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2   Advances, Challenges, and Opportunities

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14   **Conflict of Interest**

15   The authors have declared no conflicts of interest for this article.

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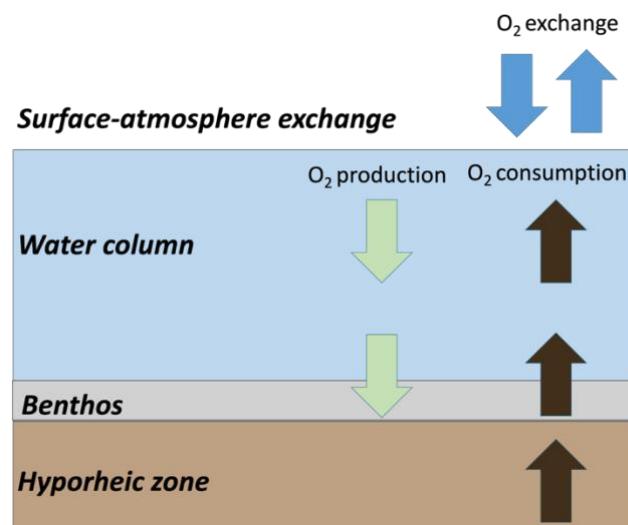
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20 **ABSTRACT**

21 Dissolved oxygen (DO) has been extensively studied in streams and rivers. Despite this breadth of  
22 research, the processes governing DO are rarely quantified concurrently with whole-ecosystem  
23 measurements. To address this gap, we synthesize 230 empirical studies (1964-2024) to evaluate  
24 how, where, and with what methods oxygen exchanges—the processes by which oxygen enters and  
25 leaves streams—have been measured. We organize oxygen exchanges into five groupings, treating  
26 whole-stream metabolism as an integrative category of four underlying process domains: (1) water  
27 column, (2) benthic, (3) the hyporheic zone and sediments, and (4) surface-atmosphere exchange.  
28 Across the literature, estimates of whole-stream metabolism and gas exchange dominate, reported in  
29 66% and 18% of studies, respectively, while underlying process domains were collectively reported in  
30 fewer than 16% of studies and jointly quantified in less than 7%. Where units permitted comparison,  
31 estimates of respiration in benthic, hyporheic, and underlying sediments were among the most  
32 variable exchanges yet least frequently measured. Together, these patterns indicate that our current  
33 understanding of stream oxygen dynamics relies heavily on integrative, reach-scale metabolism  
34 estimates, while the processes that generate these signals remain poorly resolved due to  
35 methodological and disciplinary silos. We identify key challenges and opportunities for synthesizing  
36 oxygen exchanges, emphasizing standardized reporting, concurrent measurements, and cross-site  
37 comparisons to advance integrated assessments of DO exchanges across streams and rivers.

38

39 **GRAPHICAL ABSTRACT AND CAPTION**



40

41 Dissolved oxygen in streams reflects the balance between physical exchange with the atmosphere  
42 (blue arrows) and biological processes across the water column, streambed (benthos), and hyporheic  
43 zone. Green arrows denote oxygen production, while brown arrows indicate oxygen consumption.  
44

45 **1. INTRODUCTION**

46 Dissolved oxygen (DO) is a master variable in aquatic ecosystems as it integrates physical, biological,  
47 and chemical processes occurring within a waterbody and the connected river corridor (Hutchinson,  
48 1938). DO serves as a constraint on aquatic life, an indicator of ecosystem processes and their  
49 dynamics (e.g., Davis, 1975), and regulates the cycling of both macro- and micronutrients (e.g.,  
50 Davison, 1993). In well-oxygenated environments, microbial respiration uses oxygen as a terminal  
51 electron acceptor, but as DO is depleted processes such as nitrate, iron, and sulfate reduction  
52 emerge (Morrice et al., 2000). These redox transitions influence the production of greenhouses gases  
53 (McClain et al., 2003; Beaulieu et al., 2011), mobilization of contaminants (Borch et al., 2010;  
54 O'Donnell et al., 2024), and determine the net release or retention of nutrients (Hupfer &  
55 Lewandowski, 2008). At the same time, DO constrains biological communities, where low oxygen (<5  
56 mg L<sup>-1</sup>) can impair growth and reproduction in fish and macroinvertebrates (Vaquer-Sunyer & Duarte,  
57 2008; Saari et al., 2018). Moreover, hypoxia (<2 mg L<sup>-1</sup>) has become increasingly documented in  
58 aquatic ecosystems globally driven by a combination of climate warming, nutrient enrichment, and  
59 hydrologic alteration (Jane et al., 2021; Blaszczak et al., 2023; Chan et al., 2025). While we have  
60 extensive knowledge about DO in lotic systems and its relation to other processes, our overarching  
61 hypothesis is that our present understanding of DO in space and time remains limited by  
62 methodological heterogeneity, scales of investigation, and divergence between fields and their  
63 respective norms.

64

65 Oxygen solubility in water is fundamentally controlled by temperature, pressure, and salinity, but the  
66 amount of DO and its dynamics in streams result from a complex interplay of physical (e.g.,  
67 reaeration), biological (e.g., primary production and autotrophic and heterotrophic respiration), and  
68 chemical processes (e.g., reduction of iron). In most lotic systems, however, physical (Zhi et al., 2023)  
69 and biological (Garvey et al., 2007) controls overwhelmingly dominate oxygen budgets, while  
70 chemical processes play a comparatively minor role. Streams themselves are hierarchically organized  
71 in both space and time (e.g., Frissell et al., 1986; Ward, 1989), and the dominant controls on DO shift

72 with watershed characteristics (e.g., watershed area, light availability), hydrologic attributes (e.g.,  
73 discharge, velocity, depth), surface water chemistry (e.g., nutrients, water chemistry), and subsurface  
74 conditions (e.g., residence time, soil properties, hyporheic exchange).

75

76 Early conceptual and modeling approaches sought to formalize these controls by linking them to the  
77 biological and physical components of the oxygen budget. The classic Streeter-Phelps model  
78 (Streeter & Phelps, 1925) explicitly coupled organic-matter decomposition via respiration with  
79 reaeration. Perhaps the most influential model of oxygen in lotic ecosystems, however, was  
80 developed by Odum (1956), who provided a framework to relate and quantify gross primary  
81 productivity (GPP), ecosystem respiration (ER), and gas exchange:

82

$$\frac{dO_2}{dt} = GPP - ER + k (C_s - C)$$

83 where  $\frac{dO_2}{dt}$  is the net areal oxygen flux ( $\text{g O}_2 \text{ m}^{-2} \text{ day}^{-1}$ ),  $k$  is the gas transfer velocity ( $\text{m day}^{-1}$ ),  $C_s$  is the  
84 dissolved oxygen concentration at saturation ( $\text{g O}_2 \text{ m}^{-3}$ ), and  $C$  is the in-stream dissolved oxygen  
85 concentration ( $\text{g O}_2 \text{ m}^{-3}$ ). Building on this early oxygen balance, subsequent models include analytical  
86 solutions for estimating diel DO fluctuations (O'Connor, 1967; Gulliver & Stefan, 1984; Chapra & Di  
87 Toro, 1991). Concurrently, improved sensor accessibility and reliability (Uehlinger & Naegeli, 1998;  
88 Zhang et al., 2024) enabled the application of complementary measurement and modeling  
89 approaches, ranging from the two-station method (Marzolf et al., 1994) to eddy covariance  
90 (Koopmans & Berg, 2015), and dissolved gas analysis via membrane inlet mass spectrometry  
91 (Vautier et al., 2020). The resulting growth in high resolution datasets have enabled new analytical  
92 techniques, including network synchrony (Diamond et al., 2023), entropy (Canadell et al., 2021), and  
93 frequency-domain approaches (Shin et al., 2024). Lacking, however, is the ability to disaggregate the  
94 metabolic processes in streams to better represent oxygen exchanges between ecosystem  
95 compartments based on the underlying processes.

96

97 Our review defines oxygen exchanges as processes through which oxygen enters or leaves different  
98 stream compartments. We organize oxygen exchanges into five distinct categories. First, whole-  
99 stream metabolism (GPP and ER) is an integrative measure of stream DO that emerges from the  
100 interaction of multiple underlying processes. Next, these underlying processes are grouped by where

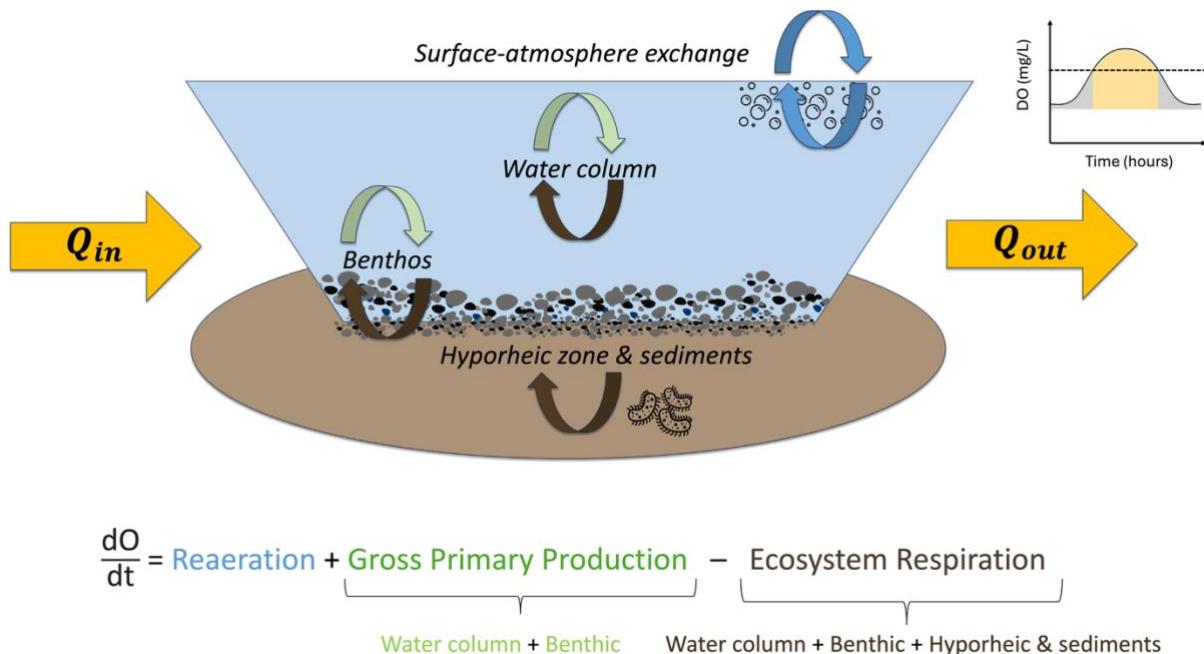
101 oxygen is consumed or produced within (1) the water column, (2) the benthos, (3) the hyporheic zone  
102 or sediments, and (4) surface-atmosphere exchange (Fig. 1). Together, these four process domain  
103 groups represent distinct yet coupled physical and biogeochemical processes that regulate the  
104 balance of oxygen, the net aggregation of which are manifest at the whole-stream scale. Different  
105 disciplines or research questions may emphasize different groups. For example, physical scientists  
106 studying channel hydraulics and gas exchange may focus on surface-atmosphere exchange using  
107 empirical tracers (O'Connor & Dobbins, 1958), sound spectral analysis (Klaus et al., 2019), and  
108 simulations rooted in fundamental physics (Zappa et al., 2007). Ecosystem ecologists may focus on  
109 whole-stream metabolism (Demars et al., 2015) or individual components and their ecological  
110 functions (Battin et al., 2016). Still others will emphasize dynamics along particular flowpaths such as  
111 those in the hyporheic zone (Zarnetske et al., 2011). Thus, a comprehensive understanding of DO  
112 and its dynamics requires bridging disciplines and linking exchanges across a host of spatial and  
113 temporal scales, as compartment dominance may vary between stream settings, and even shift  
114 seasonally, or under disturbance (Vadeboncoeur et al., 2003; Genzoli & Hall, 2016).

115

116 Our objective is to critically review how oxygen exchanges and their controls have been studied in  
117 streams and rivers. Here, we evaluate how studies quantify oxygen exchanges and how often these  
118 have been measured together to provide an integrated understanding of stream oxygen budgets. We  
119 also examine how studies of oxygen have evolved over time, identifying which exchanges and  
120 methods have been emphasized, and where persistent gaps remain. We use these patterns to  
121 answer the question: *To what extent does existing literature capture interconnected exchanges that*  
122 *comprise a stream oxygen budget, and how have these exchanges been studied?* Our results allow  
123 us to highlight biases and limitations in the literature, from which we outline four key challenges to  
124 achieving a holistic understanding of DO dynamics in streams, and propose a set of forward-looking  
125 goals for the research community.

126

127 **2. DEFINING OXYGEN EXCHANGES IN STREAMS AND RIVERS**



128

129 **Figure 1. Conceptual model of oxygen exchanges into and out of a stream channel.** The  
 130 schematic illustrates the major exchanges governing dissolved oxygen (DO) dynamics in streams  
 131 organized into water column and benthic production and respiration, hyporheic and sediment  
 132 respiration, and surface-atmosphere exchange. Arrows depict oxygen production (downward) and  
 133 consumption (upward), while inflow ( $Q_{in}$ ) and outflow ( $Q_{out}$ ) indicate advective transport through the  
 134 reach. The inset shows a characteristic diel DO signal resulting from the combination of physical and  
 135 biological exchanges, with daytime (yellow), nighttime (gray), and a dashed line indicating oxygen  
 136 saturation. The oxygen mass balance equation shown below summarizes whole-stream oxygen  
 137 dynamics ( $dO/dt$ ) as the integrated outcome of these interacting fluxes.  
 138

139 *2.1 Whole-stream metabolism*

140 Gross primary production (GPP) is inferred from the total oxygen production via photosynthesis, while  
 141 ecosystem respiration (ER) reflects the total oxygen consumption by autotrophs and heterotrophs;  
 142 their difference (i.e., GPP - ER) defines net ecosystem production (NEP), capturing the balance  
 143 between autotrophic energy fixation via photosynthesis and aerobic respiratory demand (Odum,  
 144 1956). Although typically evaluated for carbon cycling since DO is easier to measure than  $CO_2$ , GPP  
 145 and ER are fundamentally oxygen exchanges. The primary drivers of GPP are light—particularly  
 146 biologically available light within the stream channel—and flow variation, whereas ER is more tightly  
 147 coupled to GPP and flow conditions (Kirk et al., 2021; Bernhardt et al., 2022). Both GPP and ER vary  
 148 systematically in response to watershed land uses and disturbance regimes (Bernot et al. 2010;  
 149 Finlay, 2011). For example, in forested headwaters, the combination of shading, low nutrients, and  
 150 high allochthonous organic matter supply allows ER to exceed GPP, whereas downstream reaches  
 151 with greater light availability or elevated nutrients from agriculture or urbanization can increase GPP

152 and/or ER and underpin longitudinal patterns in stream ecosystems (e.g., Vannote et al., 1980, Carter  
153 et al., 2024). The balance of GPP and ER is increasingly applied as an indicator of land use change  
154 and as a tool in stream restoration and management (Young et al., 2008; Arroita et al., 2019;  
155 Diamond et al., 2021; Jankowski et al., 2021).

156

157 Early diel DO measurements to assess whole-stream GPP and ER were conducted manually using  
158 Winkler titrations, a labor-intensive approach that limited temporal resolution and duration. High-  
159 frequency *in situ* sensors have revolutionized the estimation of whole-stream GPP and ER by allowing  
160 continuous monitoring and reach-scale integration (Marzolf et al., 1994). The most widely used  
161 approach for quantifying GPP and ER is the diel oxygen technique, or whole-stream metabolism,  
162 where single- and two-station models solve oxygen mass balances to estimate GPP, ER, and gas  
163 exchange simultaneously (Riley & Dodds, 2013; Grace et al., 2015). Recent advances in modeling  
164 tools, such as streamMetabolizer, have improved parameter estimation by allowing gas exchange to  
165 vary with discharge, examining DO residuals, reducing equifinality, and enabling meta-analysis of field  
166 studies (Holtgrieve et al., 2010; Appling et al., 2018a; Appling et al., 2018b).

