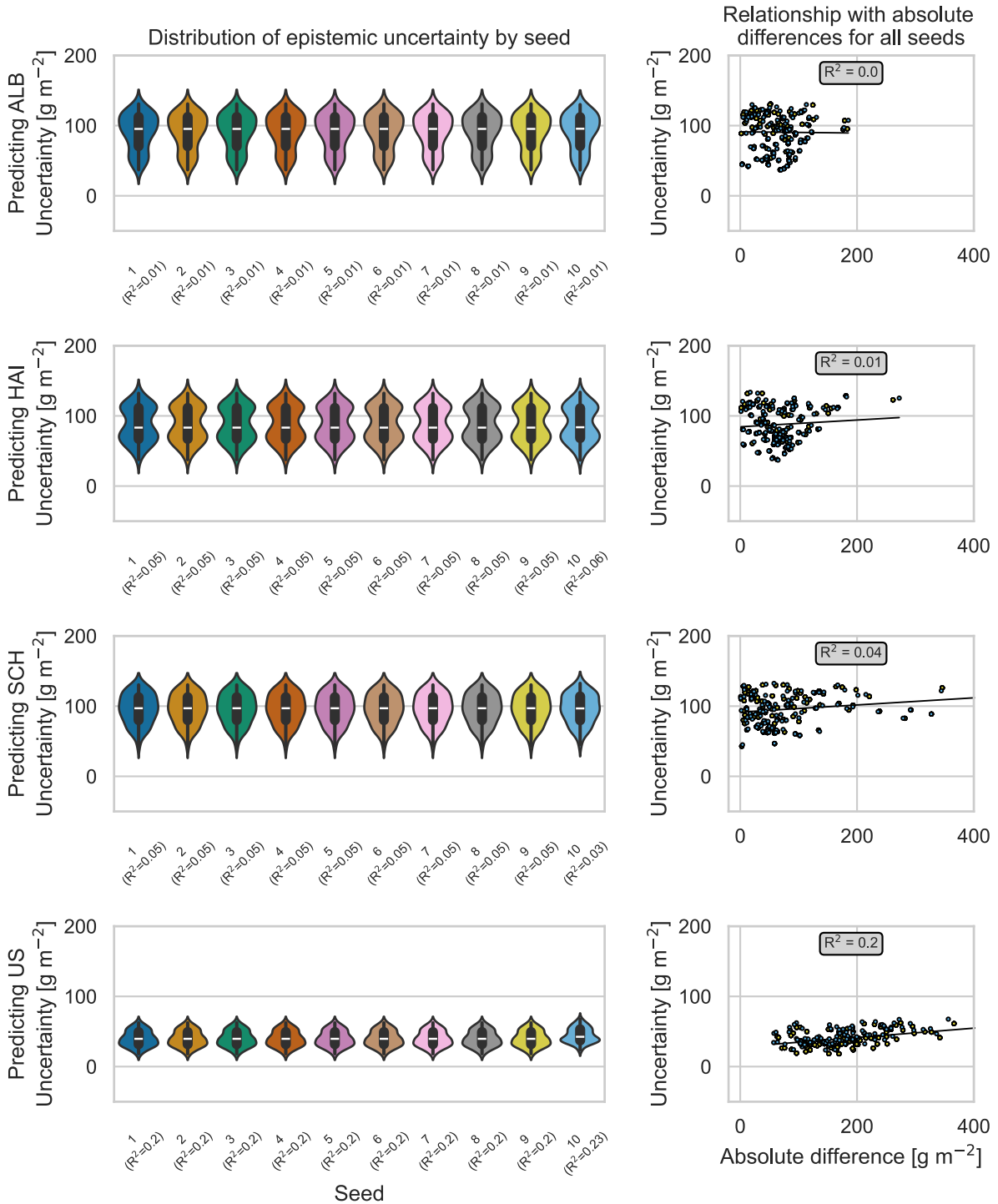
193

194

195

Figure A.10: Epistemic uncertainty of transferred empirical Gaussian process regression models for the US site. Absolute differences correspond to the absolute difference between measured and predicted biomass value. In case a model predicted a straight line (e.g., the prediction is constant), the calculation of the coefficient of determination (R^2) was not possible. The corresponding predictions were omitted for the calculation of the overall R^2 showed in the textbox in the right column. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States.

Physically-based models transferred CH model



196

197

198

199

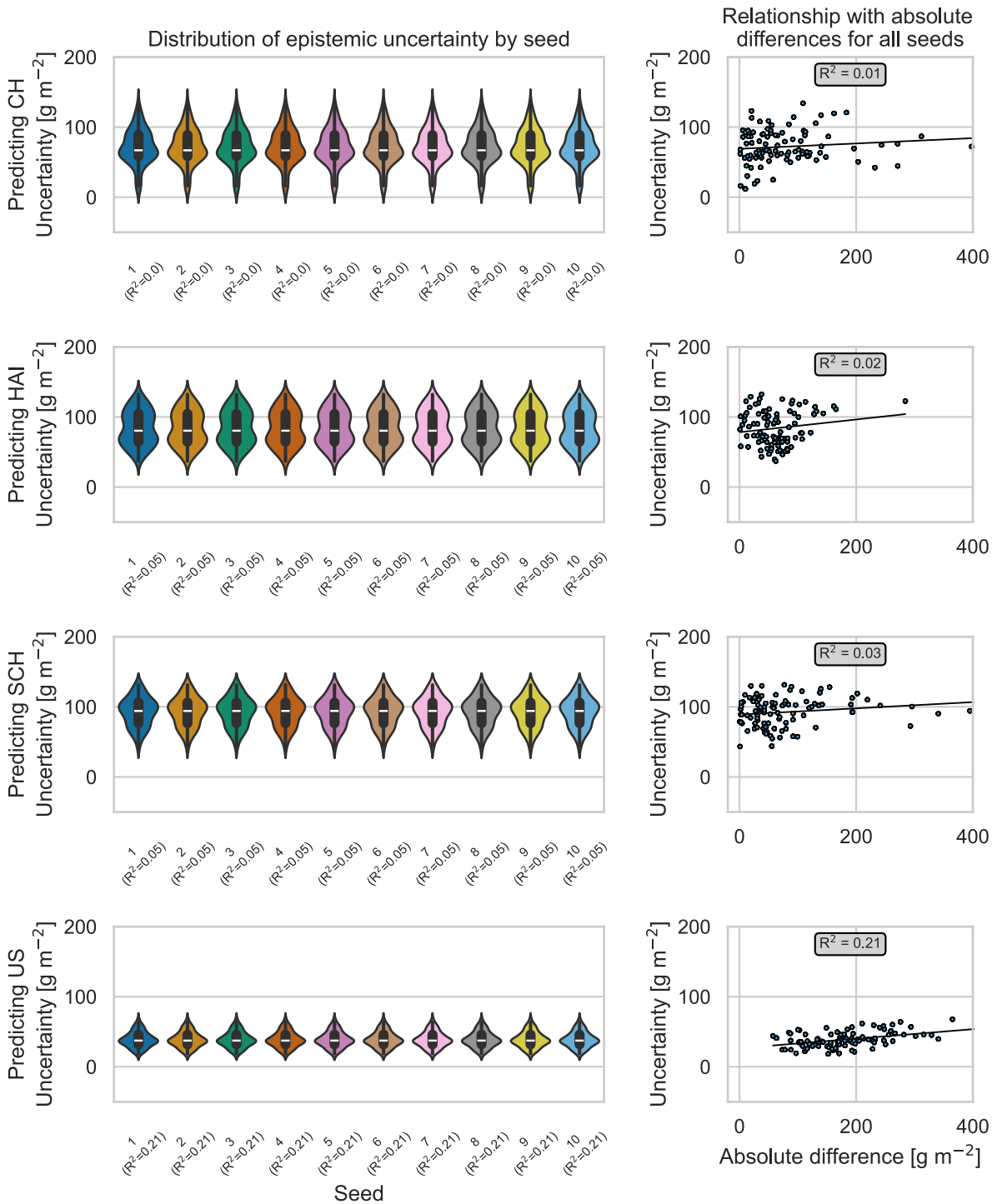
200

201

Figure A.11: Epistemic uncertainty of transferred physically-based models for the CH site. Absolute differences correspond to the absolute difference between measured and predicted biomass value. If the percentage of solutions was equal to 0.01% (e.g., only 1 sample), no standard deviation and R² could be calculated. The corresponding predictions were omitted for the calculation of the overall R² showed in the textbox in the right column. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-

Dün, SCH: Schorfheide-Chorin, US: United States.

Physically-based models transferred ALB model



202

203

Figure A.12: Epistemic uncertainty of transferred physically-based models for the ALB site. Absolute differences correspond

204

to the absolute difference between measured and predicted biomass value. If the percentage of solutions was equal to

205

0.01% (e.g., only 1 sample), no standard deviation and R^2 could be calculated. The corresponding predictions were omitted

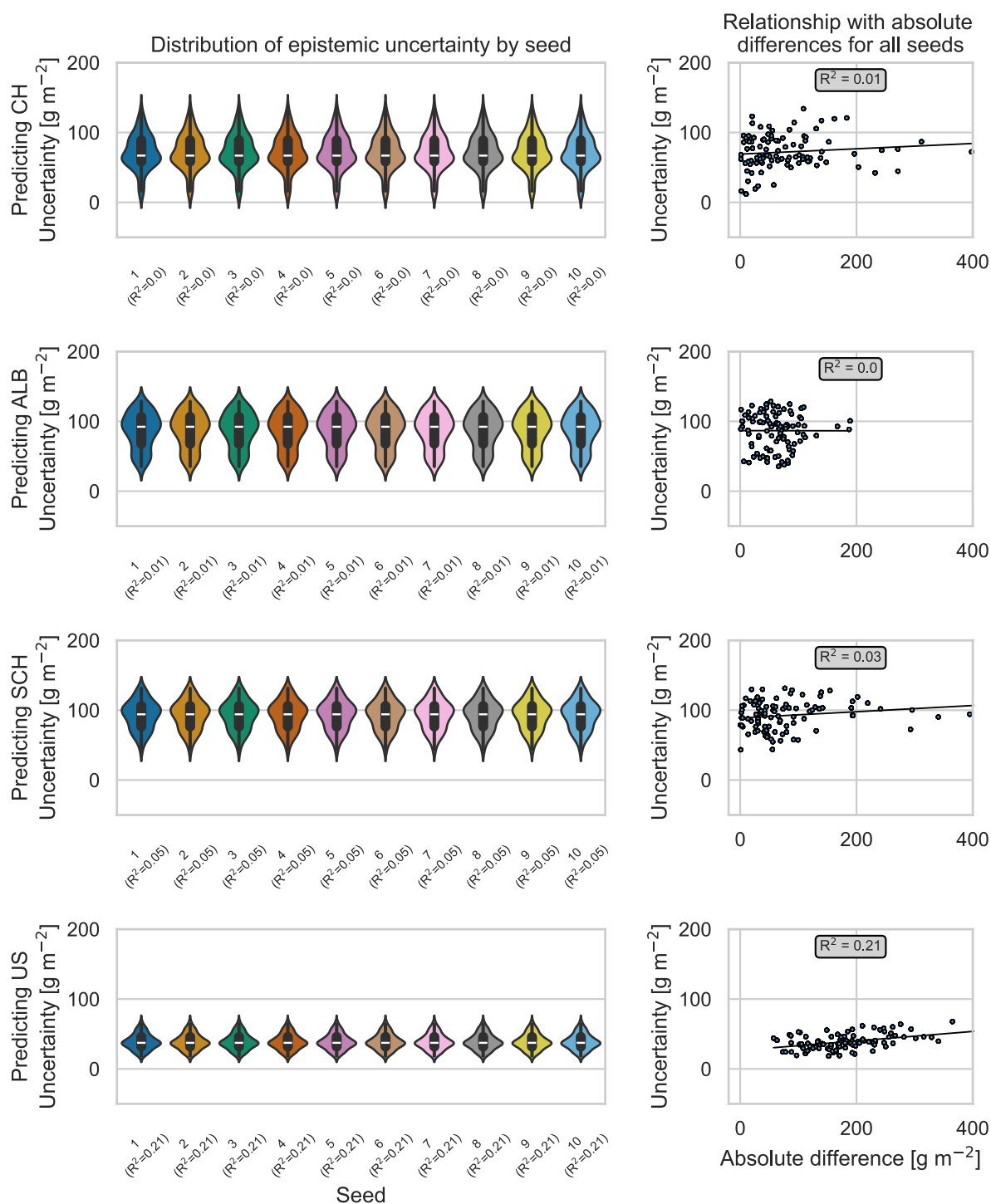
206

for the calculation of the overall R^2 showed in the textbox in the right column. CH: Switzerland, ALB: Schwäbische Alb, HAI:

207

Hainich-Dün, SCH: Schorfheide-Chorin, US: United States.

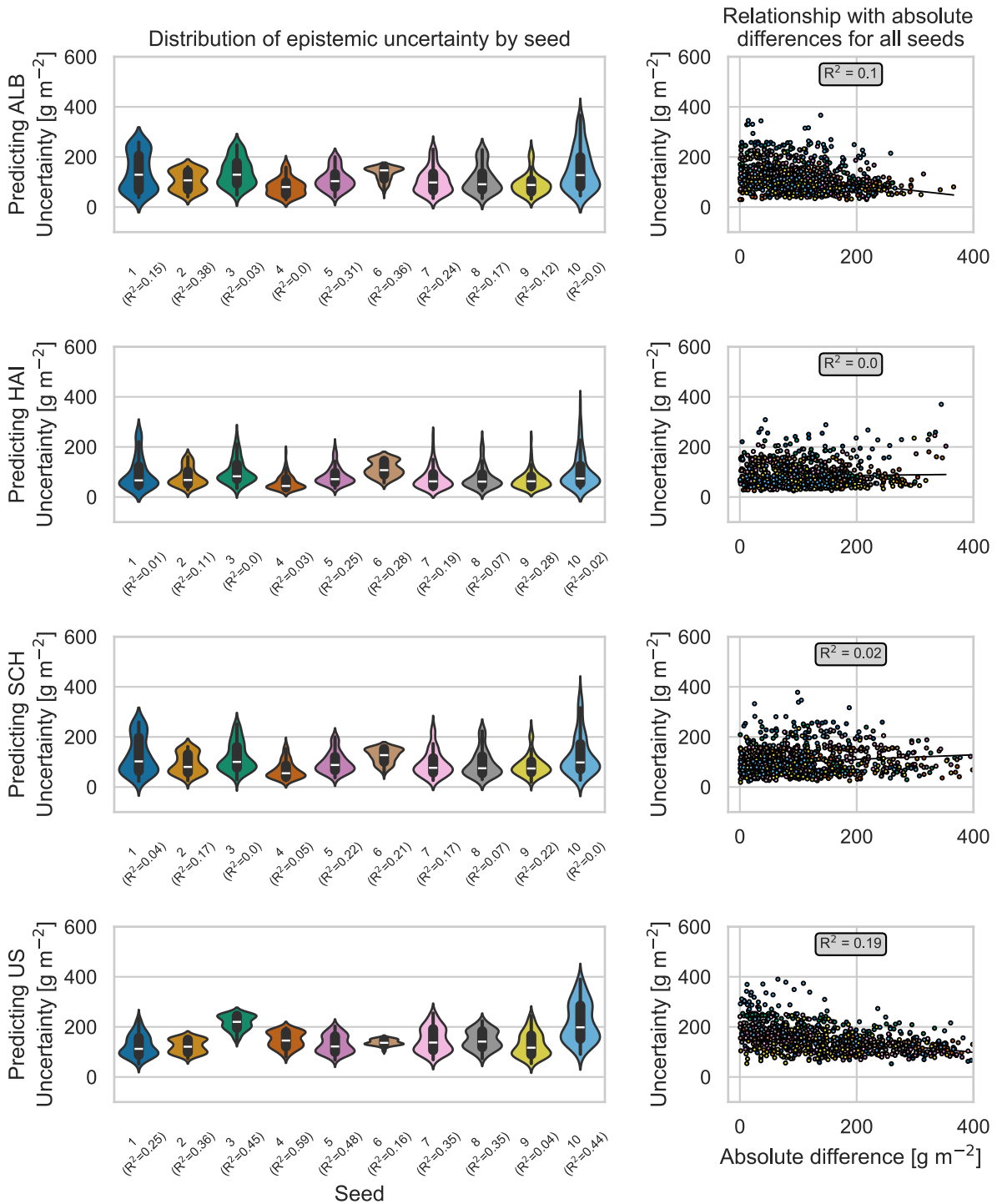
Physically-based models transferred HAI model



208
209
210
211
212
213

Figure A.13: Epistemic uncertainty of transferred physically-based models for the HAI site. Absolute differences correspond to the absolute difference between measured and predicted biomass value. If the percentage of solutions was equal to 0.01% (e.g., only 1 sample), no standard deviation and R^2 could be calculated. The corresponding predictions were omitted for the calculation of the overall R^2 showed in the textbox in the right column. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States.

