

1 **Title:** Inferring drivers of changing land surface phenology from Landsat time series
2 **Authors:** Cornelius Senf and Tobias Krueger
3 **Affiliation:** Integrative Research Institute on Transformation of Human-Environment
4 Systems (IRI THESys), Humboldt-Universität zu Berlin, Unter den Linden 6, 10099 Berlin,
5 Germany
6 **Correspondence to:** Cornelius Senf; Tel: +49 151 50652760; E-Mail:
7 corneliussenf@googlemail.com
8 **Keywords:** Phenology; Remote sensing; Climate change; Bayesian hierarchical modelling;
9 Landsat
10

11 **Abstract:** Phenology is an important ecosystem property, and monitoring and modeling of
12 phenology is particularly important for understanding climate change impacts on vegetated
13 ecosystems. However, in-situ measurements are frequently confined to a few specific
14 observation sites and species, and are thus limited for fully understanding the drivers of
15 changing phenology at broader scales. Moderate resolution remote sensing time series from the
16 Landsat archive can help overcome this limitation by delivering a consistent estimate of land
17 surface phenology over the past 30 years. Yet, methods for inferring the drivers of variation in
18 land surface phenology from these data remain scarce. We here present a new model for
19 inferring drivers of changing land surface phenology from Landsat time series. We demonstrate
20 our model using a case study comprising broadleaved and coniferous forests and estimating the
21 effects of pre-season temperature and winter-chilling on inter-annual variation in the start of
22 season. We identified significant effects of pre-season temperature on inter-annual variation in
23 start of season, with a $-3.74 \text{ d } ^\circ\text{C}^{-1}$ earlier start of season for broadleaved and a $-2.68 \text{ d } ^\circ\text{C}^{-1}$
24 earlier start of season for coniferous forests, respectively. This relationship, however, was
25 modulated by the number of chilling days, with a decreasing effect of pre-season temperature
26 with decreasing number of chilling days. The inter-annual variation in start of season predicted
27 from our model – i.e., calibrated solely from Landsat satellite time series – showed good
28 agreement with in-situ observations of bud-break (Pearson's $r = 0.79/\text{RMSE} = 4.88 \text{ d}$ for
29 broadleaved forests and $r = 0.62/\text{RMSE} = 6.57 \text{ d}$ for coniferous forests). Our model thus allows
30 for inferring drivers of changing land surface phenology directly from Landsat time series,
31 opening up phenological research in areas where in-situ measurements are unavailable, and at
32 spatial and temporal scales difficult to tackle with field and coarse-scale remote sensing data.

33

34 **Introduction**

35 Phenology is a key property of ecosystems (Forrest & Miller-Rushing, 2010). Due to its
36 sensitivity to climate, phenology is also a valuable indicator of climate change (Cleland,
37 Chuine, Menzel, Mooney, & Schwartz, 2007). Climate change effects on vegetation
38 phenology have been documented in numerous experimental and observational studies,
39 finding shifts in the start of the growing season in many plant species as a result of climate
40 change (Cleland et al., 2007; Menzel et al., 2006; Parmesan & Yohe, 2003). Those changing
41 phenological patterns have profound impacts on ecosystem functions, such as carbon uptake
42 (Keenan et al., 2014). In order to better understand the impacts of changing phenology on
43 ecosystem function, it is fundamental to monitor, model and ultimately predict vegetation
44 phenological dynamics at varying spatial and temporal scales (Pau et al., 2011; Tang et al.,
45 2016).

46 A key data source for monitoring and modeling phenological dynamics are in-situ
47 phenological measurements. Those measurements are generally taken at the level of
48 individuals (e.g., single trees or plants), and the resulting information is used to calibrate
49 species-specific phenological models that predict phenological phases (e.g., leaf unfolding)
50 based on a set of aggregated meteorological variables (Basler, 2016; Chuine, 2000; Hufkens,
51 Basler, Milliman, Melaas, & Richardson, 2018). However, such species-specific models are
52 often difficult to regionalise to the ecosystem scale (Y. Fu, Zhang, Dong, & Yuan, 2014;
53 Richardson et al., 2012; Tang et al., 2016), particularly if the exact species or community
54 composition are unknown (Jeremy I. Fisher, Richardson, & Mustard, 2007). Further,
55 calibrating species-specific models across large geographic extents is challenging, as in-situ
56 phenological observations are lacking in many regions worldwide, or are difficult to compare
57 due to varying measurement protocols and/or target species (Fitchett, Grab, & Thompson,
58 2015). An improved understand of phenological dynamics at the ecosystem scale is, however,
59 required for further developing the representation of phenology in global vegetation models

