

1 A Critical Review of Polyphosphate and
2 Polyphosphate Accumulating Organisms for
3 Agricultural Water Quality Management

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16 KEYWORDS

17 phosphorus, microbiology, agriculture, polyphosphate, polyphosphate accumulating organisms,
18 wastewater treatment, water quality

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20 **This paper is a non-peer reviewed preprint. It was submitted to *Environmental Science &*
21 *Technology* on June 2, 2020. Supporting information starts on page 80.**

22 ABSTRACT

23 Despite ongoing management efforts, phosphorus (P) loading from agricultural landscapes
24 continues to impair water quality. Wastewater treatment plant research has enhanced our
25 knowledge of microbial mechanisms influencing P cycling, especially related to microbes known
26 as polyphosphate accumulating organisms (PAOs) that store P as polyphosphate (polyP) under
27 aerobic conditions and release P under anaerobic conditions. However, there is limited
28 application of PAO research to reduce agricultural P loading and improve water quality. Herein,
29 we conducted a meta-analysis to identifying articles in Web of Science on polyP and its use by
30 PAOs. We grouped our meta-analysis by discipline—wastewater treatment, soils and sediments
31 (terrestrial), freshwater, marine, and agriculture—and discussed whether PAO behavior in
32 wastewater treatment plants can be extended to natural habitats, including agricultural settings
33 like soils, undergoing alternating anaerobic/aerobic conditions. We also synthesized knowledge
34 gaps and research opportunities. Terrestrial, freshwater, and marine disciplines had fewer polyP
35 and PAO articles compared to wastewater treatment and there was limited article overlap
36 between disciplines. There is an urgent need for interdisciplinary research linking PAOs, polyP,
37 and oxygen availability with existing knowledge of P forms and cycling mechanisms in natural
38 and agricultural environments to improve agricultural P management strategies and achieve
39 water quality goals.

40

41 1. INTRODUCTION

42 Non-point phosphorus (P) sources from agricultural landscapes constitute a substantial fraction
43 of diffuse P loading to water bodies around the world due to land application of chemical
44 fertilizer and manure (Carpenter et al., 1998; Carpenter, 2005; Dubrovsky et al., 2010; Jarvie et

45 al., 2013; Mekonnen & Hoekstra, 2018). Regional water quality models of the Mississippi River
46 basin estimated that croplands, pasturelands, and rangelands delivered about 80% of P loads to
47 the Gulf of Mexico from 1992 to 2002 (Alexander et al., 2008), and a global study of grey water
48 footprints estimated that agricultural land accounted for 38% of anthropogenic P loads to
49 freshwater from 2002 to 2010 (Mekonnen & Hoekstra, 2018). A recent watershed modeling
50 study estimated that 88% of P inputs into the Great Lakes Basin came from agricultural sources
51 (Hamlin et al., 2020), which have contributed to regional eutrophication issues for more than 50
52 years (Chapra and Dolan, 2012; Dolan and Chapra, 2012). In addition to various model
53 estimates, long-term monitoring records emphasize the impact of agriculture on P pollution. The
54 majority of stream samples taken near agricultural sites exceeded the United States
55 Environmental Protection Agency’s recommendations for P (i.e., 10 to 1000 µg/L depending on
56 ecoregion) from 1992 to 2004 (Dubrovsky et al., 2010). In the midwestern United States,
57 agricultural tile drainage contributed nearly 50% of P to Lake Erie’s tributaries (Smith, King,
58 Johnson, et al., 2015).

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60 There are several known barriers limiting progress towards effective agricultural P water quality
61 management. First, the long-term accumulation of amended P in agricultural soils—termed
62 “legacy P”—consistently contributes to P loading for decades to centuries after P fertilizer
63 application stops (Bennett et al., 2001; Carpenter, 2005; Goyette et al., 2018; Haygarth et al.,
64 2014; Powers et al., 2016; Sharpley et al., 2013). As a result, more aggressive P management
65 strategies are often required in agricultural soils with legacy P (Kleinman, Sharpley, Buda, et al.,
66 2011; Sharpley et al., 2013). Second, traditional agricultural P management strategies have
67 unintended water quality consequences. For example, no tillage and tile drains were promoted in

68 the mid-1990s for their ability to reduce sediment-bound P, also referred to as particulate P (PP,
69 Table S1), transport from farmland in the midwestern United States to Lake Erie (Jarvie et al.,
70 2017). However, until recently, no tillage and tile drainage management practices ignored the
71 transport of unbound P, also referred to as dissolved P (DP, Table S1; Kleinman, Sharpley, Buda,
72 et al., 2011; Kleinman, Sharpley, McDowell, et al., 2011). Without changes in agricultural P
73 management, tile drains may continue to transport DP and cause water quality issues such as
74 harmful algal blooms (Christianson et al., 2016; Jarvie et al., 2017; Scavia et al., 2014; Smith et
75 al., 2015). Last, projected changes in environmental conditions such as increased air and water
76 temperatures, more frequent and intense storms, and prolonged periods of drought (IPCC, 2014;
77 USGCRP, 2018) will further exacerbate P loading from agricultural lands to nearby water
78 bodies, reduce the effectiveness of existing agricultural P water quality management strategies,
79 and lead to more frequent and larger harmful algal blooms (Bieroza et al., 2019; Kaushal et al.,
80 2014; Lisboa et al., 2020; Maccoux et al., 2016; Markelov et al., 2019; Paerl & Otten, 2013;
81 Smith, King, & Williams, 2015; Williams & King, 2020).

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83 Without interdisciplinary research that leverages knowledge of microbial P forms and cycling
84 mechanisms, excess P loading due to known (and unknown) barriers may continue to cause
85 freshwater eutrophication and have ecological and economic impacts. Ecologically, excess P
86 leads to freshwater eutrophication, which causes structural changes to aquatic ecosystems such
87 as decreases in water transparency, potential growth of toxin producing cyanobacteria, hypoxic
88 (i.e., low oxygen) or anoxic (i.e., no oxygen) conditions, and fish die-offs (Bennett et al., 2001;
89 Carpenter et al., 1998; Carpenter, 2005; Dodds & Smith, 2016; Schindler, 2012). Economically,
90 freshwater eutrophication produces a range of negative outcomes, including increased spending

91 on drinking water treatment and management of threatened and endangered species, recreational
92 space closures, depreciation and/or loss of waterfront real estate, and decreased fish and wildlife
93 production (Carpenter et al., 1998; Carpenter, 2005; Dodds et al., 2009; Sekaluvu et al., 2018).
94 Furthermore, federal support of agricultural water quality management programs can cost
95 taxpayers millions of dollars (Gregory et al., 2007).

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97 Despite a great deal of research on P in agricultural systems, there remains limited focus on
98 biological (here, microbial) P forms and cycling mechanisms. Research on microbially-mediated
99 P cycling in specialized wastewater treatment plants (Section 5) and in the natural environment
100 (Section 6) may help fill this research gap. Furthermore, we see microbial P forms and cycling
101 mechanisms as a rate-limiting step in our collective knowledge and P water quality management
102 success. Therefore, the objective of this meta-analysis is to inspire a step change in the level of
103 interdisciplinary research that yields more holistic views of biogeochemical P processes,
104 overcomes persistent management challenges to reducing P pollution, and greatly improves
105 water quality.

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107 2. CRITICAL REVIEW OBJECTIVES & METHODS

108 While remaining grounded in established and effective agricultural P water quality management
109 strategies, we emphasize the need to build on research from other disciplines to examine whether
110 and how microbial P forms and cycling mechanisms impact P water quality goals. Herein, we
111 focus on the role of the microbial P form known as polyphosphate (polyP, Table S1, Section 4)
112 and P cycling mechanisms by a group of microbes known as polyphosphate accumulating
113 organisms (PAOs, Section 5). PAOs are known to store polyP and there is evidence of their

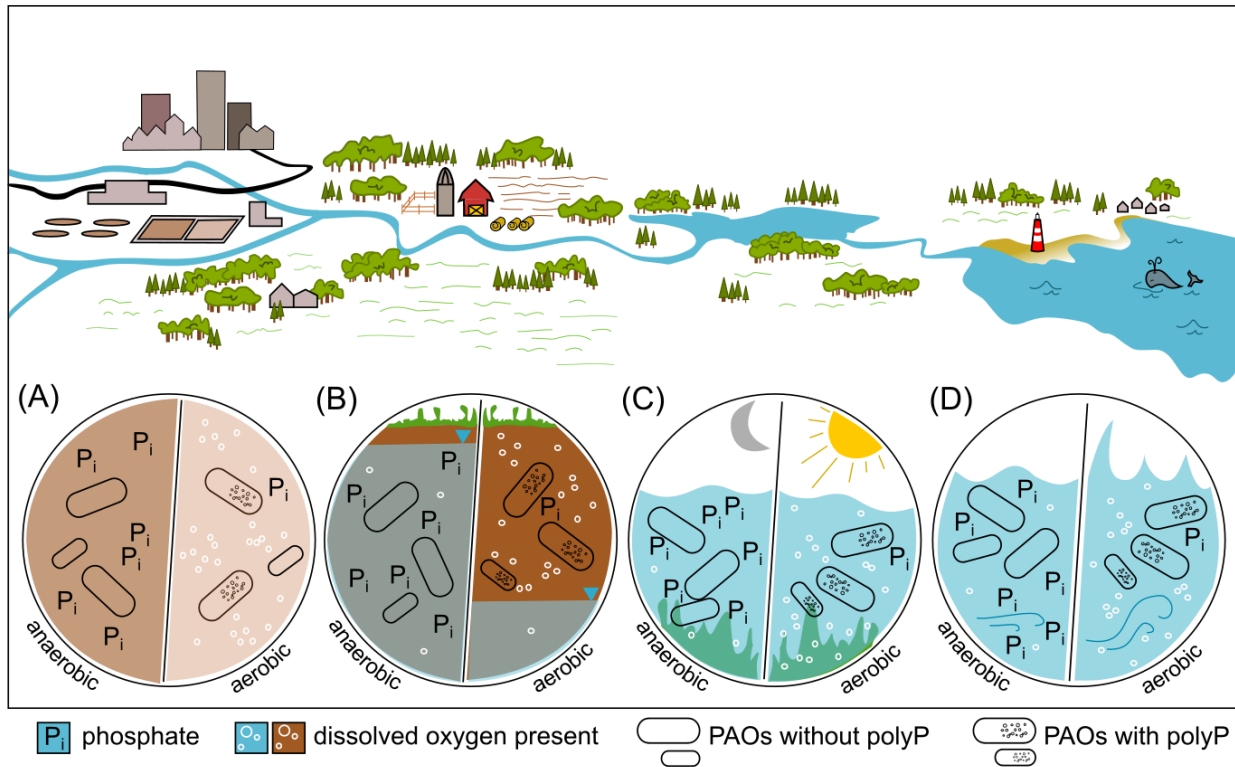
114 activity outside wastewater treatment plants (Section 6). The objectives of this critical review are
115 to: (1) synthesize established research on the role of polyP and PAOs in wastewater treatment
116 plants (Figure 1A), (2) review research on polyP and PAOs across the landscape (e.g., from soils
117 and sediments, to lakes and streams, to the ocean; Figures 1B-D) and determine whether results
118 support PAO-mediated P cycling that has been observed in wastewater treatment plants, (3)
119 discuss key knowledge gaps with respect to microbial P forms and cycling mechanisms, and (4)
120 illustrate established and emerging diagnostic tools that may encourage interdisciplinary research
121 that assesses whether knowledge of PAOs in wastewater treatment plants can be leveraged to
122 benefit agricultural P water quality management.

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Figure 1. Schematic depicting (A) known microbially-mediated phosphate (P_i) cycling in wastewater treatment plants under alternating anaerobic/aerobic conditions. Hypothesized (this study) microbially-mediated P cycling in (B) soils (including agricultural and uncultivated soils) and sediments, (C) freshwater (streams and lakes), and (D) marine water (estuaries and oceans). Blue triangles in (B) denote water table depth. Abbreviations: polyphosphate accumulating organism (PAO) and polyphosphate (polyP).

To address our objectives, we conducted a meta-analysis of articles in Web of Science (<https://webofknowledge.com>) and synthesized the results according to wastewater treatment (Figure 1A), terrestrial (i.e., soils and sediments; Figure 1B), freshwater (i.e., streams and lakes; Figure 1C), and marine (Figure 1D) disciplines. Specifically, we used the `rwos` package (<https://github.com/juba/rwos>) in R (version 3.6.2; R Core Team 2019) to query research articles

140 in the “Science Citation Index Expanded” (SCI) edition of the Web of Science Clarivate
141 Analytics web server. When web server searches resulted in less than 100,000 queries (i.e., the
142 maximum number allowed for download), we downloaded query results and analyzed them
143 using R. We summarized results of the total number of Web of Science searches as well as
144 keyword searches for five disciplinary categories from 1990 to 2019: wastewater treatment,
145 terrestrial, freshwater, marine, and agriculture (see queries in Table S2). For example, we
146 combined soil and sediment searches to form the terrestrial category. We used the total number
147 of articles in Web of Science (Figure S1) to normalize categories that were plotted versus time
148 (e.g., Figure S2). The keyword search in Web of Science queries keywords supplied by authors
149 as well as text from the article title and abstract. When possible, we included multiple versions of
150 a keyword to broaden our search (Table S2). For example, we included ‘wastewater’, ‘enhanced
151 biological phosphorus removal’, and ‘batch reactor’ when searching for articles in the
152 wastewater treatment category.

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154 In addition to the Web of Science meta-analysis, we synthesized research articles that did not
155 directly address polyP and PAOs but provided support for hypothesized microbial P mechanisms
156 outlined in Figures 1B-D (Section 6). We also relied on results from the Web of Science meta-
157 analysis and research article synthesis when summarizing knowledge gaps related to microbial P
158 forms and cycling mechanisms in agricultural water quality management research. We will not
159 synthesize research on abiotic and biotic Fe and P interactions because these subjects have been
160 thoroughly reviewed elsewhere (Chacon et al., 2006; Gerke, 2010; Lin et al., 2018; Martins et
161 al., 2011; River & Richardson, 2018; Safarzadeh-Amiri et al., 2017; Saia et al., 2017; Sulu-
162 Gambari et al., 2016; Wan et al., 2019; Wilfert et al., 2015; Wu et al., 2019) **All data and**

163 analysis scripts associated with this publication are available on GitHub at <will fill in upon
164 publication> and Zenodo (DOI: <will fill in upon publication>).

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166 3. STATE OF THE (AGRICULTURAL) FIELD

167 The majority of agricultural water quality best management practices seek to reduce DP and PP
168 loading to water bodies via human activity (e.g., reduced fertilizer applications) and abiotic
169 environmental mechanisms (i.e., chemical and physical reactions). In terms of human activity,
170 nutrient management approaches encourage farmers to use the “4Rs” when applying inorganic
171 fertilizer or manure to their fields: right source, right timing, right placement, and right rate
172 (Table S3). Chemically, soil mineral amendments increase adsorption of P to soils and soil pH
173 amendments increase soil pH to prevent P leaching (Table S3). For instance, woodchip
174 bioreactors amended with biochar can immobilize DP in surface and subsurface flows (Bock et
175 al., 2015; Sharrer et al., 2016). Many agricultural water quality best management practices are
176 designed to enhance physical retention of PP alone because over 80% of soil P is bound to
177 organic matter, clay, and minerals (Brady & Weil, 2008; Carpenter et al., 1998; Gregory et al.,
178 2007; Kleinman, Sharpley, Buda, et al., 2011; Schachtman et al., 1998; Sharpley & Menzel,
179 1987). For example, vegetated filter strips and constructed wetlands are designed to intercept
180 overland flows and allow sediments—along with PP—to settle out before they can reach
181 downstream water bodies (Table S3). Other best management practices physically retain P by
182 either minimizing soil erosion during the growing season (e.g., no tillage) and non-growing
183 season (e.g., cover crops) or increase water infiltration into the soil profile (e.g., tile drainage and
184 soil aeration). Best management practices that combine chemical and physical controls to reduce
185 DP and PP losses (e.g., mineral additions to vegetated buffers; Watts & Torbert, 2009) are

186 increasingly favored since agricultural management practices that focus on a single P form can
187 fall short of meeting P load reduction goals and have unintended water quality consequences
188 (Garcia et al., 2016; Iho et al., 2017; Jarvie et al., 2017; Smith, King, Johnson, et al., 2015;
189 Williams et al., 2016).

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191 Despite these benefits, evidence suggests that agricultural water quality best management
192 practices do not always perform as intended (Table S3). Specifically, biochar additions to
193 woodchip bioreactors have a finite adsorption capacity (Bock et al., 2015) and one study found
194 that, depending on the lumber type, woodchips leached DP during the start-up phase of
195 bioreactor development (Sharrer et al., 2016). Furthermore, PP accumulated in the topsoil of
196 vegetated filter strips and riparian buffers became an unintended P source (Lyons et al., 1998;
197 Vidon et al., 2010; Young & Briggs, 2008; Young & Ross, 2016). Use of cover crops led to the
198 accumulation of P in upper layers of the soil profile (Jarvie et al., 2017) and leached DP during
199 freeze-thaw conditions (Jarvie et al., 2015; Liu et al., 2019). No tillage practices enhanced
200 macropore development and increased subsurface transport of DP (Jarvie et al., 2017; Kleinman
201 et al., 2009). Moreover, tile drains increased subsurface DP transport (Jarvie et al., 2015; King et
202 al., 2015; Kleinman et al., 2015). These examples emphasize the need to adopt agricultural P
203 management strategies that holistically reduce P loading to nearby water bodies by
204 simultaneously considering and relying on physical, chemical, and biological mechanisms to
205 retain P.