Hybrid models transferred CH model



226

227

228

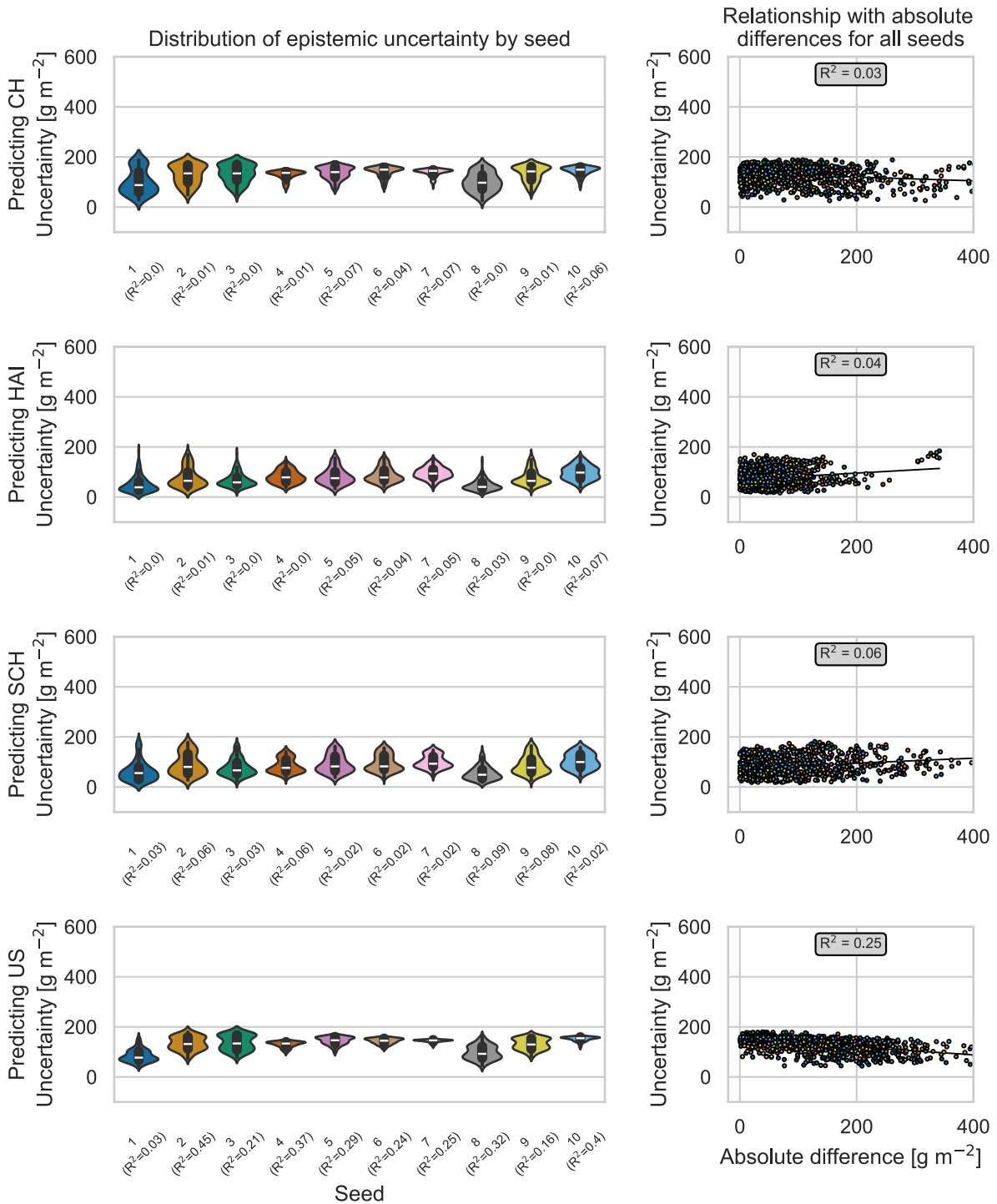
229

230

231

Figure A.16: Epistemic uncertainty of transferred hybrid models for the CH site. Absolute differences correspond to the absolute difference between measured and predicted biomass value. In case a model predicted a straight line (e.g., the prediction is constant), the calculation of the coefficient of determination (R^2) was not possible. The corresponding predictions were omitted for the calculation of the overall R^2 showed in the textbox in the right column. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States.

Hybrid models transferred ALB model



232

233

234

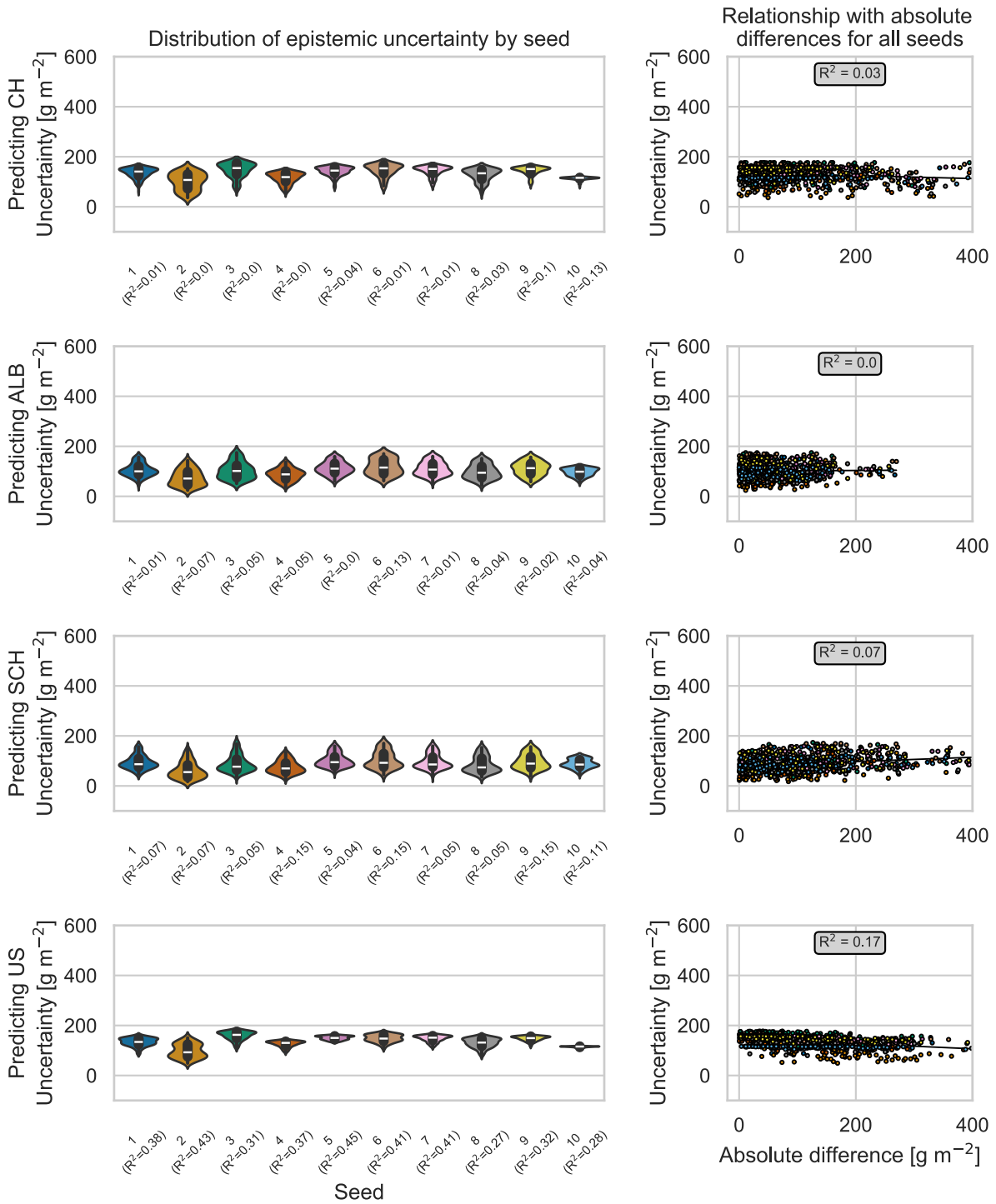
235

236

237

Figure A.17: Epistemic uncertainty of transferred hybrid models for the ALB site. Absolute differences correspond to the absolute difference between measured and predicted biomass value. In case a model predicted a straight line (e.g., the prediction is constant), the calculation of the coefficient of determination (R^2) was not possible. The corresponding predictions were omitted for the calculation of the overall R^2 showed in the textbox in the right column. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States.

Hybrid models transferred HAI model



238

239

240

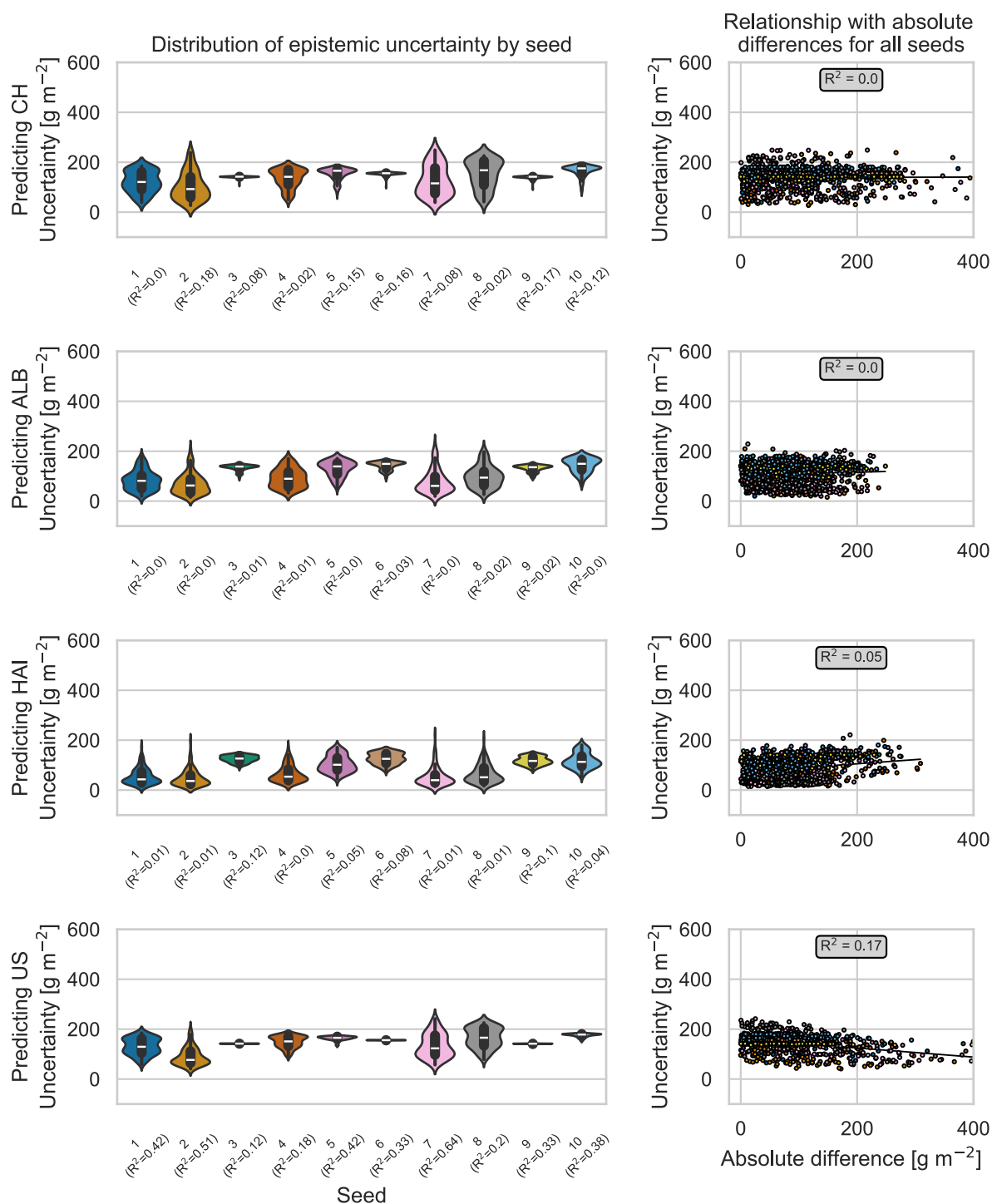
241

242

243

Figure A.18: Epistemic uncertainty of transferred hybrid models for the HAI site. Absolute differences correspond to the absolute difference between measured and predicted biomass value. In case a model predicted a straight line (e.g., the prediction is constant), the calculation of the coefficient of determination (R^2) was not possible. The corresponding predictions were omitted for the calculation of the overall R^2 showed in the textbox in the right column. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States.

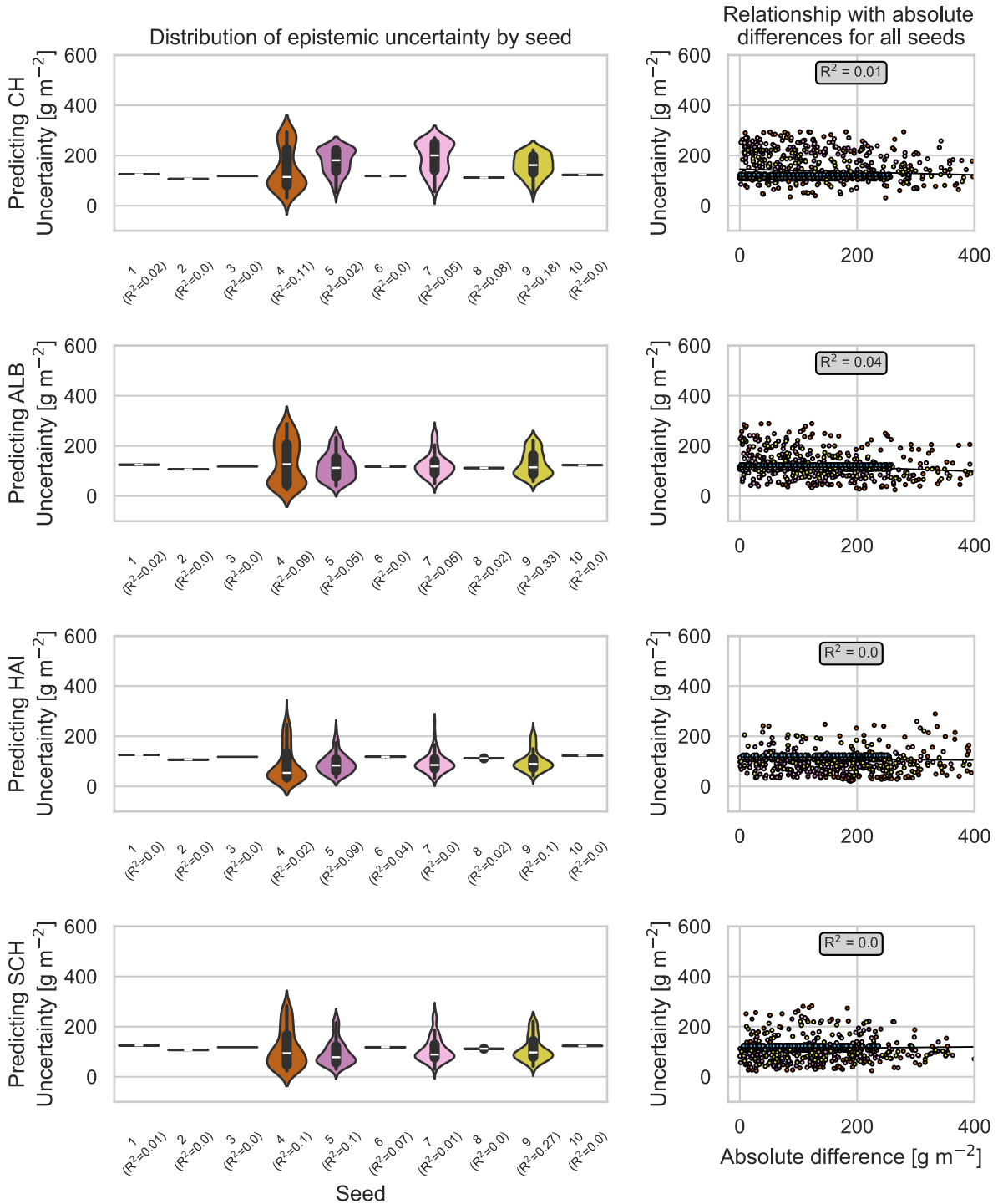
Hybrid models transferred SCH model



244
 245
 246
 247
 248
 249

Figure A.19: Epistemic uncertainty of transferred hybrid models for the SCH site. Absolute differences correspond to the absolute difference between measured and predicted biomass value. In case a model predicted a straight line (e.g., the prediction is constant), the calculation of the coefficient of determination (R^2) was not possible. The corresponding predictions were omitted for the calculation of the overall R^2 showed in the textbox in the right column. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States.

