

Climate Change and Curtailment: Evaluating Water Management Practices in the Context of Changing Runoff Regimes in a Snowmelt-Dominated Basin

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1 **Abstract**

2 Climate change directly affects the hydrologic cycle in mountainous watersheds, which has
3 consequences for downstream users. Improved water projections under diverse potential climate
4 futures are critical to improving water security and management in these watersheds. The hydro-
5 logic science researchers and water resource managers, however, often focus on different metrics
6 of flow regimes in changing climates. The research community tends to more closely focus on
7 biophysical state and flux variables of the hydrologic system. Managers, meanwhile, tend to fo-
8 cus on key administrative benchmarks that govern the operation of complex water storage and
9 distribution systems. Here, we examine potential hydrologic changes in a water supply basin in
10 the western United States in the context of both biophysical states and fluxes, as well as from
11 the perspective of how those changes map onto key variables that govern the administration of
12 water resources in the region. The study site consists of the Upper Boise River Basin, ID. This
13 snowmelt-dominated, mountainous watershed that supplies water to a semi-arid, agriculturally
14 intensive and rapidly urbanizing region. Using the Envision integrated modeling framework, we

15 created a hydrologic model and simulated hydrologic response to the year 2100 using six diverse
16 climate scenarios. Annual discharge increased from historical values by an average of 13% across
17 all climate scenarios with a range of increase of 6-24%, reflecting an increase in the precipitation
18 in the climate projections. Runoff timing was altered, with peak discharge occurring 4-33 days
19 earlier and center of timing of streamflow occurring 4-17 days earlier by midcentury. Examining
20 potential changes in the date junior water rights holders begin to be curtailed regionally (the Day
21 of Allocation), we found that the Day of Allocation occurs up to 14 days earlier by 2100 across all
22 climate scenarios, with one scenario suggesting this date could occur over a month earlier. These
23 results suggest that current methods and policies of water rights accounting and management
24 may need to be revised moving into the future.

25 **Keywords:** Climate Change; Runoff Regime; Snowmelt; Water Management; Water Rights;
26 Day of Allocation; Flood Control; Water Supply

27 **1 Introduction**

28 Climate change exerts a significant control on global hydrologic regimes by influencing the
29 timing, magnitude, phase, and seasonal variability in precipitation (Mote *et al.*, 2005; Regonda
30 *et al.*, 2005; Knowles *et al.*, 2006; Haddeland *et al.*, 2014). Changes in temperature further influence
31 how that precipitation moves through a watershed by affecting snowmelt timing, soil moisture,
32 and evapotranspiration rates (Barnett *et al.*, 2005; Li *et al.*, 2017). While there is general consensus
33 among scientists that the Earth is warming and will continue to do so, there remain significant
34 uncertainties regarding the impacts of global warming on the water cycle and how those changes
35 will be distributed regionally in the future (Huntington, 2006; Turrall *et al.*, 2011).

36 Significant changes in the water cycle can have serious consequences for water users and man-
37 agement across many sectors. It is estimated that more than two billion people currently live in
38 highly water-stressed regions (Oki, 2006), with this number projected to increase in the future
39 (Schewe *et al.*, 2014). Agriculture is vulnerable to changes in hydrologic regimes, especially in

40 regions that rely on surface water resources for irrigation and in rain-fed systems (Turrall *et al.*,
41 2011). Flooding could intensify, putting stress on current water management infrastructure as
42 well as lessening the effectiveness of hydropower generation as runoff arrives earlier (Markoff and
43 Cullen, 2008). Despite the seriousness of the potential impacts of hydrologic changes across sec-
44 tors, the effectiveness of current water management systems, practices, and policies under chang-
45 ing hydrologic regimes is not well understood.

46 Many previous modeling studies have investigated how water resources will respond to cli-
47 mate change in snowmelt-dominated systems (Adam *et al.*, 2009; Jin and Sridhar, 2011; Ficklin
48 *et al.*, 2013; Gergel *et al.*, 2017). However, results from such studies are not always presented in a
49 way that is usable to water managers and users. Here we provide an example of how hydrologic
50 modelers can generate results that may provide additional meaning for management decisions.
51 Managers of these systems tend to focus on the ways in which climate variability and change
52 will challenge existing water management protocols and practices. For example, in the American
53 West, there are often hierarchies of water rights users who may be affected differently by projected
54 changes in water availability (Vicuna *et al.*, 2007). Providing predictions more applicable to water
55 users requires more in-depth and location-specific knowledge of water management and distri-
56 bution but has the potential to provide more relevant information to a wider group of audiences.

57 Snowmelt-dominated systems, particularly those in the western U.S., are especially vulnerable
58 to climate change (Barnett *et al.*, 2005; Stewart, 2009; Li *et al.*, 2017). Significant reservoirs, in the
59 form of snow, develop at times (i.e., winter) and locations (i.e., high elevations) where that water
60 cannot be used to grow crops and produce hydroelectricity. This snowpack at high elevations
61 provides a natural reservoir that holds water in reserve and, ideally, slowly releases it into the
62 spring and summer, into downstream agricultural areas. A complex system of water rights and
63 management has been developed, and reservoir and canal systems engineered to store springtime
64 runoff, mitigate flooding, and direct it to other locations when there is a demand for irrigation.
65 This current system of water management infrastructure and protocols are set up to account for
66 the historical range of hydrologic variability; however, it may not be adequate to adapt to future

67 hydrologic regimes (Palmer *et al.*, 2008). With sufficient changes in the timing and magnitude of
68 water delivery, as is projected with climate change, current management practices may be inad-
69 equate to meet the dual needs of flood control and late-season irrigation demand (Barnett *et al.*,
70 2005). However, it is uncertain to what extent current management practices may be stressed un-
71 der future hydrologic regimes or when water management agencies can expect existing practices
72 and policies to begin coming into conflict with the reality of altered runoff regimes.

73 The overarching objective of this study is to better understand and quantify how climate
74 change will impact future water resources and water management in the context of metrics that
75 managers monitor and use to implement policy. We perform our study in the Upper Boise River
76 Basin, ID, an ideal location because it is a relatively undisturbed high mountain watershed that is
77 managed to provide water resources to an agriculturally-intensive and rapidly urbanizing region.
78 We explore this connected biophysical and social system by combining a surface water hydrologic
79 model with diverse climate projections to project potential changes in future regional hydrologic
80 regimes. Furthermore, we translate our model outputs into a metric that is directly applicable to
81 downstream water users and managers. Our specific research objectives are to:

- 82 1. Identify a range of climate projections and assess how they affect hydrologic parameters
83 such as center of timing of streamflow, volume of annual water delivery, and snowpack
84 levels through the end of the century; and
- 85 2. Identify how these changes in hydrologic regimes impact an associated metric that charac-
86 terizes water storage and is used to enforce water rights accounting policies.

87 What follows in this paper is: (1) a more detailed description of the study area, (2) an overview
88 of our methodological approach, (3) results of this study, and (4) discussion, implications, and
89 conclusions.

90 2 Methods

91 2.1 Study Area

92 The Upper Boise River Basin (UBRB) is located in southwest Idaho (Figure 1) and supplies
93 water for downstream users in the populated Boise metropolitan region. This watershed en-
94 compasses an area of 6,935 km² with elevation ranging from approximately 930 to 3,000 m. It
95 is bounded by the Sawtooth range in the east, the Payette River Basin to the north, and the Snake
96 River Plain to the southwest. We delineated the study area by combining three Hydrologic Unit
97 Code (HUC) 8 watersheds: the North and Middle Forks Boise (17050111), the South Fork Boise
98 (17050113), and Boise-Mores (17050112). Due to the large variation in topography throughout the
99 study area, regions shift from semi-arid grasslands and shrublands in the lowlands to coniferous
100 forests in the highlands. In the UBRB, the dominant land covers are forest (43.0%), shrubland
101 (34.6%) and grassland (20.9%), with sparse human development within the watershed. The cli-
102 mate in this region is a continental Mediterranean climate (Köppen *Dsb*) with cold winters, warm
103 summers, and the majority of precipitation falling in winter as snow. The overall average precip-
104 itation is ~800 mm, with averages ranging from ~400 mm at low elevations to over 1300 mm at
105 high elevations (Daly *et al.*, 2008).

106 The UBRB is the primary source of water for the downstream Treasure Valley region, which
107 contains the state's three largest cities (Boise, Nampa, and Meridian) and roughly 40% of the
108 state's total population. The Treasure Valley is an agriculturally intensive region and contains
109 approximately 1300 km² of farmlands, many of which rely on irrigation water from the UBRB.
110 Like many other snowmelt-dominated watersheds in the West, the UBRB is heavily managed via
111 three large storage reservoirs to fulfill the needs of flood control and downstream uses, especially
112 for direct consumption in the Treasure Valley. Similar to other western states, water rights in this
113 region follow the Prior Appropriation Doctrine, also known as "first in time – first in right." This
114 doctrine states that the earliest beneficial users (i.e., senior water rights) retain their full water

115 right, and those that came later (i.e., junior water rights) may retain their water rights as long
116 as they do not infringe on those that came beforehand. As such, many junior water rights are
117 curtailed during low water years, as total surface water rights in the Treasure Valley surpass 14,000
118 ft³/s, far exceeding the natural flow of the Boise River.

119 Previous studies indicate that the UBRB has already begun to respond hydrologically to cli-
120 mate change, noting an increase in summer streamflow temperatures (Isaak *et al.*, 2010), earlier
121 timing of streamflow (Clark, 2010), lengthened growing season (Kunkel, 2004), and declining ex-
122 treme low flow discharges (Kormos *et al.*, 2016). Additionally, there have been previous modeling
123 studies that have used this basin to anticipate changes in hydrology under climate change (Still-
124 water, 2008; Jin and Sridhar, 2011). However, both of the aforementioned studies used an older
125 generation of global climate models as their climate input and calibrated their models to stream-
126 flow alone. This study extends those previous works by making use of climate projections from
127 the 5th Coupled Model Intercomparison Project (CMIP5, Taylor *et al.*, 2012), calibrating the hydro-
128 logic model to multiple hydrologic metrics, and producing results that may provide additional
129 meaning to water users.

130 **2.2 Modeling Framework**

131 Here we employ the Envision framework, a multiagent-based, spatially explicit modeling
132 framework, to examine how regional hydrology may change with climate. Envision was cre-
133 ated to examine relationships between human and natural environmental systems by integrating
134 scenarios, data, and component models to assess regional landscape change (Bolte *et al.*, 2007). To
135 this end, the modeling framework and software infrastructure of Envision support the integra-
136 tion of a variety of social and biophysical models in a spatiotemporally dynamic way. It is freely
137 available and users can extend and enhance model capabilities by adding additional models as
138 plugins. It has been extensively used recently in a wide variety of studies, from understanding
139 urbanization impacts on streamflow (Wu *et al.*, 2015) to projecting climate change impacts of land
140 cover and land use (Turner *et al.*, 2015), and even to understand when fire occurrence and size is

141 'surprising' (Hulse *et al.*, 2016). Additionally, it has been used to integrate water rights to spatially
142 allocate irrigation in the agriculturally intensive region below the UBRB (Han *et al.*, 2017).

