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1 **Marjal dels Moros: a model site for the structural, functional,**  
2 **and socioeconomic assessment of managed Mediterranean**  
3 **marshes**

4

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49 **Abstract**

50 Coastal wetlands deliver critical ecosystem services but remain highly degraded by  
51 anthropogenic and climatic pressures. This study presents an integrated structural,  
52 functional, and socio-economic assessment of Marjal dels Moros, a managed  
53 Mediterranean brackish marsh in eastern Spain, to evaluate restoration effectiveness  
54 and inform climate-based management. Six subsites representing well-preserved,  
55 altered, and restored conditions were analyzed for water and sediment properties,  
56 microbial community composition, and greenhouse gas (GHG) fluxes, alongside a multi-  
57 criteria socio-economic evaluation of restoration scenarios. Results revealed strong  
58 environmental heterogeneity mostly driven by the hydroperiod and salinity gradients,  
59 with restored sites exhibiting intermediate sediment characteristics and reduced proxies  
60 of pathogenic bacterial genera. Microbial ordination highlighted hydrological control of  
61 methane-cycling guilds, while GHG fluxes showed a clear functional gradient: as such,  
62 permanently inundated, nutrient-rich sites emitted high CO<sub>2</sub> and CH<sub>4</sub> fluxes whereas  
63 subsites with seasonal drying and higher salinity acted as near-neutral or CH<sub>4</sub>-  
64 suppressing zones. These patterns confirm that hydroperiod and salinity management  
65 shape microbial guilds and carbon dynamics, directly influencing climate regulation  
66 services. Socio-economic analysis indicated stakeholder preference for measures  
67 enhancing natural hydrology, habitat diversity, and risk reduction, supporting restoration  
68 strategies. Findings underscore hydrology-first strategies, integrated monitoring of  
69 functional indicators, and inclusive governance as key to achieving ecological integrity,  
70 climate mitigation, and socio-economic co-benefits. The proposed framework offers  
71 transferable guidance for policy-relevant restoration of Mediterranean coastal wetlands  
72 under multi-use pressures.

73 **Keywords** Coastal wetlands; Mediterranean marshes; ecological restoration;  
74 greenhouse gas (GHG) fluxes; microbial communities; ecosystem services; adaptive  
75 management; socio-economic assessment; hydrology; climate-smart strategies.

76

## 77        1. Introduction

78        Coastal wetlands are priority ecosystems for environmental management due to their  
79        capacity to deliver multiple ecosystem services (Barbier, 2019) while being highly  
80        exposed to anthropogenic and climatic pressures (Newton et al., 2020). They play a key  
81        role in biodiversity conservation, hydrological regulation, nutrient retention, and carbon  
82        storage, contributing simultaneously to climate change mitigation and adaptation  
83        (Nicholls & Lowe, 2004; Morris, et al., 2012; Wigand et al., 2017; Hagger et al., 2022).  
84        These functions are increasingly recognized in international and regional policy  
85        frameworks, including the Ramsar Convention, the EU Biodiversity Strategy for 2030,  
86        and climate-neutrality targets, which emphasize the protection and restoration of  
87        wetlands as nature-based solutions (Thorslund et al., 2017; Ferreira et al., 2023).

88        Despite their recognized value, coastal wetlands remain among the most degraded  
89        ecosystems worldwide (Li et al., 2018; Newton et al., 2020). Land-use change,  
90        hydrological modification, pollution, and habitat fragmentation have led to widespread  
91        losses in ecological integrity, status and ecosystem service provision (Newton et al.,  
92        2020; Morant et al., 2021). Climate change further exacerbates these pressures through  
93        sea-level rise, altered precipitation regimes, and increased frequency of extreme events  
94        (Day et al., 2008). In response, environmental managers are increasingly required to  
95        implement adaptive, evidence-based strategies that balance conservation objectives  
96        with human uses, while aligning with regulatory instruments such as the EU Water  
97        Framework Directive (WFD) and the Habitats Directive (HD) (Verhoeven, 2014).

98        Mediterranean brackish marshes constitute a distinctive type of coastal wetland shaped  
99        by strong seasonal variability, complex freshwater–marine interactions, and limited tidal  
100       influence (Morant et al., 2020). Their ecological structure and functioning are closely  
101       regulated by water management practices, salinity gradients, and sediment processes,  
102       making them particularly sensitive to both climatic variability and human intervention  
103       (Ibñez et al., 2000). These characteristics place Mediterranean marshes at the  
104       intersection of multiple policy domains, including water management (WFD) (Pérez-  
105       Ruzafa et al., 2011), nature conservation (Natura 2000 network, HD) (De Wit & Boutin,  
106       2023), flood risk management (Estrela-Segrelles et al., 2021), and climate adaptation  
107       strategies (Losada et al., 2019). These particular coastal marshes, besides, are explicitly  
108       referenced in European policy agendas, which promote the integration of their proper  
109       structural and functioning into planning and decision-making (e.g. RD817/2015 in Spain).  
110       However, the multifunctionality of these systems also creates trade-offs, particularly in

111 densely populated coastal zones where urban, industrial, and agricultural demands  
112 compete with conservation objectives (Novoa et al., 2020).

113 Historically, Mediterranean coastal marshes like Marjal dels Moros (València, Spain)  
114 have been extensively modified through drainage, reclamation, agricultural  
115 intensification, and urban and industrial expansion (Perennou et al., 2020). Hydrological  
116 alterations—often designed to control flooding or maximize land productivity—have  
117 disrupted natural water regimes, altered salinity patterns, and reduced habitat  
118 heterogeneity (Rochera et al., 2025). These changes have led to declines in biodiversity,  
119 modifications of biogeochemical cycles, and, in some cases, increased greenhouse gas  
120 (GHG) emissions, undermining both conservation and climate-related objectives (Morant  
121 et al., 2020).

122 In recent decades, restoration and management initiatives have increasingly focused on  
123 re-establishing hydrological functionality, improving water quality, and enhancing habitat  
124 diversity, in line with the WFD's ecological status objectives and the HD conservation  
125 targets (Verhoeven, 2014; Filipe et al., 2019; Stefanidis et al., 2021). More recently,  
126 climate policies have highlighted the role of wetlands in carbon sequestration and  
127 emissions reduction, reinforcing the need to consider ecosystem functioning alongside  
128 traditional structural indicators (Seddon et al., 2020, 2021). Nevertheless, evaluating the  
129 effectiveness of restoration measures remains challenging, particularly where outcomes  
130 are influenced by external pressures such as surrounding land use, industrial activities,  
131 or changes in regional water governance.

132 A persistent limitation in wetland management is the lack of integrated assessments that  
133 link environmental structure, ecosystem functioning, and ecosystem services in a way  
134 that directly informs management planning and policy implementation. Many studies  
135 focus on individual components, such as water quality, vegetation, or biodiversity, without  
136 explicitly connecting them to functional processes or management objectives defined in  
137 official plans.

138 For managed wetlands embedded in complex socio-environmental landscapes, there is  
139 a clear need for applied frameworks that integrate structural indicators (water and  
140 sediment properties), functional processes (GHG fluxes and biological activity), and  
141 socioeconomic dimensions (ecosystem services, public use, governance). Such  
142 approaches are essential to evaluate whether restoration actions contribute effectively  
143 to policy objectives, including ecological status improvement, climate resilience, and  
144 sustainable use.

145 This study presents an integrated structural, functional, and socioeconomic assessment  
146 of the Marjal dels Moros coastal marsh (eastern Spain) as a model site for the  
147 environmental management of Mediterranean brackish wetlands. The site is designated  
148 under conservation frameworks and has been the focus of multiple restoration initiatives,  
149 while simultaneously being exposed to significant external pressures, including urban  
150 and industrial development. This combination makes it particularly suitable for evaluating  
151 management effectiveness in a policy-relevant context.

152 Building on the need for integrated evaluations, this study aims to: (i) characterize  
153 environmental heterogeneity across six representative subsites of Marjal dels Moros,  
154 focusing on water and sediment properties that underpin ecological integrity and  
155 microbial community composition; (ii) assess functional processes through sediment  
156 greenhouse gas fluxes, identifying hydrological and trophic drivers of carbon dynamics;  
157 and (iii) analyse socio-economic performance and stakeholder priorities to determine  
158 how restoration scenarios influence ecosystem service provision, governance, and  
159 financing. By linking structural, functional, and socio-economic dimensions, we seek to  
160 provide actionable evidence for adaptive management and policy implementation under  
161 European frameworks, emphasizing hydrology-centred strategies and integrated  
162 monitoring as mechanisms to enhance resilience and climate-smart outcomes in  
163 Mediterranean coastal wetlands.

164

## 165 **2. Materials and Methods**

### 166 **2.1. Study site**

167 The Marjal dels Moros, **Fig. 1**, (39.61579°N to 39.64107°N, -0.27997°E to -0.240362°E)  
168 is a Mediterranean brackish coastal wetland (6 km<sup>2</sup>; 619.44 ha protected) located  
169 between the municipalities of Puçol and Sagunt (València, Spain). It is designated as  
170 Natura 2000 (SAC/SPA; code ES0000148) and hosts eight habitat types, including  
171 brackish marshes and humid dune slacks. It is classified in EUNIS level 2 as Coastal  
172 saltmarsh and saline reedbed (A2-5). Land cover is dominated by marshes and  
173 abandoned crops. Principal pressures include hydrological alteration and water scarcity  
174 (groundwater over-exploitation). Hydrologically, the system depends on a mosaic of  
175 inflows/outflows and engineered controls. This has been historically mitigated by  
176 wastewater and irrigation return flows that, as freshwater inputs often high in nutrients,  
177 have compromised the natural salinity regime required by the salt marshes. Other threat  
178 is coastal erosion associated with the nearby Sagunt Port, urban and industrial

179 expansion (Parc Sagunt), and recurrent fire risk. Management is led by the Generalitat  
180 Valenciana, with public use, environmental education (CEACV), and light agro-pastoral  
181 activities allowed in designated zones.

182 Socio-economically, the wetland sits within a dynamic territory marked by industrial  
183 growth (car-battery gigafactory, port expansion), seasonal tourism/recreation  
184 (birdwatching, education), and constrained water resources. Annual visitation to CEACV  
185 was ~10,000 users in 2024. Restoration governance involves regional authorities,  
186 academia, NGOs, and local councils, with financing historically combining public funds  
187 (e.g., LIFE projects) and municipal commitments, while long-term maintenance funding  
188 remains fragile.