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207 Biologically-based strategies are usually production-side centered, thereby focusing on the P
208 needs of the crop. For example, fungal and bacterial amendments increase P availability to crops

209 by mobilizing soil-bound P (Hayat et al., 2010; Javot et al., 2007; P. Jeffries et al., 2003;
210 Richardson et al., 2011; Richardson & Simpson, 2011; Rodríguez & Fraga, 1999; Schachtman et
211 al., 1998; Tapia-Torres et al., 2016). There are only a few agricultural water quality best
212 management practices that facilitate P retention via biological mechanisms (Table S3). This is
213 notable given the widely accepted importance of biological mechanisms on nitrogen (e.g.,
214 nitrification and denitrification) and carbon (e.g., respiration) cycling in the environment under
215 alternating anaerobic/aerobic conditions (Bernhardt et al., 2017; McClain et al., 2003; Vidon et
216 al., 2010). Constructed wetlands are one exception because wetland vegetation and
217 microorganisms may serve as P sinks (Kellogg & Bridgham, 2003; Noe et al., 2003; Reddy et
218 al., 1999; Richardson, 1985; Scinto & Reddy, 2003). Thus, constructed wetlands use a
219 combination of physical, chemical, and biological mechanisms to retain P and improve water
220 quality (Table S3). Besides constructed wetlands, research on soil organic P has refocused
221 discussions of P retention in agricultural soils around biological mechanisms (e.g., Bünemann et
222 al., 2011; George et al., 2018). However, the development and implementation of soil organic P-
223 focused agricultural water quality management practices remains nascent.

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225 In addition to DP and PP forms, a more holistic management focus on multiple P forms may
226 prove beneficial to reducing P loading and improving water quality. As an example, soil organic
227 P—the fraction of the soil PP pool that includes P complexed with or bound to soil organic
228 matter or P within soil macroorganisms and microorganisms (Table S1)—is relevant to this
229 review because microbes mobilize a large proportion (i.e., 20-50%) of soil organic P in P-limited
230 soils (Bünemann et al., 2011; Cross & Schlesinger, 1995) and these microbial P pools may
231 increase an additional 30-240% under implementation of agricultural management strategies

232 such as no tillage and cover crops, among others (Dodd & Sharpley, 2015). Furthermore, studies
233 synthesizing data on agricultural soil organic P demonstrate an important knowledge gap: soil
234 organic P may comprise up to 80% of soil P (Cross & Schlesinger, 1995; Haygarth et al., 2018;
235 McLaren et al., 2015; Schachtman et al., 1998), yet researchers know little about why this pool is
236 so large and how microbial fractions of soil organic P may enhance or inhibit crop nutrient
237 management and water quality management (Dodd & Sharpley, 2015; George et al., 2018;
238 Haygarth et al., 2018). Researchers hypothesize that a deeper exploration of mechanisms
239 influencing soil organic P cycling may be key to (1) solving persistent water quality issues due to
240 legacy P and (2) finding a balance between food security and clean water (Dodd & Sharpley,
241 2015; George et al., 2018; Haygarth et al., 2018).

242

243 4. CONSIDERING POLYPHOSPHATE

244 Agricultural water quality management research should consider microbial P forms such as
245 polyphosphate (polyP), in addition to integrating multiple P forms and cycling mechanisms.

246 PolyP is a chain of two or more phosphate (PO_4^{3-}) molecules bound together by a high-energy

247 phosphoanhydride bond (Table S1). PolyP is an inorganic form of P, but is commonly included

248 alongside organic P analyses because of its intracellular nature (e.g., Cade-Menun, 2017) and can

249 be chemically liberated using common organic extraction procedures (e.g., Eixler et al., 2005).

250 There are several reasons why polyP may be critical component to better understanding the role

251 of microbial processes on P cycling and transport in agricultural settings. First, polyP is

252 ubiquitous; it is found intracellularly in a wide range of organisms including bacteria, archaea,

253 fungi, plants, and animals (Brown & Kornberg, 2004; Harold, 1964; Kornberg, 1995; Rao et al.,

254 2009; Seviour & Nielsen, 2010; Zhang et al., 2002). Second, it plays a role in many basic

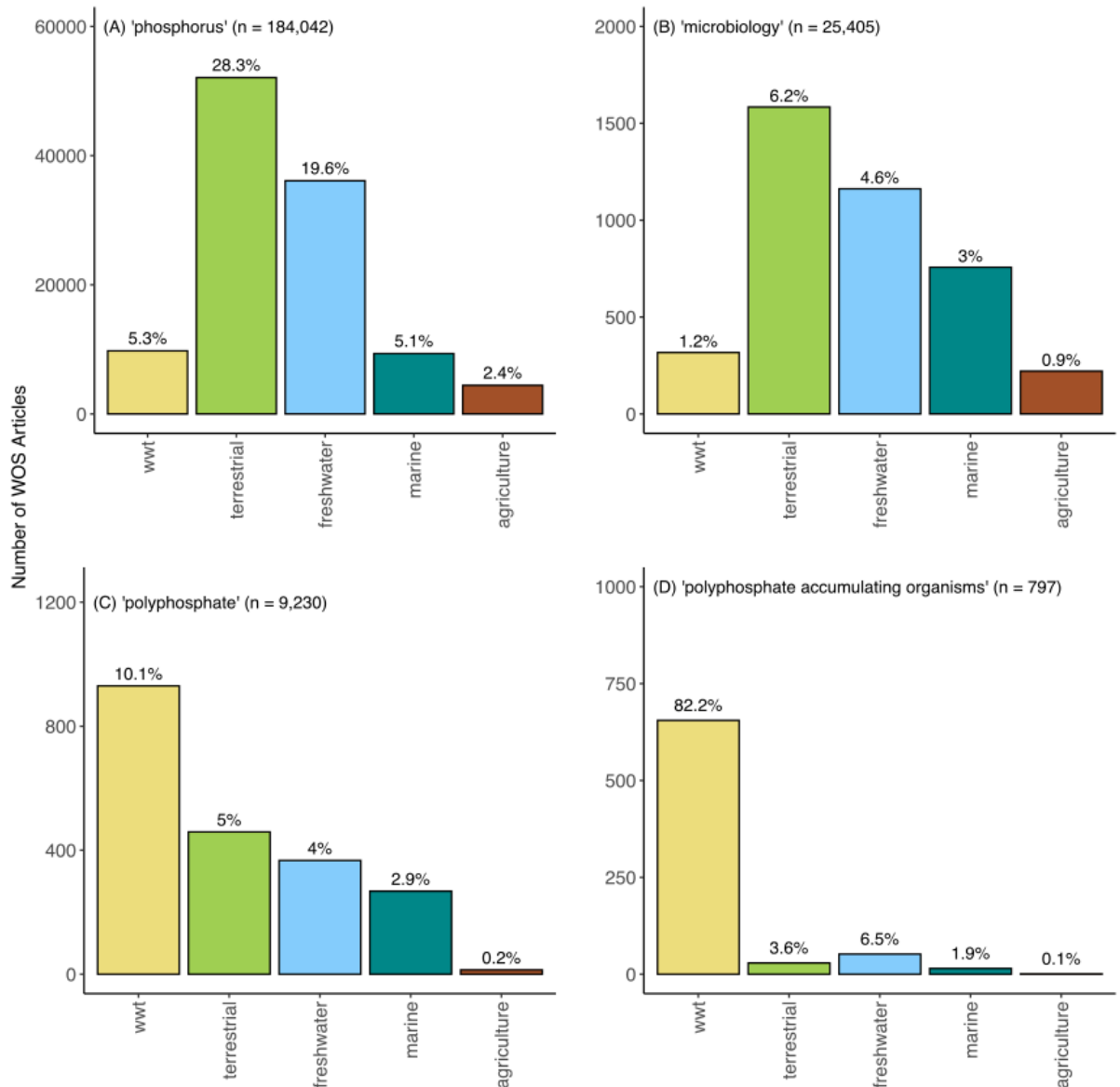
255 biological and metabolic functions such as the formation of ATP, RNA, and DNA. Specifically,
256 some microbes use polyP as an energy source, as a P reservoir, for biofilm formation, as a strong
257 ion chelator, as a buffers against alkali conditions, as a regulator of gene expression under
258 periods of stress, and as a regulator of virulence factors (Brown & Kornberg, 2004, 2008;
259 Kornberg, 1995; Rao et al., 2009; Seviour & Nielsen, 2010).

260

261 Last, while many research disciplines (individually) have expressed interest in polyP (Sections 5
262 and 6), our meta-analysis demonstrated that interdisciplinary research (i.e., research engaging
263 multiple disciplines simultaneously) is limited. We identified 9,230 articles in Web of Science
264 that included the keyword “polyphosphate” and 22.1% of these articles represented the five
265 categories included in this study (Figure 2C, Table S2). According to journal titles, the remainder
266 of polyP articles came from polymer science and biochemistry fields (e.g., from journals like
267 *Polymer Degradation and Stability* and *Journal of Biological Chemistry*). Thus, polyP-focused
268 research is being conducted independently across multiple disciplines, the majority of which are
269 not directly related to the categories we focus on in this study. Of the five categories included in
270 this study, the wastewater treatment category represented 10.1% (n = 930) of polyP articles
271 (Figure 2C, Table S2) with a consistent number of articles being published each year from 1990
272 to 2019 (Figure S2A). We expected this meta-analysis result, given that polyP storage and
273 cycling by microbes is especially well-studied in wastewater treatment plants (Section 5).

274 Terrestrial, freshwater, and marine categories combined represented 11.9% (n = 1,094) of polyP
275 articles (Figure 2C, Table S2). This research lays a foundation for microbial P cycling in the
276 natural sciences (Section 6). In the agriculture category, we found 14 polyP articles published
277 from 2001 to 2019 and found only one PAO article, which was published in 2016 (Figures 2C,

278 2D, S2B, Table S2). We discuss agricultural discipline results in further detail in Section 7.
279 Based on comparisons between searches, the majority of polyP articles (n = 838) and PAO
280 articles (n = 601) in the wastewater treatment category did not overlap with articles in another
281 (Figures 3C and 3D). In combination, these meta-analysis results highlighted an opportunity for
282 interdisciplinary research addressing P forms like polyP in the natural environment.



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284 **Figure 2.** Number and percentage of articles returned from Web of Science (WOS) by category

285 for (A) 'phosphorus', (B) 'microbiology', (C) 'polyphosphate', and (D) 'polyphosphate

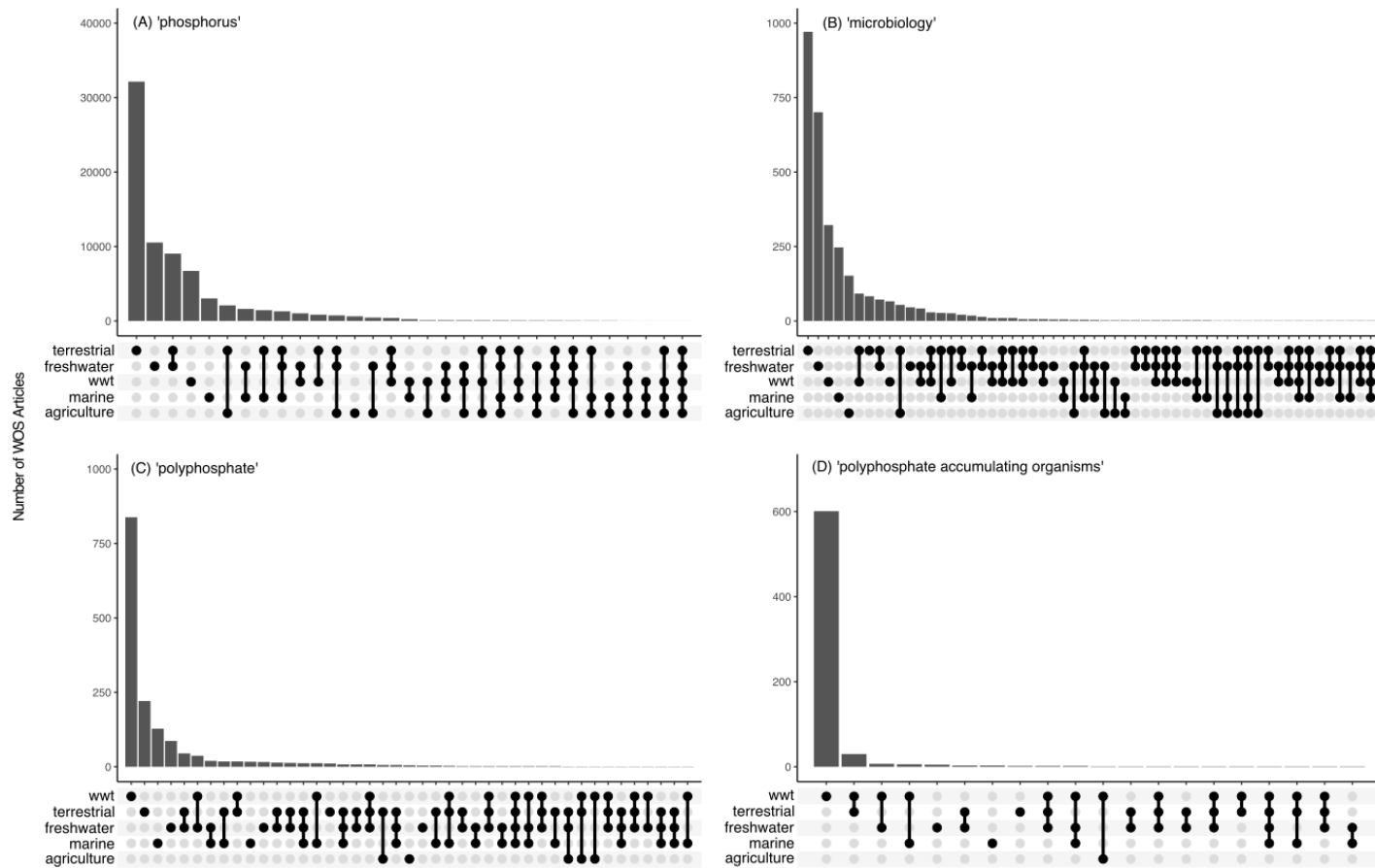
286 accumulating organisms' keyword searches. Percentages on top of each bar are relative to the

287 number of articles in the full figure subset search (i.e., n = 184,042 for 'phosphorus'); they may

288 not sum to 100% when search results are broader than the environments we focus on in this

289 study. Abbreviations: wastewater treatment (wwt). For results by specific habitat (e.g.,

290 sediments) see Figure S4.



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Figure 3. Number of overlapping Web of Science (WOS) articles by category for (A) 'phosphorus', (B) 'microbiology', (C) 'polyphosphate', and (D) 'polyphosphate accumulating organisms' keyword searches. Set diagrams under each bar plot show connections between categories; a closed circle represents articles that are not connected to other categories and circles connected with vertical lines represent articles connected to two or more categories. Abbreviations: wastewater treatment (wwt).

