

1                   **This is a non-peer review preprint submitted to EarthArXiv**

2  
3  
4                   Radiolysis *via* radioactivity is not responsible for  
5                   rapid methane oxidation in subterranean air  
6

7  
8   Arndt Schimmelmann<sup>1\*¶</sup>, Angel Fernandez-Cortes<sup>2,3¶</sup>, Soledad Cuezva<sup>2¶</sup>, Thomas Streil<sup>4&</sup>,  
9   Jay T. Lennon<sup>5&</sup>

10  
11  
12  
13   <sup>1</sup> Department of Earth and Atmospheric Sciences, Indiana University, Bloomington, Indiana,  
14   USA

15  
16   <sup>2</sup> Department of Biology and Geology, University of Almeria, Almeria, Spain

17  
18   <sup>3</sup> Department of Earth Sciences, Royal Holloway, University of London, Egham, Surrey TW20  
19   0EX, United Kingdom

20  
21   <sup>4</sup> SARAD GmbH, Dresden, Germany

22  
23   <sup>5</sup> Department of Biology, Indiana University, Bloomington, Indiana, USA  
24

25  
26   \* Corresponding author

27 E-mail: [aschimme@indiana.edu](mailto:aschimme@indiana.edu) (AS)

28

29

30 ¶ These authors contributed equally to this work.

31 & These authors also contributed equally to this work.

32

### 33 **Abstract**

34 Atmospheric methane is rapidly lost when it enters humid subterranean critical and vadose zones  
35 (e.g., air in soils and caves). Because methane is a source of carbon and energy, it can be consumed  
36 by methanotrophic methane-oxidizing bacteria. As an additional subterranean sink, it has been  
37 hypothesized that methane is oxidized by natural radioactivity-induced radiolysis that produces  
38 energetic ions and radicals, which then trigger abiotic oxidation and consumption of methane  
39 within a few hours. Using controlled laboratory experiments, we tested whether radiolysis could  
40 rapidly oxidize methane in sealed air with different relative humidities while being exposed to  
41 elevated levels of radiation (more than  $535 \text{ kBq m}^{-3}$ ) from radon isotopes  $^{222}\text{Rn}$  and  $^{220}\text{Rn}$  (i.e.,  
42 thoron). We found no evidence that radiolysis contributed to methane oxidation. In contrast, we  
43 observed the rapid loss of methane when moist soil was added to the same apparatus in the absence  
44 of elevated radon abundance. Together, our findings are consistent with the view that methane  
45 oxidizing bacteria are responsible for the widespread observations of methane depletion in  
46 subterranean environments. Further studies are needed on the ability of microbes to consume trace  
47 amounts of methane in poorly ventilated caves, even though the trophic and energetic benefits  
48 become marginal at very low partial pressures of methane.

49

50

### 51 **Introduction**

52 Energetic radiation generates ions and radicals in fluids *via* radiolysis that can trigger  
53 subsequent chemical reactions [1], including the oxidation of organics. Radiolysis has likely  
54 affected the evolution of early microbial metabolisms and is crucial for powering the deep  
55 microbial biosphere [2, 3]. However, few studies have addressed the quantitative importance of  
56 radiolysis for contemporary fluxes in the atmosphere and the critical zone, especially in  
57 comparison to processes that compete with biologically mediated transformations.

58 The concentration of methane ( $\text{CH}_4$ ) in the atmosphere has more than doubled since 1850  
59 to  $\sim 1.85$  ppmv (i.e., parts per million by volume) and now contributes  $\sim 15$  % of anthropogenic  
60 forcing of climate change [4]. The Intergovernmental Panel on Climate Change (IPCC) report [5]  
61 includes secondary greenhouse warming effects of  $\text{CH}_4$  and arrives at  $1 \text{ W m}^{-2}$  for  $\text{CH}_4$  relative to

62 1.7 W m<sup>-2</sup> for CO<sub>2</sub>, making CH<sub>4</sub> the second most important anthropogenic climate forcing agent.  
63 In the atmosphere, the removal of CH<sub>4</sub> is due primarily to oxidation *via* photochemically generated  
64 tropospheric OH• radicals [6, and refs. therein]. In spite of intense radiation in the atmosphere  
65 from sun and space, the residence time of atmospheric CH<sub>4</sub> is ~12 years. The second largest sink  
66 for atmospheric CH<sub>4</sub> is shallow subterranean environments containing aerated soils that are  
67 inhabited by CH<sub>4</sub> oxidizing bacteria, MOB [4, 7]. MOB are also found in deeper aerated  
68 subterranean environments, such as caves in the vadose zone, although their contribution to global  
69 CH<sub>4</sub> cycling has not been quantified or incorporated into earth system models [4].

70 A growing number of studies have reported that, throughout the world, concentrations of  
71 CH<sub>4</sub> are often depleted in the air of caves suggesting that subterranean environments may represent  
72 an overlooked sink for atmospheric CH<sub>4</sub> [e.g., 8-13]. Based on ventilation rates and CH<sub>4</sub> pools, it  
73 is estimated CH<sub>4</sub> is rapidly consumed in caves on time scales ranging from hours to days [14, 15].  
74 Depletion of CH<sub>4</sub> in caves is often attributed to MOB. However, a study from Spanish caves  
75 proposed that rapid CH<sub>4</sub> oxidation may be attributed to non-biological processes *via* radiolysis and  
76 ionization of subterranean air by natural radioactivity that could lead to the oxidation of CH<sub>4</sub> at a  
77 sufficiently fast rate to account for appreciable consumption of CH<sub>4</sub> [10]. It has been proposed that  
78 α-radiation (e.g., from <sup>222</sup>Rn) can radiolytically ionize, or generate radicals from, atmospheric  
79 components (e.g., H<sub>2</sub>O) including CH<sub>4</sub> [16, 17]. The study by Haynes and Kebarle [16] determined  
80 that α-radiation has a slow effect on pure CH<sub>4</sub> and mixed hydrocarbon gas in the absence of air,  
81 making it difficult to extrapolate results to CH<sub>4</sub> in air in the presence of ions and radicals from  
82 heteromolecules.

83 Some studies, however, have raised questions about the relative importance of abiotic CH<sub>4</sub>  
84 oxidation based on theoretical considerations of kinetics, the inability of α-radiation from metallic  
85 uranium and radon to trigger fast oxidation of CH<sub>4</sub> [15, 18]. Laboratory and field experiments  
86 implicated MOB with the rapid decline in cave CH<sub>4</sub> concentrations [18], while isotopically  
87 uncharacterized radon was unable to remove CH<sub>4</sub> from air in an Australian cave [15]. Studies on  
88 radon typically focus on <sup>222</sup>Rn because its longer half-life of 3.83 days facilitates quantification.  
89 No study has yet examined the radiolytic effect on CH<sub>4</sub> oxidation of the relatively more energetic  
90 decay of <sup>220</sup>Rn (called thoron, with a half-life 55.6 s), particularly in the air close to cave walls and  
91 floors where <sup>220</sup>Rn is relatively more abundant. Also, direct experiments linking the constraints of  
92 air humidity and natural radiation from specific radon isotopes to CH<sub>4</sub> oxidation in air are lacking.  
93 The current study fills these gaps with detailed independent experiments in two laboratories using  
94 energetically distinct radiation levels from isotopes of radon (<sup>222</sup>Rn and <sup>220</sup>Rn) at different  
95 humidities and contrasting the results with CH<sub>4</sub>-depletion by MOB.

96

## 97 **Materials and methods**

98 The authors of this study belong to two teams that had no knowledge of each others'  
99 experiments at Indiana University (IU) and Royal Holloway University of London (RHUL). After  
100 completion of all experiments, the two groups decided to jointly report their complementary  
101 results. Work at IU afforded superior analytical control on radon isotopes and could accurately  
102 measure higher dose rates, whereas the more gas-tight experimental setup at RHUL provided more  
103 straightforward evidence for the inability of natural radiation levels to rapidly oxidize atmospheric  
104 CH<sub>4</sub> at its natural atmospheric abundance.