Hybrid models transferred US model



250

251

252

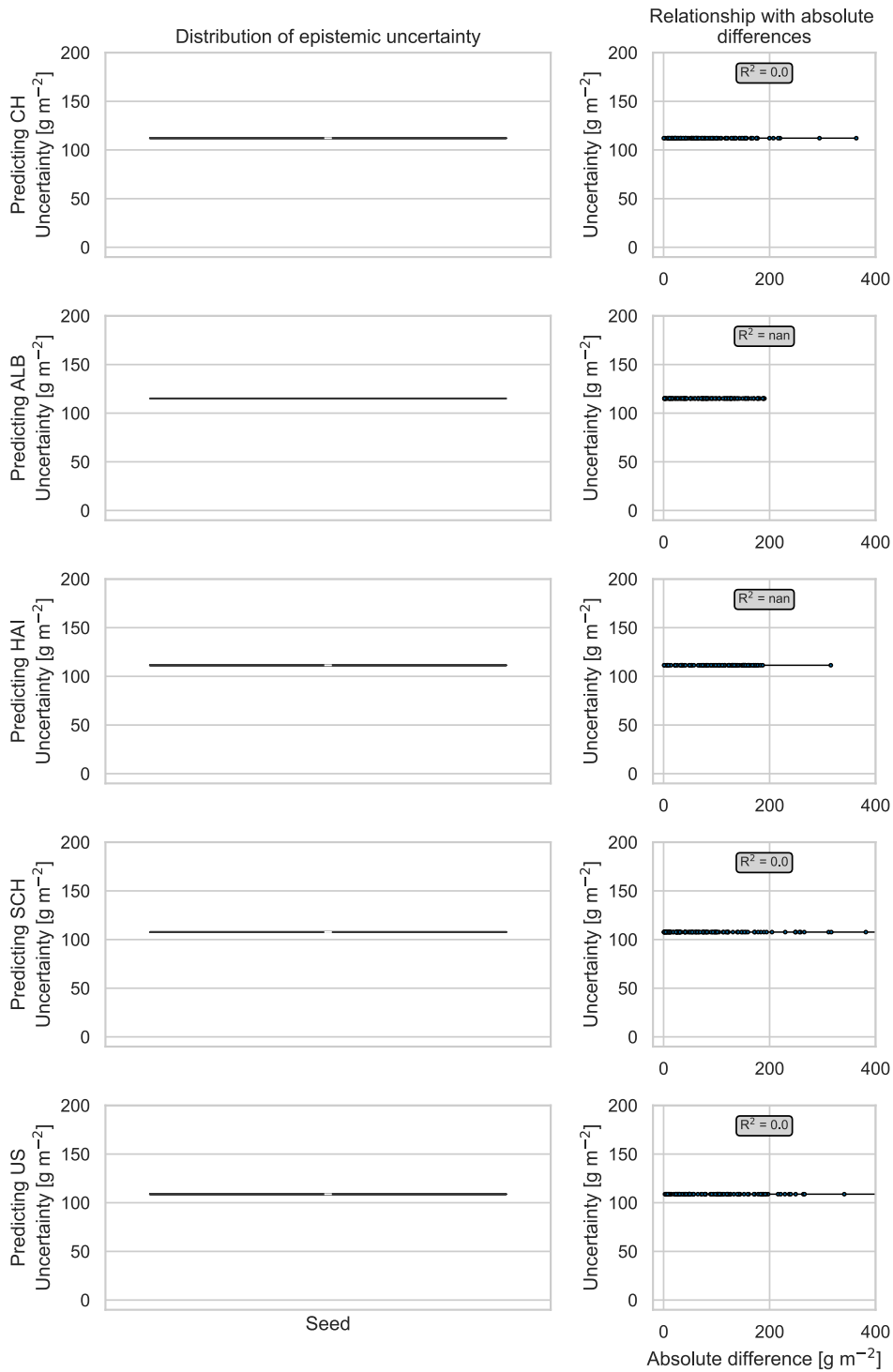
253

254

255

Figure A.20: Epistemic uncertainty of transferred hybrid models for the US site. Absolute differences correspond to the absolute difference between measured and predicted biomass value. In case a model predicted a straight line (e.g., the prediction is constant), the calculation of the coefficient of determination (R^2) was not possible. The corresponding predictions were omitted for the calculation of the overall R^2 showed in the textbox in the right column. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States.

Empirical models global



256

257

258

259

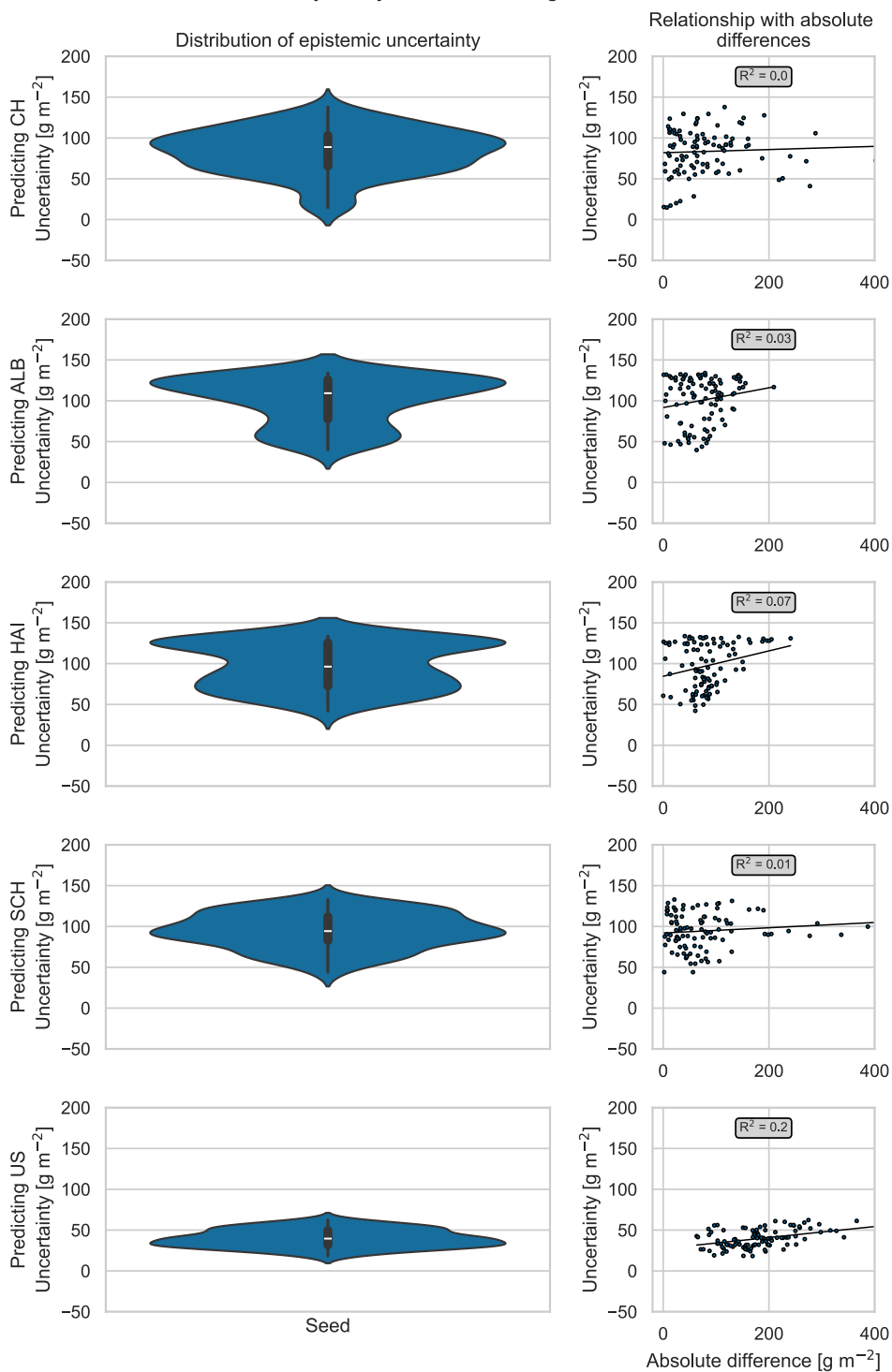
260

261

262

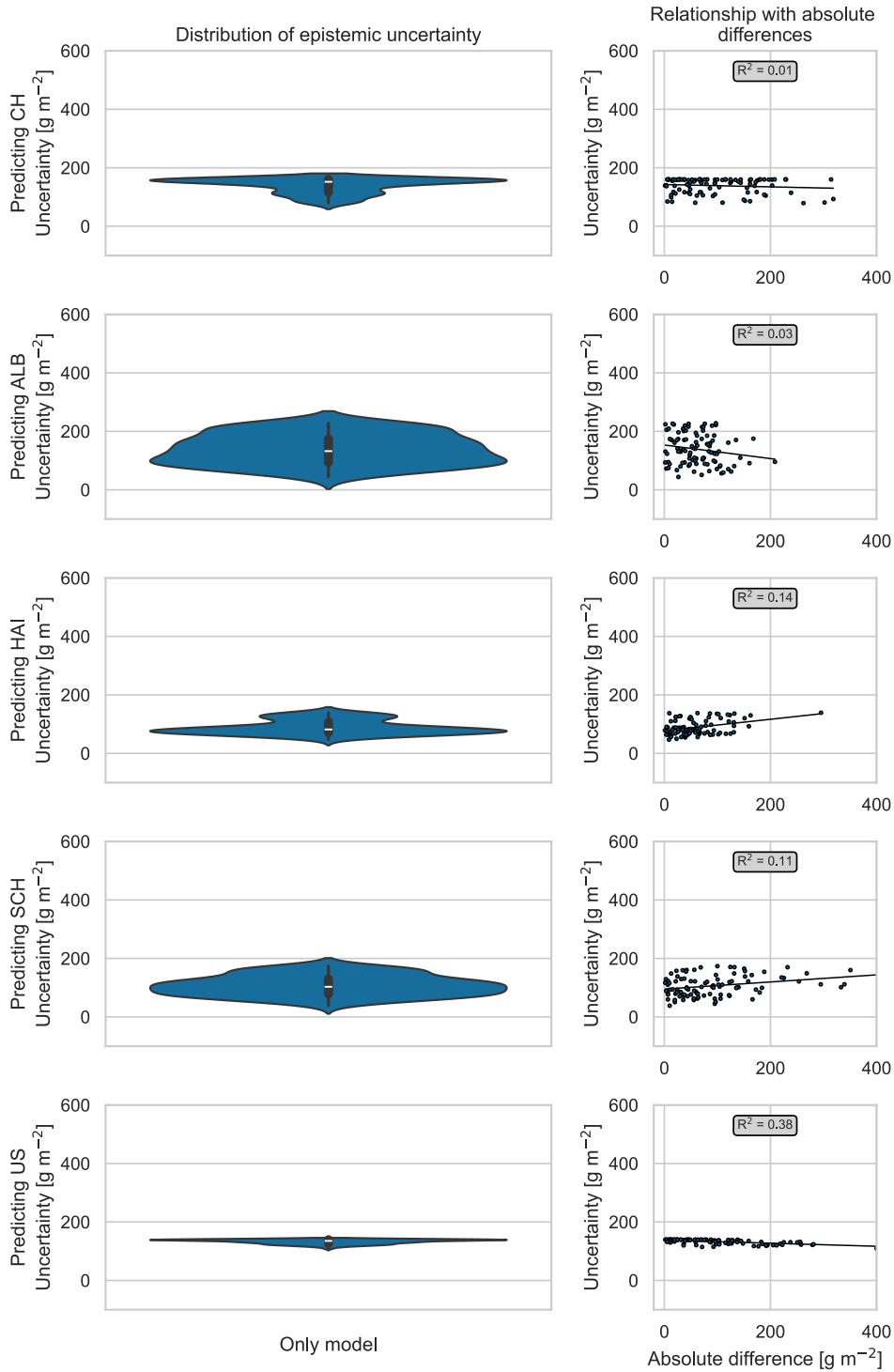
Figure A.21: Epistemic uncertainty of global empirical models. Absolute differences correspond to the absolute difference between measured and predicted biomass value. In case a model predicted a straight line (e.g., the prediction is constant), the calculation of the coefficient of determination (R^2) was not possible. The corresponding predictions were omitted for the calculation of the overall R^2 showed in the textbox in the right column. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States.

Physically-based models global



263
 264 *Figure A.22: Epistemic uncertainty of global physically-based models. Absolute differences correspond to the absolute*
 265 *difference between measured and predicted biomass value. If the percentage of solutions was equal to 0.01% (e.g., only 1*
 266 *sample), no standard deviation and R^2 could be calculated. The corresponding predictions were omitted for the calculation*
 267 *of the overall R^2 showed in the textbox in the right column. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH:*
 268 *Schorfheide-Chorin, US: United States.*

Hybrid models global



270

271

272

273

274

275

Figure A.23: Epistemic uncertainty of global hybrid models. Absolute differences correspond to the absolute difference between measured and predicted biomass value. In case a model predicted a straight line (e.g., the prediction is constant), the calculation of the coefficient of determination (R^2) was not possible. The corresponding predictions were omitted for the calculation of the overall R^2 showed in the textbox in the right column. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States.

276 **Section A.18: Biomass and spectral variability by site**

277 *Table A.14: Properties of biomass (range and coefficient of variation (CV), calculated by dividing the variance by the mean)*
 278 *and the spectral bands (CV) by site, and mean relative root-mean-square error (RRMSE) and coefficient of determination (R²)*
 279 *of best local empirical, physically-based, and hybrid models. For the empirical models, only the best-performing model in terms*
 280 *of lowest RRMSE is shown with the model specified in brackets. A missing R² indicates that its calculation was not possible due*
 281 *to predictions of constant values (=horizontal line). CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH:*
 282 *Schorfheide-Chorin, US: United States, RFR: Random Forest regression, SVR: Support Vector regression, XGB: Extreme Gradient*
 283 *Boosting regression, GPR: Gaussian Process regression.*

Property	Site				
	CH	ALB	HAI	SCH	US
Biomass range [g m ⁻²]	540.49	379.00	502.37	643.00	518.60
Biomass CV [g m ⁻²]	70.73	52.40	64.18	64.78	28.41
490 nm band CV * 100	0.94	0.37	0.36	0.24	0.49
560 nm band CV * 100	0.56	0.20	0.18	0.15	0.08
665 nm band CV * 100	1.25	1.02	0.75	0.57	0.09
705 nm band CV * 100	0.60	0.40	0.34	0.24	0.23
740 nm band CV * 100	1.72	0.71	0.57	0.56	0.19
783 nm band CV * 100	2.15	1.24	1.03	1.01	0.26
842 nm band CV * 100	2.06	1.23	1.05	0.90	0.27
865 nm band CV * 100	2.14	1.09	0.92	0.89	0.27
1610 nm band CV * 100	0.98	1.00	0.78	0.65	0.37
2190 nm band CV * 100	0.99	1.09	0.84	0.61	0.46
Mean CV of all bands * 100	1.34	0.84	0.68	0.58	0.27
Mean RRMSE of best local empirical models	0.55 (SVR)	0.43 (SVR)	0.44 (RFR)	0.42 (SVR)	0.22 (SVR)
Mean R ² of best local empirical models	0.31 (SVR)	0.54 (SVR)	0.64 (RFR)	0.42 (SVR)	0.59 (SVR)
Mean RRMSE of best local physically-based models	0.59	0.52	0.59	0.47	0.71
Mean R ² of best local physically-based models	0.26	0.43	0.66	0.30	0.30
Mean RRMSE of best local hybrid models	0.64	0.46	0.51	0.47	0.29
Mean R ² of best local hybrid models	0.23	0.46	0.54	0.27	0.23

284

285 **Section A.19: Performances of local physically-based models with US-tailored LUT**

286 *Table A.15: Value ranges and distributions of PROSAIL input parameters when adjusting the upper bound of LMA to the*
 287 *conditions of the study site in the United States.*

Parameter	Variable	Unit	Minimum value	Maximum value	Distribution
Leaf structure parameter	N	[-]	1.5	1.9	uniform
Chlorophyll content	CHL	[$\mu\text{g cm}^{-2}$]	5	75	uniform
Carotenoid content	CAR	[$\mu\text{g cm}^{-2}$]	2	60	uniform
Anthocyanin content	ANT	[$\mu\text{g cm}^{-2}$]	0	2	uniform
Brown pigment content	BROWN	[-]	0	1	uniform
Equivalent water thickness	EWT	[cm]	0.001	0.04	uniform
Leaf mass per area	LMA	[g cm^{-2}]	0.002	0.03	uniform
Angle for incident light at leaf surface	alpha	[$^{\circ}$]	40	40	fixed
Leaf inclination distribution function	TypeLidf	[-]	2	2	fixed
Average leaf angle	LIDFa	[$^{\circ}$]	40	70	uniform
Leaf area index	LAI	[-]	0.1	4	uniform
Hot spot parameter	q	[-]	0.01	0.1	uniform
Sun zenith angle	tts	[$^{\circ}$]	25	75	uniform
Observer zenith angle	tto	[$^{\circ}$]	0	0	fixed
Relative azimuth angle	psi	[$^{\circ}$]	50	180	uniform
Dry/wet soil factor	psoil	[-]	0	1	uniform

288

289

290 Table A.16: Cost function, percentage of solutions, coefficient of determination (R^2), root-mean-square error (RMSE), and
 291 relative root-mean-square error (RRMSE) of best performing local physically-based models with US-tailored look-up table
 292 (LUT) and the LUT used in this study. The last columns show the difference for the three metrics. CH: Switzerland, ALB:
 293 Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States.