60 (Richardson et al., 2012; Yang, Mustard, Tang, & Xu, 2012). Remote sensing, which
61 measures so-called land surface phenology and thus delivers an integrated view on phenology
62 independent of individual species (Morisette et al., 2009), might help overcoming this
63 knowledge gap.

64 Classical approaches tracking phenology from remote sensing data made use of the
65 synoptic view of the Advanced Very High Resolution Radiometer (AVHRR) sensor family
66 (M. A. White et al., 2009). AVHRR, however, has a spatial resolution that is too coarse (1
67 km) to infer phenological dynamics in spatially heterogeneous landscapes. The MODerate
68 Imaging Spectroradiometer (MODIS), which has a higher spatial resolution of 250 m, might
69 improve the estimation of spatial drivers (Friedl et al., 2014), yet its short time series (starting
70 in 2001) makes it difficult to infer temporal trends in phenological dynamics. Dense Landsat
71 time series were recently suggested to overcome the limitations of MODIS and AVHRR in
72 monitoring vegetation phenology (J. I. Fisher & Mustard, 2007; J. I. Fisher, Mustard, &
73 Vadeboncoeur, 2006; Melaas, Friedl, & Zhu, 2013; Melaas, Sulla-Menashe, & Friedl, 2018;
74 Senf, Pflugmacher, Heurich, & Krueger, 2017), since they have a medium spatial resolution
75 of 30 m and cover a time-span of more than 30 years (from 1984 onwards). However,
76 methods for inferring patterns and drivers of land surface phenology from those medium
77 resolution sensors remain scarce.

78 We here present a new model for inferring drivers of inter-annual variation in land
79 surface phenology from Landsat time series. Our model, which is based on a Bayesian
80 hierarchical setup (Senf et al., 2017), integrates Landsat time series and meteorological
81 observations, thus allowing for the direct assessment of drivers (e.g., pre-season temperature)
82 on inter-annual variability in phenological parameters (e.g., the start of season). We
83 demonstrate our model using a case study from southern Germany comprising broadleaved
84 and coniferous forests. Our specific objectives were to (1) estimate the effects of pre-season
85 temperature and winter chilling on inter-annual variability in start of season in broadleaved

86 and coniferous forests and to (2) compare the model, which is calibrated solely from Landsat
 87 time series, to in-situ phenological observations of leaf unfolding.

88 **Model description**

89 *Landsat phenological model*

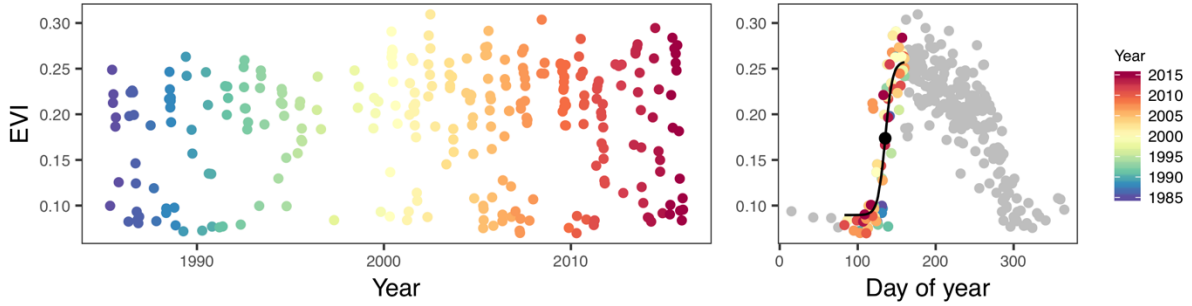
90 The underlying Landsat phenological model utilized in this study follows previous work
 91 described in Melaas et al. (2013) and Senf et al. (2017) and can be summarised as follows:
 92 Consider a time series of Landsat images for which a vegetation index has been calculated.
 93 For a pixel i , all available vegetation index values can be pooled into a vector \mathbf{y}_{ti} sorted by
 94 observation day of year t (see Fig. 1). The vector \mathbf{y}_{ti} thus describes the phenological variation
 95 of pixel i , independent of the year of observation, which may be modeled as
 96 $\mathbf{y}_{ti} \sim N(g(t; \boldsymbol{\beta}_i), \sigma^2)$. Thus, the mean phenological dynamics are described by a functional
 97 relationship $g(t; \boldsymbol{\beta}_i)$, with a set of parameters for each pixel $\boldsymbol{\beta}_i$. We here chose a logistic
 98 function as the base phenological model (Elmore, Guinn, Minsley, & Richardson, 2012; K.
 99 White, Pontius, & Schaberg, 2014):

