

¹ Potential groundwater recharge during floods

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₁₀ Potential groundwater recharge during floods

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15

Abstract

16 Groundwater constitutes 30% of fresh water reserves on Earth. It is
17 important as a source for drinking water and irrigation due to its good
18 quality. For many aquifers in arid regions, long-term groundwater
19 extraction has put in risk its sustainable use. Thus, it is relevant
20 to understand and quantify processes that contribute to sustainable
21 groundwater recharge.

22 Most recharge to aquifers in arid regions occurs during flood events
23 that happen with a frequency of a few years to decades. Paleo-climatic
24 records show that intensity and frequency of floods have been partic-
25 ularly variable during periods of climate change. Therefore, under-
26 standing how floods could impact the magnitude and occurrence of
27 groundwater recharge to aquifers in arid regions is relevant for improv-
28 ing water management strategies and assessing aquifer vulnerability
29 to pollution from surface streams.

30 Direct measurement of infiltration during flood events is difficult,
31 so it is common to complete analyses with numerical simulations. We
32 present results of detailed numerical simulations of infiltration through
33 the vadose zone during flood events. We use the results of the sim-
34 ulations to characterize infiltration patterns and quantify potential
35 for recharge to aquifers considering different subsurface conceptual-
36 izations, from simple homogeneous to more realistic multi-scale het-
37 erogeneous sediment distributions. We also make a few additional
38 general comments for practical applications.

³⁹ **keywords:** water resources, groundwater recharge, flooding, numerical
⁴⁰ simulations, vadose zone

41 1 Introduction

42 It is known that the interaction between surface water and groundwater is
43 a key process in hydrological systems that, among other things, serve to
44 dampen floods and store water in aquifers for future use [8]. Groundwater
45 is a main source of good quality water in many regions around the world,
46 being approximately a third of the fresh water reserves on Earth [8]. It can
47 be the sole water source in arid and semi-arid areas, i.e. regions with a
48 ratio of mean annual rainfall to potential evapotranspiration less than 0.5
49 [32], which cover more than 30% of the Earth surface. Approximately a
50 quarter (2.5 billion) of the world population lives in arid or semi-arid areas
51 [14], reaching almost 100% in large geographic regions such as the Middle
52 East, Southwest USA, and some places in Australia and South America, e.g.
53 Southern Peru, Eastern Argentina and Northern Chile.

54 According to a report prepared by the United Nations published in 2026
55 [24]: *"Groundwater now provides about 50% of global domestic water use
56 and over 40% of irrigation water, tying both drinking water security and
57 food production directly to rapidly depleting aquifers. Around 70% of the
58 world's major aquifers show long-term declining trends"* [24]. On the other
59 hand, it is expected that due to climate change, dry periods (droughts) will
60 have longer duration, so that demand for groundwater for use in irrigation
61 and human consumption will increase [37]. It has been estimated that over
62 1.8 billion people lived under drought conditions in 2022–2023 and that the

63 cost of those climatic anomalies amounted to about US\$307 billion per year
64 worldwide [24].

65 Intense groundwater exploitation is relatively recent, dating back only
66 to the past century (early to mid 1900s). Therefore, there is a lack of a
67 good understanding of long-term decades long dynamic behavior of aquifers
68 under intense groundwater extraction, which is essential to assess sustainable
69 water management strategies. Due to this, the use of numerical models to
70 evaluate the dynamic response of aquifers to groundwater extraction is a
71 common practice [5, 4]. Such models must take into account not only the
72 complex nature of the sediments and rocks that compose aquifers, but also
73 the fluctuating nature of climatic variables such as rainfall, temperature and
74 evapotranspiration, which are the key elements to calculate the potential
75 excess water available to recharge aquifers.

76 A common paradox in arid regions is that by definition, the calculation of
77 potential recharge based on mean annual values of rainfall and evapotranspi-
78 ration extracted from short time series up to few decades long predicts zero
79 recharge. Hence, it has been postulated and demonstrated for some arid
80 regions that currently available groundwater was recharged long ago during
81 periods of different climatic patterns with higher rainfall and/or lower evap-
82 otranspiration [40]. For example, chemical analysis of vertical water sample
83 profiles in North Africa indicate that concentrated recharge beneath streams
84 ceased about 5,000 years ago as result of a shift in climate patterns [32].
85 Geochemical analysis of water samples from aquifers located in arid areas in

86 Texas and Nevada, USA, that experience low rainfall, estimated long resi-
87 dence time for pore water stored in unsaturated zone of the order of 50,000
88 to more than 100,000 years even at relatively shallow depth (≈ 25 m) [40].
89 Flash floods are rapid (within a few hours) and significant (of the order of up
90 to ten times) increases in flow discharge of natural streams that are usually
91 accompanied by the inundation of large land areas over river banks [1]. Ac-
92 cording to existing records, frequency and magnitude of floods have shown to
93 be specially variable during periods of climate change [10, 40]. An alternative
94 hypothesis postulates that most recharge to aquifers in arid regions happens
95 during flash floods that take place periodically interspersed by a few to tens
96 years [32, 37, 22, 11]. Available water for infiltration, i.e. recharge, during
97 such events is several orders of magnitude higher than computed with mean
98 annual values [20]. For example, a detailed review of recharge estimates in
99 a semi-arid region located in New Mexico, USA; assessed that, while dis-
100 tributed recharge is usually less than 50-100 mm per year, focused recharge
101 rate beneath streams can reach up to more than 700 m/year [32]. Focused
102 recharge also results in much lower transit time of water through the unsat-
103 urated zone than for distributed recharge. The later can take up to a few
104 centuries to travel from the ground surface to the water table in places where
105 this locates a significant depth (≥ 100 m) [32].

106 The interaction between surface streams and shallow aquifers can be
107 stronger during floods. For example, a site-specific study found based on
108 observations and numerical simulations that for an aquifer located in Aus-

109 tralia connected to a stream, groundwater levels could increase very rapidly
110 and significantly (up to almost 10 m) during high river stages, while infil-
111 trated water could travel significant distances away from streams (up to 40
112 m) within short time [41]. They also found rapid variations in groundwater
113 flow direction before, during and after flooding events.

114 It is widely accepted that for most major exploited aquifers, long-term
115 water use has exceeded renewable inflows and safe depletion limits, putting
116 in risk the sustainable use of those resources for future generations [24].
117 Understanding how more intense floods can impact the interaction between
118 ephemeral surface streams and groundwater can be useful for improving water
119 management and flood mitigation systems [37, 22, 6]. It can also be relevant
120 to evaluate potential increase in aquifer vulnerability due to intense recharge
121 produced by natural floods or by dam failures during such events [7, 2, 43, 24].

122 The measurement of exchange fluxes between surface streams and aquifers
123 through direct methods: stream-flow based, groundwater based and infiltr-
124 ation based; is difficult, particularly in arid regions where most streams are
125 ephemeral and stay dry during long periods that are separated by extreme
126 peak flow events [34]. Because of the difficulties to directly measure infiltr-
127 ation, it is also common to use analytical estimates or numerical simulations
128 to complement observations [6, 33].