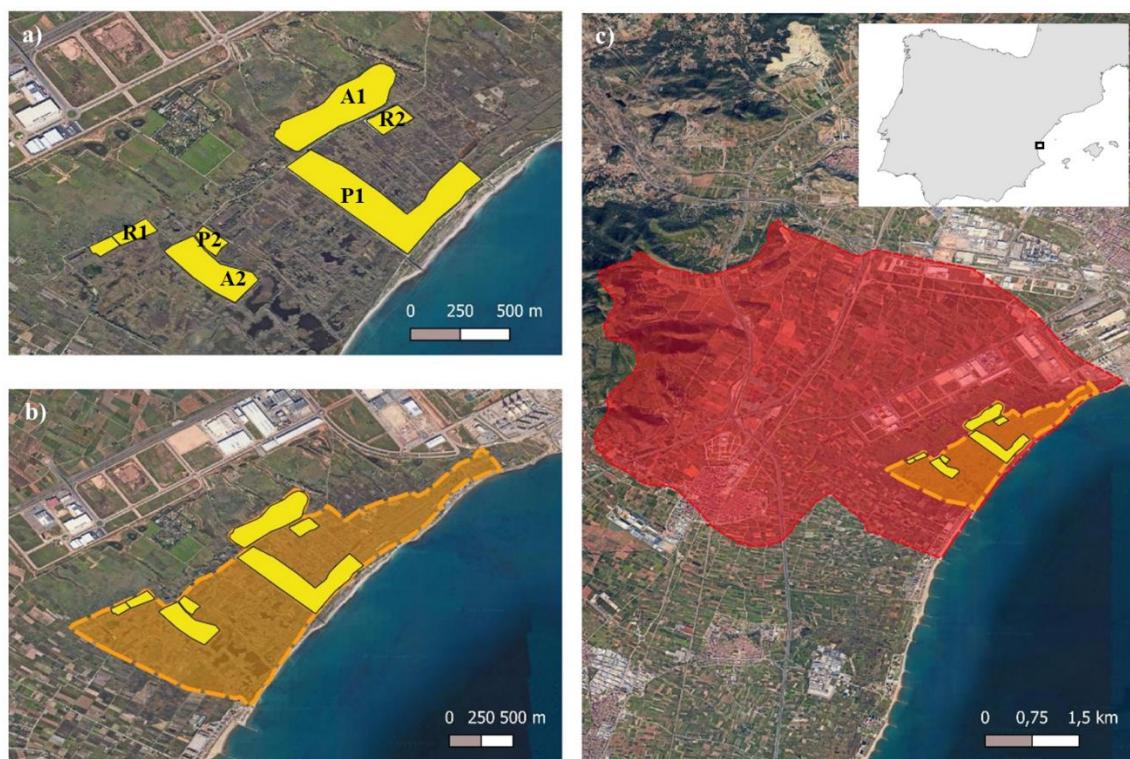
189 Six subsites were chosen to represent a gradient of hydrological and salinity conditions  
190 across three conservation categories: well-preserved, altered, and restored (two per  
191 category). Sampling was carried out in Spring 2024. Altered subsites remain artificially  
192 flooded for extended periods due to external water inputs, losing the seasonal  
193 hydroperiod and brackish character. Restored subsites exhibit partial recovery of  
194 flooding and salinity patterns approaching natural dynamics. Well-preserved subsites  
195 maintain semi-temporary hydrology with groundwater inputs and seasonal desiccation.  
196 This framework enables assessment of how salinity and hydroperiod influence  
197 ecosystem structure, function, and service provision.

198 **Altered subsites:** These marsh areas have undergone severe hydrological, trophic, and  
199 morphological disturbances. Alterations include drainage, groundwater extraction, and  
200 artificial water supply from irrigation and wastewater sources, leading to partial  
201 desalinization. Nutrient enrichment from agricultural runoff, domestic effluents, and  
202 industrial discharges has caused eutrophication and organic enrichment. Morphological  
203 changes due to land-use conversion and soil degradation have resulted in habitat loss,  
204 reduced native vegetation, proliferation of invasive species, and disruption of natural  
205 salinity gradients. Historically, agricultural exploitation and hunting further degraded  
206 vegetation structure and ecosystem services.

207 **Restored subsites:** Restoration combined active and passive measures aimed at soil,  
208 hydrology, and vegetation recovery. Actions included substrate reconstruction,  
209 topographic adjustments to restore elevation and hydrological connectivity, and active  
210 planting of native species. Hydrological improvements involved diversifying water  
211 sources (reuse of irrigation surpluses, controlled groundwater inputs) and maintaining  
212 seasonal flooding even in dry years. Vegetation recovery was achieved through planting

213 and passive recolonization following pressure reduction. Since renaturalization, four  
214 Flora Microreserves (1999–2003) and two Fauna Microreserves (2004–2006) have been  
215 established. Regular mowing of helophytic vegetation is implemented as a management  
216 measure.

217 **Well-Preserved subsites:** These sites represent intact brackish marsh habitats with  
218 stable hydrological connectivity, including natural groundwater intrusion and seasonal  
219 drying. Water quality remains high, and structural alterations are minimal. Vegetation  
220 consists of halophytic communities adapted to variable salinity regimes typical of  
221 Mediterranean coastal marshes, including reed beds, bulrush stands, and halophytic  
222 shrubs. Native biodiversity and ecological functions are largely maintained.



223

224 **Figure 1.** Marjal dels Moros subsites (R: restored, A: altered and P: well-preserved), protected  
225 site and subcatchment area.

226

## 227 **2.2. Environmental characterization**

228 Physicochemical and trophic characterization of the water and sediment was conducted  
229 adhering to the standardized methodologies of the RESTORE4Cs project (Oliveira et al.,  
230 under review). Water column analysis included *in situ* recording of depth, temperature,  
231 dissolved oxygen (DO), and conductivity. Furthermore, samples were processed to  
232 quantify alkalinity, microbial standing stocks (bacterial biomass and Chlorophyll-a), and

233 key nutrient fractions, specifically orthophosphate (PO<sub>4</sub>), ammonium (NH<sub>4</sub>), nitrate/nitrite  
234 (NO<sub>x</sub>), Total Nitrogen (TN), and Total Phosphorus (TP). Sediment assessment (Misteli et  
235 al., under review) focused on structural and organic properties, including moisture  
236 content, ash-free dry mass, Total Organic Carbon (TOC), and total nutrient  
237 concentrations (Total-N, Total-P).

### 238        **2.3.        Structural analysis (bacterial communities)**

239 In this study, a total of 28 sediment samples were analysed to characterize archaeal and  
240 bacterial communities. Genomic DNA was extracted from approximately 300–500 mg of  
241 sediment per sample using the EZNA Soil DNA isolation kit (Omega Bio-Tek, Inc.),  
242 following the protocol described by Picazo et al. (2019). The V4 hypervariable region of  
243 the 16S rRNA gene was targeted using primers 515f/806r. Amplicon libraries were  
244 prepared according to Kozich James et al. (2013), normalized using Invitrogen  
245 SequalPrep plates, and sequenced on an Illumina MiSeq platform (2x250 bp paired-end)  
246 using a v2 500-cycle reagent cartridge. Bioinformatic processing was conducted using  
247 the UPARSE pipeline within USEARCH v12b (Edgar, 2013). After merging paired-end  
248 reads and filtering for a maximum expected error rate of 0.5%, chimeric sequences were  
249 removed using UCHIME. The remaining sequences were denoised via the unoise3  
250 algorithm (Edgar, 2016) and clustered into Zero-radius Operational Taxonomic Units  
251 (ZOTUs) at 100% identity. Taxonomic assignment was performed using SINA v1.2.1152  
252 against the SILVA 138.1 database with a minimum identity threshold of 0.8 (LCA method).  
253 Non-target sequences (mitochondria, chloroplasts) and those with low alignment quality  
254 (<90%) were discarded. Finally, to normalize sequencing depth, the ZOTU table was  
255 rarefied 100 times, and the average counts were used for downstream statistical  
256 analysis.

### 257        **2.4.        Statistical analyses**

258 To evaluate spatial and temporal patterns in microbial community structure and their  
259 relationship with environmental gradients, Principal Coordinates Analysis (PCoA) was  
260 performed using the PRIMER 7 software package. Prior to analysis, ZOTU abundance  
261 tables were square-root transformed and standardized to totals to downweight highly  
262 abundant taxa (Legendre & Gallagher, 2001), and resemblance matrices were calculated  
263 based on Bray-Curtis dissimilarity. A separate PCoA was conducted for environmental  
264 variables using a square-root transformed and normalized data matrix.

265        **2.5.        Functional analysis (GHG emissions)**

266        At each subsite, six intact sediment cores were collected seasonally manually using  
267        transparent methacrylate tubes (50–100 cm length, 4 cm diameter). Each core included  
268        approximately 15 cm of sediment and an overlying water column corresponding to in situ  
269        depth. Immediately after extraction, cores were sealed, leaving a headspace of at least  
270        10 cm to enable gas measurements.

271        Cores were transported intact to the laboratory and air-purged to remove accumulated  
272        gases. During purging, cores were gently agitated without disturbing sediment structure  
273        to facilitate bubble release. An initial headspace sample was collected using a syringe  
274        with a three-way valve through a rubber septum, transferring 20 mL of air into pre-  
275        evacuated 12 mL gas-tight glass exetainers. Atmospheric pressure and temperature  
276        were recorded. After sampling, cores were resealed and incubated for ~24 h in  
277        bioclimatic chambers under temperature and light conditions approximating field  
278        conditions.

279        At the end of incubation, headspace concentrations of CO<sub>2</sub> and CH<sub>4</sub> were measured.  
280        Potential fluxes were calculated as the difference between final and initial concentrations  
281        over the incubation period. To account for ebullition, cores were vigorously shaken post-  
282        incubation to release trapped gases, and an additional headspace sample was taken.

283        Gas fluxes were computed from concentration changes using the ideal gas law and  
284        normalized to surface area:

$$286 \quad F_{\text{gas}} = \frac{dp_{\text{gas}}}{dt} \cdot \frac{V}{RTS}$$

287        where  $F_{\text{gas}}$  is expressed in mmol m<sup>-2</sup> day<sup>-1</sup>,  $\frac{dp_{\text{gas}}}{dt}$  is the rate of change in gas partial  
288        pressure (μatm s<sup>-1</sup>),  $V$  is chamber volume (m<sup>3</sup>),  $R$  is the ideal gas constant (L atm K<sup>-1</sup>  
289        mol<sup>-1</sup>),  $T$  is air temperature (K), and  $S$  is chamber surface area (m<sup>2</sup>).

290        **2.6.        Socioeconomic analysis**

291        A multi-criteria analysis (MCA) was applied to evaluate socio-economic trade-offs of  
292        restoration alternatives, integrating indicators tied to management objectives and  
293        stakeholder preferences. The approach comprised: (i) indicator definition and data  
294        collation; (ii) stakeholder weighting; and (iii) scenario scoring with multiple normalization

295 methods to test robustness. A detailed description of the method used, and the definitions  
296 of each indicator is in **Supplementary Material**, and in Anglada et al. (2025).

297 Indicators were selected through co-design (University of Valencia, Vertigo Lab,  
298 MedWet, managing authority) and derived from monitoring programs and official  
299 datasets. We retained indicators with direct management relevance and reliable data  
300 availability:

- 301 • **Socio-economic activities:** *Agriculture* (Corine-Land Cover 2018; tailored  
302 perimeter), *Industrial activities* (presence/interaction near the wetland),  
303 *Tourism/recreation* (CEACV users), *Water provisioning* (hydroperiod months).
- 304 • **Employment and costs:** *Jobs created/lost* (current staff: 6 permanent + 7  
305 temporary; planned temporary jobs per option), *Restoration costs* (maintenance  
306 mean 433,367.58€ yr<sup>-1</sup>; investments 1,047,636€ over 8 years; 2024 total  
307 ~564,308€).
- 308 • **Risk management costs:** *Flood control/drainage* (average ~1 flood event yr<sup>-1</sup>,  
309 with two events in 2024), *Coastal protection/marine submersion* (coastal barrier  
310 area, multi-temporal satellite analysis).
- 311 • **Environmental co-benefits linked to socio-economics:** *Global climate  
312 regulation* (site-type mean GWP from CO<sub>2</sub>/CH<sub>4</sub> fluxes: -2.60 t CO<sub>2</sub>eq ha<sup>-1</sup> yr<sup>-1</sup>  
313 across preserved/restored/ altered sub-sites at time of study), *Water flow  
314 improvement* (WFI), *Groundwater recharge* (E/R qualitative status >1, i.e.,  
315 extraction exceeds recharge), *Air quality* (ICQA average “Good–Fairly Good” →  
316 5.5), *Fire prevention* (regional burned area). These were included because  
317 stakeholders consistently value environmental performance as a precondition for  
318 socio-economic acceptability.