298 5. SPECIALIZED WASTEWATER TREATMENT & POLYPHOSPHATE
299 ACCUMULATING ORGANISMS

300 In response to growing eutrophication issues in lakes during the 1970s (e.g., Schindler, 1977),
301 wastewater treatment plant designs went beyond removing carbon and nitrogen to also removing
302 P (e.g., Barnard, 1976). As a result, a specialized wastewater treatment process known as
303 enhanced biological P removal (EBPR) was developed to simultaneously reduce operation costs
304 and remove P from wastewater (Seviour et al., 2003). EBPR is more economical than
305 conventional wastewater treatment plant P removal because it does not require expensive and
306 logistically complex additions of Ca, Al, or Fe to chemically precipitate out P (Oehmen et al.,
307 2007). Furthermore, EBPR wastewater treatment does not generate metal-laden waste solids, but
308 instead, P is transferred from solution to waste solids via intracellular storage of P as polyP by
309 microbes (Barnard, 1976; Seviour et al., 2003). EBPR wastewater treatment plant designs
310 include three main operating components: (1) an anaerobic reactor with an organic (carbon-
311 containing) energy source (e.g., acetate), (2) an aerobic zone, and (3) a means to recycle a
312 fraction of the settled biomass such that it is subjected to alternating anaerobic/aerobic conditions
313 (Seviour et al., 2003; Figures 1A and S3). While optimal operating conditions were originally
314 developed empirically rather than based on an understanding of microbial processes, it is now
315 commonly accepted that the characteristic alternating anaerobic/aerobic conditions of EBPR
316 selects for a group of microbes referred to as polyphosphate accumulating organisms (PAOs).
317 PAOs are capable of taking up phosphate in excess of normal cellular levels—coined “luxury
318 uptake” (Gebremariam et al., 2011; Seviour et al., 2003; Yall et al., 1972). Typically, EBPR
319 sludge is 5-7% P (dry weight) while the P content of conventional sludge ranges from 1-2%

320 (Yuan et al., 2012). PAOs play a large role in removing P from influent waters of EBPR
321 wastewater treatment plants around the world (e.g., Mao et al., 2015).
322
323 The most frequently studied (model) PAO known by the provisional scientific name *Candidatus*
324 *Accumulibacter phosphatis* (Hesselmann et al., 1999), has the ability to synthesize large amounts
325 of polyP under aerobic conditions to support the uptake and intracellular storage of organic
326 substrate (Seviour et al., 2003; Seviour and Nelson, 2010; Figures 1A and S3A). This
327 metabolism defines PAOs and enables them to outcompete other non-PAO EBPR heterotrophs
328 with less flexible metabolic capabilities (Gebremariam et al., 2011). For a detailed description
329 of *Candidatus Accumulibacter phosphatis* metabolism in EBPR wastewater treatment plants see
330 Text S1 and Figure S3. While we have a better understanding of PAO metabolism since EBPR
331 was introduced, the metabolic mechanisms separating PAOs from non-PAOs are still debated
332 and studied (e.g., Barnard et al., 2017; Skennerton et al., 2014). Genotypic and phenotypic
333 diversity of *Candidatus Accumulibacter phosphatis*, and PAOs in general, likely explain
334 observed variation in metabolic processes under anaerobic conditions. Rather than a single
335 metabolic model, many markedly different metabolic models may exist (Crocetti et al., 2000;
336 Kawakoshi et al., 2012; Kristiansen et al., 2013; Mao et al., 2014; Mino et al., 1998; Quang et
337 al., 2019; Rubio-Rincón et al., 2017; Seviour et al., 2003; Seviour & McIlroy, 2008; Skennerton
338 et al., 2014). As we learn more about the diversity of PAOs, we may be motivated to revisit
339 metabolic features that distinguish PAOs from non-PAOs as well as how EBPR wastewater
340 treatment plant microbial communities (including PAOs and non-PAOs) contribute to effective P
341 removal.
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343 There are several reasons why EBPR PAOs may serve as a starting point for further research on
344 microbially-mediated P management, including in the context of agricultural soils and
345 downstream water bodies. First, PAOs are well studied in the context of EBPR wastewater
346 treatment plants. Based on meta-analysis results, we identified 797 articles in Web of Science
347 that included the keyword “polyphosphate accumulating organisms” and 94.3% of these articles
348 represented the five categories included in this study (Figures 2C, 2D, Table S2). Of the five
349 categories discussed here, the largest number of PAO articles ($n = 655$, 82.2%) came from the
350 wastewater treatment category (Figure 2D, Table S2). Besides the magnitude of research, we
351 observed that the number of PAO articles in the wastewater treatment category from 1990 to
352 2019 increased over time; however, this increase was not as dramatic as the number of papers
353 per year in the wastewater treatment category identified using the keyword “phosphorus” from
354 1990-2019 (Figure S2A). Second, PAOs have been found in many environments around the
355 world (Table S4) and there remains an opportunity for collaborative research between
356 disciplines, especially, between EBPR wastewater treatment plant research and research in
357 natural environments. To illustrate this point further, of the 655 PAO articles in the wastewater
358 treatment category, 601 of these articles did not overlap with other categories (Figure 3D).
359 Terrestrial, freshwater, and marine categories combined represented 12% ($n = 96$) of PAO
360 articles (Figure 2D, Table S2) and it was rare for articles in terrestrial, freshwater, and marine
361 categories to overlap with articles in the wastewater treatment category (Figure 3D).
362 Furthermore, 69 PAO articles in terrestrial, freshwater, and marine categories overlapped with
363 articles in these same categories returned for the keyword “polyphosphate”. Third, there is
364 limited research of PAOs in agricultural systems, as evidenced by the one PAO article (i.e., Ota
365 et al., 2016) that we identified in the agriculture category (Section 7). Last, alternating

366 anaerobic/aerobic conditions exist in the natural environment and may serve as a means for
367 selecting PAOs; we expand upon this hypothesis in Section 6.

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369 6. LEVERAGING INTERDISCIPLINARY RESEARCH: PROPOSED EVIDENCE & ROLES 370 OF POLYPHOSPHATE CYCLING ACROSS THE LANDSCAPE

371 Concurrent to EBPR PAO studies in wastewater treatment plants, researchers hypothesized that
372 natural alterations between anaerobic/aerobic conditions is several types of natural environments
373 (e.g., soils, sediments, freshwater, and marine waters) may select for microorganism that had a
374 similar phenotypes and genotypes to EBPR PAOs (Davelaar, 1993; Diaz et al., 2012; Gächter et
375 al., 1988; Hupfer et al., 2007; Hupfer & Lewandowski, 2008; McMahon & Read, 2013;
376 McParland et al., 2015; Peterson et al., 2008; Pett-Ridge & Firestone, 2005; Reddy et al., 1999;
377 Schulz & Schulz, 2005; Uhlmann & Bauer, 1988). From the 1970s to present, research on P
378 cycling in the environment has been motivated by the understanding that anthropogenic P
379 loadings has led to accelerated eutrophication in many freshwater and marine ecosystems world-
380 wide (Schindler, 2012; Conley et al., 2014). With respect to biological P forms and cycling
381 mechanisms, researchers demonstrated that microorganism may respond to this increase by
382 storing excess P as polyP (Gächter et al., 1988; Uhlmann and Bauer, 1988; Kenney et al., 2001).
383 However, over 30 years later, few studies have addressed this hypothesis and its focus on
384 biological P forms and cycling mechanisms and there is also limited focus on the role of PAOs
385 from both a phenotypic and genotypic perspective. Therefore, in Sections 6.1 to 6.3, we
386 synthesize research in terrestrial (Figure 1B), freshwater (Figure 1C), and marine (Figure 1D)
387 environments that provides preliminary evidence for this hypothesis. This evidence draws from
388 our meta-analysis of Web of Science articles and includes discussion of studies identifying

389 known PAOs (i.e., *Candidatus Accumulibacter phosphatis*), identifying new PAOs, identifying
390 and quantifying polyP, P functional genes, and demonstrating relationships between P and
391 oxygen concentrations in terrestrial and aquatic environments.

392

393 *6.1 Terrestrial Habitats (Soils & Sediments)*

394 In the context of terrestrial environments, wetting and drying events influence the diffusion of
395 oxygen through soil and sediment pores. Soils and sediments tend to be anaerobic while
396 saturated, and aerobic while unsaturated (Burgin & Groffman, 2012; Silver et al., 1999; Smith &
397 Tiedje, 1979). Therefore, we hypothesize alternating wetting/drying events in soils and
398 sediments appear to be analogs to alternating anaerobic/aerobic conditions in EBPR wastewater
399 treatment plant. Specifically, we hypothesize soil and sediment PAOs take up P during drying
400 conditions and release P during wetting conditions (Figure 1B). Furthermore, we hypothesize
401 PAOs release phosphate under wetting soil and sediment conditions, which may negatively
402 impact P management goals for water quality protection. Our synthesis, which we highlight in
403 further detail below, supported these hypotheses as well as the need for interdisciplinary research
404 on PAOs and polyP in terrestrial habitats such as soils and sediments.

405

406 As stated previously, several researchers hypothesized that PAOs in soils and sediment may
407 release P during wet periods and take up P during dry periods, which mirrors the behavior of
408 PAOs like *Candidatus Accumulibacter phosphatis* in EBPR treatment plants (Davelaar, 1993;
409 Pett-Ridge and Firestone, 2005; Peterson et al., 2008; Lin et al., 2018; 2020). We identified
410 several studies documenting the presence of *Candidatus Accumulibacter phosphatis* (Archibald,
411 2010; Kunin et al., 2008; Martins et al., 2011; Peterson et al., 2008; Valdivia, 2009), its

412 phylogenically close relatives (DeAngelis et al., 2010; Pett-Ridge & Firestone, 2005), as well as
413 other novel PAOs (DebRoy et al., 2013; Li et al., 2013) in soils and sediments around the world
414 (Table S4). However, none of these studies directly addressed the role of PAOs in soil P cycling
415 under alternating anaerobic/aerobic conditions. With respect to polyP, there is evidence of polyP
416 accumulation by diverse bacterial species in river sediments contaminated by heavy metals
417 (Narancic et al., 2012) and evidence that microbes accumulated P as polyP under aerobic
418 conditions and released P under anaerobic conditions in freshwater lake sediments (Amirbahman
419 et al., 2013; Martins et al., 2011) and wetland sediments (Khoshmanesh et al., 1999, 2002).
420 However, only one of these polyP studies (i.e., Martins et al., 2011) confirmed the presence of
421 *Candidatus Accumulibacter phosphatis*. Consequently, research addressing the role of polyP and
422 PAOs under alternating wetting/drying conditions and their impact on water quality is still
423 needed.

424

425 Despite the limited research on the specific role of PAOs in soils under alternating
426 anaerobic/aerobic conditions, we found several studies exploring the impacts of environmental
427 perturbations such as wetting/drying events on soil organic P. Namely, soil organic P
428 mineralization (after rewetting) was positively correlated with microbial biomass (Dinh et al.,
429 2016) and microbial P (Grierson et al., 1998; Turner & Haygarth, 2001) upon rewetting. Soils
430 undergoing alternating wetting/drying events showed larger increases in microbial P over time
431 compared to soils that did not undergo alternating wetting/drying events (Grierson et al., 1998).
432 One study estimated that 41% of added phosphate was stored as microbial P upon soil rewetting
433 (Yevdokimov et al., 2016). One study subjected humid tropical soils to alternating anoxic (i.e.,
434 no oxygen) and aerobic conditions and found that biologically available P pools decreased

435 immediately following anoxic conditions (Lin et al., 2018). This finding supports PAO-mediated
436 P release under anaerobic conditions. Other studies have observed flushes in P after prolonged
437 drought (Kaushal et al., 2016; Lisboa et al., 2020) but further research is needed to determine
438 whether this is associated specifically with PAOs or other biological, physical, and chemical P
439 mechanisms. Increases in P release during saturated soil conditions have also been observed at
440 larger landscape (Dupas et al., 2015) and watershed scales (Franklin et al., 2013) but whether
441 and how much microbial P contributes to these patterns, relative to physical and chemical P
442 mechanisms, is still unknown (Blackwell et al., 2010).

443

444 Based on our meta-analysis, polyP research in soils and sediments was more common than PAO
445 research, but fewer than P and microbial research. For the terrestrial category, which included
446 soils and sediments (Table S2), we identified 459 articles in Web of Science that included the
447 keyword “polyphosphate” and 29 articles that included the keyword “polyphosphate
448 accumulating organisms”; this is in contrast to 52,076 and 1,584 articles in the terrestrial
449 category that included the keywords “phosphorus” and “microbiology”, respectively (Figures 2,
450 S4, Table S2). Upon closer inspection of results within the terrestrial category, most polyP and
451 PAO papers in the terrestrial category addressed soil environments (Figures S4C, S4D). For
452 example, 3.1% of all polyP papers included “soil” as a keyword while 1.8% included “sediment”
453 as a keyword (Figure S4C). In comparison to all categories studied here with the exception of
454 wastewater treatment, the terrestrial category made up the largest percentage of polyP articles
455 and the second largest percentage of PAO articles (Figures 2C, 2D, Table S2). Therefore, there
456 are a limited number of studies on polyP and PAOs in terrestrial habitats and none that directly
457 address the role of PAOs in P cycling.

458

459 In addition to the limited number of polyP and PAO studies relative to P and microbiology
460 research in terrestrial environments (Figure 2), our meta-analysis identified the limited overlap in
461 research between the terrestrial category and categories other than wastewater treatment. Nearly
462 half ($n = 221$, 48.1%) of polyP articles and ($n = 29$, 4.3%) of PAO articles in the terrestrial
463 category did not overlap with any other category (Figures 3C and 3D). As an example, several of
464 these polyP articles discussed organic P pools (e.g., Bünemann et al., 2008; Darch et al., 2016).
465 Specifically, one of these demonstrated that the addition of glucose (i.e., organic substrate) led to
466 accumulation of pyrophosphate (Bünemann et al., 2008), which is a polyphosphate polymer with
467 two phosphate molecules (Table S1). Another polyP article, demonstrated polyP accumulation
468 and release in sediment bacteria under aerobic and anaerobic laboratory conditions, respectively
469 (Khoshmanesh et al. 2002). This study did not identify the specific organisms involved but this
470 behavior supports PAO-mediated P cycling in sediments. In several rare instances, polyP articles
471 in the terrestrial category overlapped with 2 or more categories discussed here, but these summed
472 to a total of 98 articles. With respect to PAOs, we identified 7 PAO articles in the terrestrial
473 category that overlapped with articles in the wastewater treatment category. Several of these used
474 soil as a source for developing a sequencing batch reactor that was capable of removing P (e.g.,
475 Zhang et al., 2003 and Zhu et al., 2011), which supports the hypothesis that soil microbes may
476 offer some P removal benefits under the right conditions. Like polyP articles, there were several
477 rare instances when PAO articles in the terrestrial category overlapped with multiple other
478 categories discussed here; those added to a total of 21 articles. We found 1 article that
479 overlapped with all categories except agriculture (i.e., Krishnaswamy et al., 2011). There were
480 26 polyP and PAO articles that overlapped with one another and were both in the terrestrial

481 category (Figures 3C, 3D). For example, one of these isolated a new PAO from soils (Zhang et
482 al., 2003; Table S4). Given the nearly 10,000 articles combined on polyP and PAOs in Web of
483 Science, there remains an opportunity for interdisciplinary research addressing the role of PAOs
484 in water quality management between wastewater treatment, terrestrial, freshwater, and marine
485 disciplines.

486

487 *6.2 Freshwater Habitats (Streams & Lakes)*

488 In streams, rivers, and shallow regions of lakes, alternating anaerobic and aerobic conditions are
489 often driven by diel cycles in respiration and primary production, respectively. Specifically,
490 oxygen levels in the water column and upper sediment layers increase during the day due to
491 photosynthesis while, during the night, the cessation of photosynthesis combined with continued
492 respiration decreases oxygen levels (Cohen et al., 2013; Dodds, 2003). In the case of freshwater
493 tidal wetlands, hydrodynamics (e.g., tides) can also influence alternating anaerobic/aerobic
494 conditions. For example, the tide brings in oxygen rich water and recedes with water that has a
495 lower concentration of oxygen due to respiration within the wetland (Findlay & Fischer, 2013).
496 In deeper regions of lakes, alternating anaerobic/aerobic conditions are often driven by changes
497 in the depths of the aerobic epilimnion and anoxic hypolimnion due to either internal waves or
498 wind-induced surface waves (McMahon & Read, 2013). Given the existence of these alternating
499 anaerobic/aerobic conditions in freshwater habitats, we hypothesize diel changes in oxygen
500 availability driven by either metabolic or wave-driven hydrodynamics causes are analogs to
501 alternating anaerobic/aerobic conditions in EBPR wastewater treatment plants. Specifically,
502 freshwater PAOs may take up P during the day and/or during windy conditions and release it
503 during the evening and/or during calm conditions (Figure 1C). Our meta-analysis and synthesis

504 supported these hypotheses and highlight the need for interdisciplinary research on PAOs and
505 polyP.

506

507 A number of studies in freshwater habitats have documented the presence of EBPR PAOs such
508 as *Candidatus Accumulibacter phosphatis* and other microorganisms that can store polyP
509 intracellularly (Table S4). Microscopy-based studies found intracellular polyP granules in
510 freshwater microorganisms (Hupfer et al., 2004; Sicko-Goad & Lazinsky, 1986; Stevenson &
511 Stoermer, 1982) and stream biofilms (Rier et al., 2016; Saia et al., 2017; Taylor, 2016).

512 Consistent with EBPR PAO metabolism, there is evidence that oxygen concentrations influence
513 microbial P cycling in freshwater environments. Microbes accumulated P as polyP under aerobic
514 and released P under anaerobic in stream biofilms (Saia et al., 2017). Other researchers have
515 observed coupled P and oxygen patterns that are generally consistent with EBPR PAO
516 metabolism despite not directly measuring polyP concentrations or known EBPR PAO (i.e.,
517 *Candidatus Accumulibacter phosphatis*) genes. As an example, in freshwater streams, diel
518 stream water column phosphate cycling patterns were inversely related to oxygen availability;
519 when dissolved oxygen was high during the day, phosphate was low and vice versa at night
520 (Cohen et al., 2013; Ford et al., 2018). In a productive temperate lake (i.e., Muskegon Lake,
521 Michigan, USA), higher P concentrations have been linked to lower oxygen levels along a
522 seasonal time-series determined using *in situ* sensors (Weinke & Biddanda, 2018). Also, P
523 uptake near the water column-sediment boundary and water column-periphyton boundary
524 coincided with increasing oxygen concentrations while P release near these same interfaces
525 coincided with decreasing oxygen concentrations (Carlton & Wetzel, 1988; Fleischer, 1978;
526 Gächter et al., 1988; Gächter & Meyer, 1993; Read et al., 2014; Saia et al., 2017; Sherson et al.,

527 2015). While abiotic processes (e.g., redox sensitive dissolution/precipitation of Fe and P) cannot
528 be completely ruled out as a possible causal mechanisms in the cases described above
529 (Richardson, 1985), there is consensus that biological processes play a role in coupled P and
530 oxygen patterns, and in some cases, this role is significant. For example, biotic processes
531 accounted for 66% of stream diel water column P uptake and release while the remainder was
532 attributed to abiotic, Ca-P precipitation and dilution (Cohen et al., 2013). In freshwater wetlands,
533 biotic mechanisms accounted for over 83% of short-term (12-hour) water column P removal
534 (Scinto & Reddy, 2003). Given examples across several freshwater ecosystems, diel changes in P
535 uptake and release appear coupled with oxygen availability support the behavior of EBPR PAOs
536 in these habitats.