129 The principal objective of this work is estimating the potential for ground-
130 water recharge to aquifers located in arid regions during flash floods. Specif-
131 ically, we investigate the interaction between surface water and groundwater

132 during flash flood events based on numerical simulations. We evaluate dif-
133 ferent plausible subsurface settings, from simple homogeneous materials to
134 multi-scale heterogeneous sediment distributions. We use the simulation re-
135 sults to analyze the dynamic response of groundwater levels under focused
136 recharge that may occur during floods. Based on those results we make con-
137 clusions for practical applications such as: estimating recharge from observed
138 variations in groundwater levels, quantification of potential recharge for dif-
139 ferent flood durations and estimates for transit times of infiltrated water
140 through the vadose zone.

141 **2 Numerical simulations: Setup**

142 We use numerical simulations for understanding the dynamics of recharge
143 during flood events and assessing its potential magnitude. We employ a nu-
144 mercial simulator developed for use in supercomputers, PFLOTTRAN [17], to
145 perform detailed simulations of unsaturated flow through the vadose zone as
146 result of focused recharge induced by floods. The use of numerical simula-
147 tions for investigating stream-aquifer interaction have been applied in other
148 studies [e.g. 9, 35, 6].

149 Infiltration during floods is controlled by many parameters, e.g.: perme-
150 ability, heterogeneity of the subsurface, stream water height, depth to the
151 water table and riverbed conductance [35, 6, 18]. Here, we focus only on a
152 few of those parameters: magnitude and spatial distribution of permeabil-

153 ity, vertical extent of the receiving aquifer because of the possibility of flow
154 through the bottom boundary, and stream water height.

155 For the purpose of the analysis, we consider an idealized system composed
156 of an ephemeral river that during flood events, has a river stage that can be
157 considered almost constant and covers a well defined discharge section. In
158 addition, we assume that the water table is relatively deep so that the river
159 system is disconnected from the aquifer most of the time, which is the situ-
160 ation found in most ephemeral systems located in arid or semi-arid regions.
161 The idealized system assumes that the riverbed is almost flat so that flow
162 fluctuations due to micro-topography or vegetation can be neglected [39, 3].
163 In addition, we assume that the sediments beneath the riverbed are uncon-
164 solidated and that do not exhibit large fractures or cavities due to dessication
165 or karstic processes [22]. Since we consider floods with peak discharge that
166 can be ten times higher than the average, we assume that fine sediments
167 deposited during low discharge are removed, so that the effective vertical
168 hydraulic conductivity during peak flow does not depend on that layer [6].
169 As an additional simplification, we do not consider the potential presence of
170 cobbles or boulders in the riverbed as observed in some rivers that experience
171 high energy floods, e.g. the San José River located in northern Chile (Figure
172 1), which we used as motivation for the setting of the simulations.



Figure 1: Riverbed of San Jose river in northern Chile. Riverbed (top) and shallow stratigraphy (bottom). Picture taken from Undergraduate Thesis, G. Jimenez, U. Chile.

173 Neglecting rapid connection between the stream and the water table dur-
174 ing floods, which is a reasonable assumption for deep water table aquifers [35];
175 infiltration rates can be considered almost constant during flooding events,
176 discounting a short initial period of higher infiltration due to the sudden wet-
177 ting of the dried upper sediments [29]. Large floods can carry enough water
178 to keep a significant water height above the riverbed, so that discounting
179 preferential flow through the stream due to high slope of the river bottom
180 or low permeability of the riverbed, the infiltration problem transitions from

181 one controlled by flow rates to one controlled by a quasi-constant hydraulic
182 head boundary [13].

183 We consider three different permeability distributions for the sediments
184 beneath the riverbed: homogeneous, stratified/layered and random spatially
185 correlated. Similar distributions have been used in multiple studies to inves-
186 tigate groundwater dynamics [e.g. 26, 12, 15, 28, 31, and references therein].
187 Sediments correspond to two types of sandy soils, hereafter referred to as
188 hydrofacies, materials or units, that were characterized as part of the assess-
189 ment of an artificial recharge project in northern Chile (see Appendix for
190 details). We refer to them as *Sandy* and *Silty*, for the coarser more perme-
191 able and the finer less permeable unit, respectively. This conceptualization
192 based on two hydrofacies is reasonable to model a real site. For example, it
193 is similar to the one adopted for the highly studied Hanford Site in the USA
194 [36].

195 We model a cross-section 500 m wide and 80 m height that does not
196 include slope for the riverbanks (Figure 2), and a water table located at
197 60 m depth. The 2D cross-section was divided into 20 cm wide and 5 cm
198 high cells, resulting in a total of 4 million cells. Such discretization was
199 chosen to guarantee a good numerical resolution and accuracy, while keeping
200 the running time reasonable. The simulations were run in an instance of
201 Amazon AWS with 72 computational cores and 192 GB RAM.

202 We analyse a single flood with a duration equal to 4 days, and we further
203 simplify the problem by assuming a constant mean water stage. Thus, infil-

tration from the river was modeled as a constant pressure boundary condition applied along a central 20 m wide strip. Lateral flow through the saturated section of the aquifer was modeled using two hydrostatic pressure boundary conditions available in PFLOTTRAN. The bottom boundary condition was modeled as an open (hydrostatic pressure) or no flow condition, which allows assessing the impact that the conceptualization of the aquifer geometry may have on the simulation results. Table 1 summarizes the simulations setup.

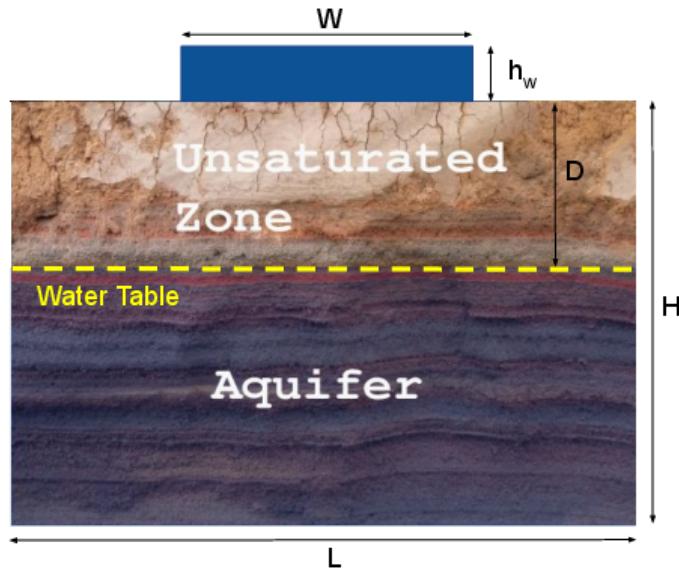


Figure 2: Schematic of cross-section considered in numerical simulations.

Parameter	Value	Symbol
Stream width [m]	20	W
Water height in stream [m]	1,0.5,0.2,0.1	h_w
Water table depth [m]	60	D
2D cross-section width [m]	500	L
2D cross-section height [m]	80	H
Hydrofacies	Sandy, Silty	-
Hydraulic conductivity [m/d]	7, 0.3	K
Permeability [m ²]	8×10^{-12} , 3×10^{-13}	k
Porosity [-]	0.33, 0.46	ϕ

Table 1: Parameters used to set up simulations.