143 In this study, we use Envision version 6.197 and utilize the Flow extension to model future
144 hydrology under various climate scenarios. In the following sections, we provide an overview of
145 the modeling structure and the inputs needed for the various components.

146 2.2.1 Spatial Coverage in Envision

147 In Envision, the most refined spatial elements where model algorithms are applied are referred
148 to as Integrated Decision Units (IDUs). The size and geometry of these polygons are dependent
149 on the type of modeling being performed and the geospatial datasets required as input to those
150 models. As such, there is no universally accepted method for creating IDU coverage. In this study,
151 we used three datasets to form the IDU geometry: surface management agency, land cover, and
152 HUC 12 stream catchments (Table 1). As such, the IDU coverage will preserve boundaries be-
153 tween HUC 12 catchments, cognizant land management agencies, as well as boundaries between
154 vegetation classes.

155 The datasets were processed in ArcMap 10.1. To shorten Envision's computation time, we
156 coarsened the land cover dataset from 30 to 100 m in increments of 10 m. We used a nearest neigh-
157 bor algorithm to resample land cover types to more accurately capture the original distribution
158 of coverage in the land cover dataset. The other two datasets were polygon geospatial datasets
159 that required very little processing besides renaming attributes to be consistent with the Envision
160 framework requirements.

161 We created our IDU coverage by intersecting the three aforementioned datasets, creating 31,625
162 polygons. We extracted the average elevation for each IDU and also assigned an elevation class
163 from 1-4, corresponding to 0-1500, 1500-2000, 2000-2500, and >2500 meters to allow binning and
164 analysis of results by elevation band. Additionally, to aid in analysis and querying we created a
165 three-tiered hierarchy of land cover classification ranging from general (e.g. Natural Vegetation)
166 to more specific (e.g. Evergreen Forest), which was formed by grouping NLCD classifications that

167 are similar (Figure 2).

168 The hydrologic model in Envision applies algorithms to Hydrologic Response Units (HRUs, Jin
169 and Sridhar, 2011; Turner *et al.*, 2016), which are an aggregation of IDUs that would theoretically
170 behave hydrologically similar. To create the HRU coverage, we grouped polygons that had the
171 same intermediate land cover (Figure 2), identical elevation class, and were located in the same
172 HUC-12 catchment. This resulted in 9,465 HRUs.

173 2.2.2 Hydrologic System Model

174 An extension in Envision called Flow provides flexibility in modeling hydrology and the use of
175 different model representations of hydrologic processes. In this study, we used a modified version
176 of the HBV (Hydrologiska Byråns Vattenbalansavdelning) rainfall-runoff model (Bergström, 1976)
177 for surface hydrology. HBV is a commonly used conceptual model (Seibert, 2000; Woodsmith *et al.*,
178 2007; Abebe *et al.*, 2010; Bergström and Lindström, 2015) but has been modified by Envision's
179 developers to be spatially distributed. Each HRU is conceptualized as a linked reservoir with five
180 layers of storage: snowpack, lakes, soil, upper groundwater, and lower groundwater (Figure 3).
181 Runoff from each HRU is routed to streams using HUC12 flowlines from NHDplus V2 (Table 1).
182 The water balance in Flow is described by the following equation:

$$P - ET - Q = \frac{d}{dt} [SP + SM + UZ + LZ + lakes] \quad (1)$$

183 where P is precipitation [mm/d], ET is evapotranspiration [mm/d], Q is runoff [mm/d],
184 SP is snow storage [mm], SM is soil moisture storage [mm], UZ is upper groundwater storage
185 [mm], LZ is lower groundwater storage [mm], and $lakes$ refers to lake storage [mm]. A more
186 thorough description of the HBV model can be found in other papers (Seibert, 1999; Bergström
187 and Lindström, 2015) and a more detailed description of Flow can be found on Envision's website
188 (<http://envision.bioe.orst.edu/>).

189 Evapotranspiration (ET) is calculated via a modified Penman-Monteith approach described in

190 the Food and Agriculture Organization’s Irrigation and Drainage paper 56 (FAO56) where a crop
191 coefficient is applied to the ET of a reference plant (Allen *et al.*, 1998) and was later developed
192 specifically for Idaho (Allen and Robison, 2007) using the following equation:

$$ET = ET_r \cdot K_c \quad (2)$$

193 where ET = evapotranspiration, ET_r = reference evapotranspiration (alfalfa, for Idaho), and K_c
194 = crop coefficient.

195 We used this equation and applied crop coefficient curves that either matched our land cover
196 type directly or estimated crop coefficient curves based upon similarities of crops to land cover
197 types (Table 2). Crop coefficients were obtained from AgriMet and (Allen and Robison, 2007),
198 with a few modified land cover coefficients from (Inouye, 2014).

199 **2.3 Climate Inputs**

200 We used statistically downscaled climate data using the MACA (Multivariate Adaptive Con-
201 structed Analogs) method version 1.0 for both historic and future simulations (Abatzoglou and
202 Brown, 2011). This data has a spatial resolution of 4 km across the continental U.S. and is avail-
203 able daily for 1950-2100. Downscaled data is available for 20 Global Climate Models (GCMs) from
204 CMIP5 for both Representative Concentration Pathway (RCP) 4.5 and 8.5 scenarios. RCPs are a
205 consistent set of projections that are named according to their additional radiating forcing level at
206 2100, such that RCP 4.5 equates to +4.5 W/m² radiative forcing relative to pre-industrial values
207 by the end of the century (van Vuuren *et al.*, 2011).

208 For future simulations, we selected GCMs based upon two criteria. First, we halved our GCM
209 selection to models that performed relatively well when ran over the historical period in the Pacific
210 Northwest region (Rupp *et al.*, 2013), meaning they produced less relative error when compared
211 across several metrics. Secondly, we selected GCMs that captured the range of variability between
212 models as it related to changes in precipitation and temperature (Figure 4). We selected three

213 climate models: CanESM2 (hotter, wetter), CNRM-CM5 (warmer, slightly wetter), and GFDL-
214 ESM2M (less warm, drier), and ran each one for RCP 4.5 and 8.5 scenarios, which resulted in
215 six total future climate scenarios (Figure 5). Table 3 provides a naming convention for these six
216 future climate scenarios to ease in discussing results and implications. For historical simulations
217 from 1980-2014, we used a historical climate dataset, METDATA (Abatzoglou, 2011), which was
218 developed using data from the North American Land Data Assimilation System Phase 2 (NLDAS-
219 2, Mitchell, 2004) and from the Parameter-elevation Regressions on Independent Slopes Model
220 (PRISM, Daly *et al.*, 2008).

221 The downscaled variables Envision requires for Flow are daily maximum, minimum, and aver-
222 age temperature, precipitation amount, specific humidity, daily downward shortwave radiation,
223 and wind speed. To format the variables for Envision, the following procedure was followed:
224 (1) subset data to the specified region, (2) convert units and rename variables where needed, (3)
225 compute average temperature as the average between minimum and maximum temperature, (4)
226 calculate overall wind speed from the eastward and northward components provided by MACA,
227 and (5) subset into annual files. Scripts created for pre-processing MACA climate data are avail-
228 able online at https://github.com/asteimke/MACA_EnvisionClimate.

229 **2.4 Calibration and Validation**

230 HBV is a semi-conceptual model, and as such, parameters required as input to the model are
231 obtained through calibration because most parameters cannot be physically measured (Bergström
232 and Lindström, 2015). Numerous combinations of parameter values can yield equally good re-
233 sults (i.e. the equifinality issue, Beven, 2006; Gupta *et al.*, 2005), which makes it difficult to select
234 the best parameter set. To combat this issue, some studies (Madsen, 2003; Inouye, 2014) build
235 an objective function to find an adequate parameter set based on the type of information they
236 want to yield from the model (e.g. streamflow volume, timing, snowpack, etc.). Typically, the
237 calibration-validation procedure takes the form of a data-denial experiment. The model is run
238 over a calibration period to select best parameter sets and then re-run over a validation period to

239 ensure that the selected parameter set performs well during this period for which data was not
 240 used to calibrate the model.

241 Fourteen parameters are included within the HBV model and govern rates of exchange be-
 242 tween reservoirs. We held five of them constant, while the remaining nine were calibrated. *CFR*
 243 and *CWH* are insensitive parameters and were held constant as is often done in HBV applications
 244 (Seibert, 1997). While many of the parameters are conceptual and cannot be measured, three of
 245 them are based on physical properties, so we fixed those parameters to better represent the reality
 246 of our study area. We used the Global Gridded Surfaces of Selected Soil Characteristics (IGBP-
 247 DIS) dataset (Hope and Peck, 1994) and took the average of values for the study area. We used the
 248 following datasets from IGBP-DIS: soil field capacity, soil profile available water capacity, and soil
 249 wilting point for the parameters *FC*, *LP*, and *WP*, respectively (Table 4). In each model run, we
 250 randomly selected the remaining nine parameters from a uniform distribution between ranges of
 251 possible values (Table 4) defined based on previous studies (Inouye, 2014; Han *et al.*, 2017).

252 We ran the model for 1000 simulations at a daily time step over the years 1988-2000 (12 years +
 253 1 spin-up year). We selected this time interval for calibration because it encompasses a reasonably
 254 long time period and includes both wet and dry years. We compared model output to historical
 255 stream discharge records from three long-term USGS gaging stations and snowpack observations
 256 from nine SNOTEL (SNOW TELEmetry) stations, omitting all leap days from these datasets (Table
 257 5). For each run, we calculated the Nash-Sutcliffe Efficiency (*NSE*, Nash and Sutcliffe, 1970),
 258 $\log NSE$, and a volume error (*VE*) using the following equations:

$$NSE = 1 - \frac{\sum_{t=1}^T (Q_{obs}^t - Q_{sim}^t)^2}{\sum_{t=1}^T (Q_{obs}^t - \overline{Q_{obs}})^2} \quad (3)$$

$$\log NSE = 1 - \frac{\sum_{t=1}^T (\ln Q_{obs}^t - \ln Q_{sim}^t)^2}{\sum_{t=1}^T (\ln Q_{obs}^t - \ln \overline{Q_{obs}})^2} \quad (4)$$

$$VE = \frac{\sum_{t=1}^t (Q_{obs}^t - Q_{sim}^t)}{\sum_{t=1}^t (Q_{obs}^t)} \quad (5)$$

259 where Q_{obs} is the observed value and Q_{sim} is the simulated value at each daily time step.