537

538 In addition to oxygen availability, there is evidence that polyP storage in freshwater
539 environments depends on P availability. For example, polyP storage by microorganisms in
540 temperate stream biofilms was greatest in nutrient-depleted headwater sites compared to
541 downstream sites that were nutrient-abundant (Taylor, 2016). Another study of temperate stream
542 biofilms demonstrated that increases in polyP storage during natural, P-abundant storm events
543 and controlled pulses of P in mesocosm experiments led to future microbial growth (Rier et al,
544 2016). An *in situ* P stream biofilm enrichment experiment carried out over four mid-Atlantic
545 ecoregions showed that P uptake rates varied predictably along a nutrient gradient (Price &
546 Carrick, 2016), such that the development of intracellular polyP granules in stream biofilms was
547 proportional to the degree of nutrient additions to stream ecosystems from the surrounding
548 landscape (Price & Carrick, 2011). Overall, proportional relationships between polyP storage and

549 P availability support/do not support the idea that polyP is a beneficial polymer for stream
550 microorganisms; it helps them conserve energy and nutrients for future use.

551

552 Similar to meta-analysis results for terrestrial habitats, polyP research in streams and lakes was
553 more common than PAO research but more limited than P and microbial research. We found 367
554 articles in Web of Science that included the keyword “polyphosphate” and 52 articles that
555 included the keyword “polyphosphate accumulating organisms” (Figures 2, S4, Table S2).

556 Compared to other keyword searches in the freshwater category, there were 36,105 articles that
557 included the keyword “phosphorus” and 1,162 articles that included the keyword “microbiology”
558 (Figures 2, S4, Table S2). When compared to terrestrial and marine categories, freshwater had
559 the most PAO articles; 6.5% of all PAO articles fell in the freshwater category while 3.6% and
560 1.9% of PAO articles came from the terrestrial and marine categories, respectively (Figure 2D,
561 Table S2). A closer look at meta-analysis results within the freshwater category revealed that
562 most polyP articles were associated with lakes while most PAO articles addressed streams
563 (Figures S4C, S4D).

564

565 Besides the limited number of polyP and PAO studies in streams and lakes, meta-analysis results
566 demonstrated limited overlap in research between the freshwater category and categories other
567 than wastewater treatment. Most frequently, polyP articles in the freshwater category did not
568 overlap with any other category ($n = 87$, Figure 3C). PAO articles most frequently overlapped
569 with the wastewater treatment category ($n = 30$, Figure 3D); however, many of these 30 articles
570 focused on EBPR PAO research. There were 45 polyP and PAO articles that overlapped with
571 one another and were both in the freshwater category. For example, one of these studied the

572 distribution of *Candidatus Accumulibacter phosphatis* P functional genes in lake water (Peterson
573 et al., 2008). Again, meta-analysis results support the need for interdisciplinary research
574 addressing the role of PAOs in water quality management between wastewater treatment,
575 terrestrial, freshwater, and marine disciplines.

576

577 *6.3 Marine Habitats (Estuaries & Oceans)*

578 Marine habitats discussed here include estuaries, coastal waters, and the open ocean. In coastal
579 estuarine ecosystems, alternating anaerobic/aerobic conditions are greatly influenced by a
580 combination of anthropogenic nutrient inputs—including nutrient-induced acceleration of
581 primary production (Diaz & Rosenberg, 2008)—and mixing of stratified layers of the water
582 column (Helm et al., 2011). In the open ocean, much like the deeper regions of lakes, alternating
583 anaerobic/aerobic conditions are driven by the mixing of stratified chemoclines typically caused
584 by waves and wind (Helm et al., 2011). Given the potential for alternating anaerobic/anaerobic
585 conditions in marine habitats, we hypothesize wave- and wind-driven oxygen gradients are
586 analogs to alternating anaerobic/aerobic conditions in EBPR wastewater treatment plants (Figure
587 1D). Our synthesis provided limited support for these hypotheses—especially compared to
588 terrestrial and freshwater habitats—but like Sections 6.1 and 6.2, our meta-analysis highlighted
589 the need for interdisciplinary research on PAOs and polyP in marine environments.

590

591 Similar to freshwater, several studies have demonstrated the storage of intracellular polyP and
592 the presence of EBPR PAOs (Table S4). Microscopy- and spectroscopy-based studies have
593 identified intracellular polyP granules in marine sediment bacteria (Diaz & Rosenberg, 2008;
594 Schulz & Schulz, 2005), and marine microorganisms (Martin et al., 2014). For example, one

595 study observed polyP accumulation by marine filamentous cyanobacterial symbionts within
596 sponges and verified the presence of P functional genes such as polyphosphate kinase (*ppk*) and
597 exopolyphosphatase (*ppx*; Text S2, Table S5) associated with polyP cycling using techniques
598 established for EBPR PAOs such as *Candidatus Accumulibacter phosphatis* (Zhang et al., 2015).
599 Another analyzed the relationships between the abundance of P functional genes in marine
600 microorganism genomes and annual water column phosphate concentrations (Temperton et al.,
601 2011). Furthermore, we identified two studies that support PAO presence in estuarine waters and
602 sediments. One noted the widespread distribution of PAOs and PAO-related P functional genes
603 (i.e., *ppk*; Text S2, Table S5) in estuarine sediments (Watson et al., 2019) and the other identified
604 bacteria in the *Rhodobacteraceae* family; *Candidatus Accumulibacter phosphatis* is also a
605 member of this family (Jeffries et al., 2016).

606

607 Only a few researchers in marine systems have linked oxygen availability in the water column
608 with P cycling, but these limited studies find support for PAO metabolism in marine
609 environments. Namely, one study found that phosphate concentrations were ~3x greater in the
610 redoxcline—a zone with a strong vertical redox gradient—of a coastal basin compared to the
611 surface (McParland et al., 2015). Another found that polyP concentrations in water samples from
612 a coastal inlet decreased as dissolved oxygen concentrations decreased (Diaz et al., 2012).

613 Elsewhere, researchers identified giant sulfur bacteria (*Thiomargarita namibiensis*) that
614 accumulated polyP under aerobic sediment conditions and released phosphate under anoxic (i.e.,
615 no oxygen) sediment conditions, a response that is functionally similar to EBPR PAOs (Schulz
616 & Schulz, 2005). Therefore, there are a number of studies identifying PAOs in marine sediments
617 but there remain opportunities to assess their role under alternating anaerobic/aerobic conditions.

618
619 In addition to oxygen availability, there is conflicting evidence in marine environments as to
620 whether P availability increases or decreases polyP storage. When taken together, polyP
621 accumulation by marine microorganisms may depend on histories of P availability. For instance,
622 phytoplankton accumulated more polyP in P-depleted regions of the Sargasso Sea compared to
623 regions that were more P-abundant (Martin et al., 2014). Similarly, in a metagenomic study of
624 marine environments, the number of P functional genes (i.e., *ppk* and *ppx*; Text S2, Table S5)
625 increased when annual P concentrations were lower (Temperton et al., 2011). In contrast to these
626 findings, one study along an urban to estuary gradient (i.e., from P-abundant inland to P-depleted
627 open ocean) demonstrated a decrease in the number of P functional genes in water column
628 microorganisms (Jeffries et al., 2016). While Jeffries et al. (2016) did not quantify intracellular
629 polyP, their findings are consistent with studies in P-abundant freshwater environments (Section
630 6.2). Additionally, a laboratory study of marine algal cultures found that intracellular
631 pyrophosphate, which is a form of polyP with only two phosphate molecules (Table S1),
632 accumulation increased under higher water column P concentrations (Cade-Menun & Paytan,
633 2010). Therefore, in P-depleted aquatic environments, polyP storage may function as an
634 adaptation to help microorganisms conserve nutrients for later use while, in P-abundant aquatic
635 environments, polyP storage may help microorganisms access alternative energy conservation
636 pathways that enable them to outcompete microorganisms with less metabolic flexibility. More
637 detailed field and laboratory research is needed to elucidate the impact of P-availability histories
638 on polyP storage by known and undiscovered PAOs.
639

640 Based on meta-analysis results, polyP and PAO articles in marine environments were less
641 common than terrestrial or freshwater categories. For marine environments, we found 367
642 articles in Web of Science that included the keyword “polyphosphate” and 15 articles that
643 included the keyword “polyphosphate accumulating organisms” (Figures 2, S4, Table S2).
644 Several of the PAO articles discussed the impact of salinity on EBPR processes (e.g., Wang et
645 al., 2018) As a comparison to other keyword searches in the marine category, there were 9,356
646 articles that included the keyword “phosphorus” and 757 articles that included the keyword
647 “microbiology” (Figures 2, S4, Table S2). By looking more closely at meta-analysis results
648 within the marine category, we found that most polyP and PAO articles were associated with
649 “marine” and “saltwater” keywords rather than “ocean” (Figures S4C, S4D). This suggests that
650 there is limited polyP and PAO research in the open ocean. Overall, polyP and PAO articles in
651 the marine category were more limited compared to all other categories except for agriculture;
652 2.9% of all polyP articles were in the marine category and 1.9% of all PAO articles were in the
653 marine category (Figures 2C, 2D).

654

655 Similar to terrestrial and freshwater habitats, we found limited evidence of interdisciplinary
656 research in marine habitats. About one-third ($n = 128$, 34.9%) of polyP articles in the marine
657 category did not overlap with other categories (Figure 3C). Most PAO articles in the marine
658 category overlapped with articles in the wastewater treatment category ($n = 6$, Figure 3D). As
659 mentioned previously, several discussed the impact of salinity on EBPR processes and provide
660 examples of how interdisciplinary research could benefit both as well as coastal agricultural
661 landscapes where saltwater intrusion is becoming increasingly problematic due to sea level rise
662 (e.g., Rabbani et al., 2013). There were 14 marine category articles that overlapped with polyP

663 and PAO searches. One of these studies discussed the use of diagnostic tools (Section 8) to study
664 PAO-related bacteria in estuarine sediments (Castle & Kirchman, 2004). Therefore, there is an
665 opportunity to leverage knowledge of polyP and PAOs in marine environments for the benefit of
666 agricultural P water quality management as well as management of P in natural environments.

667

668 7. KNOWLEDGE GAPS IN AGRICULTURAL SETTINGS

669 Our meta-analysis revealed that despite a large body of research on P in agricultural settings,
670 there remains limited focus on and exploration of the role of polyP and PAOs in agricultural
671 water quality management. We identified 184,042 articles in Web of Science that included the
672 keyword “phosphorus” and 55.4% of these articles represented the categories included in this
673 study (Figure 2A, Table S2). Of the five categories, the largest number of P articles ($n = 52,076$,
674 28.3%) came from the terrestrial category (Figure 2A, Table S2) and 30,817 of these did not
675 overlap with any other category (Figure 3A). We identified 4,429 P articles in the agriculture
676 category (Table S2). We also observed an increase in P articles in the agricultural category from
677 1990 to 2019 (Figure S2B). While interdisciplinary research is still limited, there is a large and
678 growing body of P research in each of the five disciplines studied here.

679

680 In contrast to P articles, we found only 14 polyP articles (Figure 2C, Table S2) in the agriculture
681 category. Several of these discussed the use of algal polyP as a P fertilizer source (e.g.,
682 Mukherjee et al., 2015; Siebers et al., 2019). One study used diagnostic tools highlighted in
683 Section 8 to characterize the P pools of marine sediments that were impacted by industrial and
684 agricultural P loads (Shan et al., 2016). We found only one PAO article in the agriculture
685 category (Figure 2D, Table S2), which used transmission electron microscopy (Section 8) to

686 analyze green algae cells with and without polyP and demonstrated that electron-dense bodies in
687 algal cells were sites of polyP accumulation when algae were kept under sulfur-depleted
688 conditions (Ota et al., 2016). These authors highlighted the importance of P in agriculture and
689 were interested in determining whether intracellular polyP from algal biomass could be used as a
690 renewable biologically-based fertilizer (Ota et al., 2016). Other studies have also suggested that
691 PAOs could be used to concentrate P (Acevedo et al., 2015). In addition to other proposed
692 organic P forms (e.g., Margenot et al., 2019), studies identified by our meta-analysis provide
693 preliminary evidence that polyP accumulated by microorganisms can be recycled (MacDonald et
694 al., 2016; Metson & Bennett, 2015; Withers et al., 2014) and serve as an alternative organic P
695 source for crops, but the efficacy of this has yet to be tested.

696

697 The increase in P articles in the agriculture category from 1990 to 2019 was not commensurate
698 with articles containing “polyphosphate”, “polyphosphate accumulating organisms”, or
699 “microbiology” keywords over the same period (Figure S2B). This finding highlights the limited
700 focus of biological P forms and cycling mechanisms in the agricultural literature (Section 3).
701 Overlapping meta-analysis articles between “phosphorus”, “polyphosphate”, and “polyphosphate
702 accumulating organisms” keyword searches demonstrate a need for interdisciplinary
703 collaboration that leverages knowledge of polyP and PAOs in wastewater treatments as well as
704 existing and emerging diagnostic tools to reduce non-point source pollution from agricultural
705 landscapes. For example, one PAO article we identified that overlapped with all categories
706 except agriculture (Figure 3D) isolated bacteria capable of accumulating P from eutrophic lake
707 water and forest soil samples (Krishnaswamy et al., 2011). This same study offered that these
708 bacteria isolates may be useful in remediating P contaminated environments. We identified

709 another PAO article that isolated fungi capable of accumulating P from soybean plants and
710 surrounding soil (Ye et al., 2015). Furthermore, we found one study that quantified polyP in
711 overland flows (Bourke et al., 2009) and several studies characterizing soil organic P (Cade-
712 Menun, 2017; B. Cade-Menun & Liu, 2013) but no studies directly addressing the role of PAOs
713 under alternating anaerobic/aerobic conditions. In the next decade, interdisciplinary soil
714 microbiome research is positioned to increase crop yield and resilience (NASEM, 2018), but
715 these advancements may also be extended to improve and protect water quality.

716
717 Based on our meta-analysis and research synthesis, we identified three additional knowledge
718 gaps that span all five major habitat categories discussed here (i.e., wastewater treatment,
719 terrestrial, freshwater, marine, and agricultural). First, across these categories there is limited
720 identification and quantification of non-*Candidatus Accumulbacter phosphatis* PAOs. For
721 example, studies used established molecular biology tools to assess the presence and/or quantity
722 of *Candidatus Accumulbacter phosphatis* P functional genes in wastewater treatment plants (e.g.,
723 (Albertsen et al., 2012; Mao et al., 2016), freshwater, sediments, and soils (e.g., Kunin et al.,
724 2008; Peterson et al., 2008) but we found only a few studies surveying natural environments for
725 *Candidatus Accumulbacter phosphatis* and non-*Candidatus Accumulbacter phosphatis* PAOs
726 (Table S4). We know polyP use is ubiquitous across the tree of life (Section 4) and PAOs exhibit
727 phenotypic and genetic diversity (Section 5); however, there remains limited analysis of PAOs
728 and P functional genes across bacterial, eukaryotic, and archaeal domains. Since it is likely that
729 *Candidatus Accumulbacter phosphatis* is not the only PAO in agricultural soils, additional
730 research—including the application of diagnostic tools (Section 8)—is needed to explore PAO
731 phenotypic and genetic diversity. Furthermore, soil microorganisms are diverse and rich

732 (Bardgett & Van Der Putten, 2014; Dunbar et al., 2002; Fierer & Jackson, 2006; Gans et al.,
733 2005; Hug et al., 2016; Schloss & Handelsman, 2006; Tringe et al., 2005). Therefore, new
734 discoveries in agricultural soils may benefit existing PAO research in EBPR wastewater
735 treatment as well as other natural environments by revealing additional genetic and metabolic
736 microbial diversity.

737

738 Second, there is limited identification and quantification of P functional genes (i.e., *ppk* and *ppx*;
739 Text S2, Table S5) across the five major categories. Researchers identified *Candidatus*
740 *Accumulibacter phosphatis* P functional genes (i.e., *ppk*) in EBPR wastewater treatment plants
741 around the world (Kunin et al., 2008; Albertsen et al., 2012; Mao et al., 2015). However, we
742 found only a few studies quantifying the abundance and diversity of non-*Candidatus*
743 *Accumulibacter phosphatis* PAO *ppk* in EBPR wastewater treatment plants (e.g., Mao et al.,
744 2016). We found two studies that quantified PAO-related P functional genes (i.e., *ppk*; Text S2,
745 Table S5) in marine habitats (Temperton et al., 2011; Watson et al. 2019), but no other studies
746 quantifying *ppk* abundance in agricultural, terrestrial, or freshwater environments. Furthermore,
747 we found only a few *ppx* studies (Text S2, Table S5). With the exception of one marine study
748 (Temperton et al., 2011), none quantified *ppx* abundance and diversity in the remainder of
749 categories discussed here. Due to their role in polyP formation and breakdown—an important
750 defining metabolic characteristic of potential PAOs—further study of *ppk* and *ppx* genes is
751 needed, regardless of discipline. With respect to agriculture, analysis of known P functional
752 genes in agricultural soils may lead to the isolation of novel PAOs that can then be studied in
753 wastewater treatment plants and other natural environments.

754

755 Third, there are very few studies that go beyond identification to assess the ecological role of
756 PAOs in categories other than wastewater treatment (Section 6). We synthesized many studies in
757 natural systems that either (1) identified PAOs directly (e.g., Peterson et al., 2008) or indirectly
758 (e.g., Khoshmanesh et al. 2002) or (2) assessed the role of biologically-mediated P uptake and
759 release under changing environmental conditions (e.g., Cohen et al., 2013). There were a few
760 that did both (e.g., Martins et al., 2011). In agricultural systems, most studies identified PAOs
761 (e.g., DebRoy et al., 2013), but did not go beyond this step. Agricultural soils undergo alternating
762 wetting/drying conditions that may facilitate PAO-mediated P cycling (Section 6.1). Therefore,
763 there remains an opportunity to study how PAO presence and quantity relate to the frequency
764 and duration of anaerobic/aerobic cycling and P availability in agricultural soils, nearby
765 waterbodies, and agricultural management practices such as vegetated buffers and bioreactors.
766 We also know very little about whether we can actively manage PAO-mediated P cycling in
767 agricultural settings to simultaneously achieve desired water quality goals and crop production
768 goals.