211 We simulate a total of 8 scenarios depending upon the stratigraphy of
 212 the sediments beneath the riverbed. As first scenario, we consider an aquifer
 213 composed of a single hydrogeological unit with a permeability equal to the
 214 mean value for the Sandy or Silty units (scenarios H1, H1b and H2). The
 215 next two scenarios were defined to account for the potential presence of low
 216 permeability units beneath the riverbed, which can contribute to the occur-
 217 rence of unsaturated areas beneath even permanent surface streams due to
 218 reductions in vertical infiltration rates [33]. We consider stratified aquifers
 219 composed by two hydrogeological units: Sandy-Silty or Silty-Sand according
 220 to their vertical occurrence (scenarios L1 and L2). Silty layers 3 m thick
 221 are intercalated within higher permeability Sandy layers 10 m each. This
 222 sequence is repeated from the ground surface to the bottom of the domain.
 223 To better quantify the impact of river depth on infiltration rates, three addi-
 224 tional scenarios similar to L2 were considered for which the river stage was

225 set to: 0.5 (L2b), 0.2 (L2c) and 0.1 (L2d) meters instead of the 1.0 m used
226 for the original L2 scenario.

227 Finally, for the last 4 scenarios, we consider a sandy aquifer with val-
228 ues of permeability distributed according to a random multi-scale spatially
229 correlated field [12, 15, 28, 31]. We use a multi-scale approach to generate
230 bimodal gaussian correlated fields by using an implementation of the Turn-
231 ing Bands method [25, 38]. First, we generate random fields with different
232 ratios between horizontal (λ_h) to vertical (λ_v) correlation lengths, that we
233 use as basis to create hydrofacies distributions using an indicator approach
234 [16]. The resulting hydrofacies distribution exhibits good continuity and the
235 same spatial correlation or extension described by the correlation lengths of
236 the underlying random field. Second, we generate additional random fields
237 with mean permeability equal to the one assigned to each hydrofacies and
238 short isotropic correlation lengths equal to 0.5 m. For all the *intra-facies*
239 permeability distributions we consider an exponential covariance model and
240 low variance σ_Y equal to 0.5 ($Y = \text{Ln}(k)$), which corresponds to relatively
241 mild heterogeneity for real aquifers [15, 31]. These secondary random fields
242 were used to assign saturated permeability values to each cell of the domain.
243 The resulting overall permeability field has a bimodal distribution and mul-
244 tiple correlation lengths, which are properties that have been postulated as
245 more realistic [15, 16, 28, 31]. For example, Figure 3 shows the hydrofacies
246 distribution and multi-scale permeability field assigned to scenario R3. Table
247 2 summarizes the simulated scenarios.

Run	Description
H1 ⁺	Homogeneous sandy aquifer
H2	Homogeneous silty aquifer
L1	Layered aquifer: sandy/silty
L2*	Layered aquifer: silty/sandy
R1	Random correlated permeability, with $\lambda_h/\lambda_v = 1$
R2	Random correlated permeability, with $\lambda_h/\lambda_v = 10$
R3 [†]	Random correlated permeability, with $\lambda_h/\lambda_v = 20$
R4	Random correlated permeability, with $\lambda_h/\lambda_v = 40$

Table 2: Parameters used to characterize geological settings. In all simulations with gaussian distributions we considered the vertical correlation length $\lambda_v = 3$ m and a proportion distribution of 0.23 and 0.77 for the Silty and Sandy units, respectively. ⁺H1b equal to H1, except that bottom boundary condition is set to no flow. ^{*}L2b,c,d equal to L2 but with different river stage. [†]R3b equal to R3 except that a single value of permeability is assigned to each hydrofacies.

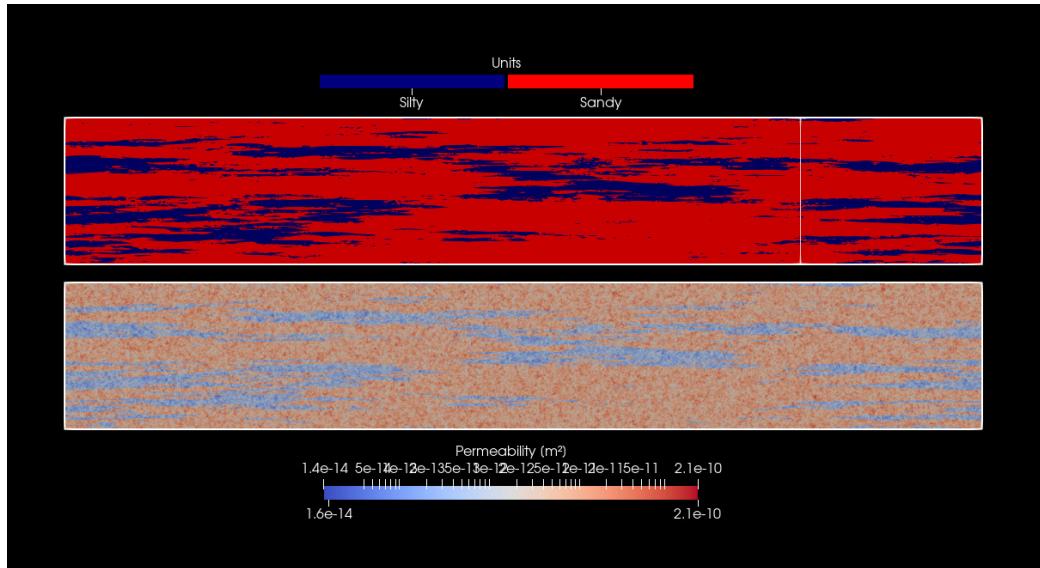


Figure 3: Hydrofacies distribution and multi-scale random permeability distribution assigned to scenario R3.

248 We use a pair of constitutive relations to characterize the unsaturated flow
249 properties (see Appendix for details), which were tied to each hydrofacies.
250 This is a reasonable assumption given that each hydrofacies is supposed to
251 have a single characteristic grain size distribution [23].

252 3 Numerical simulations: Results

253 3.1 Saturation distribution

254 Figure 4 presents the simulated saturation distribution along a vertical cross-
255 section after 4 days of infiltration through the riverbed for the 8 main sce-
256 narios considered. There are major differences in the saturation distributions
257 for the different scenarios.

258 The homogeneous (H1 and H2) and layered (L1 and L2) scenarios show a
259 wetting front with a relatively simple shape. The speed of the advancing front
260 for those scenarios is mainly controlled by the different values of permeability
261 assigned to both hydrofacies (Sandy and Silty). For the layered scenarios,
262 the Silty layers acts as low permeability barriers generating a so called *leaky-*
263 *flow*, i.e. partially-saturated permeable sediments underneath fully saturated
264 pockets or perched aquifers.

265 The saturation distribution for the scenarios with multi-scale permeabil-
266 ity, R1 to R4, shows the development of a saturation front beneath the
267 riverbed with a complex pattern and a mean width that is controlled by the
268 correlation length assigned to the less permeable hydrofacies. The front is

269 narrow for scenario R1 that consider a short horizontal correlation length,
270 while it becomes wider for the scenarios with longer horizontal correlation
271 (R3 and R4). The inclusions of low permeable Silty material act as barriers
272 for vertical flow, so that water accumulates on top of them and spills over
273 their flanks. Just below the riverbed and up to a certain variable depth, there
274 is a well connected fully saturated central zone. However, the distribution
275 of saturation becomes erratic and highly non-uniform at greater depth. This
276 means that for practical applications point-like observations, e.g. core sam-
277 ples or water samples obtained through lysimeters, may not provide enough
278 information to characterize the saturation distribution. Interestingly, even if
279 the fully saturated front does not reach the original water table for the time
280 period considered, there is clear local mounding below zones with higher
281 leakage, e.g. left central and central zones for Scenarios R3 and R4, respec-
282 tively. However, the height of the mounding is variable even within relatively
283 short distances, which may be an obstacle to correctly infer recharge rates
284 from water level measurements as it is common practice in hydrogeological
285 studies.

