

Satellite-Based Clustering of Pre-Monsoon Wildfires and Variability of Tropospheric NO₂ and CO in Nepal

Dipson Bhandari¹

¹ Independent Researcher, Nepal. Email: dipsonnec@gmail.com

¹ Corresponding author: Dipson Bhandari, Email: dipsonnec@gmail.com

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Abstract

Nepal's pre-monsoon season transforms the country's southern lowlands into one of South Asia's most active wildfire landscapes. However, spatially explicit assessments of wildfire dynamics and trace gas variability at high temporal resolution remain limited. This study uses a satellite-based methodology to quantify wildfire activity and its atmospheric impacts across Nepal during 2021–2024. Visible Infrared Imaging Radiometer Suite (VIIRS) active fire observations were processed into a daily gridded dataset to derive fire count and fire radiative power (FRP) metrics. Wildfire hotspots and their spatial dynamics were identified using percentile-based thresholding combined with connected-component labeling to delineate contiguous clusters of extreme fire activity. Tropospheric nitrogen dioxide (NO₂) and total carbon monoxide (CO) were analyzed using Sentinel-5P TROPOMI products over these clusters across distinct physiographic zones from lowlands to Himalaya. Results reveal pronounced interannual variability in wildfire extent, with the largest hotspot cluster observed in 2022 (~6,718 km²) across western and mid-western Nepal.

24 Fire activity exhibits a strong seasonal peak in April, accounting for 62% of annual detections.
25 Wildfire episodes drive substantial trace gas enhancements, with NO₂ showing extreme intra-
26 monthly variability including episodic increases exceeding 1,200% relative to baseline conditions.
27 Lagged correlation analysis indicates peak associations between fire activity and trace gas
28 concentrations at 1 day for NO₂ and 2 days for CO. Persistent hotspot regions are identified in the
29 Terai and Siwalik zones, particularly in Bardiya, Banke, Dang, Surkhet, Parsa, and Bara districts.
30 Despite limited local fire activity, enhanced seasonal and episodic trace gas patterns were also
31 observed in the Himalayan region.

32 **Keywords:**

33 Wildfires, TROPOMI, VIIRS, Tropospheric NO₂, Carbon monoxide, Nepal, Fire clustering

34 **1. Introduction**

35 Forests cover a substantial portion of the Earth's land surface and play a critical role in regulating
36 climate, biogeochemical cycles and ecosystems. In 2025, the global forest area was estimated at
37 approximately 4.14 billion hectares, representing about 32% of the total land area worldwide
38 (Global Forest Resources Assessment, 2025). However, forested regions across the globe are
39 increasingly affected by wildfires. Between 2001 and 2020, the global annual burned area averaged
40 approximately 774 million hectares, accounting for nearly 6% of all burnable land (Chen et al.,
41 2023). The frequency, intensity, and spatial extent of wildfires have increased markedly in recent
42 decades, driven primarily by climate change, land-use change, and prolonged dry conditions, even
43 in regions that were previously less fire-prone (The State of the World's Forests, 2024).

44 Wildfires are a major source of atmospheric pollutants and significantly influence tropospheric
45 chemistry. Emissions from wildland fires include greenhouse gases such as carbon dioxide (CO₂),

46 methane (CH₄), and nitrous oxide (N₂O), as well as photochemically reactive species including
47 carbon monoxide (CO), non-methane volatile organic compounds (NMVOCs), and nitrogen
48 oxides (NO_x). In addition, fires release substantial amounts of fine and coarse particulate matter
49 (PM). Through both direct emissions and secondary chemical and physical processes, wildfires
50 can severely degrade air quality and contribute to regional and long-range air pollution (Urbanski
51 et al., 2008). Exposure to smoke from wildfire significantly increases the risk of premature
52 mortality and cardiorespiratory morbidity in the general population, with particularly strong
53 associations for cardiovascular deaths and asthma-related hospitalizations and emergency visits
54 (Lei et al., 2024).

55 In South Asia, forest fires occur predominantly in India, Nepal, and Bhutan, collectively burning
56 approximately 13,000–15,000 km² of forest annually, which represents about 1.7% of the region's
57 total forest area (Aditi et al., 2025). Nepal is particularly vulnerable to seasonal wildfires due to its
58 complex topography, diverse vegetation, and monsoon-dominated climate. Forest fires in Nepal
59 are largely concentrated during the pre-monsoon season (March–May), when high temperatures,
60 low relative humidity, and dry vegetation create favorable conditions for fire ignition and spread
61 (Matin et al., 2017). Approximately 89% of all forest fires in Nepal occur during this period.
62 (Bhujel et al., 2022). Long-term analyses indicate considerable interannual variability in forest fire
63 occurrences and burned areas in Nepal between 2001 and 2020. On average, about 3,098 fire
64 incidents were recorded annually, resulting in the burning of approximately 172,041 ha of forest
65 area per year. April accounted for nearly 65% of fire incidents, followed by March (17%) and May
66 (13%). Spatially, the greatest forest damage was observed in the Terai (4.74 ha km⁻²) and Siwalik
67 (4.65 ha km⁻²) regions, followed by the Mid-hills (2.09 ha km⁻²) and the High Mountain and
68 Himalayan regions (1.57 ha km⁻²). During the same period, forest fires were estimated to emit

69 approximately 3.30 million tons of carbon annually (Bhujel et al., 2022). Forest fires have become
70 a recurring hazard in Nepal, growing increasingly intense with each passing year and influenced
71 by changes in monsoonal patterns, deforestation, land-use change, forest management practices,
72 and the impacts of climate change, including reduced rainfall and drier winter conditions
73 (ICIMOD, 2025).

74 After 2016, the highest number of forest fire incidents were recorded in 2021 between March and
75 April, with more than 6,000 forest fires detected across Nepal. These fires have been linked to an
76 exceptionally dry winter season in 2020/2021, during which the country received only 25.33% of
77 the 30-year average precipitation (1981–2010) from December 2020 to February 2021 and studies
78 conducted during this period within the Kathmandu Valley reported extreme PM_{2.5} concentrations,
79 reaching maximum values of 199 $\mu\text{g m}^{-3}$ during the pre-fire period and escalating to 371 $\mu\text{g m}^{-3}$
80 and 280 $\mu\text{g m}^{-3}$ during the first and second major fire events, respectively (Khadgi et al., 2024).
81 Despite the growing severity of wildfire impacts, air quality studies in Nepal remain limited, with
82 most research focusing primarily on particulate matter and black carbon (Gyawali et al., 2025). A
83 study on Statistical modelling identified vegetation type as a major driver, with deciduous
84 broadleaf forests contributing substantially to burned areas in the Terai and Chure, and needle-leaf
85 forests dominating fire-prone zones in the Middle and High Mountains (Joshi et al., 2025a).
86 Although the relationship between forest fires and biodiversity loss in Nepal has received
87 considerable research attention (Dahal et al., 2025), the changing spatial patterns and impacts on
88 atmospheric variability remain comparatively understudied. Recent modeling efforts have further
89 highlighted the role of wildfires in deteriorating air quality in Nepal. The WRF-Chem model was
90 applied over Nepal to simulate wildfire-generated air pollutants and their transport (Shrestha et al.,
91 2025). Model outputs were evaluated using Sentinel-5P TROPOMI observations for a series of

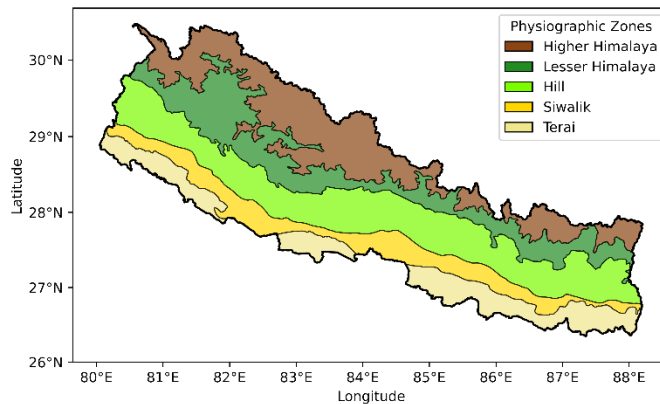
92 wildfire events between November 2020 and April 2021. Their results revealed substantial
93 wildfire-induced increases in atmospheric pollutant concentrations, including a 120% increase in
94 NO₂, followed by 48% for PM_{2.5}, 35% for PM₁₀, 32% for CO, 28% for SO₂, and 17.5% for NH₃.
95 These contributions were quantified by comparing pollutant concentrations simulated with and
96 without wildfire emissions in the WRF-Chem framework. Similar integrated observational studies
97 were identified as essential for supporting data-driven decision-making, particularly for
98 identifying wildfire hotspots and quantifying their trace gas variability. Although the studies on
99 wildfire hotspot and regional trend has been carried out, it is mostly centered towards specific
100 districts or province, trends or air pollution estimation over a certain peak episodes in major cities
101 (Joshi et al., 2025a; Karki et al., 2025) .Systematic characterization of spatial wildfire footprints
102 and quantification of tropospheric NO₂ and CO variability across multiple recent yeas remain
103 largely unaddressed .