		US-tailored LUT					LUT used in this study					Difference		
Site	Seed	Cost function	Percentage of solutions	R^2	RMSE [g m ⁻²]	RRMSE	Cost function	Percentage of solutions	R^2	RMSE [g m ⁻²]	RRMSE	R^2	RMSE [g m ⁻²]	RRMSE
CH	1	min_contrast_1	10	0.146	183.977	1.069	min_contrast_1	1	0.221	95.683	0.556	-0.075	88.294	0.513
CH	2	min_contrast_1	10	0.129	192.067	1.111	min_contrast_1	1	0.214	96.421	0.558	-0.085	95.646	0.553
CH	3	min_contrast_1	10	0.089	182.882	1.047	min_contrast_1	1	0.149	107.756	0.617	-0.066	75.126	0.436
CH	4	min_contrast_1	10	0.104	187.059	1.077	min_contrast_1	1	0.142	104.529	0.602	-0.038	82.535	0.475
CH	5	min_contrast_1	10	0.152	176.047	1.014	min_contrast_1	1	0.207	102.279	0.589	-0.055	73.768	0.425
CH	6	min_contrast_1	10	0.12	189.101	1.095	min_contrast_1	1	0.211	96.402	0.558	-0.091	92.699	0.537
CH	7	min_contrast_1	10	0.138	187.292	1.079	min_contrast_1	1	0.208	101.188	0.583	-0.077	86.112	0.496
CH	8	min_contrast_1	10	0.123	189.073	1.096	min_contrast_1	1	0.192	98.848	0.573	-0.069	90.225	0.523
CH	9	min_contrast_1	10	0.15	183.183	1.05	min_contrast_1	1	0.23	101.751	0.583	-0.088	81.432	0.467
CH	10	min_contrast_1	10	0.168	185.039	1.063	min_contrast_2	1	0.263	97.142	0.558	-0.095	87.897	0.505
ALB	1	min_contrast_1	0.01	0.004	175.498	1.258	neyman_chi_square_divergence	1	0.419	72.6	0.521	-0.415	102.898	0.737
ALB	2	min_contrast_1	0.01	0	169.204	1.225	neyman_chi_square_divergence	1	0.392	70.695	0.512	-0.392	98.509	0.713
ALB	3	min_contrast_1	0.01	-0.001	174.703	1.247	neyman_chi_square_divergence	1	0.401	73.617	0.525	-0.402	101.086	0.722
ALB	4	min_contrast_1	0.01	0	179.925	1.303	neyman_chi_square_divergence	1	0.399	71.542	0.518	-0.399	108.383	0.785
ALB	5	min_contrast_4	0.01	-0.002	166.653	1.203	neyman_chi_square_divergence	1	0.491	67.975	0.491	-0.493	98.678	0.712
ALB	6	min_contrast_1	0.01	-0.001	178.391	1.283	neyman_chi_square_divergence	1	0.519	67.184	0.483	-0.522	111.207	0.8
ALB	7	min_contrast_1	0.01	0	183.314	1.325	neyman_chi_square_divergence	1	0.488	68.575	0.496	-0.488	114.739	0.829
ALB	8	min_contrast_1	0.01	-0.001	175.728	1.258	neyman_chi_square_divergence	1	0.404	71.385	0.511	-0.405	104.343	0.747
ALB	9	min_contrast_1	0.01	0.003	165.669	1.187	neyman_chi_square_divergence	1	0.424	70.498	0.505	-0.421	95.171	0.682
ALB	10	min_contrast_1	0.01	-0.003	168.671	1.209	neyman_chi_square_divergence	1	0.406	70.858	0.508	-0.409	97.813	0.701
HAI	1	min_contrast_4	1	0.248	201.964	1.73	neyman_chi_square_divergence	1	0.535	76.289	0.653	-0.287	125.675	1.077
HAI	2	min_contrast_4	1	0.221	203.59	1.747	neyman_chi_square_divergence	1	0.502	79.265	0.68	-0.281	124.325	1.067
HAI	3	min_contrast_4	1	0.255	197.651	1.707	neyman_chi_square_divergence	1	0.488	77.544	0.67	-0.233	120.107	1.037
HAI	4	min_contrast_4	1	0.262	206.588	1.777	neyman_chi_square_divergence	1	0.521	79.296	0.682	-0.259	127.292	1.095
HAI	5	min_contrast_3	1	0.258	204.656	1.757	neyman_chi_square_divergence	1	0.471	79.608	0.683	-0.213	125.048	1.074

H AI	6	min_contrast_4	1	0.29	209.137	1.831	neyman_chi_square_divergence	1	0.56	74.294	0.65	-0.27	134.843	1.181
H AI	7	min_contrast_3	1	0.242	202.982	1.75	neyman_chi_square_divergence	1	0.496	78.714	0.679	-0.254	124.268	1.071
H AI	8	min_contrast_4	1	0.279	199.78	1.718	neyman_chi_square_divergence	1	0.494	78.26	0.673	-0.215	121.52	1.045
H AI	9	min_contrast_4	1	0.215	205.101	1.748	neyman_chi_square_divergence	1	0.513	79.514	0.678	-0.298	125.587	1.07
H AI	10	min_contrast_2	1	0.205	206.158	1.773	neyman_chi_square_divergence	1	0.465	80.354	0.691	-0.26	125.804	1.082
SC H	1	neyman_chi_square_divergence	1	0.077	152.84	0.696	laplace_distribution	10	0.232	100.658	0.459	-0.155	52.182	0.237
SC H	2	neyman_chi_square_divergence	1	0.081	151.254	0.686	laplace_distribution	10	0.258	99.444	0.451	-0.177	51.81	0.235
SC H	3	neyman_chi_square_divergence	1	0.058	157.618	0.71	laplace_distribution	10	0.253	105.421	0.475	-0.195	52.197	0.235
SC H	4	neyman_chi_square_divergence	1	0.081	156.851	0.705	laplace_distribution	10	0.252	105.525	0.475	-0.171	51.326	0.23
SC H	5	neyman_chi_square_divergence	1	0.096	150.807	0.685	laplace_distribution	10	0.25	99.303	0.451	-0.154	51.504	0.234
SC H	6	neyman_chi_square_divergence	1	0.1	154.547	0.695	laplace_distribution	10	0.265	104.778	0.471	-0.165	49.769	0.224
SC H	7	neyman_chi_square_divergence	1	0.058	156.755	0.703	laplace_distribution	10	0.237	107.39	0.481	-0.179	49.365	0.222
SC H	8	neyman_chi_square_divergence	1	0.081	156.481	0.702	laplace_distribution	10	0.249	105.4	0.473	-0.168	51.081	0.229
SC H	9	neyman_chi_square_divergence	1	0.078	157.069	0.705	laplace_distribution	10	0.234	107.052	0.48	-0.156	50.017	0.225
SC H	10	neyman_chi_square_divergence	1	0.075	152.034	0.683	laplace_distribution	10	0.235	106.999	0.481	-0.16	45.035	0.202
US	1	min_contrast_4	10	0.446	90.501	0.345	pearson_chi_square	0.01	0.11	179.487	0.684	0.336	-88.986	-0.339
US	2	min_contrast_4	10	0.463	83.106	0.32	pearson_chi_square	0.01	0.102	173.848	0.67	0.361	-90.742	-0.35
US	3	min_contrast_4	10	0.424	93.415	0.356	min_contrast_4	10	0.411	182.307	0.695	0.013	88.892	0.339
US	4	min_contrast_4	10	0.408	93.22	0.355	min_contrast_4	10	0.394	182.016	0.694	0.014	88.796	0.339
US	5	min_contrast_4	10	0.46	93.352	0.356	pearson_chi_square	0.01	0.119	176.744	0.674	0.341	-83.392	0.318
US	6	min_contrast_4	10	0.48	84.642	0.326	pearson_chi_square	0.01	0.101	174.832	0.673	0.379	-90.19	0.347
US	7	min_contrast_4	10	0.434	90.616	0.345	min_contrast_4	10	0.419	180.943	0.689	0.015	-90.327	0.344
US	8	min_contrast_4	10	0.426	91.647	0.35	min_contrast_4	10	0.417	181.294	0.692	0.009	-89.647	0.342
US	9	min_contrast_4	10	0.446	92.044	0.351	min_contrast_4	10	0.434	181.748	0.693	0.012	-89.704	0.342
US	10	min_contrast_4	10	0.356	93.236	0.357	pearson_chi_square	0.01	0.117	175.263	0.671	0.239	-82.027	0.314

294

295

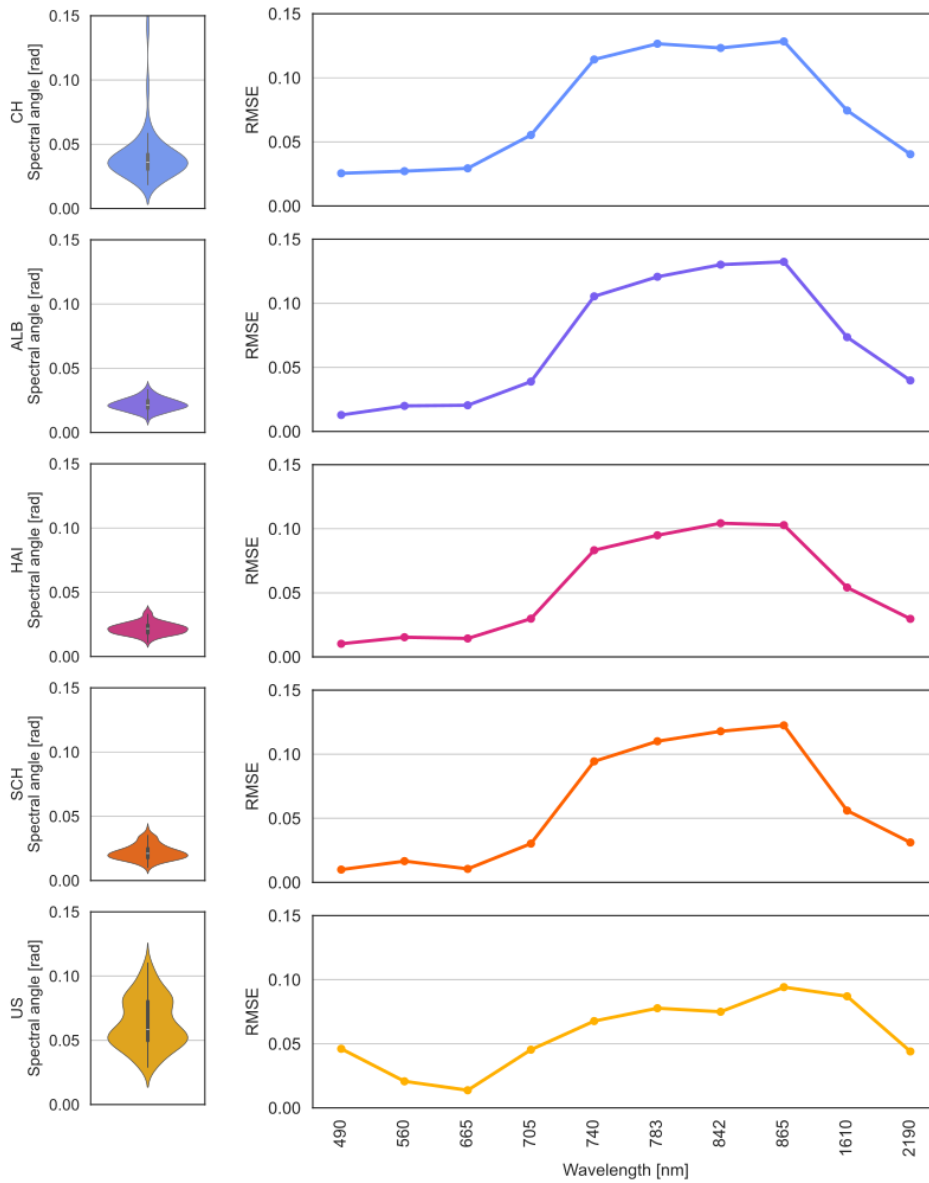
296 Table A.17: Testing performance of local and transferred physically-based models with US-tailored look-up table (LUT). The
 297 parameters correspond to the best-performing combination of parameters tested during the fivefold cross-validation. CH:
 298 Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States, R^2 : coefficient of
 299 determination, RMSE: root-mean-square error, sRMSE: systematic component of RMSE, uRMSE: unsystematic component of
 300 RMSE, RRMSE: relative root-mean-square error, MBE: mean bias error.

		Calibration site				
Prediction site	Metric	CH	ALB	HAI	SCH	US
CH	R^2	0.19 ± 0.13	0.00 ± 0.00	0.04 ± 0.01	0.08 ± 0.00	0.08 ± 0.00
CH	RMSE	178.67 ± 17.06	243.85 ± 1.43	205.20 ± 3.28	191.83 ± 0.00	209.39 ± 0.00
CH	sRMSE	174.39 ± 16.33	190.96 ± 1.14	192.41 ± 1.93	184.07 ± 0.00	197.37 ± 0.00
CH	uRMSE	50.21 ± 16.79	151.61 ± 3.64	71.24 ± 4.32	54.00 ± 0.00	69.92 ± 0.00
CH	RRMSE	1.02 ± 0.11	1.40 ± 0.01	1.18 ± 0.02	1.10 ± 0.00	1.21 ± 0.00
CH	MBE	131.38 ± 26.00	155.84 ± 1.40	157.81 ± 2.34	147.59 ± 0.00	163.87 ± 0.00
ALB	R^2	0.39 ± 0.00	0.03 ± 0.06	0.23 ± 0.00	0.28 ± 0.00	0.38 ± 0.00
ALB	RMSE	247.91 ± 0.00	184.72 ± 26.81	194.44 ± 0.70	200.32 ± 0.00	256.10 ± 0.00
ALB	sRMSE	247.01 ± 0.00	119.67 ± 19.71	195.75 ± 0.76	202.49 ± 0.00	255.04 ± 0.00
ALB	uRMSE	21.15 ± 0.00	140.06 ± 22.64	22.58 ± 0.61	29.60 ± 0.00	23.24 ± 0.00
ALB	RRMSE	1.78 ± 0.00	1.32 ± 0.19	1.40 ± 0.01	1.44 ± 0.00	1.84 ± 0.00
ALB	MBE	232.16 ± 0.00	82.11 ± 27.44	176.57 ± 0.85	184.05 ± 0.00	240.69 ± 0.00
HAI	R^2	0.48 ± 0.00	0.02 ± 0.00	0.31 ± 0.11	0.41 ± 0.00	0.46 ± 0.00
HAI	RMSE	251.88 ± 0.00	223.38 ± 0.41	195.40 ± 13.23	206.53 ± 0.00	258.13 ± 0.00
HAI	sRMSE	255.37 ± 0.00	146.79 ± 0.46	199.40 ± 11.86	213.06 ± 0.00	261.40 ± 0.00
HAI	uRMSE	42.12 ± 0.00	168.38 ± 0.15	36.22 ± 15.28	52.33 ± 0.00	41.20 ± 0.00
HAI	RRMSE	2.17 ± 0.00	1.93 ± 0.00	1.70 ± 0.13	1.78 ± 0.00	2.22 ± 0.00
HAI	MBE	240.75 ± 0.00	118.96 ± 0.56	180.96 ± 16.51	195.27 ± 0.00	247.13 ± 0.00
SCH	R^2	0.16 ± 0.00	0.01 ± 0.00	0.02 ± 0.00	0.12 ± 0.06	0.15 ± 0.00
SCH	RMSE	202.61 ± 0.00	171.44 ± 0.78	160.21 ± 0.80	153.17 ± 10.08	209.93 ± 0.00
SCH	sRMSE	204.89 ± 0.00	121.59 ± 0.04	155.64 ± 0.39	156.26 ± 10.45	211.96 ± 0.00
SCH	uRMSE	30.48 ± 0.00	120.86 ± 1.08	37.93 ± 1.77	33.74 ± 13.43	29.31 ± 0.00
SCH	RRMSE	0.91 ± 0.00	0.77 ± 0.00	0.72 ± 0.00	0.69 ± 0.04	0.95 ± 0.00
SCH	MBE	166.86 ± 0.00	-23.08 ± 0.08	100.12 ± 0.59	99.35 ± 9.47	175.46 ± 0.00
US	R^2	0.42 ± 0.00	0.24 ± 0.04	0.44 ± 0.00	0.44 ± 0.00	0.51 ± 0.12
US	RMSE	110.88 ± 0.00	125.92 ± 2.43	92.36 ± 1.66	100.72 ± 0.00	87.52 ± 13.13
US	sRMSE	123.50 ± 0.00	127.06 ± 0.43	107.97 ± 1.56	115.29 ± 0.00	104.15 ± 13.18
US	uRMSE	54.39 ± 0.00	11.22 ± 12.61	55.91 ± 0.30	56.10 ± 0.00	55.88 ± 8.20
US	RRMSE	0.42 ± 0.00	0.48 ± 0.01	0.35 ± 0.01	0.39 ± 0.00	0.33 ± 0.05
US	MBE	-89.48 ± 0.00	-94.19 ± 0.62	-66.41 ± 2.51	-77.77 ± 0.00	-61.92 ± 7.33