105 Details of materials and methods are available from *protocols.io* under  
106 <http://dx.doi.org/10.17504/protocols.io.s7aehie>. We employed two separate, complementary  
107 experimental approaches at IU and RHUL. The following two sections offer brief overviews.

## 108 **Apparatus at IU for active, time-resolved measurements of gas** 109 **concentrations with circular flow**

110 At IU, we constructed an experimental apparatus to assess the loss of CH<sub>4</sub> in an active (i.e.,  
111 with pumping of air) and time-resolved manner with or without added radiation from radon  
112 isotopes and their progeny (Fig 1). The use of pumping qualifies this method as active and time-  
113 resolved in contrast to passive measurements of radon that integrate over time [19]. Approximately  
114 6 L of air was recirculated in the sealed apparatus that included (i) a glass tube with optional  
115 thorium carbonate to generate <sup>220</sup>Rn (also called thoron), and (ii) a glass tube containing uranium  
116 ore to generate <sup>222</sup>Rn, with an overlying layer of coconut charcoal to limit the escape of co-  
117 produced, short-lived <sup>220</sup>Rn. Blank experiments without elevated radiation identified a  
118 reproducible loss of CH<sub>4</sub> (likely by diffusion through polymer tubing within the sealed analytical  
119 SARAD<sup>®</sup> RTM 2200 instrument) that was subtracted from all other experiments at IU to arrive at  
120 net CH<sub>4</sub> losses that are due to other factors, such as radiolysis or microbial methanotrophy.

121

122 **Fig 1. Experimental setup for time-resolved measurements of gas concentrations with**  
123 **circular flow at Indiana University.** Approximately 6 L of air was recirculated in a sealed  
124 apparatus to assess the loss of CH<sub>4</sub> with or without added radiation from radon isotopes and their  
125 radioactive progeny. At the beginning of each experiment, the trapped air was slightly enriched in  
126 CH<sub>4</sub> (and CO<sub>2</sub>, except for experiments with soils), followed by hourly measurements of gas  
127 concentrations over a few days to weeks. Radon <sup>222</sup>Rn was generated by uranium ore while  
128 charcoal retained <sup>220</sup>Rn (a). The air intake of the 3-neck 5-L glass flask was directed to the bottom  
129 of the flask with a plastic insert to facilitate the mixing of air (b); thorium carbonate is not shown.  
130 Depicted components of the apparatus (c) are not drawn to scale.

131

132 We conducted a number of experiments at IU to assess the importance of  $\alpha$ -radiation  
133 intensity, relative humidity, and the presence or absence of soil on CH<sub>4</sub> dynamics. Moisture is  
134 critical for the emanation efficiency of radon isotopes from solid sources (i.e., the escape of noble  
135 gas radon atoms from the interior of minerals into H<sub>2</sub>O-containing pore space *via* recoil subsequent  
136 to radioactive decay of parental nuclides; e.g., Markkanen and Arvela [20]; Morawska and Phillips  
137 [21]) and for stabilizing ions and radicals in air. Individual experiments differed in terms of their  
138 optional use of elevated humidity, thorium carbonate, and gas flowing through the tube containing  
139 uranium ore. The trapped ~6-L volume of air was initially spiked with CH<sub>4</sub> from natural gas to  
140 ~70 ppmv and with CO<sub>2</sub> to ~5,000 ppmv (except for experiments with soils) to distinguish it from  
141 room air and to increase the analytical precision during the time-series of measurements that lasted  
142 over a few days to weeks. Elevated CO<sub>2</sub> concentrations are typical for many cave environments  
143 [10].

144 Most experiments at IU discriminated between  $\alpha$ -radiation from radon <sup>222</sup>Rn *versus*  
145 thoron <sup>220</sup>Rn. Whereas radon <sup>222</sup>Rn with a half-life of 3.83 days is relatively homogeneously  
146 distributed in cave air (also in our apparatus), the much shorter lived thoron <sup>220</sup>Rn with a half-life

147 of only 55.6 s [22] cannot travel far from its parent nuclei residing in minerals [19], thus thoron's  
148 highest concentrations in cave air are near cave walls and the floor. The higher  $\alpha$ -decay energy of  
149  $^{220}\text{Rn}$  (6.3 MeV) relative to  $^{222}\text{Rn}$  (5.49 MeV) prompted us to design experiments for separate  
150 examinations of the ability of both radon isotopes to trigger the oxidation of  $\text{CH}_4$ . The more  
151 energetic  $\alpha$ -decay of thoron  $^{220}\text{Rn}$  should ionize air more efficiently than  $^{222}\text{Rn}$ . Thoron was  
152 generated from thorium carbonate that was optionally loaded into a glass tube attached to the  
153 round-bottom flask. In other experiments,  $^{222}\text{Rn}$  decay measuring up to 327  $\text{kBq m}^{-3}$  was  
154 produced *in-situ* in the glass apparatus by uranium ore chips (Fig 1). Escape of co-produced  
155 thoron from ore was reduced by using a layer of coconut charcoal in the upper part of the glass  
156 tube as a filter [23]. The resulting adsorption of  $^{220}\text{Rn}$  on charcoal increased the residence time in  
157 the glass tube and let  $^{220}\text{Rn}$  decay before it could enter the 5-L glass flask.

158 We quantified the concentrations of  $^{222}\text{Rn}$ ,  $^{220}\text{Rn}$ ,  $\text{CH}_4$  and  $\text{CO}_2$  during experiments at IU  
159 at an air flow rate of  $\sim 0.2 \text{ L min}^{-1}$  once every hour while operating the diffusion pump in the  
160 SARAD<sup>®</sup> RTM 2200.  $^{220}\text{Rn}$  radiation intensity was either measured *via*  $\alpha$ -spectroscopy at a faster  
161 flow rate of  $1 \text{ L min}^{-1}$  in 10-min increments ( $n \geq 10$ ) while temporarily operating the more powerful  
162 membrane pump, or values from flow rates  $\leq 0.2 \text{ L min}^{-1}$  with the diffusion pump were doubled to  
163 adjust for fast  $^{220}\text{Rn}$  decay (see supporting information for detailed control experiments and  
164 graphed data). Elevated relative humidity fosters the stabilization of ions in air *via* attachment to  
165 clusters of water molecules and may enhance the ability of ions to trigger oxidative degradation of  
166  $\text{CH}_4$  (discussed by Fernandez-Cortes et al. [10]). Therefore, at IU we recorded humidity in the  
167 apparatus along with temperature, air pressure, flow, and battery voltage on an hourly basis. The  
168 accuracy of data from the SARAD<sup>®</sup> RTM 2200 was independently evaluated *via* direct comparison  
169 with a newly manufactured and factory-calibrated Thoron Scout instrument (SARAD<sup>®</sup> GmbH,  
170 Dresden, Germany; details available in supporting information).

171 We conducted a number of experiments at IU to test for the effects of radiation and  
172 microbial activity on  $\text{CH}_4$  dynamics in our experimental apparatus. Multi-day time-series of data  
173 were collected in closed-circuit air reflux mode (i) as duplicated blank experiments without added  
174 radon or thoron, (ii) with enhanced  $^{220}\text{Rn}$  concentration in dry or moist air, (iii) with enhanced  
175  $^{222}\text{Rn}$  concentration in dry or moist air, and (iv) with jointly enhanced  $^{220}\text{Rn}$  and  $^{222}\text{Rn}$   
176 concentrations in moist air to depict an extreme scenario where cave air had a highly elevated  $\alpha$ -  
177 radiation level. Furthermore, (v) we tested for  $\text{CH}_4$  oxidation after placing moist soils, which we  
178 assumed contained methanotrophic bacteria (MOB), into the 5-L glass flask, without elevated  
179 radioactivity. Certain impurities in industrially conditioned natural gas may act as MOB inhibitors,  
180 for example acetylene and carbon monoxide [24, p. 335]. As a precaution, the  $\text{CH}_4$  spikes in  
181 experiments employing two different soils were derived from gas that was collected from a natural  
182 seepage of shale gas in New York State [25]. Natural shale gas is not known to contain acetylene  
183 or carbon monoxide.