167

168 Spatial patterns in river network metabolism are well-established (Vannote et al., 1980; Mulholland et  
169 al., 2001; Mejia et al., 2019). Most whole-stream metabolism models and empirical estimates operate  
170 at the reach-scale, where accuracy improves with longer time series and well-constrained gas  
171 exchange (Appling, et al., 2018a). However, scaling reach-scale metabolism estimates to network or  
172 basin scales remains challenging due to equifinality among parameters, site-specific characteristics,  
173 and event-driven disturbance that disrupt diel patterns including sediment pulses, storms, and flow  
174 intermittence (Hotchkiss & Hall, 2014). In response, a growing body of literature is advancing network-  
175 scale modeling frameworks (e.g., Segatto et al., 2023) and synthesizing metabolic patterns that  
176 emerge across regions 28/01/2026 22:34:00. Emerging work further integrates stream metabolism  
177 with process-based approaches such as hydraulic routing (Payn et al., 2017), partitions autotrophic  
178 and heterotrophic production (Hall & Beaulieu, 2013), and links metabolic rates to CO<sub>2</sub> cycling and  
179 emissions (Hotchkiss et al., 2015). Despite these advances, challenges remain when interpreting DO  
180 signals (e.g., Gomez-Velez et al. 2026). Most metabolism models assume observed DO dynamics  
181 reflect in-stream biological processes and surface-atmosphere exchange. In reality, streams lose and

182 gain oxygen from multiple hydrologic processes including hyporheic exchange, groundwater inputs or  
183 losses, and overland flow, each likely with potentially different signatures than stream water.  
184 Groundwater inputs are known to bias metabolism estimates by altering DO concentrations without  
185 corresponding in-stream biological production or consumption (Hall & Tank, 2005).

186

187 While whole-stream estimates of ER and GPP have advanced our understanding of DO (and carbon)  
188 dynamics at the scale of river segments and networks, improved reach-scale understanding requires  
189 quantifying the compartment-specific processes that contribute to integrated whole-stream ER and  
190 GPP observations (sections 2.2 and 2.3). Resolving these compartment-specific exchanges is  
191 essential because oxygen dynamics arise from the superposition of water column, benthos, hyporheic  
192 zone, and groundwater processes, the combination of which can create signals that obscure clear  
193 inferences of whole-stream metabolism (Fig. 1).

194

#### 195 *2.2 Water column production & respiration*

196 Water column metabolism refers to the combination of photosynthesis and respiration occurring within  
197 the water column of the stream, often by phytoplankton and microbes on suspended particles and  
198 organic matter. Although conceptually distinct from benthic metabolism (section 2.4 below), the two  
199 are tightly coupled as they continuously exchange material and energy (Fisher et al., 1998; Battin et  
200 al., 2008). Water column contributions to whole-stream metabolism broadly increase with channel  
201 size, where greater depth, residence time, and light availability favor planktonic communities (Vannote  
202 et al., 1980; Yang et al., 2021). While water column metabolism tends to be minimal in most low-order  
203 streams, often an order of magnitude lower than whole-stream estimates (Fisher & Likens, 1973; Bott  
204 et al., 2006; Plont et al., 2022), it can represent a substantial fraction of ecosystem metabolism in mid-  
205 sized, nutrient enriched rivers (Vink et al., 2005; Oliver & Merrick, 2006). For example, Reisinger et al.  
206 (2016) showed that water column contributions to GPP exceeded those to ER in mid-order rivers, with  
207 their importance largely dependent on organic matter and nutrients.

208

209 Early studies of stream metabolism relied on chamber and incubation techniques, such as light-dark  
210 bottle incubations and *in situ* chambers, which isolated water column and benthic processes without  
211 the confounding effect of gas exchange (Hargrave, 1972; Bott et al., 1978). While these techniques

212 remain useful, they are labor-intensive and capture only a snapshot in time. Recent advances such as  
213 autonomous *in situ* devices that automate light-dark incubations with optode-based oxygen sensors  
214 have been applied in marine environments (Collins et al., 2018). These instruments are costly and are  
215 not designed for deployment in smaller, more dynamic headwaters, but could be deployed in larger  
216 rivers. Remote sensing of phytoplankton via pigment optical effects provides another pathway; while  
217 these approaches work well in lakes, they often fail in rivers due to optical complexity (Spyrakos et al.,  
218 2018). Beyond sensor technology, integrative approaches that combine metabolic measurements with  
219 biogeochemical assays (e.g., nutrient uptake, denitrification) have been used to link water column  
220 metabolism to broader biogeochemical cycling, particularly across land use gradients (Reisinger et  
221 al., 2021). Long-term monitoring efforts combined with process-based approaches can also track  
222 temporal shifts in water column contributions (Van De Bogert et al., 2007), but scaling across sites,  
223 seasons, or networks is rare, constraining our ability to generalize rates and controls on water column  
224 contributions to whole-ecosystem fluxes.

225

### 226 2.3 *Benthic production & respiration*

227 Benthic metabolism, the photosynthesis and respiration occurring on the streambed, is driven by  
228 biofilms, periphyton, macrophytes, and sediment microbial communities. Benthic production is  
229 regulated by light availability and substrate properties including surface roughness, stability, and  
230 disturbance history (scour, burial, or grazing), with algal biomass accrual and disturbance regimes  
231 further modulating autotrophic production, particularly in nutrient-limited headwaters (Francoeur et al.,  
232 1999; Atkinson et al., 2008). By contrast, benthic respiration is largely driven by organic matter  
233 supply, temperature, and oxygen delivery to sediments (Acuña et al., 2004). In most streams and  
234 rivers, benthic processes dominate relative to the water column and may be amplified by changes in  
235 food web dynamics and external energy sources. For instance, Rutherford (1991) found that benthic  
236 metabolism far outweighed water column contributions in a stream receiving paper mill effluent, while  
237 Hall et al. (2003) demonstrated how the invasive snail *Potamopyrgus antipodarum* consumed most of  
238 the benthic autotrophic biomass produced by GPP and drove nutrient cycling in a highly productive  
239 stream.

240

241 Benthic metabolism is typically measured using chambers that enclose patches of the streambed  
242 (Mulholland et al., 2001; Rüegg et al., 2015; Lopez et al., 2025). Partitioning contributions of different  
243 benthic materials has been a research focus with studies targeting macrophyte production (Fisher &  
244 Carpenter, 1976; Hensley & Cohen, 2020), algal photosynthesis (Mulholland et al., 1994; Acuña et al.,  
245 2011), and microbial respiration in sediments (Hill et al., 2000). Recently, aquatic eddy covariance—a  
246 technique widely applied in wetlands and marine systems—has been applied to measure *in situ*  
247 benthic oxygen fluxes in streams (Rovelli et al., 2017), though measurements integrate turbulent  
248 exchange near the bed and may include water column contributions. Scaling benthic metabolism in  
249 space is difficult due to sediment heterogeneity (Lee-Cullin et al., 2018), with fluxes varying up to two  
250 orders of magnitude depending on substrate type, catchment land use, and disturbance regime  
251 (Fellows et al., 2006), and organic matter quantity (Hill et al., 2017). This variation implies that efforts  
252 to quantify these fluxes, especially when estimates are conjoined with open-channel measurements,  
253 will improve our understanding of how whole-system fluxes are partitioned.

254

#### 255 *2.4 Hyporheic respiration and sediment oxygen demand (SOD)*

256 The hyporheic zone is the interface within stream sediments where surface water exchanges with  
257 shallow groundwater, encompassing flowpaths where parcels of water temporarily leave the stream  
258 channel and return after undergoing biogeochemical transformations within the sediment (White,  
259 1993; Krause et al., 2011). Within this zone, hyporheic respiration—the consumption of oxygen along  
260 subsurface flowpaths—can represent a major sink of DO, driven primarily by microbial and faunal  
261 respiration (Triska et al., 1989; Findlay, 1995; Boulton et al., 2010). The magnitude and variability of  
262 hyporheic respiration are shaped by a combination of hydrologic, biogeochemical, and geomorphic  
263 factors. Residence time, hyporheic exchange fluxes, and the reaction rates ultimately control solute  
264 transport and transformation within the hyporheic zone, ultimately determining the efficiency with  
265 which the hyporheic zone processes oxygen and other constituents (Gomez et al., 2012). Although  
266 the hyporheic zone plays a central role in stream oxygen dynamics, much of the literature has  
267 focused on carbon and nitrogen transformations, reflecting its function as a biogeochemical reactor  
268 (Findlay et al., 1993; Turečkaia et al., 2023).

269

270 Hyporheic respiration has been quantified by vertical oxygen profiles, pore water sampling, and *in situ*  
271 or laboratory incubations (Revsbech et al., 1986; Uzarski et al., 2001; Nelson et al., 2019).  
272 Experimental flumes have been used under controlled flow and sediment conditions to isolate  
273 hyporheic exchange and determine how reaction rates and biogeochemical processes vary across  
274 redox conditions (Reeder et al., 2018), while streamside mesocosms packed with bed sediment have  
275 been used to replicate *in situ* flow paths to the maximum extent practicable (Serchan et al., 2024;  
276 Herzog et al., 2025). High-resolution oxygen sensors have been deployed directly into hyporheic  
277 sediment to measure oxygen along hyporheic flowpaths through time (Malcolm et al., 2006; Gooseff  
278 et al., 2023), complemented by dissolved gas time series that reveal hyporheic exchange and redox  
279 transitions (Mächler et al., 2013).

280

281 Sediment oxygen demand (SOD) represents the total biological and chemical consumption of oxygen  
282 in the subsurface and is conceptually complementary to hyporheic respiration. While hyporheic  
283 research is largely focused on understanding subsurface ecosystem processes (Ward, 2016), SOD  
284 emerged as a stream-centric concept to explain oxygen deficits not accounted for by in-stream  
285 processes alone, particularly in the context of predicting oxygen sag and organic loading along rivers  
286 (Thomann 1972; Lee et al., 1978). SOD has been widely used to assess the biogeochemical  
287 consequences of disturbance such as urbanization, logging, and eutrophication that elevate organic  
288 matter inputs, alter stream temperature and flow regimes, and change microbial assemblages (Uttley  
289 et al., 2008). SOD is typically measured using mesocosms or chambers to incubate buried streambed  
290 sediment and quantify oxygen consumption (Bowman & Delfino, 1980), while recent reviews have  
291 called for standardized *in situ* chamber methods to improve transferability among systems (Coenen et  
292 al., 2019). Although rarely measured simultaneously, hyporheic respiration and SOD both quantify net  
293 oxygen consumption in the stream-connected subsurface (e.g., the hyporheic zone), yet *ex situ*  
294 incubations may not capture the full complexity of *in situ* transport and exchange processes. Because  
295 these subsurface fluxes are central to both the oxygen budget of a stream, and to the genesis of  
296 open-channel DO signals (Gomez-Velez et al., 2026), renewed attention to quantifying them as part  
297 of whole-stream measurements is likely to be critical to improved understanding of metabolic variation  
298 in time and space.

299

300 2.5 Surface-atmosphere exchange

301 Reaeration is the flux of oxygen across the stream-atmosphere interface and is a positive flux into the  
302 stream if the water is undersaturated relative to the atmosphere (negative if the water is  
303 supersaturated). In streams, gas exchange reflects the interaction between flow and channel form.  
304 The efficiency of exchange is controlled by turbulent mixing at the air-water interface, which depends  
305 on channel hydraulics such as velocity, depth, and slope (Melching & Flores, 1999; Ulseth et al.,  
306 2019). The rate of gas exchange depends on both chemical and physical controls, including water  
307 temperature (Wanninkhof, 1992; Hall & Ulseth, 2020), gas solubility (Garcia & Gordon, 1992), and the  
308 strength of the concentration gradient between the water and atmosphere (Raymond & Cole, 2001;  
309 Zappa et al., 2007).

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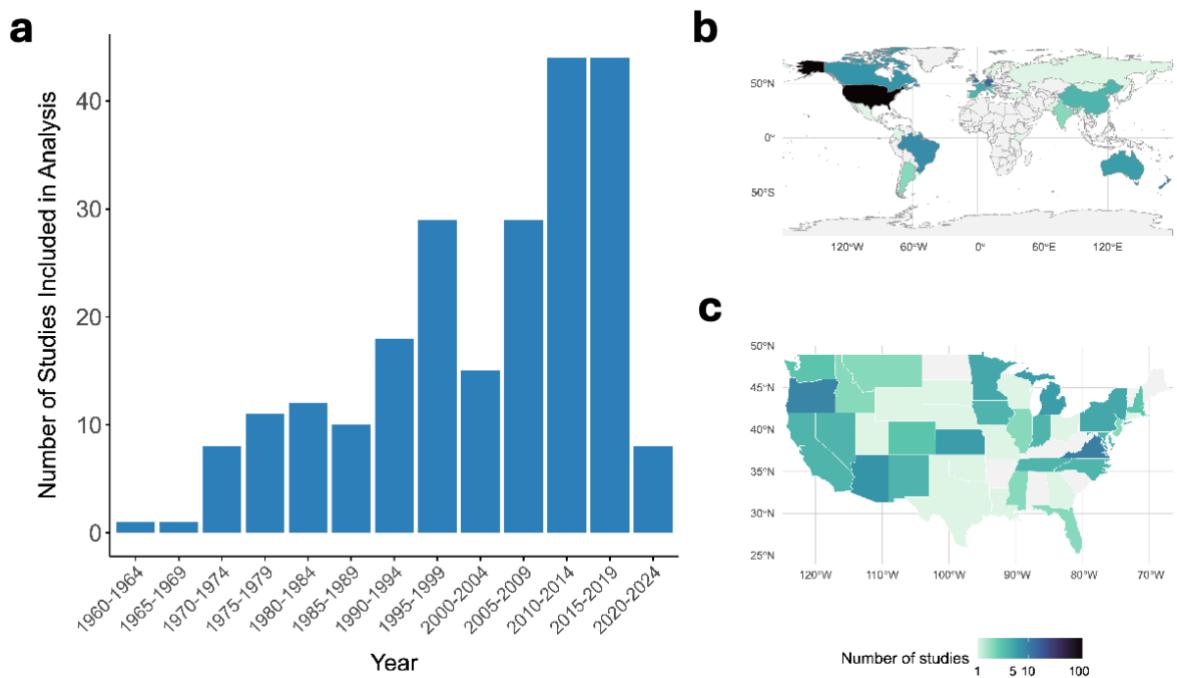
311 The study of gas exchange in rivers has advanced along three primary methodological pathways.  
312 First, empirical tracer injections using a specific gas tracer (e.g., propane, SF<sub>6</sub>, Ar) are used to  
313 measure degassing and inform empirical (e.g., Schmidt-number) relationships for estimating oxygen  
314 exchange (Tsivoglou & Wallace, 1972; Rathbun et al., 1978; Genereux & Hemond, 1992). However,  
315 scaling gas transfer across gases can introduce error under bubble-mediated conditions because of  
316 differences in gas solubility and diffusivity (Klaus et al., 2022). To reduce this uncertainty, argon has  
317 emerged as an ideal tracer due to its similarity in solubility and diffusivity to oxygen, minimizing  
318 scaling uncertainty (Hall & Madinger, 2018). Floating dome chambers offer a direct, non-invasive, and  
319 easily scalable method for measuring gas exchange, avoiding the logistical challenges of gas tracers  
320 but also provide smaller-scale measurements prone to bias when deployments alter stream  
321 turbulence (Lorke et al., 2015). Second, analyses of DO timeseries are used to infer gas exchange.  
322 The most widely used techniques include night-time regression, which isolates gas exchange from  
323 nighttime declines in DO when photosynthesis ceases and requires sufficient daytime oxygen  
324 production to generate a resolvable nighttime slope (Hornberger & Kelly, 1975); the solar noon lag,  
325 which examines the temporal offset between peak light and peak DO (McBride, 2002); and the delta  
326 method, which quantifies fluxes by partitioning observed changes in DO storage during both day and  
327 night (Chapra & Di Toro, 1991). Each of these approaches relies on detectable diel signals generated  
328 by the dynamics of photosynthesis and performs best in medium to large rivers where turbulence is  
329 lower and oxygen dynamics are less spatially heterogenous (Bailey & Ahmadi, 2014). Finally, based

330 on tracer injection and other field studies, empirical relationships to predict gas exchange as a  
331 function of river hydraulics have been derived, allowing for gas exchange estimates without direct  
332 measurements of DO. Early models related gas exchange rates to velocity and depth (O'Connor &  
333 Dobbins, 1958), while later work incorporated turbulence metrics such as the energy dissipation  
334 method (Tsivoglou & Neal, 1976; Chu & Jirka, 2003) and used prior tracer experiments to develop  
335 physical models of gas exchange for small streams (Raymond et al., 2012; Aho et al., 2025). More  
336 recently, these models have been applied in scaling gas exchange across river networks and are  
337 widely applied in regional and global syntheses (Raymond et al., 2013; Hotchkiss et al., 2015).

