

1 **Insights into agricultural influences and weathering processes from major**
2 **ion patters in the Wiesent River catchment (Germany)**

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22 **ABSTRACT**

23 A study of the geochemical downriver evolution of the Wiesent River and its tributaries
24 revealed unexpected decreases of nitrate downstream from the source, with maximum values
25 of 31.6 mg L⁻¹ to minimum values of 10 mg L⁻¹ near the mouth. This trend persisted over the
26 length of the river even though increased agricultural activities are evident in the downstream
27 section of the catchment. This pattern is caused by fertilizer inputs via fast fracture flow and
28 fast conduits from karst lithology in the upstream area that may have reached the river's
29 source even from beyond the hydrological catchment boundaries. Further downstream, these
30 influences became diluted by tributary inputs that drain subcatchments dominated by
31 claystone and sandstone lithologies that increased potassium and sulfate concentrations. In
32 addition, except for inputs in the source area, groundwater contributions along the river
33 course were found to be minor as attested by seasonal temperature trends of the river, with
34 values ranging from 18.9 °C during summer to 0.9 °C during winter. This seasonality was
35 absent at the source where water temperatures were consistently 9 °C throughout the year,
36 which is a typical value for central European groundwater.

37

38 **KEYWORDS**

39 watershed, hydrology, geochemistry, weathering, karst, fertilizers

40

41 **INTRODUCTION**

42 The downriver evolution of various river systems has been examined in several studies
43 worldwide (Meybeck, 1982; Meybeck, 1993; Yang *et al.*, 1996; Richey *et al.*, 2002). Most of
44 these were aimed at weathering processes and continental fluxes of major element and carbon
45 phases. Following these studies, work on smaller river systems were also carried out (Barth *et*
46 *al.*, 2003; Kanduč *et al.*, 2007; Stögbauer *et al.*, 2008; Brunet *et al.*, 2011), and most of these
47 showed increasing conductivities and fluxes with closer proximity to confluences to larger
48 river systems. This trend corresponds to increasing element loads further downstream and is
49 usually related to increasing anthropogenic activities including traffic infrastructure,
50 settlements and agricultural practices in the downstream sections of catchments.

51 Here we present a small river system of 72.4 km length, the Wiesent River (Figure 1).
52 Being dominated by karst lithology, it was first studied as an ideal natural laboratory for
53 investigations of the aqueous carbon cycle in karst areas and its response to excessive carbon
54 loss to the atmosphere, particularly in its source area (van Geldern *et al.*, 2015). The
55 catchment is situated in a dolomite and calcite-dominated terrain that trends towards shale and
56 sandstone terrains in the lower reaches of the river. This allows for interpretations of spatial
57 heterogeneity in river water chemistry, while sampling over almost one year enabled temporal
58 interpretations of seasonal variations in riverine solutes, carbon transport, CO₂ evasion and
59 associated changes in the carbon isotope distribution of dissolved inorganic carbon (DIC).
60 The Wiesent River is also influenced by anthropogenic activities including farming, fish
61 farming and tourism. In particular, farming activities along the river course were expected to
62 significantly influence the water chemistry. These influences may be stronger in a karst-
63 dominated lithology as compared to catchments with clastic sediments, due to the fast conduit
64 systems present in karst. These conduits also make karst aquifers highly vulnerable to

65 anthropogenic pollution because of the virtual absence of filter effects that are common in
66 pore aquifers. Moreover, it is known that fertilizers can lower pH values and thus cause
67 additional weathering that should, in turn, be particularly pronounced in karst that is
68 susceptible to acidity changes (Perrin *et al.*, 2008; Gandois *et al.*, 2011).
69 The objectives of this study were to assess the spatial and temporal downriver evolution of
70 major ion chemistry in a typical karst river. Specifically, we investigated whether the
71 hypothesized continuous increase of solutes along the flow path of the river could be
72 observed. With this in mind, we evaluate relative influences of natural versus anthropogenic
73 impacts together with the role of groundwater contributions along the river course.
74 Disentangling the relative importance of these factors advance the understanding of source-
75 related inputs and at the same time, provide information on the dilution of anthropogenic and
76 weathering inputs in space and time. Regarding the high vulnerability of karst terrains by
77 agricultural land use and anthropogenic pollution, this is of particular interest with regard to
78 the ecological functioning of rivers with large proportions of limestone in their catchment
79 lithology.

80 **STUDY AREA**

81 The study area was located in southern Germany, in Northern Bavaria between the
82 cities of Bamberg, Bayreuth and Nuremberg (Figure 1). The total catchment area of the
83 Wiesent River is 1040 km². The source of the Wiesent River is located in the northwest of the
84 catchment area, at the hamlet of Steinfeld at 445 meters above sea level (m.a.s.l.). From this
85 point the river flows in a southeast to south direction. Near the town of Goeßweinstein, the
86 general flow direction changes westward until the river finally reaches the Regnitz River at
87 240 m.a.s.l. The total length of the Wiesent River is 72.4 km and the most important
88 tributaries are the rivers Kainach, Truppach, Puettlach, Aufsess, Leinleiter, and Trubach

89 (Figure 2). Two discharge gauges are located along the main river at the villages of Hollfeld
90 and Muggendorf. They show averaged long-term discharge values of $1.04 \text{ m}^3 \text{ s}^{-1}$ and 7.11 m^3
91 s^{-1} , respectively (LfU, 2014).

92 The study area is located in a karst region with lithologies of calcites and dolomites
93 that are occasionally interbedded with marly layers. These sediments form the Franconian
94 Jura uplands, which belong to the Franconian Alb that is part of the South German
95 Scarplands. Overall, late Jurassic carbonates cover the majority (~67%) of the catchment area
96 (Figure 2). These comprise a limestone sequence of up to 200 m thickness that is in part
97 dolomitized, and forms a typical karst landscape of the area with incised rivers and other dry
98 valleys, dolines and caverns. This carbonate sequence is overlain by relict patches of Late
99 Cretaceous sandstone with interbedded marl and shale layers.