## 184 **Gas-tight terrarium experiments at RHUL**

185 Experiments at RHUL at atmospheric  $\text{CH}_4$  abundance used a gas-tight glass terrarium (i.e.,  
186 an aquarium without water holding a volume of 13.45 L; Fig 2) with a hermetically sealing glass  
187 lid. Two air-tight gas ports allowed the withdrawal of 1-L air samples into Tedlar<sup>®</sup> bags without  
188 changes in atmospheric pressure. Fragments of uraninite-bearing pitchblende served as a source  
189 of radioactivity. An AlphaLab Air Ion Counter with an integrated fan was placed in the terrarium

190 to measure the abundance of ions in air in 30-s intervals. The  $\alpha$ -radiation was quantified on 1-h  
191 intervals with a Canary Pro monitor (Airthings, Oslo, Norway) *via*  $\alpha$ -spectrometry. Gas samples  
192 in Tedlar<sup>®</sup> bags were analyzed for CH<sub>4</sub> mole fractions with a Picarro G1301 CRDS (Cavity Ring-  
193 Down Spectrometer, Picarro Inc., Santa Clara, California, USA).

194

195 **Fig 2. Gas-tight terrarium experiments at Royal Holloway University of London.** (a) A  
196 hermetically sealed glass terrarium was filled with laboratory air containing atmospheric CH<sub>4</sub>. A  
197 radon monitor provided data on <sup>222</sup>Rn abundance, while an AlphaLab Air Ion Counter measured  
198 the concentration of negative ions. After ~6 h into the RHUL experiment #1, the placement of a  
199 beaker filled with deionized, warm water elevated the relative humidity to > 85 %. At the same  
200 time, two fragments of pitchblende (containing uraninite as a radiation source) were placed into  
201 the terrarium to generate <sup>222</sup>Rn. Tedlar<sup>®</sup> bags in the terrarium are not shown in the photograph. (b)  
202 Diagram of the sampling procedure to collect <1-L aliquots of air from the terrarium in RHUL  
203 experiment #2. This experiment lasted for 76 h and 50 min and reached a <sup>222</sup>Rn-based radiation  
204 level in excess of 50 kBq m<sup>-3</sup> after 5 h.

205

206 The initial RHUL experiment #1 (Fig 2a) assessed the production of negative ions and the  
207 abundance of <sup>222</sup>Rn over ~6 h (i.e., stage 1) without either pitchblende or a beaker with water in  
208 the terrarium that had been flushed initially with laboratory air, and subsequently for ~15 h in the  
209 presence of pitchblende and a beaker with 130 mL of 38 °C warm water in the terrarium (stage 2).

210 The subsequent RHUL experiment #2 (Fig 2b) in the same terrarium included monitoring  
211 of the CH<sub>4</sub> mole fraction of laboratory air sealed in the terrarium where pitchblende and a beaker  
212 with 130 mL of water (initially at 38 °C) had been placed to provide for elevated radioactivity and  
213 relative humidity. Elevated relative humidity was needed to simulate cave conditions. The  
214 AlphaLab Air Ion Counter failed to provide useful data due to static interference with Tedlar<sup>®</sup>  
215 bags. The second experiment lasted for 76 h and 50 min and reached a <sup>222</sup>Rn-based radiation level  
216 in excess of 50 kBq m<sup>-3</sup> after 5 h. Approximate 1-L aliquots of air sampled from the terrarium were  
217 analytically compared with aliquots of exterior laboratory air on four occasions.

218

## 219 **Results and discussion**

### 220 **Active time-series measurements with circular flow at IU**

221 Our controlled experiments with and without <sup>220</sup>Rn and/or <sup>222</sup>Rn were designed to directly  
222 test whether or not radiation can oxidize CH<sub>4</sub> in cave air on ecologically relevant time scales (i.e.,  
223 hours to days). We relied on comparisons of CH<sub>4</sub> inventories in experiments with (i) high radiation  
224 intensity from *in-situ* generated <sup>220</sup>Rn and/or <sup>222</sup>Rn with those from (ii) duplicate blank experiments  
225 with no artificially enhanced radiation to demonstrate the sensitivity of our setup to detect CH<sub>4</sub>-  
226 losses. In addition, we conducted (iii) two experiments with moist soils in the absence of added  
227 radon isotopes to assess the potential for environmental microorganisms (i.e., MOB) to remove  
228 CH<sub>4</sub> as has been demonstrated elsewhere by members of our research team [14, 18]. The

229 comparisons among experiments covered a common range of CH<sub>4</sub> concentration and thus only  
230 differed in the lengths of their time windows needed to lower the CH<sub>4</sub> concentration from the upper  
231 to the lower threshold (i.e. yellow rectangle in Fig 3a). The ‘common window’ of CH<sub>4</sub> decline for  
232 all 11 experiments maximized the data available for comparison.

233

234 **Fig 3. Time-series experiments at Indiana University.** Results from time-series experiments to  
235 test for abiotic (radiation) and biotic (CH<sub>4</sub> oxidizing bacteria) factors on CH<sub>4</sub> dynamics. **(a)** The  
236 decline of CH<sub>4</sub> concentrations in experiments #1 to #9 (without soil) followed similar trajectories  
237 depending on original concentrations, despite major differences in radiation intensity (see the  
238 yellow rectangle that identifies a window of CH<sub>4</sub> concentrations that is common to all  
239 experiments). Declining CH<sub>4</sub> concentrations are independent of the intensity of  $\alpha$ -radiation. Blank  
240 experiments #1 to #4 without elevated radiation identify a reproducible loss of CH<sub>4</sub> by diffusion  
241 that was subtracted from all other experiments to arrive at net losses that are due to other factors,  
242 such as radiolysis or microbial methanotrophy. **(b)** The decline of CO<sub>2</sub> concentrations in a range  
243 of experiments without soil followed similar patterns. In addition to loss due to diffusion through  
244 plastic, it was likely influenced by adsorption, solution in water, or possible chemical uptake. In  
245 addition, CO<sub>2</sub> was generated from moist soils in experiments #10 and #11. **(c)** Radon <sup>222</sup>Rn and  
246 **(d)** thoron <sup>220</sup>Rn concentrations partially depended on relative humidity; soil no. 2 in experiment  
247 #11 generated low levels of <sup>222</sup>Rn over time presumably due to traces of uranium in minerals; <sup>220</sup>Rn  
248 concentrations are original data from low flow rates at <0.2 L min<sup>-1</sup> when laminar flow conditions  
249 in the 5-L glass flask caused heterogeneity and occasional spikes. **(e)** Noise in relative humidity  
250 data partially derived from the automatic battery recharge cycle that influenced the internal  
251 temperature of the SARAD<sup>®</sup> RTM 2200 and the algorithm to calculate humidity. **(f)** Experiments  
252 #10 and #11 with moist soils without added <sup>222</sup>Rn or <sup>220</sup>Rn resulted in a long-term exponential  
253 decline of CH<sub>4</sub> concentrations while CO<sub>2</sub> was generated biologically.