$$g(t; \boldsymbol{\beta}_i) = \beta_{1[i]} + \frac{\beta_{2[i]}}{\left(1 + e^{-\beta_{3[i]} * (t - \beta_{4[i]})}\right)} \quad (1)$$

100 with $\boldsymbol{\beta}_i$ presenting a four-dimensional vector of model parameters describing the shape of the
 101 logistic function for each pixel i (see Figure 1 and Table 1). The parameter vector $\boldsymbol{\beta}_i$ is
 102 modelled as a multivariate normal distribution $MVN(\boldsymbol{\mu}_{[i]}, \boldsymbol{\Sigma}_\beta)$, with $\boldsymbol{\mu}_{[i]}$ being the vector of
 103 mean model parameters and $\boldsymbol{\Sigma}_\beta$ being the variance-covariance matrix of the four model
 104 parameters. Following Senf et al. (2017) we re-parameterized the multivariate normal
 105 distribution as $\boldsymbol{\beta}_i = (\text{diag}(\boldsymbol{\sigma}_\beta^2 * \boldsymbol{\tau}) * L_\beta * \mathbf{z}_\beta)^T + \boldsymbol{\mu}_{[i]}$ to numerically improve sampling
 106 (Monnahan, Thorson, & Branch, 2017). Thereby, \mathbf{z}_β represents a vector of $N(0, 1)$ random
 107 variables, $\boldsymbol{\sigma}_\beta^2$ is a vector of variances, $\boldsymbol{\tau}$ represents a vector with scaling factors for the
 108 variance (which is by default set to $\boldsymbol{\tau} = \{0.1, 0.1, 0.1, 1\}$), and L_β is the Cholesky

109 decomposition of the correlation matrix C_β with $C_\beta = L_\beta * L_\beta^T$. The vector $\mu_{[i]}$ centers the
 110 multivariate normal distribution on the approximate location of each phenological parameter
 111 in β_i and must be given *a priori*. We introduce an automatic algorithm for determining those
 112 prior choices in Section 2.3. A weakly-informative *half – cauchy*(0, 1) prior is assigned to
 113 σ_β^2 and a weakly-informative *LKJ*(2) prior is assigned to the correlation matrix C_β .

114



115

116 **Figure 1:** Example pixel time series showing the annual observations (left panel) and the
 117 pooled time series (right panel). Shown here is the Enhanced Vegetation Index (EVI). The
 118 pooled time series is truncated to spring and summer observations as described in Section 2.3,
 119 and only those observations are colored. A phenological model as described in Section 2.1
 120 was fit to the data. The black dot represents the start of season estimated from the model.

121

122 **Table 1:** Phenological parameters of the model.

Parameter	Name	Description
$\beta_{1[i]}$	Minimum	The minimum spectral value.
$\beta_{2[i]}$	Magnitude	The magnitude of spectral change during the year.
$\beta_{3[i]}$	Change rate	The change rate in the inflection point.
$\beta_{4[i]}$	Start of season	The day of year of the inflection point.