100 All sedimentary layers dip at a shallow angle in an eastward direction. As a
101 consequence, lower Jurassic and Late Triassic sandstones and claystones that underlie the
102 thick Late Jurassic carbonate succession crop out in the south-western part of the catchment.
103 Additional sandstone and claystone lithologies of the same Triassic sequence also appear in
104 the eastern part of the main basin, in subcatchments of the tributaries Truppach and Puettlach.
105 Here, the sandstones and claystones were lifted to the surface along a northwest-southeast
106 striking fault.

107 The vegetation in the south-western part of the catchment is characterized by broad-leaved
108 forest dominated by beeches and scattered small fruit orchards. Coniferous trees largely forest
109 northern and central parts of the catchment, while the eastern regions consist of grassland
110 pastures. Numerous small villages are present in the basin and, particularly in the downstream
111 part of the catchment, agriculture plays an important role in land use.

112 **METHODS**

113 Eight locations (sites 1 to 8 in Figure 2) were sampled during eight field campaigns
114 between January and November 2010 in order to cover all seasons. Starting in April 2010, the
115 six most important tributaries were additionally sampled during the last six sampling
116 campaigns. All tributaries were sampled near their confluences with the Wiesent River (sites
117 9 to 14 in Figure 2), with water samples collected manually from the middle of the rivers. All
118 bottles were rinsed several times with sample water before filling. Samples destined for major
119 ion analyses were filled without headspace in double closure 100 mL high-density
120 polyethylene bottles with an inner stopper (product item LR100-2, Gosselin SAS, Borre,
121 France). These were then stored in the dark and at 4°C until analysis.

122 Measurements of temperature, pH, electrical conductivity (EC), and dissolved oxygen (DO)
123 were conducted in the field by a multi parameter instrument (WTW Multi 350i, WTW GmbH,
124 Weilheim, Germany). All probes were calibrated prior to each field campaign. The pH and
125 conductivity probes used built-in automatic temperature compensation functions, and
126 temperature precision is quoted as ± 0.1 °C by the manufacturer. Conductivity values were
127 referenced to 25 °C by a non-linear correction function (nLF) for natural waters. For pH
128 measurements precision was better than ± 0.05 pH units (1σ) and for conductivity
129 determinations it was better than ± 5 $\mu\text{S cm}^{-1}$. Dissolved oxygen was determined by a
130 galvanic sensor and automatically corrected for ambient air pressure by a built-in pressure
131 sensor.

132 Total alkalinity (TA) was determined from 100 mL water samples directly in the field by a
133 Hach Digital Titrator (Model 1690001, Hach Company, Loveland, CO, USA). The titration
134 endpoint was determined by colour changes of a pH indicator (bromcresol green-methyl red).
135 At a pH < 9, hydroxide (OH^-) does not contribute to TA at an appreciable concentration and

136 other TA contributing species (BOH_4^- , H_3SiO_4^-) can be regarded as negligible in the waters
137 examined by this study. Therefore, TA essentially contains only carbonic alkalinity and the
138 titration value was used for the calculation of carbonate species concentrations, which is
139 mainly bicarbonate (HCO_3^-) in the observed pH range (see Verma *et al.*, 2015)
140 Major cations (Na^+ , K^+ , Li^+ , NH_4^+ , Ca^{2+} , Mg^{2+}) and anions (F^- , Cl^- , NO_3^- , NO_2^- , SO_4^{2-} , PO_4^{3-})
141 were determined by ion chromatography (ICS 2000, Thermo Dionex, Sunnyvale, CA, USA).
142 Prior to ion chromatography analysis, samples were filtered with 0.45 μm disk filters, which
143 were flushed with sample before the autosampler vials were filled. Limit of quantification
144 (LOQ) for major ions was 0.1 mg L^{-1} with a typical precision of <5% (1σ) relative standard
145 deviation (RSD) based on the repeated analysis of two control standards treated as unknowns
146 in the lower and upper calibration range. Typical charge errors derived from equivalent
147 concentrations of ion chromatography and field titration data were between 5 and 12%.

148 **RESULTS**

149 *Major ion chemistry of the Wiesent River*

150 Major ion concentrations and the described trends along the river course are shown in
151 Figure 3. Detailed analytical data are available in the supplementary material to this article
152 from the website of the host journal, and are additionally archived in PANGAEA.[†]

153 Calcium showed the highest concentrations amongst all cations, followed by
154 magnesium, sodium, and potassium. Lithium concentrations were always below LOQ (<0.1
155 mg L^{-1}). Ammonium was detected in only two samples during February, with values of 0.11

[†] www.pangaea.de (DOI for data set pending)

156 and 0.28 mg L^{-1} , and was below the limit of detection (LOD) in all others. In contrast to most
157 other parameters, calcium and magnesium concentrations showed little variability at the
158 source during the various sampling campaigns, although their general downstream patterns
159 were almost identical for all sampling campaigns. Calcium exhibited a range of 90 to 145 mg
160 L^{-1} , whereas magnesium concentrations ranged between 15 and 45 mg L^{-1} (Figure 3).
161 Calcium concentrations showed an initial decrease between the source and sampling site 2,
162 whereas magnesium increased within the same direction. Further downstream, calcium
163 concentrations remained fairly stable and ranged between 105 and 130 mg L^{-1} , with exception
164 of the November campaign that was characterized by lower values further downstream. After
165 the initial rise, magnesium concentrations tended to decrease until sampling site 4 and showed
166 only minor variations thereafter. In particular, magnesium concentrations showed a clear
167 correlation with corresponding conductivity values for different sampling campaigns, which
168 in turn were correlated with precipitation events and resulting increased discharge values
169 (data shown in van Geldern *et al.*, 2015). This trend was also obvious to a lower extent for
170 calcium, especially during the November campaign.