254

255 Multiple trials in our experimental apparatus revealed that CH<sub>4</sub> dynamics were unaffected  
256 by radiation within the precision of measurements. Repeat blank experiments with dry  
257 (experiments #1 and #2) or moist air (experiments #3 and #4) without artificially elevated radon  
258 or thoron concentrations resulted in reproducible and systematic small losses of both CH<sub>4</sub> and CO<sub>2</sub>  
259 over time (Fig 3a, b; Table 1A). Although radon isotopes, CH<sub>4</sub> and CO<sub>2</sub> could not diffuse through  
260 glass and metal in our apparatus, the SARAD<sup>®</sup> RTM 2200 and its Axetris<sup>®</sup> laser OEM Module  
261 LGC F200 methane detector were internally and externally connected to glass and metal  
262 components with short segments of various types of clear polymer tubing (Fig 1) that resulted in  
263 slow losses *via* gas diffusion through polymers. The rate of diffusion across a layer of polymer is  
264 dependent on the difference in partial pressures between the interior and exterior air, and hence  
265 the rates of CH<sub>4</sub> and CO<sub>2</sub> losses *via* diffusion over time follow curves that asymptotically approach  
266 equilibria (Fig 3f). At a CH<sub>4</sub> concentration of ~59 ppmv in the apparatus (i.e., the midpoint of the  
267 common CH<sub>4</sub> range; Fig 3a) and outside air with ~1.85 ppmv, the mean CH<sub>4</sub> diffusive loss from  
268 air in the apparatus during blank experiments #1 to #4 consistently amounted to 0.39 ppmv h<sup>-1</sup>  
269 regardless of humidity and small variations in room temperature and air pressure (Table 1A;  
270 supporting information). Such a loss of gas over time could theoretically result from a small  
271 internal leak in the system. However, the non-parallel pattern of CO<sub>2</sub> losses in blank experiments  
272 (Fig 3b) is inconsistent with a leak and instead argues for varying diffusivity of the polar molecule

273 CO<sub>2</sub> through permeable material at different humidities. The observed degree of CH<sub>4</sub> loss from  
274 the system was unavoidable and had to be subtracted from the observed bulk CH<sub>4</sub> losses in  
275 experiments with enhanced radiation and soils to arrive at any specific losses that are due to  
276 radioactivity or presumed microbial methanotrophy.

277

278 **Table 1. Overview on individual experiments performed at Indiana University to constrain**  
279 **the consumption of methane over time.**



Experiment #	Overall duration (h)	Common window (h)	Mean values across common window of CH <sub>4</sub> decrease from 67.2 to 50.9 ppmv						
			<sup>220</sup> Rn (kBq m <sup>-3</sup> )	<sup>222</sup> Rn (kBq m <sup>-3</sup> )	CH <sub>4</sub> loss (ppmv h <sup>-1</sup> )	CO <sub>2</sub> loss (ppmv h <sup>-1</sup> )	Corrected temperature (°C)	Relative humidity (%)	Pressure (kPa)
<i>(A) Duplicated blank experiments without added radon or thoron at low or high relative humidity</i>									
#1, no added radiation, dry	86	42.0	<0.01*	<0.01	0.39 ± 0.06	25.8 ± 14.1	23.2 ± 0.9	49.7 ± 1.5	97.2 ± 0.4
#2, no added radiation, dry	52	42.7	<0.01*	<0.01	0.39 ± 0.07	47.6 ± 21.3	22.8 ± 0.4	54.7 ± 0.8	98.0 ± 0.1
#3, no added radiation, wet	69	42.5	<0.2*	~0.4	0.39 ± 0.08	22.9 ± 23.0	23.4 ± 0.7	103 ± 2.7**	98.1 ± 0.2
#4, no added radiation, wet	43	42.3	<0.3*	<0.2	0.39 ± 0.07	n.d.	25.4 ± 0.7	95.3 ± 4.1	99.1 ± 0.1
<i>(B) Experiments with enhanced <sup>220</sup>Rn and/or <sup>222</sup>Rn concentrations at low or high relative humidity</i>									
#5, <sup>220</sup> Rn added, dry	82	42.4	~50*	<0.01	0.38 ± 0.07	18.4 ± 14.2	22.9 ± 0.6	53.3 ± 2.7	98.6 ± 0.2
#6, <sup>220</sup> Rn added, wet	75	41.8	143	<0.01	0.39 ± 0.07	10.6 ± 15.0	22.8 ± 0.7	100.4 ± 2.1	98.9 ± 0.1
#7, <sup>222</sup> Rn added, dry	87	43.8	<0.01*	153	0.37 ± 0.07	34.7 ± 35.3	24.3 ± 0.8	56.3 ± 2.2	98.8 ± 0.2
#8, <sup>222</sup> Rn added, wet	88	44.5	<0.04*	216	0.37 ± 0.09	22.0 ± 29.9	25.1 ± 0.9	95.5 ± 4.1	98.4 ± 0.3
#9, <sup>222</sup> Rn and <sup>220</sup> Rn added, wet	87	43.5	208	327	0.37 ± 0.11	10.3 ± 27.9	25.8 ± 1.1	92.8 ± 3.4	98.4 ± 0.2
<i>(C) Experiments with moist soils without added radon or thoron</i>									
#10, no added radiation, 45 g soil #1,	456	33.8	<0.03*	<0.2	0.49 ± 0.09	CO <sub>2</sub> was generated	24.3 ± 0.8	100.6 ± 3.9	98.6 ± 0.2
#11, no added radiation, 112 g soil #2	172	34.7	~0.1*	~4	0.47 ± 0.11	CO <sub>2</sub> was generated	24.7 ± 0.9	97.9 ± 4.3	98.8 ± 0.1

280 Analytical data are mean values with standard deviations for the time window when CH<sub>4</sub> concentrations declined from 67.2 to 50.9 ppmv in each experiment.  
281 This window represents the maximum range of methane concentrations that is common to all experiments. The length of time needed to deplete methane from  
282 the upper to the lower threshold across the common window (i.e., yellow rectangle in Fig 3a) was interpolated from hourly spaced data.

283 \*  $^{220}\text{Rn}$  radiation values were doubled when measured at flow rates  $\leq 0.2 \text{ L min}^{-1}$  to adjust for fast  $^{220}\text{Rn}$  decay, instead of measurements without doubling of  
284 values in 10-min intervals ( $n \geq 10$ ) at a flow rate of  $1 \text{ L min}^{-1}$ . See justification in supporting information.

285 \*\* High values of relative humidity are affected by analytical errors in excess of standard deviation.

286 n.d. = not determined.

288 There was a comparable loss of CH<sub>4</sub> in recirculating air for all experiments without soil,  
289 regardless of the absence or presence of radiation from <sup>220</sup>Rn, <sup>222</sup>Rn, or both <sup>220</sup>Rn and <sup>222</sup>Rn, in  
290 dry or moist air (Fig 3a). The time needed to cross the ‘common window’ of CH<sub>4</sub> decline from  
291 67.2 to 50.9 ppmv was not shorter when radiation from <sup>220</sup>Rn and/or <sup>222</sup>Rn was added (Table 1A,  
292 B). The slopes of lines representing CH<sub>4</sub> decline within the common window in Fig 3a were not  
293 higher for experiments with elevated radiation (mean ~0.38 ppmv h<sup>-1</sup>) than for blank experiments  
294 without added radon isotopes (mean ~0.39 ppmv h<sup>-1</sup>; Table 1A, B). The mean levels of added  
295 radiation from <sup>220</sup>Rn, and especially the cumulative radiation in experiment #9 from  
296 simultaneously added <sup>220</sup>Rn and <sup>222</sup>Rn, ranged between ~50 and 535 kBq m<sup>-3</sup> after doubling of  
297 experimental <sup>220</sup>Rn values that were measured at ≤0.2 L min<sup>-1</sup> (Table 1B) and thus always exceeded  
298 the radiation levels reported in cave air [26], including the air in all Spanish caves where abiotically  
299 driven CH<sub>4</sub> oxidation due to radiolysis has been reported [10]. For example, the average rate of  
300 CH<sub>4</sub> consumption in Spanish Altamira Cave air of -0.03 ppmv h<sup>-1</sup> occurred at a maximum <sup>222</sup>Rn  
301 radiation level of ~6 kBq m<sup>-3</sup>, which is roughly one to three orders of magnitude less than the  
302 radiation in any of our experiments with added radon isotopes (#5 through #9). Thus, in terms of  
303 radiation intensity, our experiments represent an extreme test of the radiolysis hypothesis. Only  
304 the air in shafts of underground uranium mines has been observed to reach even higher radiation  
305 levels of one million or more Bq m<sup>-3</sup> [27].