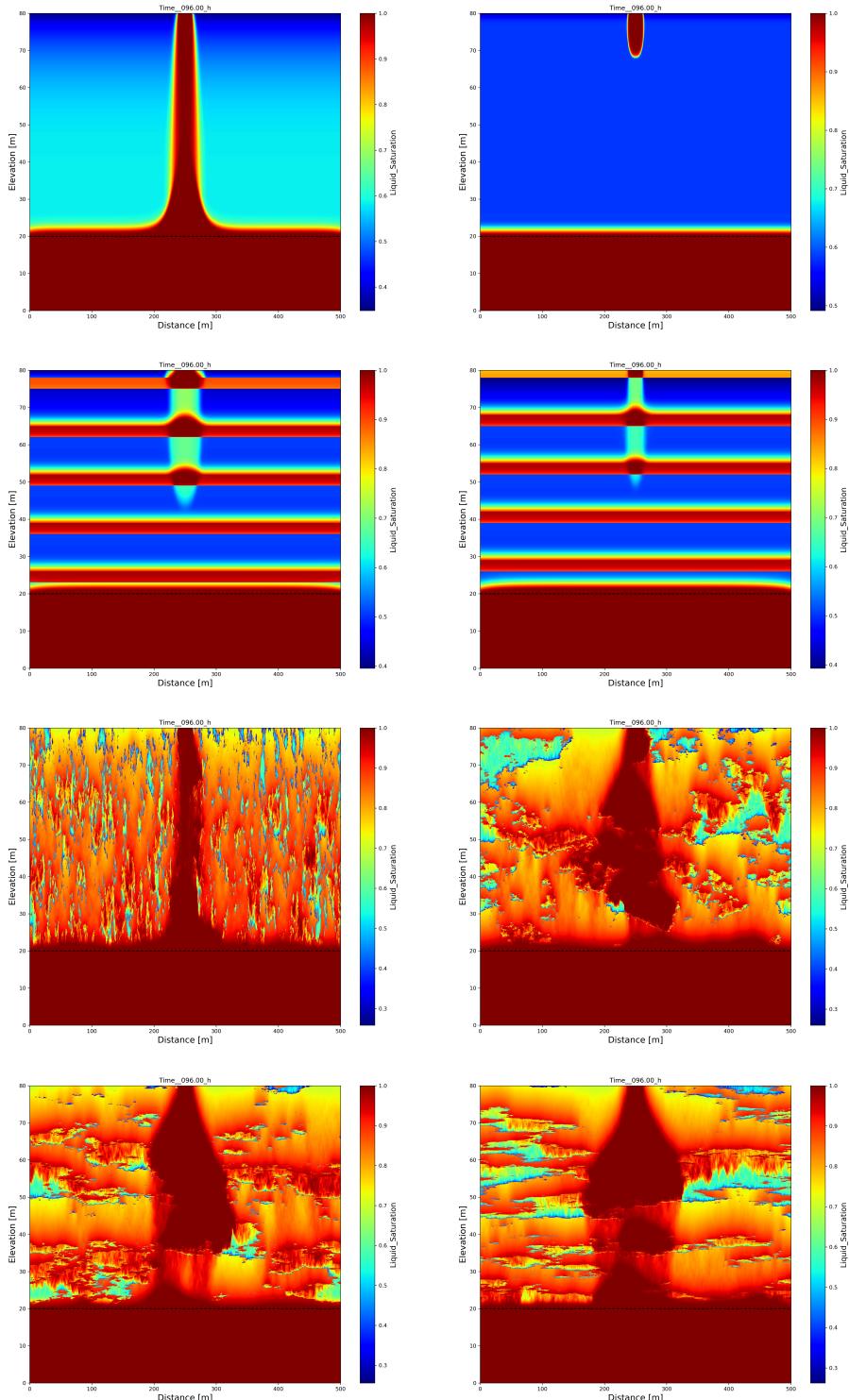


Figure 4: Simulated saturation after 96 hrs of infiltration along a cross-section located beneath the riverbed. River centerline located at 250 m in the horizontal axis. From left to right and top to bottom: H1, H2, L1, L2, R1, R2, R3 and R4.

286 Figure 5 shows simulated saturation at the end of the 4 days for Scenario
 287 R3 and R3b, thus it allows assessing the impact of including *intra-facies*
 288 heterogeneity. There are only small differences that are almost undistinguish-
 289 able in the figure, which indicates that the saturation distribution is mainly
 290 controlled by the distribution of the hydrofacies, i.e. large-scale heterogene-
 291 ity.

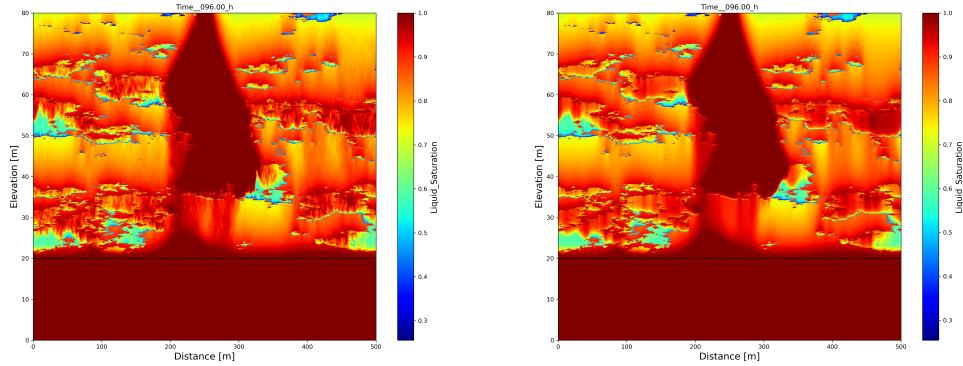


Figure 5: Simulated saturation after 96 hrs of infiltration for R3 (left) and R3b (right).

292 In all scenarios discussed so far, we considered a permeable boundary
 293 condition at the bottom of the receiving aquifer. Hence, there is potential
 294 for a large part of the water that reaches the main aquifer to leave the domain
 295 through the bottom. Figure 6 shows a comparison between Scenario H1b that
 296 includes a no-flow bottom boundary condition and the reference Scenario
 297 H1, which considers an open bottom boundary. There is a large difference
 298 in the wetting front that develops below the riverbed for both scenarios. For
 299 Scenario H1b, water can only leave the domain though the lateral faces, so

300 that there is significant rise of the original water table within the central part
 301 of the model domain for accommodating lateral flow to balance the inflow
 302 due to recharge. This distinction can be potentially important to infer the
 303 existence of lower impermeable limits based on the observation of changes in
 304 groundwater levels during and after flood events. On the other hand, this
 305 difference can lead to errors in the interpretation of observed mounding for
 306 estimating recharge rates. The infiltration for both scenarios, R3 and R3b,
 307 is the same (see Table 3) despite the large differences observed in the shape
 308 of the water table.