171 Sodium concentrations ranged from 8 to 15 mg L^{-1} , stabilizing at a value of 11.2
172 $(\pm 0.6) \text{ mg L}^{-1}$ at the source followed by a slight initial rise [arithmetic average $(\pm 1\sigma)$]. Values
173 declined further downstream to concentrations of about 8 mg L^{-1} at the lower river course.
174 Potassium concentrations increased from $1.04 (\pm 0.24) \text{ mg L}^{-1}$ at the source to values of
175 around 2.0 mg L^{-1} downriver. Maximum values of 3.5 mg L^{-1} were recorded during the
176 November sampling campaign at site 4 and also at locations further downstream. Note that
177 during this campaign most other parameters, as for example calcium and magnesium
178 concentrations or conductivity, showed lower values, a trend that was most likely related to
179 heavy rains prior to that campaign.

180 The ranking of anions in order of decreasing concentrations at the Wiesent River
181 source was bicarbonate, nitrate, chloride, and sulfate with concentrations of 357 (± 7), 29.6
182 (± 2.7), 22.7 (± 1.6), and 15.8 (± 1.5) mg L⁻¹, respectively. Interestingly, the downstream
183 behavior revealed different patterns for these anions. Nitrate showed an almost linear
184 monotonic decline over the river course to 17.6 (± 1.6) mg L⁻¹ at the final sampling site and
185 bicarbonate decreased to an average of 307 (± 22) mg L⁻¹. The lowest bicarbonate values were
186 measured along the lower course during the November sampling campaign. A pronounced
187 decrease was observed after the confluence with the Truppach, whereby bicarbonate
188 concentrations dropped below 250 mg L⁻¹ in November. The minimum concentration of 228
189 mg L⁻¹ was recorded at site 6, located downstream of the Puettlach tributary. During other
190 campaigns, bicarbonate concentrations along the lower course (site 4 and downstream),
191 ranged from 340 to 280 mg L⁻¹ (Figure 3).

192 The chloride downstream evolution was similar to sodium or magnesium, with a
193 distinct increase from the source to sampling site 2 and subsequent decreasing values.
194 Chloride concentrations along the lower course ranged from 15 to 25 mg L⁻¹. In contrast,
195 sulfate behaved similar to potassium with a continuous increase along the Wiesent River up to
196 values of 20 to 30 mg L⁻¹ near the mouth of the Wiesent at its confluence to the Regnitz. A
197 larger variability of sulfate could be observed downstream of sampling site 4. Here, maximum
198 concentrations of up to 42 mg L⁻¹ were recorded during the November sampling campaign.
199 However, an exception from this behavior was observed in February, when the second highest
200 sulfate concentrations were measured at the middle and lower course while the downstream
201 conductivity curve showed maximum values (Figure 3).

202 Low concentrations for fluoride and phosphate above LOD were detected in 12 and 11
203 out of 64 analyses, respectively. When detected, typical concentrations of F⁻ and PO₄³⁻ were

204 below 0.3 mg L^{-1} and never exceeded 0.5 mg L^{-1} . Nitrite was not detected in any of the
205 sampling campaigns in the main stem of the Wiesent River (cf. supplementary material)

206 The tributary and Wiesent River samples are characterized by their major ions in a
207 Piper diagram (Figure 4). Here the tributaries are marked with their numbers in brackets
208 according to Figure 2. The tributaries Kainach (9), Aufsess (11), and Trubach (14) were
209 largely similar to the water chemistry of the main stem of the Wiesent River. The only
210 exception was the Leinleiter tributary (13) with its lower magnesium concentrations, but most
211 other major ion concentrations were similar to those in the main river. Note that the two
212 eastern tributaries Truppach (10) and Puettlach (12) plot differently in Figure 4. For instance,
213 magnesium concentrations were below 12 mg L^{-1} in all samples taken from the Truppach
214 tributary. Moreover, both the Truppach and the Puettlach, were characterized by lower
215 bicarbonate, increased potassium, and high sulfate concentrations. These patterns were also
216 transferred to the middle and lower course of the Wiesent River. This was particularly evident
217 after precipitation events, for instance during November. Finally, low concentrations of
218 phosphate (0.11 to 0.72 mg L^{-1}) were found during all campaigns in the Truppach and
219 Puettlach, but not in the other tributaries.

220 **DISCUSSION**

221 *Geological and anthropogenic controls*

222 The downstream evolution of the parameters discussed above indicates that
223 groundwater is one of the major contributors to the Wiesent River and its tributaries. The
224 expectation is that groundwater infiltration to the river occurs primarily by the piston-flow
225 effect and minor contributions should be derived from overland runoff or soil water. The
226 latter processes are normally observed during snowmelt and more intensive precipitation
227 events. This would particularly hold true for precipitation in tributaries that drain poorly

228 permeable sediments such as in the eastern part of the catchment (i.e., Truppach and
229 Puettlach). Mixing of these tributary waters with the main river is the most important process
230 for dilution of total dissolved solids (TDS), as indicated by lower conductivity values. In
231 contrast, in the upstream part of the Wiesent Catchment, runoff generation should play a more
232 important role than in permeable karst lithology with fast infiltration rates.

233 When tributaries are chemically different as compared to the main stem of the Wiesent
234 River, their influences are expected to be obvious. For instance, lower magnesium
235 concentrations in the Truppach likely caused the pronounced drop in magnesium after
236 sampling sites 3 in the Wiesent River. This decline was even more pronounced during periods
237 of low conductivity after precipitation events.

238 On the other hand, contributions of tributaries that drain a similar lithology as the
239 Wiesent River are harder to detect by natural tracers. For instance, near the catchment outlet,
240 the lithology along the Wiesent River valley changes from carbonate rocks to claystone. The
241 formation is known as the Feuerletten Formation from the Late Tertiary. Except for small
242 changes in calcium, magnesium, and sodium the measured parameters did not respond to this
243 prominent change in riverbed lithology. This indicates that groundwater contributions at the
244 final river stretch likely become less important due to the less permeable nature of the
245 underlying thick claystone formation.