306 The consistent pattern of CH<sub>4</sub> decline in our experiments without soils can be better  
307 appreciated in light of the observed CO<sub>2</sub> dynamics (Fig 3b). CO<sub>2</sub> is more polar than CH<sub>4</sub>, can be  
308 more easily adsorbed on surfaces, and is more water-soluble and reactive than CH<sub>4</sub>. Therefore, it  
309 is possible that changes in room temperature (21.1 to 27.5 °C) and atmospheric pressure (96.7 to  
310 99.3 kPa) may have affected adsorption and solubility of CO<sub>2</sub> during our experiments. Moreover,  
311 after one week of measurements with a wet paper tissue in the 4-L glass flask without soil, fungi  
312 had discolored the paper tissue and metabolically generated CO<sub>2</sub>, thus partially stabilizing the CO<sub>2</sub>  
313 partial pressure (experiment #6, Fig 3b), apparently without affecting the CH<sub>4</sub> decline (Fig 3a).  
314 The paper tissue had been hung by a thread from the central glass stopcock to maximize surface  
315 area and to avoid any anoxic microenvironments that could facilitate biological methanogenesis  
316 (Fig 1). Subsequent experiments in moist air without soil replaced the wet paper tissue with added  
317 deionized water at the bottom of the 4-L glass flask. Experiments with soils initially generated  
318 CO<sub>2</sub> *via* microbial and fungal remineralization of soil organic matter, followed after several days  
319 by a decline due to diffusive loss of CO<sub>2</sub>.

320 In the two experiments with moist soils, we documented a CH<sub>4</sub> loss of of ~0.09 ppmv h<sup>-1</sup>  
321 within the common window of CH<sub>4</sub> concentration decline (Fig 3a), as determined by subtracting  
322 the diffusive CH<sub>4</sub> loss in blank experiments from the bulk CH<sub>4</sub> loss in experiments #10 and #11  
323 with soils (Table 1). It is well established that heterogeneously distributed methanotrophic biofilms  
324 in the subsurface [28] are capable of scavenging CH<sub>4</sub> from the atmosphere [e.g., 29, 30]. Soil gas  
325 can often reach <sup>222</sup>Rn radiation levels of many thousand Bq m<sup>-3</sup>, depending on local geology [31,  
326 32]. If radiolysis would indeed be able to trigger fast oxidative decay of CH<sub>4</sub> in soil gas, such an  
327 important CH<sub>4</sub> sink in dry soils without abundant methanotrophic activity would likely have been

328 documented. Also, radiolysis would compete with methanotrophs in moist soils for CH<sub>4</sub> and would  
329 have been identified as a factor in soil CH<sub>4</sub> studies.

330

## 331 **Experiments in gas-tight terrarium at RHUL**

332 The first stage of experiment #1 at RHUL (Figs 2a and 4) established background  
333 conditions for the abundance of negative ions (~3800 ions cm<sup>-3</sup>) and the concentration of <sup>222</sup>Rn  
334 (17 to 51 Bq m<sup>-3</sup>) in laboratory air at temperatures from 21.4 to 21.7 °C and relative humidities  
335 from 26.6 to 29.0 %. After the onset of stage 2, the placement of pitchblende and a beaker with  
336 130 mL, 38 °C warm water into the sealed terrarium strongly increased the abundance of negative  
337 ions in air (up to ~200,000 ions cm<sup>-3</sup>) and the concentration of <sup>222</sup>Rn (~118 kBq m<sup>-3</sup>). The relative  
338 humidity exceeded 85 %, and the air temperature intermittently rose by 5 °C. The measurement  
339 uncertainty of the Canary Pro radon monitor increased with the <sup>222</sup>Rn radiation level (Fig 4).  
340 However, the factory-documented uncertainty at the highest measured radiation level and the  
341 steadily increasing abundance of negative ions in air suggested that after a run time of ~17 h, the  
342 <sup>222</sup>Rn-based radiation level exceeded 100 kBq m<sup>-3</sup> (Fig 4; data shown in supporting information).

343

344 **Fig 4. Experiment #1 at Royal Holloway University of London.** The experiment proceeded in  
345 two stages where the first stage established background conditions in a hermetically closed  
346 terrarium at low relative humidity and in the absence of artificially elevated concentrations of  
347 <sup>222</sup>Rn. In the second stage, the addition of pitchblende and a beaker with hot water provided a  
348 source of <sup>222</sup>Rn and high humidity to simulate conditions characteristic of cave environments.  
349 Although the final <sup>222</sup>Rn concentration exceeded 100 kBq m<sup>-3</sup> and was thus higher than in most  
350 caves, the elevated radiation and ionization of air in the terrarium was unable to lower the  
351 atmospheric abundance of CH<sub>4</sub> over 77 h in the subsequent RHUL experiment #2 (Table 2).

352

353 **Table 2. Results of experiment #2 at RHUL to assess the consumption of CH<sub>4</sub> over time in**  
354 **the presence of elevated <sup>222</sup>Rn concentrations and high relative humidity.**

Date	Time	Sample	CH <sub>4</sub> (ppm)	CH <sub>4</sub> st. dev. (ppm)	H <sub>2</sub> O (vol. %)
March 4, 2016	11:00	Laboratory air	1.9760	0.0006	0.85
March 4, 2016	11:15	Initial terrarium air	1.9971	0.0122	1.75
March 4, 2016	13:15	Laboratory air	1.9616	0.0005	0.81
March 4, 2016	13:15	Terrarium air	1.9873	0.0066	1.10
March 4, 2016	16:00	Laboratory air	1.9643	0.0006	0.85
March 4, 2016	16:00	Terrarium air	1.9818	0.0054	1.64
March 7, 2016	16:05	Laboratory air	1.9693	0.0007	0.77
March 7, 2016	16:05	Final terrarium air	1.9941	0.0036	1.84

355

356 Experiment #2 at RHUL (Fig 2b) used the same sealed terrarium with pitchblende and high  
357 humidity to monitor and compare the CH<sub>4</sub> mole fractions in the air of both the terrarium and the  
358 outside laboratory air over ~77 h. The Canary Pro radon monitor in the terrarium indicated an  
359 increase in <sup>222</sup>Rn over time parallel to RHUL experiment #1. After 5 h into RHUL experiment #2,  
360 the <sup>222</sup>Rn-based radiation in the terrarium was consistently > 50 kBq m<sup>-3</sup>. Despite high levels of  
361 ionization and <sup>222</sup>Rn-based radiation in the terrarium, the CH<sub>4</sub> mole fraction of 1.9941 ± 0.0036  
362 ppm in terrarium air after being sealed for ~77 h was indistinguishable from the starting value of  
363 1.9971 ± 0.0122 ppm within the uncertainty of measurements (Table 2).