260 NSE coefficients range from $-\infty$ to 1, with 1 indicating a perfect fit of the model to the ob-
 261 served data, and a value of $NSE > 0$ indicating the model is a better predictor than the historically
 262 observed mean. Typically, a model is deemed satisfactory if the NSE is larger than 0.5 (Moriasi
 263 *et al.*, 2007). The logarithmic form of the NSE also ranges from $-\infty$ to 1, but is more sensitive to
 264 low flow and still reacts to peak flows (Krause *et al.*, 2005). The volume error provides insight into
 265 whether the model overestimates ($VE < 0$) or underestimates ($VE > 0$) total volume, with a value
 266 closest to 0 being ideal.

267 We created an objective function to select the best-performing parameter set and was devel-
 268 oped based on work by (Inouye, 2014):

$$Obj = \frac{1}{3} (NSE_G) + \frac{1}{3} (\log NSE_G) + \frac{1}{3} (NSE_S) - 0.2 \cdot |VE_G| \quad (6)$$

269 where NSE_G is the Nash-Sutcliffe Coefficient of discharge weighted by an areal average of the
 270 gauges, VE_G is the volume error for the gauges weighted by an areal average, and NSE_S is the
 271 averaged Nash-Sutcliffe Coefficient for SWE (snow water equivalent) for all SNOTEL sites.

272 The objective function ideally is as close to 1 as possible, as we wish to maximize NSE and
 273 minimize volume bias. The top 1% best performing parameter sets were run over the eight-year
 274 validation period (2001-2008) and the set that performed on average the best in both calibration
 275 and validation years was chosen for our model. Results of the calibration/validation exercises are
 276 reported in the Results section of this manuscript.

277 2.5 Evaluating Climate Change Impacts

278 To assess the potential impact of climate change on hydrologic regimes, we examined three
 279 broad metrics: streamflow, snowpack, and water management. A more detailed description of

280 methods for these metrics is described here.

281 2.5.1 Streamflow

282 While Envision has the capability to examine discharge values anywhere along its stream net-
283 work, we focused here on the aggregation of streamflows for the basin. In all cases, unless men-
284 tioned otherwise, streamflow results are for the unregulated discharge on the Boise River occur-
285 ring at the location of Lucky Peak Dam’s outlet, i.e. the pourpoint of the watershed (Figure 1).
286 This modeled streamflow, as well as daily values for the three major tributaries, can be obtained
287 online (Steimke *et al.*, 2017).

288 To assess climate change impacts on streamflow, we looked at changes in the amount and
289 timing of discharge. An additional metric we used was the center of timing (CT) of streamflow,
290 which is the date when half of the annual volume of water during the water year has arrived
291 at a specified location. We calculated the CT for historical data and future simulations with the
292 following equation (Stewart *et al.*, 2005):

$$CT = \frac{\sum (t_i Q_i)}{\sum Q_i} \quad (7)$$

293 where t_i is the time in days from the start of the water year (October 1) and Q_i is the discharge
294 for that date.

295 2.5.2 Snowpack

296 To assess climate impacts on the basin’s snowpack, we looked at averaged values over three
297 elevation zones: low (1500-2000 m), medium (2000-2500 m), and high (2500+ m) zones. These
298 zones cover 43.4%, 25.8%, and 6.9% of the area of interest, respectively. We do not show results
299 for elevations less than 1500 m as the lowest SNOTEL station to aid in calibration is the Prairie site
300 at 1463 m. Within these three zones, we examine the dates and magnitudes of when SWE is at its
301 maximum, as well the April 1 SWE amount. Water managers have historically used the amount

302 of SWE on this date as an indicator for water availability in the upcoming year, as it has correlated
303 well with maximum SWE at many SNOTEL sites in the West historically (Bohr and Aguado, 2001).

304 2.5.3 Water Management

305 Since 1986, water managers annually declare a Day of Allocation (DOA) in the Lower Boise
306 River Basin for the purpose of water rights accounting during the irrigation season (April – Oc-
307 tober). This day is declared on or after the date of maximum reservoir fill and once natural flow
308 is less than irrigation demand (Memo from IDWR Technical Hydrologist Liz Cresto to IDWR Di-
309 rector Gary Spackman, November 4, 2014, Subject: Accounting for the distribution of water to
310 the federal on-stream reservoirs in Water District 63). The DOA occurs after peak runoff and has
311 been shown historically to typically occur once the natural flow of the Boise River at Lucky Peak
312 reaches below 4000 ft³/s (Garst, 2017), or 113.3 m³/s (Figure 6), which is roughly equivalent to
313 the diversion demand of the river. It is beneficial for farmers if the DOA occurs later in the season
314 because after the DOA is declared water rights begin to be curtailed, starting with the junior-
315 most water rights holders. While the term DOA is unique to three major river basins in Idaho
316 (i.e. Boise, Payette, and Upper Snake river basins), many western states have similar methods for
317 appropriating water as the irrigation season begins.

318 To predict how the DOA may change in our modeled scenarios, we assume that diversion
319 rights will continue to be approximately 113.3 m³/s. We model our DOA date by finding the last
320 day during peak runoff during the irrigation season that flow is greater than 113.3 m³/s and select
321 the day after. We then manually observe the hydrographs and the DOA selected to ensure we are
322 capturing a date on the downfalling limb of peak runoff and not a later season event. If a later
323 season event was modeled, then we manually select the date on which modeled flow falls below
324 113.3 m³/s during the recession limb of spring runoff. We ran the model during the historical
325 period to investigate how well the model reproduces historical DOA using this definition, which
326 provides confidence in our interpretation of DOA changes in modeled future scenarios.

327 **3 Results**

328 **3.1 Calibration and Validation**

329 We calibrated and validated the model using historical records from three USGS gauges and
330 nine SNOTEL sites. The parameter set that performed best had an objective function score of 0.63
331 and 0.62 for calibration and validation periods, respectively (Table 6). We averaged the NSE for
332 each gauge by its respective drainage area, which resulted in a *NSE* of 0.71 and 0.70 for calibration
333 and validation, respectively. However, it should be noted that Mores Creek on its own achieved
334 a lesser *NSE* of 0.58, which is potentially due to this smaller watershed exhibiting some major
335 differences from the other two (notably lower elevation, less precipitation, and less steepness).

336 Among all gauges, we see relatively good agreement between the model simulations and ob-
337 served flow for the historic period (Figure 7), although the model frequently under predicts the
338 magnitude of peak flows at all gauge sites and over predicts baseflow at Mores Creek. While the
339 unregulated flow for the Boise River at Lucky Peak (Table 5) was not used to calibrate the model,
340 we used this as an additional verification dataset to ensure accuracy of the model. With the cho-
341 sen parameter set, we achieved a *NSE* at this site of 0.74 and *VE* of -0.01 averaged over the entire
342 calibration and validation period, providing additional confidence in our model.

343 **3.2 Streamflow**

344 **3.2.1 Annual Discharge**

345 In all future climate scenarios, we see an increase in the median annual discharge from the
346 Boise River (Figure 8). By midcentury (2040-2069), all climate scenarios showed an increase in an-
347 nual discharge over historical (1950-2009) averages, with an average increase of 13% and ranges
348 of increase from 6-24%. RCP 8.5 climate scenarios showed a greater rate of increase over RCP
349 4.5 scenarios. Because our hydrologic model did not perform well historically in accurately cap-

350 turing the magnitude of peak discharges, we do not have adequate confidence to predict future
351 magnitudes in peak or low flows.

352 **3.2.2 Timing of Discharge**

353 While we see some changes in the volume of annual discharge, streamflow is also projected to
354 arrive at significantly different times than in the historical past. However, these arrival times vary
355 greatly between different climate models.

356 In most future climate scenarios, the date of peak discharge occurs earlier in the season, with
357 an increase in early winter flooding events (Figure 9). In extreme climate cases (i.e. C-85), the
358 average peak discharge occurs approximately 45 days earlier in the period 2040-2060 relative to
359 1980-2009. In a conservative climate model (i.e. A-45), peak discharge may only be on average
360 about 5 days earlier by midcentury.

361 To get an understanding of the shift in seasonality and variance between climate scenarios,
362 we can look at the multi-decadal averaged hydrographs between two endmember climate models
363 predicting the least and most amount of change from historical averages (Figure 10). With the
364 coolest climate scenario (A-45), there is little discernible deviation from the historical average hy-
365 drograph. However, if we look at the warmest climate scenario (C-85), we see obvious differences
366 in the average hydrograph, where by 2050-2070 the average peak of the hydrograph is over a
367 month and a half earlier. Overall, this warmest scenario shows a shift in seasonality through time,
368 where we see flows occurring earlier in the season with an additional increase in early-season,
369 mid-winter discharge events.

370 **3.2.3 Center of Timing**

371 The historical average (1980-2009) center of timing (CT) of streamflow for the UBRB is April 22.
372 In our simulations, we see this date shift earlier in most of our climate scenarios (Figure 11). Three
373 scenarios (C-45, B-45, and A-85) behave similarly and begin deviating from the historical range
374 of variability between 2040 and 2050, showing a CT date that is 13-17 days earlier on average

375 between 2070 and 2099. Both C-85 and B-85 begin to deviate from historical averages around 2030
376 and exhibit an average a CT date 27-30 days earlier than the historical average during the 2070-
377 2099 period. A-45 remains relatively similar to historical ranges through the century, although its
378 CT date shifts a few days earlier, resulting in fewer occurrences of exceeding the historical 75th
379 percentile of CT date.

380 **3.3 Snowpack**

381 **3.3.1 April 1 SWE**

382 Our results (Figure 12) show a substantial decrease in April 1 SWE in five of the climate sce-
383 narios, with lower elevations essentially experiencing no April 1 SWE by midcentury. Higher ele-
384 vations remain less affected across all RCP 4.5 scenarios but begin substantially decreasing around
385 2050 in B-85 and C-85 where they experience virtually no April 1 SWE from 2080-2100. Under the
386 A-45 scenario, April 1 SWE experiences variability, but has no discernible downward trend.

387 **3.3.2 Dates and amounts of maximum SWE**

388 The previous section suggests that April 1 SWE will, at some point in the future, cease to be a
389 good indicator of maximum SWE. In terms of evaluating potential climate change impacts on SWE
390 in the context of water supply, therefore, it is necessary to examine additional metrics. Specifically,
391 we see the date of maximum SWE happening earlier across most scenarios (Figure 13). Both C-85
392 and B-85 show maximum SWE occurring more than two months earlier on average by the end
393 of the century. Three scenarios, A-85, C-45, and B-45 behave similarly with maximum SWE date
394 happening between 38 and 42 days earlier than historically observed averages. A-45 produces
395 little change in timing by the end of the century (7 days earlier on average).