338

339 The predictability of gas exchange is scale- and system-dependent regardless of which approach is  
340 used. Despite their utility, hydraulic equations are fundamentally reach-averaged estimates and  
341 perform poorly in morphologically complex headwaters, where turbulence and bubble entrainments  
342 vary over short spatial scales, and are usually generated from log-transformed data, leading to high  
343 predictive uncertainty (Rovelli et al., 2018; Vautier et al., 2020). Consequently, site-specific calibration  
344 or empirical measurements remain essential for accurate flux estimates in many streams and rivers  
345 (Hall & Ulseth, 2020). Methodological advancements including acoustic Doppler velocimetry (Gålfalk  
346 et al., 2013), stable isotopes (Tobias et al., 2007; Jamieson et al., 2013), and sensor-equipped  
347 floating chambers (Vachon et al., 2010; Arce et al., 2021) have increased our ability to resolve gas  
348 fluxes at higher spatial and temporal resolution, yet scaling these approaches across heterogeneous  
349 systems remains a significant challenge (Botter et al., 2022). We note that the goal of empirically  
350 constraining gas exchange to ensure robust metabolism estimates has been widely recognized, with  
351 conjoined measurements of  $k_{600}$  and open-channel DO time series now serving as method best  
352 practice (Appling et al., 2018a). This recognition highlights our central premise that attention to other  
353 compartment-specific fluxes will likely enhance the validity of metabolism measurements.

354

355 **3. LITERATURE ANALYSIS: METHODS**

356

357 **Figure 2. Temporal and geographic distribution of studies included in the synthesis.** (a)  
 358 Number of studies by five-year intervals. (b) Global distribution of study locations, with darker shading  
 359 representing regions with higher numbers of included studies. (c) Spatial density of studies across the  
 360 continental United States by state (log-scale color bar).

361

362 To assess the extent to which oxygen exchanges and methods have been applied, we conducted a  
 363 search on Web of Science on August 26th, 2024, with the key words ["stream" OR "river" OR  
 364 "aquatic"] AND ["dissolved oxygen" OR "O<sub>2</sub>"] and listed the processes of DO independently (see  
 365 Table S1 of search strings), yielding a total of 4,086 studies. In Phase 1 of our extraction, we  
 366 excluded synthesis papers (leaving n = 3,929) and filtered by relevant field of study to focus on  
 367 streams and rivers (n = 2,832 after excluding studies of lakes, wetlands, marshes, and saltwater  
 368 systems). The top fields of study represented in the body of literature were environmental sciences  
 369 (1,487), water resources (622), and environmental engineering (578).

370

371 In Phase 2, a total of 2,832 articles were then exported and processed in Covidence, a systematic  
 372 review management tool (Covidence, 2024). We screened each study by reviewing the title and  
 373 abstract, retaining papers that met two criteria: (1) papers retained must include the presence of  
 374 empirical data, excluding literature reviews, remaining syntheses, and modeling-only studies; and (2)

375 papers retained must report an oxygen exchange as a rate or flux. We define oxygen exchanges  
376 broadly to encompass all quantified transfers of oxygen into or out of stream compartments,  
377 regardless of how rates and fluxes are reported. For consistency, we distinguish a flux as the mass  
378 per unit area per unit time and a rate as the mass per unit time (after Stauffer, 2006). A total of 914  
379 studies were retained based on our review of the title and abstract. Next, we reviewed the full text of  
380 each study to screen for explicit reporting of a DO rate or flux, finding 321 papers that met our criteria.  
381 Studies reporting only carbon metabolism in non-O<sub>2</sub> units (e.g., CO<sub>2</sub> production as a measure of  
382 respiration) were excluded, as were studies that reported a flux but did not have a standardized unit  
383 such as compartment-specific biofilm studies (e.g., Glud et al., 1992 on cyanobacteria vs. diatoms) or  
384 with results standardized to a unit that would not be comparable to other literature (e.g., ash-free dry  
385 weight; McIntire, 1966). We excluded 115 studies on this basis, leaving 204 papers. In review of these  
386 204 papers, only 8 included explicit observations of hyporheic processes. Consequently, we relaxed  
387 the requirement for hyporheic-centric studies to retain those that included subsurface observations of  
388 dissolved oxygen concentration (retaining them even if they did not specifically report a flux or rate),  
389 and those for which empirical work in a mesocosm or column was conducted as a proxy for hyporheic  
390 conditions. This yielded an additional 26 papers that described hyporheic concentrations of dissolved  
391 oxygen. Thus, our searches yielded a total of 230 papers from which data were extracted (Fig. 2a).

392  
393 For each of the 230 papers, we first recorded background information on the study period and  
394 geographic location (Fig. 2b). Over half of studies were conducted in the U.S. (51%), with  
395 substantially fewer in Germany (7%) and the United Kingdom (6%; Fig. 2c). Within the U.S., Virginia  
396 had the most studies (9%), followed by Oregon (8%), and Alaska and Arizona (6% each). There were  
397 few studies in some continents (e.g., n=1 for Africa, n=0 Antarctica) and large countries based on land  
398 area (e.g., Russia). Studies were predominantly located in mid-latitude regions (90%), with fewer in  
399 tropical (6%) and high-latitude (4%) regions.

400  
401 Next, we extracted information on stream size, land use, study objectives, watershed attributes, and  
402 environmental conditions. Stream size was classified using watershed area when provided: small  
403 (<100 km<sup>2</sup>), mid-sized (100–10,000 km<sup>2</sup>), and large (>10,000 km<sup>2</sup>) (after Lyche Solheim et al., 2019).  
404 When watershed area was not available, we used stream size binned as small (1-3), mid (4-6), and

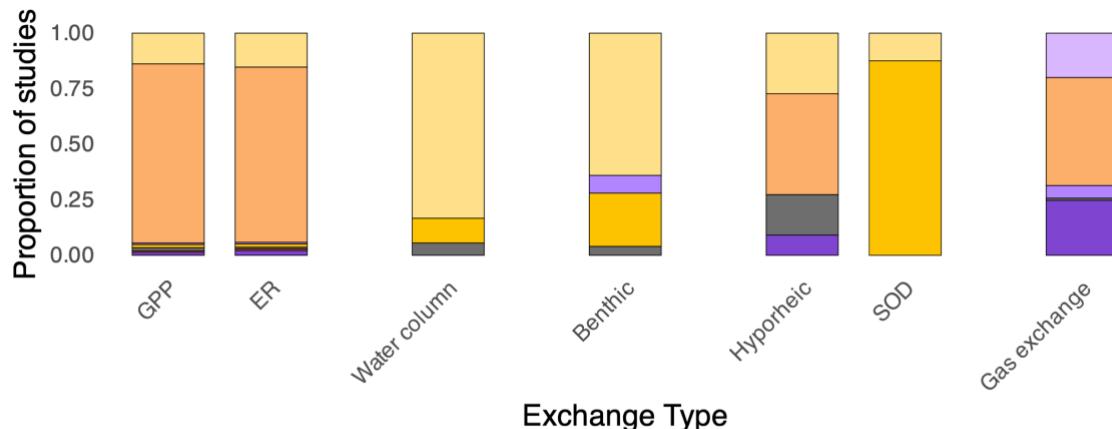
405 large (>7) stream orders (after Vannote et al., 1980). If neither metric was provided, we relied on  
406 study maps, qualitative descriptors, or conducted additional research on the study location to estimate  
407 stream size (n=23 studies). Land use was categorized based on qualitative descriptions in the site or  
408 methods sections using indicator terms: “pasture” or “cropland” for agricultural, “dams” or “impervious  
409 surfaces” for urban, and “mixed” for sites with multiple influences. Sites described as “pristine” or  
410 “forested” were classified as undisturbed. When land use percentages were reported, thresholds of  
411 >10% urban and >30% agricultural were used for classification, as these levels have been associated  
412 with negative impacts to stream condition (Allan, 2004). Without reported land use, additional  
413 research was conducted to characterize that study reach (n=16 studies) or left blank when not readily  
414 available from additional research (n=21 studies). Study objectives were categorized as: (1)  
415 measurements at the same site or location under varying conditions (e.g., changes between flows or  
416 seasons), (2) replicates in different streams, or (3) replicates along a river network. We also  
417 documented watershed attributes (canopy cover, elevation, watershed area), hydrologic metrics  
418 (gradient, stream size, discharge, and velocity), flow regime (e.g., perennial vs. non-perennial),  
419 surface water chemistry and environmental conditions (e.g., turbidity, temperature, nutrient  
420 concentrations, light, standing stock) and subsurface characteristics (e.g., hydraulic conductivity,  
421 transit time). Nutrient measurements included carbon (TOC, POM, POC, DOC, DIC), nitrogen (NO<sub>3</sub><sup>-</sup>,  
422 NH<sub>4</sub><sup>+</sup>, TN, DON, N<sub>2</sub>), and phosphorus (SRP/PO<sub>4</sub><sup>3-</sup>, TP, TDP). Light was characterized using  
423 illuminance (lux), photosynthetically active radiation (PAR), solar radiation, and light-attenuation  
424 coefficients. Whenever possible, mean values were taken directly from tables; when unavailable, we  
425 extracted ranges from tables or figures.

426  
427 Because each oxygen exchange is measured and reported using different conventions (e.g.,  
428 volumetric rates, areal fluxes, mass transfer coefficients), we systematically documented methods  
429 used, numerical values and associated units, and standard deviations when reported. We recorded  
430 values reported as rates (e.g., day<sup>-1</sup> or m day<sup>-1</sup> for gas exchange, and g O<sub>2</sub> L<sup>-1</sup> day<sup>-1</sup> for volumetric  
431 metabolic rates) and metabolic fluxes (g O<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup>). For the comparison of exchange magnitudes  
432 in section 4.3, we analyzed only exchanges standardized to areal fluxes, which is the norm for  
433 reporting whole-stream metabolism. This requirement omitted 25 studies that did not report an areal

434 flux. Gas exchange was excluded from this conversion because it is conventionally reported as an  
435 exchange rate (time<sup>-1</sup>) or gas transfer velocity (length time<sup>-1</sup>) that is normalized by depth, rather than a  
436 flux. For gas exchange, a host of conventions were used to report raw values (see Fig. S3). Here, we  
437 analyzed reported gas exchange as  $k_{600}$ , where the  $k$  value is normalized to a temperature-dependent  
438 Schmidt number (including n=64 studies for which sufficient data were presented for us to compute  
439  $k_{600}$  if it was not reported by the authors).

440

441 Finally, to evaluate how oxygen exchanges have been measured across studies, we classified  
442 methods into eight categories that capture the dominant approaches (Box 1; Fig. 3). Tracer methods  
443 include conservative gas additions (e.g., Ar, SF<sub>6</sub>, N<sub>2</sub>, propane) as well as isotope tracers (e.g., <sup>16</sup>O  
444 and <sup>18</sup>O), which directly quantify gas exchange and transport processes. In-stream DO measurements  
445 were used to estimate multiple oxygen exchanges, including discrete *in situ* measurements (e.g.,  
446 point measurements in wells to quantify hyporheic respiration) and continuous DO observations to  
447 estimate GPP, ER, and gas exchange from diel DO dynamics (see Fig. S3). Empirical equations  
448 represent velocity-depth or shear stress formulations that predict gas exchange coefficients from  
449 hydraulic variables. Interface flux methods incorporate floating probes, eddy covariance, sound  
450 velocity, and impermeable surface techniques that directly measure fluxes or turbulence. Standing  
451 stock quantifies accumulated biomass or chlorophyll-a as a proxy for metabolic activity. Chamber  
452 approaches are divided into field chambers (e.g., *in situ* benthic, mesocosms, light-dark bottles) and  
453 lab chambers collected from *in situ* samples (sediment or mesocosm incubations). Lastly, other  
454 modeling approaches include coupled biogeochemical (e.g., OTIS, PFLOTRAN) and water quality  
455 (e.g., QUAL2K, WASP) models, although these were required to have an empirical component to be  
456 retained in our database (i.e., modeling-only studies were excluded).

**BOX 1.** Oxygen fluxes and their typical methods and scales

**Figure 3. Methods used to quantify oxygen fluxes in streams.** Stacked bars show the proportional contribution of different method categories used to estimate each oxygen exchange, while colors denote method categories. See Fig. 5 for temporal breakdown of methods.

Oxygen dynamics in streams emerge from multiple fluxes operating across distinct compartments and spatial scales, each quantified using different measurement approaches (Fig. 3). As summarized below, these fluxes integrate processes over spatial domains, spanning centimeter-scale sediment measurements to reach-scale estimates of metabolism and gas exchange, highlighting the ongoing challenge of integrating oxygen fluxes across compartments.