319 A stakeholder workshop, held in Port de Sagunt, on 21st January 2025, gathered 16  
320 participants (local/regional government, academia/projects, NGOs). Stakeholders rated  
321 themes (environmental, socio-economic, socio-cultural), categories, and criteria; weights  
322 were computed from these ratings. We document the imbalance (no industry present in  
323 the main workshop) as a limitation.

324 Three restoration scenarios were scored against all indicators:

- 325 1. Option 1 – Recovery of traditional irrigation network (Master Plan measures):  
326 reinstate/upgrade ditches and hydraulic connectivity; includes wastewater  
327 diversion and heritage irrigation structures.

- 328        2. Option 2 – Land-use changes and nature-based solutions (NbS):  
 329            expand/reconfigure natural areas, dune/barrier reinforcement, extensive  
 330            grazing/vegetation management to improve resilience and public access.  
 331        3. Option 3 – Vegetation/soil management for carbon markets: implement actions  
 332            aligned with offset/compensation schemes to maximize GWP reduction and  
 333            long-term financing.

334        For each indicator, expert elicitations specified expected trends per option (e.g.,  
 335            hydroperiod months, employment, costs, etc.).

336        Scenario scores were normalized using Min–Max, Max, and Vector methods to test  
 337            sensitivity of rankings to scaling assumptions.

338        The Max method normalizes each indicator by dividing its value by the maximum  
 339            observed value across scenarios:

$$344 \quad v_{ij} = \frac{a_{ij}}{\max_i a_{ij}}$$

340        where  $a_{ij}$  is the raw value of indicator  $i$  for scenario  $j$ . This approach preserves  
 341            proportionality among scenarios, meaning normalized values reflect relative differences  
 342            (e.g., a value twice as large remains twice as significant after normalization). The highest  
 343            value is scaled to 1, and others are expressed as fractions of this maximum.

345        The Min-Max method rescales indicator values between 0 and 1 based on the minimum  
 346            and maximum observed values:

$$350 \quad v_{ij} = \frac{a_{ij} - \min_i a_{ij}}{\max_i a_{ij} - \min_i a_{ij}}$$

347        Here, the minimum value is always normalized to 0 and the maximum to 1, regardless  
 348            of their relative magnitude. This method facilitates comparison across indicators but does  
 349            not capture the intensity of differences between scenarios.

351        The vector method normalizes each indicator by dividing its value by the square root of  
 352            the sum of squared values across scenarios:

$$355 \quad v_{ij} = \frac{a_{ij}}{\sqrt{\sum_i a_{ij}^2}}$$

353        This converts all attributes into dimensionless units, enabling inter-attribute comparison.  
 354        However, the resulting scale length varies, which can complicate interpretation.

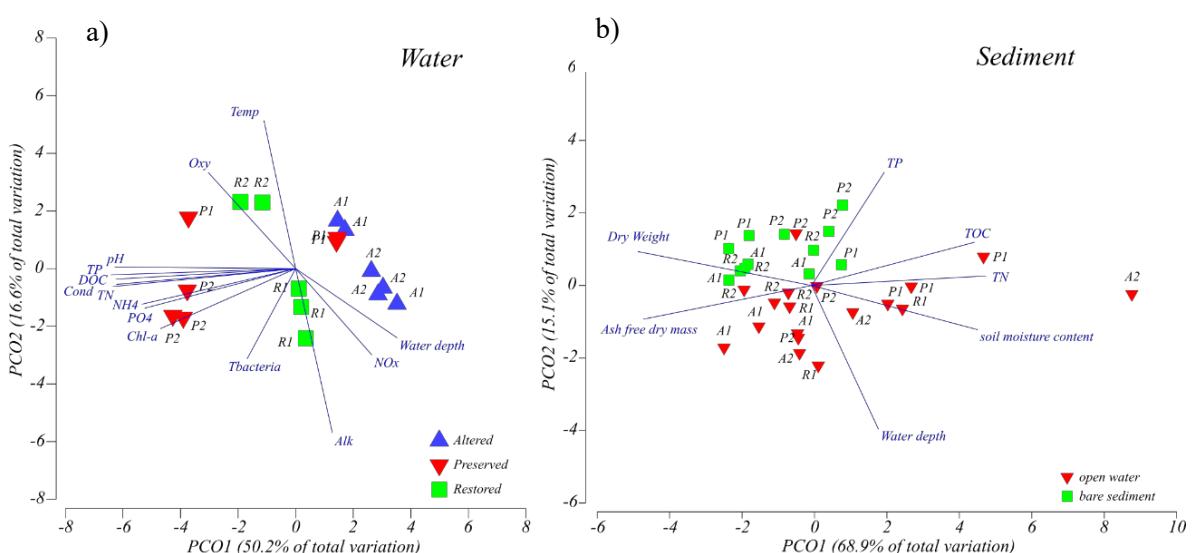
356 Weighted sums across indicators produced final MCA scores per scenario. We also ran  
 357 a simple cost-sensitivity check (vector method) to explore when higher costs would erode  
 358 Option 1's advantage.

359

360 **3. Results**

361 **3.1. Environmental Characterization of water and sediment**

362 To assess differences in environmental conditions among the studied subsites, Principal  
 363 Coordinates Analyses (PCoA) based on physicochemical variables of water and  
 364 sediment were performed for the spring samples. The PCoA analysis of water variables  
 365 revealed a clear spatial ordination of the samples (**Fig. 2a**), where the first two axes  
 366 jointly explained 66.8% of the total variation. The first axis (PCO1, 50.2%) defined the  
 367 main environmental gradient, clearly separating sites according to their conservation  
 368 status. Well-preserved sites (P1 and P2) clustered at the negative end of the axis,  
 369 strongly associated with vectors for nutrients (TP, TN, PO4, NH4), dissolved organic  
 370 carbon (DOC), conductivity (Cond), chlorophyll-a (Chl-a), and pH. In contrast, altered  
 371 (A1, A2) and Restored (R1, R2) sites were mostly positioned towards the positive end of  
 372 PCO1, correlating with greater water depth and nitrate/nitrite (NOx). The second axis  
 373 (PCO2, 16.6%) reflected internal variability within each category, separating specific  
 374 subsites. This axis notably distinguished R2 (positive values, associated with  
 375 temperature and oxygen) from R1, as well as P1 from P2, and A1 from A2, highlighting  
 376 environmental heterogeneity between replicates of the same management type.



378 **Figure 2.** Principal Coordinates Analysis (PCoA) **a.** for water, **b.** for sediment illustrating the  
 379 environmental variability across subsites in Marjal del Moros. The ordination is based on  
 380 Euclidean distances calculated from normalized physicochemical variables measured in water

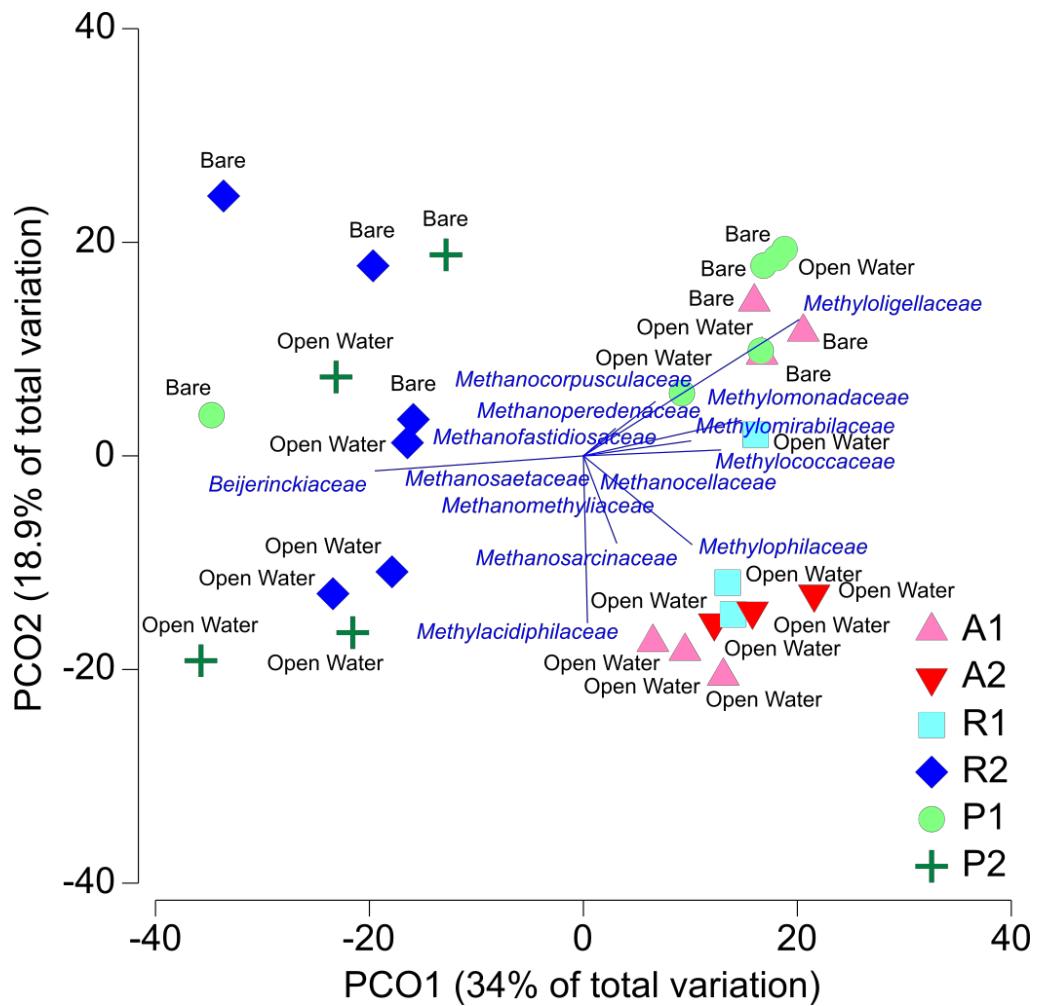
381 and sediment. Vectors (arrows) indicate the direction and magnitude of the environmental  
382 parameters driving the separation between sites along the first two axes. Environmental Variables  
383 abbreviations: Temp: Temperature; Cond: Conductivity; Oxy: Dissolved oxygen; Chl-a:  
384 Chlorophyll-a; Tbacteria: Total bacterial abundance, DOC: Dissolved organic carbon, NH4:  
385 Ammonium, NOx: nitrate + nitrite, PO4: Orthophosphate, TN: Total Nitrogen; TP: Total Phosphorus;  
386 TOC: Total organic carbon.