104 Building upon these advances, the present study utilizes VIIRS data and Sentinel-5P TROPOMI
105 observations to identify wildfire footprints across Nepal and to investigate the spatiotemporal
106 variability of NO₂ and CO over the hotspots during 2021–2024. Raw swath-level VIIRS
107 observations were filtered and gridded to generate fire metrics for Nepal spanning 2021–2024.
108 Percentile-based thresholding and connected component labeling are applied to identify the spatial
109 footprints of persistent wildfire hotspot clusters, and daily TROPOMI observations are used to
110 investigate the spatiotemporal column variability of tropospheric NO₂ and total CO over these
111 hotspots during 2021–2024. The findings of this study are intended to serve a broad range of
112 applications, including forest fire risk planning and management, evidence-based policy
113 formulation at government and community levels, assessment of fire impacts on biodiversity-

114 sensitive landscapes and protected areas, and support for public health researchers estimating
115 population exposure to wildfire smoke across Nepal's most fire-affected regions.

116 **2. Methods**

117 This study covers the entire country of Nepal, a landlocked nation in South Asia, bordered by
118 China (Tibet) to the north and India to the south, east, and west. Nepal has an area of about 147,516
119 km² and lies between approximately 26°22'–30°27' N latitude and 80°40'–88°12' E longitude. The
120 country has steep range in elevation, from around 60 m above sea level in the southern plains to
121 8,848 m at the peak of Mount Everest in the north. Nepal's diverse topography and can be classified
122 into five physiographic north–south zones (Figure 1) extending from east to west: Terai (below
123 300 m), Siwalik (300–700 m), Hill (700–2,000 m), Lesser Himalaya (2,000–2,500 m), and Higher
124 Himalaya (2,500–8,848 m). The Terai (plains) region is a flat lowland zone. The Siwalik region,
125 also known as the Churia Range, has highly rugged terrain. The Hill region consists of
126 mountainous terrain with valleys and mid-elevation forests. The Himalayan region lies at higher
127 elevations and includes some of the world's highest mountains. The year is generally divided into
128 spring or pre-monsoon (March–May), summer (June–August), autumn (September–November),
129 and winter (December–January). The monsoon season lasts from June to September, bringing most
130 of the annual rainfall. For data development and extraction, this study uses a spatial domain
131 spanning 26.3°–30.5° N latitude and 80.0°–88.5° E longitude (Fig. 1), which includes Nepal and
132 some portions of southern China and northern India. However, all analyses are confined to the
133 geographical boundary of Nepal.



134

135

Fig. 1. Physiographic zones of Nepal.

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2.1. Sentinel 5P TROPOMI Products from the Sentinel-5P Product Algorithm Laboratory

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(S5P-PAL)

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TROPOMI is a nadir-viewing satellite sensor launched on October 13, 2017, aboard the Sentinel-

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5 Precursor satellite. It operates in a near-polar, sun-synchronous orbit at an altitude of 817 km,

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crossing the equator at approximately 13:30 local time in its ascending orbit and completing a full

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global coverage cycle every 17 days. With a high spatial resolution of up to $3.5 \times 5.5 \text{ km}^2$,

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TROPOMI can detect localized pollution plumes and pinpointing small-scale emission sources

143

and biomass burning events.

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The Sentinel-5P Product Algorithm Laboratory (S5P-PAL), integrated into the Copernicus

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Dataspace Ecosystem in 2025, is a cloud-based platform for developing and testing Level 2 (L2)

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algorithms across the full Sentinel-5P data record (S5P-PAL, 2025). It offers distributed processing

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tools to advance algorithms to pre-operational stage. All pre-operational and operational L2

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products are freely accessible through S5P-PAL data Portal. The platform also generates Level 3

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gridded datasets at temporal averages ranging from daily to annual, produced using HARP, a data

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harmonization toolset for scientific Earth observation data (S&T Corporation, 2024). For this

151 study, Level 3 tropospheric NO₂ and CO products were utilized, which are available at a spatial
152 resolution of 0.022° and 0.044° respectively across Nepal. Daily data of tropospheric NO₂ and total
153 CO column density for the period 2021–2024 were obtained from the SP5-PAL data portal using
154 Spatiotemporal Asset Catalog Application Programming Interface (STAC-API). The Copernicus
155 Sentinel-5P level 3 data for NO₂ and CO data are available pre-filtered for cloud contamination
156 following the guidelines of the Product, with only measurements having a quality assurance value
157 (qa_value) > 0.75 for tropospheric NO₂ and qa_value > 0.5 for Total CO column. The data were
158 extracted to the study domain shown in Figure 1, corresponding to 26.3°–30.5° N latitude and
159 80.0°–88.5° E longitude.

160 **2.2. Processing and Gridding the VIIRS (VNP14IMG) Swath Product**

161 The NASA Suomi-NPP Visible Infrared Imaging Radiometer Suite (VIIRS) Active Fire product
162 provides high-resolution observations of global fire activity at ~375 m spatial resolution and ~12-
163 hour temporal intervals. For this study, the Level 2 VIIRS 6-Minute Swath 375 m product
164 (VNP14IMG) was utilized, which uses all five I-channels (I1–I5) and the 750 m M13 channel,
165 applying a multispectral contextual algorithm based on the MODIS heritage to detect sub-pixel
166 fires and thermal anomalies using fixed and contextual tests for day and night observations. The
167 product provides geolocated fire detections along with fire radiative power (FRP), Fire Detection
168 confidence (FP_confidence), Fire pixel’s latitude and longitude, day/night flag, and other fire
169 metrics (NASA, 2018).

170 The VNP14IMG product has an 8-bit image classification product (“fire mask”) consisting of a
171 two-dimensional array with $[x, y]$ dimensions. The fire mask contains 10 classes (0–9), where class
172 8 represents nominal-confidence fire pixels and class 9 represents high-confidence fire pixels. Fire
173 pixel locations are provided as Fire Pixel Latitude (FP_latitude) and Fire Pixel Longitude

174 (FP_longitude) in the original product. VIIRS also provides the Level 3 product VNP14A1, which
 175 offers a 1-km gridded daily composite of VNP14 fire pixels detected within each grid cell.
 176 However, because this product only reports the maximum Fire Radiative Power (FRP) per grid
 177 cell, and the present study deals with total FRP counts per grid cell, the Level 2 Swath product
 178 (VNP14IMG) was used to reproduce a daily gridded product with total FRP counts and other
 179 required parameters. Daily VIIRS active fire observations were obtained from the 375-m Level-2
 180 Swath product (VNP14IMG) for the region of Nepal (latitude 26.3°–30.5° N; longitude 80.0°–
 181 88.5° E) for the years 2021-2024. All swath files were first grouped by date. Each swath was
 182 processed individually, and fire detections were filtered using fire mask values of 8 and 9 only,
 183 corresponding to nominal- and high-confidence fire pixels.