**Sediment oxygen demand:** Total oxygen consumption by streambed sediments due to biological respiration and abiotic oxidation of reduced compounds

- *Typical methods:* Quantified using *ex situ* sediment cores or *in situ* chambers
- *Typical spatial scale:* Patch scale (centimeter to meter)

**Benthic metabolism:** Photosynthesis and respiration occurring on the streambed, often by biofilms, periphyton, macrophytes, and sediment microbial communities

- *Typical methods:* Measured using *in situ* benthic chambers or incubations of rocks or sediments
- *Typical spatial scale:* Patch scale (centimeter to meter)

**Water column metabolism:** Photosynthesis and respiration occurring within the water column, often by phytoplankton and microbes

- *Typical methods:* Estimated using *in situ* chambers such as light-dark bottles
- *Typical spatial scale:* Patch scale (decimeter to meter)

**Hyporheic respiration:** Biologically mediated oxygen consumption along subsurface flow paths where surface water exchanges with shallow groundwater

- *Typical methods:* Quantified using *in situ* DO, *in situ* hyporheic chambers, or inferred from residence-time modeling, and tracer- or time-series approaches (e.g., argon, conductivity)
- *Typical spatial scale:* Patch scale (centimeter to meter) to sub-reach (1 to 10 meters)

**Gross Primary Production (GPP) and Ecosystem Respiration (ER):** Total oxygen produced via photosynthesis and consumed by autotrophs, heterotrophs, sediments, and subsurface pathways

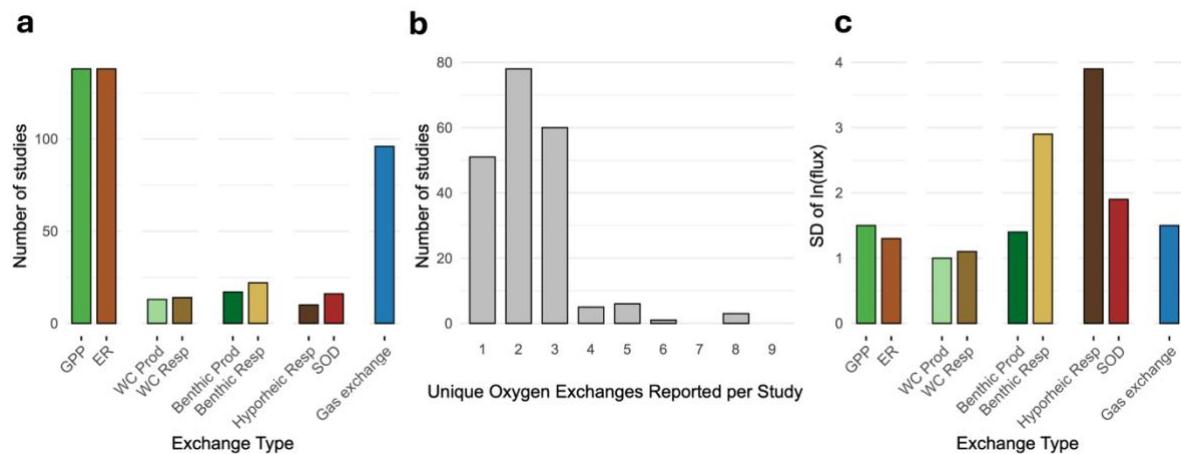
- *Typical methods:* Inferred from diel DO signals using diel metabolism models or *in situ* chambers
- *Typical spatial scale:* Reach-scale (10s of meters to kilometer)

**Surface-atmosphere exchange:** Physical exchange of oxygen between the stream and the atmosphere

- *Typical methods:* Estimated using *in situ* DO time series, tracer gas injections (e.g., SF<sub>6</sub>, propane, argon), empirical equations, or floating probes
- *Typical spatial scale:* Reach-scale (10s of meters to kilometer)

458 **4. LITERATURE ANALYSIS: RESULTS**

459 *4.1 Which oxygen exchanges are most frequently reported and co-reported?*



460

461 **Figure 4. Reporting frequency and variability of oxygen exchange measurements across**  
 462 **studies.** (a) Number of studies reporting each oxygen exchange type, including GPP and ER, water  
 463 column and benthic production and respiration, hyporheic respiration and SOD, and gas exchange.  
 464 (b) Distribution of the number of unique oxygen exchanges reported per study. (c) Variability in  
 465 reported oxygen exchanges across studies, expressed as the standard deviation of the log-  
 466 transformed flux values for each exchange type.

467

468 Across all studies and exchange groups, reporting effort was highly variable (Fig. 4a). Whole-stream  
 469 processes overwhelmingly dominated, with GPP and ER each reported in 33% of (66% combined),  
 470 followed by the surface-atmosphere exchange needed to estimate metabolism (18%). Together,  
 471 these three exchanges comprise the essential elements of Odum's (1956) whole-stream metabolism  
 472 model and accounted for roughly 84% of all measurements, underscoring the emphasis on whole-  
 473 stream metabolism within the literature. In contrast, reporting of underlying process domains was  
 474 comparatively limited, with benthic exchanges reported in 7% of studies, sediment oxygen demand in  
 475 5%, water column processes in 4%, and hyporheic respiration in only 1%.

476

477 Co-reported exchanges further illustrate this pattern. GPP and ER were jointly reported in 65% of  
 478 studies, and gas exchange was commonly paired with one or the other (31% with ER; 30% with GPP;  
 479 Fig. S5). Beyond whole-stream and surface-atmosphere exchanges, co-reporting declined sharply:  
 480 benthic production and benthic respiration were the next most frequently paired exchanges at 7%,  
 481 followed by water column production with gas exchange (5%), and water column production with  
 482 water column respiration (4%). Notably, hyporheic respiration and SOD were never reported together,

483 and when they were measured, they were typically paired alongside benthic or water column  
484 exchanges. The distribution of unique exchanges per study further emphasizes this imbalance, where  
485 94% of studies reported three or fewer oxygen exchanges, indicating that most investigations focus  
486 on whole-stream metabolism, often paired with gas exchange, while omitting the water column,  
487 benthic, and hyporheic and sediment measurements contributing to whole-stream oxygen signals  
488 (Fig. 4b).

489

490 Only a handful of studies characterize different oxygen exchanges within streams—Acuña et al.  
491 (2011), Hogsett & Goel (2013), and Oliver & Merrick (2006) reported eight exchanges (excluding  
492 SOD), and one study (Chessman, 1985) reported six. By measuring multiple oxygen and carbon  
493 exchanges simultaneously, these studies gained important information on the spatial and temporal  
494 variability of different exchanges, with controls shifting across hydrologic conditions, disturbance, and  
495 position along the river continuum. Key insights included the importance of specific processes and  
496 compartments, such as the hyporheic zone accounting for a majority of ER and floating macroalgae  
497 dominating GPP (Acuña et al., 2011), as well as identifying ecosystems where net water column  
498 production exceeded net benthic production (Oliver & Merrick, 2006). Importantly, ecosystems with  
499 different metabolic pathways can exhibit similar whole-stream signals, including similar GPP:ER  
500 ratios. For example, Oliver & Merrick (2006) showed that sites with different dominant carbon sources  
501 (e.g., phytoplankton vs benthic) can have similar NEP values, underscoring the importance of  
502 knowing how different zones are contributing to whole-ecosystem productivity and food web  
503 energetics.

504

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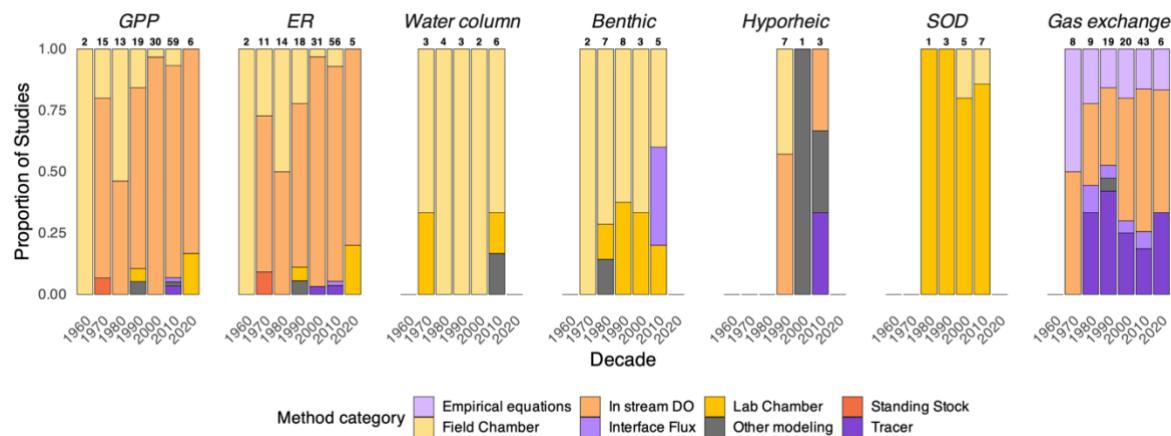
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508

509

510 4.2 How are oxygen exchanges being measured?



511

512 **Figure 5. Temporal evolution of methods used to quantify oxygen exchanges across studies.**  
513 Stacked bar plots show the proportional contribution of different methodological approaches used to  
514 estimate oxygen exchanges by decade for each exchange type (GPP, ER, water column, benthic,  
515 hyporheic, SOD, and gas exchange). Colors indicate method categories, including empirical  
516 equations, field, lab, and flux chambers, standing stock, tracers, in-stream DO, and other modeling  
517 approaches. Numbers above bars denote the total number of studies contributing data in each  
518 decade. See Fig. 3 for collective methods by exchange type.  
519

520 Across oxygen exchanges, methodological differences reflected both the spatial scale of  
521 measurement and the process of interest (Box 1). In-stream DO was the most common approach for  
522 quantifying gas exchange (49% of studies), GPP (81%), ER (79%), and hyporheic respiration (46%),  
523 reflecting both discrete and continuous measurements that have evolved through time (Fig. 3). DO-  
524 based estimates of hyporheic respiration often relied on the difference between ecosystem and  
525 benthic metabolism rather than being directly measured (e.g., Naegeli & Uehlinger, 1997; Acuña et  
526 al., 2011; see section 5.2 for limitations). After in-stream DO, gas exchange was most commonly  
527 quantified using tracers (25% of studies), encompassing a variety of gases—including argon,  
528 propane, helium, xenon, ethylene, nitrogen, sulfur hexafluoride, and krypton (e.g. Young & Huryn,  
529 1999; Roley et al., 2023)—with additional tracers used to link gas exchange and metabolic rates  
530 (Piatka et al., 2021; Barbosa et al., 2023), followed by empirical equations (20%). In contrast, water  
531 column, benthic, and SOD measurements were dominated by chamber-based approaches, with field  
532 chambers accounting for 83%, 64%, 12% of studies and lab chambers comprising 11%, 24%, 88%,  
533 respectively.

534

535 Beyond in-stream DO, tracers, and chamber-based approaches, a suite of less frequently applied  
536 methods has been used to quantify oxygen exchanges. Interface flux approaches have expanded  
537 exchange measurements across flow regimes by capturing near-continuous gradients *in situ* such as  
538 the soluble floating probe method using oxalic acid as a reactive tracer (De Souza Ferreira et al.,  
539 2020), floating probes with O<sub>2</sub> sensors (Hogsett & Goel, 2013), and aquatic eddy covariance  
540 (Koopmans & Berg, 2015). The interface flux category was highest for benthic exchanges (8% of  
541 studies), then gas exchange (6%), and GPP/ER (1%). Other approaches included the standing stock  
542 method—applied exclusively to GPP and ER (1%)—relating photosynthetic or respiratory activity to  
543 algal biomass or organic carbon content. Lastly, process-based modeling studies informed by  
544 empirical observations, including frameworks such as the PROSE model (Even et al., 2004) or  
545 reactive transport models (Diem et al., 2013) accounted for 18% of hyporheic, 6% of water column,  
546 4% of benthic, and less than 1% of GPP, ER, and gas exchange studies.

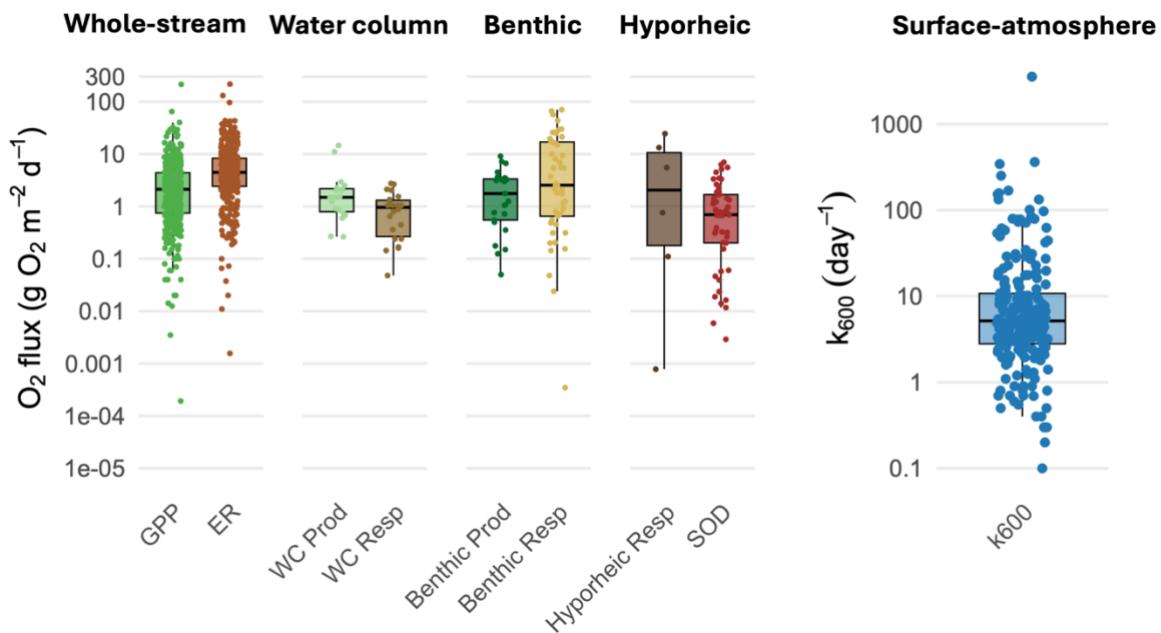
547

548 Methods to quantify exchanges have also shifted through time, both in frequency and diversity (Fig.  
549 5). After relatively few studies in the 1960s, research on most oxygen exchanges increased in the  
550 1970s (except for hyporheic exchange and SOD) and peaked in the 2010s. The apparent decline in  
551 the number of studies in 2020 is reflected by limitations of our analysis only considering studies with  
552 DOIs assigned before August 2024. Field chambers dominated early studies, comprising all studies in  
553 the 1960s and about half in the 1970s, but declining to 8% by the 2010s. In contrast, in-stream DO  
554 methods have greatly increased in number and gradually increased by percentage from 46% in the  
555 1970s to 56% in the 2020s, reflecting the transition from discrete sampling to continuous sensor  
556 deployments as sensor technology advanced after 2000. Whole-stream metabolism methods have  
557 remained relatively stable over time, with single-station approaches predominating, while  
558 methodological diversity increased in the 1990s-2000s with greater use of two-station and delta  
559 approaches and continues today through use of delta, extreme value, and oxygen-carbon methods  
560 (Fig. S4). Tracer techniques were first applied in the 1980s, comprising about 10% of studies,  
561 increasing in the 1990s, decreasing in the 2000s-2010s and rebounding to 22% in the 2020s with the  
562 emergence of new field methods (e.g., argon; Hall & Madinger, 2018), while empirical equations  
563 plateaued during the 1990s at 23% and have remained relatively stable since. Interface flux and  
564 modelling approaches only became common after 2010 (10-15%), mainly for benthic and hyporheic

565 studies. Overall, the data methods for GPP, ER, and gas exchange have remained stable since the  
566 1980s, comprised by in-stream DO and tracers (although computational methods have evolved),  
567 while longer-term methodological shifts reflect a decline in chamber-based studies and increasing  
568 diversification in approaches used to quantify benthic and hyporheic exchanges.

569

570 *4.3 How variable are the oxygen exchanges in lotic ecosystems?*



571

572 **Figure 6. The magnitude of oxygen fluxes reported across all studies.** Boxplots summarize the  
573 distribution of reported oxygen fluxes for whole-stream metabolism (GPP and ER), water column and  
574 benthic production and respiration, hyporheic respiration and sediment oxygen demand, and surface-  
575 atmosphere exchange ( $k_{600}$ ), shown on log-scaled and standardized axes. Boxes represent  
576 interquartile ranges with medians, whiskers indicate 1.5 $\times$  the interquartile range, and points show  
577 mean values across individual sites and studies.

578

579 Among all oxygen exchanges standardized to fluxes (except gas exchange, which is standardized to  
580  $k_{600}$ ), site-level mean magnitude varied among groupings (Fig. 6; Table S2). Within whole-stream  
581 metabolism, ER exceeded GPP on average (means = 8.0 and 4.4 g O<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup>; n = 364 and 371,  
582 respectively), meaning most streams were heterotrophic (consistent with Bernhardt et al., 2018). For  
583 water column observations, production was higher than respiration on average (means = 2.6 and 1.0  
584 g O<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup>; n = 19 and 22, respectively). In contrast, the maximum benthic respiration (range = 3.5  
585  $\times$  10<sup>-4</sup> to 70 g O<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup>; n = 49) was roughly an order of magnitude higher and far more variable  
586 than the maximum benthic production (range = 0.05 to 9.1 g O<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup>; n = 25). Hyporheic

587 respiration was markedly higher on average than SOD (means = 7.4 and 1.3 g O<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup>; n = 6 and  
588 54), implying methodological dissonance in this critical compartment flux.