123

124

125

126 ***Hierarchical formulation to model inter-annual variability***

127 To account for temporal variation in the model parameters β_i among years j , we follow Senf
128 et al. (2017) and replace β_i by a hierarchical level: $\beta'_{ij} = \beta_i + \phi_j$. This hierarchical
129 formulation allows the model parameters to simultaneously vary in space and time. We model
130 ϕ_j as a multivariate normal distribution $MVN(\gamma_{[j]}, \Sigma_\phi)$, re-parameterized as described above:
131 $\phi_j = (\text{diag}(\sigma_\phi^2 * \tau) * L_\phi * \mathbf{z}_\phi)^T + \gamma_{[j]}$. While previous studies assumed the temporal
132 variation in model parameters to be purely random (Senf et al., 2017), that is setting $\gamma_{[j]}$ to
133 $\{0,0,0,0\}$, we here test the hypothesis that the temporal variation in the start of season $\gamma_{4[j]}$ is
134 described by a linear combination of annual predictor variables $\gamma_{4[j]} = \mathbf{M}_j \boldsymbol{\rho}$; with $\boldsymbol{\rho}$ being a
135 vector of model coefficients and \mathbf{M}_i a design matrix of predictors. The sub-model thus
136 determines the direction and strength of influence of a set of temporal drivers on the inter-
137 annual variability in the start of season. The other model parameters (minimum, maximum or
138 green-up rate; see Figure 1 and Table 1) could also be modeled dependent on temporal
139 drivers. Weakly-informative *half-Cauchy*(0, 1) priors were assigned to the variance
140 parameter σ_ϕ^2 , a weakly-informative *LKJ*(2) prior was put on the correlation matrix C_ϕ , and
141 weakly-informative *student-t*(3, 0, 1) priors were assigned to $\boldsymbol{\rho}$.

142 ***Ensemble implementation and Bayesian inference***

143 Bayesian hierarchical models are computational intensive and hence restricted to a limited
144 number of pixels. However, previous research has shown that already a small number of
145 pixels (e.g., 100 pixels) are sufficient for estimating the ecosystem-scale temporal dynamics
146 in phenology (Senf et al., 2017). Hence, we here again utilize a sample of Landsat time series
147 for estimating temporal dynamics and drivers of land surface phenology instead of creating
148 wall-to-wall maps. However, to reduce bias stemming from sampling variability, we
149 implemented an ensemble sampling strategy that draws m samples of size n , and later
150 average over the m joint posterior probability distributions of each model parameter of

151 interest (i.e., ϕ_j and ρ). That way, we reduce bias in parameter estimates stemming from
152 sampling variability, while simultaneously allowing for efficient and fast inference. The
153 associated cost of greater imprecision of parameter estimates due to the averaging of the
154 posterior distributions of the ensemble was mitigated by a sufficient number of ensemble
155 members (Figure S1).

156 For each sampled time series, there was first need for selecting only spring and
157 summer observations. We follow a method suggested by Melaas et al. (2013), which excludes
158 all observations before day of year 80 (winter observations) and then uses an iterative
159 algorithm identifying the transition from summer to autumn. In essence, the algorithm fits a
160 linear model to a running window of 21 observations and identifies the day of year with the
161 first negative slope as the summer-autumn transition date.

162 Prior choices for $\mu_{[i]}$ are estimated by iteratively fitting $g(t; \beta_i)$ using non-linear least
163 squares until 10 models were collected (models that did not converge were dropped). We then
164 averaged parameter estimates over those 10 models and used the averages for $\mu_{[i]}$. The prior
165 choices could also be based on previous studies or knowledge of the local phenological
166 system, if available.

167 Finally, full Bayesian inference was made by sampling the joint posterior distribution
168 using Markov Chain Monte Carlo (MCMC) methods implemented in the free software Stan
169 (Carpenter et al., 2017). The joint posterior distribution was sampled independently for each
170 of the m samples. Sampling was done using four chains à 4,000 iterations, of which 2,000
171 were later dropped as warm-up iterations. Convergence of chains was tested by comparing the
172 between- and within-chain variance using the \hat{R} statistic described in Gelman, Carlin, Stern,
173 and Rubin (2014). Subsequently, we calculated average posterior distributions over all m
174 joint posterior distributions. We tested the fit of the ensemble model by means of posterior
175 predictive checking. Posterior predictive checking generates replicated data drawn randomly
176 under the model, which can then be compared to the observed data (Gelman et al., 2014). If