246 At the source of the Wiesent, Ca^{2+} , Mg^{2+} , and HCO_3^- account for 86 % of the TDS and
247 along the river course, this value varies only slightly between 81 and 88 %. This shows that
248 carbonate weathering of calcite (CaCO_3) and dolomite ($\text{CaMg}(\text{CO}_3)_2$) mainly accounts for the
249 high concentrations of these three ions. The stoichiometry of carbonate weathering reactions
250 demands that carbonate-derived calcium and magnesium should be equal to those of
251 bicarbonate when expressed in (milli)equivalents per liter (meq/L). Such a plot of HCO_3^-

252 versus ($\text{Ca}^{2+} + \text{Mg}^{2+}$) concentrations reveals a general excess of calcium plus magnesium over
253 bicarbonate (Figure 5A). This excess of cations is partially balanced by Cl^- , SO_4^{2-} and NO_3^- ,
254 and suggests that fractions of calcium or magnesium concentrations in the watershed were not
255 exclusively derived from carbonate weathering. Other potential natural sources of calcium are
256 gypsum or silicate weathering of Ca-bearing minerals such as anorthite. However, both
257 minerals are of minor importance in the Jurassic limestone formations and are therefore
258 unlikely candidates for leaching Ca^{2+} and Mg^{2+} .

259 Other sources for this excess include mineral fertilizers that may contain calcium- or
260 magnesium-bearing compounds. This should also correlate with other solutes that are
261 typically derived from these fertilizers, such as nitrate. Further anthropogenic sources that
262 may exist in the watershed include organic fertilizers or domestic sewage. Typical
263 anthropogenic compounds are phosphate, sulfate and nitrogen compounds (NO_3^- , NO_2^- , NH_4^+)
264 from manure and mineral fertilizers. In particular, elevated nitrate concentrations are often
265 related to the use of such fertilizers. During this study, the highest concentrations of nitrate
266 were found at the source of the Wiesent River with a stable value of $29.6 (\pm 2.7) \text{ mg L}^{-1}$. This
267 stability of nitrate concentrations at the source suggests that nitrate may have been
268 continuously transferred into the feeding karst aquifer that acted as a large reservoir.
269 Subsequently, nitrate likely has accumulated over decades and seasonal variations became
270 negligible.

271 This supports the interpretation that the observed excess of calcium and magnesium is
272 a result of long-term anthropogenic influences on the groundwater, particularly at the Wiesent
273 River spring and the upper reaches. It thus indicated the influences from farming activities in
274 the upper part of the catchment. This is surprising, because the Wiesent River has its source in
275 a rural, largely natural region where forests dominate large portions of the catchment and

276 industry does not play a significant role. Therefore, anthropogenic signals in river water
277 chemistry were expected to be minimal, with inorganic aquatic chemistry mainly mirroring
278 weathering processes or changes in lithology. However, particularly in karst systems,
279 anthropogenic influences may directly enter the river via fast conduit systems and fracture
280 flow.

281 As opposed to the expected trend of increasing nitrate concentrations further
282 downstream where agriculture and urbanization is more prevalent, a contrary pattern was
283 observed: nitrate concentrations were found to be highest at the source of the Wiesent River
284 (see discussion above) and subsequently decreased with distance from the headwaters during
285 all sampling periods (Figure 3). This pattern further suggests the long-term influences of
286 agricultural land use on the groundwater, particularly in the upstream part of the watershed.
287 Even though agricultural activities are also prevalent in the downstream part of the catchment,
288 their associated nitrate releases may not reach the river as rapidly because the lithology
289 contains a high proportion of fine sandstones and shales that slows or hampers groundwater-
290 related transport to the river.

291 As for other anthropogenic parameters, road salt and domestic sewage are a primary
292 source of sodium and chloride. Increases in chloride or sodium concentrations were not
293 observed at the source during colder periods. This could either mean that they were dampened
294 in the large feeding karst reservoir or that road salting had a minor influence on the river
295 water chemistry. Since both chloride and potassium showed no distinct seasonality over the
296 entire course of the river, a pronounced influence of this source is likely small in this
297 catchment. It is more likely that these elements reach the river via agricultural activities that
298 are more dominant in the source region, as previously stated. At this juncture, it is unclear
299 why both Cl^- and Na^+ show significant increases from the spring to the second sampling

300 station (site 2) although this trend may be related to the presence of point sources such as
301 sewage release from private housing or farms.

302 Cl^- was also found to be in excess of Na^+ and K^+ and their cation-anion plot does not follow a
303 1:1 ratio (Figure 5B). The Cl^- excess could in part be balanced by excess Ca^{2+} and Mg^{2+} that
304 does not account for carbonate weathering. When assuming that this cation excess stems from
305 fertilizer use in the upstream part of the catchment, it supports the hypothesis that fertilizers
306 are more pronounced in upstream section and may have entered the river via fracture flow of
307 the present karst system.

308 Other fertilizers derived from potash with compounds such as KCl or K_2SO_4 may also
309 have added to the fertilizer load of the river. However, both K^+ and SO_4^{2-} have their lowest
310 values in the source region and show increases further downstream. At first sight, this pattern
311 does not appear to support the hypothesis of more pronounced fertilizer input in the upstream
312 part of the catchment. However, it is possible that both ions actually enter the river as a result
313 of fertilizer transport in the upstream, karst-dominated section of the catchment. The baseline
314 values at the source may then have increased further downstream mainly as a result of K^+ and
315 SO_4^{2-} released from weathering processes in the clay-rich shale formations in the eastern
316 subcatchments, and discharged into the main river stem via tributaries. These and other
317 mechanisms of tributary influences are discussed in the following section.

318 *Downstream geochemical evolution*

319 After setting of the initial values for physico-chemical parameters at the source, many
320 parameters show changes along the river course. This may be either related to processes that
321 occur within the main river itself such as seasonal temperature changes, or due to the inflow
322 and mixing of waters from different sources.