589

590 Oxygen exchanges differed markedly in their variability, expressed as the standard deviation of log-  
591 transformed fluxes ( $\sigma_{ln}$ ) to account for strong right skew (Fig. 3c; Table S2). Hyporheic respiration  
592 was the most variable ( $\sigma_{ln} = 3.9$ ), followed by benthic respiration ( $\sigma_{ln} = 2.4$ ), SOD ( $\sigma_{ln} = 1.9$ ), and gas  
593 exchange ( $k_{600}$ ;  $\sigma_{ln} = 1.8$ ). Whole-stream processes, benthic production, and water column production  
594 and respiration were less variable, with  $\sigma_{ln}$  values of 1.5 for GPP, 1.3 for ER, 1.4 for benthic  
595 production, and 1.0 and 1.1 for water column production and respiration. This implies that the most  
596 variable compartment-specific fluxes are also the least frequently measured.

597

598 GPP was lowest in small streams (mean = 2.8 g O<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup>), peaked in mid-sized streams (mean =  
599 6.5), and declined in large streams (mean = 5.2; Fig. S1; Table S3). ER followed a similar pattern but  
600 had consistently higher magnitudes, averaging 6.8 g O<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup> in small streams, 10 g O<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup>  
601 in mid-sized, and 5.6 g O<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup> in large streams. Benthic production was relatively consistent  
602 across stream sizes (means = 2.7, 2.6, and 2.3 g O<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup> for small, mid, and large streams),  
603 whereas average benthic respiration was highest in mid-sized streams (15 g O<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup>). SOD was  
604 relatively consistent across stream size (means ranging from 0.95 to 1.6 g O<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup>). Gas  
605 exchange coefficients ( $k_{600}$ ) generally decreased with stream size, averaging 66 d<sup>-1</sup> in small streams  
606 and 7.0 d<sup>-1</sup> in both mid-sized and large streams. Limited reporting of water column and hyporheic  
607 fluxes across stream size constrained our ability to distinguish trends. Overall, oxygen exchanges  
608 tended to peak in mid-sized streams, with higher variability in small streams and lower magnitudes in  
609 larger rivers (Fig. S1).

610

611 GPP and ER showed similar changes with land use, with the highest values in agricultural streams  
612 (means = 7.1 and 10.5 g O<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup>) followed by urban (means = 5.4 and 8.8 g O<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup>),  
613 mixed (means = 4.4 and 9.2 g O<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup>), and lowest in undisturbed ecosystems (means = 2.8 and  
614 5.8 g O<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup>), although streams in all land use categories were heterotrophic (Fig. S1; Table  
615 S3). Water column production was highest in agricultural (mean = 3.4 g O<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup>) and urban

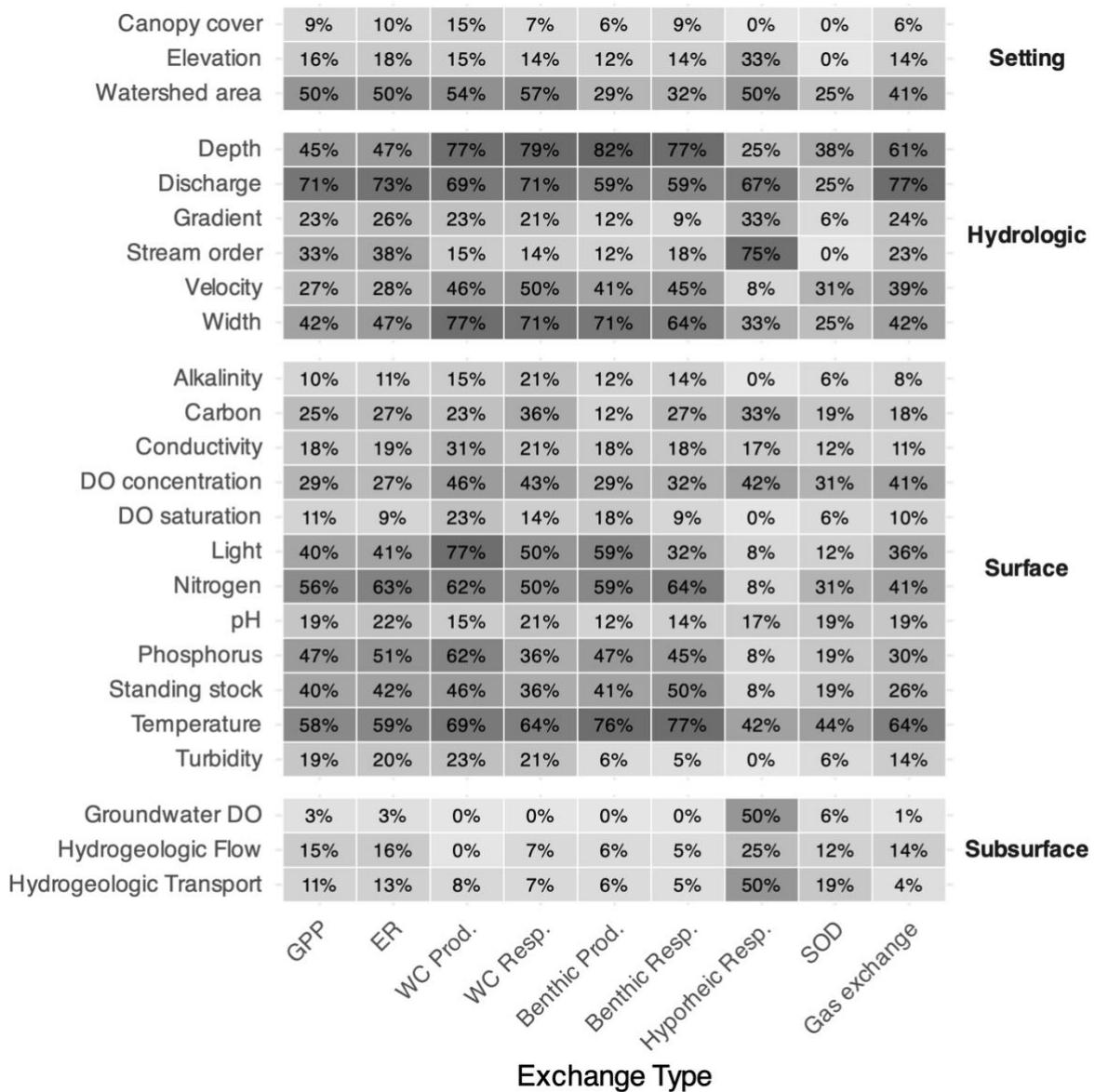
616 streams (mean = 3.1 g O<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup>), whereas water column respiration was highest in undisturbed  
617 (mean = 2.0 g O<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup>). Benthic fluxes had a stronger land use response, where benthic  
618 respiration was markedly elevated in urban streams averaging 23.6 g O<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup>, followed by mixed  
619 (mean = 7.9 g O<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup>), agricultural (mean = 2.8 g O<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup>), and undisturbed ecosystems  
620 (mean = 1.1 g O<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup>), while benthic production also peaked in urban streams (mean = 4.0) but  
621 with smaller contrasts among categories. Hyporheic respiration lacked enough data to infer patterns  
622 across land use, whereas SOD was most frequently measured and highest in urban and mixed  
623 streams.  $k_{600}$  was highest in undisturbed streams (127.9 d<sup>-1</sup>) and substantially lower in agricultural  
624 (16.4 d<sup>-1</sup>), mixed (8.2 d<sup>-1</sup>), and urban ecosystems (5.4 d<sup>-1</sup>). Overall, whole-stream, water column,  
625 benthic, and hyporheic and sediment exchanges peaked in agricultural, urban, and mixed land use  
626 streams, while surface-atmosphere exchange was highest in undisturbed ecosystems (Fig. S1).

627

628 Together, these patterns highlight a shift in the dominant drivers of oxygen fluxes across stream size  
629 and land use: while natural gradients predict dominant subsurface metabolism in headwaters and  
630 increasing water column activity downstream, elevated respiration in mid-sized and urban streams  
631 indicates that human modification can disrupt this expected metabolic hierarchy along the river  
632 continuum (Ledford et al., 2021).

633

634 *4.4 What complementary information & metadata are we reporting alongside oxygen exchange data?*



635

636 **Figure 7. Reporting frequency of metadata across oxygen exchange types across studies.** The  
 637 heatmap shows the percentage of studies reporting key metadata for each oxygen exchange type,  
 638 grouped by setting, hydrologic properties, surface water chemistry and light, and subsurface  
 639 attributes. Darker cell shading indicates a greater percentage of studies reporting a given covariate.  
 640 The number of studies reported for each cell, rather than the percent, are presented in Fig. S2.  
 641

642 Oxygen dynamics in streams are shaped by complex interactions among physical, chemical, and  
 643 biological factors, with dominant controls expected to vary as we consider different exchanges and  
 644 settings (Fig. 7). Despite this heterogeneity, some variables consistently emerge as more commonly  
 645 reported regardless of which oxygen exchange(s) were included in a study. Discharge and  
 646 temperature were the most common covariates, reported in 70% and 56% of all studies, respectively.  
 647 Among total, particulate, and dissolved nutrient concentrations (see methods in section 3 for the

648 included criteria), nitrogen was most often reported (47% of studies), followed by phosphorus (37%),  
649 and carbon (26%). Light-related measurements and standing stock indicators were included in about  
650 30% of studies, while turbidity (13%), alkalinity (10%), and canopy cover (6%), were less frequently  
651 reported.

652

653 Among surface water variables, temperature, nitrogen, and light were most commonly reported,  
654 particularly for whole-stream, water column, and benthic exchanges, aligning with expectations from  
655 decades of metabolism literature (Bernhardt et al., 2022). Light was reported in 77% of water column  
656 production studies and 59% of benthic production, demonstrating the importance of light in  
657 photosynthesis. Standing stocks were also frequently reported in water column, benthic and whole-  
658 stream metabolism studies (36-50%), but less reported in gas exchange (26%), SOD (19%), and  
659 hyporheic respiration (8%).

660

661 By contrast, subsurface variables were rarely reported across all groupings. Groundwater DO – an  
662 essential baseline for subsurface oxygen conditions and the impacts of lateral inputs on metabolic  
663 estimates – had highest reporting for hyporheic respiration (50%) and SOD (6%), followed by whole-  
664 stream (3% for GPP and ER) and surface-atmosphere (1%) measurements, and was not reported at  
665 all for water column and benthic exchanges. Hydrogeologic transport metrics which characterize  
666 subsurface water fluxes and timescales - such as residence time, hyporheic exchange fluxes,  
667 dispersivity, and diffusion coefficients - were reported in about 15% of whole-stream metabolism, 12%  
668 of SOD, and 25% of hyporheic respiration studies. Hydrogeologic flow properties, such as hydraulic  
669 conductivity and transmissivity, which describe the ease with which water moves through subsurface  
670 materials, were included in fewer than 13% of studies, though were highest in hyporheic-focused  
671 studies (50%). Organic and inorganic species of carbon were the least reported nutrient across all  
672 studies, but most often included with measurements of water column, benthic, and hyporheic  
673 respiration, reflecting the coupling between organic carbon consumption and inorganic carbon  
674 production during heterotrophic metabolism.

675

676 While the influence of the subsurface on in-stream processes is widely acknowledged, our synthesis  
677 demonstrates that subsurface variables remain the least reported when studying DO dynamics in

678 flowing waters. This gap is not a result of oversight but rather reflects methodological challenges of  
679 measuring processes that are spatially heterogenous, transient, and operate across several orders of  
680 magnitude (Strayer et al., 2003). Variables describing subsurface transport and transformation are  
681 inherently system-dependent; for example, hyporheic fluxes may be easier to measure and influence  
682 in-stream processes more in a steep headwater stream than a low-gradient river (González-Pinzón et  
683 al., 2015). Similarly, carbon pools encompass a wide range of forms (DIC, DOC, POC) and the most  
684 relevant reporting metric depends on the dominant pathway of heterotrophic metabolism and the  
685 research objective.

686

687 The heterogenous reporting of metadata across studies is not unique to stream oxygen exchanges  
688 but reflects a broader challenge shared across interdisciplinary fields. Understanding oxygen  
689 dynamics requires moving beyond single-domain perspectives and instead considering multiple,  
690 interacting controls that operate simultaneously (Wohl, 2017; Wymore et al., 2023). Metadata  
691 reporting standards that follow from this perspective will enable future synthesis of DO dynamics to be  
692 far more inclusive.

693

## 694 **5. CURRENT CHALLENGES AND OPPORTUNITIES TO ADVANCE OUR 695 UNDERSTANDING OF LOTIC OXYGEN DYNAMICS**

696 *5.1 Ecosystems are not uniformly sampled, biasing our basis for understanding and predicting DO  
697 exchanges in streams and rivers*

698 Global research on stream and river DO has been spatially and temporally biased to-date. Most  
699 studies have been conducted in North America and Europe with underrepresentation in tropical and  
700 high-latitude systems. Study designs are also skewed toward local scales, with a majority quantifying  
701 DO within a single stream (51%), then by comparisons among streams (32%), and lastly longitudinally  
702 along the river corridor (17%), limiting understanding of how variability scales from reach to watershed  
703 and across sites. Temporal bias further compounds these biases as a majority (54%) of the studies  
704 included in our synthesis spanned less than one year, 39% lasted two to three years, and 6%  
705 extended beyond four years. Although maximum study duration increases substantially in later years,  
706 reflecting the emergence of long-term monitoring programs and sensor capability, mean study

707 duration remains persistently low. While short-term studies can provide valuable mechanistic insight  
708 under specific conditions, transferring these findings across diel, seasonal, and interannual scales will  
709 be key to understanding ecosystem responses across the full flow, light, temperature, and  
710 disturbance regimes of river networks.

711

712 Expanding and continuing existing work in under-represented regions is therefore critical. Initiatives  
713 such as the U.S. Department of Energy's Next-Generation Ecosystem Experiments and field-based  
714 programs at high latitudes—such as the McMurdo Dry Valleys LTER and the Arctic LTER—are key to  
715 addressing these gaps. Similarly, cross-site comparisons and coordinated efforts like the Lotic  
716 Intersite Nitrogen eXperiments (LINX; Mulholland et al., 2008; Bernot et al., 2010; Wymore et al.,  
717 2019) demonstrate the value of standardized, collaborative approaches for identifying patterns across  
718 diverse hydrologic and geographic gradients. Tools such as streamMetabolizer (Appling et al., 2018a)  
719 demonstrate how analytical standardization can enable cross-site synthesis of lotic oxygen dynamics,  
720 highlighting a clear opportunity to extend similar coordination to field-based measurements and  
721 experimental design.

722

723 Hydrologic regime bias remains one of the most significant gaps and opportunities in oxygen  
724 research. Although roughly half of global rivers are intermittent (Messager et al., 2021), only 4% of  
725 studies in our synthesis include non-perennial rivers. Our current methods used to characterize  
726 whole-stream metabolism were primarily developed for perennial waterways and may not accurately  
727 represent non-perennial streams. This underrepresentation overlooks stream networks that are both  
728 highly vulnerable to climate-driven hydrologic variation (Ward et al., 2020) and disproportionately  
729 important for biogeochemical processing (Datry et al., 2014; Stegen et al., 2025). A growing body of  
730 evidence indicates that intermittency introduces pronounced spatial and temporal heterogeneity in  
731 ecosystem function, particularly through repeated cycles of drying and rewetting that can stimulate  
732 organic matter decomposition and respiration. For example, soil research has long documented the  
733 “Birch effect,” in which rewetting of dry soils triggers large, short-lived respiration pulses (Manzoni et  
734 al., 2020), and recent studies indicate that similar dynamics occur in aquatic systems, where  
735 respiration from dry stream sediments increases following rewetting (Arce et al., 2021). However,  
736 Bretz et al. (2023) found that rewetting can initially suppress observable CO<sub>2</sub> signals via dilution and

737 degassing, with concentrations rebounding as surface water contracts, while Ward et al. (2019) found  
738 little variation in biogeochemical function through diurnal wetting and drying cycles. Together, these  
739 studies highlight that respiration signals in non-perennial streams are highly variable and can be  
740 amplified, delayed, masked by, or insensitive to changes in hydrologic connectivity, indicating a fruitful  
741 area for future research.