769

770 8. PROMISING DIAGNOSTIC TOOLS & RESEARCH APPLICATIONS FOR 771 AGRICULTURAL SETTINGS AND BEYOND

772 Opportunities exist to apply established and emerging interdisciplinary diagnostic tools such as
773 microscopy, molecular biology techniques, and other measurements to overcome barriers and
774 knowledge gaps presented previously. See Table S6 for a full description of these tools.

775 Microscopy tools can be used to identify the size, location, and amount of intracellular polyP
776 granules and can also be used to label known PAOs in agricultural soils and downstream
777 environments. For example, the 4',6-diamidino-2-phenylindole (DAPI) stain can be used in

778 combination with an epi-fluorescence microscope (e.g., Aschar-Sobbi et al., 2008; Eixler et al.,
779 2005) or a fluorescence spectrophotometer (e.g., Martin & Van Mooy, 2013) to identify and
780 quantify polyP storage in microbial cells. Molecular biology tools such as quantitative real-time
781 polymerase chain reaction (qPCR) and next generation sequencing—including shotgun and
782 amplicon metagenomics—can be used to quantify P functional genes and identify undiscovered
783 PAOs in agricultural fields and best management practices such as vegetated buffers. For
784 example, one study developed and used qPCR probes to quantify different genetically similar
785 sub-groups of *Candidatus Accumulibacter phosphatis* in nine wastewater treatment plants
786 (Zhang et al., 2016). Shotgun metagenomics data from a global ocean dataset was used to
787 quantify P functional genes (Temperton et al., 2011). In addition to microscopy and molecular
788 biology techniques, tools like ³¹P-nuclear magnetic resonance (NMR) spectroscopy can be used
789 to measure the concentration of polyP and other organic P forms in agricultural soils and
790 sediments (e.g., Kenney et al., 2015; Cade-Menun, 2017). Finally, high-frequency sensors can be
791 used to measure oxygen concentrations in soils and phosphate concentrations in the water
792 column or in tile drainage. These sensors can help capture environmental variables at time scales
793 that are more closely aligned with microbial processes.

794

795 When using the tools discussed above, care must be taken to ensure that microbial and
796 environmental measurement time and spatial scales are compatible (Battin et al., 2016; Bier et
797 al., 2015; Blackwell et al., 2010; Hall et al., 2018), measurement bias for/against active microbial
798 community members is understood (Carini et al., 2016; Jones & Lennon, 2010; Lennon & Jones,
799 2011; J. Schimel & Gullledge, 1998), and environmental variables (e.g., pH and temperature) that
800 may influence microbial communities are accounted for in the experimental design (Battin et al.,

801 2016; Bier et al., 2015; Dinh et al., 2016; Fierer & Jackson, 2006; Lauber et al., 2009; Oliverio et
802 al., 2016; Rousk et al., 2011; Schimel & Gullledge, 1998). Researchers must also take care to
803 design controls that consider abiotic processes which may mimic PAO-mediated P cycling (e.g.,
804 reductive dissolution of Fe-P).

805

806 Beyond specific tools, several general frameworks exist to link microbial with ecosystem-
807 scale—or potentially, watershed-scale—processes (Bier et al., 2015; Hall et al., 2018; Martiny et
808 al., 2015; Nemergut et al., 2014; Prosser, 2013; Schimel et al., 1999; Wallenstein & Hall, 2012).
809 These frameworks have been applied to research on microbially-mediated nitrogen and carbon
810 cycling but have not been applied to study microbially-mediated P cycling in natural and
811 agricultural settings. As researchers establish studies to explore microbial P cycling in new
812 habitats, they can refer to previous work for guidance on microbial-scale ecological theories
813 (Choudoir et al., 2017; Prosser et al., 2007), statistical approaches and considerations (Bernhardt
814 et al., 2017; Bier et al., 2015; Buttigieg & Ramette, 2014; Rocca et al., 2015; Schimel &
815 Gullledge, 1998; Willis, 2016; Willis et al., 2017), method overviews (Ekblom & Wolf, 2014;
816 Howe et al., 2014; Howe & Chain, 2015; Kozich et al., 2013; Pallen, 2016; Riesenfeld et al.,
817 2004; Schloss, 2020; Zimmerman et al., 2014), bioinformatics (Cock et al., 2009; Howe et al.,
818 2014; Loman & Watson, 2013; Shade & Teal, 2015; Wilson et al., 2016), reproducible research
819 (da Veiga Leprovost et al., 2014; Perez-Riverol et al., 2016; Schloss, 2017; Shade & Teal, 2015;
820 Wilson et al., 2016), and modelling (Graham et al., 2014, 2016; Manzoni et al., 2014; Powell et
821 al., 2015; Reed et al., 2014; Todd-Brown et al., 2012; Wieder et al., 2013).

822

823 9. RESEARCH NEEDS FOR AGRICULTURAL SETTINGS & BEYOND

824 Our meta-analysis demonstrated a need for research assessing the degree to which polyP and
825 PAOs impact P agricultural water quality management efforts, and ultimately, whether and how
826 these impacts influence the achievement of short- and long-term water quality goals. Research on
827 polyP and PAOs in agricultural landscapes can be combined with established physical and
828 chemical P controls to initiate the development and testing of agricultural P water quality
829 management strategies that overcome what we highlight as persistent barriers to reducing P
830 pollution (Section 1): legacy P, unintended consequences of existing management strategies, and
831 changing environmental conditions. Meta-analysis results revealed important opportunities for
832 interdisciplinary research on polyP and PAO research in the agricultural sciences and beyond. In
833 addition to advancing agricultural water quality management, studies of polyP and PAOs in
834 agricultural settings may benefit the treatment of wastewater and knowledge of P cycling in
835 natural environments. Specifically, we see expanding the known diversity of PAOs and refining
836 diagnostic tools for characterizing PAO community composition and function as important first
837 steps in this effort. Below, we summarize specific research priorities that we believe will fill key
838 polyP and PAO knowledge gaps while alleviating some of the challenges stymying meaningful
839 reductions in P pollution.

840

841 *Quantify Abiotic versus Biotic Phosphorus Pools* – Comparisons between the relative magnitude
842 of abiotic and biotic P pools is a necessary first-order data need. These comparisons could be
843 made along gradients of legacy P in soils and sediments that capture shorter (years) to longer
844 (decades) histories of legacy P. In addition, P pool comparisons (e.g., inorganic P forms versus
845 organic P forms) between undisturbed soils, cultivated soils, and soils associated with

846 agricultural management practices (e.g. riparian buffers and bioreactors) are needed. Last, P pool
847 comparisons along temperature and soil moisture gradients are needed to address whether and
848 how biotic P retains external and internal P sources compared to abiotic P retention.

849 *Identify and Quantify PAOs and P Functional Genes* – Studies identifying and quantifying
850 known PAOs (e.g., *Candidatus Accumulibacter phosphatis*), new PAOs (i.e., microbes
851 exhibiting PAO metabolism where P release and uptake corresponds with anaerobic and aerobic
852 conditions, respectively), and P functional genes in agricultural soils and agricultural
853 management practices are needed.

854 *Describe the Role of PAOs* – Assessment of statistically significant relationships between PAO
855 abundance, P functional gene abundance, and soil P forms (e.g., microbial P) in agricultural soils
856 and management practices undergoing alternating wetting/drying conditions are needed.

857 Furthermore, there is a need for these studies to also examine how the intensity, duration, and
858 frequency of temperature changes and alternating wetting/drying cycles affect PAO-mediated P
859 uptake and release.

860 *Implement Next Generation Management Practices* – Studies that design and test the
861 performance and feasibility of in-field or edge-of-field, EBPR-inspired, PAO-friendly
862 agricultural management practices are needed. Furthermore, there is an opportunity to augment
863 the effective physical and chemical P retention mechanisms of existing agricultural management
864 strategies with biological P retention. For example, an edge-of-field bioreactor intercepting tile
865 drainage to for nitrate removal could be retrofitted according to EBPR wastewater treatment
866 design (Section 5) to also treat DP. Hypothetically, the retrofitted bioreactor could promote PAO
867 community growth and removal of DP from tile drain effluent while establishing new methods to
868 recycle P from settled microbial biomass. For agricultural management practices that rely more

869 on rainfall events to drive alternating anaerobic/aerobic conditions (i.e., saturation of riparian
870 buffer soils), feasibility testing will likely be necessary to address whether the frequency of
871 alternating anaerobic/aerobic conditions is adequate to ensure PAO community stability and P
872 retention.

873 *Interdisciplinary Research* – There is a need for interdisciplinary studies that test and leverage
874 diagnostic tools (Section 8) as well as results of research needs listed above (e.g., identification
875 of new PAOs) from agricultural settings to address broader questions concerning the origin and
876 role of polyP and PAOs in wastewater treatment plants and natural environments undergoing
877 alternating anaerobic/aerobic conditions (Figure 1).

878

879 ACKNOWLEDGEMENTS

880 Author contributions: SMS designed the meta-analysis and analyzed the data. All authors
881 interpreted the data. SMS drafted the manuscript and prepared data and code for GitHub/Zenodo.
882 All authors provided critical revision.

883

884 The authors would like to thank Jennifer Rocca, Edward Hall, Brian Rahm, Miranda Stockton,
885 and Claudia Rojas for their critical feedback. This project was supported by funding from the US
886 Department of Agriculture (UDSA) grant 2014-67019-21636. This publication was also
887 developed under STAR Fellowship Assistance Agreement no. FP917670-01-0 awarded by the
888 US Environmental Protection Agency (EPA). It has not been formally reviewed by the EPA. The
889 views expressed in this publication are solely those of SMS and EPA does not endorse any
890 products or commercial services mentioned in this publication. This article is contribution
891 number <will fill in upon publication> of the Institute for Great Lakes Research at Central

892 Michigan University. All data and analysis scripts associated with this publication are available
893 on GitHub at <will fill in upon publication> and Zenodo (DOI: <will fill in upon publication>).

894 A preprint of this publication is available on the EarthArXiv at <https://eartharxiv.org/ge95h/>.

895

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Supporting Information for

**A Critical Review of Polyphosphate and Polyphosphate Accumulating Organisms
for Agricultural Water Resources Management**

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File Contents

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Information Text, Tables, and Figures. *Number of Pages*: 34.

Contents Metadata

This document includes the supplementary figures and tables for this study as referred to in the
main text of the article. All data and analysis scripts associated with this publication are available
on GitHub at <will fill in upon publication> and Zenodo (DOI: <will fill in upon publication>).

A preprint of this publication is available on the EarthArXiv at <https://eartharxiv.org/ge95h/>.

1756 TEXT

1757 **Text S1.** *Candidatus* Accumulibacter phosphatis Metabolism

1758 The metabolism of model polyphosphate accumulating organisms (PAO), *Candidatus*
1759 Accumulibacter phosphatis, in operational EBPR wastewater treatment plants is described as
1760 follows. During anaerobic conditions (Figure S3A), *Candidatus* Accumulibacter phosphatis
1761 takes up short chain volatile fatty acids (VFAs) such as acetate and store them as poly- β -
1762 hydroxyalkanoates (PHAs) like poly- β -hydroxybutyrate (PHB; Seviour et al. 2003; Seviour and
1763 Nielsen 2010). Intracellular polyphosphate (polyP) and glycogen concentrations decrease
1764 because they are used to convert VFA to PHA (Seviour et al., 2003; Seviour and Nielsen, 2010).
1765 Phosphate cleaved from the terminal end of a polyP chain during this process is exported from
1766 the cell, thereby contributing to an increase in the concentration of phosphate (i.e., P_i in Figure
1767 S3) in the bulk wastewater (Seviour et al., 2003; Seviour and Nielsen, 2010). *Candidatus*
1768 Accumulibacter phosphatis uses the energy released from the respiration of PHAs to replace
1769 polyP and glycogen (Seviour et al., 2003; Seviour and Nielsen, 2010) during aerobic periods
1770 (Figure S3B). As a result, *Candidatus* Accumulibacter phosphatis uptakes phosphate to build
1771 polyP chains, thereby drawing down bulk water phosphate concentrations in the wastewater prior
1772 to its discharge from the wastewater treatment plant (Seviour et al., 2003; Seviour and Nielsen,
1773 2010).

1774

1775 **Text S2.** Phosphorus Cycling Functional Genes

1776 There are several known functional genes associated with PAO-mediated P uptake and release in
1777 EBPR (Table S5). Polyphosphate kinases PPK1 and PPK2—coded for by *ppk1* and *ppk2*,
1778 respectively—catalyze the reversible reaction of ATP to ADP to form intracellular polyP (Table
1779 S5). The nucleotide sequence for *ppk1* was first isolated from *Escherichia coli* (Akiyama et al.,
1780 1992) and has since been identified in a wide range of bacterial, archaeal, and eukaryotic
1781 organisms (Trelstad et al., 1999; Zhang et al., 2002; Rao et al., 2009; Kawakoshi et al., 2012)
1782 and PAOs (McMahon et al., 2002; He et al., 2007; Zhang et al., 2016). PPK1 is likely a
1783 membrane-bound protein with four domains concentrated in regions where the inner and outer
1784 cell membranes come together (Ahn and Kornberg, 1990). The nucleotide sequence of *ppk2* was
1785 first isolated from *Pseudomonas aeruginosa* (Zhang et al., 2002). PPK2 differs from PPK1 in its
1786 ability to catalyze the formation of polyP from both GTP and ATP as well as enzyme cofactors.
1787 Also, PPK2 has an affinity for Mn^{2+} while PPK1 has an affinity for Mg^{2+} (Zhang et al., 2002;
1788 Rao et al., 2009). Because they are highly conserved (Zhang et al., 2002; Rao et al., 2009), *ppk*'s
1789 such as *ppk1* and *ppk2* are an ideal gene marker for bacterial strain diversity. Some microbes
1790 possess two *ppk*'s (*ppk1* and *ppk2*) while others only have one (Zhang et al., 2002; Rao et al.,
1791 2009; Temperton et al., 2011; Kawakoshi et al., 2012). Specific to PAOs, the *Candidatus*
1792 *Accumulibacter phosphatis* genome has a single copy of *ppk1* and it evolves faster than
1793 *Candidatus Accumulibacter phosphatis* 16S rRNA genes (Kunin et al., 2008; He and McMahon,
1794 2011a).

1795

1796 Exopolyphosphatase PPX1 and PPX2/GPPA (coded for by *ppx1* and *ppx2/gppA*) catalyze the
1797 breakdown of polyP. PPX1 breaks off the terminal phosphate molecules of a polyP chain when

1798 excess phosphate is present (Table S5). PPX1 preferentially acts on longer chains of polyP (i.e.,
1799 500 phosphate molecules or longer), does not act on ATP, and cannot be inhibited by ADP or
1800 ATP (Akiyama et al., 1993). PPX2/GPPA, also referred to as pentaphosphate phosphohydrolase,
1801 inhibits polyP accumulation at the enzymatic level by hydrolyzing stress response nucleotides
1802 pppGpp to ppGpp or catalyzes the release of phosphate by breaking polyP chains (Table S5).
1803 PPX2/GPPA is thought to be less active than PPX1, prefers longer polyP chains (i.e., 1000
1804 residues or longer), and is inhibited by the presence of short- and medium-length polyP chains
1805 (Keasling et al., 1993). Some organisms have both *ppx1* and *ppx2/gppA* (Keasling et al., 1993;
1806 Alcántara et al., 2014), but this trend is not well characterized for PAOs.