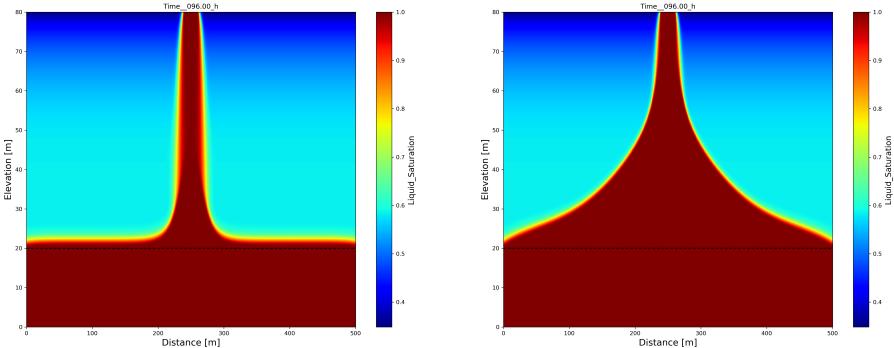


Figure 6: Simulated saturation after 96 hrs of infiltration for H1 (left) and H1b (right).

309 3.2 Water table variation

310 Figure 7 shows simulated piezometric head at the end of the simulated 4 days,
 311 along a horizontal profile located at the original position of the water table
 312 at $z=20$ m (60 m deep). Most of the simulated scenarios show a significant
 313 increase of the water table in the central zone of the domain beneath the river

314 bed. The magnitude of the change is particularly important for Scenario H1b
315 that considers no-flow through the bottom of the domain. The increase in
316 piezometric levels is also important for the most permeable homogeneous
317 scenario (H1) and all scenarios that consider heterogeneous sediments (R).
318 For the latter, the variation in piezometric head along the profile is irregular,
319 reaching a maximum at some location near the center of the domain, but
320 not at the center as in the scenarios with homogeneous or layered materials.
321 This can be potentially important for practical applications that require the
322 interpretation of observed groundwater levels at a single or few boreholes.

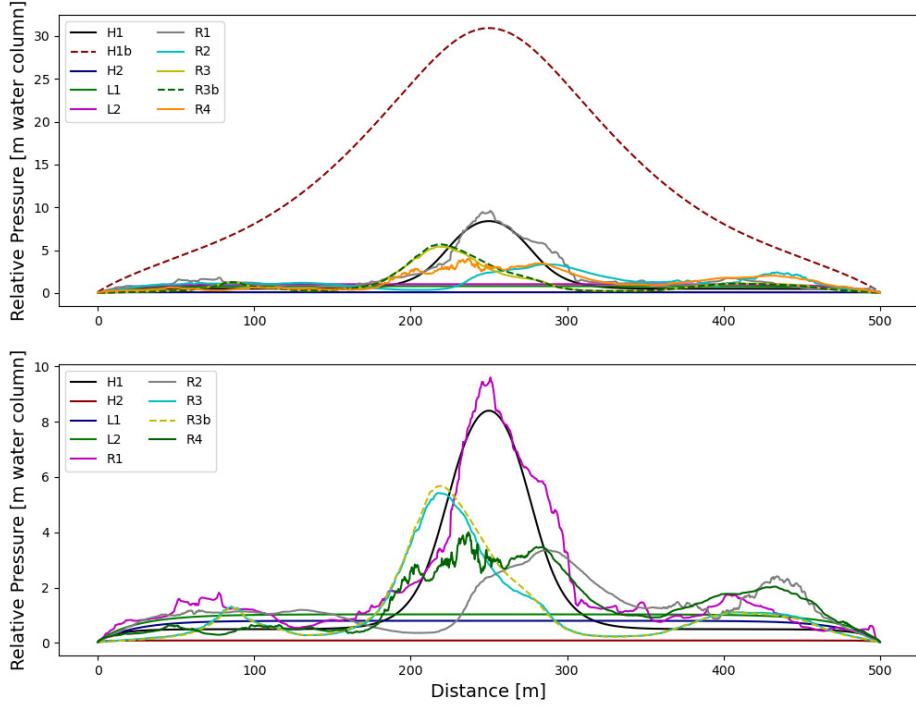


Figure 7: Simulated change in piezometric head along a horizontal profile at $z=20$ m, i.e. initial water table position. Maximum increases in piezometric head for different scenarios expressed as meters of water column are: 8.4(H1), 30.9(H1b), 0.1(H2), 0.8(L1), 1.0(L2), 9.6(R1), 3.4(R2), 5.4(R3), 5.7(R3b) and 4.0(R4). Bottom plot is similar to the top one, except for H1b results.

3.3 Infiltration rates

Figure 8 shows simulated infiltration rates through the riverbed. As predicted by existing theory, the infiltration rate decreases asymptotically from an initial peak value to a final constant value determined by the hydraulic gradient controlled by the simulated river stage and the saturated hydraulic conductivity. The final simulated infiltration rate through the 20 m wide

329 riverbed ranges between 0.1 to 2.4 L/s per meter of the river section (Table
330 3). The time to reach the final value is short for the homogeneous and lay-
331 ered scenarios, but it is significantly longer (≥ 20 hours) for the cases that
332 consider heterogeneous sediments. The exception to the previous statement
333 is Scenario H1b (no flow bottom boundary) that shows a late deviation in
334 infiltration as result of hydraulic control from the outflow boundaries.

335 As expected, the highest infiltration rate corresponds to the scenario with
336 homogeneous Sandy subsurface and the lowest to the scenarios where vertical
337 water flow is controlled by the occurrence of the Silty unit. The scenarios
338 with heterogeneous sediments have intermediate values of infiltration rate,
339 which can be explained by the presence of cells with values of saturated
340 permeability that are higher and/or lower than the mean value assigned to
341 each hydrofacies.

342 It is useful for the interpretation of the final infiltration rates to convert
343 them to equivalent distributed recharge as computed for rainfall. Assuming
344 a 5 km wide valley, the final infiltration rates are equivalent to an areal
345 distributed recharge of up to 160 mm, with a mean value between 60 to 80
346 mm. Such equivalent recharge must be compared to the mean annual rainfall
347 recorded in arid regions where ephemeral streams occur, which is usually less
348 than 100 mm/year for semi-arid areas and even less for arid regions, e.g. less
349 than 5 mm/year in the case of the lower section of the San José River valley in
350 northern Chile. Therefore, the potential recharge to the groundwater during
351 a single flood event can be significant for aquifers in arid regions.

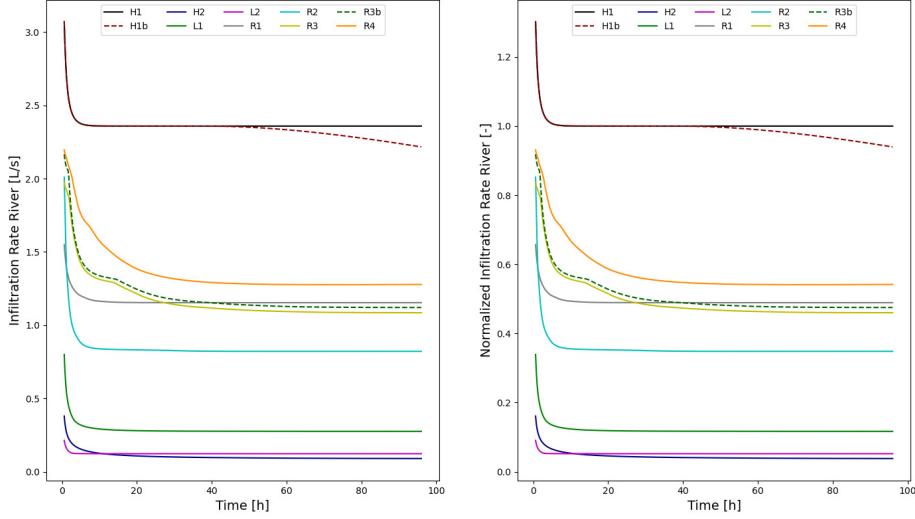


Figure 8: Simulated infiltration rate across the 20 m wide riverbed versus time.