742

743 To address persistent hydrologic regime biases and improve cross-system inference, coordinated  
744 initiatives are underway such as the Dry Rivers Research Coordination Network (Datry et al., 2023;  
745 Price et al., 2024) and the Aquatic Intermittency effects on Microbiomes in Streams (Open Science  
746 Framework, 2025) that integrate cross-biome comparisons, coordinated sampling, the development of  
747 new methods and models appropriate for non-perennial ecosystems, and synthesis of ecosystem  
748 responses to non-perennial flow.

749

750 *5.2 Reporting units, methodological silos, and scaling barriers challenge intercomparison*  
751 Intercomparison of oxygen exchanges among studies is fundamentally constrained by reporting units.  
752 Gas exchange, for example, is typically reported as a rate constant ( $d^{-1}$ ) or as a velocity ( $m d^{-1}$ ), but  
753 the literature contains at least 26 distinct gas exchange coefficients (Fig. S3), although recent years  
754 have seen a convergence toward using a normalized  $k_{600}$  ( $day^{-1}$ ). Whole-stream metabolism is  
755 typically expressed in areal units ( $g O_2 m^{-2} d^{-1}$ ), whereas water column, benthic, and hyporheic  
756 studies are often reported as volumetric rates ( $g O_2 L^{-1} d^{-1}$ ), with some hyporheic and transport-  
757 reaction studies additionally expressing respiration as first-order DO uptake or decay coefficients  
758 ( $d^{-1}$ ), further complicating direct comparison across studies and exchange types without appropriate  
759 stream attributes.

760

761 Oxygen exchanges are typically studied in isolation using methods optimized for specific spatial  
762 scales, reinforcing methodological and disciplinary silos that hinder integration. These approaches  
763 span patch-scale chambers and incubations for sediment and benthic exchanges to sub-reach water  
764 column metabolism assays and reach-scale sensors and associated measurements for GPP, ER, and  
765 gas exchange (Box 1). Because these methods differ not only in spatial and temporal resolution, but

766 also in the processes they are designed to resolve, direct comparison and synthesis across exchange  
767 types remains challenging.

768

769 Scaling isolated measurements to the reach-scale introduces additional uncertainty, particularly  
770 related to reach geometry, depth normalization, and underlying spatial variability. Nevertheless, such  
771 standardization remains necessary to compare fluxes across sites and systems and understand their  
772 sources. In practice, researchers routinely apply scaling assumptions for whole-stream metabolism,  
773 accepting a degree of error as a trade-off for interpretability and synthesis. Water column and benthic  
774 chambers can be standardized from volumetric rates (e.g.,  $\text{g O}_2 \text{ L}^{-1} \text{ d}^{-1}$ ) to an areal basis (e.g.,  $\text{g O}_2$   
775  $\text{m}^{-2} \text{ d}^{-1}$ ) by normalizing rates to the chamber footprint area. However, hyporheic metabolism is far less  
776 frequently scaled to reach-scale fluxes, as doing so requires additional assumptions about  
777 metabolically active sediment depth, exchange volume, water residence time, or subsurface  
778 connectivity that are rarely constrained empirically. As a result, comparable standardization efforts  
779 remain uncommon for hyporheic processes, despite their central role in oxygen dynamics and their  
780 potential to dominate whole-stream fluxes. While variation in substrate type, organic matter quantity  
781 and quality, and uncertainty in delineating hyporheic boundaries add complexity, these challenges  
782 should not preclude systematic attempts at harmonization. Coenen et al. (2019) emphasized the  
783 importance of standardizing SOD measurements to enhance comparability – an approach that we call  
784 for extending to all oxygen exchanges to foster integration and synthesis.

785

786 As a result of methodological fragmentation and scaling uncertainty, studies of whole-stream  
787 metabolism, water column, benthic, hyporheic, SOD, and gas exchange have largely advanced in  
788 parallel rather than in an integrated fashion. Although some efforts have attempted to bridge these  
789 gaps—for example, by comparing patch-scale and whole-stream metabolism estimates despite  
790 substantial methodological and scaling challenges (e.g., Tromboni et al., 2017) or by linking  
791 metabolism and nutrient uptake dynamics (e.g., Hoellein et al., 2009)—these integrative approaches  
792 remain rare in our review of the literature. This pattern likely reflects both logistical constraints and  
793 disciplinary incentives that prioritize measurements directly aligned with specific research questions,  
794 rather than coordinated multi-flux designs. Differences in training, terminology, and research

795 questions among hydrologists, biogeochemists, and ecologists have further limited integration across  
796 subsurface and whole-ecosystem perspectives, resulting in a lack of “connective tissue” that  
797 constrains synthesis across spatial scales and disciplines (Box 1). The integration of pathway-specific  
798 fluxes, specifically via local measurements of gas exchange, has already been shown to be  
799 enormously valuable for constraining open-channel metabolism methods. We submit that similar  
800 improvements in the accuracy of and controls on whole-system metabolism will emerge from routinely  
801 connecting benthic and water column fluxes to the common DO concentration inferences.

802  
803 These methodological and scaling challenges also complicate the application of the Odum (1956)  
804 oxygen budget framework, whose simplifying assumptions are increasingly challenged in ecosystems  
805 where subsurface processes strongly influence DO dynamics. We caution against differencing whole-  
806 stream metabolism estimates with independently measured benthic and water column exchanges, as  
807 this approach can introduce statistical uncertainty and process error in systems where hyporheic  
808 respiration contributes substantially to DO dynamics and reach-scale GPP and ER partially reflect  
809 unresolved subsurface processes (Naegeli & Uehlinger, 1997; Gomez-Velez et al., 2026). Rather  
810 than treating methods as interchangeable, complementary approaches should be combined to better  
811 constrain which ecosystem components are contributing to whole-stream metabolism. Modified two-  
812 station metabolism models, for example, are required to account for groundwater influences on  
813 oxygen dynamics that may bias estimates of GPP and ER (Hall and Tank, 2005). Yet, the use of two  
814 station models (with or without groundwater corrections) has declined from a peak in the 1990s—  
815 when 50% of whole-stream metabolism studies employed a two-station design—to ~25% in the 2000s—  
816 and 2010s. Re-diversifying beyond single-station models is therefore particularly important for  
817 systems that violate the core assumptions underlying those approaches.

818  
819 *5.3 DO should be coupled with other biogeochemical cycles to move from inference to process-based*  
820 *understanding*  
821 Explicitly coupling DO with other biogeochemical cycles presents an opportunity to move towards a  
822 more integrated understanding of ecosystem processes. While whole-stream metabolism provides  
823 valuable insight into aerobic carbon cycling (i.e., presuming oxygen consumption is predominantly  
824 related to aerobic respiration), it represents an integrated metric that cannot inform all underlying

825 transformations occurring within streams. Integrating DO with complementary measures such as other  
826 naturally occurring dissolved gases (e.g., CO<sub>2</sub>, CH<sub>4</sub>, Ar), nutrient spiraling, and study of individual  
827 fluxes (e.g., water column, benthic, and hyporheic zones) can directly link oxygen dynamics with  
828 nutrient and carbon transformations. For example, by coupling DO-derived metabolism with organic  
829 carbon and gas exchange, Hall et al. (2016) showed how oxygen dynamics can be used to infer  
830 large-scale patterns in carbon metabolism and transport across river networks. Similar cross-scale  
831 measurement approaches have illustrated how DO dynamics relate to nitrification, denitrification, and  
832 DOC metabolism across land use gradients (Hanrahan et al., 2018; Plont et al., 2022), collectively  
833 revealing that oxygen analyses offer enhanced insights into ecosystem processes when considering  
834 its tight coupling to carbon and nutrient transformations.

835  
836 Moving forward, advancing multi-solute and cross-disciplinary frameworks will be critical to link DO-  
837 based metabolism with CO<sub>2</sub>, nutrient, and redox dynamics. Discrepancies between CO<sub>2</sub>- and O<sub>2</sub>-  
838 based estimates of stream metabolism (Vachon et al., 2020, Shangguan et al., 2024) highlight  
839 processes exist that are not captured in DO measurements alone and different spatial scales  
840 associated with different gas dynamics. Decoupling may arise from anaerobic respiration that does  
841 not impact DO concentration, or hyporheic pathways that induce long lags in water column DO  
842 signals (Gomez-Velez et al. 2026). Sources of error are also compounded by the widespread use of a  
843 1:1 O<sub>2</sub>-CO<sub>2</sub> stoichiometric ratio when converting DO-based metabolism to carbon units as this ratio is  
844 known to vary (Trentman et al., 2023), particularly in ecosystems with strong groundwater influence  
845 (Rocher-Ros et al., 2025). Together, these patterns underscore that aerobic and anaerobic respiration  
846 reflect distinct redox pathways that together shape organic matter decomposition and nutrient  
847 turnover, yet quantifying anaerobic respiration remains challenging due to its spatial heterogeneity.  
848 Studies such as Son et al., (2022) demonstrate that hyporheic exchange can drive large-scale  
849 variability in anaerobic metabolism, underscoring the need to explicitly represent subsurface  
850 pathways in metabolic frameworks. Developing modeling approaches capable of partitioning  
851 autotrophic, heterotrophic, and anaerobic signals (Hall & Beaulieu, 2013; Diamond & Bertuzzo, 2025)  
852 will be key to linking empirical data and process-based understanding.

853

854     *5.4 Opportunities for integrating oxygen exchanges across scales through metadata and emerging*  
855     *modeling*  
856     Translating process-based insights into cross-study synthesis requires consistent reporting of core  
857     metadata that enable integration across spatial scales and datasets. At a minimum, reporting the  
858     mean or median oxygen exchange alongside location (latitude and longitude), discharge, water  
859     temperature, channel depth, and width for the study period would substantially improve comparability  
860     across studies. Where subsurface processes are expected to influence DO dynamics, reporting basic  
861     indicators of hyporheic or groundwater exchange (e.g., groundwater DO, sediment depth, residence  
862     time, or tracer-based exchange metrics) would further interpretation. Importantly, variables such as  
863     discharge and depth are difficult to predict accurately at the reach scale, and gas exchange estimates  
864     derived from empirical equations or normalized by depth often require direct measurement for reliable  
865     constraint. In contrast, others such as stream order, land cover, and catchment slope can often be  
866     estimated reliably from existing geospatial products, including the Stream Catchment Attributes  
867     dataset, the National Hydrography Dataset Plus, and the National Land Cover Database (Hill et al.,  
868     2016; USGS, 2023; EPA, 2025). Clearly distinguishing between variables that require direct  
869     measurement and those that can be inferred would facilitate data reuse, reduce duplication of effort,  
870     and strengthen synthesis across spatial scales.  
871  
872     Integrating disciplinary and methodological approaches requires combining complementary methods  
873     across spatial and temporal scales, rather than measuring oxygen exchanges in isolation. Advances  
874     in time series analysis and the use of naturally occurring tracers (e.g., argon, radon, electrical  
875     conductivity) have made it increasingly feasible to quantify hyporheic exchange and respiration  
876     without the need for intensive tracer injections or laboratory incubations (e.g., Vieweg et al., 2016).  
877     Emerging tools such as low-cost floating chambers can be used to independently constrain gas  
878     exchange assumptions in whole-stream oxygen models (Mannich et al., 2019). When combined with  
879     synoptic light-dark incubations and benthic chamber measurements and deployed alongside  
880     continuous sensor networks (e.g., two-station), these approaches enable process-based attribution of  
881     oxygen fluxes across compartments and spatial scales.  
882

883 Since logistical constraints will always limit direct measurement of all oxygen exchanges, advancing  
884 synthesis will increasingly depend on data-driven and modeling approaches that harmonize existing  
885 observations across scales. For instance, given the logistical demands of wells, chambers, or  
886 mesocosms hyporheic exchange and respiration cannot be quantified ubiquitously, so many studies  
887 will focus on a single component of the oxygen budget at a particular location or scale, leaving  
888 substantial gaps in cross-scale understanding and assessment of heterogeneity. Looking ahead, data  
889 science and modeling tools offer new opportunities to harmonize existing datasets and advance  
890 synthesis across exchanges and scales. Large-scale databases and modeling approaches – similar  
891 to those developed by Appling et al. (2018b) for whole-stream metabolism – could be expanded to  
892 include individual exchanges, while emerging basin-scale models demonstrate growing capacity to  
893 infer hyporheic contributions from reach- or network-scale signals (Regier et al., 2025). New time  
894 series-based approaches and machine learning frameworks further show promise for resolving  
895 overlapping metabolic signals and identifying spatial drivers at network scales that are inaccessible  
896 through site-specific studies alone (Segatto et al., 2021; Bertuzzo et al., 2022).

897

## 898 **CONCLUSION**

899 Advancing dissolved oxygen science requires moving beyond single-flux measurements toward the  
900 concurrent quantification of oxygen exchanges across process domains that collectively give rise to  
901 whole-stream signals. Evidence from the relatively few studies that have quantified multiple oxygen  
902 exchanges demonstrates that this added complexity yields disproportionate insight. These studies  
903 reveal strong spatial and temporal variability in compartment-specific processes, shifts in the relative  
904 dominance of individual exchanges (e.g., hyporheic respiration or water column production), and  
905 cases where fundamentally different underlying pathways produce similar whole-stream fluxes.  
906 Importantly, even partial resolution of multiple exchanges has been shown to improve process  
907 attribution by identifying hyporheic exchange as a hidden control on sub-daily oxygen dynamics,  
908 explaining apparent variability and bias in GPP, ER, and gas exchange estimates that would  
909 otherwise be attributed to noise or equifinality. As a result, whole-stream estimates of GPP, ER, or  
910 NEP alone cannot resolve where oxygen and carbon are primarily produced or consumed.  
911 Distinguishing whether oxygen dynamics are dominated by water column, benthic, or hyporheic and  
912 sediment processes has direct implications for the supply and demand of energy in food webs,

913 aquatic ecosystem health, management decisions, and the interpretation of stream DO signals from  
914 which metabolic fluxes are inferred.

915

916 At the same time, boundaries between compartments are often diffuse and temporally dynamic,  
917 further complicating attribution. Sediments may integrate benthic and hyporheic signals, while under  
918 certain hydrologic conditions subsurface flowpaths can dominate metabolic activity influencing in-  
919 stream diel oxygen concentrations. Although oxygen dynamics retain distinct process signatures, their  
920 effective integration length makes them sensitive to small-scale changes; for instance, more  
921 turbulence and mixing can reset oxygen signals and either amplify or dampen hyporheic contributions  
922 depending on flow and season. These sensitivities are especially pronounced in intermittent streams,  
923 where dynamic subsurface flowpaths contribute to streamflow yet remain underrepresented in the  
924 literature. While our review highlights the need for compartment-based measurements, we also  
925 encourage a shift toward a more integrated perspective, in which oxygen production, consumption,  
926 and transport are quantified along hydrologic pathways spanning the water column, benthos,  
927 hyporheic zone, sediments, and groundwater. Such an approach better reflects how oxygen moves  
928 through and is processed within stream networks, with hydrologic connectivity acting as a master  
929 variable that mediates fluxes among compartments and regulates when and where specific  
930 exchanges may dominate.