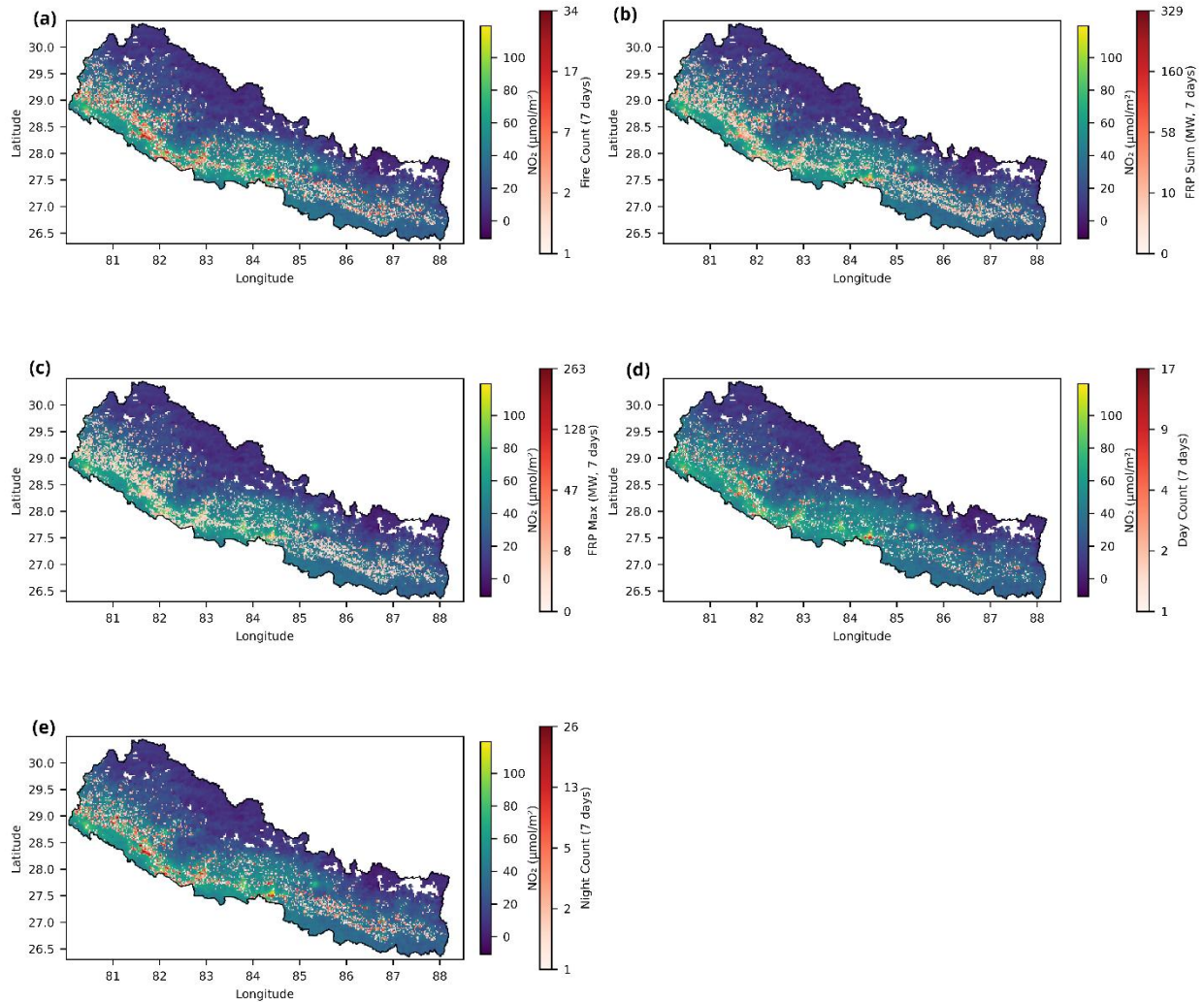
184 A binning approach was then used to grid the level 2 VIIRS data into a regular latitude-longitude
 185 grid at $\sim 0.022^\circ$ resolution. Since, the original dataset is not on a uniform geographic grid, this
 186 approach enables calculation of fire activity within each grid cell. For a given day, let $\mathcal{F} =$
 187 $\{(lat_k, lon_k, P_k, d_k)\}_{k=1}^N$ denote the set of all fire detections, where lat_k and lon_k are the latitude
 188 and longitude of fire k , P_k is its fire radiative power (FRP), and d_k is the day/night flag (1 for day,
 189 0 for night). The spatial domain is divided into a regular latitude-longitude grid with edges
 190 $\{lat_0, \dots, lat_{n_{lat}}\}$ and $\{lon_0, \dots, lon_{n_{lon}}\}$. Each fire is assigned to a grid cell (i, j) such that $lat_i \leq$
 191 $lat_k < lat_{i+1}$ and $lon_j \leq lon_k < lon_{j+1}$. For each grid cell, the total fire count is computed as the
 192 number of fires in that cell, $fire_count_{i,j} = |\mathcal{F}_{i,j}|$, while the sum and maximum of FRP are given
 193 by $frp_sum_{i,j} = \sum_{k \in \mathcal{F}_{i,j}} P_k$ and $frp_max_{i,j} = \max_{k \in \mathcal{F}_{i,j}} P_k$. The number of daytime and nighttime
 194 fires are calculated as $day_count_{i,j} = \sum_{k \in \mathcal{F}_{i,j}} \mathbf{1}_{\{d_k=1\}}$ and $night_count_{i,j} = \sum_{k \in \mathcal{F}_{i,j}} \mathbf{1}_{\{d_k=0\}}$,
 195 and the dominant detection time is the mode of the day/night flags within the cell,

196 $\text{day_night_mode}_{i,j} = \text{mode}\{d_k \mid k \in \mathcal{F}_{i,j}\}$. Grid cells with no detections are initialized to zero for
 197 counts and sums and to undefined for FRP max and day/night mode. The gridded VIIRS variables
 198 and Sentinel-5P tropospheric NO₂ and CO column densities were analyzed on a grid over Nepal.
 199 Both datasets were processed for the period 2021–2024. Table 1 provides descriptions of the
 200 variables calculated over grid cells, while (Fig. 2) shows example values of this dataset for the
 201 period 1-7 April 2021 for NO₂, and additional illustration for CO is provided in Fig. S1.

202 **Table 1.** Names and descriptions of variables in the gridded VIIRS fire products over Nepal

Data Variables	Descriptions
Fire Count (Fire_count)	The total number of fire detections in each grid cell for the given day. Each fire detection is counted once per grid cell.
FRP sum (FRP_sum)	The sum of fire radiative power (FRP) of all fires within a grid cell. FRP is a measure of fire intensity in megawatts (MW).
FRP max (FRP_max)	The maximum fire radiative power detected in a grid cell. This identifies the single most intense fire in that cell
Day Count (Day_Count)	Number of fires detected during daytime in a grid cell
Night Count (Night_Count)	Number of fires detected during nighttime in a grid cell
Latitude Bounds (Lat_bnds)	The latitude edges of each grid cell

203



204

205 **Fig. 2.** Illustration of Seven-day mean (1–7 April 2021) tropospheric NO₂ column density
 206 (μmol/m²) from TROPOMI overlaid with VIIRS-derived fire variables across Nepal: (a) fire
 207 count, (b) FRP sum, (c) FRP max, (d) daytime fire count, and (e) nighttime fire count

208 **2.3. Hotspot Algorithm**

209 **2.3.1. Hotspot Identification (Percentile-Based Thresholding)**

210 Pre-Monsoon fire hotspot regions were identified using a combination of percentile-based
211 thresholding and spatial connected component analysis applied to daily VIIRS gridded fire
212 observations. Such files containing fire count and fire radiative power (FRP) were first aggregated
213 over the full study time to produce pre-monsoon cumulative fire count and cumulative FRP fields
214 for each grid cell.

215 For each grid cell within Nepal, hotspot was defined as a location where fire activity is specifically,
216 90th percentile thresholds of annual fire count or annual fire radiative power (FRP) across all cells
217 in Nepal. A grid cell was marked as a hotspot if its annual fire count or FRP exceeded the
218 corresponding 90th percentile threshold. Cells outside Nepal were not considered. Percentile-
219 based thresholds allow identification of extreme fire activity relative to the regional background,
220 avoiding the limitations of fixed absolute thresholds. fire count and FRP were combined to ensure
221 that both frequently burning areas and regions with fewer but more intense fires are captured

222 **2.3.2 Connected Component Labeling**

223 Connected component labeling is one of the most important processes for image analysis, image
224 understanding, pattern recognition, and computer vision (He et al., 2017). Connected component
225 labelling is defined as a method where unique clusters in a image are identified based on
226 connectivity to their sides or edges(Gonzalez & Woods, 2003). In 4-connected labeling, a pixel at
227 coordinates (x, y) is considered connected only to the pixels that share an edge with it, namely the
228 pixels at positions $(x \pm 1, y)$ and $(x, y \pm 1)$. In contrast, 8-connected labeling also considers diagonal
229 neighbor pixels that share a corner, so connections can additionally occur with pixels at $(x \pm 1,$

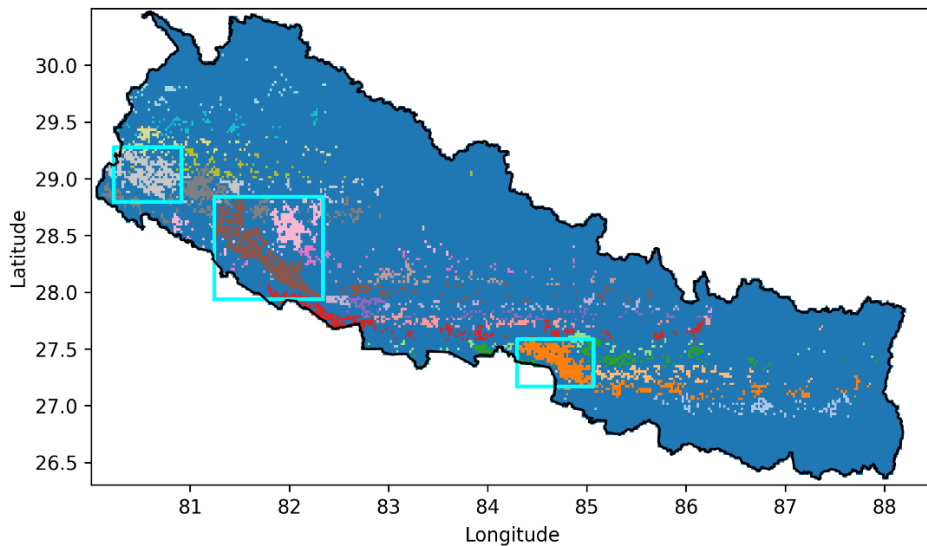
230 $y \pm 1$). Two-dimensional connected component labeling (CCL) is widely used in image analysis
231 and has been applied to group fire pixels in images (Kantzas et al., 2015; Koltunov et al., 2012).
232 Once hotspot pixels were identified pixel clustering was performed using a connected-component
233 labeling (CCL) algorithm with 4-connected labelling. This method groups neighboring hotspot
234 pixels into contiguous regions based on spatial adjacency, where pixels sharing a common edge
235 (north, south, east, or west) are considered connected. Each connected group of hotspot pixels was
236 assigned to a unique cluster identifier. The resulting hotspot clusters were ranked according to their
237 spatial extent, quantified as the total number of largest interconnected grid cells within each cluster.
238 Cluster size serves as a proxy for the areal extent of high fire activity. This study focuses on the
239 top three largest contiguous clusters, which correspond to the areas or clusters of pixels where
240 annual fire activity is extreme (top 10%). Each connected group of hotspot pixels was assigned
241 with a unique integer cluster ID. For visualization purposes, cluster IDs are mapped to a discrete
242 colormap (Fig. 3. Illustration of the top three wildfire clusters over Nepal in 2021.. The colors
243 serve solely as visual separators to distinguish spatially distinct clusters and carry no physical
244 meaning. Where the number of clusters exceeds the number of available colors, colors may repeat
245 across non-adjacent clusters.