301

302

303 **Section A.20: PROSAIL forward model fit**



304

305 *Figure 24: Violin plots show the smallest spectral angle (Yuhás et al., 1992) between the Sentinel-2 spectra and the PROSAIL*
 306 *simulations by site. Line plots show the root-mean-square error (RMSE) for the Sentinel-2 spectra and the closest PROSAIL*
 307 *simulations in terms of smallest spectral angle for each spectral band. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-*
 308 *Dün, SCH: Schorfheide-Chorin, US: United States.*

309

310 **Section A.21: Precipitation at US site**

311 *Table A.1818: Summarized precipitation data in millimeters [mm] at the Norman (OK) weather station as provided by Mesonet*
 312 *(2024) for the summer months (June to August) from 2004-2023. * denotes the months in which the field measurements used*
 313 *in this study were conducted.*

Year	Monthly precipitation [mm]			Cumulative precipitation
	June	July	August	
2004	8.11	4.37	3.55	16.03
2005	3.51	2.60	4.92	11.03
2006	2.63	2.05	2.15	6.83
2007	10.82	8.08	6.09	24.99
2008	5.90	0.76	10.26	16.92
2009	1.27	3.56	4.59	9.42
2010	4.03	5.55	0.72	10.30
2011	2.35	0.34	2.06	4.75
2012	0.82	0.02	3.14	3.98
2013	4.16	9.56	2.73	16.45
2014	4.58	3.76	1.34	9.68
2015	5.95	7.46	1.74	15.15
2016	2.93	6.48	0.51	9.92
2017	0.66	2.43	8.43	11.52
2018	6.55	2.28	6.87	15.70
2019	5.67	0.18	6.61	12.46
2020	2.19	2.65	5.15	9.99
2021	6.54	2.70	1.46	10.70
2022	5.39	1.19 *	0.84 *	7.42
2023	5.13	7.62	0.59	13.34
Average	4.46	3.68	3.69	11.83

314

315

316 **Section A.22: C3 and C4 species per study site**

317 Information about the photosynthetic pathway of the species in the study sites investigated in
318 this study were estimated from the TRY database (Cornelissen, 1996; Cornelissen et al., 2004;
319 Cornwell et al., 2017, 2008; Craine et al., 2005; Diaz et al., 2004; Fitter and Peat, 1994; Flowers
320 et al., 2023, 2017; Iversen et al., 2017; Kapralov et al., 2012; Kattge et al., 2020, 2009; Laughlin
321 et al., 2011, 2010; Lin et al., 2015; Meir et al., 2002; Munroe et al., 2021; Poschlod et al., 2003;
322 Quested et al., 2003; Reich et al., 2009, 2008; Smith and Dukes, 2017; Wang et al., 2017;
323 Wright et al., 2004). We point out that the percentages indicated in Table A.19 relate to species
324 occurrence and not species abundance. In fact, the abundance of C4 species was considerably
325 higher for the US site. However, species abundance data was not available for all sites which is
326 why we did not add a separate column.

327 *Table A.19: Number of C3 and C4 species, and relative share of C4 species per study site. CH: Switzerland, ALB: Schwäbische*
328 *Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States.*

Study site	Number of C3 species	Number of C4 species	Relative share of C4 species [%]
CH	152	4	2.56
ALB	42	0	0.00
HAI	80	1	1.23
SCH	174	8	4.40
US	156	23	12.85

329

330 **Section A.23: Transferability of local hybrid models with varying initial training set size**

331 *Table A.20: Testing performance of local and transferred hybrid models. The parameters correspond to the best-performing*
 332 *combination of parameters tested during the fivefold cross-validation using Active Learning with an initial training set size of*
 333 *1%. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States, R²: coefficient of*
 334 *determination, RMSE: root-mean-square error, sRMSE: systematic component of RMSE, uRMSE: unsystematic component of*
 335 *RMSE, RRMSE: relative root-mean-square error, MBE: mean bias error.*

Initial training set size = 1%		Calibration site				
Prediction site	Metric	CH	ALB	HAI	SCH	US
CH	R ²	0.25 ± 0.19	0.03 ± 0.02	0.06 ± 0.06	0.02 ± 0.02	0.01 ± 0.02
CH	RMSE	103.69 ± 18.85	129.44 ± 8.99	134.83 ± 21.99	142.12 ± 9.30	191.62 ± 63.09
CH	sRMSE	113.31 ± 9.08	119.74 ± 9.18	122.02 ± 12.26	121.00 ± 9.74	171.73 ± 52.16
CH	uRMSE	47.11 ± 27.93	45.47 ± 18.60	52.98 ± 38.71	72.81 ± 15.75	71.14 ± 58.50
CH	RRMSE	0.59 ± 0.10	0.75 ± 0.05	0.78 ± 0.13	0.82 ± 0.05	1.10 ± 0.36
CH	MBE	-2.72 ± 13.49	-41.46 ± 23.45	-40.36 ± 35.62	5.02 ± 50.57	126.33 ± 63.89
ALB	R ²	0.08 ± 0.10	0.49 ± 0.17	0.32 ± 0.10	0.40 ± 0.07	0.06 ± 0.06
ALB	RMSE	126.21 ± 16.94	62.87 ± 10.57	74.22 ± 7.02	111.74 ± 12.28	162.29 ± 23.31
ALB	sRMSE	108.82 ± 16.12	85.15 ± 4.63	85.61 ± 0.94	120.57 ± 8.51	143.02 ± 14.02
ALB	uRMSE	59.38 ± 24.24	54.70 ± 14.70	41.52 ± 11.89	43.36 ± 9.64	61.41 ± 49.65
ALB	RRMSE	0.91 ± 0.12	0.45 ± 0.07	0.53 ± 0.05	0.80 ± 0.09	1.17 ± 0.17
ALB	MBE	65.03 ± 25.47	0.44 ± 13.69	-12.72 ± 6.51	85.76 ± 11.92	114.26 ± 18.83
HAI	R ²	0.06 ± 0.06	0.37 ± 0.10	0.51 ± 0.13	0.26 ± 0.08	0.08 ± 0.09
HAI	RMSE	135.73 ± 8.34	70.17 ± 7.51	60.93 ± 11.56	112.65 ± 10.12	188.68 ± 31.60
HAI	sRMSE	118.33 ± 13.62	86.30 ± 0.99	82.65 ± 9.90	116.60 ± 4.41	167.83 ± 26.43
HAI	uRMSE	62.44 ± 20.15	48.49 ± 10.83	54.56 ± 10.28	36.40 ± 12.60	70.53 ± 52.63
HAI	RRMSE	1.17 ± 0.07	0.60 ± 0.06	0.53 ± 0.09	0.97 ± 0.09	1.63 ± 0.27
HAI	MBE	79.76 ± 21.29	1.45 ± 13.36	-5.70 ± 6.34	79.31 ± 6.29	143.04 ± 31.57
SCH	R ²	0.04 ± 0.06	0.25 ± 0.05	0.17 ± 0.07	0.29 ± 0.12	0.07 ± 0.08
SCH	RMSE	139.32 ± 13.87	128.55 ± 9.52	143.68 ± 8.58	108.25 ± 12.61	151.20 ± 25.32
SCH	sRMSE	127.16 ± 10.54	141.13 ± 6.90	150.62 ± 4.83	120.60 ± 11.47	125.22 ± 2.90
SCH	uRMSE	54.08 ± 19.89	57.44 ± 6.97	42.75 ± 15.06	58.95 ± 11.45	69.11 ± 55.27
SCH	RRMSE	0.63 ± 0.06	0.58 ± 0.04	0.65 ± 0.04	0.49 ± 0.05	0.68 ± 0.11
SCH	MBE	-33.08 ± 31.21	-75.39 ± 12.86	-92.23 ± 8.13	0.41 ± 12.44	24.95 ± 28.13
US	R ²	0.11 ± 0.12	0.10 ± 0.08	0.18 ± 0.13	0.19 ± 0.10	0.30 ± 0.25
US	RMSE	226.44 ± 69.67	204.36 ± 51.18	197.08 ± 52.07	148.20 ± 37.87	72.09 ± 14.27
US	sRMSE	210.86 ± 61.49	200.61 ± 51.46	185.05 ± 60.75	125.16 ± 38.69	84.44 ± 12.64
US	uRMSE	76.37 ± 45.34	37.57 ± 20.66	58.73 ± 27.86	72.22 ± 36.25	34.16 ± 27.01
US	RRMSE	0.87 ± 0.27	0.78 ± 0.20	0.75 ± 0.20	0.57 ± 0.14	0.28 ± 0.05
US	MBE	-190.17 ± 68.45	-179.42 ± 57.72	-159.16 ± 72.14	-79.48 ± 58.63	1.16 ± 10.21

336

337

338 Table A.21: Testing performance of local and transferred hybrid models. The parameters correspond to the best-performing
 339 combination of parameters tested during the fivefold cross-validation using Active Learning with an initial training set size of
 340 2%. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States, R^2 : coefficient of
 341 determination, RMSE: root-mean-square error, sRMSE: systematic component of RMSE, uRMSE: unsystematic component of
 342 RMSE, RRMSE: relative root-mean-square error, MBE: mean bias error.

Initial training set size = 2%		Calibration site				
Prediction site	Metric	CH	ALB	HAI	SCH	US
CH	R^2	0.23 ± 0.14	0.02 ± 0.02	0.02 ± 0.03	0.03 ± 0.03	0.01 ± 0.01
CH	RMSE	112.22 ± 31.13	127.07 ± 7.94	125.43 ± 9.89	136.69 ± 13.00	164.02 ± 35.27
CH	sRMSE	113.87 ± 9.18	116.51 ± 7.64	117.34 ± 6.45	125.35 ± 13.19	146.02 ± 14.56
CH	uRMSE	54.73 ± 39.46	49.79 ± 13.76	44.33 ± 15.06	53.38 ± 17.36	58.53 ± 64.41
CH	RRMSE	0.64 ± 0.18	0.73 ± 0.05	0.72 ± 0.06	0.79 ± 0.07	0.94 ± 0.20
CH	MBE	-3.46 ± 17.01	-31.66 ± 21.59	-36.42 ± 18.12	36.96 ± 48.54	93.61 ± 20.14
ALB	R^2	0.07 ± 0.07	0.46 ± 0.18	0.34 ± 0.07	0.41 ± 0.09	0.03 ± 0.02
ALB	RMSE	122.05 ± 8.44	64.39 ± 11.93	72.30 ± 4.12	107.21 ± 8.09	158.60 ± 28.14
ALB	sRMSE	103.03 ± 10.46	84.85 ± 4.76	86.83 ± 1.25	118.81 ± 4.89	147.49 ± 33.86
ALB	uRMSE	61.51 ± 21.43	54.13 ± 12.30	47.60 ± 5.58	50.35 ± 6.74	42.55 ± 41.87
ALB	RRMSE	0.88 ± 0.06	0.46 ± 0.08	0.52 ± 0.03	0.77 ± 0.06	1.14 ± 0.20
ALB	MBE	55.91 ± 19.27	-0.37 ± 11.47	-19.74 ± 5.34	83.55 ± 6.82	117.33 ± 42.37
HAI	R^2	0.03 ± 0.03	0.32 ± 0.04	0.54 ± 0.11	0.25 ± 0.06	0.02 ± 0.03
HAI	RMSE	131.67 ± 12.72	72.81 ± 1.65	58.51 ± 11.47	110.02 ± 5.09	179.92 ± 31.30
HAI	sRMSE	112.79 ± 15.17	86.15 ± 0.83	83.22 ± 9.76	115.72 ± 4.17	171.60 ± 35.17
HAI	uRMSE	65.35 ± 16.62	45.90 ± 3.40	58.49 ± 6.67	35.08 ± 7.02	40.01 ± 41.06
HAI	RRMSE	1.13 ± 0.11	0.63 ± 0.01	0.51 ± 0.09	0.95 ± 0.04	1.55 ± 0.27
HAI	MBE	70.71 ± 24.06	0.42 ± 12.18	-9.55 ± 8.77	78.01 ± 6.21	146.89 ± 38.34
SCH	R^2	0.02 ± 0.02	0.23 ± 0.04	0.17 ± 0.07	0.27 ± 0.15	0.02 ± 0.03
SCH	RMSE	142.90 ± 12.30	129.56 ± 8.21	144.81 ± 6.85	104.35 ± 14.22	144.01 ± 25.57
SCH	sRMSE	129.70 ± 11.57	141.33 ± 6.42	152.09 ± 3.38	120.30 ± 11.46	132.76 ± 16.64
SCH	uRMSE	58.91 ± 12.09	56.03 ± 4.71	44.91 ± 12.62	56.42 ± 18.15	46.69 ± 44.79
SCH	RRMSE	0.64 ± 0.06	0.58 ± 0.04	0.65 ± 0.03	0.47 ± 0.06	0.65 ± 0.12
SCH	MBE	-44.35 ± 27.74	-75.77 ± 12.38	-94.74 ± 5.45	-0.14 ± 8.90	39.24 ± 43.46
US	R^2	0.13 ± 0.10	0.12 ± 0.10	0.14 ± 0.11	0.19 ± 0.05	0.23 ± 0.26
US	RMSE	184.27 ± 27.00	161.10 ± 38.01	155.40 ± 9.64	118.22 ± 23.46	76.40 ± 20.77
US	sRMSE	171.80 ± 24.72	159.96 ± 38.99	148.05 ± 10.26	106.73 ± 19.36	84.04 ± 13.67
US	uRMSE	63.41 ± 27.27	33.24 ± 13.22	45.41 ± 12.44	52.36 ± 23.27	22.45 ± 24.84
US	RRMSE	0.70 ± 0.10	0.62 ± 0.15	0.59 ± 0.04	0.45 ± 0.09	0.29 ± 0.07
US	MBE	-148.09 ± 28.52	-132.65 ± 46.57	-120.22 ± 12.54	-48.41 ± 44.96	0.49 ± 7.92

343

344

345 Table A.22: Testing performance of local and transferred hybrid models. The parameters correspond to the best-performing
 346 combination of parameters tested during the fivefold cross-validation using Active Learning with an initial training set size of
 347 5%. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States, R^2 : coefficient of
 348 determination, RMSE: root-mean-square error, sRMSE: systematic component of RMSE, uRMSE: unsystematic component of
 349 RMSE, RRMSE: relative root-mean-square error, MBE: mean bias error.