931

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938

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942

943 **Data Availability**

944 Data are hosted on HydroShare and are currently discoverable at:

945 Krause, J., and A. S. Ward (2026). A Systematic Review of Oxygen Exchanges in Streams and

946 Rivers. HydroShare. <http://www.hydroshare.org/resource/56ce89df533a4d928f0919b92def79aa>

947 The dataset will be made publicly accessible upon publication, at which time a DOI will be issued.

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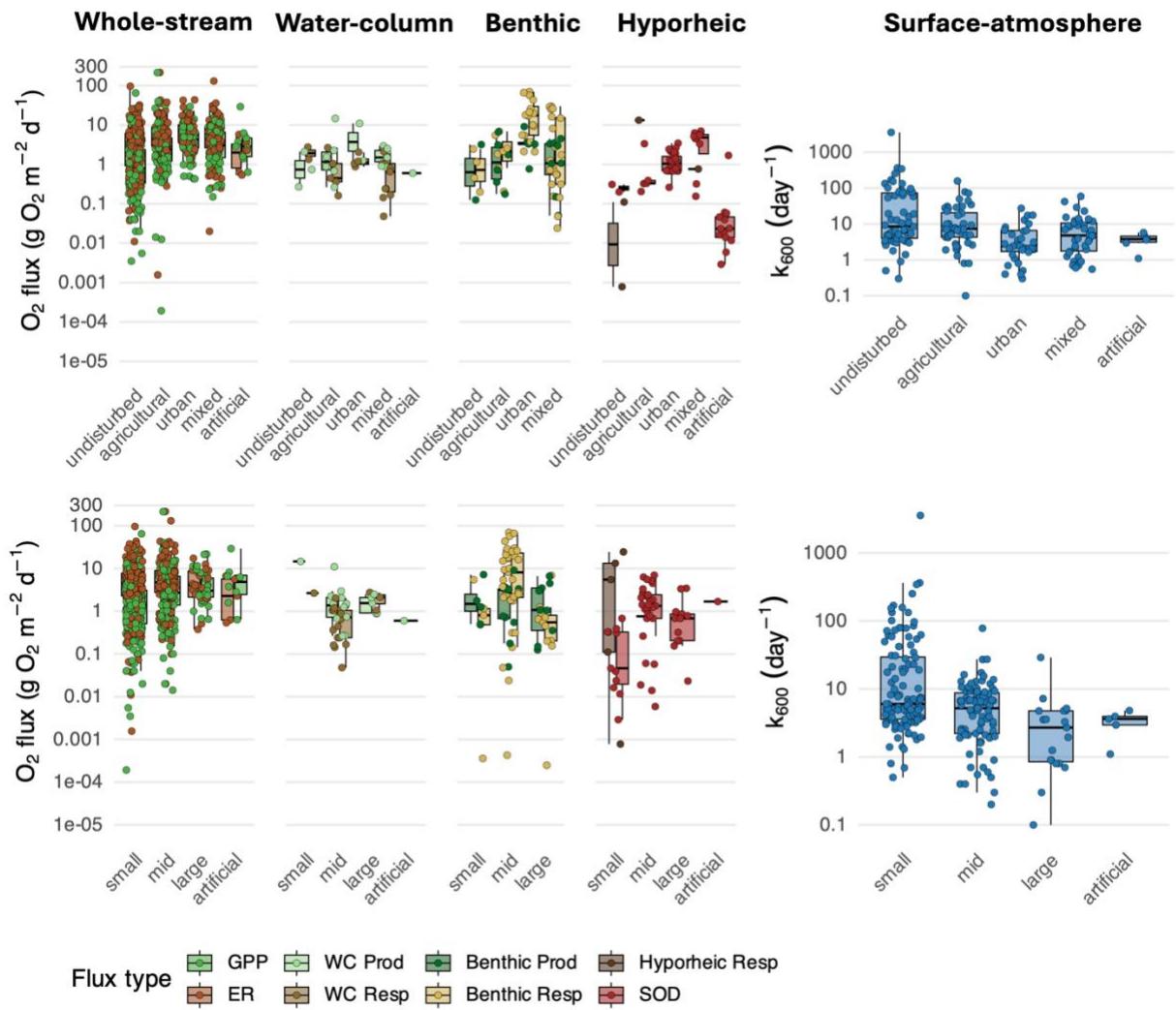


Figure S1. Comparison of flux magnitudes across stream size (top) and land-use (bottom) gradients. Each panel shows oxygen fluxes standardized to units of  $\text{day}^{-1}$  or  $\text{g O}_2 \text{ m}^{-2} \text{ day}^{-1}$  on a log-scaled axis. The top plot shows exchanges by land use (undisturbed, agricultural, urban, mixed, or artificial), while the bottom plot shows exchanges by stream size (small, mid, large, and artificial). Boxes represent the interquartile range with medians, whiskers indicate  $1.5 \times \text{IQR}$ , and points show means across study sites.

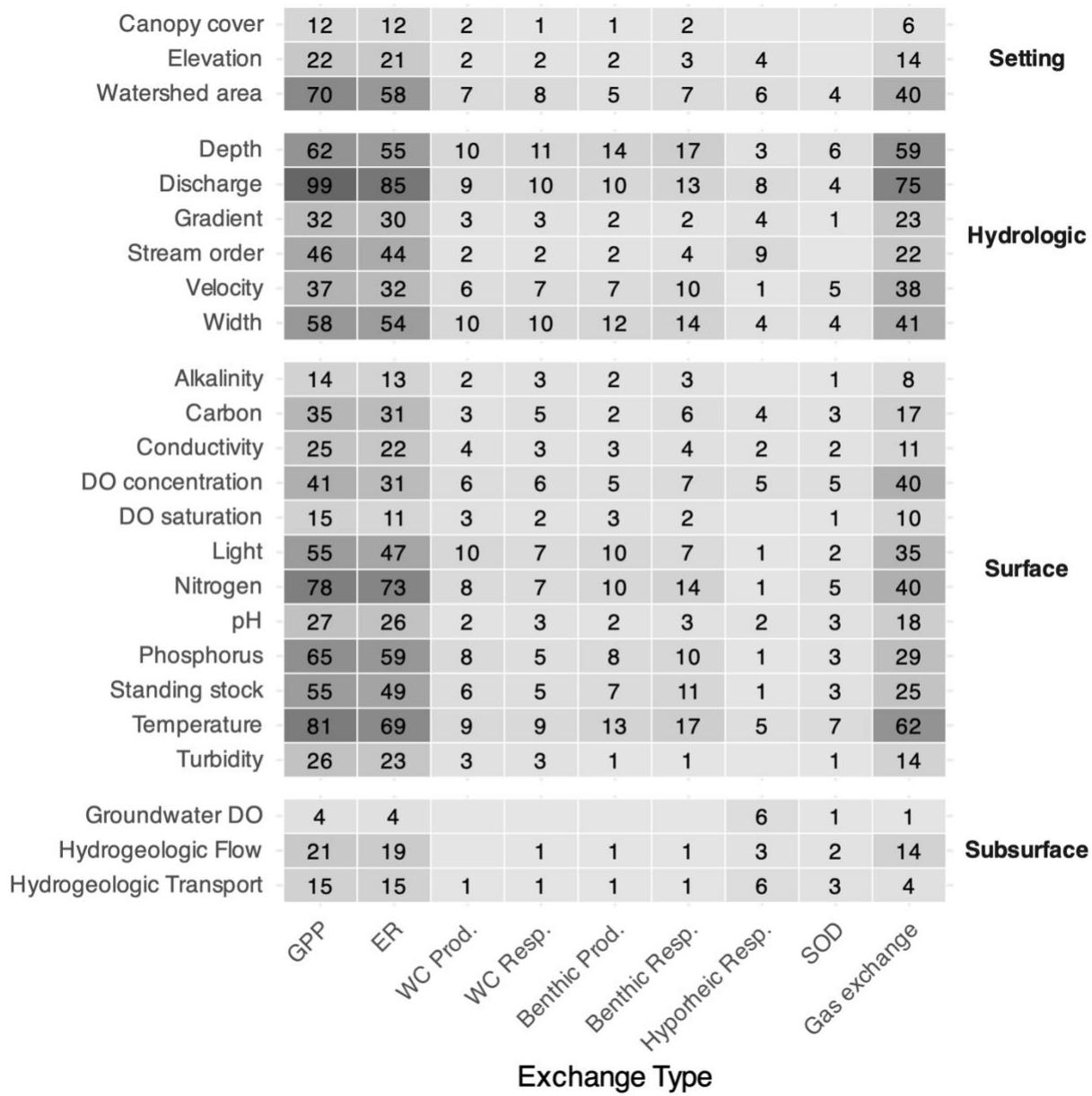


Figure S2. Heatmap showing the number of studies that reported each metadata variable, grouped into four categories (setting, hydrologic, surface, and subsurface) by measured flux type (gas exchange, GPP, ER, water column production and respiration, benthic production and respiration, sediment oxygen demand, and hyporheic respiration). Darker shading indicates a greater number of studies reporting that covariate.

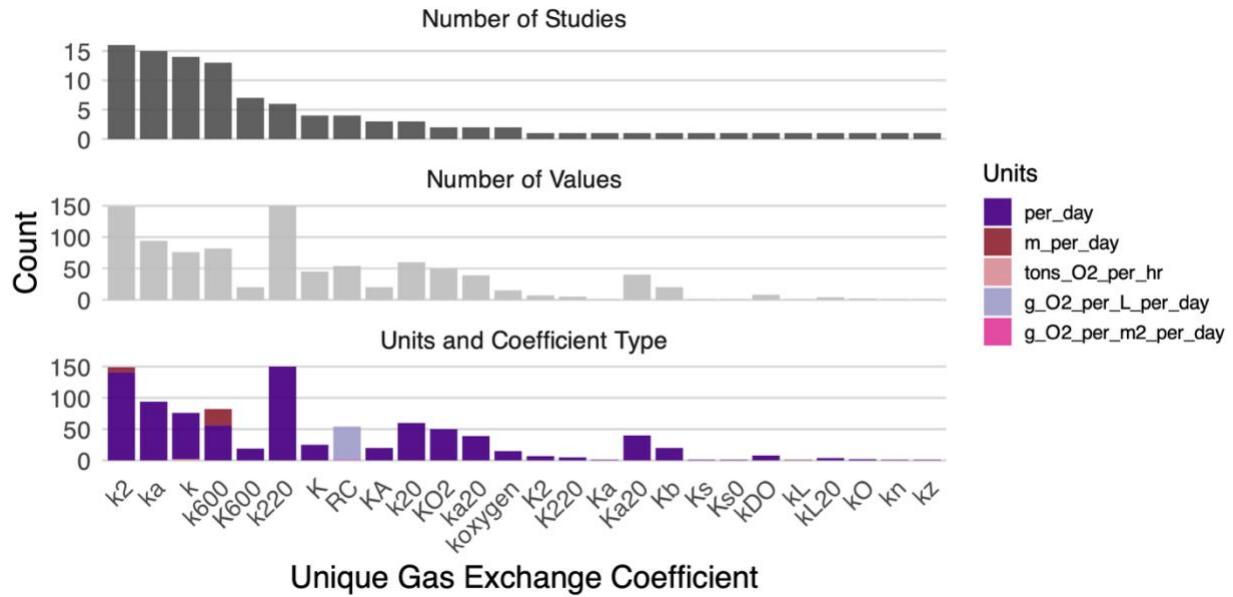


Figure S3. Diversity of gas exchange coefficient reporting. The top panel shows the number of studies reporting each coefficient type, the middle panel shows the total number of reported values, and the bottom panel shows the units associated with each coefficient type.

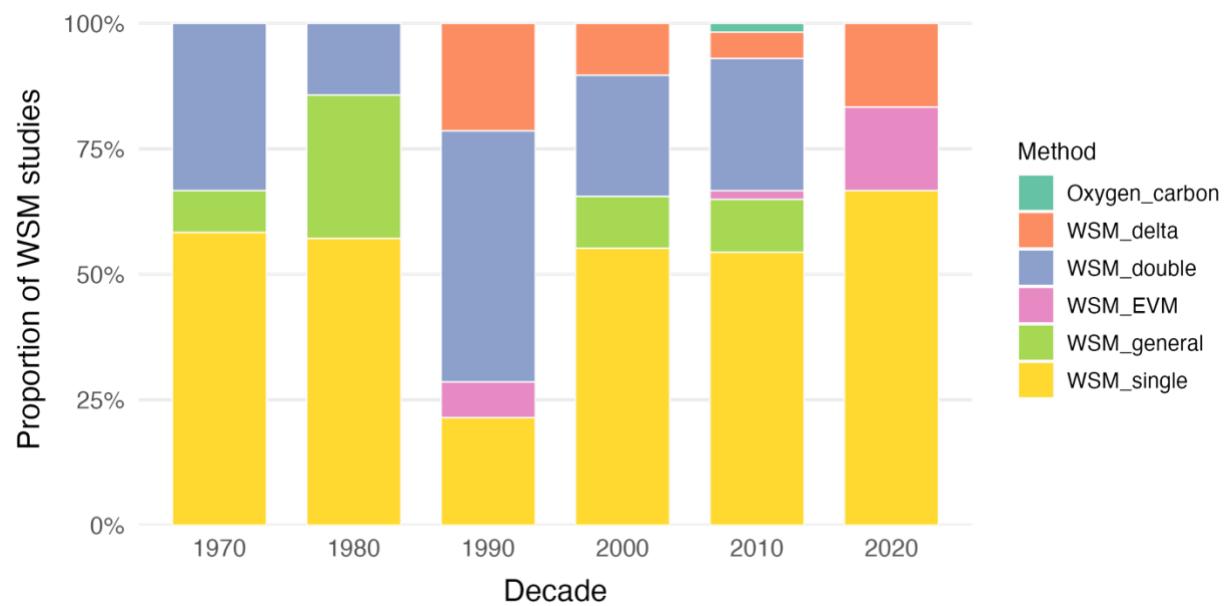


Figure S4. Temporal evolution of whole-stream metabolism (WSM) methods by decade. Stacked bars show the proportional use of single-station, two-station, delta, extreme value method, oxygen-carbon, and other WSM approaches.

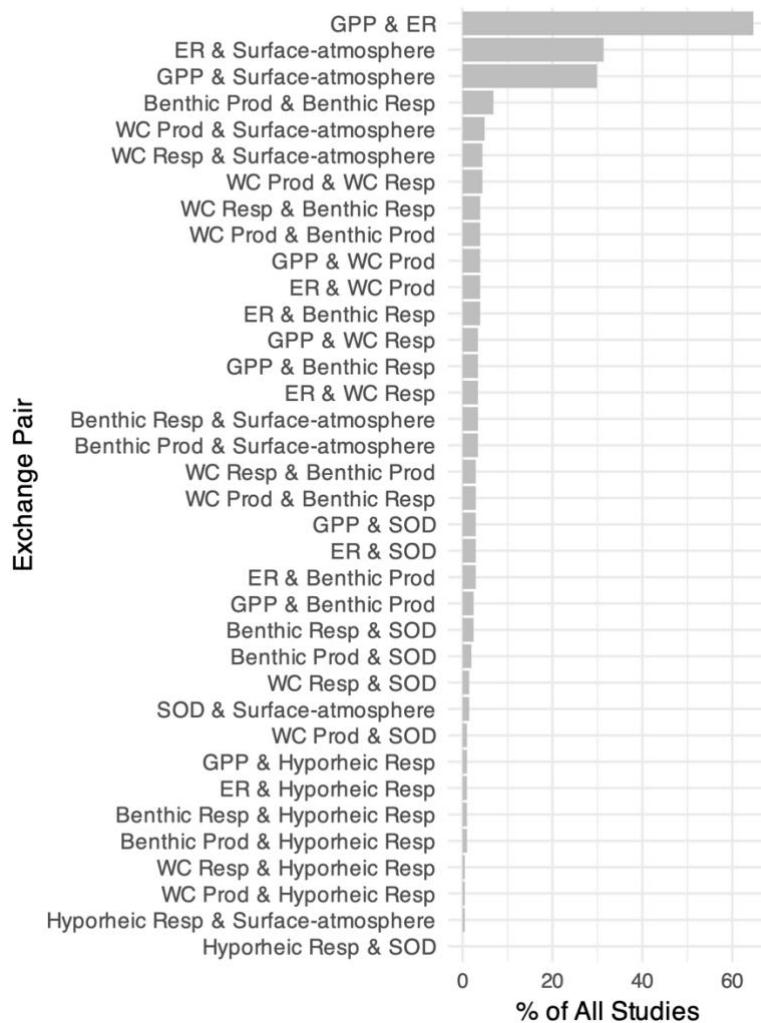


Figure S5. Co-occurrence of oxygen exchange metrics across reviewed across studies. Bars show the number of studies reporting each pair of exchange types, based on unique study-level co-reporting.