396 The magnitudes of maximum SWE may change as well (Figure 14). Within mid-elevation
397 zones (2000-2500 m), we see a drastic decrease in the occurrence of annual amounts above the
398 historical 75th percentile in five of our climate scenarios. Furthermore, from 2050 onward, we see

399 that 80% (C-85) and 84% (B-85) of the time the maximum SWE is falling below the historical 25th
400 percentile. As with many of the metrics previously mentioned, A-45 shows very little change from
401 historical trends.

402 **3.4 Water Management**

403 **3.4.1 Day of Allocation**

404 The developed model reasonably reproduces the DOA in the historical period ($R^2=0.90$), al-
405 though it over-predicted the date on average 4.8 days later (Figure 6). Thus, the defined metric for
406 the DOA provides a reasonably robust vehicle to analyze how the DOA may shift under different
407 climate scenarios. Our results show the DOA occurring much earlier under four of our scenarios
408 (Figure 15), ranging from 11 to 33 days earlier on average by the end of the century. Scenarios A-45
409 and B-45 resulted in little to no change in the trend of DOA. While the DOA remains variable on
410 an interannual basis, we do not see significant changes in variability of DOA through time (Table
411 7).

412 **4 Discussion**

413 **4.1 Trends in Future Hydrologic Regimes**

414 We calibrated our model using metrics that included historic snowpack levels, daily stream-
415 flow, logarithmic transformation of streamflow, and streamflow volume. Choosing multiple met-
416 rics to select the best parameter set provides some additional confidence that the model is sim-
417 ulating key attributes of historical hydrologic regimes and, therefore, strengthens confidence in
418 the robustness of our interpretations of future climate change impacts on hydrologic regimes pre-
419 dicted by the model.

420 We have shown that a variety of hydrologic regime characteristics within the UBRB could ex-

hibit significant changes, depending on which climate model and RCP scenario is used. However, certain trends are consistent across several considered climate scenarios and are consistent with other projections (Adam *et al.*, 2009; Inouye, 2014; Gergel *et al.*, 2017). Our results suggest an increase in annual water discharge, but with significantly altered timing, with flows arriving much earlier than historically. Our modeled results also show a decrease in the total amount of snowpack, an earlier melting date, and earlier dates of peak snowpack. In order to reconcile how annual discharge can increase while the snowpack is consistently smaller in volume and more ephemeral in time, we examined the seasonality of the precipitation input to the model. This allows us to better understand whether observed changes in discharge volume are primarily related to changes in the seasonality of input precipitation, changes in the seasonal dynamics of snowpacks, or some combination of both. Typically, however, the precipitation exhibits increases across all seasons rather than large shifts between seasons in precipitation. Accordingly, this may indicate that the basin could begin transitioning from being snowmelt-dominated to a regime that is mixed rain- and snow-dominated watershed moving forward.

4.2 Management Implications

Our modeled scenarios support previous studies (Pederson *et al.*, 2011; Klos *et al.*, 2014) that April 1 SWE is not likely to remain a reliable metric for estimating maximum SWE (and therefore snow water storage) in the future for water resource prediction and management. This work suggests declines in the amount of SWE on April 1 and a maximum SWE date over a month earlier than historically observed in five of the six considered scenarios. Rather than choosing a static date to estimate peak SWE across a vast region, managers may need to more closely monitor the relationship between hydrologic regimes and the timing of peak SWE in their regions, potentially necessitating increased investment in monitoring and modeling of snow conditions.

There is little evidence to conclude that we will experience future water shortages from the UBRB in an absolute sense, as most models suggest at least a small increase in annual discharge. However, we will likely experience hydrologic shifts that are outside of our current range of vari-

447 ability. All climate scenarios show peak discharge occurring earlier in the year. This is problematic
448 for reservoir managers who primarily manage dams to provide storage for flood mitigation. Man-
449 agers might have to release more 'usable' water from reservoirs in preparation for these events,
450 which potentially could equate to shortages later in the irrigation season. Such outcomes could be
451 viewed as an "operational deficit" that arises because of a mismatch between the release of water
452 from storage for flood mitigation and the timing of water allocation as codified in water rights
453 laws.

454 At the same time, in this region agricultural land is increasingly transitioning to urban areas
455 (Dahal *et al.*, 2017), which could indicate that future water demand may be substantially differ-
456 ent from the past. With warmer climates, farmers might plant earlier in the season, which would
457 change the timing of water demand. Recent modeling efforts have shown that current water rights
458 are not always able to support irrigation demand (Han *et al.*, 2017). Agricultural water use effi-
459 ciency, however, is likely to increase with technological advances like genetically modified crops,
460 which could change spatiotemporal patterns of water demand. A more comprehensive examina-
461 tion of how, when, and where water is being used downstream and how that may change in the
462 future will help managers understand to what extent regional water infrastructure is vulnerable
463 and the potential policies that might help to mitigate effects.

464 Our results show that under most climate scenarios, the Day of Allocation occurs much earlier
465 than it has historically, with two models showing the date moving by over a month earlier. If this
466 projection becomes reality, then there is an earnest need for exploring potential conflicts between
467 water users in the future as curtailments may come increasingly early and impact more water
468 rights holders than in previous decades. It may be necessary, for instance, to incentivize farmers
469 to transition to more efficient irrigation practices (e.g. switching from flood to drip irrigation)
470 and to diversify with crops that require less water, or expand other solutions like water bank-
471 ing and water markets. If junior water rights holders are curtailed over a month earlier without
472 any mitigation practices set in place, it may result in substantial repercussions to Idaho's agricul-
473 tural sector. These effects are compounded if other mountain water supply basins exhibit similar

474 changes to hydrologic regimes.

475 **4.3 Study Limitations**

476 It is worth noting that this study did not simulate reservoir operations. There are three dams
477 present in the study area that are located close to the outlet of the basin. For purposes of simplicity,
478 the present work focuses on evaluating the ramifications of climate change on natural flows in the
479 UBRB and capturing reservoir operations is outside the scope of this study. A significant challenge
480 in future work will arise from the need to develop plausible scenarios by which water managers
481 from federal agencies, irrigation districts, environmental groups, and utility companies can create
482 strategies to adapt to potential changes in hydrologic regimes similar to those simulated here.
483 Given the complexities in both biophysical and social responses to climate change, such studies
484 will likely need to be region- and context-specific.

485 An additional source of uncertainty in this study lies in the land cover data used in the hy-
486 drologic model, which was treated as static. Specifically, the land cover dataset used represents
487 a snapshot estimated based on Landsat reflectances from 2011. Vegetation along ecotones is sen-
488 sitive to changes in climate, and there are likely to be additional large-scale vegetation and land
489 cover changes that occur after wildfire events or through land management actions. Future mod-
490 eling studies should incorporate plausible shifts in vegetation to understand the sensitivity of
491 changes in hydrologic regimes to associated changes in land cover as well as climate change. This
492 might be best accomplished using a physically-based model, rather than the conceptual model
493 used in this study, to be able to better capture complex interactions between climate, hydrology,
494 vegetation dynamics, and changing land cover.

495 **4.4 Conclusions**

496 In this study, we used an integrated modeling framework, Envision, to simulate future hy-
497 drology in a mountainous watershed that supports an urban and agriculturally intensive region
498 below it. We calibrated the hydrologic model to metrics of both streamflow and snowpack, and it

499 performed well under historical conditions. We ran the model to the year 2100 under six climate
500 scenarios (three GCMs and two RCP scenarios) to analyze future possible hydrologic regimes.

501 Our results suggest that overall annual streamflow will increase, and five of six scenarios sug-
502 gest hydrologic regimes that will deliver runoff substantially earlier than historically observed.
503 This could lead to operational water shortages later in the season as water managers balance re-
504 lease of water from storage in reservoirs to mitigate flooding hazards with retention of water for
505 supplying irrigation in the warm, dry summers. Without changes in existing policies, these hydro-
506 logic regimes could have repercussions to late-season irrigation demand, hydropower operations,
507 recreational flows, and municipal water supply.

508 Mountainous, snowmelt-dominated watersheds have already begun responding to climate
509 change, which will almost certainly continue in the future. The degree to which the runoff re-
510 sponse of these watersheds changes in association with climate change is uncertain, and will
511 depend heavily on the nature of the change in the climatic forcing variables. Increasingly so-
512 phisticated comparisons with climate model predictions and observations, as well as regionally
513 focused and contextual modeling of coupled hydrologic and social systems, will improve our abil-
514 ity to constrain how hydrologic regimes will change in the future. This may increase the efficacy
515 of efforts to respond to changes and potential conflicts between potentially competing demands
516 for water.

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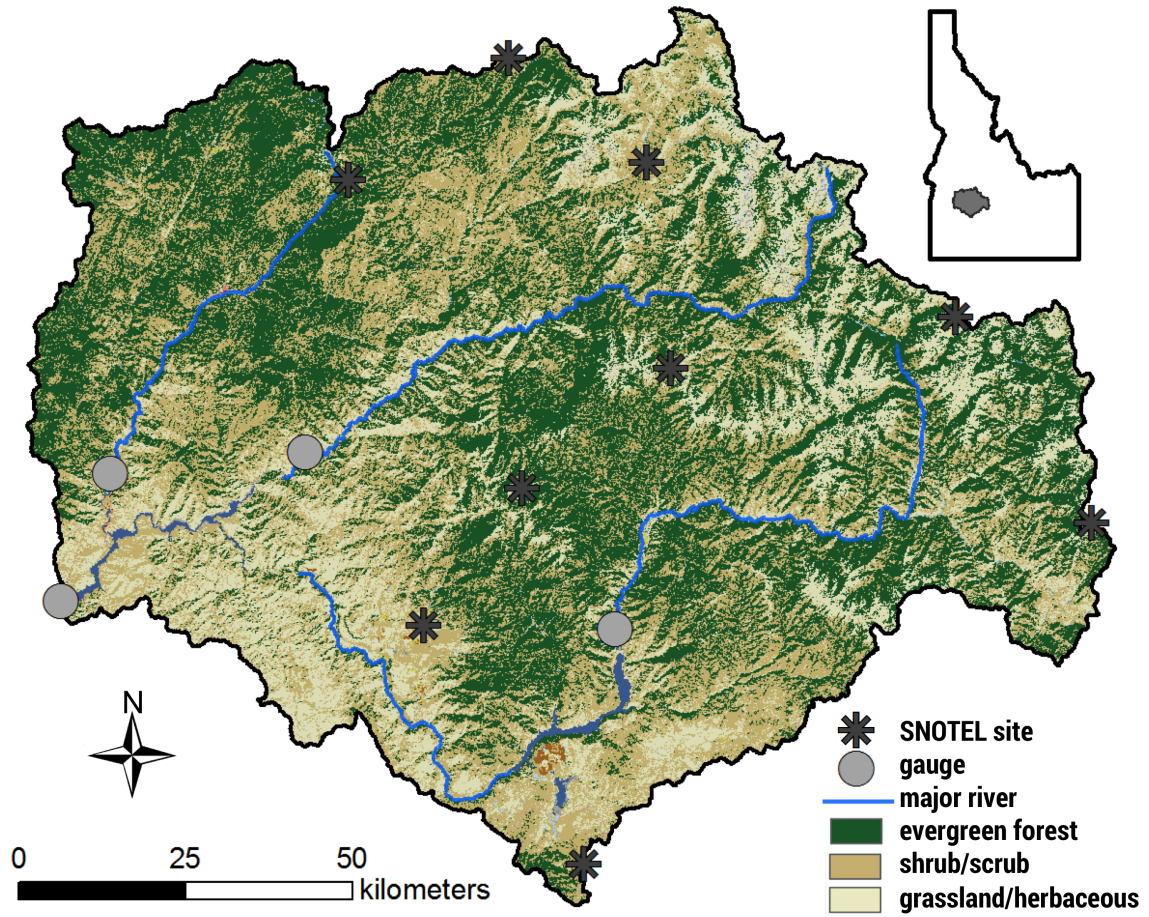


Figure 1: Overview of the study area with major land cover types and locations of SNOTEL stations and gauge locations (see Table 5 for names of gauges).