246 The cluster size S_k corresponds to the total number of grid cells belonging to that cluster. Only the
247 hotspot cells themselves are counted; surrounding non-hotspot cells, cells from other clusters, or
248 empty space inside the bounding box are not included (Fig. 4). For example, if 100 neighboring
249 hotspot cells are connected, the cluster size is $S_k = 100$. The spatial extent of each cluster was
250 described using a latitude–longitude bounding box defined by the minimum and maximum
251 coordinates of the cluster cells. Clusters were ranked according to S_k , and the three largest clusters
252 were selected for detailed spatial analysis and visualization. The area for affected region in the

253 cluster is calculated using equation (1) where A is the area of affected regions in the cluster. R is
254 the radius of the earth, $(\lambda_2 - \lambda_1)$ is the longitudinal width of the region, φ_1 and φ_2 represents
255 latitudes.

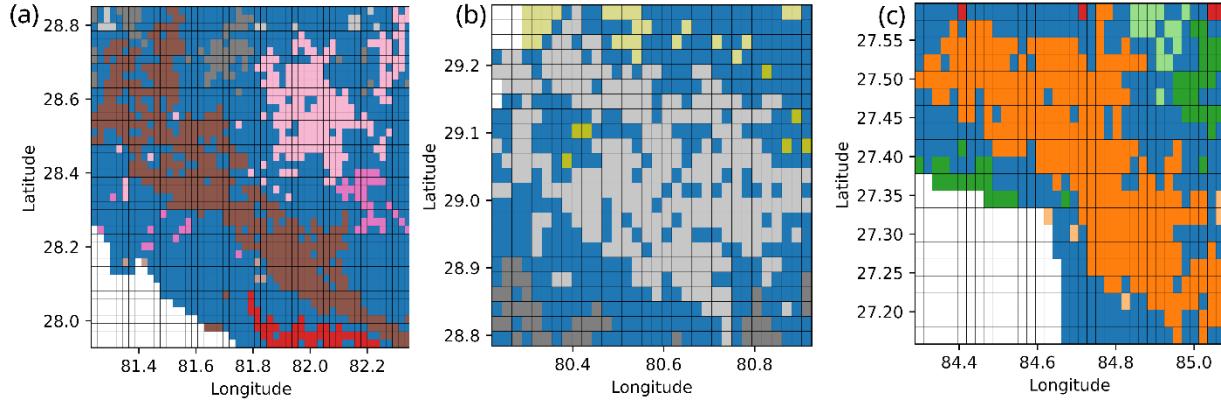
$$A = R^2 (\lambda_2 - \lambda_1) [\sin(\varphi_2) - \sin(\varphi_1)] \quad (1)$$

256 For analysis of column density variability, the entire boundary of each cluster was used to ensure
257 that all regions potentially affected by the fires were included. This approach accounts not only for
258 the main hotspot area but also for smaller internal clusters that may exist within a larger cluster,
259 which could contribute to fire emissions and impact the surrounding region.



260

261 **Fig. 3.** Illustration of the top three wildfire clusters over Nepal in 2021.



262

263 **Fig. 4.** Zoomed-in views of the top three fire clusters over Nepal in 2021 for pixel-level
 264 illustration. Panels (a), (b), and (c) correspond to the first, second, and third largest clusters.

265 **2.4. Statistical Analysis of Trace Gas Variability**

266 **2.4.1. Seasonal Enhancement and Intra-Monthly Episodic amplitude**

267 Daily cluster-averaged NO₂ and CO column densities were computed by spatially averaging all
 268 valid TROPOMI retrievals within each cluster bounding box. Monthly means and standard errors
 269 were computed from these daily values for each cluster, year, and month over the pre-monsoon
 270 period (March–May). Two metrics were used to characterize fire-driven trace gas variability. The
 271 first, seasonal enhancement (E_A), expresses the percentage increase of the monthly mean above
 272 the January–February pre-fire baseline for the same year and cluster:

273 Metric A: Seasonal Enhancement (%)

$$E_A = (\bar{X}_{month} - \bar{X}_{baseline}) / \bar{X}_{baseline} \times 100 \quad (2)$$

274 where \bar{X}_{month} is the monthly mean column density and $\bar{X}_{baseline}$ is the January–February baseline
 275 mean for the same cluster and year. January–February was selected as the background period
 276 because it precedes the fire season and exhibits minimal fire activity across all study years. The

277 second metric, intra-monthly episodic amplitude (E_B), captures the intensity of short-lived fire
278 events within a month:

279 Metric B: Intra-monthly episodic amplitude (%)

$$E_B = (X_{max} - X_{min})/X_{min} \times 100 \quad (3)$$

280 where X_{max} and X_{min} are the maximum and minimum daily cluster-averaged column density
281 within the month.

282 **2.4.2. Monthly Spearman Correlation**

283 To assess whether fire activity drives the observed trace gas enhancements, Spearman's rank
284 correlation coefficient (ρ) was computed between the monthly total fire count and column
285 densities across all cluster–year–month combinations. For each of the 3 clusters, 4 years (2021–
286 2024), and 3 fire-season months (March–May), the total VIIRS-detected fire count within the
287 cluster bounding box was aggregated, yielding a pooled dataset of up to 36 observations per
288 correlation. Fewer than 3% of days were missing from the archive across all years and both trace
289 gases, and these days were excluded from monthly mean calculations. Spearman's non-parametric
290 rank correlation was selected as fire count distributions are strongly right-skewed (many low-fire
291 months with occasional extreme events).

292 **2.4.3. Lagged Daily Spearman Correlation**

293 Since Metric B is computed once per month using only the monthly maximum and minimum
294 values, a daily NO₂ anomaly was used as a proxy for the episodic signal to check only the day-to-
295 day fluctuation components.

296

$$X'(d) = X(d) - X_{Baseline}$$

297

where $X'(d)$ is the NO₂ anomaly on day d , $X(d)$ is the daily TROPOMI-measured column density,

298

and $X_{Baseline}$ is the pre-fire-season baseline defined as the January–February mean for the

299

corresponding cluster and year.

300

Despite its improved spatial resolution, TROPOMI's ability to capture short-lived atmospheric

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gases is constrained by its limited temporal resolution, particularly in mid-latitude regions where

302

only one observation is typically available per day. This limitation can hinder accurate

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characterization of temporal variations in gas concentrations and introduce time-lag effects

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(Atalay et al., 2025). To investigate this effect, a lagged daily Spearman correlation was

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performed between daily fire counts and $X'(d + lag)$ at lags of 0 to 3 days:

306

$$\rho_{lag} = Spearman(Fire(d), X'(d + lag))$$

307

where ρ_{lag} is the Spearman rank correlation coefficient at a given lag, $Fire(d)$ is the total VIIRS-

308

detected fire count within the cluster on day d , $X'(d + lag)$ is the NO₂ anomaly observed lag days

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after the fire, and lag is the temporal offset in days (0–3) representing the delay between fire and

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satellite-detected column enhancement.