323 An important question in river research is the interaction between river water and
324 surface water within the hyporheic and riparian zones along the river course (Krause *et al.*,
325 2009; Krause *et al.*, 2014). Depending on the hydraulic system, the river can lose water to the
326 groundwater by outflow through the riverbed. This situation is denoted as a losing stream.
327 Alternatively, water from the riparian zone can feed the river, which is denoted as a gaining
328 stream. Different environmental tracers, for example temperature, stable isotopes, or
329 wastewater-related pollutants, have been used to investigate these interactions in detail
330 (Engelhardt *et al.*, 2014; Fox *et al.*, 2016). Furthermore, tributaries with different water
331 chemistry that mix with the main river can also induce changes in water chemistry when
332 differences and fluxes are large enough to cause a measurable shift.

333 The downstream temperature development of the Wiesent River varies seasonally,
334 with higher temperatures in summer and lower temperatures in winter (Figure 3).
335 Groundwater seeps would shift the warm or cold temperature curve locally towards the
336 average groundwater temperature of about 9 ° C for this region. However, the spatial
337 resolution along the river is not detailed enough to detect point-source groundwater inflow.
338 This is because groundwater temperature can be expected to adjust quickly to the temperature
339 of the river water. Small groundwater contributions would therefore only be detectable close
340 to the point of inflow. However, a steady and spatially ubiquitous groundwater inflow along
341 the whole river course would likely dampen the seasonal changes of the river. Such trends
342 could only be observed at the source and for the remainder of the river it is more likely that
343 groundwater contributions are of minor importance.

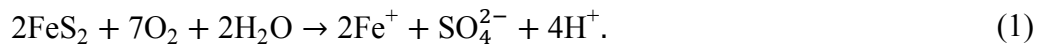
344 A closer examination of the downstream curves showed a notable increase in
345 magnesium concentrations of 7 to 13 mg L⁻¹ from the source to sampling site 2 (Figure 3). A
346 larger fraction of dolomite weathering might have increased the portion of magnesium over

347 calcium. However, this shift was also correlated with increases of sodium and chloride. To a
348 large degree, both elements are likely of anthropogenic origin and this pattern suggests that
349 the Mg^{2+} increase has an anthropogenic component. This is supported by the fact that the
350 increase in chloride was larger than the corresponding increase in sodium, so that chloride
351 could have additionally buffered the magnesium increase. However, even when accounting
352 for Cl^- , a negative charge deficit persisted and Mg^{2+} must have been additionally balanced by
353 other anions. As no other major anions (SO_4^{2-} , NO_3^-) showed a corresponding increase, the
354 surplus of magnesium is most likely buffered by organic compounds that were not measured
355 in this study.

356 Potassium and sulfate downstream evolutions in the Wiesent River showed opposing
357 patterns as compared to the other major ions along the course (see above). Both ion
358 concentrations gradually increased downstream, whereas other major ion concentrations
359 decreased most likely due to dilution effects by the tributaries Truppach and Puettlach that
360 drain a different lithology. Their lithology is dominated by claystone and sandstone (see
361 Figure 2) and increased concentrations of potassium and sulfate were observed in these two
362 eastern tributaries (supplementary material). A potential anthropogenic source could be
363 potassium sulfate, a common inorganic fertilizer that is often used in fruit orchards. Both ions
364 show a fairly strong correlation in the two subcatchments (Pearson correlation coefficient $r =$
365 0.69 , $p = 0.012$) but only about one-tenth of the sulfate is balanced by potassium in this
366 scenario.

367 One plausible natural source for potassium is weathering of orthoclase and other
368 potassium-bearing silicate minerals in the clastic sediments. As a result of these weathering
369 processes, higher potassium concentrations may have been transported into the Wiesent
370 River.

371 The higher sulfate concentrations could have been caused by the dissolution of
372 gypsum and/or weathering of pyrite. Both minerals are present in the Lower and Middle
373 Jurassic formations (Krumbeck, 1956; Schröder, 1968) and pyrite is especially abundant in
374 the thick claystone sequence of the 'Amaltheenton' and 'Opalinuston' formation that cover
375 large areas of the eastern subcatchments (Wippert, 1955). The principles of pyrite oxidation
376 follow a sequence of reactions that involve microbiologically-mediated and oxic
377 decomposition of pyrite followed by an anoxic decomposition of pyrite by ferric iron
378 (Lowson, 1982; Evangelou, 1995). The overall stoichiometry is represented by



379 Pyrite weathering reactions release acidity and result in low pH values. However, lower pH
380 values were neither observed in the Wiesent nor in any of the tributaries. This absence of
381 expected pH decrease could be explained by the presence up to 10% carbonates in the
382 claystones (Schröder, 1968). This means that acidity derived from equation (1) could become
383 directly neutralized by reaction with carbonates.

384 Unexpectedly higher concentrations of potassium and sulfate in the Truppach and
385 Puettlach tributaries was observed in November: samples from this month showed increased
386 concentrations of potassium and sulfate that were also found in the middle and lower reaches
387 of the Wiesent River, downstream of the Truppach and Puettlach confluences (Figure 3). In
388 contrast, most other ions (e.g., calcium or magnesium) were diluted. This dilution can be
389 attributed to increased fluxes of low-mineralized surface runoff following a heavier period of
390 rainfall in November. Surface runoff was also expected to also dilute the concentrations of
391 potassium and sulfate. However, the data suggest that potassium and sulfate were largely
392 derived from weathering processes at the land surface or in the soil zone. They were then

393 mobilized and flushed with increased concentrations along the eastern tributaries into the
394 Wiesent River during precipitation events.

395 **CONCLUSIONS**

396 The Wiesent River has an unusual chemical downriver evolution, particularly for
397 solutes that are often derive from agricultural fertilizers. While increased anthropogenic
398 activities normally occur in the downriver sections of the catchment, an unexpectedly stronger
399 fertilizer-derived signal was found in its upstream area. This is particularly true for nitrate.
400 These agricultural inputs may have been introduced through fast conduit and fracture flow in
401 the karst lithology of the upstream catchment. Agricultural influences may even stem from
402 beyond the catchment surface water boundaries. After leaving the source, these inputs became
403 diluted further downstream, specifically via tributaries that were not sourced from carbonate
404 lithologies. While Ca^{2+} , HCO_3^- and NO_3^- decreased over the course of the river, we found
405 increases of K^+ and SO_4^{2-} in the downstream section that is more influenced by clay
406 lithologies. This enables a division of the river into two sections: an upstream part that is
407 dominated by karst and receives strong fertilizer inputs, and a downstream section that
408 receives less groundwater inputs but is subject to increased contributions by tributaries that
409 drain sandstone and shale lithologies. This highlights the importance of including tributaries
410 in river studies.