387

388 Principal Coordinates Analysis (PCoA) based on physico-chemical variables of sediment  
389 samples revealed a spatial ordination of the sites, in which the first axis (PCO1) alone  
390 explained most of the total variance (68.9%) (**Fig. 2b**). This axis defined a marked trophic  
391 and mineralization gradient: nutrient (TP, TN) and organic content (TOC, soil moisture  
392 content) vectors exhibited a strong positive correlation with PCO1, closely associated  
393 with subsites A2 and P1, which were positioned at the positive end of the axis. In contrast,  
394 sites A1, P1, and R2 (bare sediment samples) clustered on the opposite side (negative  
395 PCO1 values), correlated with higher dry weight and ash-free mass, indicative of more  
396 consolidated and less organic sediments. Notably, the Restored subsites (R1, R2)  
397 exhibited intermediate values distributed along the axis, indicating that these sediments  
398 are more heterogeneous. The second axis (PCO2, 15.1%) clearly separated sediments  
399 collected beneath the water column (open water samples) from bare sediment samples  
400 collected at points without standing water.

401 **3.2. Microbial characterization of Sediment**

402 Taking spring season as a model for the community composition in sediment, the  
403 Principal Coordinates Analysis of microbial community at level family reveals that  
404 methane-cycling functional guilds had a complex pattern driven by both management  
405 and hydrology (**Fig. 3**). The first axis (PCO1, 34%) clearly separated subsites P2 and  
406 R2—which were located at the negative end and associated with groups such as  
407 *Beijerinckiaceae*—from the rest of the samples. At the opposite end (positive PCO1  
408 values), the vast majority of vectors for methanotrophic bacteria (e.g.,  
409 *Methyloligellaceae*, *Methylomonadaceae*) and methanogenic archaea (e.g.,  
410 *Methanosaerincaceae*) were projected, coinciding with a clear cluster of samples  
411 collected under the water column ("Open Water") from the Altered (A1, A2) and Restored  
412 (R1) sites. This arrangement suggests that permanent inundation in these sites,  
413 regardless of their restoration status, favors the development of more diverse and  
414 abundant microbial consortia linked to methane metabolism.

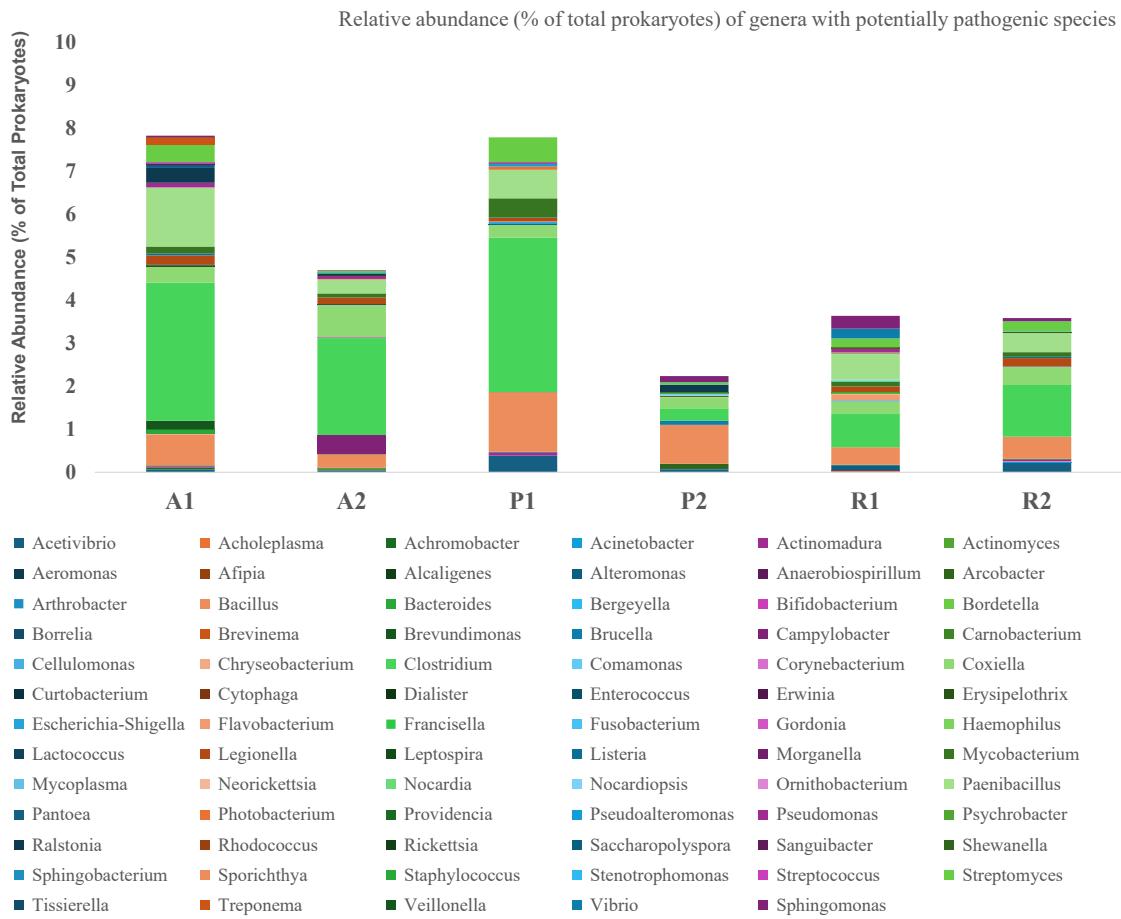


415

416 **Figure 3.** Principal Coordinates Analysis (PCoA) illustrating the microbial community variability  
 417 (Family level) across subsites y Marjal del Moros. The ordination is based on Bray-Curtis  
 418 distances calculated from standardized microbial composition (complete ZOTU table) measured  
 419 in water and sediment. Subsites: A1/A2: Altered; P1/P2: Well-Preserved; R1/R2: Restored. Only  
 420 taxa methane related are plotted.

421

422 The second axis (PCO2, 18.9%) reflected the determining effect of water cover at the  
 423 time of sampling, separating submerged samples (negative values) from bare sediment  
 424 samples ("Bare"). In this regard, samples from P1 and the emerged fraction of A1  
 425 clustered in the positive quadrant of PCO2, distant from the main vectors of  
 426 methanotrophs typical of aquatic environments. This highlights that seasonal desiccation  
 427 exerts a strong selective pressure on the functional community, differentiating it from  
 428 permanently flooded sediments.



429

430 **Figure 4.** Relative abundance (%) of prokaryotic genera containing potentially  
 431 pathogenic species in sediment samples.

432

433 The cumulative relative abundance of genera harboring potentially pathogenic species  
 434 was moderate, ranging between 2% and 8% of the total prokaryotic community (**Fig. 4**).  
 435 Most of this fraction was dominated by genera ubiquitous in soils and sediments, such  
 436 as *Clostridium*, *Bacillus*, and *Paenibacillus*. These groups, although harboring  
 437 pathogenic strains, comprise mostly free-living species typical of the natural dynamics of  
 438 organic matter decomposition in wetlands, frequently acting only as opportunistic  
 439 pathogens rather than as a direct sanitary threat.

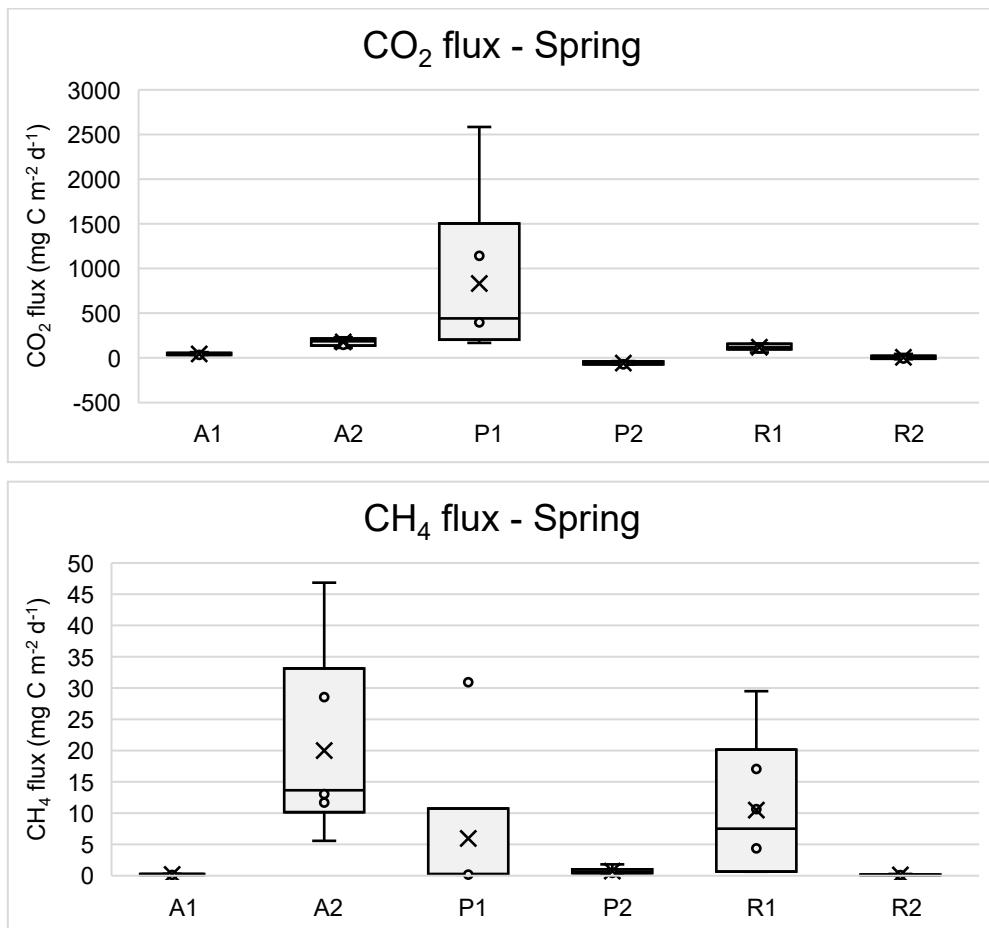
440 Nevertheless, from the perspective of management and ecosystem services, the  
 441 widespread detection of specific indicators of fecal contamination is relevant. Although  
 442 their relative abundance was low, their ubiquitous presence in all samples—regardless  
 443 of conservation status—signals a diffuse anthropogenic pressure throughout the marsh.  
 444 Specifically, marker genera such as *Escherichia-Shigella* appeared in all subsites, albeit  
 445 always maintaining marginal relative abundances below 0.05%. Similarly, *Arcobacter*, a

446 genus frequently associated with wastewater, was detected across the entire wetland,  
447 yet with values consistently lower than 0.1%.

448 When examining differences between treatments, it was observed that the relative  
449 abundance of potentially pathogenic genera was not distributed uniformly but rather  
450 appeared linked to substrate characteristics. The highest percentages of these groups  
451 were consistently recorded in subsites A1, A2, and P1 (Fig. 4), coinciding with sediments  
452 presenting a higher organic and nutrient load. In contrast, the Restored subsites (R1 and  
453 R2) exhibited a notably different pattern: not only did they present the lowest abundance  
454 percentages in the study, but they also displayed high similarity to one another. This  
455 homogeneity in the restored sites indicates that management measures have succeeded  
456 in stabilizing sediment conditions, reducing intra-class variability and mitigating the  
457 presence of risk-associated communities linked to the organic degradation observed in  
458 altered zones and specific points of the preserved zones.