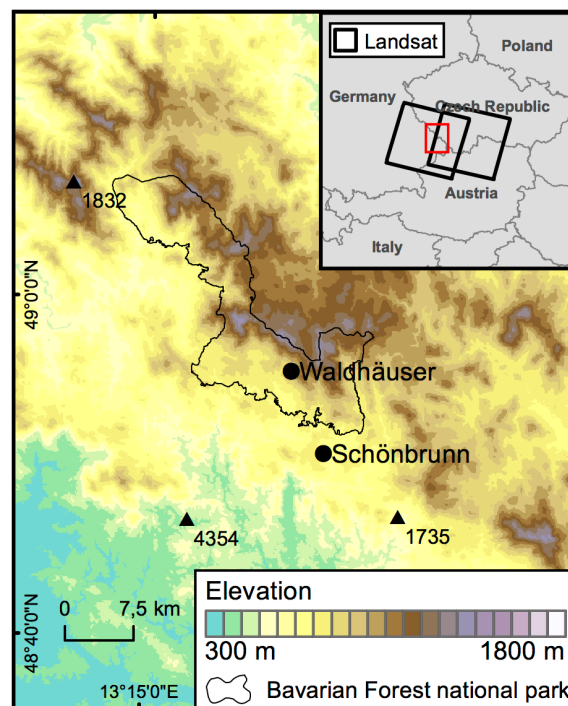
177 the model fits the data well, there should be no systematic differences between the replicated
178 data and the observed vegetation index time series.

179 **Case study: Drivers of changing phenology in broadleaved and coniferous forests**

180 *Study landscape*

181 The study system is located in the Bavarian Forest National park, with moderate topography
182 ranging from approximately 300 to 1,800 meters in elevation (Figure 1). The ecosystem is
183 characterized by mountain beech forests with *Fagus sylvatica* (European beech) being the
184 leading species. Other broad-leaved tree species include *Acer pseudoplatanus* (Sycamore
185 maple) or *Quercus robur* (English oak), but *F. sylvatica* is by far the most abundant. The
186 higher elevation areas of the ecosystem, in turn, are covered by mountain spruce forests,
187 consisting mainly of *Picea abies* (Norway spruce), with some *Abies alba* (European silver fir)
188 intermixed.

189



190

191 **Figure 2:** Study landscape with national park boundaries, the three meteorological stations
192 (triangles), and the two in-situ phenological observation sites (points) from the International
193 Phenological Gardens of Europe network.

194 ***Landsat data***

195 We identified suitable sampling locations for both broadleaved forests (i.e., mountain beech
196 forests) and coniferous forests (i.e., mountain spruce forests) using an existing land cover map
197 generated by National Park authorities from aerial imagery. To exclude young and/or
198 disturbed forests we only sampled mature stands, that is, stands with a minimum age of 60
199 years. The sample size was set to $n = 25$ and the number of samples in the ensemble was set
200 to $m = 10$. We also tested alternative sample and ensemble sizes, but found no substantial
201 differences in the results (data not shown). For each pixel sampled, cloud-, snow-, and
202 shadow-free EVI time series were generated from the full Landsat archive (see Senf et al.
203 (2017) for details on the Landsat processing). We preferred EVI over other vegetation indices
204 as it has been shown to allow better estimation of key phenological dates from remote sensing
205 data (Klosterman et al., 2014).

206 ***Temporal drivers of inter-annual variability in start of season***

207 Past research indicates substantial shifts in leaf unfolding of the two major broadleaved and
208 coniferous tree species found in our study site under climate change (Menzel et al., 2006).
209 However, the underlying drivers are not fully understood. While warmer pre-season
210 temperatures likely support earlier leaf unfolding, warmer winters might have an inverse
211 effect on spring phenology due to the absence of winter chilling (Y. H. Fu, Zhao, et al., 2015).
212 We hence chose two climatological variables hypothesized to influence broadleaved and
213 coniferous land surface phenology in our study ecosystem: pre-season temperature, winter
214 chilling, and their interaction accounting for potential modulating effects of winter chilling on
215 pre-season temperature effects (Y. H. Fu, Piao, et al., 2015). Pre-season temperature was
216 defined as the average mean daily temperature T_{mean} in the months April and May, being the
217 approximate month of vegetation green-up in our study region (Senf et al., 2017). Winter
218 chilling was defined as the number of days with $0 < T_{mean} \leq 5^\circ C$ as suggested by previous
219 studies (Y. H. Fu, Zhao, et al., 2015). Both climatic variables were generated from three