Initial training set size = 5%		Calibration site				
Prediction site	Metric	CH	ALB	HAI	SCH	US
CH	R^2	0.15 ± 0.15	0.02 ± 0.02	0.01 ± 0.01	0.02 ± 0.02	0.01 ± 0.01
CH	RMSE	111.07 ± 19.79	120.25 ± 4.85	120.99 ± 5.25	140.38 ± 6.01	143.60 ± 7.80
CH	sRMSE	114.20 ± 9.45	112.32 ± 1.75	115.70 ± 2.65	134.66 ± 4.40	141.40 ± 1.41
CH	uRMSE	44.88 ± 23.47	39.92 ± 16.58	31.77 ± 16.23	37.52 ± 13.52	10.75 ± 24.04
CH	RRMSE	0.63 ± 0.11	0.69 ± 0.03	0.70 ± 0.03	0.81 ± 0.03	0.83 ± 0.04
CH	MBE	-13.64 ± 13.37	-18.65 ± 10.31	-34.13 ± 8.57	76.89 ± 7.50	88.51 ± 2.21
ALB	R^2	0.04 ± 0.04	0.44 ± 0.16	0.30 ± 0.08	0.37 ± 0.09	0.02 ± 0.04
ALB	RMSE	113.02 ± 16.52	64.96 ± 10.37	73.35 ± 4.82	106.73 ± 8.13	152.55 ± 10.94
ALB	sRMSE	97.54 ± 10.32	84.72 ± 4.70	86.16 ± 1.01	118.23 ± 5.02	144.84 ± 12.28
ALB	uRMSE	53.11 ± 24.62	55.35 ± 8.69	44.39 ± 7.09	50.06 ± 6.30	16.82 ± 44.59
ALB	RRMSE	0.81 ± 0.12	0.46 ± 0.07	0.53 ± 0.03	0.77 ± 0.06	1.10 ± 0.08
ALB	MBE	44.89 ± 19.52	1.22 ± 10.29	-16.04 ± 6.17	82.66 ± 7.32	116.72 ± 16.84
HAI	R^2	0.05 ± 0.06	0.26 ± 0.08	0.46 ± 0.13	0.22 ± 0.07	0.02 ± 0.02
HAI	RMSE	125.35 ± 18.84	76.21 ± 4.91	63.10 ± 11.90	107.83 ± 7.36	171.78 ± 8.86
HAI	sRMSE	106.93 ± 9.33	85.54 ± 0.22	83.12 ± 9.67	113.64 ± 5.75	166.19 ± 8.06
HAI	uRMSE	61.60 ± 28.97	37.05 ± 10.57	52.57 ± 10.79	33.82 ± 11.05	16.02 ± 40.76
HAI	RRMSE	1.08 ± 0.16	0.66 ± 0.04	0.55 ± 0.09	0.93 ± 0.06	1.48 ± 0.08
HAI	MBE	62.55 ± 15.42	1.85 ± 5.40	-4.17 ± 11.12	74.69 ± 8.72	142.14 ± 9.74
SCH	R^2	0.02 ± 0.03	0.19 ± 0.05	0.13 ± 0.04	0.22 ± 0.08	0.01 ± 0.01
SCH	RMSE	142.94 ± 9.99	134.32 ± 4.59	147.07 ± 5.50	108.28 ± 11.79	131.14 ± 16.47
SCH	sRMSE	130.85 ± 5.26	143.28 ± 3.57	152.39 ± 3.78	121.01 ± 11.26	125.12 ± 1.57
SCH	uRMSE	54.97 ± 22.31	49.01 ± 8.91	38.37 ± 10.96	52.54 ± 12.04	15.61 ± 39.65
SCH	RRMSE	0.64 ± 0.05	0.61 ± 0.02	0.66 ± 0.02	0.49 ± 0.04	0.59 ± 0.07
SCH	MBE	-51.63 ± 16.42	-79.81 ± 6.50	-95.12 ± 6.11	-5.58 ± 14.29	34.03 ± 16.48
US	R^2	0.09 ± 0.09	0.18 ± 0.06	0.20 ± 0.05	0.19 ± 0.03	0.11 ± 0.17
US	RMSE	147.63 ± 26.06	163.08 ± 11.32	152.60 ± 7.63	102.45 ± 8.77	82.00 ± 16.11
US	sRMSE	142.80 ± 19.56	151.49 ± 11.25	145.50 ± 7.02	89.77 ± 5.75	82.40 ± 10.50
US	uRMSE	43.32 ± 20.47	59.97 ± 7.16	45.30 ± 8.60	48.85 ± 9.79	10.56 ± 16.63
US	RRMSE	0.56 ± 0.10	0.62 ± 0.04	0.58 ± 0.03	0.39 ± 0.03	0.31 ± 0.06
US	MBE	-113.41 ± 23.40	-124.34 ± 13.64	-117.09 ± 8.69	-18.94 ± 17.36	-1.07 ± 5.78

350

351 Table A.23: Testing performance of local and transferred hybrid models. The parameters correspond to the best-performing
 352 combination of parameters tested during the fivefold cross-validation using Active Learning with an initial training set size of
 353 10%. CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United States, R^2 : coefficient of
 354 determination, RMSE: root-mean-square error, sRMSE: systematic component of RMSE, uRMSE: unsystematic component of
 355 RMSE, RRMSE: relative root-mean-square error, MBE: mean bias error.

Initial training set size = 10%		Calibration site				
Prediction site	Metric	CH	ALB	HAI	SCH	US
CH	R^2	0.17 ± 0.17	0.02 ± 0.01	0.02 ± 0.03	0.01 ± 0.02	0.01 ± 0.03
CH	RMSE	104.26 ± 15.57	117.78 ± 3.98	118.99 ± 2.31	138.46 ± 3.69	141.06 ± 0.68
CH	sRMSE	113.18 ± 8.79	111.64 ± 0.99	115.54 ± 2.58	135.89 ± 4.20	140.93 ± 0.76
CH	uRMSE	34.21 ± 26.09	35.90 ± 13.64	26.58 ± 11.38	25.48 ± 7.16	0.89 ± 2.34
CH	RRMSE	0.60 ± 0.08	0.68 ± 0.02	0.69 ± 0.01	0.80 ± 0.02	0.81 ± 0.00
CH	MBE	-4.20 ± 10.30	-17.59 ± 9.13	-33.37 ± 8.68	79.04 ± 7.44	87.96 ± 1.08
ALB	R^2	0.03 ± 0.04	0.32 ± 0.21	0.27 ± 0.09	0.29 ± 0.10	0.00 ± 0.00
ALB	RMSE	102.02 ± 10.17	71.02 ± 12.27	74.43 ± 3.86	110.90 ± 5.44	149.05 ± 0.90
ALB	sRMSE	94.03 ± 5.34	84.82 ± 4.60	85.93 ± 0.66	119.22 ± 3.30	148.91 ± 0.95
ALB	uRMSE	34.53 ± 24.13	41.66 ± 20.56	41.94 ± 8.43	42.80 ± 7.97	0.08 ± 0.20
ALB	RRMSE	0.73 ± 0.07	0.51 ± 0.08	0.53 ± 0.03	0.80 ± 0.04	1.07 ± 0.01
ALB	MBE	38.65 ± 12.74	1.15 ± 9.86	-15.28 ± 4.10	84.09 ± 4.60	122.46 ± 1.10
HAI	R^2	0.04 ± 0.06	0.19 ± 0.07	0.39 ± 0.15	0.18 ± 0.07	0.00 ± 0.01
HAI	RMSE	119.38 ± 14.55	81.05 ± 4.65	67.97 ± 13.90	111.40 ± 5.62	169.02 ± 0.99
HAI	sRMSE	105.85 ± 9.46	85.95 ± 0.67	82.95 ± 9.52	115.88 ± 3.43	168.84 ± 1.10
HAI	uRMSE	45.78 ± 32.95	29.93 ± 9.05	45.05 ± 11.36	29.84 ± 11.23	1.37 ± 2.93
HAI	RRMSE	1.03 ± 0.13	0.70 ± 0.04	0.59 ± 0.11	0.96 ± 0.05	1.46 ± 0.01
HAI	MBE	60.19 ± 17.35	1.81 ± 11.26	-4.00 ± 9.09	78.19 ± 5.01	145.55 ± 1.23
SCH	R^2	0.01 ± 0.01	0.18 ± 0.06	0.10 ± 0.04	0.19 ± 0.12	0.00 ± 0.01
SCH	RMSE	138.26 ± 10.83	139.34 ± 4.02	149.11 ± 4.27	110.25 ± 11.89	125.74 ± 0.36
SCH	sRMSE	130.54 ± 9.74	147.00 ± 3.88	153.22 ± 2.61	120.60 ± 11.67	125.65 ± 0.38
SCH	uRMSE	40.52 ± 21.39	47.67 ± 11.13	33.54 ± 10.32	45.77 ± 18.46	1.70 ± 2.33
SCH	RRMSE	0.62 ± 0.05	0.63 ± 0.02	0.67 ± 0.02	0.50 ± 0.05	0.57 ± 0.00
SCH	MBE	-47.97 ± 24.76	-86.04 ± 6.35	-96.47 ± 4.13	-7.42 ± 8.90	39.67 ± 1.17
US	R^2	0.05 ± 0.04	0.18 ± 0.06	0.15 ± 0.07	0.17 ± 0.04	0.08 ± 0.20
US	RMSE	142.09 ± 13.04	156.92 ± 11.19	158.49 ± 7.94	94.97 ± 3.44	84.34 ± 12.95
US	sRMSE	140.19 ± 10.94	148.43 ± 9.00	153.74 ± 8.54	87.67 ± 1.33	82.16 ± 10.98
US	uRMSE	24.30 ± 16.72	51.18 ± 16.01	37.17 ± 9.65	35.88 ± 7.41	1.63 ± 4.32
US	RRMSE	0.54 ± 0.05	0.60 ± 0.04	0.61 ± 0.03	0.36 ± 0.01	0.32 ± 0.04
US	MBE	-110.53 ± 13.95	-120.95 ± 10.53	-127.21 ± 10.27	-13.80 ± 8.37	0.28 ± 5.41

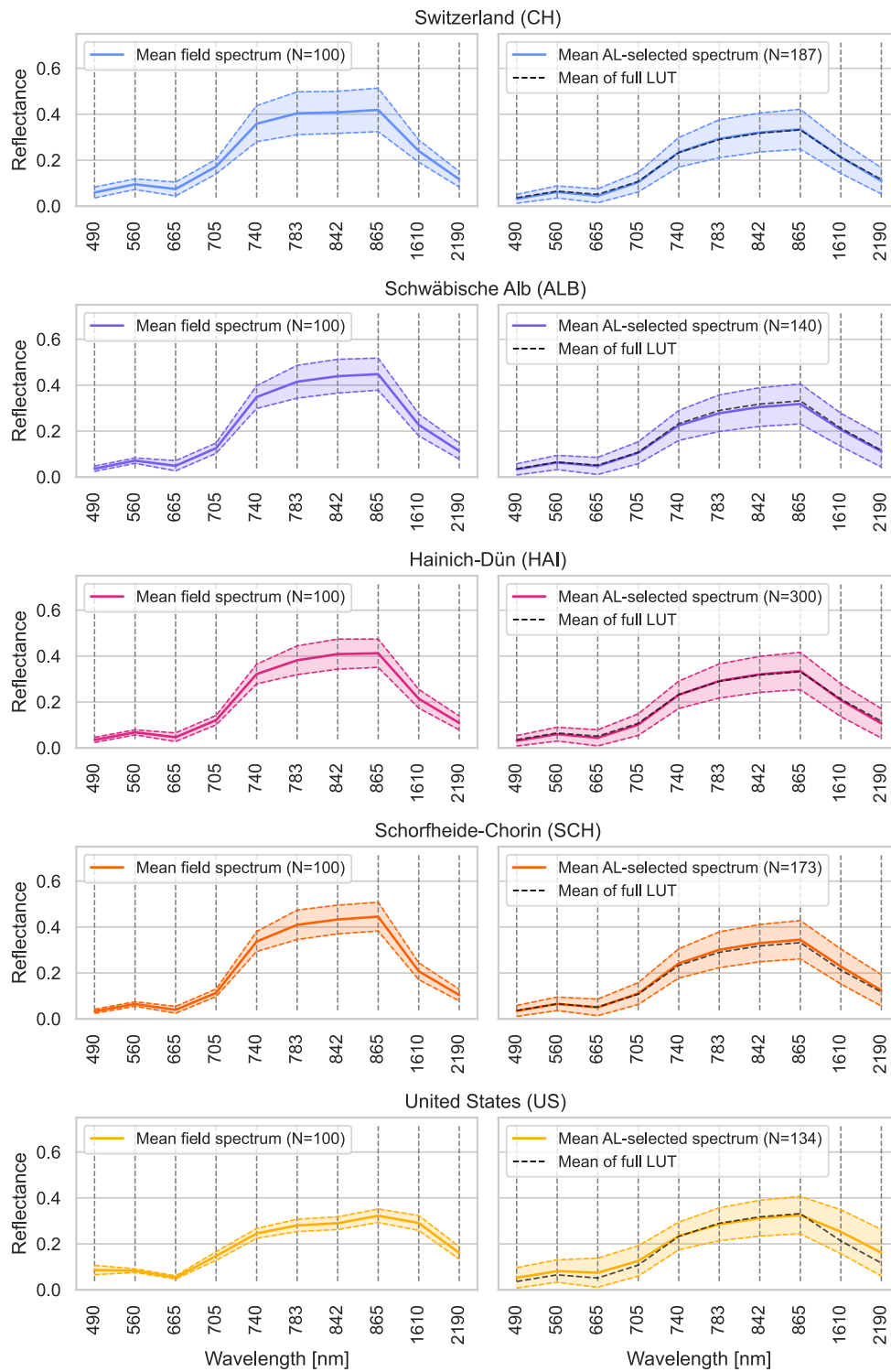
356

357 Table A.24: Testing performance of local and transferred hybrid models. The parameters correspond to the best-performing
 358 combination of parameters tested during the fivefold cross-validation using Active Learning with an initial training set size of
 359 100% (= no Active Learning). CH: Switzerland, ALB: Schwäbische Alb, HAI: Hainich-Dün, SCH: Schorfheide-Chorin, US: United
 360 States, R^2 : coefficient of determination, RMSE: root-mean-square error, sRMSE: systematic component of RMSE, uRMSE:
 361 unsystematic component of RMSE, RRMSE: relative root-mean-square error, MBE: mean bias error.

Initial training set size = 100% (= no Active Learning)		Calibration site				
Prediction site	Metric	CH	ALB	HAI	SCH	US
CH	R^2	0.07 ± 0.07	0.03 ± 0.00	0.03 ± 0.00	0.03 ± 0.00	0.03 ± 0.00
CH	RMSE	112.71 ± 8.67	109.87 ± 0.00	109.87 ± 0.00	109.87 ± 0.00	109.87 ± 0.00
CH	sRMSE	112.96 ± 8.67	110.10 ± 0.00	110.10 ± 0.00	110.10 ± 0.00	110.10 ± 0.00
CH	uRMSE	7.59 ± 4.63	7.18 ± 0.00	7.18 ± 0.00	7.18 ± 0.00	7.18 ± 0.00
CH	RRMSE	0.64 ± 0.04	0.63 ± 0.00	0.63 ± 0.00	0.63 ± 0.00	0.63 ± 0.00
CH	MBE	0.07 ± 3.45	0.99 ± 0.00	0.99 ± 0.00	0.99 ± 0.00	0.99 ± 0.00
ALB	R^2	0.06 ± 0.00	0.09 ± 0.05	0.06 ± 0.00	0.06 ± 0.00	0.06 ± 0.00
ALB	RMSE	89.47 ± 0.00	89.24 ± 3.93	89.47 ± 0.00	89.47 ± 0.00	89.47 ± 0.00
ALB	sRMSE	90.72 ± 0.00	90.41 ± 3.59	90.72 ± 0.00	90.72 ± 0.00	90.72 ± 0.00
ALB	uRMSE	14.99 ± 0.00	13.53 ± 5.02	14.99 ± 0.00	14.99 ± 0.00	14.99 ± 0.00
ALB	RRMSE	0.64 ± 0.00	0.64 ± 0.02	0.64 ± 0.00	0.64 ± 0.00	0.64 ± 0.00
ALB	MBE	32.42 ± 0.00	32.03 ± 2.86	32.42 ± 0.00	32.42 ± 0.00	32.42 ± 0.00
HAI	R^2	0.03 ± 0.00	0.03 ± 0.00	0.08 ± 0.06	0.03 ± 0.00	0.03 ± 0.00
HAI	RMSE	100.17 ± 0.00	100.17 ± 0.00	98.26 ± 7.91	100.17 ± 0.00	100.17 ± 0.00
HAI	sRMSE	100.91 ± 0.00	100.91 ± 0.00	99.17 ± 7.33	100.91 ± 0.00	100.91 ± 0.00
HAI	uRMSE	12.20 ± 0.00	12.20 ± 0.00	14.74 ± 2.59	12.20 ± 0.00	12.20 ± 0.00
HAI	RRMSE	0.86 ± 0.00	0.86 ± 0.00	0.85 ± 0.05	0.86 ± 0.00	0.86 ± 0.00
HAI	MBE	53.30 ± 0.00	53.30 ± 0.00	54.24 ± 2.70	53.30 ± 0.00	53.30 ± 0.00
SCH	R^2	0.08 ± 0.00	0.08 ± 0.00	0.08 ± 0.00	0.09 ± 0.07	0.08 ± 0.00
SCH	RMSE	128.15 ± 0.00	128.15 ± 0.00	128.15 ± 0.00	129.58 ± 12.06	128.15 ± 0.00
SCH	sRMSE	130.20 ± 0.00	130.20 ± 0.00	130.20 ± 0.00	131.46 ± 12.23	130.20 ± 0.00
SCH	uRMSE	23.03 ± 0.00	23.03 ± 0.00	23.03 ± 0.00	20.49 ± 8.84	23.03 ± 0.00
SCH	RRMSE	0.58 ± 0.00	0.58 ± 0.00	0.58 ± 0.00	0.58 ± 0.04	0.58 ± 0.00
SCH	MBE	-52.78 ± 0.00	-52.78 ± 0.00	-52.78 ± 0.00	-53.14 ± 5.05	-52.78 ± 0.00
US	R^2	0.02 ± 0.00	0.02 ± 0.00	0.02 ± 0.00	0.02 ± 0.00	0.07 ± 0.08
US	RMSE	122.36 ± 0.00	122.36 ± 0.00	122.36 ± 0.00	122.36 ± 0.00	121.22 ± 12.12
US	sRMSE	122.41 ± 0.00	122.41 ± 0.00	122.41 ± 0.00	122.41 ± 0.00	121.32 ± 12.08
US	uRMSE	3.69 ± 0.00	3.69 ± 0.00	3.69 ± 0.00	3.69 ± 0.00	4.87 ± 1.51
US	RRMSE	0.47 ± 0.00	0.47 ± 0.00	0.47 ± 0.00	0.47 ± 0.00	0.46 ± 0.04
US	MBE	-87.21 ± 0.00	-87.21 ± 0.00	-87.21 ± 0.00	-87.21 ± 0.00	-86.97 ± 4.31