459 **3.3. Functional Assessment**

460



461

462

463 **Figure 5.** Boxplots showing seasonal variability and conservation-status differences in  
464 greenhouse gas fluxes ( $\text{CO}_2$  and  $\text{CH}_4$  in  $\text{mg C m}^{-2} \text{ d}^{-1}$ ) across subsites of Marjal dels  
465 Moros during the Spring campaign (Altered: A1, A2; Restored: R1, R2; Well-preserved:  
466 P1, P2).

467

468 Greenhouse gas fluxes measured in spring revealed strong spatial variability among  
469 subsites, reflecting the combined influence of hydroperiod, sediment properties, and  
470 microbial composition (**Fig. 5**).  $\text{CO}_2$  dominated carbon efflux across all conditions, while  
471  $\text{CH}_4$  fluxes were comparatively lower but provided a sensitive indicator of reduced  
472 conditions and methanogenic activity.

473 Altered subsites exhibited contrasting behaviours. A2 showed elevated  $\text{CO}_2$  emissions  
474 ( $178.14 \pm 43.56 \text{ mg C m}^{-2} \text{ d}^{-1}$ ) and a pronounced  $\text{CH}_4$  hotspot ( $19.99 \pm 15.18 \text{ mg C m}^{-2}$   
475  $\text{d}^{-1}$ ), consistent with its position in the environmental ordination (**Fig. 2**) as nutrient-rich  
476 and waterlogged, and its microbial profile (**Fig. 3**) dominated by methanogenic archaea  
477 (*Methanosaerincaceae*) and aquatic methanotrophs. In contrast, A1 displayed much lower  
478  $\text{CO}_2$  fluxes ( $43.01 \pm 11.91 \text{ mg C m}^{-2} \text{ d}^{-1}$ ) and negligible  $\text{CH}_4$  release ( $0.19 \pm 0.07 \text{ mg C}$   
479  $\text{m}^{-2} \text{ d}^{-1}$ ), aligning with its drier sediment characteristics and reduced abundance of  
480 methane-cycling guilds.

481 Well-preserved sites showed the widest functional divergence. P1 recorded extremely  
482 high  $\text{CO}_2$  fluxes ( $831.77 \pm 927.27 \text{ mg C m}^{-2} \text{ d}^{-1}$ ), likely driven by transient aerobic  
483 mineralization during exposure, whereas P2 acted as a net  $\text{CO}_2$  sink ( $-58.82 \pm 16.42 \text{ mg}$   
484  $\text{C m}^{-2} \text{ d}^{-1}$ ) with minimal  $\text{CH}_4$  emissions ( $0.73 \pm 0.55 \text{ mg C m}^{-2} \text{ d}^{-1}$ ). This contrast mirrors  
485 their structural differences: P1 sediments were organic-rich and moist, while P2 exhibited  
486 consolidated, mineral substrates. Microbial ordination confirms this pattern, with P2  
487 clustering alongside R2 in the negative quadrant of PCO1 (**Fig. 3**), associated with  
488 *Beijerinckiaceae* and low methanogen abundance, explaining its negligible  $\text{CH}_4$  flux.

489 Restored subsites displayed intermediate and more stable fluxes. R1 emitted moderate  
490  $\text{CO}_2$  ( $119.81 \pm 36.72 \text{ mg C m}^{-2} \text{ d}^{-1}$ ) and  $\text{CH}_4$  ( $10.47 \pm 11.27 \text{ mg C m}^{-2} \text{ d}^{-1}$ ), reflecting  
491 partial recovery of hydroperiod and salinity but persistent organic enrichment. In contrast,  
492 R2 exhibited near-neutral  $\text{CO}_2$  exchange ( $6.61 \pm 21.57 \text{ mg C m}^{-2} \text{ d}^{-1}$ ) and negligible  $\text{CH}_4$   
493 release ( $0.12 \pm 0.08 \text{ mg C m}^{-2} \text{ d}^{-1}$ ), paralleling its structural position as a drier site and  
494 its microbial similarity to P2, both characterized by reduced methanogenic guilds and  
495 dominance of aerobic taxa. This convergence between R2 and P2 underscores how  
496 hydroperiod and sediment consolidation jointly constrain methane cycling, reinforcing the  
497 role of structural recovery in shaping functional outcomes.

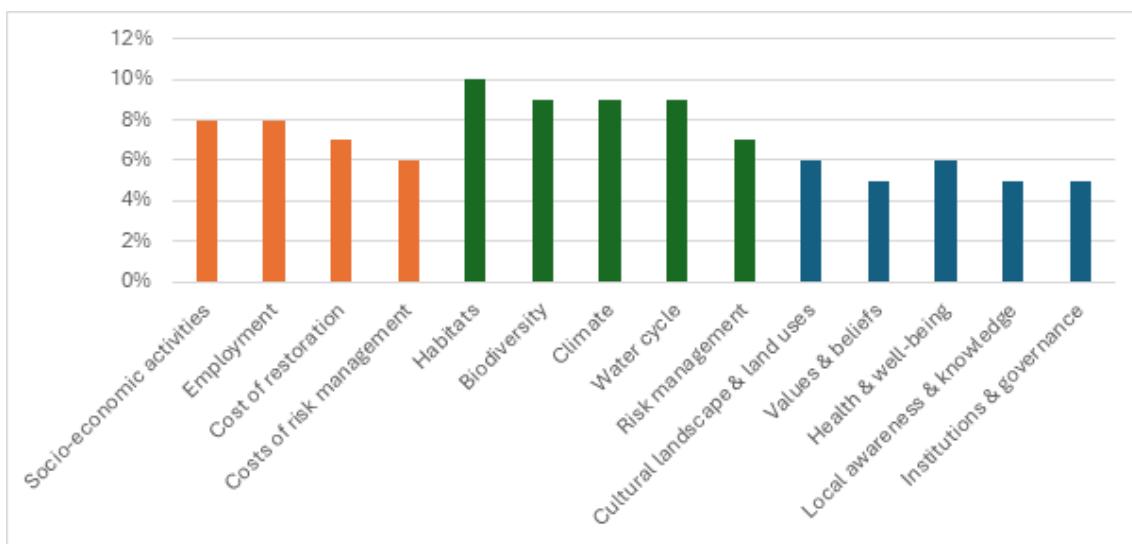
498 Overall, the integrated evidence indicates that hydrological state and substrate quality  
499 govern both microbial assemblages and GHG fluxes. Permanently inundated, nutrient-  
500 rich zones (e.g., A2) favour methanogenesis and elevated CO<sub>2</sub> emissions, while sites  
501 with seasonal exposure and higher salinity (e.g., P2, R2) tend toward lower or even  
502 negative CO<sub>2</sub> fluxes and minimal CH<sub>4</sub> production. These findings highlight hydroperiod  
503 management and salinity control as critical levers for climate-relevant functioning in  
504 Mediterranean coastal wetlands.

505

### 506 3.4. Socioeconomic Dimension

#### 507 Stakeholder priorities and indicator performance

508 **Fig. 6** compiles the results from the stakeholders' criteria for a restoration in Marjal dels  
509 Moros. Local stakeholders assigned the highest priority to the Habitats category, which  
510 received a weight of 10%. Within this category, emphasis was placed on the preservation  
511 of aquatic and terrestrial habitats that may be affected during or after restoration  
512 interventions.



513

514 **Figure 6.** Comparison of the weight of criteria when considering a restoration  
515 project, according to Marjal dels Moros' stakeholders (16 respondents)

516

517 The categories Biodiversity (species richness), Climate (global climate regulation), and  
518 Water Cycle ranked closely behind, each with a weight of 9%. Although the remaining  
519 categories scored below 9%, they are not considered negligible; rather, they are  
520 perceived as less critical compared to environmental and socio-economic dimensions.

521 Notably, socio-cultural categories received lower weights relative to environmental and  
522 economic ones.

523 Within the Socio-economic category, criteria were distributed relatively evenly, indicating  
524 balanced consideration of activities. However, Water Provisioning emerged as slightly  
525 more important (30%) compared to Industrial Activities (20%).

526 For categories comprising two criteria—such as Costs of Risk Management (flood control  
527 and drainage; erosion and coastal protection), Water Cycle (water flow improvement;  
528 groundwater recharge), Risk Management (air quality improvement; fire prevention),  
529 Cultural Landscape and Land Uses (accessibility to public blue/green areas; cultural  
530 heritage), and Local Awareness and Knowledge (scientific research; educational and  
531 recreational interest)—weights were nearly equivalent between paired criteria. This  
532 suggests that both aspects within each category should be considered equally in  
533 restoration planning. Categories represented by a single criterion should be interpreted  
534 based on their individual weight.

535 Detailed weights for all criteria, as reported by 16 stakeholders from Marjal dels Moros,  
536 are provided in **Supplementary Table 1**.

### 537 Scenario comparison

538 • **Min–Max normalization:** Option 2 (Land-use changes & NbS) ranked highest  
539 (0.68), outperforming Option 1 (0.43) and Option 3 (0.40). This result indicates  
540 broader, positive contributions across many indicators—particularly those  
541 weighted by hydrology and accessibility/resilience.

542 • **Max normalization:** Option 1 (Master Plan irrigation recovery) ranked slightly  
543 higher (0.66) than Option 2 (0.63) and Option 3 (0.59), driven by strong benefits  
544 for Global climate regulation, Cultural heritage (restoration of historic ditches),  
545 and Employment (temporary jobs).

546 • **Vector normalization:** Option 1 again led (0.64) with Option 2 and Option 3 very  
547 close (0.62 each), confirming that practical hydraulic reconnection plus heritage  
548 works deliver a favourable mix of stakeholder-weighted benefits despite higher  
549 costs.

550

551 **Table 1:** Normalized scores per criterion and per option, and global MCA score Marjal  
552 dels Moros case pilot.