220 meteorological stations operated by the German Weather Service (station-ids: 1735, 1832,
221 4354; Figure 1; [ftp://ftp-](ftp://ftp-cdc.dwd.de/pub/CDC/observations_germany/climate/daily/kl/historical/)
222 [cdc.dwd.de/pub/CDC/observations_germany/climate/daily/kl/historical/](ftp://ftp-cdc.dwd.de/pub/CDC/observations_germany/climate/daily/kl/historical/)). We calculated
223 average time-series of both meteorological variables for the complete study site and there is
224 hence no spatial variation in the meteorological variables among sampling locations. Finally,
225 we restricted the time-series to our study period (1985 – 2015) and z-transformed the data to
226 represent anomalies in units of standard deviation.

227 ***Comparison to in-situ measurements***

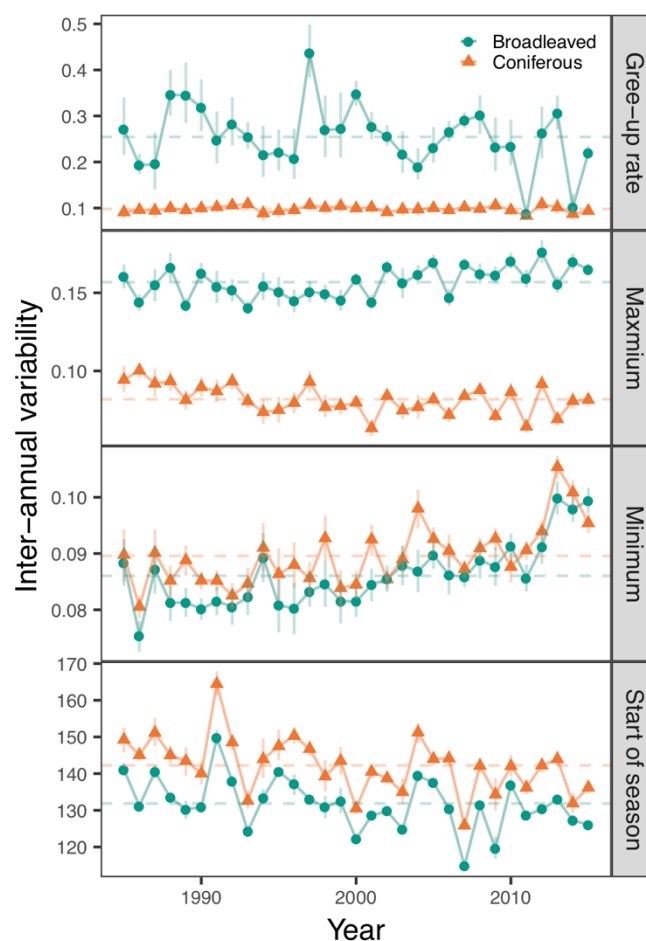
228 After sampling the joint posterior distribution for each model in the ensemble and
229 subsequently averaging over the ensemble, we evaluated the predictive power of the model by
230 comparing the predicted inter-annual variability in the start of season against in-situ
231 observations of leaf unfolding recorded for each three *F. sylvatica* trees and three *P. abies*
232 trees at two phenological gardens within and in close proximity to the National Park,
233 respectively (see Figure 1). The phenological observations were acquired from the
234 International Phenological Gardens of Europe network (stations: Waldhäuser and
235 Schönbrunn; Figure 1; <http://ipg.hu-berlin.de/>).

236 **Results**

237 All models in the ensemble showed good convergence with \hat{R} smaller 1.01 for 99% of the
238 model parameters. Further, uncertainty in the posterior could be substantially reduced by our
239 ensemble approach (Figure S1), with little additional changes in the posterior after seven
240 iterations. Posterior predictive checks indicated no systematic deviation between simulated
241 and raw values (Figure S2), and 96% of the raw values were within the 95% credible interval
242 of the posterior simulations.