352 The results of the simulations show that the water table rise due to fo-
 353 cused recharge is not necessarily correlated with infiltration rates. For ex-
 354 ample, Scenarios H1 and H1b have similar final infiltration rates (2.4 and
 355 2.2 L/s), but quite different maximum increases in water table: 8.4 and 31.0
 356 m, respectively. On the other hand, final infiltration rates for the scenarios
 357 with heterogeneous sediments show low variability, ranging between 0.8 to
 358 1.2 L/s. However, the water table increase is significantly higher for Scenario
 359 R1 (9.6 m) in comparison to the other three with increases between 3.4 and
 360 5.7 m. This calls for caution when estimating recharge rates from observed
 361 variations in groundwater levels.

362 Assuming a constant river depth during a flood is a simplifying assump-

Run	Final Vertical Infiltration [m/d]	Final Infiltration [L/s/m]	Total Infiltration in 96 hrs [m ³ /m]	5 km [†]	10 km [†]	20 km [†]	Equivalent distributed recharge (mm)
H1	10.20	2.36	816	4.1	8.2	16.3	163
H1b	9.59	2.22	767	3.8	7.7	15.3	153
H2	0.39	0.09	31	0.2	0.3	0.6	6
L1	1.21	0.28	97	0.5	1.0	1.9	19
L2	0.52	0.12	41	0.2	0.4	0.8	8
R1	4.97	1.15	397	2.0	4.0	7.9	79
R2	3.54	0.82	283	1.4	2.8	5.7	57
R3	4.67	1.08	373	1.9	3.7	7.5	75
R3b	4.84	1.12	387	1.9	3.9	7.7	77
R4	5.53	1.28	442	2.2	4.4	8.8	88

Table 3: Final vertical infiltration rate and total infiltration and equivalent distributed infiltration rate assuming a 5 km wide valley. [†]Columns show total infiltration volume in million cubic meters for river sections of different length.

363 tion, since a flood usually consists of two stages: a raising water level period,
 364 followed by one of declining water elevation. Nevertheless, ready water avail-
 365 ability is the factor that facilitates infiltration. Higher or lower water levels
 366 result only in minor changes in surface infiltration as the result of changes
 367 in hydraulic gradient given a higher hydraulic head below the river bed. For
 368 example, Table 4 reports the final infiltration rate for Scenario L2 as a func-
 369 tion of river stage. These results demonstrate that decreasing the river depth
 370 by a factor of 10 produces a $\sim 30\%$ decrease in infiltration rate.

Run	River Depth [m]	Final Infiltration Rate [L/s/m]
L2	1.0	0.12
L2b	0.5	0.10
L2c	0.2	0.09
L2d	0.1	0.08

Table 4: Total infiltration rate considering different river depths.

371 The analysis we present is based on 2D simulations, it is likely that in-
 372 filtration for a 3D setting would be higher because of the extra degree of
 373 freedom for the infiltrated water to flow [19] or, a higher connection of fast
 374 flow pathways in heterogeneous sediments [21].

375 4 Conclusions

376 Based on the results of numerical simulations that considered a few varia-
 377 tions of a unique disconnected aquifer, we can make the following general

378 statements about aquifer recharge in arid regions during flooding events that
379 usually occurred with a few years to decades return period:

380 1. The recharge that may occur during a few days duration event can
381 be significant, of the order of a few times the typical mean annual
382 precipitation registered in those regions. This could explain estimates
383 for recent/modern recharge to aquifers located in arid regions where
384 excess water calculated from mean rainfall and evapotranspiration is
385 negligible or zero.

386 2. Transit times of infiltrated water during focused recharge events, even
387 for aquifers with relatively deep water table (≥ 50 m), can be short,
388 of the order of a few hours. This must be considered when evaluating
389 potential groundwater pollution due to infiltration of poor quality water
390 as results of spills to surface streams or dam failures, e.g. tailings dams
391 located on or near surface streams.

392 3. Infiltration patterns during focused recharge, as it occurs during flood-
393 ing events, may be quite complex to analyze or model for real settings.
394 In particular, numerical simulations that rely on simple representa-
395 tions of the subsurface sediments stratigraphy can provide estimates
396 of recharge, saturation distribution, water table mounding and transit
397 times to the water table; that can be misleading. Therefore, results of
398 such simulations should be used with caution when evaluating water
399 management options or aquifer vulnerability to pollution.

400 4. The complexity of the saturation patterns observed under more realistic
401 assumptions, e.g. multi-scale heterogeneous sediments, may be quite
402 difficult to interpret based on a few point-like measurements such as
403 core samples or water samples taken from cup lysimeters.

404 5. Estimates of recharge based on water level rise, which relies on the
405 concept of specific yield or storage coefficient, may be bogus due to
406 incorrect assumptions about the distribution of the infiltrated water
407 and/or shape of the resulting wetting front.

408 6. Different boundary conditions, e.g. permeable or impermeable bot-
409 tom boundary, may produce large differences in results of numerical
410 simulations of this type of processes. This should be considered as a
411 source of uncertainty in studies that rely on the use of numerical mod-
412 els. Moreover, the presence of that type of boundary conditions in real
413 aquifers should be considered when analyzing field collected data, e.g.
414 piezometric head for estimation of recharge rates.

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419 **References**

420 [1] K. Ali, R. M Bajracharyar, and N. Raut. Advances and challenges in
421 flash flood risk assessment: A review. *Journal of Geography & Natural*
422 *Disasters*, 7(2):1–6, 2017.

423 [2] M. Ascott, D. Lapworth, D. Goody, R. Sage, and I. Karapanos. Impacts
424 of extreme flooding on riverbank filtration water quality. *Science of the*
425 *Total Environment*, 554:89–101, 2016.

426 [3] E. Bonanno, G. Blöschl, and J. Klaus. Discharge, groundwater gradi-
427 ents, and streambed micro-topography control the temporal dynamics
428 of transient storage in a headwater reach. *Water Resources Research*,
429 59(7):e2022WR034053, 2023.

430 [4] J.D. Bredehoeft. The water budget myth revisited: Why hydrogeologists
431 model. *Groundwater*, 40(4):340–345, 2002.

432 [5] J.D. Bredehoeft, S.S. Papadopoulos, and H.H. Cooper. Groundwater:
433 The water budget myth. *Scientific Basis of Water Resource Manage-
434 ment*, 51:57, 1982.

435 [6] P. Brunner, R. Therrien, P. Renard, C.T. Simmons, and H.H. Franssen.
436 Advances in understanding river-groundwater interactions. *Reviews of*
437 *Geophysics*, 55(3):818–854, 2017.