CRITERIA	Results Op1 Normalised value	Results Op2 Normalised value	Results Op3 Normalised value	Results Op1 Normalised value "Min Max"	Results Op2 Normalised value "Min Max"	Results Op3 Normalised value "Min Max"	Results Op1 Normalised value	Results Op2 Normalised value	Results Op3 Normalised value	
	1,00	0,62	0,73	1,00	0,00	0,30	0,72	0,44	0,53	
Agriculture	1,00	0,62	0,73	1,00	0,00	0,30	0,64	0,58	0,50	
Water provisioning (Water availability for nature conservation)	1,00	0,92	0,79	1,00	0,60	0,00	0,00	0,00	1,00	
Interactions with industrial sector	0,00	0,00	1,00	0,00	0,00	1,00	0,00	0,00	1,00	
Tourism / Recreational activities	0,92	1,00	0,88	0,33	1,00	0,00	0,57	0,62	0,54	
Jobs created or lost during/ following restoration	1,00	0,55	0,51	1,00	0,09	0,00	0,80	0,44	0,41	
Restoration costs (investments, maintenance...)	0,00	0,14	0,14	0,00	1,00	1,00	0,36	0,45	0,45	
Flood control/drainage	0,43	0,62	0,00	0,69	1,00	0,00	0,53	0,69	0,18	
Erosion / runoff, Coastal protection / marine submersion regulation	1,00	1,00	0,95	1,00	1,00	0,00	0,59	0,59	0,56	
Land and aquatic habitats created/preserved or lost	1,00	1,00	0,95	1,00	1,00	0,00	0,59	0,59	0,56	
Species richness	0,85	1,00	0,85	0,00	1,00	0,00	0,54	0,64	0,54	
Global climate regulation	0,00	0,19	0,39	0,00	0,49	1,00	1,48	1,57	1,67	
Water flow improvement	1,00	0,92	0,83	1,00	0,50	0,00	0,63	0,58	0,52	
Groundwater recharge	0,10	0,10	0,00	1,00	1,00	0,00	0,44	0,44	0,38	
Air quality improvement	0,91	1,00	0,95	0,00	1,00	0,00	0,55	0,60	0,58	
Fire prevention	0,00	0,05	0,24	0,00	0,22	1,00	0,36	0,40	0,52	
Accessibility to public blue/green areas	0,67	1,00	0,67	0,00	1,00	0,00	0,49	0,73	0,49	
Cultural heritage	1,00	0,14	0,14	1,00	0,00	0,00	0,98	0,14	0,14	
Place attachment for spiritual, aesthetic, cultural reasons...	0,77	0,91	1,00	0,00	0,61	1,00	0,49	0,58	0,64	
Mental health influence	0,92	1,00	1,00	0,00	1,00	1,00	0,54	0,59	0,59	
Scientific research	0,75	0,86	1,00	0,00	0,43	1,00	0,49	0,57	0,66	
Education & recreational interest	1,00	1,00	0,79	1,00	1,00	0,00	0,62	0,62	0,49	
Participation in decision making and trust in institutions	0,67	0,67	1,00	0,00	0,00	1,00	0,49	0,49	0,73	
	MCA VALUE	0,66	0,63	0,59	0,43	0,68	0,40	0,64	0,62	0,62

553

554

## 555 4. Discussion

### 556 Integrating environmental structure, ecosystem function, and socio-economic 557 outcomes

558 This study provides an integrated appraisal of a managed Mediterranean brackish  
559 marsh, showing how restoration modifies environmental structure (water and sediment  
560 properties; **Fig. 2**), reshapes benthic microbial assemblages (**Fig. 3**, **Fig. 4**), and alters  
561 greenhouse gas (GHG) emissions ((**Fig. 5**), with measurable implications for  
562 socio-economic priorities and decision-making (**Table 1**). Across the gradient of  
563 well-preserved, altered, and restored subsites, hydrology and salinity emerged as the  
564 principal organizers of both biogeochemical dynamics and stakeholder-valued services.  
565 This finding supports restoration strategies in coastal wetlands that elevate hydroperiod  
566 design, water-quality controls, and habitat heterogeneity as foundational levers for  
567 resilience and multifunctionality (Yang et al., 2017; Vélez-Martín et al., 2018). In the  
568 Marjal dels Moros, where historical engineering, water scarcity, and surrounding  
569 urban/industrial land uses complicate governance, the integrated assessment clarifies  
570 how targeted actions can move the system toward ecological integrity, climate benefits,  
571 and socially acceptable uses, consistent with EU policy frameworks, WFD and HD,  
572 (Latron et al., 2022) and the nature-based solutions agenda (Thorslund et al., 2017).

573 From a management perspective, the results support the view that restoring and  
574 maintaining environmental heterogeneity is central to enhancing wetland resilience and  
575 multifunctionality. Differences among sites in water characteristics, sediment properties,

576 and biological communities appear to structure key functional processes, including GHG  
577 dynamics, and influence the capacity of the marsh to support ecosystem services  
578 prioritized in the Management Plan.

### 579 **Environmental heterogeneity as a resilience asset**

580 The ordination of physicochemical variables (**Fig. 2**) revealed a strong  
581 trophic/mineralization axis (PCO1; 68.9% variance) along which altered (A2) and one  
582 preserved subsite (P1) clustered with high TOC, TN, TP, and moisture, while other  
583 preserved and restored subsites aligned with drier, more mineral substrates. The second  
584 axis (PCO2; 15.1%) separated submerged open water, from bare sediments, highlighting  
585 water cover at sampling as a proximate driver of variability. Restored subsites (R1, R2)  
586 occupied intermediate, more dispersed positions, consistent with re-established  
587 heterogeneity under recovery. Ecologically, this mosaic limits single-state dominance  
588 (e.g., permanently inundated, eutrophic patches) and supports multiple functions and  
589 services (Vélez-Martín et al., 2018). In managed Mediterranean marshes, such  
590 heterogeneity buffers interannual climate variability (e.g., drought) and aligns with  
591 adaptive management prescriptions that favour diverse hydroperiods and salinity  
592 regimes (Rochera et al., 2025).

### 593 **Microbial community structure: hydrological control of methane-cycling guilds**

594 The microbial PCoA (**Fig. 3**) showed methane-cycling guilds tracking hydrological state  
595 rather than restoration label alone. Methanotrophic (*Methyloligellaceae*,  
596 *Methylomonadaceae*) and methanogenic (*Methanosaerincaceae*) families co-occurred  
597 predominantly in “open water” samples from altered sites (A1, A2) and one restored site  
598 (R1), while P2 and R2 grouped with taxa such as *Beijerinckiaceae* on the opposite side  
599 of the gradient. The second axis separated submerged from bare sediments,  
600 emphasizing seasonal desiccation as a selective filter that diminishes typical aquatic  
601 methanotroph signatures. These patterns indicate that water level and residence time  
602 modulate redox conditions and substrate supply, thereby governing methane-related  
603 metabolism, mechanistically consistent with the functional flux differences observed  
604 across conservation classes (Morant et al., 2020; Rochera et al., 2025).

605 From an ecosystem-services standpoint, **Fig. 4** shows low but ubiquitous detection of  
606 fecal-associated genera (e.g., *Escherichia-Shigella* <0.05%; *Arcobacter* <0.1%),  
607 implying diffuse anthropogenic (or waterfowl) inputs across all subsites. The cumulative  
608 relative abundance of potentially pathogenic genera was consistently higher in altered  
609 (A1, A2) and one preserved subsite (P1), while restored sites (R1, R2) had the lowest  
610 values and highest mutual similarity, evidence that restoration is stabilizing sediment

611 conditions and limiting risk-associated microbiota tied to organic enrichment. These  
612 results argue for sustained management to curb nutrient-rich return flows and for  
613 hydroperiod designs that avoid prolonged, stagnant inundation in organic-rich sectors,  
614 thus supporting both ecological function and socially valued uses (Barbier, 2013).

615 **Functional processes: sediment GHG fluxes and hydrology-first control**

616 Our spring analysis demonstrates that GHG fluxes in Marjal dels Moros are tightly  
617 coupled to hydrological state, salinity gradients, and sediment structure, which in turn  
618 shape microbial community composition. This mechanistic linkage (hydrology-salinity-  
619 microbial guilds-GHG emissions) emerges as a central axis of ecosystem functioning  
620 and management relevance. Therefore, two complementary levers emerge for climate  
621 restoration: (i) preserve or recover salinity gradients that constrain methanogenesis by  
622 microbial guilds (Hartman-Wyatt et al., 2024); and (ii) manage hydroperiod and  
623 vegetation structure to enhance autotrophic uptake while avoiding persistent  
624 waterlogging in nutrient-rich zones that foster CH<sub>4</sub> hotspots (Beaulieu et al., 2019).  
625 Evidence from the Marjal dels Moros and related work suggests that saline conditions,  
626 intermittent exposure, and vigorous helophyte stands can yield lower net emissions or  
627 even net ecosystem carbon uptake, depending on nutrient status and vegetation  
628 productivity (Morris et al., 2012; Hagger et al., 2022).

629 Sites with permanent inundation and nutrient enrichment (e.g., A2) exhibited elevated  
630 CO<sub>2</sub> emissions and pronounced CH<sub>4</sub> hotspots, consistent with microbial profiles  
631 dominated by methanogenic archaea and aquatic methanotrophs. Conversely, subsites  
632 with seasonal exposure and higher salinity (e.g., P2 and R2) showed near-neutral or  
633 negative CO<sub>2</sub> fluxes and negligible CH<sub>4</sub> release, paralleling their consolidated sediments  
634 and microbial assemblages dominated by aerobic taxa and low methanogen abundance.  
635 These patterns confirm that restoration measures that re-establish hydroperiod variability  
636 and salinity gradients can suppress methanogenesis, stabilize carbon dynamics, and  
637 reduce climate-relevant emissions (Morris et al., 2012; Hagger et al., 2022).

638 Methodologically, the sediment-focused analyses are complemented by the study of  
639 Cabrera-Brufau et al. (under review), which was conducted concurrently with the present  
640 work within the same RESTORE4Cs project framework. While our study utilizes  
641 sediment core incubations to isolate the benthic biogeochemical drivers, Cabrera-Brufau  
642 et al. employed an ecosystem-scale approach using static chambers to measure net  
643 fluxes. This methodological duality offers a robust validation of carbon dynamics in Marjal  
644 dels Moros: the core assays elucidate the potential production capacity of the microbial  
645 community, whereas the chamber measurements integrate the modulating effects of the

646 water column and vegetation structure. Consequently, the combination of both datasets  
647 bridges the gap between benthic metabolic potential and the realized atmospheric  
648 exchange.

649 This recognition of carbon cycle processes, spanning sediment fluxes, water column  
650 dynamics, and vegetative capture as illustrated by the combined findings of this study,  
651 (Morant et al., 2020), and Cabrera-Brufau et al. (under review), provides the essential  
652 baseline for Nature-based management for climate neutrality. Thus, restored subsites  
653 may mitigate climate change under Global Warming Potential (GWP units), whereas  
654 altered subsites have a warming effect. Within this framework, the microbiological  
655 characterization presented here serves as a crucial mechanistic addition, offering the  
656 process-level understanding to support interventions. This integration of functional  
657 ecology and microbial drivers may facilitate the shift from passive conservation to active  
658 mitigation, a strategy exemplified by the LIFE Wetlands4Climate project (LIFE  
659 Wetlands4Climate, 2024), which established Marjal dels Moros as a key pilot site. This  
660 project demonstrated that targeted adjustments in hydroperiod, vegetation, and soil  
661 management can effectively optimize the carbon balance and reduce the warming  
662 capacity, thereby translating scientific evidence into concrete emission abatement  
663 capacity.

664 The integration of structural and functional evidence underscores that hydrology-first  
665 management is not only a biophysical imperative but also a socio-economic lever. By  
666 controlling water regimes and salinity, managers influence microbial guild composition  
667 and GHG fluxes, which directly affect ecosystem services prioritized in policy  
668 frameworks, such as climate regulation, water quality, and biodiversity support. For  
669 example, reducing  $\text{CH}_4$  emissions through salinity maintenance and intermittent  
670 drawdown enhances the wetland's contribution to global climate mitigation targets, while  
671 stabilizing sediment conditions improves habitat quality and reduces sanitary risks linked  
672 to pathogenic proxies.