362

363 **Section A.24: Mean field spectrum and AL-selected spectra by site**



364
 365 *Figure A.25: Line plots of selected field spectra (left columns) and Active Learning (AL) selected training spectra of the look-up*
 366 *table (LUT) of the best performing local hybrid model with an initial training set size of 2% in terms of lowest relative root-*
 367 *mean-square error (right column) by study site. Bold lines represent mean spectra and colored areas represent the mean*
 368 *spectra ± 1 standard deviation.*

369 **References**

- 370 Chen, T., Guestrin, C., 2016. XGBoost: A scalable tree boosting system. Proceedings of the ACM
371 SIGKDD International Conference on Knowledge Discovery and Data Mining 13-17-August-2016,
372 785–794. <https://doi.org/10.1145/2939672.2939785>
- 373 Combal, B., Baret, F., Weiss, M., Trubuil, A., Macé, D., Pragnère, A., Myneni, R., Knyazikhin, Y., Wang,
374 L., 2003. Retrieval of canopy biophysical variables from bidirectional reflectance. Remote Sens
375 Environ 84, 1–15. [https://doi.org/10.1016/S0034-4257\(02\)00035-4](https://doi.org/10.1016/S0034-4257(02)00035-4)
- 376 Cornelissen, J.H.C., 1996. An Experimental Comparison of Leaf Decomposition Rates in a Wide Range
377 of Temperate Plant Species and Types. *J Ecol* 84, 573–582. <https://doi.org/10.2307/2261479>
- 378 Cornelissen, J.H.C., Quested, H.M., Gwynn-Jones, D., Van Logtestijn, R.S.P., De Beus, M.A.H.,
379 Kondratchuk, A., Callaghan, T. V., Aerts, R., 2004. Leaf Digestibility and Litter Decomposability
380 Are Related in a Wide Range of Subarctic Plant Species and Types. *Funct Ecol* 18, 779–786.
- 381 Cornwell, W.K., Cornelissen, J.H.C., Amatangelo, K., Dorrepaal, E., Eviner, V.T., Godoy, O., Hobbie,
382 S.E., Hoorens, B., Kurokawa, H., Pérez-Harguindeguy, N., Quested, H.M., Santiago, L.S., Wardle,
383 D.A., Wright, I.J., Aerts, R., Allison, S.D., Van Bodegom, P., Brovkin, V., Chatain, A., Callaghan, T.
384 V., Díaz, S., Garnier, E., Gurvich, D.E., Kazakou, E., Klein, J.A., Read, J., Reich, P.B., Soudzilovskaia,
385 N.A., Vaieretti, M.V., Westoby, M., 2008. Plant species traits are the predominant control on
386 litter decomposition rates within biomes worldwide. *Ecol Lett* 11, 1065–1071.
387 <https://doi.org/10.1111/j.1461-0248.2008.01219.x>
- 388 Cornwell, W.K., Wright, I., Turner, J., Maire, V., Barbour, M., Cernusak, L., Dawson, T., Ellsworth, D.,
389 Farquhar, G., Griffiths, H., Keitel, C., Knohl, A., Reich, P., Williams, D., Bhaskar, R., Cornelissen,
390 J.H.C., Richards, A., Schmidt, S., Valladares, F., Körner, C., Schulze, E., Buchmann, N., Santiago,
391 L., 2017. A global dataset of leaf D13 C values.
- 392 Craine, J.M., Lee, W.G., Bond, W.J., Williams, R.J., Johnson, L.C., 2005. Environmental constraints on a
393 global relationship among leaf and root traits of grasses. *Ecology* 86, 12–19.
394 <https://doi.org/10.1890/04-1075>
- 395 Diaz, S., Hodgson, J.G., Thompson, K., Cabido, M., Cornelissen, J.H.C., Jalili, A., Montserrat-Martí, G.,
396 Grime, J.P., Zarrinkamar, F., Asri, Y., Band, S.R., Basconcelo, S., Castro-Díez, P., Funes, G.,
397 Hamzehee, B., Khoshnevi, M., Pérez-Harguindeguy, N., Pérez-Rantomé, M.C., Shirvany, F.A.,
398 Vendramini, F., Yazdani, S., Abbas-Azimi, R., Bogaard, A., Boustani, S., Charles, M., Dehghan, M.,
399 de Torres-Espuny, L., Falczuk, V., Guerrero-Campo, J., Hynd, A., Jones, G., Kowsary, E., Kazemi-
400 Saeed, F., Maestro-Martínez, M., Romo-Díez, A., Shaw, S., Siavash, B., Villar-Salvador, P., Zak,
401 M.R., 2004. The plant traits that drive ecosystems: Evidence from three continents. *Journal of*
402 *Vegetation Science* 15, 295–304. <https://doi.org/10.1111/j.1654-1103.2004.tb02266.x>
- 403 ESA, 2015. Sentinel-2 User Handbook Issue 1 Revision 2. Noordwijk.
- 404 Fitter, A.H., Peat, H.J., 1994. The Ecological Flora Database. *J Ecol* 82, 415–425.
405 <https://doi.org/10.2307/2261309>
- 406 Flowers, T., Santos, J., Jahns, M., Warburton, B., Reed, P., 2023. eHALOPH - Halophytes Database
407 (version 3.11) [WWW Document]. URL <http://www.sussex.ac.uk/affiliates/halophytes>,
408 (accessed 11.28.23).

- 409 Flowers, T., Santos, J., Jahns, M., Warburton, B., Reed, P., 2017. eHALOPH - Halophytes Database
 410 (version 3.11) [WWW Document]. URL <http://www.sussex.ac.uk/affiliates/halophytes>,
 411 (accessed 1.1.17).
- 412 Huete, A.R., Liu, H., van Leeuwen, W.J.D., 1997. The use of vegetation indices in forested regions:
 413 issues of linearity and saturation, in: IGARSS'97. 1997 IEEE International Geoscience and
 414 Remote Sensing Symposium Proceedings. Remote Sensing - A Scientific Vision for Sustainable
 415 Development. IEEE, pp. 1966–1968. <https://doi.org/10.1109/IGARSS.1997.609169>
- 416 Iversen, C.M., McCormack, M.L., Powell, A.S., Blackwood, C.B., Freschet, G.T., Kattge, J., Roumet, C.,
 417 Stover, D.B., Soudzilovskaia, N.A., Valverde-Barrantes, O.J., van Bodegom, P.M., Violle, C., 2017.
 418 A global Fine-Root Ecology Database to address below-ground challenges in plant ecology. *New*
 419 *Phytologist* 215, 15–26. <https://doi.org/10.1111/nph.14486>
- 420 Jehle, M., Hueni, A., Damm, A., D'Odorico, P., Weyermann, J., Kneubihler, M., Schlapfer, D.,
 421 Schaepman, M.E., Meuleman, K., 2010. APEX - current status, performance and validation
 422 concept, in: 2010 IEEE Sensors. IEEE, pp. 533–537.
 423 <https://doi.org/10.1109/ICSENS.2010.5690122>
- 424 Kapralov, M. V., Smith, J.A.C., Filatov, D.A., 2012. Rubisco Evolution in C4 Eudicots: An Analysis of
 425 *Amaranthaceae Ssensu Lato*. *PLoS One* 7, e52974.
 426 <https://doi.org/10.1371/journal.pone.0052974>
- 427 Kattge, J., Bönisch, G., Díaz, S., Lavorel, S., Prentice, I.C., Leadley, P., Tautenhahn, S., Werner, G.D.A.,
 428 Aakala, T., Abedi, M., Acosta, A.T.R., Adamidis, G.C., Adamson, K., Aiba, M., Albert, C.H.,
 429 Alcántara, J.M., Alcázar C, C., Aleixo, I., Ali, H., Amiaud, B., Ammer, C., Amoroso, M.M., Anand,
 430 M., Anderson, C., Anten, N., Antos, J., Apgaua, D.M.G., Ashman, T., Asmara, D.H., Asner, G.P.,
 431 Aspinwall, M., Atkin, O., Aubin, I., Baastrup-Spohr, L., Bahalkeh, K., Bahn, M., Baker, T., Baker,
 432 W.J., Bakker, J.P., Baldocchi, D., Baltzer, J., Banerjee, A., Baranger, A., Barlow, J., Barneche, D.R.,
 433 Baruch, Z., Bastianelli, D., Battles, J., Bauerle, W., Bauters, M., Bazzato, E., Beckmann, M.,
 434 Beeckman, H., Beierkuhnlein, C., Bekker, R., Belfry, G., Belluau, M., Beloiu, M., Benavides, R.,
 435 Benomar, L., Berdugo-Lattke, M.L., Berenguer, E., Bergamin, R., Bergmann, J., Bergmann
 436 Carlucci, M., Berner, L., Bernhardt-Römermann, M., Bigler, C., Bjorkman, A.D., Blackman, C.,
 437 Blanco, C., Blonder, B., Blumenthal, D., Bocanegra-González, K.T., Boeckx, P., Bohlman, S.,
 438 Böhning-Gaese, K., Boisvert-Marsh, L., Bond, W., Bond-Lamberty, B., Boom, A., Boonman,
 439 C.C.F., Bordin, K., Boughton, E.H., Boukili, V., Bowman, D.M.J.S., Bravo, S., Brendel, M.R.,
 440 Broadley, M.R., Brown, K.A., Bruelheide, H., Brumnich, F., Bruun, H.H., Bruy, D., Buchanan, S.W.,
 441 Bucher, S.F., Buchmann, N., Buitenwerf, R., Bunker, D.E., Bürger, J., Burrascano, S., Burslem,
 442 D.F.R.P., Butterfield, B.J., Byun, C., Marques, M., Scalon, M.C., Caccianiga, M., Cadotte, M.,
 443 Cailleret, M., Camac, J., Camarero, J.J., Company, C., Campetella, G., Campos, J.A., Cano-
 444 Arboleda, L., Canullo, R., Carbognani, M., Carvalho, F., Casanoves, F., Castagnyrol, B., Catford,
 445 J.A., Cavender-Bares, J., Cerabolini, B.E.L., Cervellini, M., Chacón-Madrigal, E., Chapin, K.,
 446 Chapin, F.S., Chelli, S., Chen, S., Chen, A., Cherubini, P., Chianucci, F., Choat, B., Chung, K.,
 447 Chytrý, M., Ciccarelli, D., Coll, L., Collins, C.G., Conti, L., Coomes, D., Cornelissen, J.H.C.,
 448 Cornwell, W.K., Corona, P., Coyea, M., Craine, J., Craven, D., Crowsigt, J.P.G.M., Cseceserits, A.,
 449 Cufar, K., Cuntz, M., da Silva, A.C., Dahlin, K.M., Dainese, M., Dalke, I., Dalle Fratte, M., Dang-Le,
 450 A.T., Danihelka, J., Dannoura, M., Dawson, S., de Beer, A.J., De Frutos, A., De Long, J.R., Dechant,
 451 B., Delagrangé, S., Delpierre, N., Derroire, G., Dias, A.S., Diaz-Toribio, M.H., Dimitrakopoulos,
 452 P.G., Dobrowolski, M., Doktor, D., Dřevojan, P., Dong, N., Dransfield, J., Dressler, S., Duarte, L.,
 453 Ducouret, E., Dullinger, S., Durka, W., Duursma, R., Dymova, O., E-Vojtkó, A., Eckstein, R.L.,
 454 Ejtehadi, H., Elser, J., Emilio, T., Engemann, K., Erfanian, M.B., Erfmeier, A., Esquivel-Muelbert,

455 A., Esser, G., Estiarte, M., Domingues, T.F., Fagan, W.F., Fagúndez, J., Falster, D.S., Fan, Y., Fang,
456 J., Farris, E., Fazlioglu, F., Feng, Y., Fernandez-Mendez, F., Ferrara, C., Ferreira, J., Fidelis, A.,
457 Finegan, B., Firn, J., Flowers, T.J., Flynn, D.F.B., Fontana, V., Forey, E., Forgiarini, C., François, L.,
458 Frangipani, M., Frank, D., Frenette-Dussault, C., Freschet, G.T., Fry, E.L., Fyllas, N.M.,
459 Mazzochini, G.G., Gachet, S., Gallagher, R., Ganade, G., Ganga, F., García-Palacios, P.,
460 Gargaglione, V., Garnier, E., Garrido, J.L., de Gasper, A.L., Gea-Izquierdo, G., Gibson, D., Gillison,
461 A.N., Giroldo, A., Glasenhardt, M., Gleason, S., Gliesch, M., Goldberg, E., Gödel, B., Gonzalez-
462 Akre, E., Gonzalez-Andujar, J.L., González-Melo, A., González-Robles, A., Graae, B.J., Granda, E.,
463 Graves, S., Green, W.A., Gregor, T., Gross, N., Guerin, G.R., Günther, A., Gutiérrez, A.G.,
464 Haddock, L., Haines, A., Hall, J., Hambuckers, A., Han, W., Harrison, S.P., Hattingh, W., Hawes,
465 J.E., He, T., He, P., Heberling, J.M., Helm, A., Hempel, S., Hentschel, J., Hérault, B., Hereş, A.,
466 Herz, K., Heuertz, M., Hickler, T., Hietz, P., Higuchi, P., Hipp, A.L., Hiron, A., Hock, M., Hogan,
467 J.A., Holl, K., Honnay, O., Hornstein, D., Hou, E., Hough-Snee, N., Hovstad, K.A., Ichie, T., Igić, B.,
468 Illa, E., Isaac, M., Ishihara, M., Ivanov, L., Ivanova, L., Iversen, C.M., Izquierdo, J., Jackson, R.B.,
469 Jackson, B., Jactel, H., Jagodzinski, A.M., Jandt, U., Jansen, S., Jenkins, T., Jentsch, A., Jespersen,
470 J.R.P., Jiang, G., Johansen, J.L., Johnson, D., Jokela, E.J., Joly, C.A., Jordan, G.J., Joseph, G.S.,
471 Junaedi, D., Junker, R.R., Justes, E., Kabzems, R., Kane, J., Kaplan, Z., Kattenborn, T., Kavelenova,
472 L., Kearsley, E., Kempel, A., Kenzo, T., Kerkhoff, A., Khalil, M.I., Kinlock, N.L., Kissling, W.D.,
473 Kitajima, K., Kitzberger, T., Kjoller, R., Klein, T., Kleyer, M., Klimešová, J., Klipel, J., Kloeppel, B.,
474 Klotz, S., Knops, J.M.H., Kohyama, T., Koike, F., Kollmann, J., Komac, B., Komatsu, K., König, C.,
475 Kraft, N.J.B., Kramer, K., Kreft, H., Kühn, I., Kumarathunge, D., Kuppler, J., Kurokawa, H.,
476 Kurosawa, Y., Kuyah, S., Laclau, J., Lafleur, B., Lallai, E., Lamb, E., Lamprecht, A., Larkin, D.J.,
477 Laughlin, D., Le Bagousse-Pinguet, Y., le Maire, G., le Roux, P.C., le Roux, E., Lee, T., Lens, F.,
478 Lewis, S.L., Lhotsky, B., Li, Y., Li, X., Lichstein, J.W., Liebergesell, M., Lim, J.Y., Lin, Y., Linares, J.C.,
479 Liu, C., Liu, D., Liu, U., Livingstone, S., Llusà, J., Lohbeck, M., López-García, Á., Lopez-Gonzalez,
480 G., Lososová, Z., Louault, F., Lukács, B.A., Lukeš, P., Luo, Y., Lussu, M., Ma, S., Maciel Rabelo
481 Pereira, C., Mack, M., Maire, V., Mäkelä, A., Mäkinen, H., Malhado, A.C.M., Mallik, A., Manning,
482 P., Manzoni, S., Marchetti, Z., Marchino, L., Marcilio-Silva, V., Marcon, E., Marnani, M.,
483 Markesteijn, L., Martin, A., Martínez-Garza, C., Martínez-Vilalta, J., Mašková, T., Mason, K.,
484 Mason, N., Massad, T.J., Masse, J., Mayrose, I., McCarthy, J., McCormack, M.L., McCulloh, K.,
485 McFadden, I.R., McGill, B.J., McPartland, M.Y., Medeiros, J.S., Medlyn, B., Meerts, P., Mehrabi,
486 Z., Meir, P., Melo, F.P.L., Mencuccini, M., Meredieu, C., Messier, J., Mészáros, I., Metsaranta, J.,
487 Michaletz, S.T., Michelaki, C., Migalina, S., Milla, R., Miller, J.E.D., Minden, V., Ming, R., Mokany,
488 K., Moles, A.T., Molnár, A., Molofsky, J., Molz, M., Montgomery, R.A., Monty, A., Moravcová, L.,
489 Moreno-Martínez, A., Moretti, M., Mori, A.S., Mori, S., Morris, D., Morrison, J., Mucina, L.,
490 Mueller, S., Muir, C.D., Müller, S.C., Munoz, F., Myers-Smith, I.H., Myster, R.W., Nagano, M.,
491 Naidu, S., Narayanan, A., Natesan, B., Negoita, L., Nelson, A.S., Neuschulz, E.L., Ni, J., Niedrist,
492 G., Nieto, J., Niinemets, Ü., Nolan, R., Nottebrock, H., Nouvellon, Y., Novakovskiy, A., Nystuen,
493 K.O., O'Grady, A., O'Hara, K., O'Reilly-Nugent, A., Oakley, S., Oberhuber, W., Ohtsuka, T.,
494 Oliveira, R., Öllerer, K., Olson, M.E., Onipchenko, V., Onoda, Y., Onstein, R.E., Ordonez, J.C.,
495 Osada, N., Ostonen, I., Ottaviani, G., Otto, S., Overbeck, G.E., Ozinga, W.A., Pahl, A.T., Paine,
496 C.E.T., Pakeman, R.J., Papageorgiou, A.C., Parfionova, E., Pärtel, M., Patacca, M., Paula, S.,
497 Paule, J., Pauli, H., Pausas, J.G., Peco, B., Penuelas, J., Perea, A., Peri, P.L., Petisco-Souza, A.C.,
498 Petraglia, A., Petritan, A.M., Phillips, O.L., Pierce, S., Pillar, V.D., Pisek, J., Pomogaybin, A.,
499 Poorter, H., Portsmouth, A., Poschlod, P., Potvin, C., Pounds, D., Powell, A.S., Power, S.A.,
500 Prinzing, A., Puglielli, G., Pyšek, P., Raevel, V., Rammig, A., Ransijn, J., Ray, C.A., Reich, P.B.,
501 Reichstein, M., Reid, D.E.B., Réjou-Méchain, M., de Dios, V.R., Ribeiro, S., Richardson, S., Riibak,
502 K., Rillig, M.C., Riviera, F., Robert, E.M.R., Roberts, S., Robroek, B., Roddy, A., Rodrigues, A.V.,