438 [7] D.A. Cenderelli et al. Floods from natural and artificial dam failures.

439 *Inland Flood Hazards: Human, riparian and aquatic communities*, pages
440 73–103, 2000.

441 [8] V.T. Chow, D.R. Maidment, and L.W. Mays. *Applied Hydrology*.
442 McGraw-Hill, 1988.

443 [9] J. Constantz, A.E. Stewart, R. Niswonger, and L. Sarma. Analysis of
444 temperature profiles for investigating stream losses beneath ephemeral
445 channels. *Water Resources Research*, 38(12):52–1, 2002.

446 [10] P.G. Cook, W.M. Edmunds, and C.B. Gaye. Estimating paleorecharge
447 and paleoclimate from unsaturated zone profiles. *Water Resources Re-*
448 *search*, 28(10):2721–2731, 1992.

449 [11] M.O. Cuthbert, R.I. Acworth, M.S. Andersen, J.R. Larsen, A.M. Mc-
450 Callum, G.C. Rau, and J.H. Tellam. Understanding and quantifying fo-
451 cused, indirect groundwater recharge from ephemeral streams using wa-
452 ter table fluctuations. *Water Resources Research*, 52(2):827–840, 2016.

453 [12] G. Dagan. *Flow and transport in porous formations*. Springer-Verlag,
454 1989.

455 [13] G.A. Fox and D.S. Durnford. Unsaturated hyporheic zone flow in
456 stream/aquifer conjunctive systems. *Advances in Water Resources*,
457 26(9):989–1000, 2003.

458 [14] M.K. Gaur and V.R. Squires. Geographic extent and characteristics of

459 the world's arid zones and their peoples. In *Climate variability impacts*
460 *on land use and livelihoods in drylands*, pages 3–20. Springer, 2017.

461 [15] L.W. Gelhar. *Stochastic subsurface hydrology*. Prentice-Hall, 1993.

462 [16] J. Gómez-Hernández and X. Wen. To be or not to be multi-gaussian?
463 A reflection on stochastic hydrogeology. *Advances in Water Resources*,
464 21(1):47–61, 1998.

465 [17] G.E. Hammond, P.C. Lichtner, and R.T. Mills. Evaluating the perfor-
466 mance of parallel subsurface simulators: An illustrative example with
467 pflotran. *Water Resources Research*, 50(1):208–228, 2014.

468 [18] J. Hauser, F. Wellmann, and M. Trefry. Water table uncertainties due
469 to uncertainties in structure and properties of an unconfined aquifer.
470 *Groundwater*, 56(2):251–265, 2018.

471 [19] P.A. Herrera. Numerical simulations of saturated and unsaturated flow
472 in porous media (*Simulaciones numéricas de flujo en medios porosos*
473 *saturados y no saturados*). Undergraduate Thesis, Universidad de Chile,
474 2000.

475 [20] P.A. Herrera and P. Lichtner. Potential groundwater recharge during
476 floods: First-order estimates. *Tecnología y Ciencias Agua*, To appear,
477 2026.

478 [21] A.G. Hunt. Applications of percolation theory to porous media with

479 distributed local conductances. *Advances in Water Resources*, 24(3-
480 4):279–307, 2001.

481 [22] H. Jourde, A. Lafare, N. Mazzilli, G. Belaud, L. Neppel, N. Dörfliger,
482 and F. Cernesson. Flash flood mitigation as a positive consequence of
483 anthropogenic forcing on the groundwater resource in a karst catchment.
484 *Environmental Earth Sciences*, 71:573–583, 2014.

485 [23] X. Li, J.H. Li, and L.M. Zhang. Predicting bimodal soil–water charac-
486 teristic curves and permeability functions using physically based param-
487 eters. *Computers and Geotechnics*, 57:85–96, 2014.

488 [24] K. Madani. Global Water Bankruptcy: Living beyond our hydrological
489 means in the post-crisis era. Technical report, United Nations University
490 Institute for Water, Environment and Health (UNU-INWEH), 2026.
491 DOI: 10.53328/INR26KAM001.

492 [25] A. Mantoglou and J.L. Wilson. The Turning Bands method for simula-
493 tion of random fields using line generation by a spectral method. *Water
494 Resources Research*, 18(5):1379–1394, 1982.

495 [26] G. Matheron and G. de Marsily. Is transport in porous media always
496 diffusive? A counterexample. *Water Resources Research*, 16(5):901–917,
497 1980.

498 [27] Y. Mualem. A new model for predicting the hydraulic conductivity of

499 unsaturated porous media. *Water Resources Research*, 12(3):513–522,
500 1976.

501 [28] S.P. Neuman and V. Di Federico. Multifaceted nature of hydrogeologic
502 scaling and its interpretation. *Reviews of Geophysics*, 41(3), 2003.

503 [29] J. Philip. Theory of infiltration. In *Advances in Hydroscience*, volume 5,
504 pages 215–296. Elsevier, 1969.

505 [30] L.A. Richards. Capillary conduction of liquids through porous mediums.
506 *Physics*, 1(5):318–333, 1931.

507 [31] Y. Rubin. *Applied stochastic hydrogeology*. Oxford University Press,
508 2003.

509 [32] B.R. Scanlon, K.E. Keese, A.L. Flint, L.E. Flint, C.B. Gaye, W.M.
510 Edmunds, and I. Simmers. Global synthesis of groundwater recharge
511 in semiarid and arid regions. *Hydrological Processes*, 20(15):3335–3370,
512 2006.

513 [33] O.S. Schilling, D.J. Irvine, H.J. Hendricks Franssen, and P. Brunner. Es-
514 timating the spatial extent of unsaturated zones in heterogeneous river-
515 aquifer systems. *Water Resources Research*, 53(12):10583–10602, 2017.

516 [34] M. Shanafield and P.G. Cook. Transmission losses, infiltration and
517 groundwater recharge through ephemeral and intermittent streambeds:
518 A review of applied methods. *Journal of Hydrology*, 511:518–529, 2014.

519 [35] M. Shanafield, P.G. Cook, P. Brunner, J. McCallum, and C.T. Simmons.
520 Aquifer response to surface water transience in disconnected streams.
521 *Water Resources Research*, 48(11), 2012.

522 [36] X. Song, X. Chen, M. Ye, Z. Dai, G. Hammond, and J.M. Zachara. De-
523 lineating facies spatial distribution by integrating ensemble data assim-
524 ilation and indicator geostatistics with level-set transformation. *Water*
525 *Resources Research*, 55(4):2652–2671, 2019.

526 [37] R. Taylor, B. Scanlon, P. Döll, M. Rodell, R. Van Beek, Y. Wada,
527 L. Longuevergne, M. Leblanc, J. Famiglietti, M. Edmunds, et al. Ground
528 water and climate change. *Nature Climate Change*, 3(4):322–329, 2013.

529 [38] A.F.B. Tompson, R. Ababou, and L.W. Gelhar. Implementation of the
530 three-dimensional turning bands random field generator. *Water Re-*
531 *sources Research*, 25(10):2227–2243, 1989.

532 [39] D. Tonina and J.M. Buffington. Hyporheic exchange in mountain rivers
533 i: Mechanics and environmental effects. *Geography Compass*, 3(3):1063–
534 1086, 2009.