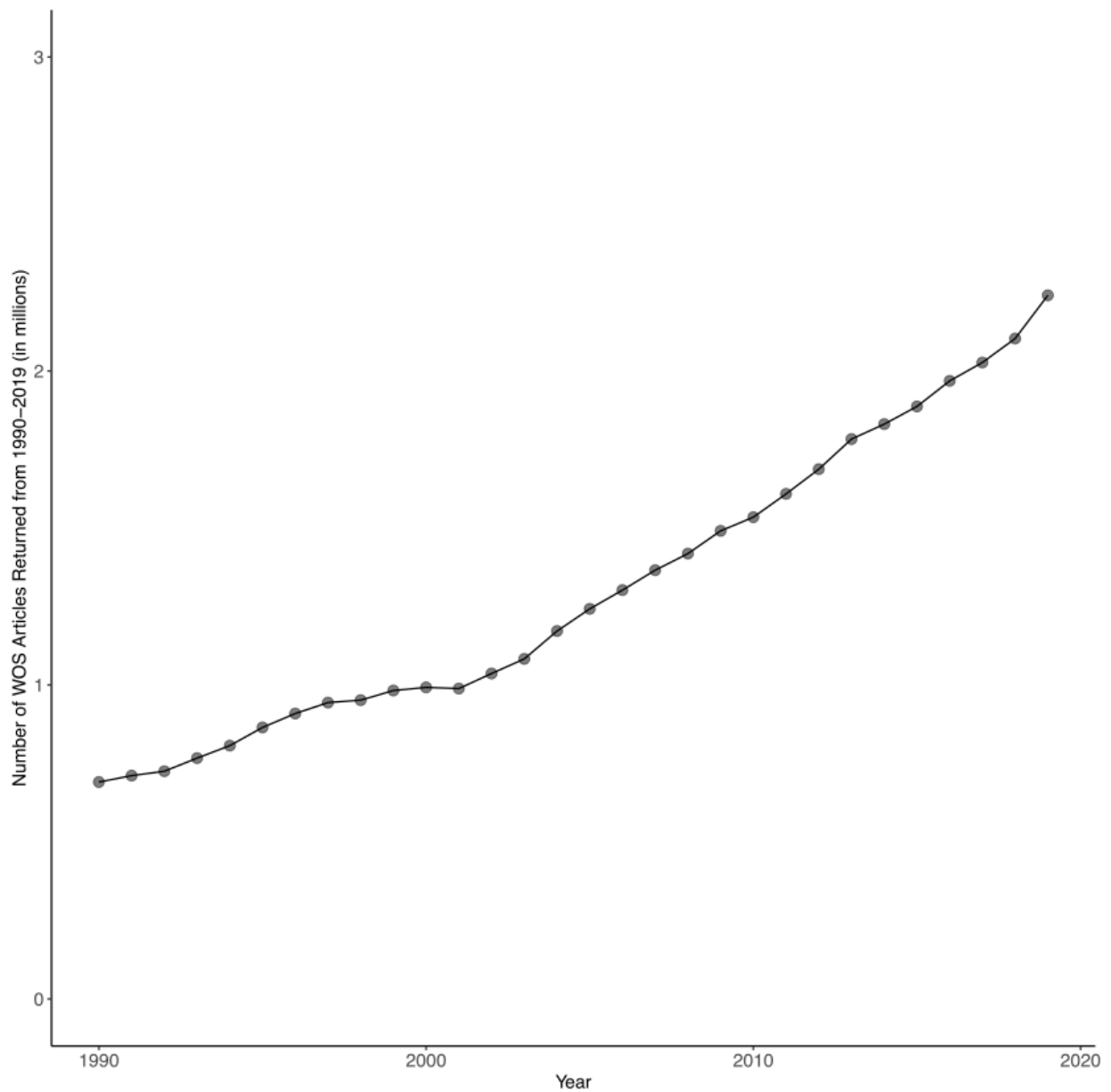
1807

1808 Other P cycling genes of interest include *pap*, *phoX*, *phoD*, *phoA*, *pit*, *pst*, and *ppn*. Associated
1809 protein functionality and key traits are summarized in Table S5. As we mentioned previously for
1810 *ppk* and *ppx*, more work needs to be done to characterize the abundance, diversity, and role of
1811 these genes in P cycling with respect to engineered systems. While we only focus on P-related
1812 genes here, little is known about functional genes regulating other important PAO polymers:
1813 PHA and glycogen (Seviour et al., 2003).

1814

1815

1816 FIGURES



1817

1818 **Figure S1.** Number of the total number of Web of Science (WOS) articles published per year

1819 from 1990 to 2019.

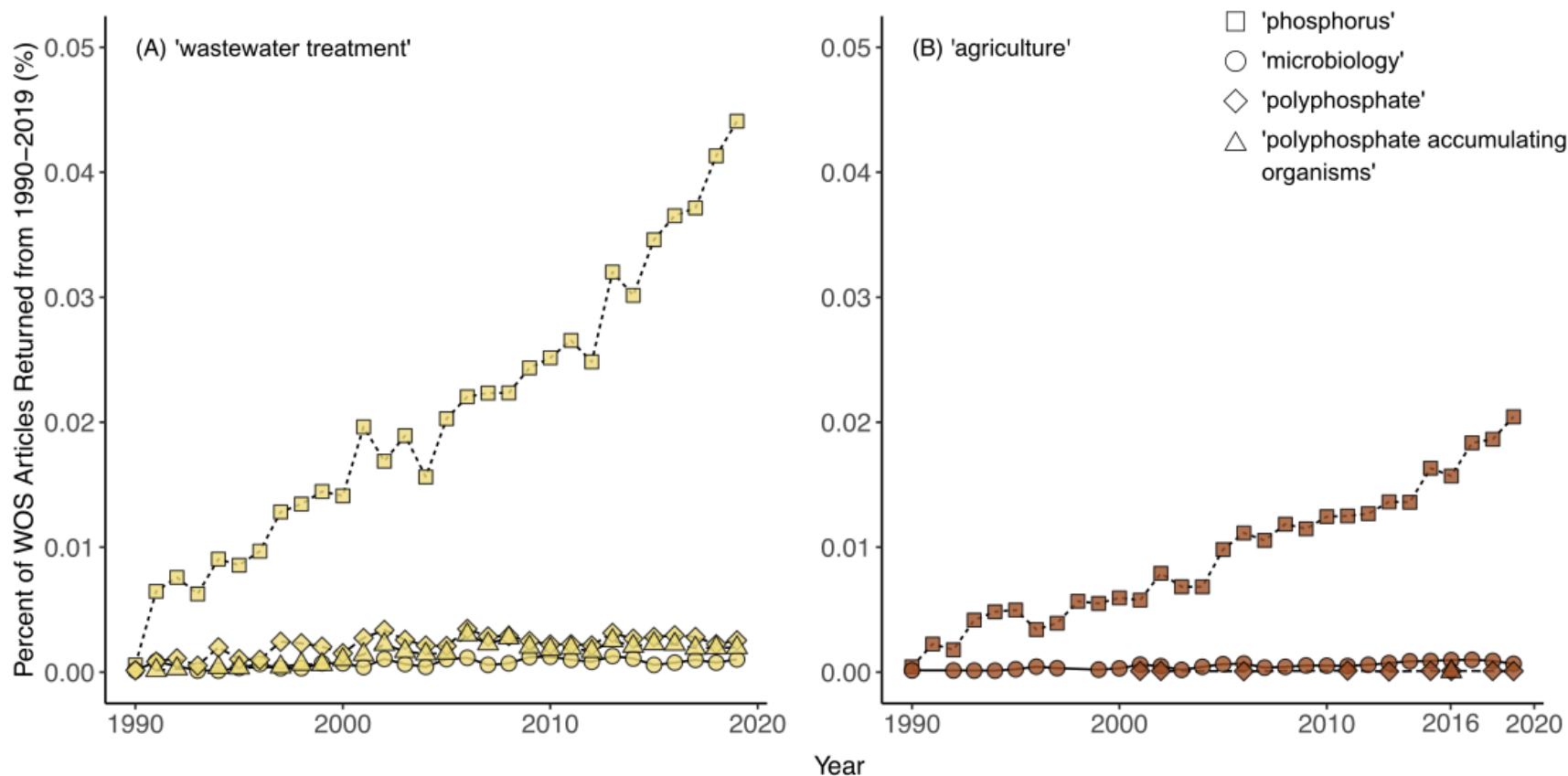
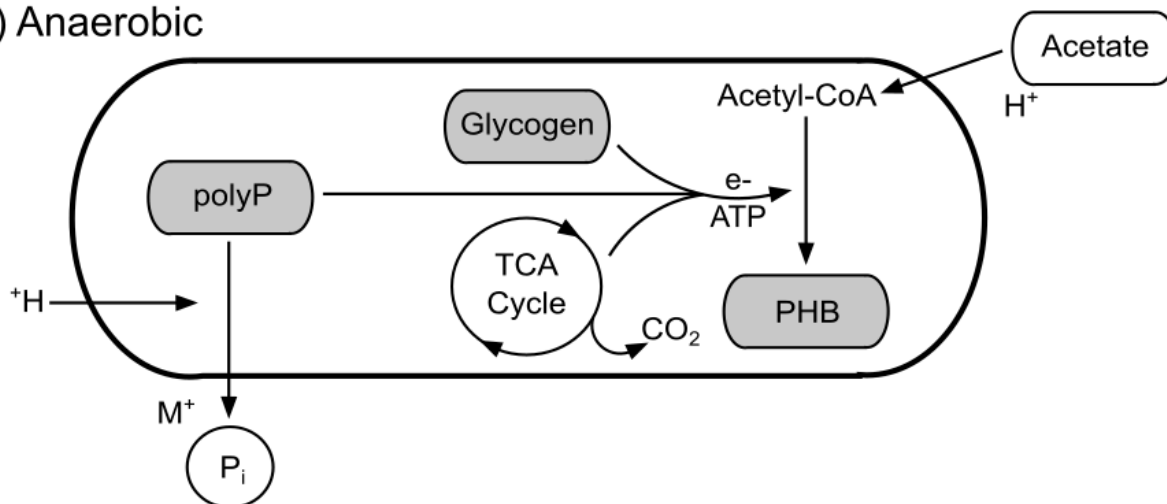


Figure S2. Percent of Web of Science (WOS) articles returned from 1990-2019 for various (A) 'wastewater treatment' and (B) 'agriculture' keyword searches. For example, the open box line represents the WOS search “ 'wastewater treatment' AND 'phosphorus' ”. Note: There is only one “ 'agriculture' AND 'polyphosphate accumulating organisms' ” article in 2016.

(A) Anaerobic



(B) Aerobic

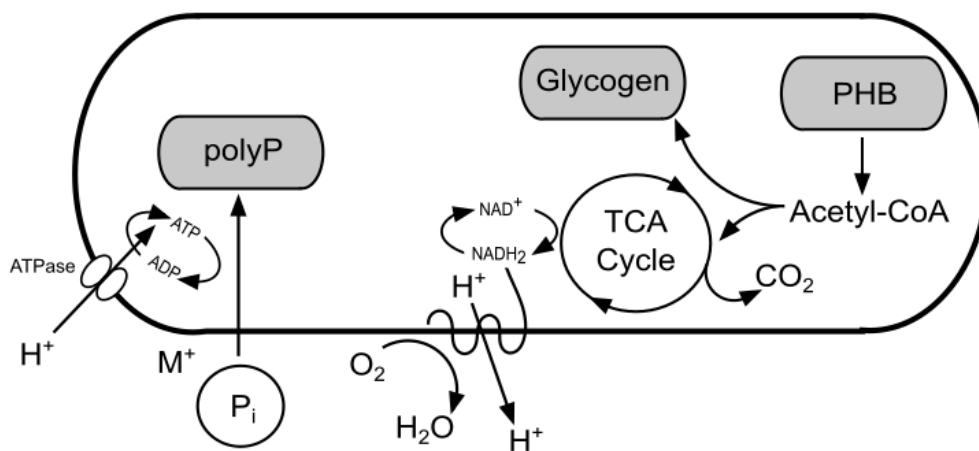


Figure S3. (A) Anaerobic and (B) aerobic metabolism of model polyphosphate accumulating organism (PAO), *Candidatus Accumulibacter phosphatis*. The poly- β -hydroxyalkanoate (PHA) known as poly- β -hydroxybutyrate (PHB) is specific to CAP. Abbreviations: metal cations (M⁺), phosphate (P_i). Adapted from Seviour et al. (2003), Seviour and Nielsen (2010), and Skennerton et al. (2014).

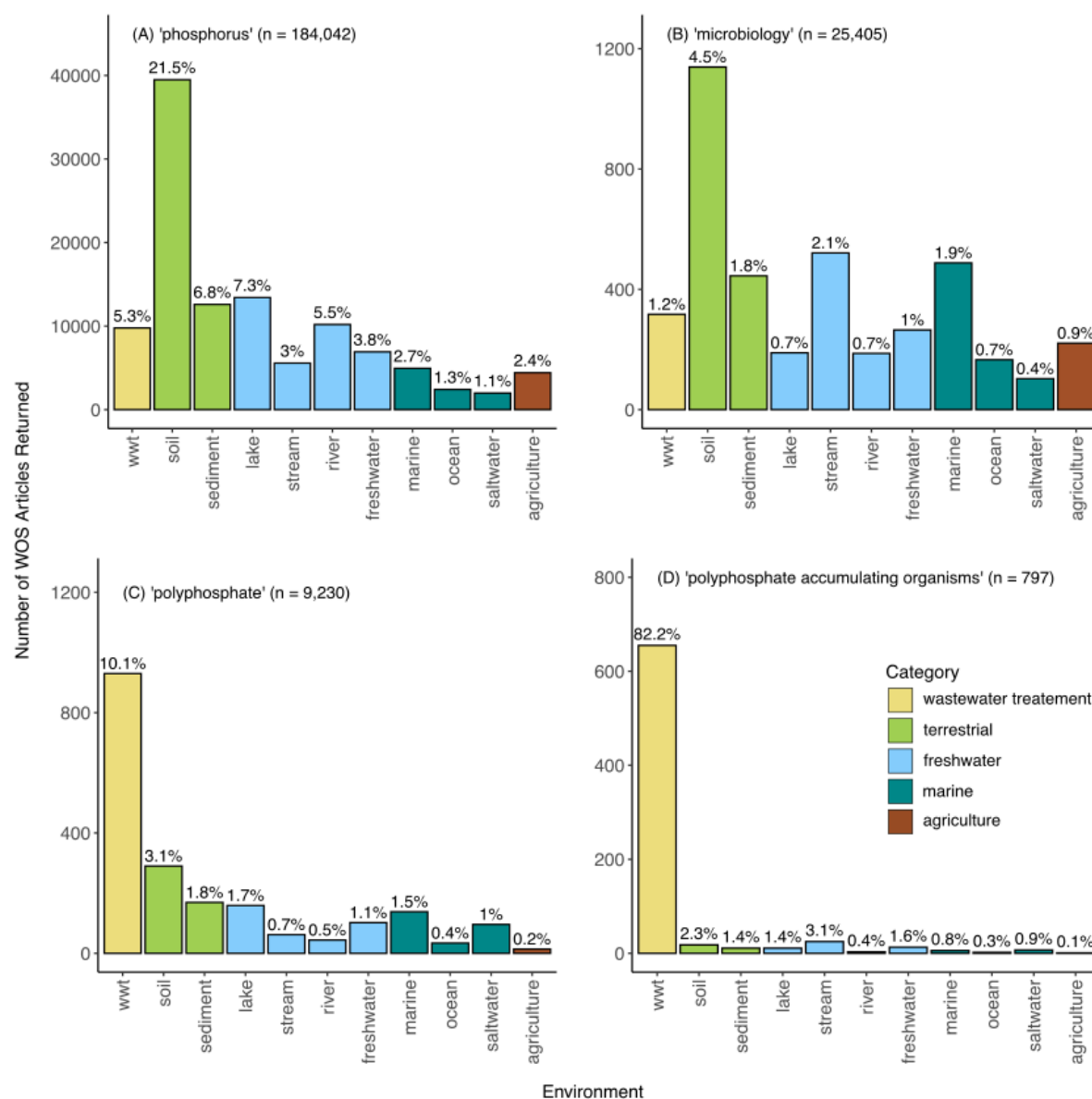


Figure S4. Number and percentage of articles returned from Web of Science (WOS) by environment for (A) ‘phosphorus’, (B) ‘microbiology’, (C) ‘polyphosphate’, and (D) ‘polyphosphate accumulating organisms’ keyword searches. Environments are grouped into five categories: wastewater treatment (wwt), terrestrial, freshwater, marine, and agriculture. Percentages on top of each bar are relative to the number of articles in the full figure subset search (i.e., n = 184,042 for ‘phosphorus’); they may not sum to 100% when search results are broader than the environments we focus on in this study.

TABLES

Table S1. Common phosphorus (P) forms and their descriptions.

| Form | Common Abbreviation | Description | References |
|--|--------------------------------------|---|---|
| dissolved phosphorus | DP | Operationally defined form of P that passes through 0.45um filter. | USEPA, 1978; Zeckoski et al., 2013 |
| particulate phosphorus | PP | Operationally defined form of P that is obtained by subtracting DP from TP. | Zeckoski et al., 2013 |
| total phosphorus | TP | Total amount of P in a soil or water sample. For water samples TP is measured on an unfiltered sample and for soil samples TP is determined after digestion using strong acids and bases like fluoric acid, hydrogen peroxide, hydrochloric and nitric acids, sodium hydroxide. | USEPA, 1978; Zeckoski et al., 2013 |
| soluble reactive phosphorus, | SRP | Portion of the DP fraction (i.e., P passing through 0.45 um filter) that can be detected with the molybdenum blue assay. SRP consists of primarily Pi but may also include hydrolyzed Po due to the required acidity of the molybdenum blue assay. | USEPA, 1978; Benitez-Nelson, 2000; Zeckoski et al., 2013 |
| dissolved orthophosphate inorganic phosphorus | Pi | P that is not associated with organic (carbon) molecules. Examples include phosphate, polyphosphate, and phosphate bound directly to metals (e.g., apatite - $\text{Ca}_{10}(\text{PO}_4)_6(\text{OH})_2$). | Zeckoski et al., 2013 |
| dissolved inorganic phosphorus phosphate, orthophosphate | DIP PO ₄ ³⁻ | The Pi fraction of DP. The most basic form of Pi in the environment. | USEPA, 1978; Benitez-Nelson, 2000 Zeckoski et al., 2013 |
| organic phosphorus | Po | P associated with organic (carbon containing) material/molecules. This may include phosphate bound to organic matter as well as phosphate inside plants, animals, and microorganisms (e.g., as DNA or polyP). | Zeckoski et al., 2013; Cade-Menun et al., 2005; Cade-Menun, 2015; 2017 |
| polyphosphate | polyP | Two or more phosphate molecules bound together by a high energy phosphoanhydride bond (i.e., tetrahedral phosphate groups are linked via O ₂ bonds). It is a Pi on its own, but because it is typically stored intracellularly, it is often considered a Po. | Harold, 1964; Kornberg, 1995; Cade-Menun et al., 2005; Cade-Menun, 2015; 2017 |
| pyrophosphate | -- | A polyP with only two phosphate molecules bound together by a phosphoanhydride bond. It is a Pi on its own, but because it is typically stored intracellularly, it is often considered a Po. | Cade-Menun et al., 2005; Cade-Menun, 2015; 2017 |
| dissolved organic phosphorus orthophosphate monoesters | DOP -- | The Po fraction of DP. Extracted from organic P. Includes sugar phosphates (e.g., glucose 1-phosphate). | USEPA, 1978; Benitez-Nelson, 2000 Cade-Menun et al., 2005; Cade-Menun 2015 |

| | | | |
|-------------------------|----|---|--|
| orthophosphate diesters | -- | Extracted from organic P and they include DNA, RNA, and phospholipids. | Cade-Menun et al., 2005; Cade-Menun, 2015 |
| phosphonate | -- | Extracted from organic P and they are most commonly found as free molecules or membrane phosphonolipids. | Cade-Menun et al., 2005; Cade-Menun, 2015 |
| microbial phosphorus | -- | Pi and Po stored intracellularly by microorganisms; therefore, it is considered a form of Po. Calculated based on the difference in P detected before and after fumigation of soil or sediment samples. | Hedley et al., 1982 |
| bound phosphorus | -- | A general term that refers to P (usually inorganic P) that is attached to soil or sediment via bonds or associations with organic matter or metals such as Ca, Fe, and Al. | -- |
| labile phosphorus | -- | P that is easy converted into other forms. | Zeckoski et al., 2013 |
| bioavailable phosphorus | -- | P that can easily be taken up by plants, animals, microorganisms. Determined by summing SRP and what is extracted from PP using NaOH. | Sharpley et al., 1991; Zeckoski et al., 2013 |