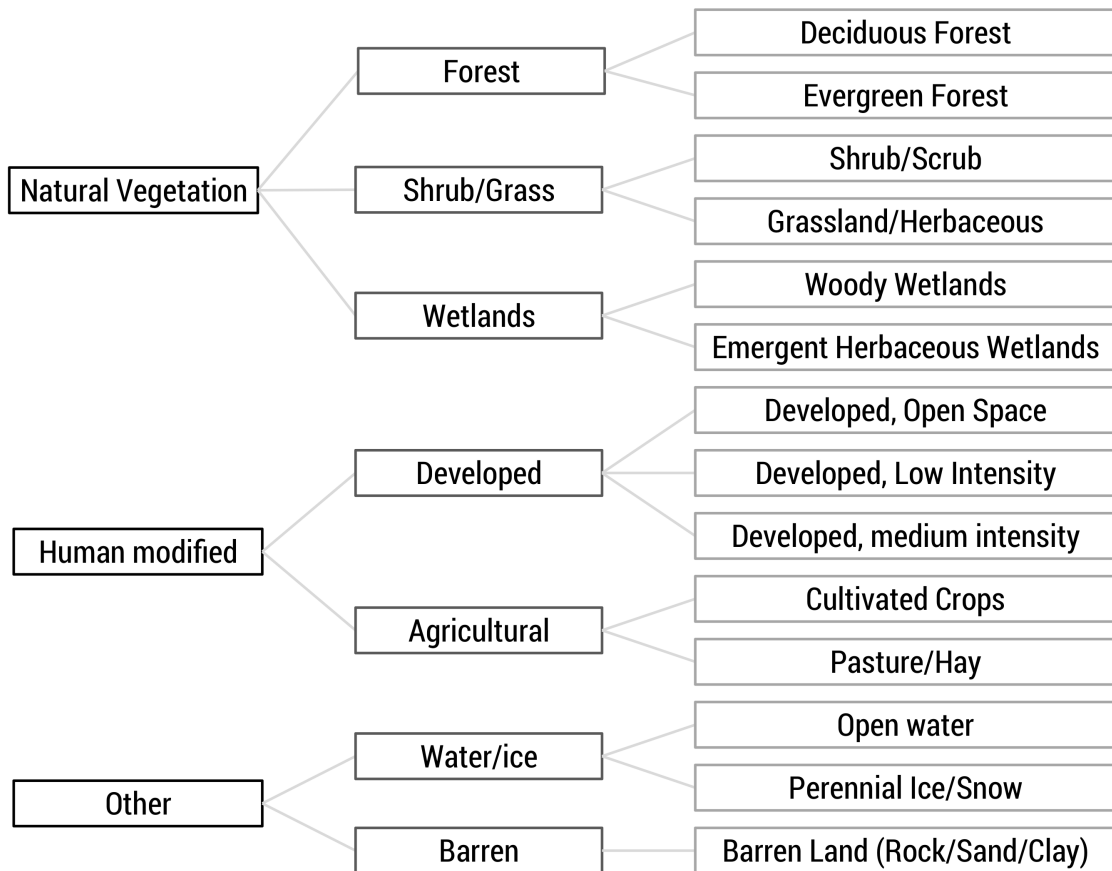


Figure 2: Land use/land cover tree developed for Envision. The tree allows for modeling algorithms to be applied at different hierarchy levels, from more general to more specific land types. The finest categories on the right correspond to the NLCD land classification system.

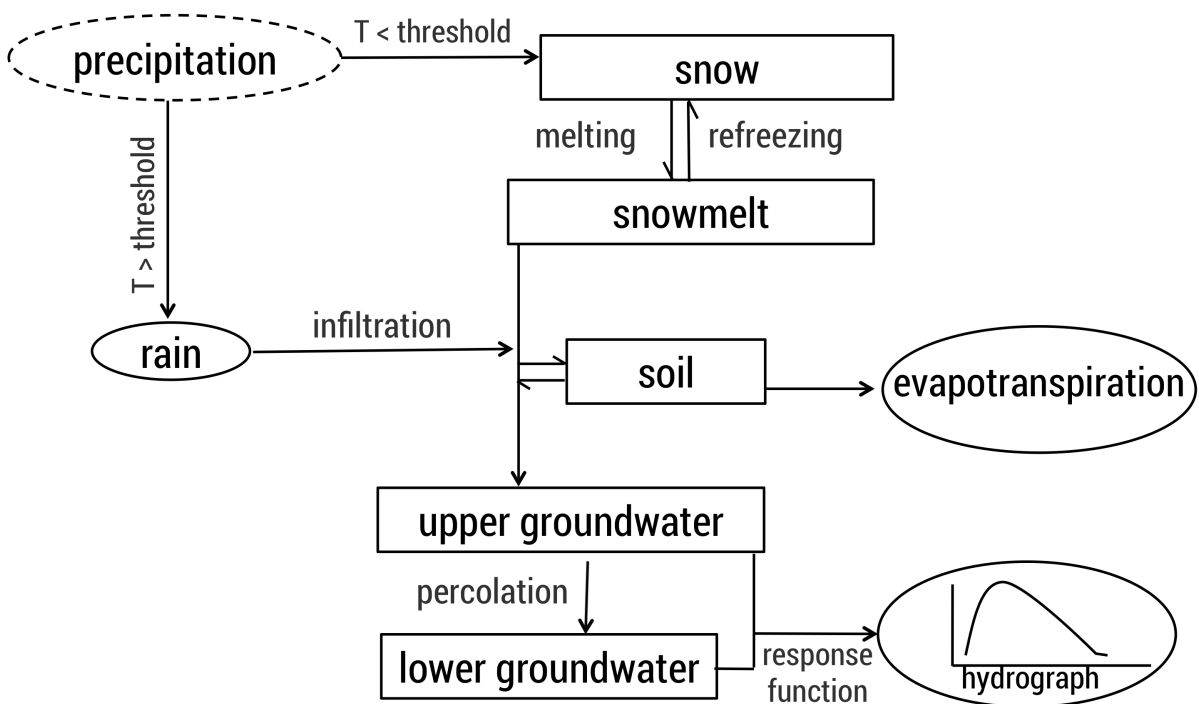


Figure 3: Flowchart of the different hydrologic processes and reservoirs within the Flow model in Envision, (modified from Han *et al.*, 2017)

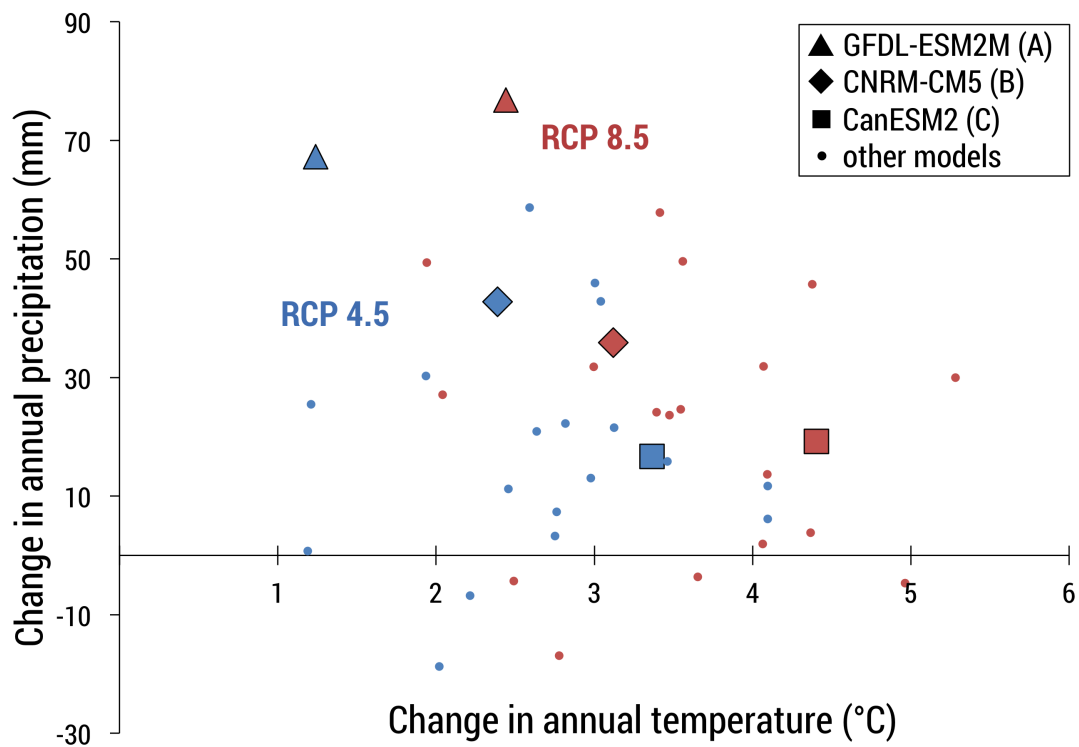


Figure 4: Change in climate variables from 1979-2000 to 2040-2069 for MACA downscaled GCMs (Abatzoglou and Brown, 2011). Blue and red points represent RCP 4.5 and 8.5 scenarios, respectively. The larger icons represent the GCMs selected for this study.