503 Rogers, A., Rollinson, E., Rolo, V., Römermann, C., Ronzhina, D., Roscher, C., Rosell, J.A.,
504 Rosenfield, M.F., Rossi, C., Roy, D.B., Royer-Tardif, S., Rüger, N., Ruiz-Peinado, R., Rumpf, S.B.,
505 Rusch, G.M., Ryo, M., Sack, L., Saldaña, A., Salgado-Negret, B., Salguero-Gomez, R., Santa-
506 Regina, I., Santacruz-García, A.C., Santos, J., Sardans, J., Schamp, B., Scherer-Lorenzen, M.,
507 Schleuning, M., Schmid, B., Schmidt, M., Schmitt, S., Schneider, J. V., Schowanek, S.D., Schrader,
508 J., Schrod, F., Schuldt, B., Schurr, F., Selaya Garvizu, G., Semchenko, M., Seymour, C., Sfair, J.C.,
509 Sharpe, J.M., Sheppard, C.S., Sheremetiev, S., Shiodera, S., Shipley, B., Shovon, T.A., Siebenkäs,
510 A., Sierra, C., Silva, V., Silva, M., Sitzia, T., Sjöman, H., Slot, M., Smith, N.G., Sodhi, D., Soltis, P.,
511 Soltis, D., Somers, B., Sonnier, G., Sørensen, M.V., Sosinski, E.E., Soudzilovskaia, N.A., Souza,
512 A.F., Spasojevic, M., Sperandii, M.G., Stan, A.B., Stegen, J., Steinbauer, K., Stephan, J.G., Sterck,
513 F., Stojanovic, D.B., Strydom, T., Suarez, M.L., Svenning, J., Svitková, I., Svitok, M., Svoboda, M.,
514 Swaine, E., Swenson, N., Tabarelli, M., Takagi, K., Tappeiner, U., Tarifa, R., Tauougourdeau, S.,
515 Tavsanoğlu, C., te Beest, M., Tedersoo, L., Thiffault, N., Thom, D., Thomas, E., Thompson, K.,
516 Thornton, P.E., Thuiller, W., Tichý, L., Tissue, D., Tjoelker, M.G., Tng, D.Y.P., Tobias, J., Török, P.,
517 Tarin, T., Torres-Ruiz, J.M., Tóthmérész, B., Treurnicht, M., Trivellone, V., Trolliet, F., Trotsiuk, V.,
518 Tsakalos, J.L., Tsiripidis, I., Tysklind, N., Umehara, T., Usoltsev, V., Vadeboncoeur, M., Vaezi, J.,
519 Valladares, F., Vamosi, J., van Bodegom, P.M., van Breugel, M., Van Cleemput, E., van de Weg,
520 M., van der Merwe, S., van der Plas, F., van der Sande, M.T., van Kleunen, M., Van Meerbeek,
521 K., Vanderwel, M., Vanselow, K.A., Vårhammar, A., Varone, L., Vasquez Valderrama, M.Y.,
522 Vassilev, K., Vellend, M., Veneklaas, E.J., Verbeeck, H., Verheyen, K., Vibrans, A., Vieira, I.,
523 Villacís, J., Violle, C., Vivek, P., Wagner, K., Waldram, M., Waldron, A., Walker, A.P., Waller, M.,
524 Walther, G., Wang, H., Wang, F., Wang, W., Watkins, H., Watkins, J., Weber, U., Weedon, J.T.,
525 Wei, L., Weigelt, P., Weiher, E., Wells, A.W., Wellstein, C., Wenk, E., Westoby, M., Westwood,
526 A., White, P.J., Whitten, M., Williams, M., Winkler, D.E., Winter, K., Womack, C., Wright, I.J.,
527 Wright, S.J., Wright, J., Pinho, B.X., Ximenes, F., Yamada, T., Yamaji, K., Yanai, R., Yankov, N.,
528 Yguel, B., Zanini, K.J., Zanne, A.E., Zelený, D., Zhao, Y., Zheng, Jingming, Zheng, Ji, Ziemińska, K.,
529 Zirbel, C.R., Zizka, G., Zo-Bi, I.C., Zotz, G., Wirth, C., 2020. TRY plant trait database – enhanced
530 coverage and open access. *Glob Chang Biol* 26, 119–188. <https://doi.org/10.1111/gcb.14904>

531 Kattge, J., Knorr, W., Raddatz, T., Wirth, C., 2009. Quantifying photosynthetic capacity and its
532 relationship to leaf nitrogen content for global-scale terrestrial biosphere models. *Glob Chang*
533 *Biol* 15, 976–991. <https://doi.org/10.1111/j.1365-2486.2008.01744.x>

534 Laughlin, D.C., Fulé, P.Z., Huffman, D.W., Crouse, J., Laliberté, E., 2011. Climatic constraints on trait-
535 based forest assembly. *Journal of Ecology* 99, 1489–1499. <https://doi.org/10.1111/j.1365-2745.2011.01885.x>

537 Laughlin, D.C., Leppert, J.J., Moore, M.M., Sieg, C.H., 2010. A multi-trait test of the leaf-height-seed
538 plant strategy scheme with 133 species from a pine forest flora. *Funct Ecol* 24, 493–501.
539 <https://doi.org/10.1111/j.1365-2435.2009.01672.x>

540 Lin, Y.-S., Medlyn, B.E., Duursma, R.A., Prentice, I.C., Wang, H., Baig, S., Eamus, D., de Dios, V.R.,
541 Mitchell, P., Ellsworth, D.S., de Beeck, M.O., Wallin, G., Uddling, J., Tarvainen, L., Linderson, M.-
542 L., Cernusak, L.A., Nippert, J.B., Ocheltree, T.W., Tissue, D.T., Martin-StPaul, N.K., Rogers, A.,
543 Warren, J.M., De Angelis, P., Hikosaka, K., Han, Q., Onoda, Y., Gimeno, T.E., Barton, C.V.M.,
544 Bennie, J., Bonal, D., Bosc, A., Löw, M., Macinins-Ng, C., Rey, A., Rowland, L., Setterfield, S.A.,
545 Tausz-Posch, S., Zaragoza-Castells, J., Broadmeadow, M.S.J., Drake, J.E., Freeman, M.,
546 Ghannoum, O., Hutley, L.B., Kelly, J.W., Kikuzawa, K., Kolari, P., Koyama, K., Limousin, J.-M.,
547 Meir, P., Lola da Costa, A.C., Mikkelsen, T.N., Salinas, N., Sun, W., Wingate, L., 2015. Optimal

548 stomatal behaviour around the world. *Nat Clim Chang* 5, 459–464.
549 <https://doi.org/10.1038/nclimate2550>

550 Meir, P., Kruijt, B., Broadmeadow, M., Barbosa, E., Kull, O., Carswell, F., Nobre, A., Jarvis, P.G., 2002.
551 Acclimation of photosynthetic capacity to irradiance in tree canopies in relation to leaf nitrogen
552 concentration and leaf mass per unit area. *Plant Cell Environ* 25, 343–357.
553 <https://doi.org/10.1046/j.0016-8025.2001.00811.x>

554 Mesonet, 2024. Monthly Rainfall Table: Norman (NRMN) Monthly Actual Oklahoma Mesonet Rainfall
555 Table (in inches) [WWW Document]. URL [https://www.mesonet.org/weather/rainfall/monthly-](https://www.mesonet.org/weather/rainfall/monthly-rainfall-table?ref=1210)
556 [rainfall-table?ref=1210](https://www.mesonet.org/weather/rainfall/monthly-rainfall-table?ref=1210) (accessed 1.4.24).

557 Munroe, S.E.M., McInerney, F.A., Andrae, J., Welti, N., Guerin, G.R., Leitch, E., Hall, T., Szarvas, S.,
558 Atkins, R., Caddy-Retalic, S., Sparrow, B., 2021. The photosynthetic pathways of plant species
559 surveyed in Australia’s national terrestrial monitoring network. *Sci Data* 8, 97.
560 <https://doi.org/10.1038/s41597-021-00877-z>

561 Pedregosa, F., Varoquaux, G., Gramfort, A., Michel, V., Thirion, B., Grisel, O., Blondel, M., Müller, A.,
562 Nothman, J., Louppe, G., Prettenhofer, P., Weiss, R., Dubourg, V., Vanderplas, J., Passos, A.,
563 Cournapeau, D., Brucher, M., Perrot, M., Duchesnay, É., 2012. Scikit-learn: Machine Learning in
564 Python. *Journal of Machine Learning Research* 12, 2825–2830.

565 Poschlod, P., Kleyer, M., Jackel, A.-K., Dannemann, A., Tackenberg, O., 2003. BIOPOP — A database of
566 plant traits and internet application for nature conservation. *Folia Geobot* 38, 263–271.
567 <https://doi.org/10.1007/BF02803198>

568 Quested, H.M., Cornelissen, J.H.C., Press, M.C., Callaghan, T. V., Aerts, R., Trosien, F., Riemann, P.,
569 Gwynn-Jones, D., Kondratchuk, A., Jonasson, S.E., 2003. Decomposition of sub-arctic plants with
570 differing nitrogen economies: A functional role for hemiparasites. *Ecology* 84, 3209–3221.
571 <https://doi.org/10.1890/02-0426>

572 Reich, P.B., Oleksyn, J., Wright, I.J., 2009. Leaf phosphorus influences the photosynthesis–nitrogen
573 relation: a cross-biome analysis of 314 species. *Oecologia* 160, 207–212.
574 <https://doi.org/10.1007/s00442-009-1291-3>

575 Reich, P.B., Tjoelker, M.G., Pregitzer, K.S., Wright, I.J., Oleksyn, J., Machado, J., 2008. Scaling of
576 respiration to nitrogen in leaves, stems and roots of higher land plants. *Ecol Lett* 11, 793–801.
577 <https://doi.org/10.1111/j.1461-0248.2008.01185.x>

578 Richter, R., Schläpfer, D., 2002. Geo-atmospheric processing of airborne imaging spectrometry data.
579 Part 2: Atmospheric/topographic correction. *Int J Remote Sens* 23, 2631–2649.
580 <https://doi.org/10.1080/01431160110115834>

581 Schaepman, M.E., Jehle, M., Hueni, A., D’Odorico, P., Damm, A., Weyermann, J., Schneider, F.D.,
582 Laurent, V., Popp, C., Seidel, F.C., Lenhard, K., Gege, P., Küchler, C., Brazile, J., Kohler, P., De Vos,
583 L., Meuleman, K., Meynart, R., Schläpfer, D., Kneubühler, M., Itten, K.I., 2015. Advanced
584 radiometry measurements and Earth science applications with the Airborne Prism Experiment
585 (APEX). *Remote Sens Environ* 158, 207–219. <https://doi.org/10.1016/j.rse.2014.11.014>

586 Schiefer, F., Schmidtlein, S., Kattenborn, T., 2021. The retrieval of plant functional traits from canopy
587 spectra through RTM-inversions and statistical models are both critically affected by plant
588 phenology. *Ecol Indic* 121, 107062. <https://doi.org/10.1016/j.ecolind.2020.107062>

- 589 Schläpfer, D., Richter, R., 2002. Geo-atmospheric processing of airborne imaging spectrometry data.
 590 Part 1: Parametric orthorectification. *Int J Remote Sens* 23, 2609–2630.
 591 <https://doi.org/10.1080/01431160110115825>
- 592 Schweiger, A.K., Schütz, M., Anderwald, P., Schaepman, M.E., Kneubühler, M., Haller, R., Risch, A.C.,
 593 2015. Foraging ecology of three sympatric ungulate species – Behavioural and resource maps
 594 indicate differences between chamois, ibex and red deer. *Mov Ecol* 3, 6.
 595 <https://doi.org/10.1186/s40462-015-0033-x>
- 596 Smith, N.G., Dukes, J.S., 2017. LCE: leaf carbon exchange data set for tropical, temperate, and boreal
 597 species of North and Central America. *Ecology* 98, 2978–2978.
 598 <https://doi.org/10.1002/ecy.1992>
- 599 Stevens, A., Ramirez-Lopez, L., 2022. An introduction to the prospectr package [WWW Document].
 600 URL <https://CRAN.R-project.org/package=prospectr>
- 601 Van Der Meer, F., Bakker, W., Scholte, K., Skidmore, A., De Jong, S., Clevers, J., Addink, E., Epema, G.,
 602 2001. Spatial scale variations in vegetation indices and above-ground biomass estimates:
 603 Implications for MERIS. *Int J Remote Sens* 22, 3381–3396.
 604 <https://doi.org/10.1080/01431160152609227>
- 605 Verrelst, J., Dethier, S., Rivera, J.P., Munoz-Mari, J., Camps-Valls, G., Moreno, J., 2016. Active Learning
 606 Methods for Efficient Hybrid Biophysical Variable Retrieval. *IEEE Geoscience and Remote*
 607 *Sensing Letters* 13, 1012–1016. <https://doi.org/10.1109/LGRS.2016.2560799>
- 608 Wang, H., Harrison, S.P., Prentice, I.C., Yang, Y., Bai, F., Furstenau Togashi, H., Wang, M., Zhou, S., Ni,
 609 J., 2017. The China Plant Trait Database.
- 610 Wright, I.J., Reich, P.B., Westoby, M., Ackerly, D.D., Baruch, Z., Bongers, F., Cavender-Bares, J.,
 611 Chapin, T., Cornelissen, J.H.C., Diemer, M., Flexas, J., Garnier, E., Groom, P.K., Gulias, J.,
 612 Hikosaka, K., Lamont, B.B., Lee, T., Lee, W., Lusk, C., Midgley, J.J., Navas, M.-L., Niinemets, Ü.,
 613 Oleksyn, J., Osada, N., Poorter, H., Poot, P., Prior, L., Pyankov, V.I., Roumet, C., Thomas, S.C.,
 614 Tjoelker, M.G., Veneklaas, E.J., Villar, R., 2004. The worldwide leaf economics spectrum. *Nature*
 615 428, 821–827. <https://doi.org/10.1038/nature02403>
- 616 Yuhas, R., Goetz, A., Boardman, J., 1992. Discrimination among semi-arid landscape endmembers
 617 using the spectral angle mapper (SAM) algorithm, in: *Summaries of the Third Annual JPL*
 618 *Airborne Geoscience Workshop. Volume 1: AVIRIS Workshop. JPL*, pp. 147–149.
- 619