673 These findings position hydroperiod and salinity management as foundational strategies  
674 for climate-based restoration. Measures such as diversifying water sources, curbing  
675 nutrient-rich inflows, reinstating seasonal drawdown, and maintaining saline gradients  
676 will optimize microbial processes toward aerobic pathways, minimizing methane  
677 production and promoting carbon sequestration. In turn, these functional improvements  
678 strengthen the delivery of ecosystem services valued by stakeholders, risk reduction,  
679 recreational opportunities, and cultural heritage, creating a direct bridge to socio-  
680 economic performance and governance considerations addressed in the next section.

681 **Socio-economics and governance: stakeholder preferences and scenario  
682 trade-offs**

683 Stakeholders prioritized environmental outcomes (44% aggregate weight), with habitats  
684 (10%), biodiversity (9%), climate regulation (9%), and water cycle (9%) leading.  
685 Socio-economic (29%) and socio-cultural (27%) themes were valued but ranked second,  
686 reflecting local expectations that ecological integrity is a precondition for acceptable  
687 public use and long term social license. Within socio-economics, water provisioning  
688 outranked industrial activity, and risk-management criteria (flood control/coastal  
689 protection) carried meaningful weights, consistent with coastal hazard exposure and  
690 regional water scarcity (Geijzendorffer et al., 2019). Multi-criteria analysis (**Table 1**)  
691 revealed near-parity among restoration scenarios, with the top option depending on the  
692 normalization method: Option 2 (land-use changes & NbS) ranked first under Min–Max  
693 scaling (0.68), whereas Option 1 (recovery of traditional irrigation network) led under Max  
694 (0.66) and Vector (0.64); Option 3 (vegetation/soil management for carbon markets) was  
695 consistently close behind (0.59–0.62).

696 Workshop representation skewed toward government, academia, and NGOs, with no  
697 direct industry participation, an imbalance that may have amplified biocentric  
698 preferences and elevated environmental weights. Subsequent co-design should correct  
699 this governance gap to ensure durable social license in an industrial hinterland, pairing  
700 ecological evidence with risk reduction (floods, coastal erosion) and transparent  
701 cost-sharing. Aligning local actions with basin-scale water governance (e.g., wastewater  
702 and agricultural return-flow management) is essential to sustain function and public  
703 acceptance (Turner et al., 2000).

704 **Management and policy implications**

705 The convergence of structural, microbial, and flux evidence argues for hydroperiod  
706 management and restoration as the foundation of adaptive plans. Priority actions include:  
707 (a) diversifying water sources while curbing nutrient-rich inputs; (b) reinstating seasonal  
708 drawdown in selected cells to disrupt methanogenesis; (c) maintaining salinity gradients  
709 that suppress CH<sub>4</sub>; and (d) optimizing hydraulic connectivity to balance residence time  
710 and oxygenation. These interventions reduce CO<sub>2</sub>/CH<sub>4</sub> emissions, mitigate microbial  
711 contamination proxies, and enhance habitat heterogeneity, arose explicitly valued in the  
712 management plan and supportive of WFD/HD objectives.

713 Restored subsites' lower emissions and reduced risk-associated microbiota suggest that  
714 vegetation management (helophyte establishment, rotational mowing) coupled with  
715 sediment stabilization can consolidate functional gains. Managers should target mosaics

716 that couple high autotrophic production with intermittent sediment air-exposure, avoiding  
717 extensive, nutrient-rich permanent waters that harbor methanogenic guilds (Morant et  
718 al., 2020; Hartman-Wyatt H. et al., 2024).

719 Routine programs should complement traditional structural/biological metrics with  
720 functional ones, such as seasonal CO<sub>2</sub> and CH<sub>4</sub> fluxes determinations, sediment TOC,  
721 and microbial guild markers, to quantify restoration success beyond habitat structure.  
722 Embedding such indicators into policy reporting (e.g., site-level climate regulation  
723 metrics used in the MCA) strengthens funding cases and aligns with climate-resilience  
724 objectives (Thorslund et al., 2017; Seddon et al., 2020).

725 MCA results justify combining hydraulic heritage recovery (Option 1) with NbS land-use  
726 reconfiguration (Option 2), while selectively piloting carbon-market actions (Option 3)  
727 wherever monitoring capacity and verification are robust. Managers should treat the cost  
728 sensitivity as a planning signal and pair it with stakeholder willingness-to-pay an  
729 acceptable cost sharing instruments to avoid under-resourced maintenance (Anglada et  
730 al., 2025).

### 731 **Limitations of the biophysical approach and future directions**

732 Our sediment-focused flux approach isolates benthic drivers but does not close the  
733 ecosystem carbon budget; thus combining core incubations with parallel chamber  
734 measurements over vegetation/water columns would refine source/sink attribution and  
735 better represent the wetland/atmospheric exchange. Temporal coverage underscores  
736 strong seasonality; multi-year monitoring is needed to detect interannual trends and  
737 management effects. Microbial guild inference from 16S rRNA gene taxonomic patterns  
738 should be complemented with functional gene assays (e.g., *mcrA* for methanogenesis,  
739 *pmoA* for methanotrophy) to track process-level responses to specific restoration  
740 actions. Finally, diffuse anthropogenic inputs detected via microbial proxies appeal to  
741 basin-scale interventions (wastewater diversion, agricultural return-flow reductions)  
742 beyond site boundaries to secure long-term gains in ecological integrity and social  
743 acceptance (Turner et al., 2000).

744

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759 **CONFLICTS OF INTEREST**

760 None

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763 During the preparation of this work the authors used AI (COPILOT) in order to identify  
764 potential improvements in text readability and for coding syntax support during data  
765 processing. After using this tool/service, the authors reviewed and edited the content as  
766 needed and take full responsibility for the content of the published article.

767 **AUTHOR CONTRIBUTION STATEMENT**

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769 Methodology: AC, DM, AP, CR, ACS, (biophysical-biogeochemistry part), AC, CR, MJP,  
770 CA, JM, ML, EC, JR, NP, AB, LS (socio-economic part)

771 Investigation (field and lab work): AC, DM, AP, CR, ZA, KA, MCB, ACS, RC, MJC, PC,  
772 EMGR, AIL, FMJ, CM, BM, JSM, BO, MP, JP, RO, JRS, CS, MDS, DvS, JMS, CT, DV,  
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775 Socio-economic analysis: AC, CR, MJP, CA, JM, ML, EC, JR, NP, AB, LS

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778 manuscript.

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## 944      **Supplementary material**

945

### 946      **Supplementary Material 1: Workshop methodology (Source: Anglada et al. 2025)**

947      The selection of criteria and indicators resulted from a collaborative work between  
948      Vertigo Lab, MedWet and the case pilot leader at University of Valencia. Representatives  
949      of the managing authority were particularly involved in this case pilot given their interest  
950      in updating the draft Master Plan with inputs from stakeholders. Moreover,  
951      other local stakeholders were involved, as 14 interviews were led to find out more about  
952      the context and issues of the site, and to collect data that would be useful for the MCA.

953      Definition of indicators

954      **Agriculture:** agricultural surface was selected as an indicator, based on Corine Land-  
955      Cover 2018 data Class 2 “Agricultural Areas”, since the data was easily available. Thanks  
956      to WP6 work, it was possible to tailor it to the exact perimeter of the study. According to  
957      CLC 2018 data, **2,700 hectares** of land are classified as “Agricultural Areas”. According  
958      to expert knowledge, this indicator is expected to moderately increase in case of option  
959      1, slightly decrease in case of option 2 and very slightly decrease in case of option 3.

960      **Water provisioning and availability for nature conservation:** the number of  
961      months per year when the wetland remains inundated (hydroperiod) was selected as an  
962      indicator for this criterion and attained **9** in 2024. This is determined using the inundation  
963      level in the wetland, measured at strategic points using limnimeters (water-level gauges)  
964      to obtain a continuous record of water-level variations (e.g., meters above a reference  
965      point) over time. This information was found using IZONASH, a Wetland Monitoring  
966      Program. According to expert knowledge, this indicator is expected to slightly increase  
967      in case of option 1, very slightly increase in case of option 2 and very slightly decrease  
968      in case of option 3.

969      **Interactions with the industrial sector:** next to the wetland – just outside the limits of  
970      the protected areas – works are underway to build Europe’s biggest gigafactory while  
971      extending the size of an existing industrial park. The construction of these facilities has  
972      a severe impact on the catchments’ hydrological and ecological connectivity and is  
973      expected to reduce water availability for the wetland. The lack of involvement of the  
974      private sector – particularly the industrial sector – is therefore considered one of the main  
975      barriers to preserve and restore the wetlands. With the goal of involving the private sector  
976      in restoration, the University of Valencia, through the LIFE Wetlands4Climate project has  
977      been working to create a carbon offset or carbon compensation scheme. This option is

978 captured by the scenario of option 3, as opposed to option 1 and 2, which reflect on the  
979 current scenario where **no scheme** is set up.

980 **Tourism / Recreational activities:** the number of users of the protected area per year  
981 was selected as an indicator to represent this criterion, based on data published in the  
982 annual report of the Centre of Environmental Education of the Valencia Region (CEACV).  
983 On their website, it is estimated that approximately **10,000 people** visited the centre in  
984 2024. According to expert knowledge, this indicator could experience a moderate  
985 increase in the case of option 2 (linked to the development of wilder natural  
986 areas), a rather slight increase with option 1, because of the recovery of traditional  
987 irrigation ditches, while barely increasing in the case of option 3.

988 **Jobs created or lost during/following restoration:** At Marjal dels Moros, there are 6  
989 permanent staff members involved in the restoration of the site, as well as 7 temporary  
990 workers, making up to a total of **13 persons**. According to the data shared by the  
991 managing authority, the plan is to create additional 15 to 20 temporary jobs in case of  
992 option 1, a little less in case of option 2 and even less in case of option 3, although the  
993 indicator is still increasing in all options.

994 **Restoration costs (investments, maintenance...):** According to the draft Master Plan  
995 of Marjal dels Moros, €433,367.58 per year are needed on average for the maintenance  
996 of the area, whereas investments over the last eight years accounted for €1,047,636 in  
997 total, so €130,940.50 per year on average. Thus, for 2024, the overall costs  
998 for Marjal dels Moros have been estimated at **€564,308.08**. Currently, the investments  
999 planned for 2025-2026 are estimated at €978,125.76 meaning that the average per year  
1000 of the overall costs in case of option 1, which follows the Master plan, are expected to  
1001 reach €922,430.46. High maintenance and investment costs may be perceived as an  
1002 obstacle to some SH when implementing a restoration project. This may depend on the  
1003 site capacity to finance or on the source of financing. Even if costs can be seen as a  
1004 positive contribution, enabling projects with benefits for SH and the ecosystems,  
1005 financing capacities may be limited. High costs might thus be perceived as too important  
1006 in comparison to what could be socially accepted. This indicator's value is then  
1007 considered as negative for the purpose of this MCA. According to the local team, these  
1008 costs are expected to still fairly increase in case of option 2 and option 3, but less than  
1009 for option 1.

1010 **Flood control / drainage:** Since 2021, one flood event has happened per year between  
1011 Sagunto and Puçol, according to the local news. Two happened in 2024, one in July and  
1012 one in October. The estimated average for the last years is **one flood event** per year.

1013 Floods evidently cause harm to populations; therefore, an increase of this indicator's  
1014 value is considered as negative for this case pilot. According to expert knowledge, option  
1015 1 and 2 will play a moderate role in reducing flood occurrence; although more important  
1016 in the case of option 2, while barely increasing in the case of option 3.