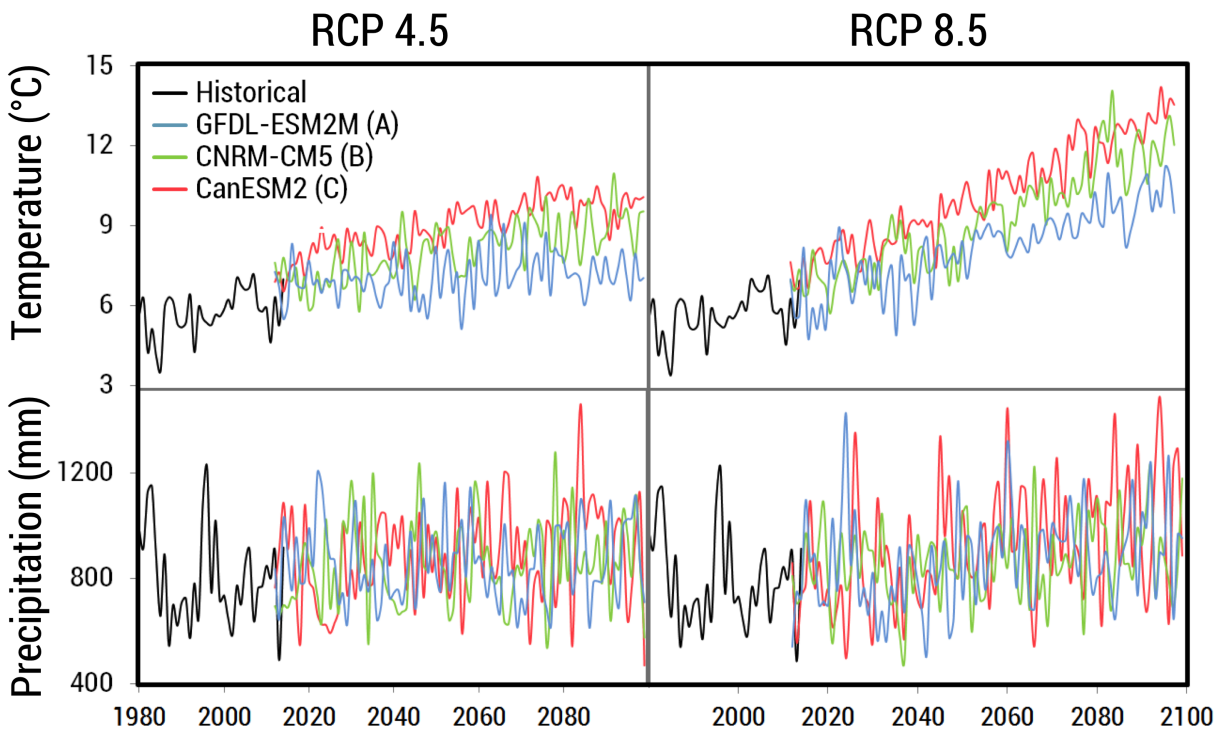


Figure 5: Temporal projections for annual mean temperature and precipitation for the six climate scenarios used in this study. Temperature increases in all scenarios, but precipitation is more variable.

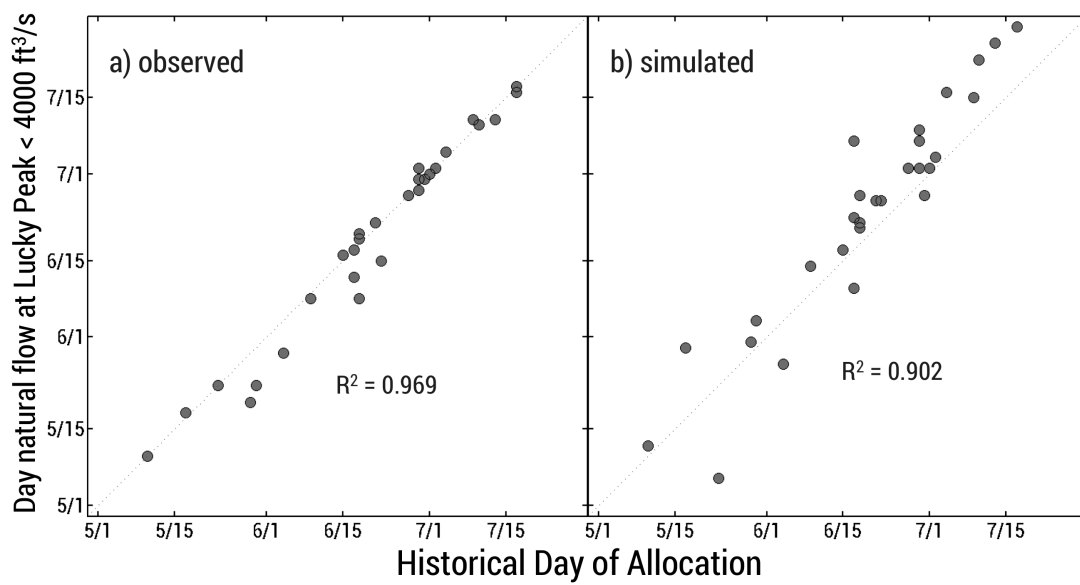


Figure 6: (a) Relationship between the day natural flow at Lucky Peak reaches below 4000 ft³/s and the date the Day of Allocation is declared, modified from (Garst, 2017). (b) Our modeled historical Day of Allocation using the same method as (a). Dashed line is 1:1 in both plots.

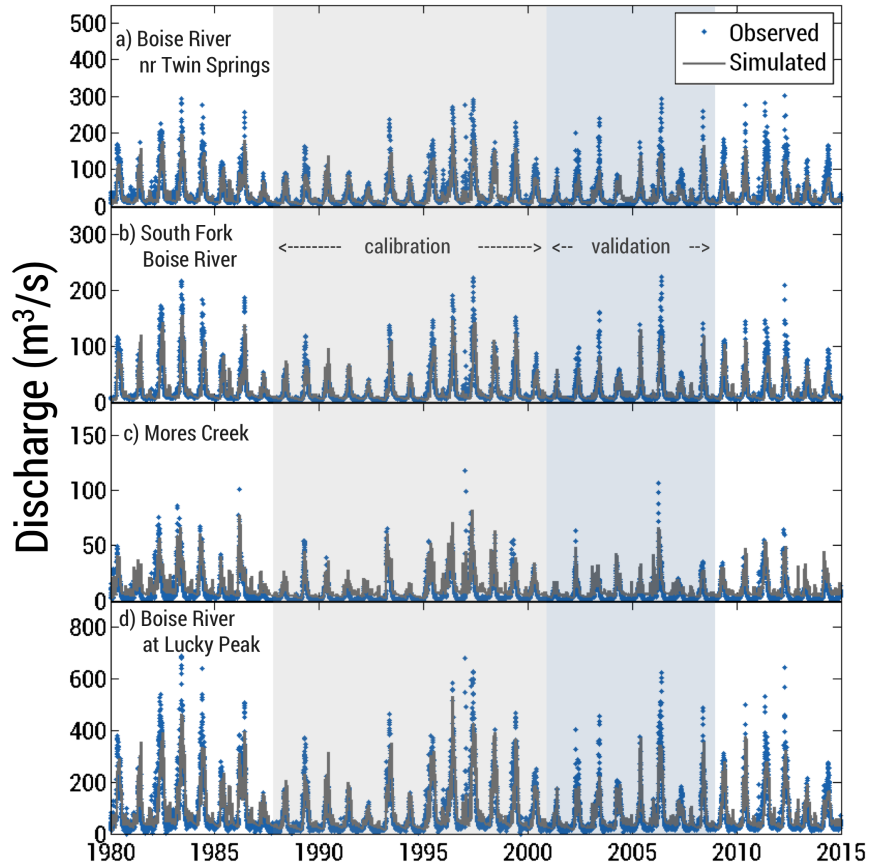


Figure 7: Observed and simulated streamflows during the historical period from 1980 to 2014. See Figure 1 for locations of sites. The model does a good job at simulating historical flows, but under estimates magnitude of peak flows and over estimates baseflow at Mores Creek.

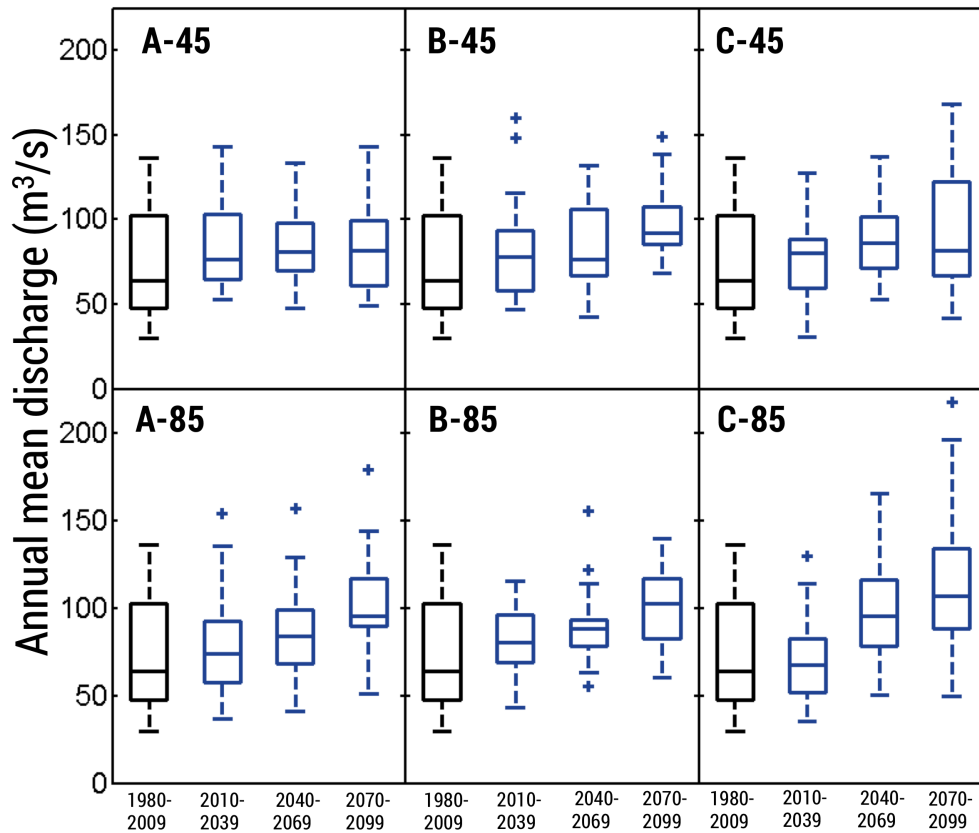


Figure 8: Average annual discharge of the UBRB. Values for 1980-2009 are observed. In most scenarios, we see an increase in overall discharge throughout the century. Boxes represent upper and lower quartiles and lines inside are the median.