411 Overall, our findings indicate that bedrock geology remains the dominant control on
412 the major ion chemistry of the Wiesent River, and that agricultural influences are the
413 strongest near the headwaters despite increased land-use further downstream, due to long-
414 term storage and accumulation in karst aquifers. Future studies should include ecological

415 indicators that may be more sensitive to fertilizer input, such as counts of specific organisms
416 and quantification of sediment loads.

417

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426

427

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- 494

495 **FIGURE CAPTIONS**

496 Figure 1. Location of the study area in Southern Germany. The total area of the Wiesent
497 Catchment is 1040 km².

498
499 Figure 2. Sampling points and lithology of the Wiesent catchment. The upper and central
500 parts of the catchment are dominated by Late Jurassic limestone formations, whereas Late
501 Triassic to Middle Jurassic claystones and sandstones emerge in the southwestern and eastern
502 part of the catchment that lie further downstream. Sampling points are indicated for the
503 Wiesent main river (numbers in circles) and tributaries (numbers in square boxes).

504
505 Figure 3. Major ion chemistry, electric conductivity (EC), and temperature (T) along the
506 Wiesent River for bi-monthly sampling campaigns. Bicarbonate was calculated from total
507 alkalinity (TA) field titration assuming that TA is comprised by carbonic species only and
508 CO₃²⁻-contributions are negligible at the observed pH values. Other major ion concentrations
509 were determined by ion chromatography.

510
511 Figure 4. Piper diagram for the Wiesent River and sampled tributaries. Numbers in
512 parentheses refer to sampling locations. The tributaries Truppach (10) and Puettlach (12)
513 mainly drain claystone and sandstone dominated lithology. These tributaries are characterized
514 by lower magnesium, increased potassium and higher sulfate concentrations. This different
515 water chemistry could also have been caused by anthropogenic inputs.

516
517 Figure 5. (A) Bicarbonate plotted against calcium plus magnesium concentrations of the
518 Wiesent River expressed in meq L⁻¹. The dashed line indicates the stoichiometric relation of

519 carbonate weathering. (B) Chloride versus sodium plus potassium expressed in meq L^{-1} . The

520 dashed line indicates the stoichiometric relationship for salt dissolution.