Table S1. Web of Science search string used to identify studies included in the oxygen synthesis.

Category	Search Terms
Ecosystem terms	stream; river; aquatic
Oxygen terms	dissolved oxygen; O <sub>2</sub>
Exchange terms	gas exchange; reaeration; prod*; photo*; metab*; respir*; hyporheic*; benth*
Exclusions	lake; sea; ocean; marine; marsh; wetland; swamp; pond; estuary

Search Field	Web of Science Topic Search (TS)
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Table S2. Summary statistics for gas exchange, volumetric metabolic rates, and areal oxygen fluxes standardized across all reviewed studies. Reported values include units, sample size (number of values), number of studies (number of studies), distributional statistics (mean, standard deviation or SD, minimum, and maximum), and log-transformed summaries for skewed metrics (mean<sub>ln</sub> and SD<sub>ln</sub>).

Exchange	Rate or Flux	Units	# of Values	# of Studies	Mean	SD	Min	Max	Mean (ln)	SD (ln)
<b>Whole-stream GPP</b>	Flux	g O <sub>2</sub> / m <sup>2</sup> / day	371	113	4.4	12.5	1.9E-04	2.1E+02	0.5	1.5
<b>Whole-stream GPP</b>	Rate	g O <sub>2</sub> / L / day	78	27	2.4	5.2	1.9E-04	2.5E+01	-3.0	3.3
<b>Whole-stream ER</b>	Flux	g O <sub>2</sub> / m <sup>2</sup> / day	364	108	8.0	15.5	1.6E-03	2.2E+02	1.4	1.3
<b>Whole-stream ER</b>	Rate	g O <sub>2</sub> / L / day	87	31	1.7	6.1	4E-04	3.1E+01	-4.2	2.4
<b>Water column production</b>	Flux	g O <sub>2</sub> / m <sup>2</sup> / day	19	10	2.6	3.8	2.6E-01	1.5E+01	0.3	1.0
<b>Water column production</b>	Rate	g O <sub>2</sub> / L / day	2	2	0.0	0.0	1.6E-03	2E-03	-6.3	0.1
<b>Water column respiration</b>	Flux	g O <sub>2</sub> / m <sup>2</sup> / day	22	10	1.0	0.8	4.8E-02	2.8E+00	-0.4	1.1
<b>Water column respiration</b>	Rate	g O <sub>2</sub> / L / day	39	4	0.2	0.4	8E-04	1.2E+00	-3.9	2.7
<b>Benthic production</b>	Flux	g O <sub>2</sub> / m <sup>2</sup> / day	25	14	2.5	2.4	5E-02	9.1E+00	0.3	1.4
<b>Benthic production</b>	Rate	g O <sub>2</sub> / L / day	3	3	3.9	4.3	4E-03	8.4E+00	-0.7	4.2
<b>Benthic respiration</b>	Flux	g O <sub>2</sub> / m <sup>2</sup> / day	49	21	11.4	17.2	3.5E-04	7E+01	0.9	2.4
<b>Benthic respiration</b>	Rate	g O <sub>2</sub> / L / day	5	2	0.3	0.7	1.4E-02	1.5E+00	-3.2	2.0
<b>Hyporheic respiration</b>	Flux	g O <sub>2</sub> / m <sup>2</sup> / day	6	5	7.4	9.8	7.8E-04	2.4E+01	-0.4	3.9
<b>Hyporheic respiration</b>	Rate	g O <sub>2</sub> / L / day	8	5	0.4	0.9	2E-03	2.5E+00	-2.6	2.1
<b>Sediment oxygen demand</b>	Flux	g O <sub>2</sub> / m <sup>2</sup> / day	54	13	1.3	1.7	2.9E-03	7E+00	-0.8	1.9
<b>Sediment oxygen demand</b>	Rate	g O <sub>2</sub> / L / day	2	2	0.0	0.0	1.3E-04	1.4E-02	-6.6	3.3
<b>Gas exchange (K<sub>600</sub>)</b>	Rate	1 / day	300	94	37.9	218.6	2.6E-03	3.6E+03	1.8	1.8

Table S3. Summary statistics for standardized gas exchange (K<sub>600</sub>) rates (day<sup>-1</sup>) and areal oxygen fluxes (g O<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup>) across stream size (small, mid, large, artificial) and land use (undisturbed, agricultural, urban, and mixed). Reported values include sample size (number of values), number of studies (number of studies), distributional statistics (mean, minimum, maximum), and log-transformed standard deviation (SD<sub>ln</sub>).

Group	Exchange	Units	# of Values	# of Studies	Mean	Min	Max	SD (ln)
<b>small</b>	Whole-stream GPP	g O <sub>2</sub> / m <sup>2</sup> / day	207	56	3.0	2E-04	7E+01	1.6
<b>small</b>	Whole-stream ER	g O <sub>2</sub> / m <sup>2</sup> / day	206	56	7.0	2E-03	1E+02	1.4

<b>small</b>	Water column production	g O <sub>2</sub> / m <sup>2</sup> / day	1	1	14.7	1E+01	1E+01	NA
<b>small</b>	Water column respiration	g O <sub>2</sub> / m <sup>2</sup> / day	1	1	2.7	3E+00	3E+00	NA
<b>small</b>	Benthic production	g O <sub>2</sub> / m <sup>2</sup> / day	4	3	2.7	5E-01	7E+00	1.1
<b>small</b>	Benthic respiration	g O <sub>2</sub> / m <sup>2</sup> / day	6	5	1.3	0E+00	5E+00	NA
<b>small</b>	Hyporheic respiration	g O <sub>2</sub> / m <sup>2</sup> / day	5	4	8.7	8E-04	2E+01	4.3
<b>small</b>	Sediment oxygen demand	g O <sub>2</sub> / m <sup>2</sup> / day	11	2	0.2	3E-03	7E-01	1.8
<b>small</b>	Gas exchange (K600)	1 / day	106	28	66.0	0E+00	4E+03	NA
<b>mid</b>	Whole-stream GPP	g O <sub>2</sub> / m <sup>2</sup> / day	126	51	6.5	1E-02	2E+02	1.5
<b>mid</b>	Whole-stream ER	g O <sub>2</sub> / m <sup>2</sup> / day	129	50	10.2	2E-02	2E+02	1.3
<b>mid</b>	Water column production	g O <sub>2</sub> / m <sup>2</sup> / day	14	8	1.8	0E+00	1E+01	NA
<b>mid</b>	Water column respiration	g O <sub>2</sub> / m <sup>2</sup> / day	18	8	0.7	5E-02	2E+00	1.0
<b>mid</b>	Benthic production	g O <sub>2</sub> / m <sup>2</sup> / day	12	7	2.6	5E-02	9E+00	1.5
<b>mid</b>	Benthic respiration	g O <sub>2</sub> / m <sup>2</sup> / day	35	13	15.4	4E-04	7E+01	2.5
<b>mid</b>	Hyporheic respiration	g O <sub>2</sub> / m <sup>2</sup> / day	1	1	0.8	8E-01	8E-01	NA
<b>mid</b>	Sediment oxygen demand	g O <sub>2</sub> / m <sup>2</sup> / day	35	6	1.6	0E+00	7E+00	NA
<b>mid</b>	Gas exchange (K600)	1 / day	81	3	7.0	0E+00	8E+01	NA
<b>large</b>	Whole-stream GPP	g O <sub>2</sub> / m <sup>2</sup> / day	28	16.	5.1	7E-01	2E+01	0.9
<b>large</b>	Whole-stream ER	g O <sub>2</sub> / m <sup>2</sup> / day	22	14	5.6	4E-01	2E+01	1.0
<b>large</b>	Water column production	g O <sub>2</sub> / m <sup>2</sup> / day	5	2	1.6	9E-01	2E+00	0.5
<b>large</b>	Water column respiration	g O <sub>2</sub> / m <sup>2</sup> / day	3	1	2.0	1E+00	3E+00	0.5
<b>large</b>	Benthic production	g O <sub>2</sub> / m <sup>2</sup> / day	9	4	2.3	1E-01	7E+00	1.5
<b>large</b>	Benthic respiration	g O <sub>2</sub> / m <sup>2</sup> / day	11	6	1.1	3E-04	7E+00	2.6
<b>large</b>	Sediment oxygen demand	g O <sub>2</sub> / m <sup>2</sup> / day	13	7	1.0	2E-02	3E+00	1.3
<b>large</b>	Gas exchange (K600)	1 / day	15	7	2.4	1E-01	7E+00	1.2
<b>artificial</b>	Whole-stream GPP	g O <sub>2</sub> / m <sup>2</sup> / day	9	6	5.9	6E-01	3E+01	1.1
<b>artificial</b>	Whole-stream ER	g O <sub>2</sub> / m <sup>2</sup> / day	8	6	7.4	6E-01	3E+01	1.1
<b>artificial</b>	Whole-stream ER	g O <sub>2</sub> / m <sup>2</sup> / day	6	4	2.4	5E-01	6E+00	0.9
<b>artificial</b>	Whole-stream ER	g O <sub>2</sub> / m <sup>2</sup> / day	5	4	3.2	5E-01	7E+00	1.2
<b>artificial</b>	Water column production	g O <sub>2</sub> / m <sup>2</sup> / day	1	1	0.6	6E-01	6E-01	NA
<b>artificial</b>	Sediment oxygen demand	g O <sub>2</sub> / m <sup>2</sup> / day	13		0.2	3E-03	2E+00	1.5
<b>artificial</b>	Sediment oxygen demand	g O <sub>2</sub> / m <sup>2</sup> / day	1	1	1.7	2E+00	2E+00	NA
<b>artificial</b>	Gas exchange (K600)	1 / day	6	4	3.7	1E+00	6E+00	0.6
<b>artificial</b>	Gas exchange (K600)	1 / day	5	3	3.3	1E+00	5E+00	0.6
<b>undisturbed</b>	Whole-stream GPP	g O <sub>2</sub> / m <sup>2</sup> / day	122	47	2.8	4E-03	7E+01	1.7
<b>undisturbed</b>	Whole-stream ER	g O <sub>2</sub> / m <sup>2</sup> / day	122	47	5.8	1E-02	1E+02	1.5
<b>undisturbed</b>	Water column production	g O <sub>2</sub> / m <sup>2</sup> / day	3	3	1.0	3E-01	2E+00	1.0

<b>undisturbed</b>	Water column respiration	g O <sub>2</sub> / m <sup>2</sup> / day	2	2	2.0	1E+00	3E+00	0.5
<b>undisturbed</b>	Benthic production	g O <sub>2</sub> / m <sup>2</sup> / day	2	2	1.7	1E-01	3E+00	2.3
<b>undisturbed</b>	Benthic respiration	g O <sub>2</sub> / m <sup>2</sup> / day	4	4	1.1	2E-01	2E+00	1.1
<b>undisturbed</b>	Hyporheic respiration	g O <sub>2</sub> / m <sup>2</sup> / day	2	2	0.1	8E-04	1E-01	3.5
<b>undisturbed</b>	Sediment oxygen demand	g O <sub>2</sub> / m <sup>2</sup> / day	2	2	0.3	2E-01	3E-01	0.3
<b>undisturbed</b>	Gas exchange (K600)	1 / day	48	19	127.9	3E-01	4E+03	1.9
<b>agricultural</b>	Whole-stream GPP	g O <sub>2</sub> / m <sup>2</sup> / day	79	32	7.1	2E-04	2E+02	1.8
<b>agricultural</b>	Whole-stream ER	g O <sub>2</sub> / m <sup>2</sup> / day	78	32	10.5	2E-03	2E+02	1.4
<b>agricultural</b>	Water column production	g O <sub>2</sub> / m <sup>2</sup> / day	6	4	3.4	3E-01	1E+01	1.4
<b>agricultural</b>	Water column respiration	g O <sub>2</sub> / m <sup>2</sup> / day	5	3	0.9	2E-01	3E+00	1.1
<b>agricultural</b>	Benthic production	g O <sub>2</sub> / m <sup>2</sup> / day	6	4	2.1	2E-01	7E+00	1.4
<b>agricultural</b>	Benthic respiration	g O <sub>2</sub> / m <sup>2</sup> / day	7	5	2.8	2E-01	7E+00	1.2
<b>agricultural</b>	Hyporheic respiration	g O <sub>2</sub> / m <sup>2</sup> / day	1	1	13.3	1E+01	1E+01	NA
<b>agricultural</b>	Sediment oxygen demand	g O <sub>2</sub> / m <sup>2</sup> / day	5	3	0.9	2E-01	3E+00	1.1
<b>agricultural</b>	Gas exchange (K600)	1 / day	50	17	16.4	0E+00	2E+02	NA
<b>urban</b>	Whole-stream GPP	g O <sub>2</sub> / m <sup>2</sup> / day	54	22	5.4	4E-01	3E+01	0.9
<b>urban</b>	Whole-stream ER	g O <sub>2</sub> / m <sup>2</sup> / day	53	21	8.8	1E+00	4E+01	0.8
<b>urban</b>	Water column production	g O <sub>2</sub> / m <sup>2</sup> / day	4	3	3.1	0E+00	1E+01	NA
<b>urban</b>	Water column respiration	g O <sub>2</sub> / m <sup>2</sup> / day	4.0	3	1.3	9E-01	2E+00	0.4
<b>urban</b>	Benthic production	g O <sub>2</sub> / m <sup>2</sup> / day	5	3	4.0	8E-01	9E+00	0.9
<b>urban</b>	Benthic respiration	g O <sub>2</sub> / m <sup>2</sup> / day	16	5	23.6	2E+00	7E+01	1.2
<b>urban</b>	Sediment oxygen demand	g O <sub>2</sub> / m <sup>2</sup> / day	26	7	1.3	3E-01	3E+00	0.6
<b>urban</b>	Gas exchange (K600)	1 / day	29	14	5.4	3E-01	3E+01	1.2
<b>mixed</b>	Whole-stream GPP	g O <sub>2</sub> / m <sup>2</sup> / day	73	34	4.4	1E-01	2E+01	1.1
<b>mixed</b>	Whole-stream ER	g O <sub>2</sub> / m <sup>2</sup> / day	73	33	9.2	2E-02	1E+02	1.3
<b>mixed</b>	Water column production	g O <sub>2</sub> / m <sup>2</sup> / day	7	4	1.8	9E-01	3E+00	0.5
<b>mixed</b>	Water column respiration	g O <sub>2</sub> / m <sup>2</sup> / day	11	4	0.7	5E-02	2E+00	1.2
<b>mixed</b>	Benthic production	g O <sub>2</sub> / m <sup>2</sup> / day		5	1.9	5E-02	4E+00	1.6
<b>mixed</b>	Benthic respiration	g O <sub>2</sub> / m <sup>2</sup> / day	20	8	7.9	2E-02	3E+01	2.2
<b>mixed</b>	Hyporheic respiration	g O <sub>2</sub> / m <sup>2</sup> / day	1	1	0.8	8E-01	8E-01	NA
<b>mixed</b>	Sediment oxygen demand	g O <sub>2</sub> / m <sup>2</sup> / day	14	4	2.3	0E+00	7E+00	NA
<b>mixed</b>	Gas exchange (K600)	1 / day	41	15	8.2	6E-01	6E+01	1.2