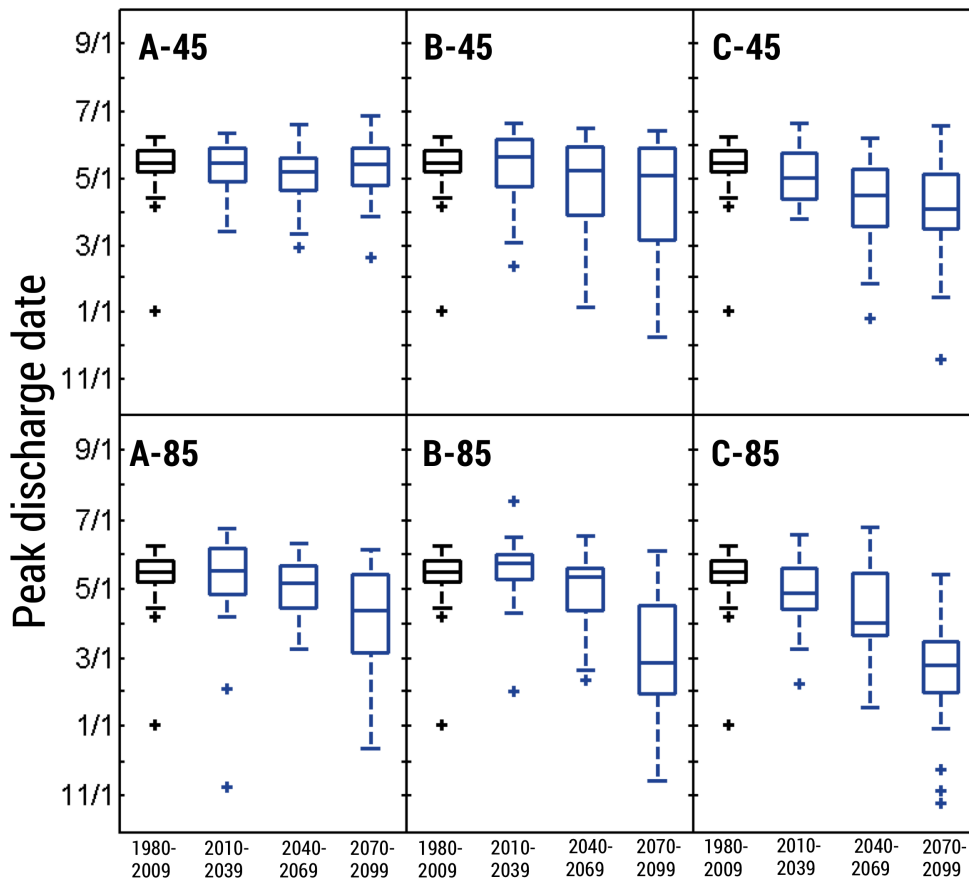


Figure 9: Date when peak discharge occurs for the Boise River at Lucky Peak. Values for 1980-2009 are observed. Overall, we see peak discharge date moving substantially earlier in five scenarios.

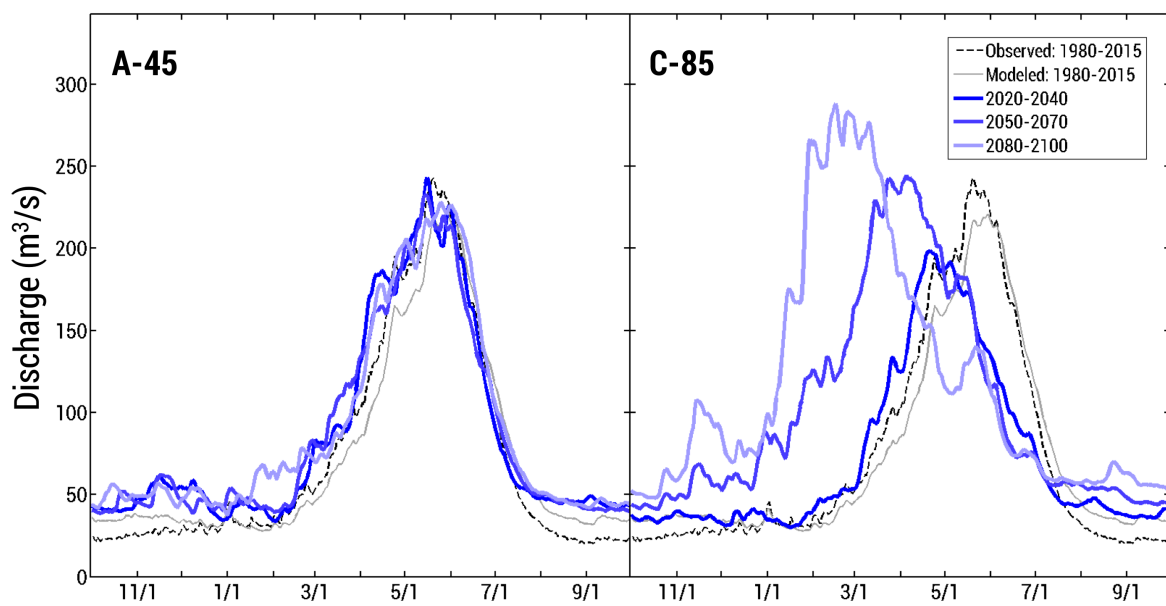


Figure 10: Hydrographs averaged over 2-decadal timespans for scenarios predicting the least amount of change (A-45) and the greatest amount of change (C-85) from historical.



Figure 11: Center of timing of streamflow for historic and future simulations. Dashed lines show the upper and lower quartile ranges from 1980-2009.

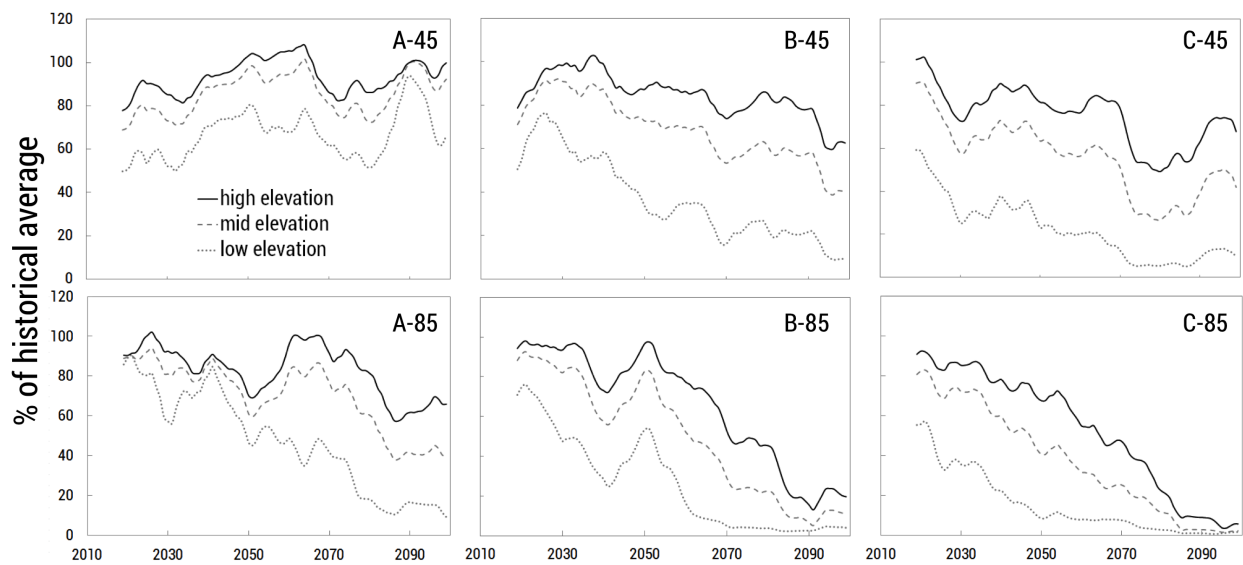


Figure 12: 10-year moving average percentage of April 1 SWE from historical simulated averages (1980-2009) for low, medium, and high elevation zones, corresponding to 1500-2000, 2000-2500, and 2500+ m, respectively.

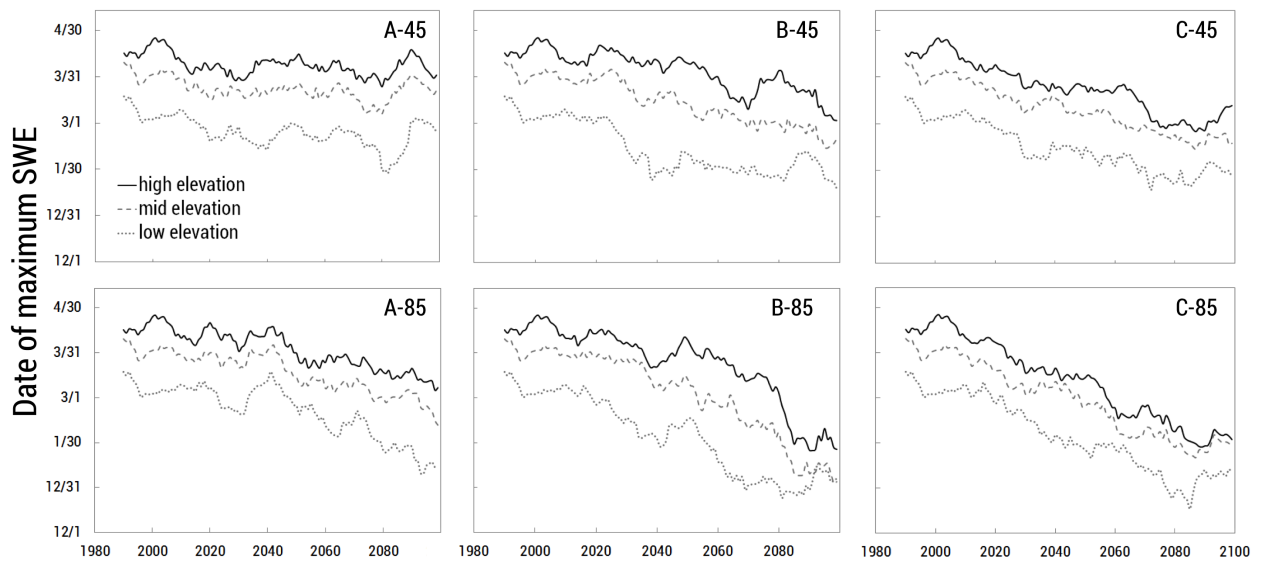


Figure 13: 10-year moving average of dates of maximum SWE for three elevation zones. Values for 1980-2009 are simulated with MACA METDATA.