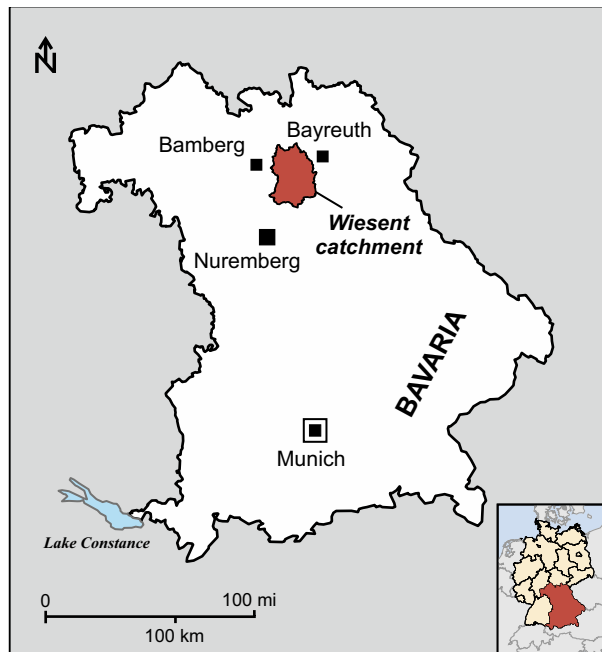


Figure 1

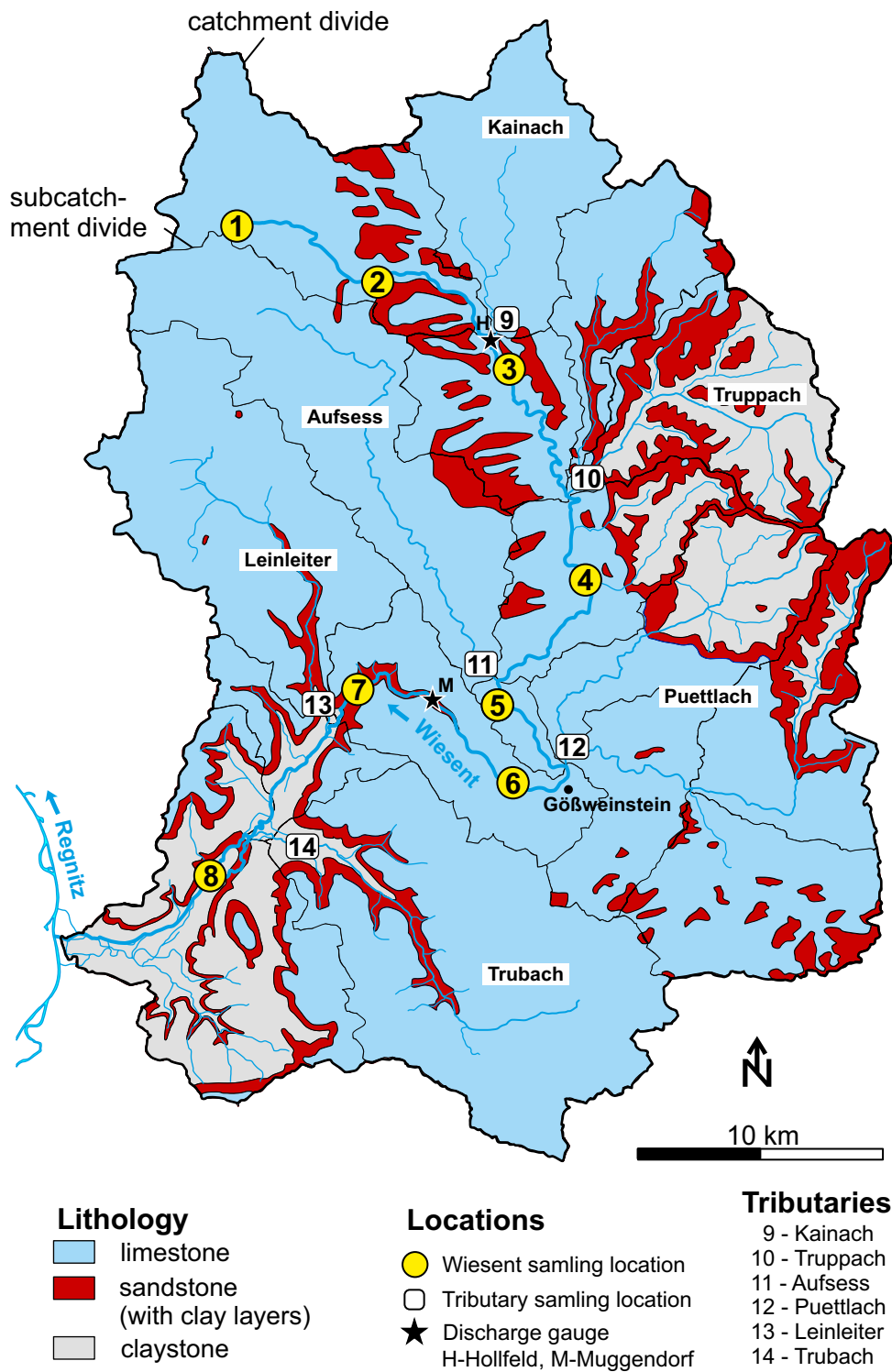


Figure 2

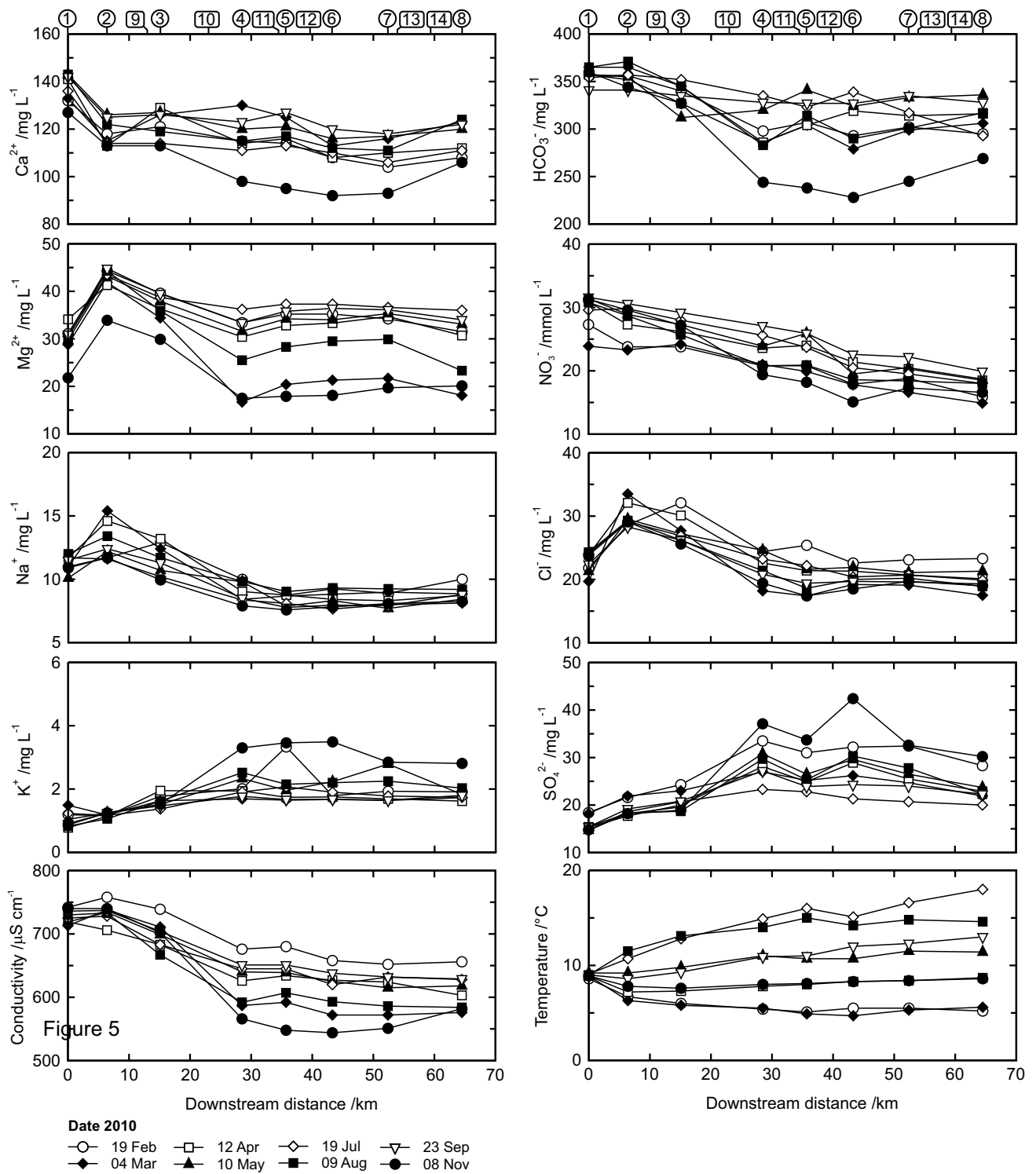