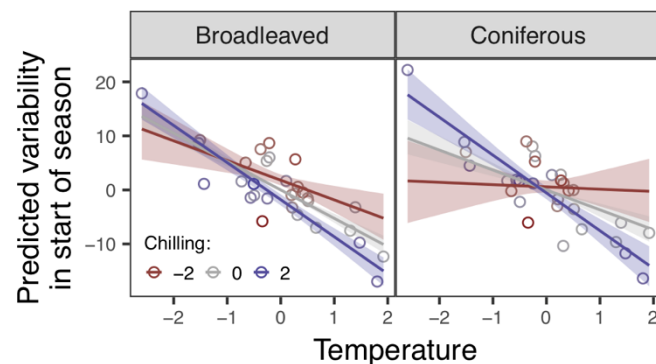
243 We identified substantial inter-annual variability in three out of four phenological
244 parameters (Figure 3). The start of season and minimum EVI both showed similar trends
245 across broadleaved and coniferous forests, with an earlier start of season and a higher

246 minimum EVI over time. There were also some very late years (e.g., 1991) and some very
 247 early years (e.g., 2007 and 2009) identifiable for both broadleaved and coniferous forests. No
 248 substantial inter-annual variability could be detected for the maximum EVI, yet there was a
 249 slight decreasing trend over time for coniferous forests and a slight increasing trend over time
 250 for broadleaved forests. Substantial inter-annual variability could also be detected for the
 251 green-up rate of broadleaved forests, whereas coniferous forests had more stable green-up
 252 rates over time.



253
 254 **Figure 3:** Ecosystem-scale inter-annual variability in the four phenological parameters
 255 estimated from Landsat time series for two forest types. Dots represent the median of the
 256 posterior distribution and error-bars extent from the 2.5% to the 97.5% quantile of the
 257 posterior (95% credible interval). The dashed horizontal lines indicate the long-term average.
 258

259 The driver analysis revealed that the inter-annual variability in the start of season is
 260 highly sensitive to pre-season temperature (Figure 4), with a $-3.74 \text{ d } ^\circ\text{C}^{-1}$ earlier start of
 261 season for broadleaved and a $-2.68 \text{ d } ^\circ\text{C}^{-1}$ earlier start of season for coniferous forests,
 262 respectively. The effect of pre-season temperature on inter-annual variability in start of season
 263 was, however, modulated by the total number of chilling days in the preceding winter, with a
 264 decreasing effect of pre-season temperature on inter-annual variability in the start of season
 265 with a decreasing number of chilling days.
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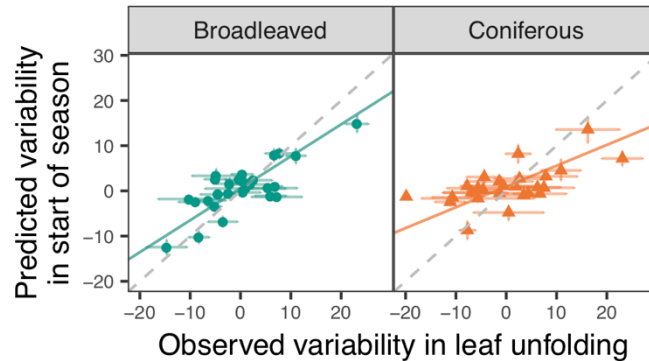


267
 268 **Figure 4:** Ecosystem-scale inter-annual variability in start of season predicted from our model
 269 using pre-season temperature and winter chilling days as predictors. Both predictors are z-
 270 transformed and thus represent anomalies in standard deviations from the long-term average.
 271 Ribbons indicate the 95% credible interval.

272
 273 Comparing predictions from the model (i.e., based on the regression relationship
 274 shown in Figure 4) to in-situ observations of bud-break (Figure 5), we found strong
 275 agreement for broadleaved forests (Pearson's $r = 0.79$ [0.74 – 0.80] and RMSE = 4.88 [4.60 –
 276 5.21] days) and moderate agreement for coniferous forests (Pearson's $r = 0.69$ [0.63 – 0.75]
 277 and RMSE = 6.57 [6.00 – 7.13] days). Hence, our model – based only on pre-season
 278 temperature and winter chilling and calibrated solely from Landsat time series – was able to

279 predict the general inter-annual variability in spring phenology as recorded in in-situ
280 measurements.