## 364 **Synopsis of combined results**

365 The absence of any experimental evidence for accelerated loss of CH<sub>4</sub> in the presence of  
366 elevated radiation makes it highly unlikely that radiation from radon isotopes is important in nature  
367 where <sup>220</sup>Rn and <sup>222</sup>Rn concentrations are typically much lower. Our data indicate that natural  
368 radiation in cave air cannot be responsible for the rapid consumption of CH<sub>4</sub> in air on time-scales  
369 of days, even in caves with high relative humidity. The same conclusion had been reached from  
370 earlier laboratory experiments [18] and from observations in Australian cave air [15].

371 Many caves experience seasonally different degrees of venting and even reversals of air  
372 flow, which results in differences in air temperature and humidity and is difficult to simulate in  
373 laboratory experiments. Still, most cave environments at sufficient distances from cave entrances  
374 and vent holes are thermally buffered by surrounding rock and therefore do not express the  
375 relatively high diurnal and seasonal temperature and humidity variations as outside environments.  
376 Our experiments in laboratories were conducted at relatively constant room temperatures similar  
377 to many cave environments. Room temperatures in air-conditioned laboratory buildings are similar  
378 to actual temperatures in sub-tropical and tropical caves [14]. The use of water and moist soil in  
379 many of our experiments simulated the range of humidity in natural cave air. One possible caveat  
380 in terms of dissimilarity between our laboratory settings and actual caves may be the fact that our  
381 experiments allowed daylight to reach our experimental setups. However, the amount and timing  
382 of indirect light (no direct sunshine) was insufficient to let any photoautotrophs (algae) observably  
383 grow in our experiments. A necessary difference between air in our experiments at IU and actual  
384 cave air was the presence of traces of CH<sub>4</sub> in our experiments. Some CH<sub>4</sub> was needed to test for  
385 possible radiolytic destruction of CH<sub>4</sub>. In contrast, most natural cave air is depleted in CH<sub>4</sub> relative  
386 to outside air. We conclude that the experimental conditions during experiments at IU and RHUL  
387 were reasonable approximations to simulate cave conditions. In the open atmosphere, solar  
388 radiation is mainly responsible for the generation of OH• radicals [6, and refs. therein] that are the  
389 longest-lived potential radical reactant with CH<sub>4</sub> in air. Subterranean radiolysis by radioactivity  
390 involves far more energy than photochemical dissociation of molecules by solar radiation, hence  
391 the speciation of resulting ions and radicals is different. A host of highly energetic, short-lived ions  
392 and radicals other than OH• is generated in subterranean air. The first abstraction of an H atom  
393 from CH<sub>4</sub> requires a far higher activation energy than those of H atoms from methyl CH<sub>3</sub> and  
394 methylene CH<sub>2</sub> moieties. We argue that cave environments with elevated radioactivity may host  
395 short-lived, yet highly energetic radicals and ions that can supply the needed activation energy for  
396 first H-abstraction from CH<sub>4</sub> more efficiently than OH• in the open atmosphere. Thus, the

397 application of kinetic and energetic findings of photochemical CH<sub>4</sub> oxidation in the open  
398 atmosphere may not be warranted for subterranean environments.

399 The  $\alpha$ -radiation level in cave air is typically higher than in the open atmosphere because  
400 cave air is relatively close to rock and sediment surfaces with minerals harboring radioactive  
401 nuclides. The ionization rate in air *via* <sup>222</sup>Rn radon decay is larger close to the ground, as reported  
402 for a Finnish forest [33], a spa [34], and in houses [35]. The effect is due to (i) strongly elevated  
403 radon concentrations in the air in porous, uranium-containing substrates and the rapid dilution of  
404 radon above surfaces upon mixing with the open atmosphere, especially during windy conditions.  
405 In contrast, cave air far from cave entrances is typically less turbulent and allows for a more even  
406 distribution of <sup>222</sup>Rn in cave air. (ii) Short-lived <sup>220</sup>Rn will always exhibit a greater abundance in  
407 air close to its parent nuclides in soil, rock, cave walls and floors [19]. Regardless, even  
408 exceptionally high combined radiation levels of <sup>220</sup>Rn and <sup>222</sup>Rn provided no evidence for  
409 accelerated CH<sub>4</sub> oxidation in our experiments.

410 A plausible reason for slow radiolytic reaction kinetics is the mismatch between the large  
411 number of CH<sub>4</sub> molecules in 1 m<sup>3</sup> of atmosphere containing 1.85 ppmv CH<sub>4</sub> at standard conditions  
412 (i.e.,  $\sim 4.55 \cdot 10^{19}$  molecules CH<sub>4</sub>) relative to the small number of radon-related nuclear decay  
413 events in the same volume of air (e.g., 10 kBq m<sup>-3</sup> from <sup>222</sup>Rn resulting from the decay of 10,000  
414 atoms of <sup>222</sup>Rn per second). The following simplistic numerical example illustrates the lack of  
415 feasibility of radiation-induced rapid oxidation of CH<sub>4</sub>. If we assume that 1 m<sup>3</sup> of atmosphere  
416 entering a cave with 10 kBq m<sup>-3</sup>, even if every decay of <sup>222</sup>Rn leads to the oxidation of one molecule  
417 CH<sub>4</sub>, it would require a geologic time period of  $\sim 144$  million years to oxidize all CH<sub>4</sub>. In reality,  
418 the nuclide-specific radiation from the decay of <sup>222</sup>Rn alone is dwarfed by the total radiation from  
419 radon, thoron, their radioactive progeny, and any other radioactive nuclides present in a given  
420 environment [19]. The supporting information offers alternative calculations based on the  
421 assumptions that either (i) all energy from  $\alpha$ -decay is exclusively invested in radiolytic dissociation  
422 of CH<sub>4</sub> and results in the oxidation of multiple molecules of CH<sub>4</sub> per decay event, or (ii) that only  
423 a fraction of the energy from  $\alpha$ -decay is dissociating CH<sub>4</sub> in the overwhelming presence of other  
424 molecules and atoms. The calculated time periods needed to degrade 1.85 ppmv CH<sub>4</sub> at a <sup>222</sup>Rn  
425 radiation level of 10 kBq m<sup>-3</sup> range from 45.1 to 153,000 years, respectively. Even the most  
426 optimistic assumptions cannot speed up the radiolytic reaction kinetics to consume atmospheric  
427 CH<sub>4</sub> within hours to days.

428 We can use the most optimistic scenario for consumption of 1.85 ppmv CH<sub>4</sub> during 45.1  
429 years at 10 kBq m<sup>-3</sup> and calculate a radiation level of  $\sim 165$  MBq m<sup>-3</sup> that would be required to  
430 perform the same task in 24 h, which would be commensurate with kinetic CH<sub>4</sub> observations in  
431 caves. Natural radiation levels of a few MBq m<sup>-3</sup> have been measured in air where <sup>222</sup>Rn emanates  
432 through geologic faults from underlying uranium minerals [36]. Radiation levels in the range of  
433 MBq m<sup>-3</sup> have been observed in the air of uranium mines [27]. Still, no location is known to offer  
434 values close to the required  $\sim 165$  MBq m<sup>-3</sup>. We conclude that there is no natural cave environment  
435 on earth where the  $\alpha$ -radiation level is strong enough to rapidly degrade CH<sub>4</sub>. The same conclusion  
436 was recently described in a study that included arguments based on radiolytic kinetics of ion-  
437 induced reactions [15] that complement our calculations using  $\alpha$ -decay and activation energy.