535 [40] S.W. Tyler, J.B. Chapman, S.H. Conrad, D.P. Hammermeister, D.O.
536 Blout, J.J. Miller, M.J. Sully, and J.M. Ginanni. Soil-water flux in the
537 southern Great Basin, United States: Temporal and spatial variations
538 over the last 120,000 years. *Water Resources Research*, 32(6):1481–1499,
539 1996.

540 [41] N.P. Unland, I. Cartwright, E. Daly, B.S. Gilfedder, and A.P. Atkinson.

541 Dynamic river–groundwater exchange in the presence of a saline, semi-

542 confined aquifer. *Hydrological Processes*, 29(23):4817–4829, 2015.

543 [42] M.Th. van Genuchten. A closed-form equation for predicting the hy-

544 draulic conductivity of unsaturated soils. *Soil Science Society of Amer-*

545 *ica Journal*, 44(5):892–898, 1980.

546 [43] A.M. Youssef, M.M. Abu-Abdullah, E.A. AlFadail, H.D. Skilodimou,

547 and G.D. Bathrellos. The devastating flood in the arid region a conse-

548 quence of rainfall and dam failure: Case study, Al-Lith flood on 23th

549 November 2018, Kingdom of Saudi Arabia. *Zeitschrift Für Geomor-*

550 *phologie*, 63:115–136, 2021.

551 **Appendix**

552 PFLOTRAN [17], available at <https://pflotran.org/>, simulates unsaturated
 553 flow solving the following form of Richards Equation [30]:

$$\frac{\partial}{\partial t} (\phi s \eta) + \nabla \cdot (\eta \mathbf{q}) = Q_w, \quad (\text{A.1})$$

554 with porosity ϕ [m^3/m^3], saturation s [-], molar water density η [kmol/m^3],
 555 source/sink term Q_w [$\text{kmol}/\text{m}^3/\text{s}$], and Darcy flux \mathbf{q} [m/s]

$$\mathbf{q} = -\frac{kk_r(s)}{\mu} \nabla (P - \rho g z) \quad (\text{A.2})$$

556 where k is permeability [m^2], k_r is relative permeability [-], μ is dynamic
 557 viscosity [Pa s], P is pressure [Pa], ρ is mass water density [kg/m^3], g is
 558 gravity [m/s^2] and z is elevation. Solving (A.1) requires the definition of
 559 a couple of closure models for linking water content θ or saturation $s =$
 560 $(\theta - \theta_r)/(\phi - \theta_r)$, where θ_r is residual water content; to relative permeability
 561 and pressure or capillary pressure. Those relations are commonly referred to
 562 as unsaturated flow properties.

563 We considered the properties of two soils samples collected in sites in
 564 northern Chile considered in the past for implementing artificial recharge
 565 projects [19, and references therein] in the numerical simulations. One of the
 566 samples referred to as Silty, was collected near the lower section of the San
 567 José River Valley; while the second one (Sandy) was collected in a valley

568 located 1000 km south, near the city of Copiapó. Both sites are located in
 569 fluvial-alluvial valleys in extreme arid areas: the mean annual rainfall in
 570 Arica is < 5 mm/yr and < 24 mm/yr in Copiapó according to data taken
 571 from the Climate Explorer at <https://explorador.cr2.cl/>.

572 The unsaturated flow properties of the samples were characterized through
 573 laboratory tests. The interpretation of the results considered the van Genuchten
 574 model for the water content versus suction relation [42],

$$\theta(h) = \theta_r + \frac{\theta_s - \theta_r}{[1 + |\alpha h|^n]^m} \quad (\text{A.3})$$

575 and the Mualem model for relative permeability [27],

$$k_r(h) = Se^{0.5}[1 - (1 - Se^{1/m})^m]^2 \quad (\text{A.4})$$

576 where $Se = \frac{\theta - \theta_r}{\theta_s - \theta_r}$ is known as residual saturation and $m = 1 - \frac{1}{n}$. Table A.1
 577 summarizes the parameter values considered in the numerical simulations,
 578 while Figure A.1 shows the resulting curves.

Parameter	Sandy	Silty
Saturated water content, θ_s [-]	0.33	0.46
Residual water content, θ_r [-]	0.19	0.01
Inverse of entry pressure, α [Pa ⁻¹]	4.9x10 ⁻⁵	1.8x10 ⁻⁵
Distribution index, n [-]	2.45	1.35
$m = 1 - 1/n$	0.59	0.26
Saturated hydraulic conductivity, K_{sat} [m/day]	7.0	0.3

Table A.1: Unsaturated flow parameters considered in the numerical simulations.

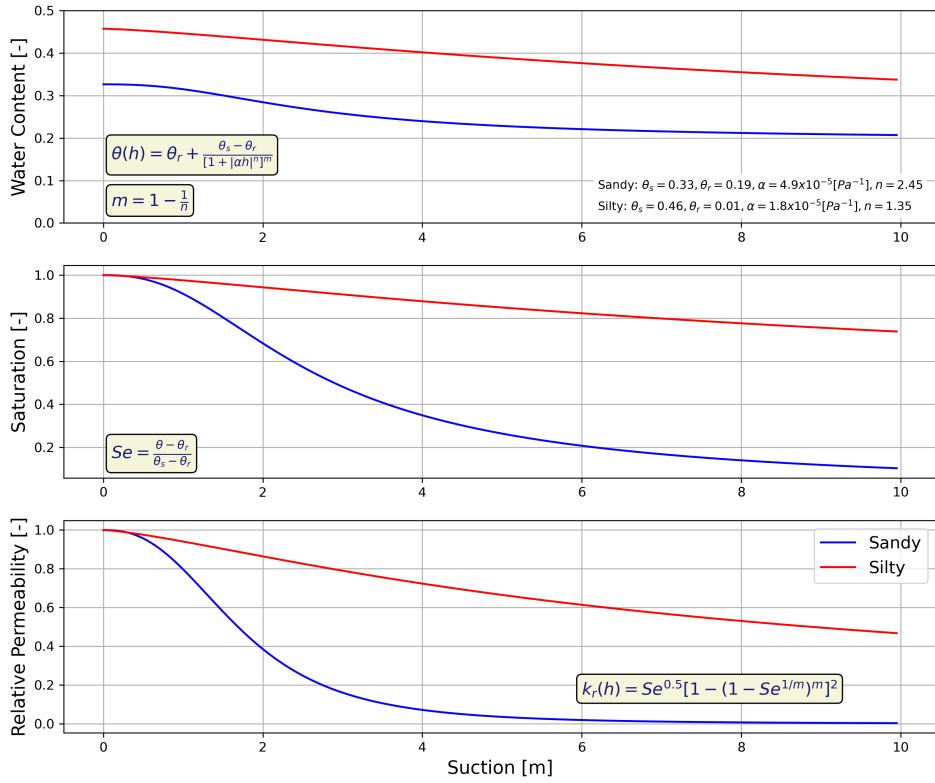


Figure A.1: Water retention and relative permeability for sandy and silty soils considered in numerical simulations [19].

579 **Conflict of interests**

580 The authors declare that they have no known competing financial interests
581 or personal relationships that could have appeared to influence the work
582 reported in this paper.

583 **Author contribution**

584 P.A.H. was in charge of initial conceptualization, numerical simulations and
585 writing of initial draft; P.C.L. provided comments to conceptualization and
586 review initial draft.