311

3. Results and Discussion

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3.1. Wildfire Trends

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Over the period 2021–2024, fire activity in Nepal showed a pronounced seasonal cycle, with the

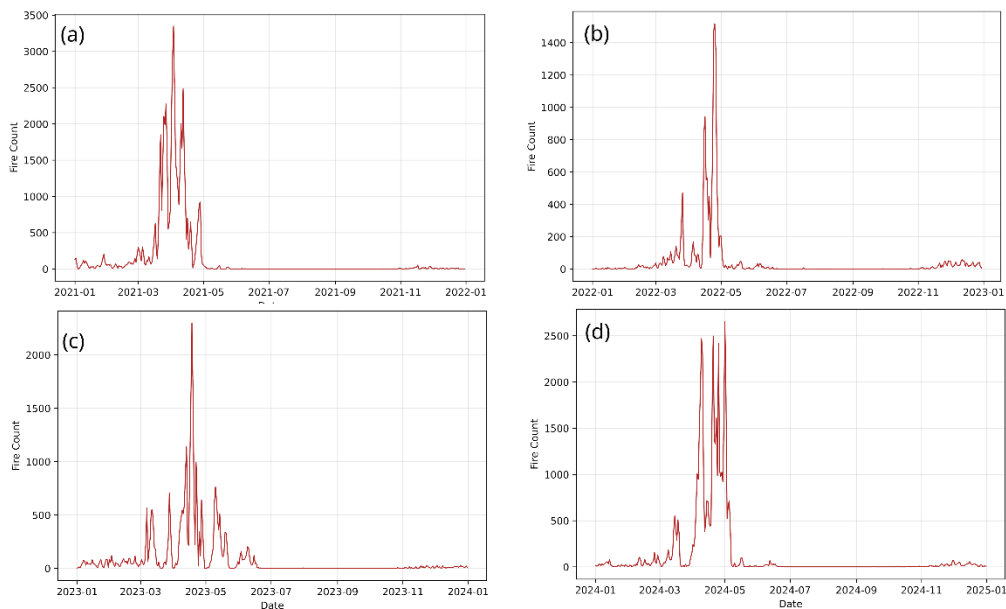
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highest occurrence between March and June, coinciding with the country's main dry season. The

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daily variation in VIIRS-detected fires from 2021 to 2024 (see Fig. 5) showed consistently low

316 fire activity during the winter and monsoon seasons, followed by a rapid increase during the pre-
317 monsoon period. This seasonal peak can be driven by vegetation, topographic, and meteorological
318 factors that create favorable conditions for forest fires (Joshi et al., 2025b). In 2021 (Fig. 5a), fire
319 activity increased sharply from early March and peaked in April before declining with the onset of
320 the monsoon. This period coincided with severe air-quality deterioration, prompting nationwide
321 school closures when Kathmandu ranked among the most polluted cities globally. A similar but
322 weaker seasonal pattern occurred in 2022 (Fig. 5b). In contrast, 2023 (Fig. 5c) exhibited a shorter
323 yet more intense fire season, characterized by sharp peaks concentrated in April, while 2024 (Fig.
324 6d) showed sustained high fire activity throughout the pre-monsoon period, which indicates a
325 longer and more severe fire season. Averaged over the four-year period, April alone accounted for
326 approximately 62% of annual fire activity, consistently appearing as the peak month, with the
327 highest contribution observed in 2022 (~70%). March and May contributed approximately 19%
328 and 9%, respectively. These results show the dominant role of the pre-monsoon season in shaping
329 fire activity across Nepal.

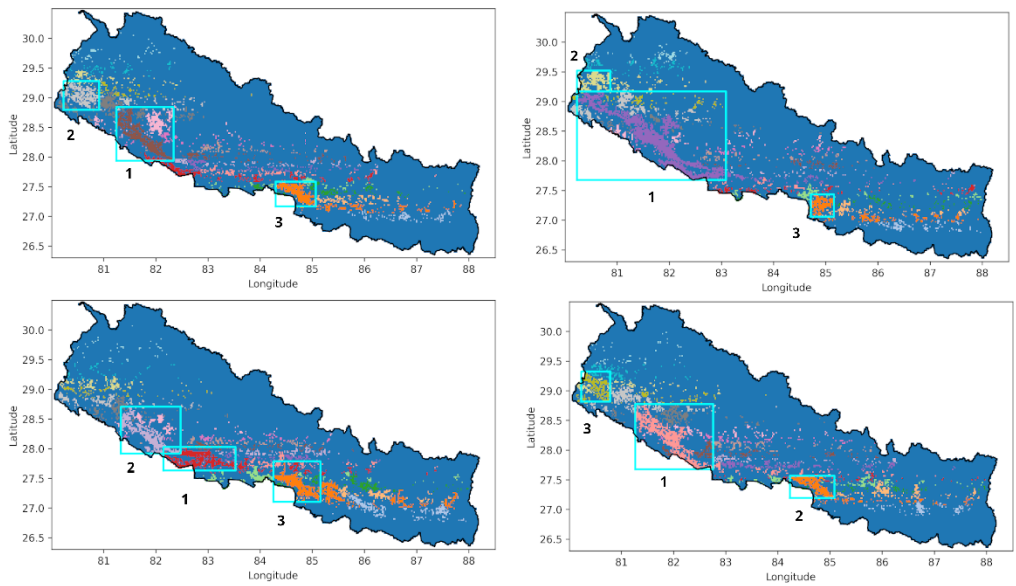


331 **Fig. 5.** Daily VIIRS fire counts over Nepal for the full year for (a) 2021, (b) 2022, (c) 2023, and
332 (d) 2024.

333 **3.2. Wildfire Hotspot and Affected Regions 2021-2024**

334 Strong interannual and spatial variability in wildfire hotspots was observed across Nepal, with
335 most fire activity concentrated in the western, mid-western, and central regions of the southern
336 Terai belt (Fig. 6). In the southern plains, even slight slopes enhance drainage, leading to drier
337 conditions and increasing the susceptibility of vegetation to fire, while prolonged precipitation
338 deficits also contribute to extreme fire conditions in this region (Joshi et al., 2025b). In 2021, The
339 largest cluster consisted of 425 connected hotspot grid cells, covering $\sim 2,232$ km², mainly in the
340 mid-western districts of Bardiya, Banke, Dang, Surkhet, and Salyan. Smaller clusters were
341 observed in the far-western districts (Dadeldhura, Doti, Kanchanpur, Kailali) and in the central–
342 southern districts (Chitwan, Parsa, Bara), indicating localized wildfire hotspots (see Table 2). The
343 year 2022 recorded the highest wildfire spatial extent. The largest cluster size of 1,279 connected
344 grid cells fire and covered approximately 6,718 km², spreading across western and mid-western
345 districts such as Kanchanpur, Kailali, Bardiya, Surkhet, Banke, Dang, and Arghakhanchi. This
346 suggests widespread and intense wildfire conditions during that year. Smaller clusters were found
347 in Dadeldhura and Baitadi, and in the southern plains (Parsa and Bara), which indicates that some
348 areas remained repeatedly affected even during extreme years. In 2023, wildfire activity decreased
349 compared to 2022 with cluster size of 377 over about 1,990 km², mainly in Dang, Arghakhanchi,
350 Kapilbastu, Rupandehi, and Palpa. Another large cluster (375 connected hotspot grid; 1,972 km²)
351 was concentrated in Bardiya, Banke, Dang, and Surkhet, while Chitwan, Parsa, and Bara continued
352 to appear as persistent hotspots. In 2024, wildfire activity increased again but did not reach the
353 2022 level. The largest cluster had 598 fire events covering around 3,146 km², located mainly in

354 Bardiya, Banke, Dang, Surkhet, and Salyan. Overall, Bardiya, Banke, Dang, Surkhet, Parsa, and
355 Bara appeared repeatedly as wildfire-prone districts across all four years (Fig. 6), indicating areas
356 of consistent high risk. Substantial fire activity was also observed in northern India, particularly in
357 regions adjacent to Nepal's southern border, which could influence atmospheric composition
358 across the broader region. Indian cities and districts in these border areas, including Gorakhpur,
359 Muzaffarnagar, and Varanasi in Uttar Pradesh, have been independently identified as wildfire
360 hotspot zones during the pre-monsoon season (Majumdar, 2025).



361

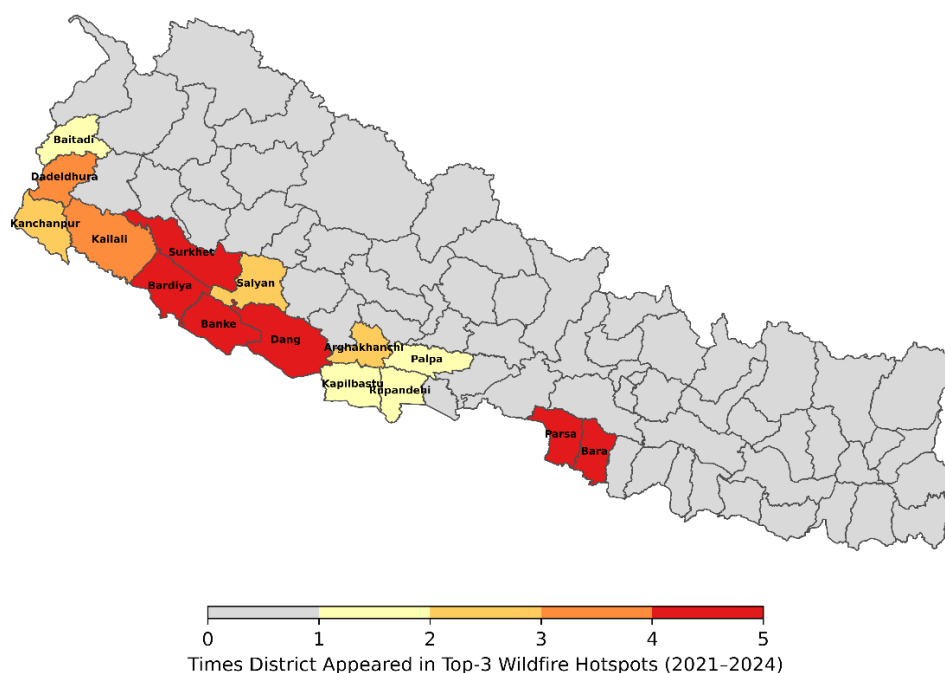
362 **Fig. 6.** Spatial extent of the three largest contiguous wildfire hotspot clusters across Nepal for
363 each year from 2021 to 2024: (a) 2021, (b) 2022, (c) 2023, and (d) 2024.