438 Subterranean radiation does not provide a mechanism for a fast-acting sink of atmospheric  
439 CH<sub>4</sub> that would extend to arid and hyperarid environments, unlike microbial methanotrophy. Our  
440 study does not invalidate the geochemical data from previous studies documenting CH<sub>4</sub> dynamics  
441 in subterranean ecosystems [10]. We do not call into question the fundamental importance of

442 radiolysis of H<sub>2</sub>O (and other air components) and subsequent redox reactions that are documented  
443 in the geologic record [e.g., 37] or the long-term subterranean radiolytic impact on sedimentary  
444 organic matter [1]. However, the exceedingly slow chemical rates of reaction caused by natural  
445 rates of radiolysis would likely take years to geologic time periods in cave environments to deplete  
446 trace amounts of atmospheric CH<sub>4</sub> in cave air. As long as no alternative mechanisms have been  
447 identified, microbial methanotrophy serves as the only known fast-acting sink for subterranean  
448 CH<sub>4</sub> in the critical and vadose zones.

449

## 450 **Conclusions**

451 Strong radiation from radon isotopes and subsequent radiolysis of air proved unable to  
452 rapidly oxidize methane in dry or moist air. In the absence of a feasible alternative methane  
453 oxidation mechanism other than microbial methanotrophy, further studies are needed on the ability  
454 of microbes to consume trace amounts of methane in poorly ventilated caves, even though the  
455 trophic and energetic benefits become marginal at very low partial pressures of methane.

456

## 457 **Acknowledgments**

458 We thank Dave Matthey and the staff of the Greenhouse Gas Research Group in the  
459 Department of Earth Science at RHUL for their support and technical suggestions during  
460 experiments under project SMACKS (IEF-Marie Curie Actions n° 624204, FP7/2007-2013) and  
461 projects CGL2016-78318-C2-1R and 2R AEI/FEDER/UE. This study is based upon work at  
462 Indiana University supported by the U.S. Department of Energy, Office of Science, Office of Basic  
463 Energy Sciences, Chemical Sciences, Geosciences, and Biosciences Division under Award  
464 Number DESC0006978. In addition, financial support was provided by the National Science  
465 Foundation (0851143, 0851113) and the US Army Research Office Grant W911NF-14-1-0411.  
466 We are indebted to Scott Ensminger for guidance to a natural methane seep in New York State,  
467 Dương Nguyễn-Thùy and Hương Nguyễn-Văn at Vietnam National University Hanoi for a <sup>220</sup>Rn-  
468 emitting mud brick, and Greg Crouch at Indiana University for technical and radiation safety  
469 advice. Steffen Kolb, Tibor Kovacs, Leonie Anna Mueck, and two anonymous reviewers provided  
470 expert advice to improve the manuscript.

471

472

## 473 **References**

474

- 475 1. Greenwood PF, Shan C, Holman AI, Grice K. The composition and radiolysis impact on  
476 aromatic hydrocarbons in sedimentary organic matter from the Mulga Rock (Australia)  
477 uranium deposit. *Org. Geochem.* 2018; 123: 103-112.  
478 <https://doi.org/10.1016/j.orggeochem.2018.06.013>
- 479 2. Colwell FS, D'Hondt S. Nature and extent of the deep biosphere, *Rev. Mineral. Geochem.*  
480 2013; 75: 547-574. <http://dx.doi.org/10.2138/rmg.2013.75.17>
- 481 3. Colman DR, Poudel S, Stamps BW, Boyd ES, Spear JR. The deep, hot biosphere: Twenty-  
482 five years of retrospection. *PNAS* 2017; 114 (27): 6895-6903.

- 483 <https://dx.doi.org/10.1073/pnas.1701266114>
- 484 4. Xu X, Yuan F, Hanson PJ, Wullschleger SD, Thornton PE, Riley WJ, et al. Reviews and  
485 syntheses: Four decades of modeling methane cycling in terrestrial ecosystems.  
486 *Biogeosciences* 2016; 13: 3735-3755. <https://dx.doi.org/10.5194/bg-13-3735-2016>
- 487 5. IPCC. Climate Change 2013: The Physical Science Basis, Figure SPM.5 of the Summary for  
488 Policymakers; 2013. [http://www.ipcc.ch/pdf/assessment-](http://www.ipcc.ch/pdf/assessment-report/ar5/wg1/WG1AR5_SPM_FINAL.pdf)  
489 [report/ar5/wg1/WG1AR5\\_SPM\\_FINAL.pdf](http://www.ipcc.ch/pdf/assessment-report/ar5/wg1/WG1AR5_SPM_FINAL.pdf) (accessed on September 4<sup>th</sup>, 2018)
- 490 6. Prather MJ, Holmes CD. Overexplaining or underexplaining methane's role in climate  
491 change. *PNAS* 2017; 114 (21): 5324-5326. <https://dx.doi.org/10.1073/pnas.1704884114>
- 492 7. Edwards CR, Onstott TC, Miller JM, Wiggins JB, Wang W, Lee CK, et al. Draft genome  
493 sequence of uncultured upland soil cluster Gammaproteobacteria gives molecular insights  
494 into high-affinity methanotrophy. *Genome Announc.* 2017; 5: e00047-17.  
495 <https://doi.org/10.1128/genomeA.00047-17>
- 496 8. Waring CL, Griffith DWT, Wilson S, Hurry S. Cave atmosphere: A guide to calcification  
497 and a methane sink. *Geochim. Cosmochim. Acta* 2009; 73 (13): A1419-A1419.  
498 <http://apo.ansto.gov.au/dspace/handle/10238/2461?mode=full> (accessed on September  
499 4<sup>th</sup>, 2018)
- 500 9. Matthey DP, Fisher R, Atkinson TC, Latin JP, Durrell R, Ainsworth M, et al. Methane in  
501 underground air in Gibraltar karst. *Earth Planet. Sci. Lett.* 2013; 374: 71-80.  
502 <http://dx.doi.org/10.1016/j.epsl.2013.05.011>
- 503 10. Fernandez-Cortes A, Cuezva S, Alvarez-Gallego M, Garcia-Anton E, Pla C, Benavente D, et  
504 al. Subterranean atmospheres may act as daily methane sinks. *Nat. Commun.* 2015; 6:  
505 7003. <https://dx.doi.org/10.1038/ncomms8003>
- 506 11. McDonough LK, Iverach CP, Beckmann S, Manefield M, Rau GC, Baker A, et al. Spatial  
507 variability of cave-air carbon dioxide and methane concentrations and isotopic  
508 compositions in a semi-arid karst environment. *Environ. Earth Sci.* 2016; 75: 700.  
509 <http://dx.doi.org/10.1007/s12665-016-5497-5>
- 510 12. Webster KD, Mirza A, Deli JM, Sauer PE, Schimmelmann A. Consumption of atmospheric  
511 methane in a limestone cave in Indiana, USA. *Chem. Geol.* 2016; 443: 1-9.  
512 <https://dx.doi.org/10.1016/j.chemgeo.2016.09.020>
- 513 13. Webster KD, Drobniak A, Etiope G, Mastalerz M, Sauer PE, Schimmelmann A.  
514 Subterranean karst environments as a global sink for atmospheric methane. *Earth Planet.*  
515 *Sci. Lett.* 2018; 485: 9-18. <https://doi.org/10.1016/j.epsl.2017.12.025>
- 516 14. Nguyễn-Thùy D, Schimmelmann A, Nguyễn-Văn H, Drobniak A, Lennon JT, Tạ PH, et al.  
517 Subterranean microbial oxidation of atmospheric methane in cavernous tropical karst.  
518 *Chem. Geol.* 2017; 466: 229-238. <https://doi.org/10.1016/j.chemgeo.2017.06.014>
- 519 15. Waring CL, Hankin SI, Griffith DWT, Kertesz MA, Kobylski V, Wilson NL, et al. Seasonal  
520 total methane depletion in limestone caves. *Sci. Rep.* 2017; 7: 8314.  
521 <https://dx.doi.org/10.1038/s41598-017-07769-6>
- 522 16. Haynes RM, Kebarle P. Mass-spectrometric study of ions at near-atmospheric pressure. III.  
523 The alpha radiolysis of methane. *J. Chem. Phys.* 1966; 45 (11): 3899-3906.  
524 <https://dx.doi.org/10.1063/1.1727435>
- 525 17. Rasul G, Prakash GKS, Olah GA. Comparison of structures and energies of CH<sub>5</sub><sup>2+</sup> with  
526 CH<sub>4</sub><sup>+</sup> and their possible role in superacidic methane activation. *PNAS* 1997; 94 (21):  
527 11159-11162. <https://dx.doi.org/10.1073/pnas.94.21.11159>
- 528 18. Lennon JT, Nguyễn-Thùy D, Phạm TM, Drobniak A, Tạ PH, et al. Microbial contributions