Table S2. Summary of Web of Science queries results.

| Theme | Category | Query Keywords | Number of Results Returned |
|----------------------|----------------------|---|----------------------------|
| Phosphorus | | | |
| | all | "phosphorus" | 184,042 |
| | wastewater treatment | "phosphorus" AND ("wastewater" OR "enhanced biological phosphorus removal" OR "batch reactor") | 9,774 |
| | terrestrial | "phosphorus" AND "soil" | 39,481 |
| | terrestrial | "phosphorus" AND "sediment" | 12,595 |
| | freshwater | "phosphorus" AND "lake" | 13,421 |
| | freshwater | "phosphorus" AND "stream" | 5,572 |
| | freshwater | "phosphorus" AND "river" | 10,190 |
| | freshwater | "phosphorus" AND ("freshwater" OR "fresh water") | 6,922 |
| | marine | "phosphorus" AND "marine" | 4,949 |
| | marine | "phosphorus" AND "ocean" | 2,419 |
| | marine | "phosphorus" AND ("salt water" OR "saltwater") | 1,988 |
| | agriculture | "phosphorus" AND "agriculture" | 4,429 |
| Microbiology | | | |
| | all | "microbiology" | 25,405 |
| | wastewater treatment | "microbiology" AND ("wastewater" OR "enhanced biological phosphorus removal" OR "batch reactor") | 317 |
| | terrestrial | "microbiology" AND "soil" | 1,139 |
| | terrestrial | "microbiology" AND "sediment" | 445 |
| | freshwater | "microbiology" AND "lake" | 189 |
| | freshwater | "microbiology" AND "stream" | 521 |
| | freshwater | "microbiology" AND "river" | 187 |
| | freshwater | "microbiology" AND ("freshwater" OR "fresh water") | 265 |
| | marine | "microbiology" AND "marine" | 488 |
| | marine | "microbiology" AND "ocean" | 166 |
| | marine | "microbiology" AND ("salt water" OR "saltwater") | 103 |
| | agriculture | "microbiology" AND "agriculture" | 221 |
| Polyphosphate | | | |
| | all | "polyphosphate" | 9,230 |
| | wastewater treatment | "polyphosphate" AND ("wastewater" OR "enhanced biological phosphorus removal" OR "batch reactor") | 930 |
| | terrestrial | "polyphosphate" AND "soil" | 290 |
| | terrestrial | "polyphosphate" AND "sediment" | 169 |
| | freshwater | "polyphosphate" AND "lake" | 159 |
| | freshwater | "polyphosphate" AND "stream" | 62 |
| | freshwater | "polyphosphate" AND "river" | 44 |

| | | |
|----------------------|--|-----|
| freshwater | "polyphosphate" AND ("freshwater" OR "fresh water") | 102 |
| marine | "polyphosphate" AND "marine" | 138 |
| marine | "polyphosphate" AND "ocean" | 34 |
| marine | "polyphosphate" AND ("salt water" OR "saltwater") | 96 |
| agriculture | "polyphosphate" AND "agriculture" | 14 |
| PAOs | | |
| all | "polyphosphate accumulating organisms" | 797 |
| wastewater treatment | "polyphosphate accumulating organisms" AND ("wastewater" OR "enhanced biological phosphorus removal" OR "batch reactor") | 655 |
| terrestrial | "polyphosphate accumulating organisms" AND "soil" | 18 |
| terrestrial | "polyphosphate accumulating organisms" AND "sediment" | 11 |
| freshwater | "polyphosphate accumulating organisms" AND "lake" | 11 |
| freshwater | "polyphosphate accumulating organisms" AND "stream" | 25 |
| freshwater | "polyphosphate accumulating organisms" AND "river" | 3 |
| freshwater | "polyphosphate accumulating organisms" AND ("freshwater" OR "fresh water") | 13 |
| marine | "polyphosphate accumulating organisms" AND "marine" | 6 |
| marine | "polyphosphate accumulating organisms" AND "ocean" | 2 |
| marine | "polyphosphate accumulating organisms" AND ("salt water" OR "saltwater") | 7 |
| agriculture | "polyphosphate accumulating organisms" AND "agriculture" | 1 |

Table S3. Targeted forms, advantages, and disadvantages for agricultural phosphorus (P) best management practices (BMPs).

Abbreviations: particular P (PP), dissolved P (DP).

| BMP | Targeted P Form(s) | Advantages | Disadvantages | References |
|---|--------------------|---|--|--|
| tile drains (including other controlled drainage) | PP, DP | Increases infiltration and reduces DP and PP losses in overland flow. Intercepts lateral, subsurface flows. DP in infiltrating or lateral flows adsorbs to soil. Increases crop yield. | PP builds up in the top layers of the soil profile. Serves as a conduit (i.e., artificial preferential flow path) facilitating subsurface transport of DP to nearby streams and water bodies. Medium cost to operations. | Rittenberg et al., 2015; McDowell et al., 2014b; 2016; King et al., 2015; Williams et al., 2016 |
| soil aeration | PP, DP | Decrease of PP and DP loss by increasing infiltration of overland flows. Increased airflow to plant leads to increases in plant growth and nutrient uptake. Airflow also increases soil carbon respiration and leads to subsequent mineralization of organic P to DP. | Loss of PP to atmosphere through wind erosion. Increases DP and PP loss via overland flow and erosion by disturbing soil structure. | Saif, 1981; Schachtman et al., 1998; Johnson et al., 2011; Baumhardt et al., 2015 |
| soil amendments (e.g., pH- or mineral-based) | DP | Decreases DP loss via overland flow and subsurface leaching. Includes adding lime or red mud to raise soil pH in acidic soils or additions of minerals (e.g., aluminum sulfate) to aid DP adsorption to field soils or other BMPs (e.g., riparian buffer soils). | Amendments cannot remove P once it reaches P adsorption capacity. Limited research on the ecological impacts of mineral accumulation downstream. High cost to operations. | Callahan et al., 2012; Buda et al., 2012; McDowell and Nash, 2012; McDowell et al., 2014b; 2016 |
| woodchip bioreactors with biochar additions | DP | Post-processing of tile drainage that removes DP. Edge-of-field practice that does not take agricultural land out of production. | No longer removes DP when biochar amendments reach P adsorption capacity. Select woodchip types leach P during the start-up phase. High cost to operations. | Christianson et al., 2012; McDowell et al., 2014b; Bock et al., 2015; Sharrer et al., 2016; |
| vegetated filter strip & riparian buffers | PP, DP | Decreases PP losses by slowing down overland flow. Reduces DP losses by facilitating infiltration and adsorption in soil profile. | Accumulation of P in the top layers of the soil profile. Soil cannot remove P once it reaches P adsorption capacity. Preferential flow paths develop if not properly maintained. Medium cost to operations. | Hoffmann et al., 2009; Tomer et al., 2011; Rittenberg et al., 2015; McDowell et al., 2014b; 2016; Liu et al., 2017 |
| constructed wetlands | PP, DP | Decreases PP losses by collecting and slowing down overland flows. Wetland vegetation and organisms can take up DP. Wetland sediments may adsorb DP. | PP collects in the wetland. High cost to operations. | Hill and Robinson, 2012; Kaushal et al., 2014; Rittenberg et al., 2015; McDowell et al., 2014b; 2016; Liu et al., 2017 |
| reduced tillage (e.g., no till) | PP | Preserves soil structure, which holds PP in place during subsequent overland flow runoff events. Limited tillage decreases mineralization of organic P to DP. | Promotes formation of preferential flow paths, which can serve as DP conduits. May cause vertical stratification of P. | Sharpley et al., 2011; Tomer et al., 2011; Baumhardt et al., 2015; Rittenberg et al., 2015; Christianson et al., 2016; |
| terracing & contour farming | PP | Decreases PP losses by minimizing soil erosion. May be coupled with tile drainage. | Limited impact on DP. Requires considerable alteration of field slopes. | Gassman et al., 2006; Rittenberg et al., 2015 |

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| | | | | |
|---|--------|---|---|--|
| cover crops (also referred to as catch crops) | PP, DP | Preserves soil structure during non-growing seasons, which holds PP in place during subsequent overland flow runoff events. Increases infiltration capacity, which reduces DP losses. | Freeze and thaw events trigger DP losses from cover crops. | Kleinman et al., 2005; Tomer et al., 2011; Bergström et al. 2015; Rittenberg et al., 2015; Cober et al., 2019; Liu et al., 2019 |
| stream fencing | PP, DP | Decreases DP and PP loads by preventing direct defecation of livestock into the stream and ensuring streambank stability. Low cost to operations. | Alternative supplies are required to water livestock. | Tomer et al., 2011; McDowell et al., 2014b; 2016 |
| land conservation (e.g., conservation reserve program; CRP) | PP, DP | Reduces PP losses and P fertilizer applications by removing erodible land from cropping and pasturing. Crop land is lost but producers enter 10-year rental contract with the US Department of Agriculture; they are compensated for this loss. | High cost to operations. Duration of practice implementation may depend on economic viability for the producer. | Gregory et al., 2007; Tomer et al., 2011; Rittenberg et al., 2015; Smith et al., 2015b |
| nutrient management (e.g., 4Rs: right source, right timing, right placement, and right rate). | PP, DP | Decreases DP and PP losses by using low solubility P fertilizers, avoiding fertilizer application when probability of precipitation is high, reducing amount of fertilizer applied, encouraging mixing of fertilizer into soil, and balancing crop needs with P available through fertilizer and existing soil P. Low cost to operations. | P is an essential and typically limited nutrient for plant growth; decreases in P availability decrease crop yield. | Schachtman et al., 1998; Good and Beatty, 2011; Tomer et al., 2011; Bergström et al. 2015; Rittenberg et al., 2015; McDowell et al., 2014b; Christianson et al., 2016; McDowell et al., 2016 |

Table S4. A summary of polyphosphate accumulating organisms (PAOs) identified in engineered and natural environments. Abbreviations: full-scale wastewater treatment plant (WWTP), sequencing batch reactor (SBR), uncharacterized wastewater (Mixed).

| Phylum/Class | Genus | Carbon Source | Reactor Type /Environment | Reference |
|----------------------------|------------------------|---------------|---------------------------|--|
| <i>Actinobacteria</i> | -- | Mixed | WWTP | Auling et al., 1991; Beer et al., 2006 |
| <i>Actinobacteria</i> | -- | Mixed | SBR | Bark et al., 1993 |
| <i>Actinobacteria</i> | <i>Arthrobacter</i> | Mixed | Rhizosphere soil | Li et al., 2013 |
| <i>Actinobacteria</i> | <i>Gordonia</i> | Mixed | WWTP | Beer et al., 2006 |
| <i>Actinobacteria</i> | <i>Microlunatus</i> | Mixed | Pure culture | Kawakoshi et al., 2012 |
| <i>Actinobacteria</i> | <i>Microlunatus</i> | Mixed | SBR | Nakamura et al., 1991; 1995a; 1995b; Kawaharasaki et al., 1998 |
| <i>Actinobacteria</i> | <i>Microlunatus</i> | Mixed | WWTP | Beer et al., 2006 |
| <i>Actinobacteria</i> | <i>Tetrasphaera</i> | Mixed | WWTP | Maszenan et al., 2000; Nguyen et al., 2011; Kong et al., 2005; Albertsen et al., 2012; Kristiansen et al., 2013 |
| <i>Actinobacteria</i> | <i>Tetrasphaera</i> | Mixed | SBR | Günther et al., 2009 |
| <i>Alphaproteobacteria</i> | -- | Mixed | WWTP | Zilles et al., 2002a |
| <i>Alphaproteobacteria</i> | -- | Mixed | SBR | Kawaharasaki et al., 1998 |
| <i>Alphaproteobacteria</i> | <i>Defluviicoccus*</i> | Acetate | SBR | Nobu et al., 2014 |
| <i>Bacilli</i> | <i>Bacillus</i> | Mixed | Rhizosphere soil | DebRoy et al., 2013 |
| <i>Betaproteobacteria</i> | -- | Mixed | WWTP | Zilles et al., 2002a |
| <i>Betaproteobacteria</i> | -- | Acetate | SBR | Stante et al., 1997 |
| <i>Betaproteobacteria</i> | -- | Mixed | SBR | Ge et al., 2015 |
| <i>Betaproteobacteria</i> | <i>Accumulibacter</i> | Mixed | WWTP | Beer et al., 2006; Albertsen et al., 2012; Nguyen et al., 2012; Mao et al., 2015 |
| <i>Betaproteobacteria</i> | <i>Accumulibacter</i> | Mixed | SBR | Liu et al., 2001; Günther et al., 2009; Mao et al., 2014; Skennerton et al., 2014 |
| <i>Betaproteobacteria</i> | <i>Accumulibacter</i> | Acetate | SBR | Hesselmann et al., 1999; Crocetti et al., 2000; McMahan et al., 2002; García Martin et al., 2006; Kim et al., 2010 |
| <i>Betaproteobacteria</i> | <i>Accumulibacter</i> | Propionate | SBR | García Martin et al., 2006 |
| <i>Betaproteobacteria</i> | <i>Accumulibacter</i> | Mixed | Estuary | Kunin et al., 2008; Peterson et al., 2008 |
| <i>Betaproteobacteria</i> | <i>Accumulibacter</i> | Mixed | Freshwater | Kunin et al., 2008; Peterson et al., 2008 |
| <i>Betaproteobacteria</i> | <i>Accumulibacter</i> | Mixed | Soil | Kunin et al., 2008; Valdivia, 2009; Archibald, 2010 |
| <i>Betaproteobacteria</i> | <i>Accumulibacter</i> | Mixed | Sediment | Kunin et al., 2008; Watson et al., 2019 |
| <i>Betaproteobacteria</i> | <i>Dechloromonas</i> | Mixed | WWTP | Zilles et al., 2002b; Kong et al., 2007 |
| <i>Betaproteobacteria</i> | <i>Propinoibacter</i> | Acetate | SBR | Crocetti et al., 2000 |
| <i>Betaproteobacteria</i> | <i>Rhodocyclus</i> | Mixed | WWTP | Zilles et al., 2002a; 2002b; Kong et al., 2004; 2005 |
| <i>Betaproteobacteria</i> | <i>Rhodocyclus</i> | Acetate | SBR | Hesselmann et al., 1999; Crocetti et al., 2000; McMahan et al., 2002 |

| | | | | |
|----------------------------|----------------------|---------|------------------------|---|
| <i>Cyanobacteria</i> | <i>Leptolyngbya</i> | -- | Marine, SBR | Zhang et al., 2015, Oyserman et al., 2017 |
| <i>Gammaproteobacteria</i> | -- | Mixed | SBR | Liu et al., 2001 |
| <i>Gammaproteobacteria</i> | <i>Acinetobacter</i> | Mixed | WWTP | Wagner et al., 1994; Streichan et al., 1990 |
| <i>Gammaproteobacteria</i> | <i>Halomonas</i> | Mixed | WWTP | Nguyen et al., 2012 |
| <i>Gammaproteobacteria</i> | <i>Pseudomonas</i> | Mixed | WWTP | Nguyen et al., 2012 |
| <i>Gammaproteobacteria</i> | <i>Pseudomonas</i> | Mixed | SBR | Günther et al., 2009 |
| <i>Gammaproteobacteria</i> | <i>Thiomargarita</i> | Mixed | Sediment pore water | Schulz and Schulz, 2005 |
| <i>Gammaproteobacteria</i> | <i>Beggiatoa</i> | Acetate | Pure culture | Brock and Schulz-Vogt, 2011 |
| <i>Gammaproteobacteria</i> | <i>Thiothrix</i> | Acetate | SBR | Rubio-Ricón et al., 2017 |
| <i>Gemmatimonadetes</i> | <i>Gemmatimonas</i> | Mixed | SBR | Zhang et al., 2003 |
| <i>Melainabacteria</i> | | Mixed | WWTP | Soo et al., 2014 |
| <i>Chlorophyta</i> | <i>Parachlorella</i> | -- | Pure culture | Ota et al., 2016 |

*This organism was described as a competitor to PAOs, a glycogen accumulating organism (GAO), but had genes coding for intracellular polyphosphate storage.