364 **Table 2.** Cluster rank, number of connected hotspot grid cells, spatial extent (km²), and districts
365 covered by each cluster.

Year Rank	Cluster Size (Sk)	Area (km²)	Districts	Region
2021 1	425	2,232.20	Bardiya, Surkhet, Salyan, Banke, Dang	Mid-Western Terai
2	260	1,356.85	Dadeldhura, Doti, Kanchanpur, Kailali	Far-Western Terai
3	254	1,346.16	Chitwan, Parsa, Bara	Central Terai
2022 1	1,279	6,718.14	Kanchanpur, Kailali, Bardiya, Surkhet, Banke, Dang, Arghakhanchi	Far-Western to Mid-Western Terai
2	202	1,051.17	Dadeldhura, Baitadi	Far-Western Hills
3	166	880.85	Parsa, Bara	Central Terai
2023 1	377	1,990.47	Dang, Arghakhanchi, Kapilbastu, Rupandehi, Palpa	Mid-Western Terai–Hill Belt
2	375	1,971.90	Bardiya, Banke, Dang, Surkhet	Mid-Western Terai
3	295	1,563.41	Chitwan, Parsa, Bara	Central Terai
2024 1	598	3,145.65	Bardiya, Surkhet, Banke, Salyan, Dang	Mid-Western Terai

Year Rank	Cluster Size (Sk)	Area (km ²)	Districts	Region
2	230	1,218.84	Chitwan, Parsa, Bara, Makwanpur	Central Terai
3	189	986.15	Dadeldhura, Kailali, Doti	Far-Western Terai

366



367

368 **Fig. 7.** District map of Nepal showing districts at risk based on the number of times they
369 appeared in the Top 3 wildfire hotspot footprint during 2021–2024.

370 3.3. Tropospheric NO₂ variability

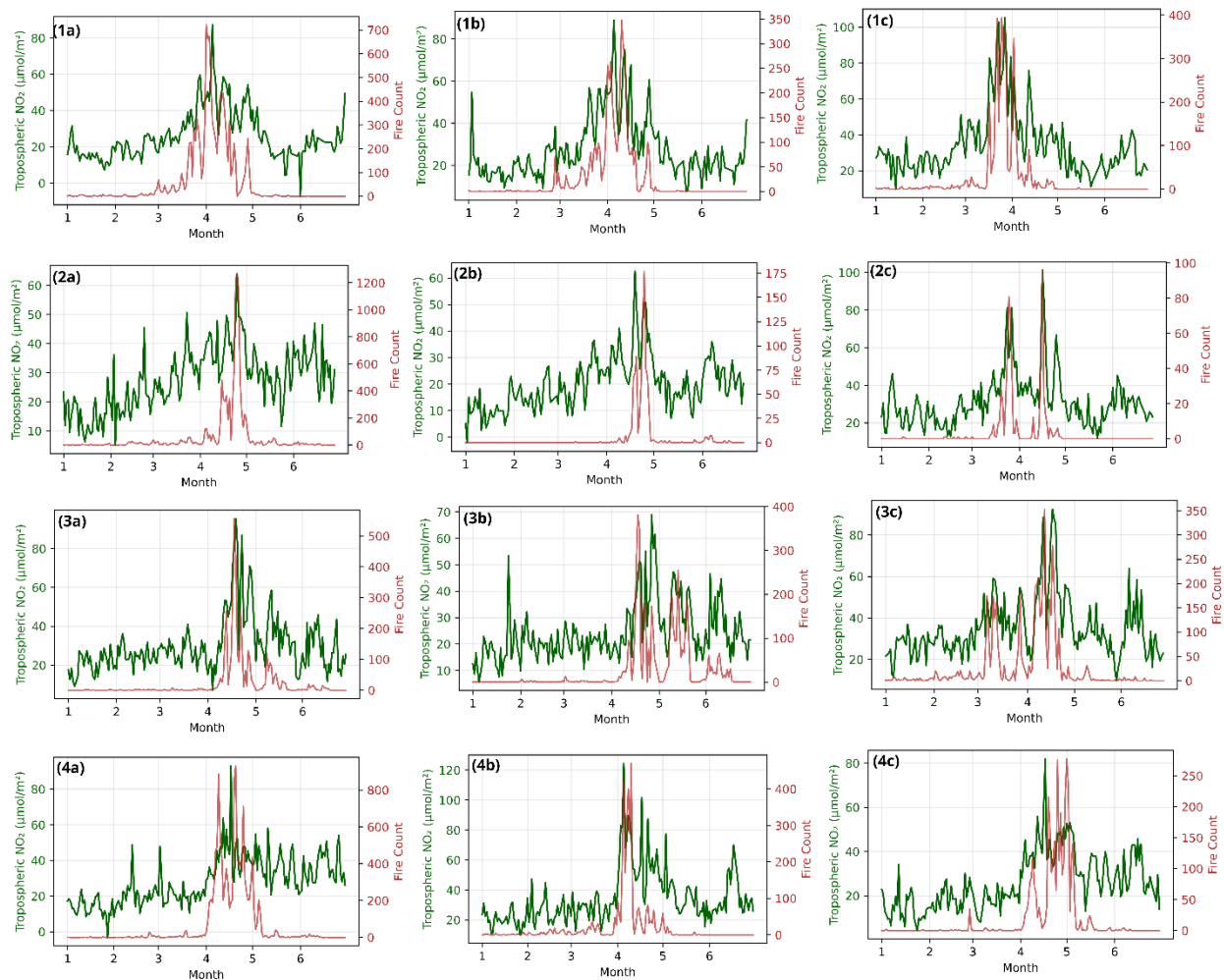
371 Tropospheric NO₂ column densities over Nepal during 2021–2024 showed a strong and consistent
372 pre-monsoon signal, with April consistently emerging as the peak month for seasonal enhancement
373 across all clusters and years (see Fig. 8). Mean seasonal enhancements in April were 130% higher

374 than the January–February baseline during 2021–2024. This finding is consistent with, and slightly
375 exceeds, the 120% wildfire-attributable NO₂ increase reported by Shrestha et al. (2025) for Nepal
376 using WRF-Chem simulations. However, Shrestha et al.(2025) derived their estimate by
377 comparing model simulations with and without fire emissions. The highest values was recorded in
378 2024, Cluster 2 which includes central-southern terai belt (Chitwan parsa and Bara) reached
379 +177.6% (mean: $62.92 \pm 3.81 \mu\text{mol m}^{-2}$) with a single-day maximum of $124.46 \mu\text{mol m}^{-2}$ on April,
380 the highest cluster-averaged daily value of the study period.

381 Intra-monthly episodic amplitudes were substantially larger, with the most extreme event recorded
382 in April 2023 in Cluster 1 (~1,250%), where the daily maximum of $95.34 \mu\text{mol m}^{-2}$ in April
383 exceeded the monthly minimum by approximately 13.5-fold, indicating an intense short-lived fire
384 plume. This includes western Terai and adjoining hill districts (Kapilbastu, Rupandehi, Dang,
385 Arghakhanchi, and Palpa. The western Terai corridor , particularly the Bardiya–Banke–Dang–
386 Surkhet belt, appears as a persistent and dominant NO₂ hotspot, as the primary cluster in three out
387 of four years. The central Terai (Chitwan–Parsa–Bara) consistently forms a secondary cluster and
388 recorded the highest baseline NO₂ values among all clusters. This central Terai region contains
389 extensive urban and industrial areas and has been identified as an NO₂ hotspot in previous studies,
390 contributed by the high concentration of cement and brick factories (Gyawali et al., 2025) , which
391 can drive its persistently elevated baseline values. The region also reached the highest monthly
392 mean NO₂ column density of all clusters in April, with a value of $62.9 \pm 3.8 \mu\text{mol m}^{-2}$. The highest
393 daily maximum NO₂ column density was recorded in April 2024, reaching $124.5 \mu\text{mol m}^{-2}$

394 Spearman rank correlation between monthly VIIRS fire counts and Metric A (seasonal
395 enhancement) yielded a strong and statistically significant association ($\rho = 0.820$, $p < 0.0001$),
396 which indicates that the months with higher fire activity consistently corresponded to greater NO₂

397 enrichment above the pre-fire baseline across all clusters and years (see Fig.S2). A lagged daily
 398 Spearman analysis was conducted for lags of 0 to 3 days as mentioned in section 2.4.3. Daily fire
 399 counts were significantly correlated with same-day NO₂ at lag-0 (mean $\rho = 0.661$, $p < 0.0001$),
 400 with a marginal but consistent improvement at lag-1 (mean $\rho = 0.678$, $p < 0.0001$), followed by
 401 decline at lags 2 and 3 with $\rho = 0.658$ and 0.616 , respectively (Fig.S2, Fig.S3).