Figure 3

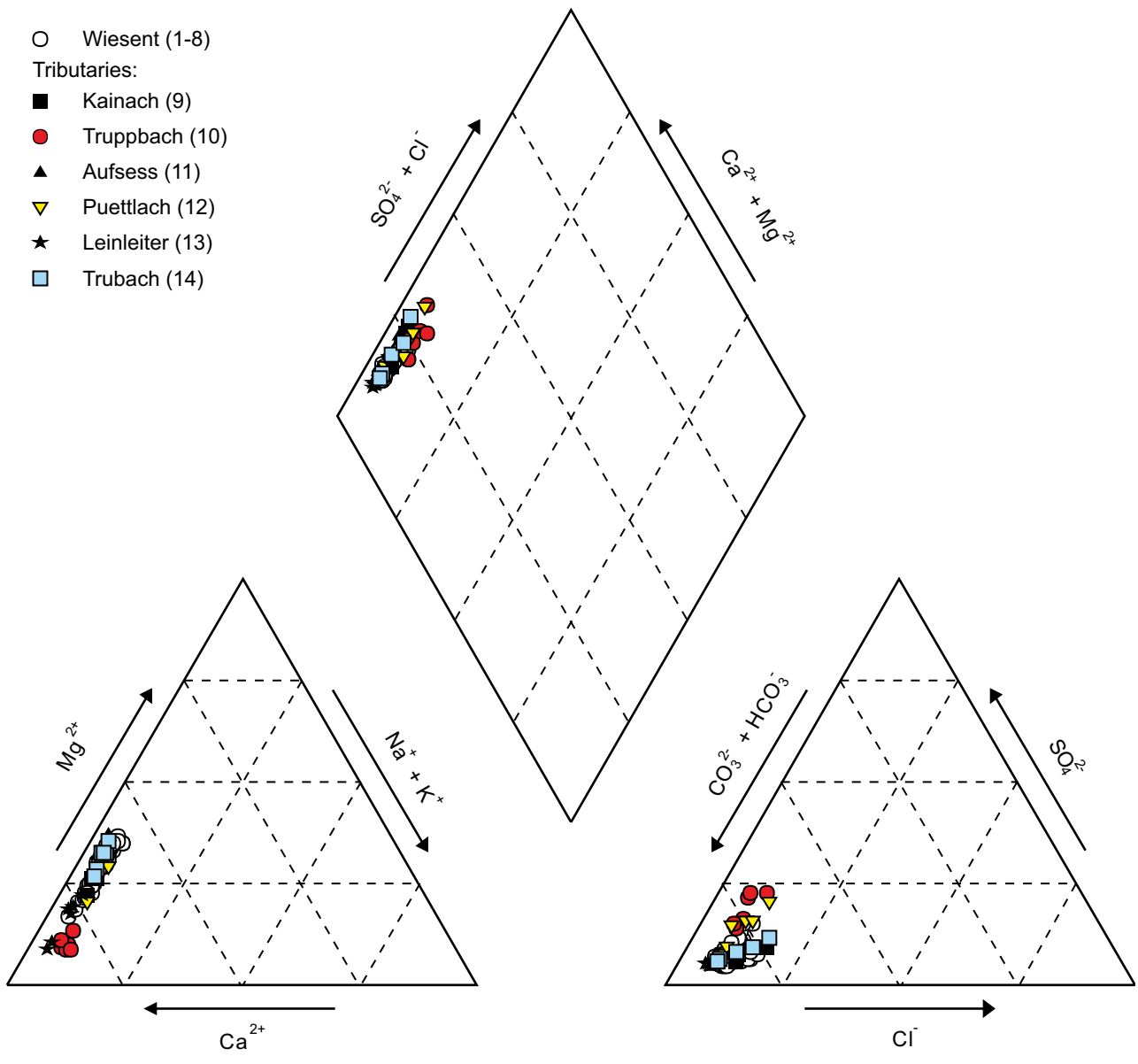


Figure 4

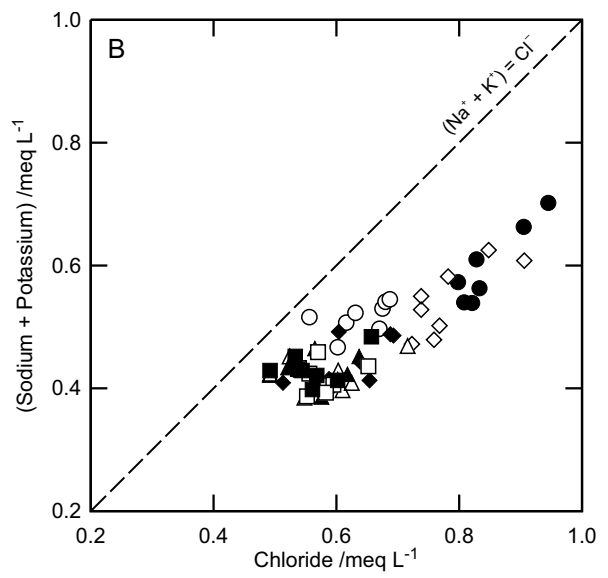
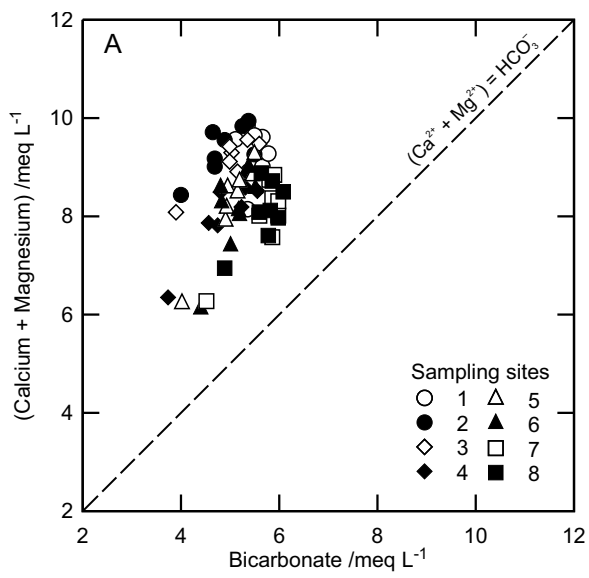


Figure 5