1017 **Erosion / runoff and Coastal protection / marine submersion regulation:** Changes  
1018 in the shoreline considering the surface area of the coastal barrier is the indicator  
1019 selected to represent this criterion. According to Multitemporal satellite imagery analysis  
1020 (e.g., Sentinel-2, Landsat), they represent **28 hectares** in 2024. According to expert  
1021 knowledge, this indicator is expected to very slightly increase in case of option 1 and  
1022 option 2 and to remain stable in case of option 3. This is because only options 1 and 2  
1023 are expected to have a direct effect on the wetland's hydrological functioning, which can  
1024 promote increase and/or stabilization of the coastal barrier. As pointed out by  
1025 stakeholders throughout the different participatory moments (workshops, interviews),  
1026 effective measures to reduce this risk require stronger investments by the coastal  
1027 management authority.

1028 **Land and aquatic habitats created/preserved or lost:** the surface of natural habitats  
1029 was chosen as an indicator to represent this criterion and is estimated to cover **600**  
1030 **hectares** of the area. The data was collected thanks to WP6 work, based on the CLC  
1031 2018 classification "Natural drylands" and "Natural wetlands". According to expert  
1032 knowledge, this indicator is expected to very slightly increase in case of option 1 and  
1033 option 2 and to increase a little less in case of option 3. In addition, option 1 – in particular  
1034 the diversion of wastewater – is expected to improve the status of brackish habitats,  
1035 potentially helping abate carbon and GHG emissions.

1036 **Species richness:** the number of species of communal interest under annex II of Habitat  
1037 and Bird Directive was chosen as an indicator as these species are the most important  
1038 ones and their diversity is a good indicator of a healthy ecosystem. As shown by the  
1039 Natura 2000 viewer, **9** of these species are identified in Marjal dels Moros. According to  
1040 expert knowledge, this indicator is expected to very slightly increase in case of option 1  
1041 and option 3 and to moderately increase in case of option 2.

1042 **Global climate regulation:** to represent this criterion, GWP derived from CO<sub>2</sub> and  
1043 CH<sub>4</sub> fluxes was chosen as an indicator since it was measured by WP4 and is very  
1044 specific to the case pilot. As GWP values were recorded only for the subsites – and not  
1045 for the whole Marjal dels Moros - the value taken as a reference for this analysis is the  
1046 mean GWP of two well-preserved sub-sites, two restored sub-sites and two altered sub-  
1047 sites. The value may differ slightly to significantly depending on the case pilot, and

1048 whether the sub-site was altered, restored, or well-preserved. For the selected sites, it is  
1049 currently estimated at **-2,60 tons of CO<sub>2</sub>eq/yr/ha** from WP4 data available at the time of  
1050 the study. The goal of this project being to help climate mitigation by restoring wetlands,  
1051 the release of GHG and therefore the increase of this indicator's value, is considered as  
1052 negative for this case pilot. According to expert knowledge, this indicator is expected  
1053 to very slightly decrease in case of option 1, slightly decrease in case of option 2 and  
1054 fairly decrease in case of option 3.

1055 **Water flow improvement:** to represent this criterion, the local team has developed a  
1056 custom indicator: the Water Flow Index (WFI). It reflects the number and equilibrium of  
1057 active inflows and outflows. It prioritizes quantity but is adjusted for balance between the  
1058 components. The formulation penalizes scenarios in which inflows and outflows are  
1059 highly unbalanced, even if the total number is high. Accordingly, calculation is described  
1060 as follow: **WFI = (nb Inflows + nb Outflows) × [1 - |nb Inflows - nb Outflows| /**  
1061 **(nb Inflows + nb Outflows)]** and is currently evaluated at **6**. To interpret this index  
1062 value, it should be considered that it is not based on a predefined scale (like 0 to 10), so  
1063 it can take any positive value depending on the characteristics of each site. Therefore, it  
1064 is feasible to assess trends over time or to compare similar systems, rather than for  
1065 absolute benchmarking. An increase in WFI is positive, as it indicates either an increase  
1066 in total water connections, which generally improves hydrological function and ecological  
1067 resilience, or a more balanced flow configuration, which is desirable for the site's stability.  
1068 In Marjal dels Moros, a WFI of 6 already reflects a relatively good configuration, but  
1069 further increases would improve hydrological setup on the site. Again, we do not expect  
1070 that option 3 has a direct effect in this hydrological setup. According to the local team,  
1071 this indicator is expected to slightly increase in case of option 1, very slightly increase in  
1072 case of option 2 and remain stable in case of option 3.

1073 **Groundwater recharge:** the aquifer overexploitation has been selected as an indicator  
1074 to represent this criterion, as a qualitative assessment based on a quotient between  
1075 extraction (E) and recharge (R), defined as: E/R < 1 (i.e., E < R); E/R = 1 (i.e., E = R);  
1076 E/R > 1 (i.e., E > R). Currently, and according to the data available from the Cartographic  
1077 Institute of Valencia, this value stands at E/R > 1, since extraction exceeds recharge. In  
1078 the context of recurrent extreme weather events, including droughts, it is seen as  
1079 sensible to not overexploit aquifers; therefore, an increase of this indicator's value is  
1080 considered as negative for this case pilot. According to expert knowledge, this indicator  
1081 is expected to very slightly decrease in case of option 1 and option 2 and remain stable  
1082 in case of option 3.

1083 **Air quality improvement:** the ICQA (Spanish Air Quality Index) is based on a maximal  
1084 subindex of PM10, PM2.5, NO<sub>2</sub>, O<sub>3</sub>, SO<sub>2</sub> and CO – scaled by health thresholds. Spain  
1085 uses its own national Air Quality Index (ICQA), which is based on EU legislation and  
1086 differs from the US AQI in scale and classification. The average value for 2024 was  
1087 estimates to be **Good-Fairly Good** which translates into a **5.5** for the purposes of the  
1088 MCA<sup>29</sup>. According to expert knowledge, this indicator is expected to remain stable in case  
1089 of option 1, very slightly increase in case of option 2 and increase a little less in case of  
1090 option 3.

1091 **Fire prevention:** the historical fire occurrence measured as burned area was selected  
1092 as an indicator for this criterion. According to the Cartographic Institute of  
1093 Valencia, **21.89km<sup>2</sup>** were burned on average per year from 2013 to 2022. Fire  
1094 incidents evidently cause harm to populations and nature; therefore, an increase of this  
1095 indicator's value is considered as negative for this case pilot. According to expert  
1096 knowledge, this indicator is expected to very slightly decrease in case of option 1,  
1097 decrease a little more in case of option 2 and slightly decrease in case of option 3.  
1098 Moreover, data collected on fire events suggests a concentration of provoked fires in  
1099 former agricultural land, and from which land farmers were expropriated.

1100 **Accessibility to public blue green areas:** there are three access points to Marjal dels  
1101 Moros, only **two** being currently in good condition and allowing accessibility. According  
1102 to the draft Master plan, only in the case of option 2 would the third access point be  
1103 renovated. In the case of the other two options, they would remain as they are.

1104 **Cultural heritage:** So far, only **one historical construction** has been restored in the  
1105 last decades, i.e. the building hosting the Centre for Environmental Education. As per the  
1106 plans shared by the management authority, option 1 would enable restoration of 7  
1107 traditional irrigation ditches along 7,800 meters, leading to a significant increase in the  
1108 number of heritage-related facilities that are restored. Options 2 and 3 aren't expected  
1109 to have any influence on this criterion.

1110 **Place attachment for spiritual, aesthetic, cultural reasons:** The indicator chosen to  
1111 measure this criterion was the number of civil society organisations that deploy their  
1112 activities in the area of the wetland – currently only **two**. The number of associations  
1113 would be expected to increase very slightly in case of option 1 (potentially linked to an  
1114 increase in agricultural activity) and moderately for option 2 – as a result of increased  
1115 wilderness - and option 3 (related to restoration activities).

1116 **Mental health influence:** to assess this criterion, a custom survey was developed by  
1117 the University of Valencia, to find out about the perceived impact of the Marjal dels

1118 Moros, considering its broad environmental context and nearby industrial areas, on the  
1119 mental well-being of the local population. This survey was done face to face over a day,  
1120 asking the questions to randomly selected visitors of Marjal dels Moros. According to this  
1121 survey of 10 respondents, **60% of the people** who were interrogated value the role of  
1122 the wetland for mental health. According to University of Valencia, this indicator is  
1123 expected to very slightly increase in case of option 1 and slightly increase in case of  
1124 option 2 and option 3.

1125 **Scientific research:** Currently, there are **four ongoing research projects** linked  
1126 with Marjal dels Moros. These include the LIFE project Wetlands4Climate, and three  
1127 Horizon projects - BlueGreenGovernance, Soteria, and RESTORE4Cs. According to the  
1128 local team, this indicator is expected to very slightly increase in case of option 1, slightly  
1129 increase in case of option 2 and fairly increase in case of option 3.

1130 **Education & recreational interest:** Currently, there are **two ongoing educational**  
1131 **projects** linked with Marjal dels Moros: one focused on birds study led by the CEACV  
1132 and another one for school-age children at Camp de Morvedre. According to the local  
1133 team, this indicator is expected to fairly increase in case of option 1 and option 2 and  
1134 very slightly increase in case of option 3.

1135 **Participation in decision making and trust in institutions:** Before the launch of  
1136 RESTORE4Cs, the wetland did not count on any participatory committee. Thanks to the  
1137 stakeholder engagement activities organised by the project, **a first** Living Lab is being  
1138 developed with the support of other research projects and local actors. A significant  
1139 increase in the number of participatory units is only expected in the case of option 3 (as  
1140 a result of the activities to develop offset or compensation schemes), whereas option 1  
1141 and 2 would have a minor impact.

1142

### 1143 **Supplementary Table 1:**

1144 The weight that criteria should have when considering a restoration project, according to  
1145 Marjal dels Moros' stakeholders (16 respondents)

Level (Themes) 1	Weight	Level (Categories) 2	Weight	Level (Criteria) 3	Weight
Socio- economics	29%	Socio- economic activities	8%	Agriculture	27%
				Industrial activities	20%
				Tourism/recreational activities	23%
				Water provisioning	30%

		Employment	8%	Jobs created or lost during/following restoration	
		Cost of restoration	7%	Investment, maintenance costs...	
		Costs of risk management	6%	Flood control / drainage	51%
				Erosion/Runoff / Coastal protection / marine submersion regulation	49%
Environment	44%	Habitats	10%	Aquatic and terrestrial habitats created/preserved or lost	
		Biodiversity	9%	Species richness	
		Climate	9%	Global climate regulation	
		Water cycle	9%	Water flow improvement	47%
				Groundwater recharge	53%
		Risk management	7%	Air quality improvement	52%
				Fire prevention	48%
Socio-cultural	27%	Cultural landscape and land uses	6%	Accessibility to public blue green areas	49%
				Cultural heritage	51%
		Values and beliefs	5%	Place attachment for spiritual, aesthetic, cultural reasons...	
		Health and well-being	6%	Mental health influence	
		Local awareness and knowledge	5%	Scientific research	53%
		Institutions and governance	5%	Education & recreative interest	47%
				Participation in decision making and trust in institutions	