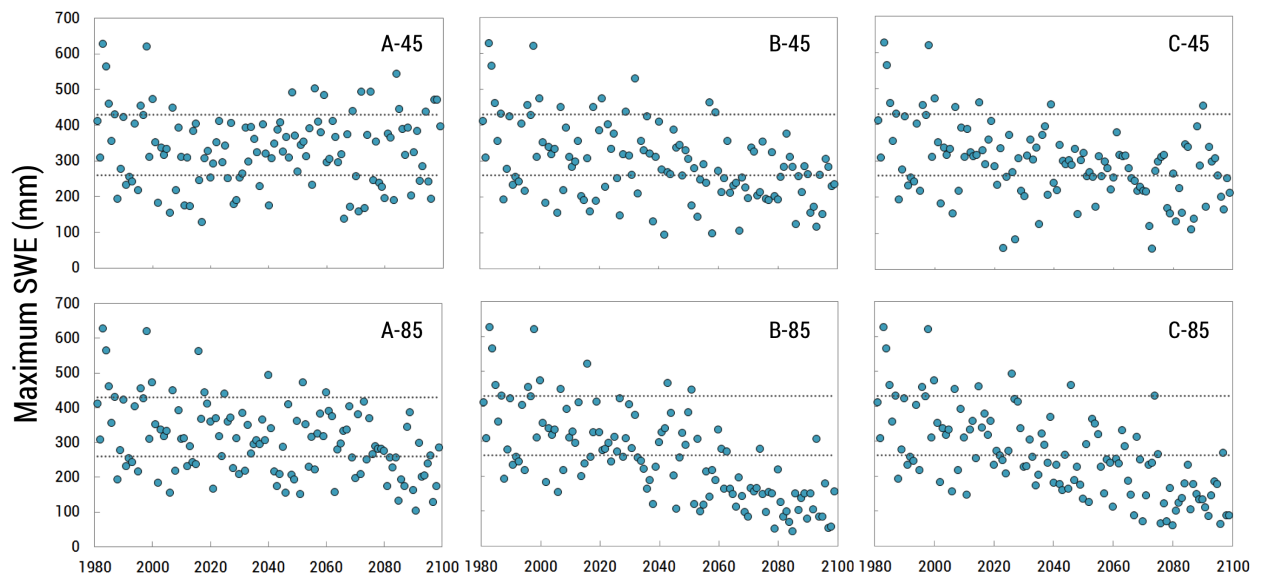


Figure 14: Maximum SWE amount (mm) for mid-elevations (2000-2500 m). Values for 1980-2009 are simulated with MACA METDATA. Dashed lines show upper and lower quartile ranges for 1980-2009.

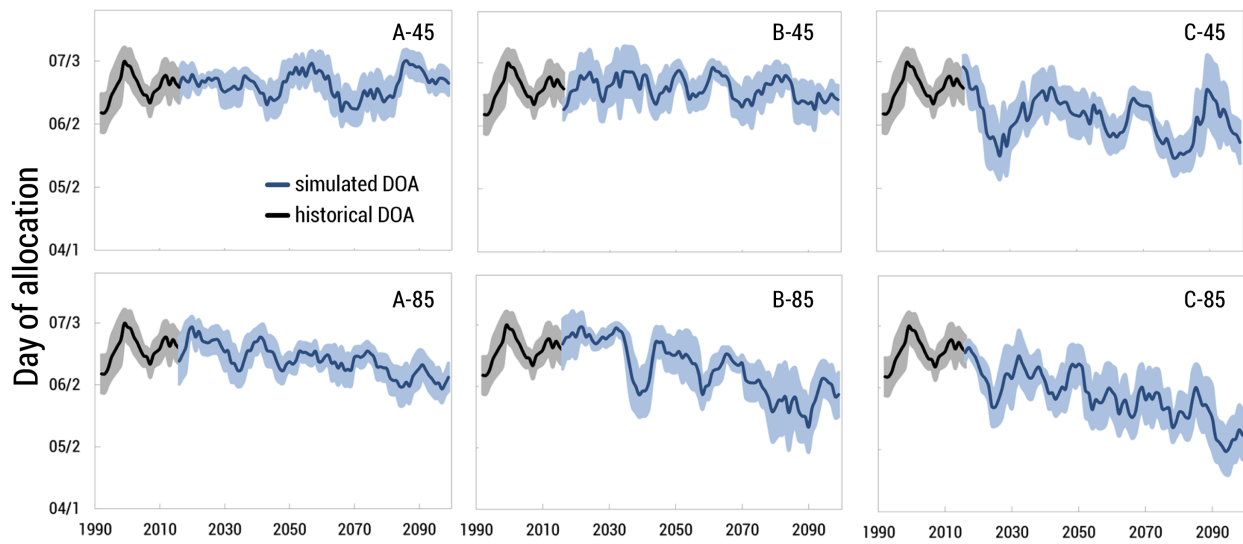


Figure 15: Simulated future (2010-2099) and historical (1986-2014) Day of Allocation with a 7-year moving average. Shaded area is $\pm 0.5\sigma$ of 7-year moving average values.

Table 1: Data sources used for spatial coverage in Envision

| Input Data (<i>resolution</i>) | Data Sources | Used In |
|----------------------------------|------------------------------------|----------|
| Surface Management Agency | Bureau of Land Management | IDU |
| Land Cover (30 m) | National Landcover Database (2011) | IDU, ET |
| Streams & Catchments (HUC12) | NHD Plus V2 | IDU, HBV |
| Elevation (30 m) | National Elevation Dataset | HRU |

Table 2: Land cover type in Envision and the associated crop used to calculate evapotranspiration

| Land Cover | Crop substituted for land cover | Source |
|--------------|--|--------------------------|
| Forest | 3 rd year poplar \times 3 | Agrimet, Inouye (2014) |
| Shrubland | Sagebrush | Allen and Robison (2007) |
| Grassland | Bunch grass | Allen and Robison (2007) |
| Wetlands | Poplar \times 3 | Agrimet, Inouye (2014) |
| Developed | Lawn \times 0.21 | Agrimet, Inouye (2014) |
| Agricultural | Alfalfa (mean) | Agrimet |

Table 3: Naming convention for the six climate scenarios used in this study

| | GFDL-ESM2M (warm) | CNRM-CM5 (warmer) | CanESM2 (warmest) |
|--------|----------------------|----------------------|----------------------|
| RCP4.5 | A-45 | B-45 | C-45 |
| RCP8.5 | A-85 | B-85 | C-85 |

Table 4: Parameters for Flow and the ranges/values considered for calibration

| Routine | Parameter | Description | Units | Range | Value |
|----------------------------------|-----------|--|---|--------------|-------|
| Snow Routine | TT | Threshold temperature | °C | -0.5 – 2.0 | 1.335 |
| | CFMAX | Degree-day factor | mm°C ⁻¹ day ⁻¹ | 1.0 – 6.0 | 1.489 |
| | SFCF | Snowfall correction factor | - | 0.7 – 1.2 | 0.568 |
| | CFR | Refreeze coefficient | - | - | 0.05 |
| | CWH | Water holding capacity of snowpack | - | - | 0.1 |
| Soil and Evaporation Routine | FC* | Max depth of water in soil water reservoir | mm | - | 399.7 |
| | LP* | Soil moisture value where actual ET=PET | mm | - | 247.2 |
| | WP* | Wilting point in soil for ET to occur | mm | - | 156.2 |
| | BETA | Shaping coefficient | - | 1.0 – 6.0 | 2.015 |
| | PERC | Percolation coefficient | day ⁻¹ | 0.1 – 2.0 | 1.272 |
| Groundwater and Response Routine | UZL | Threshold for K0 to outflow | mm | 1.0 – 400.0 | 365.4 |
| | K0 | Recession coefficient | day ⁻¹ | 0.1 – 1.0 | 0.339 |
| | K1 | Recession coefficient | day ⁻¹ | 0.01 – 0.5 | 0.079 |
| | K2 | Recession coefficient | day ⁻¹ | 0.001 – 0.15 | 0.004 |

* values obtained from ORNL DAAC SDAT

Table 5: Data sites used for calibration and validation. See Figure 1 for locations of gauges.

| Type | Name | Drainage Area (km ²) | Record Length | Site ID |
|--------|-----------------------------------|-------------------------------------|----------------|----------|
| Gauge | a) Boise River nr Twin Springs | 2154.9 | 1911 – present | 13185000 |
| | b) SF Boise River nr Featherville | 1660.2 | 1945 – present | 13186000 |
| | c) Mores Creek abv Robie Creek | 1028.2 | 1950 – present | 13200000 |
| | d) Boise River at Lucky Peak* | 6571 | 1895 – present | LUC |
| Type | Name | Elevation (m) | Record Length | Site ID |
| SNOTEL | Atlanta Summit | 2310 | 1981 – present | 306 |
| | Camas Creek | 1740 | 1992 – present | 382 |
| | Dollarhide Summit | 2566 | 1981 – present | 450 |
| | Graham Guard Station | 1734 | 1981 – present | 496 |
| | Jackson Peak | 2155 | 1981 – present | 550 |
| | Mores Creek | 1859 | 1981 – present | 637 |
| | Prairie | 1463 | 1987 – present | 704 |
| | Trinity | 2368 | 1981 – present | 830 |
| | Vienna Mine | 2731 | 1979 – present | 845 |

*not an actual gauge, but a calculated daily average of runoff at this location if dams were not present. Obtained from the US Bureau of Reclamation.

Table 6: Calibration and validation results for the chosen parameter set for this study.

| Calibration | | | | | Validation | | | | |
|-------------|--------------|--------|---------|--------|------------|--------------|--------|---------|--------|
| NSE_G | $\log NSE_G$ | VE_G | NSE_S | $Obj.$ | NSE_G | $\log NSE_G$ | VE_G | NSE_S | $Obj.$ |
| 0.71 | 0.61 | -0.03 | 0.59 | 0.63 | 0.70 | 0.66 | -0.06 | 0.52 | 0.62 |

Table 7: Simulated mean Day of Allocation (DOA) and standard deviation (*italicized, in parentheses*) over three future time intervals. Historical (1986-2014) average DOA is 6/19.

| Time Period | A-45 | B-45 | C-45 | A-85 | B-85 | C-85 |
|-------------|-------------|-------------|-------------|-------------|-------------|-------------|
| 2010-2039 | 6/22 (12.0) | 6/21 (20.0) | 6/10 (24.3) | 6/19 (17.1) | 6/20 (20.6) | 6/10 (19.0) |
| 2040-2069 | 6/20 (17.3) | 6/20 (15.3) | 6/7 (16.8) | 6/15 (13.1) | 6/15 (17.2) | 5/30 (23.5) |
| 2070-2099 | 6/23 (15.1) | 6/18 (16.1) | 5/29 (24.0) | 6/8 (14.9) | 5/27 (25.6) | 5/17 (23.5) |