Table S5. Functional genes and associated proteins related to phosphorus (P) cycling. Abbreviations: polyphosphate (polyP).

| Gene | Protein (Abbreviation) | Function and Key Traits | References |
|---|--|---|---|
| <i>ppk1</i> | polyphosphate kinase 1 (PPK1) | Catalyzes the de-phosphorylation of ATP to make polyP. Has an affinity for Mg ²⁺ . | Ahn and Kornberg, 1990; Akiyama et al., 1992; Trelstad et al., 1999; Zhang et al., 2002; Zhu et al., 2005; Rao et al., 2009; Kawakoshi et al., 2012 |
| <i>ppk2</i> | polyphosphate kinase 2 (PPK2) | Catalyzes the de-phosphorylation of ATP and GTP to make polyP. Has an affinity for Mn ²⁺ . | Ishige et al., 1998; 2002; Zhang et al., 2002; Nocek et al., 2008; Rao et al. 2009; Kawakoshi et al. 2012 |
| <i>ppx1</i> | exopolyphosphatase (PPX1) | Liberates the terminal phosphate molecule in longer (>500) polyP chain. Requires Mg ²⁺ and KCl. Does not act on ATP and cannot be inhibited by ADP or ATP. | Reizer et al., 1992; Akiyama et al., 1993; Wurst and Kornberg, 1994; Kornberg, 1995; Zago et al., 1999; Ohtake et al., 2001; Rangarajan et al., 2006; Lichko et al., 2006; Rao et al., 2009; Kawakoshi et al., 2012 |
| <i>ppx2/gppA</i> | exopolyphosphatase/ pentaphosphate phosphohydrolase (PPX/GPPA) | Liberates the terminal phosphate in a polyP chain and important in cellular stress response by hydrolyzing pppGpp to ppGpp. | Keasling et al., 1993; Reizer et al., 1993; Rao et al., 2009; Kawakoshi et al., 2012; Alcántara et al., 2014 |
| <i>ppn</i> | endopolyphosphatase (PPN) | Cleaves phosphate from polyP chains (not selective to terminal phosphate). Has only been found in archaea, fungi, and mammals. | Kumble and Kornberg, 1996; Shi and Kornberg, 2005; Lichko et al., 2006; Rao et al., 2009 |
| <i>pap</i> | AMP phosphotransferase (PAP) | Catalyzes reaction of AMP to ADP using energy from breakdown of polyP. | Rao et al., 2009; Mao et al., 2014; Skennerton et al., 2014 |
| <i>phoX</i> , <i>phoD</i> , <i>phoA</i> | alkaline phosphatase (APase) | Under P limited conditions, catalyzed the breakdown of Po to Pi. | Temperton et al., 2011; von Sperber et al., 2014; Morrison et al., 2016; Ragot et al., 2015; 2016; George et al., 2018; Margenot et al., 2018 |
| <i>pit</i> | low-affinity phosphate transporter (PIT) | Binds phosphate and brings it into the cell. | Mao et al., 2014 |
| <i>pst</i> | high-affinity phosphate transporter (PST) | Binds phosphate and brings it into the cell. | Mao et al., 2014 |

Table S6. Descriptions, references, and examples of tools that may benefit the study of polyphosphate (polyP) and polyphosphate accumulating organisms (PAOs) in agricultural landscapes and beyond.

| Tool | Description/Purpose | Use Category | Use References | Example in Wastewater Treatment | Example in the Environment |
|------------------------------------|---|---------------------------------|---|--|--|
| <i>Microscopy</i> | | | | | |
| TEM | Transmission electron microscopy (TEM) is used with energy dispersive x-ray analysis to identify intracellular polyP granules. | metabolism | Sick-Goad and Lazinsky, 1986 | Günther et al., 2009 | Uhlmann and Bauer, 1988 |
| NanoSIMS | Nanometer-scale second ion mass spectrometry (NanoSIMS) is a microscopy-based spectroscopy that is used to identify intracellular polyP granules as well as the 2-D and 3-D spatial distribution of P and various cations such as Fe, Al, Ca, and Mg. | metabolism | Herrmann et al., 2007; Hoppe et al., 2013; Mueller et al., 2013; Kruse et al., 2015; Gao et al., 2016 | -- | Sulu-Gambari et al., 2016 |
| <i>Light Microscopy</i> | | | | | |
| Lead staining | Light microscopy technique used to identify intracellular polyP granules | metabolism | Stevenson and Stoermer, 1982 | -- | Stevenson and Stoermer, 1982 |
| Neisser's (methylene) blue | Light microscopy stain used to identify intracellular polyP granules. | metabolism | Crocetti et al., 2000 | Crocetti et al., 2000 | Schulz and Schultz, 2005 |
| Toluidine blue | Light microscopy stain used to identify intracellular polyP granules. | metabolism | Schulz and Schulz, 2005 | Streichan et al., 1990 | Schulz and Schultz, 2005 |
| <i>Epi-fluorescence Microscopy</i> | | | | | |
| Tetracycline | Fluorescent stain used to identify intracellular polyP granules. | metabolism | Günther et al., 2009 | Günther et al., 2009 | -- |
| DAPI | 4',6-diamidino-2-phenylindole (DAPI) is a fluorescent stain used to identify and quantify intracellular polyP granules. | metabolism | Hung et al., 2002; Aschar-Sobbi et al., 2008; Diaz and Ingall, 2010, Martin and Van Mooy, 2013 | Hung et al., 2002 | Rier et al., 2016 |
| FISH/qFISH | Fluorescence in-situ hybridization (FISH) is a fluorescent oligonucleotide probe that hybridizes to 16S rRNA or 23S rRNA sequences. Quantitative FISH (qFISH) is used to quantify PAOs and study the diversity of PAO communities. FISH is | microbial community, metabolism | Amann et al., 1990; 2001; Seviour et al., 2003; Seviour and Nielsen, 2010 | metabolism: Hesselman et al., 1999; diversity: Beer et al., 2006; quantification: Albertsen et al., 2012 | Castle and Kirchman, 2004; Sulu-Gambari et al., 2016 |

used in combination with epi-fluorescence microscopy or fluorescence spectroscopy. Can also be used with DAPI to co-locate polyP in PAO cells. See Table S8.

| <i>Molecular Biology</i> | | | | | |
|--------------------------|---|---|--|-------------------------|---|
| PCR/qPCR | Polymerase chain reaction (PCR) is used to amplify and identify PAO 16S rRNA and functional genes (e.g., <i>ppk</i>). Quantitative real-time PCR (qPCR) is used to amplify and quantify PAO 16S rRNA and functional genes. Both can be used to determine PAO strain diversity and PAO community functional gene diversity. See Table S7. | metabolism, functional genes, microbial community | He et al., 2010; Seviour and Nielsen, 2010 | Zhang et al., 2016 | Peterson et al., 2008 |
| Shotgun metagenomics | Non-targeted next generation sequencing technique used to identify potential PAOs, study the metabolic potential of potential PAOs, and quantify the relative abundance of PAO functional genes. | metabolism, functional genes, microbial community | Riesenfeld et al., 2004; Howe et al., 2014; Zimmerman et al., 2014; Howe and Chain, 2015; Wang et al., 2015; Choi et al., 2016, Menzel and Krogh, 2016 | Skennerton et al., 2014 | Temperton et al., 2011 |
| Amplicon metagenomics | Targeted (16S rRNA, 23S rRNA, internal transcribed spacer region) next generation sequencing technique used to quantify known PAOs and study PAO communities. | microbial community | Riesenfeld et al., 2004, Kozich et al., 2013, Zimmerman et al., 2014 | Oyserman et al., 2017 | Locke, 2015 |
| Flow cytometry | Cell sorting technique often used in conjunction with DAPI to identify potential PAOs. | metabolism, microbial community | Zilles et al., 2002a; 2002b; Hung et al., 2002 | Kim et al., 2010 | Locke, 2015 |
| <i>Other</i> | | | | | |
| ³¹ P-NMR | ³¹ P-nuclear magnetic resonance (NMR) spectroscopy is used to measure the concentration of various P-species (e.g., polyphosphate) in water and soil/sediments. | metabolism | Cade-Menun, 2015 | Peng et al., 2010 | Bourke et al., 2009; Read et al., 2014, McDowell et al., 2015; Cade-Menun, 2017 |
| Sensors | Used to measure environmental variables (e.g., P concentration, dissolved oxygen concentration) along a range of time scales. | metabolism | Pellerin et al., 2016; Rode et al., 2016; Fares et al., 2016 | Lanham et al., 2013 | Cohen et al., 2013 |

Table S7. A summary of polyphosphate accumulating organism (PAO) polymerase chain reaction (PCR) primer sequences.

| Gene Target | Primer Pair | Sequence (5'-3') | Reference |
|------------------------------|---------------------------------|--|--|
| Most <i>ppk1</i> | NLDE-0199F TGNV-1435R | CGTATGAATTTTCTTGGTATTTATTGTACTAATCTngaygarttyt GTCGAGCAGTTTTGCATGAwartncngt | McMahon et al., 2002; 2007 |
| CAP <i>ppk1</i> | ACCppk1-254F ACCppk1-1376R | TCACCACCGACGGCAAGAC ACGATCATCAGCATCTTGGC | McMahon et al., 2002; 2007; Kunin et al., 2008 |
| CAP <i>ppk1</i> | ppk274f ppk1156r | ACCGACGGCAAGACSG CGGTAGACGGTCATCTTGAT | Kunin et al., 2008 |
| CAP <i>ppk1</i> | ppk734f ppk1601r | CTCGGCTGCTACCAGTTCCG GATSCCGGCGACGACGTT | Kunin et al., 2008 |
| CAP Clade 1A <i>ppk1</i> | Acc-ppk1-763f Acc-ppk1-1170r | GACGAAGAAGCGGTCAAG AACGGTCATCTTGATGGC | He et al., 2007; He and McMahon, 2011b |
| CAP Clade 1A <i>ppk1</i> | Acc-ppk1-974f Acc-ppk1-1113r | TGATGCGCGACAATCTCAAATTCAA AATGATCGGATTGAAGCTCTGGTAG | Zhang et al., 2016 |
| CAP Clade 1B <i>ppk1</i> | Acc-ppk1-372f Acc-ppk1-653r | TGAAGGCATTCGCTTCCCT AAGCAGTATTCGCTGTC | Zhang et al., 2016 |
| CAP Clade 1C <i>ppk1</i> | Acc-ppk1-362f Acc-ppk1-758r | AGCTGGCGAGTGAAGCATTTCG AACAGTTGCTGTTGCGCGTGA | Zhang et al., 2016 |
| CAP Clade 1D <i>ppk1</i> | Acc-ppk1-634f Acc-ppk1-848r | TGCGACAGCGAATACAG ACTTCGAGGCGGACG | Zhang et al., 2016 |
| CAP Clade 2A <i>ppk1</i> | Acc-ppk1-893f Acc-ppk1-997r | AGTTCAATCTCACCGACAGC GGAACCTCAGGTCGTTGC | He et al., 2007; He and McMahon, 2011b |
| CAP Clade 2B <i>ppk1</i> | Acc-ppk1-870f Acc-ppk1-1002r | GATGACCCAGTTCCTGCTCG CGGCACGAACTTCAGATCG | He et al., 2007 |
| CAP Clade 2C <i>ppk1</i> | Acc-ppk1-254f Acc-ppk1-460r | TCACCACCGACGGCAAGAC CCGGCATGACTTCGCGGAAG | He et al., 2007 |
| CAP Clade 2D <i>ppk1</i> | Acc-ppk1-375f Acc-ppk1-522r | GGGTATCCGTTTCTCAAGCG GAGGCTCTTGTGAGTACACGC | He et al., 2007 |
| CAP Clade 2E <i>ppk1</i> | Acc-ppk1-757f Acc-ppk1-1129r | TTCGTGGACGAGGAAGA ATTGTTTCGAGCAACTCGATG | Zhang et al., 2016 |
| CAP Clade 2G <i>ppk1</i> | Acc-ppk1-410f Acc-ppk1-514r | CCGAGCAACGCGAATGG TGTTGAGTACGCGCGGA | Zhang et al., 2016 |
| CAP Clade 2H <i>ppk1</i> | Acc-ppk1-701f Acc-ppk1-928r | ACTCCTTCGTATTCTCTCT TCATCGCTTCGGAGCA | Zhang et al., 2016 |
| CAP Clade 2I <i>ppk1</i> | Acc-ppk1-688f Acc-ppk1-946r | AGTGATTATGCTTTCGTCTTTC TGAAGTGTCCGAGCAGGA | Zhang et al., 2016 |
| CAP 16S | CAP438f CAP846r | GGTTAATACCCTGWGTAGAT GTTAGCTACGGCACTAAAAGG | Zhang et al., 2016 |
| CAP 16S | PAO-518f PAO-846r | CCAGCAGCCGCGGTAAT GTTAGCTACGGCACTAAAAGG | He et al., 2007; He and McMahon, 2011b |
| CAP Clade 1A 16S | 16S-Acc-1Af 16S-Acc-1Ar | TTGCTTGGGTTAATACCCTGA CTGCCAAACTCCAGTCTTGC | He et al., 2010 |
| CAP Clade 2A 16S | 16S-Acc-2Af 16S-Acc-2Ar | TTGCACGGGTTAATACCCTGT CTCTGCCAAACTCCAGCCTG | He et al., 2010 |
| <i>Halomona</i> -related 16S | Pse136f 1492R | TAGTAGTGGGGGATAACGTC GCYTACCTTGT TACGACTT | Lane, 1991; Nguyen et al., 2012 |

Table S8. A summary of fluorescence *in-situ* hybridization (FISH) probes used to identify polyphosphate accumulating organism (PAO)-related organisms (updated from Seviour and Nielsen 2010). Abbreviations: not determined (ND).

| Probe Name | Sequence (5'-3') | Formamide (%) | Target | Reference |
|-------------------------|-------------------------|---------------|--------------------------------------|---|
| ALF1b | GCTGCCTCCCGTAGGAGT | 20 | Alphaproteobacteria | Manz et al., 1992 |
| BET42 ^a | GCCTTCCCACATTCGTTT | 35 | Betaproteobacteria | Manz et al., 1992 |
| GAM42 ^a | GCCTTCCCACATTCGTTT | 35 | Gammaaproteobacteria | Manz et al., 1992 |
| RHC175 | TGCTCACAGAATATGCGG | 30 | <i>Rhodocyclus</i> /Accumulibacter | Hesselmann et al., 1999 |
| RHC439 | CNATTCTTCCCGCCGA | 30 | Most <i>Rhodocyclus</i> | Hesselmann et al., 1999 |
| Re988 | AGGATTCTGACATGTCAAGGG | ND | <i>Rhodocyclus</i> group | Crocetti et al., 2000 |
| PAO462 ^c | CCGTATCTACWCAGGGTATTAAC | 35 | Most Accumulibacter | Crocetti et al., 2000 |
| PAO651 ^c | CCCTTGCCAAACTCCAG | 35 | Most Accumulibacter | Crocetti et al., 2000 |
| PAO846 ^c | GTTAGCTACGGCACTAAAAGG | 35 | Most Accumulibacter | Crocetti et al., 2000 |
| Acc-I-444 | CCCAAGCAATTTCTTCCCC | 35 | PAO clade IA and other Type I clades | Flowers et al., 2009 |
| Acc-II-444 | CCGTGCAATTTCTTCCCC | 35 | PAO clade IIA, IIC, and IID | Flowers et al., 2009 |
| Actino-1011 | TTGCGGGGCACCCATCTCT | 30 | <i>Tetrasphaera</i> -relatives | Liu et al., 2001 |
| Actino-221 ^a | CGCAGGTCCATCCCAGAC | 30 | <i>Tetrasphaera</i> -relatives | Kong et al., 2005 |
| Actino-658 ^a | TCCGGTCTCCCCTACCAT | 40 | <i>Tetrasphaera</i> -relatives | Kong et al., 2005 |
| Tet1-266 | CCCCTCGTCGCCTGTAGC | 25 | <i>Tetrasphaera</i> -relatives | Nguyen et al., 2011 |
| Tet2-892 | TAGTTAGCCTTGCGGCCG | 5 | <i>Tetrasphaera</i> -relatives | Nguyen et al., 2011 |
| Tet2-174 | GCTCCGTCTCGTATCCGG | 20 | <i>Tetrasphaera</i> -relatives | Nguyen et al., 2011 |
| Tet3-654 | GGTCTCCCCTACCATACT | 35 | <i>Tetrasphaera</i> -relatives | Nguyen et al., 2011 |
| Tet3-19 | CAGCGTTCGTCCTACACA | 0 | <i>Tetrasphaera</i> -relatives | Nguyen et al., 2011 |
| BET135 | ACGTTATCCCCCACTCAATGG | 45 | <i>Dechloromonas</i> -relatives | Kong et al., 2007 |
| MIC179 | GAGCAAGCTCTTCTGAAACCG | 10 | <i>Microlunatus phosphovor</i> | Kawaharasaki et al., 1998 |
| G123T | CCTCCGATCTCTATGCA | 40 | <i>Thiothrix</i> -relatives | Kanagawa et al., 2000; Rubio-Rincón et al., 2017 |
| EUB338 | GCTGCCTCCCGTAGGAGT | 60 | Most Bacteria ^b | Amann et al., 1990 |
| EUB338-II | GCAGCCACCCGTAGGTGT | 60 | Most Bacteria ^b | Daims et al., 1999 |
| EUB338-III | GCTGCCACCCGTAGGTGT | 60 | Most Bacteria ^b | Daims et al., 1999 |

^a Competitor probes required.

^b Use EUB338, EUB338-II, and EUB338-III together to obtain an estimate of total bacteria.

^c Use PAO462, PAO651, PAO846 together to obtain an estimate of total CAP PAOs.

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