281



282

283 **Figure 5:** Comparing the ecosystem-scale inter-annual variability in start of season predicted
284 from our model to in-situ observations of leaf unfolding derived from two phenological
285 gardens in and around the national park (see Figure 2). Dots represent the median of the
286 posterior (prediction; y-axis) and the arithmetic mean (leaf unfolding; x-axis). Uncertainty in
287 the prediction is expressed by the 95% credible interval, whereas error-bars for the leaf
288 unfolding represent 95% confidence intervals.

289

290 Discussion

291 We here presented a new model for inferring drivers of changing land surface phenology from
292 Landsat time series. Our study is among the first to explicitly integrate Landsat time series
293 with meteorological observations in order to deliver insights into the drivers underlying
294 changing phenological patterns. Hence, instead of creating phenological products from remote
295 sensing data that are subsequently used in analysis, we here suggest to calibrate phenological
296 models directly from the millions of spectral observations available through long-term
297 satellite archives such as Landsat. The advantage is a reduced number of, and redundancy in,
298 processing steps, as well as a better propagation of uncertainty, which is otherwise largely
299 neglected in remotely sensed phenological products. Moreover, given that the Landsat data
300 used in this study have global coverage and are free to access, and considering emerging

301 cloud-computing capabilities like the Google Earth Engine (Gorelick et al., 2017), it is at the
302 researchers fingertips to use Landsat for understanding phenological dynamics globally.

303 Integrating Landsat time series with meteorological observations allowed us to infer
304 drivers of inter-annual variability in the start of season, building a predictive model that is
305 solely calibrated from Landsat time series. While we here only modeled the start of season,
306 any of the other parameters (i.e., minimum spectral index value, maximum spectral index
307 value or green-up rate) could be modeled, too. As shown in our case study, the model was
308 capable of predicting inter-annual variability in leaf-unfolding observed in-situ, despite the
309 relatively simple model structure. This finding reinforces previous studies showing that more
310 complex models do not necessarily lead to better predictions (Basler, 2016; Yang et al.,
311 2012). In fact, RMSE values obtained by our model are in a similar range or even lower than
312 comparable or more complex phenological models (Basler, 2016; Hufkens et al., 2018),
313 emphasizing the predictive power of our model. Further, as our model is calibrated on land
314 surface phenology (Morisette et al., 2009), it might better represent the ecosystem-scale
315 drivers of phenology than models based on in-situ observations. This is mainly due to a more
316 aggregated view going beyond individual species (Jeremy I. Fisher et al., 2007). This
317 difference might be particularly important for calibrating phenological models used in global
318 vegetation models (Richardson et al., 2012; Yang et al., 2012). Finally, in-situ phenological
319 measurements themselves are often highly uncertain, with large differences among
320 observations taken in close proximity (i.e., large confidence-intervals along the *x*-axis in
321 Figure 5). This uncertainty, however, is often neglected in phenological models.

322 The drivers of inter-annual variability in start of season identified in this study largely
323 corroborate the current literature, with pre-season temperature being a major control of spring
324 phenology (Basler, 2016; Menzel et al., 2006; Yang et al., 2012). However, recent studies
325 also showed a modulating effect of winter chilling on pre-season temperature effects (Y. H.
326 Fu, Piao, et al., 2015), for which we found support in our study (Figure 4). Thus, even though

327 spring temperatures might further increase in the future, warming-related reductions in
328 chilling days might slow down the advance of spring phenology under climate change (Y. H.
329 Fu, Zhao, et al., 2015). Yet, the exact processes are still elusive and additional drivers such as
330 precipitation or the previous year's phenology are discussed in the current literature (Y. H.
331 Fu, Piao, et al., 2015; Y. S. Fu et al., 2014). Our model – in conjunction with global long-term
332 satellite records from the Landsat archive – can be extended to include those drivers and thus
333 offers great potential for understanding phenological responses under climate change in more
334 detail, and for areas/time scales yet hardly covered alternative data sources.

335

336 **Acknowledgements**

337 Cornelius Senf and Tobias Krueger are both funded through IRI THESys by the German
338 Excellence Initiative. We thank Prof. Dr. Frank-M. Chmielewski of Humboldt-Universität zu
339 Berlin for access to the phenological observations.

340

341 **Data and code accessibility**

342 All data and code used in this publication are available for review at
343 https://github.com/corneliussenf/phenoBayes_drivers and will be archived using
344 <https://zenodo.org/> after the paper has been accepted for publication.

345

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