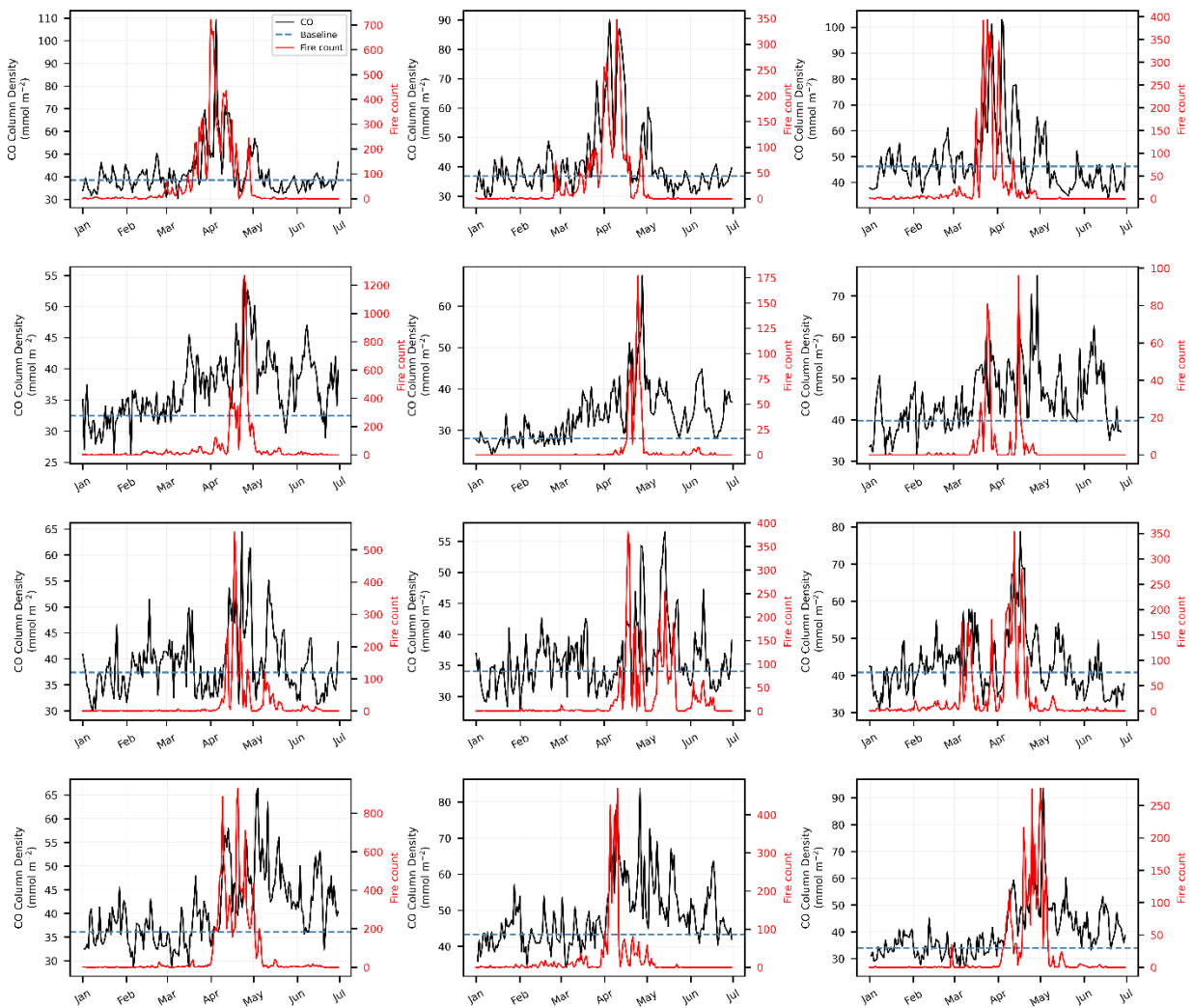
402
 403 **Fig. 8.** Daily tropospheric NO₂ ($\mu\text{mol}/\text{m}^2$) and fire counts for the three largest wildfire clusters in
 404 Nepal (January–June, 2021–2024). Panels show each year (rows 1–4) and cluster rank (columns
 405 a–c), where column 'a' is the largest cluster each year.

406 3.4. Total CO Column Variability over Hotspots

407 Total CO column concentrations over Nepal's Terai clusters during the pre-monsoon fire season
408 (March–May, 2021–2024) showed consistent seasonal enhancement above the January–February
409 baseline, with April dominating as the peak month across most clusters and years (Fig. 9). Cluster-
410 averaged April enhancements reached approximately 39% in 2021, 35% in 2022, 19% in 2023,
411 and 35% in 2024. The western Terai corridor (Bardiya–Surkhet–Banke–Salyan–Dang)
412 consistently recorded the most intense episodic CO variability, with the most extreme event
413 occurring in April 2021 when the daily maximum reached $109.4 \text{ mmol m}^{-2}$ on April which is
414 approximately 3.3 times the within-month minimum and showed an episodic amplitude of $\sim 230\%$
415 .This indicates a concentrated, high-intensity regional-scale fire episode. The central-southern
416 Terai corridor (Chitwan–Parsa–Bara) consistently maintained the highest baseline CO
417 concentrations across all years (up to 46.2 mmol m^{-2} in 2021) and repeatedly exhibited strong April
418 episodic amplitudes peaking at $\sim 189\%$ in April 2021 and $\sim 130\%$ in April 2023. This can be
419 attributed to the region's accumulative nature, compounded by the year-round anthropogenic
420 emissions and high concentration of mines, brick kilns, and cement industries in this region
421 (Gyawali et al., 2025; Joshi et al., 2025b). As this region is located near densely populated urban
422 areas, targeted policy interventions and strengthened wildfire management during the pre-
423 monsoon season are urgently needed.

424 Monthly Spearman correlation between VIIRS fire counts and CO seasonal enhancement (Metric
425 A) was statistically significant (Fig.S4) but moderate ($\rho = 0.532$, $p < 0.001$), weaker than the
426 corresponding NO_2 result ($\rho = 0.820$, $p < 0.0001$), and the lagged daily Spearman analysis (Fig.S5)
427 indicated a progressive strengthening of the fire–CO relationship from lag-0 (mean $\rho = 0.457$, $p =$
428 0.0163) to a peak at lag-2 (mean $\rho = 0.566$, $p = 0.0001$), with consistently stronger correlations

429 observed at lag 1–2 days across all four individual years .CO seasonal enhancements were
430 substantially more moderate than NO₂ across all clusters and years (CO Metric A range: ~15–53%
431 vs NO₂ range: ~83–178%), and this weaker seasonal signal alongside the broader 2-day lag in
432 contrast to the sharper 1-day peak observed for NO₂ (lag-1: $\rho = 0.678$, $p < 0.0001$). This is
433 physically consistent with CO's longer atmospheric lifetime of several weeks, which maintains
434 elevated background concentrations throughout the fire season, compresses the baseline-to-peak
435 ratio, and allows fire-emitted CO to accumulate gradually in the tropospheric column over multiple
436 days before reaching peak detectable concentrations at the TROPOMI overpass.