- 529 to subterranean methane sinks. *Geobiology* 2017; 15 (2): 254-258.  
530 <https://dx.doi.org/10.1111/gbi.12214>
- 531 19. Meisenberg O, Mishra R, Joshi M, Gierl S, Rout R, Guo L, et al. Radon and thoron  
532 inhalation doses in dwellings with earthen architecture: Comparison of measurement  
533 methods. *Sci. Total Environ.* 2017; 579: 1855-1862.  
534 <http://dx.doi.org/10.1016/j.scitotenv.2016.11.170>
- 535 20. Markkanen M, Arvela H. Radon emanation from soils. *Radiat. Prot. Dosimetry* 1992; 45 (1-  
536 4), 269-272. <https://doi.org/10.1093/oxfordjournals.rpd.a081541>
- 537 21. Morawska L, Phillips CR. Dependence of the radon emanation coefficient on radium  
538 distribution and internal structure of the material. *Geochim. Cosmochim. Acta* 1993; 57  
539 (8): 1783-1797. [https://doi.org/10.1016/0016-7037\(93\)90113-B](https://doi.org/10.1016/0016-7037(93)90113-B)
- 540 22. Hursh JB. Thoron half-life. *J. Inorg. Nucl. Chem.* 1966; 28 (12): 2771-2776.  
541 [https://doi.org/10.1016/0022-1902\(66\)80001-5](https://doi.org/10.1016/0022-1902(66)80001-5)
- 542 23. Wang J, Meisenberg O, Chen Y, Karg E, Tschiersch J. Mitigation of radon and thoron decay  
543 products by filtration. *Sci. Total Environ.* 2011; 409 (19): 3613-3619.  
544 <https://doi.org/10.1016/j.scitotenv.2011.06.030>
- 545 24. Oremland RS, Capone DG. Use of “specific” inhibitors in biogeochemistry and microbial  
546 ecology. In: Marshall KC, editor. *Advances in Microbial Ecology* 10 (ch. 8), Plenum  
547 Publishing Corporation; 1988. pp. 285-383. [https://dx.doi.org/10.1007/978-1-4684-5409-  
548 3\\_8](https://dx.doi.org/10.1007/978-1-4684-5409-3_8)
- 549 25. Farhan UI Haque M, Crombie AT, Ensminger SA, Baciu C, Murrell JC. Facultative  
550 methanotrophs are abundant at terrestrial natural gas seeps. *Microbiome* 2018; 6: 118.  
551 <https://doi.org/10.1186/s40168-018-0500-x>
- 552 26. Cigna AA. Radon in caves. *Int. J. Speleol.* 2005; 34 (1–2): 1-18.  
553 <http://scholarcommons.usf.edu/cgi/viewcontent.cgi?article=1203&context=ijs> (accessed  
554 on September 4<sup>th</sup>, 2018)
- 555 27. Wismut GmbH and MECSEK-ÖKO. Annual Report 2010 on the Cooperation between  
556 Wismut GmbH and MECSEK-ÖKO. 2011; p. 24-26.  
557 [https://www.google.com/url?sa=t&rct=j&q=&esrc=s&source=web&cd=3&ved=2ahUKE  
558 wjm7rrq5JrdAhWF34MKHScrBTQQFjACegQICBAC&url=http%3A%2F%2Fgov.pecs.  
559 hu%2Fdownload%2Findex.php%3Fid%3D309232&usg=AOvVaw0u7IRhg7ZYhYU2IGY  
560 shAXv](https://www.google.com/url?sa=t&rct=j&q=&esrc=s&source=web&cd=3&ved=2ahUKEwjm7rrq5JrdAhWF34MKHScrBTQQFjACegQICBAC&url=http%3A%2F%2Fgov.pecs.hu%2Fdownload%2Findex.php%3Fid%3D309232&usg=AOvVaw0u7IRhg7ZYhYU2IGYshAXv) (accessed on September 4<sup>th</sup>, 2018)
- 561 28. Yan Z, Liu C, Liu Y, Bailey VL. Multiscale investigation on biofilm distribution and its  
562 impact on macroscopic biogeochemical reaction rates. *Water Resour. Res.* 2017; 53 (11):  
563 8698-8714. <https://dx.doi.org/10.1002/2017WR020570>
- 564 29. Mancinelli RL. The regulation of methane oxidation in soil. *Annu. Rev. Microbiol.* 1995;  
565 49: 581-605. <https://doi.org/10.1146/annurev.mi.49.100195.003053>
- 566 30. Conrad R. Microbial ecology of methanogens and methanotrophs. *Adv. Agron.* 2007; 96: 1-  
567 63. [https://doi.org/10.1016/S0065-2113\(07\)96005-8](https://doi.org/10.1016/S0065-2113(07)96005-8)
- 568 31. Lindmark A, Rosen B. Radon in soil gas — Exhalation tests and in situ measurements. *Sci.*  
569 *Total Environ.* 1985; 45: 397-404. [https://doi.org/10.1016/0048-9697\(85\)90243-8](https://doi.org/10.1016/0048-9697(85)90243-8)
- 570 32. Papastefanou C. Measuring radon in soil gas and groundwaters: a review. *Annals Geophys.*  
571 2007; 50 (4): 569-578.  
572 <http://www.annalsofgeophysics.eu/index.php/annals/article/download/3070/3113>  
573 (accessed on September 4<sup>th</sup>, 2018)
- 574 33. Tammet H, Hörrak U, Laakso L, Kulmala M. Factors of air ion balance in a coniferous

- 575 forest according to measurements in Hyytiälä, Finland. Atmos. Chem. Phys. 2006; 6:  
576 3377-3390. <https://doi.org/10.5194/acp-6-3377-2006>
- 577 34. Cosma C, Suci I, Jäntschi L, Bolboacă SD. Ion-molecule reactions and chemical  
578 composition of emanated from Herculane Spa geothermal sources. Int. J. Mol. Sci. 2008;  
579 9 (6): 1024-1033. <https://dx.doi.org/10.3390/ijms9061024>
- 580 35. Kolarz PM, Filipović DM, Marinković BP. Daily variations of indoor air-ion and radon  
581 concentrations. Appl. Radiat. Isot. 2009; 67 (11): 2062-2067.  
582 <https://doi.org/10.1016/j.apradiso.2009.07.023>
- 583 36. Seyis C, İnan S, Streil T. Ground and indoor radon measurements in a geothermal area. Acta  
584 Geophys. 2010; 58 (5): 939-946. [https://link.springer.com/article/10.2478/s11600-010-  
585 0012-y](https://link.springer.com/article/10.2478/s11600-010-0012-y)
- 586 37. Leticariu L, Pratt LA, LaVerne JA, Schimmelmann A. Anoxic pyrite oxidation by water  
587 radiolysis products — A potential source of biosustaining energy. Earth Planet. Sci. Lett.  
588 2010; 292 (1-2): 57-67. <https://dx.doi.org/10.1016/j.epsl.2010.01.020>

589

590

## 591 **Supporting information**

592 An Excel file “S1\_File.xlsx” contains a first sheet “read me” with instructions and an  
593 overview on additional sheets offering analytical details and radiolysis calculations.