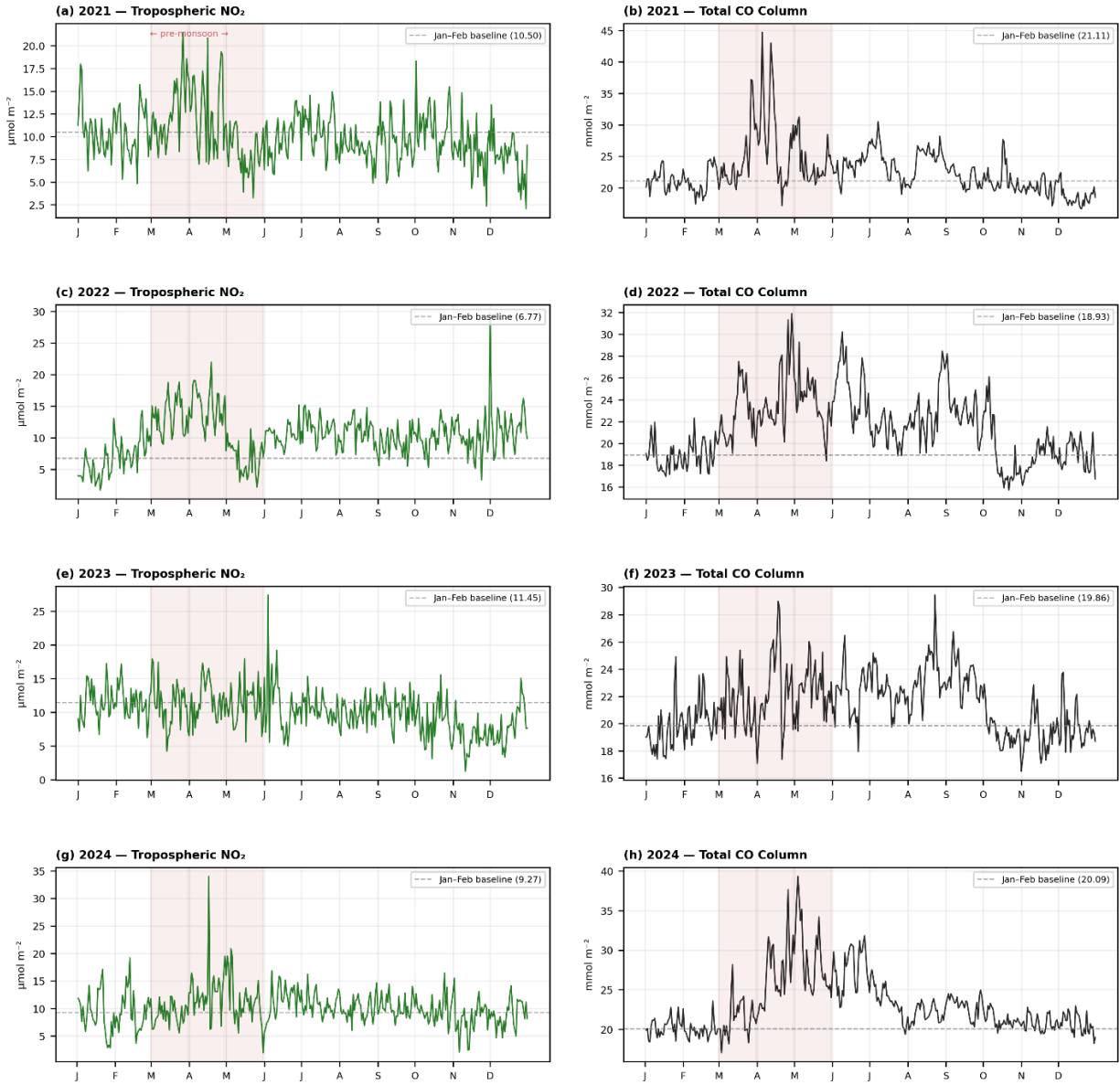


438 **Fig. 9.** Daily total column CO (mmol/m²) and fire counts for the three largest wildfire clusters in
439 Nepal (January–June, 2021–2024). Panels show each year (rows 1–4) and cluster rank (columns
440 a–c), where column 'a' is the largest cluster each year.

441 **3.5. Trace Gas Variability in the Himalayan Region**

442 Despite the Higher and Lesser Himalayan zones combined contributing approximately 7% of total
443 VIIRS-detected fire activity during the pre-monsoon season, clear enhancements in both
444 tropospheric NO₂ and total CO were observed during March–May across all four years (Fig. 10).
445 Fire-season NO₂ enhancement over the Himalayan zone showed strong interannual variability,
446 with the most extreme enhancements recorded in 2022 (March: +96.4%; April: +121.9%),
447 coinciding with the anomalously large and spatially merged western Terai cluster spanning nearly
448 6,718 km² in that year (Fig. 11). Episodic NO₂ amplitudes over the Himalayan zone were large,
449 with the most extreme events recorded in May 2024 (960.3%) and May 2022 (419.6%). These
450 were months when the seasonal mean enhancement was near zero, which suggest that the
451 individual fire plume transport events drive extreme single-day concentration spikes over the
452 Himalayan zone rather than producing sustained baseline elevation. CO enhancements over the
453 Himalayan zone were more moderate and temporally sustained than NO₂, with the highest values
454 recorded in 2024 (April: +34.3%; May: +44.1%), making 2024 the most CO-impacted fire season
455 over the Himalaya in the study period, while 2023 remained the least affected year across both
456 gases. CO episodic amplitudes over the Himalayan zone (34–160%) were consistently and
457 substantially lower than the corresponding NO₂ amplitudes (116–960%) and maintained elevated
458 background concentrations in the Himalaya region.

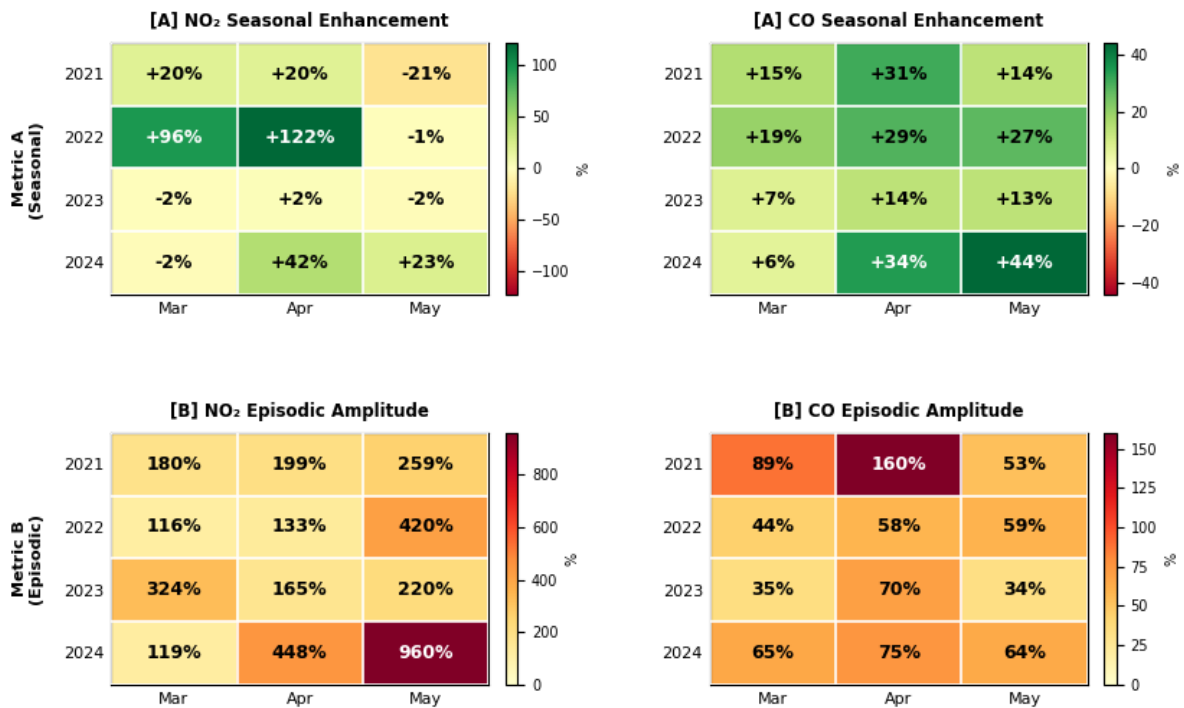
459 The entire territory of Nepal lies within the Hindu Kush Himalaya region, a transboundary
460 mountain system encompassing partial or complete territories of Afghanistan, Bangladesh,
461 Bhutan, China, India, Myanmar, and Pakistan. Nepal's Himalayan zones are geographically
462 proximate to the Indian and Bhutanese portions of this system, both of which are known to have
463 high forest fire susceptibility (Ghale et al., 2026). Tracing the actual transport trajectories of these
464 pollutants would require dedicated atmospheric transport modeling, which represents a potential
465 avenue for future research to quantify the relative contributions of individual fire cluster events to
466 trace gas enhancement in the Himalayan zone. In this study, a statistical approach is employed to
467 provide a foundational understanding of this relationship. A pooled same-day Spearman rank
468 correlation computed from the averaged NO₂ anomaly across all three Terai fire clusters and entire
469 Himalayan zone NO₂ anomaly on the same day (Fig.S6), across all fire season days (March–May)
470 of 2021–2024 combined, yielded $\rho = 0.585$ ($p < 0.0001$, $n = 361$). When computed separately for
471 each year (Fig.S7), the correlation was significant in all four individual years: 2021 ($\rho = 0.788$),
472 2022 ($\rho = 0.740$), 2023 ($\rho = 0.401$), and 2024 ($\rho = 0.623$), which indicates that the linkage between
473 the Terai lowlands and the Himalayan zone pre-monsoon enhancement is a consistent feature of
474 Nepal. This finding is consistent with the preliminary assessment of transboundary transport of
475 fire-induced air pollution (PM_{2.5}) from the lowland plains into the Himalayan region, which has
476 been identified as an urgent regional air quality concern requiring integrated and transboundary
477 approach to address air pollution episodes in Himalayan region (Mahapatra & Dhital, 2025).



478

479 **Fig. 10.** Temporal variability of tropospheric NO₂ and CO column densities averaged over the

480 Lower and Higher Himalaya zones of Nepal.



481

482 **Fig. 11.** Seasonal and Episodic Variability of Tropospheric NO₂ and CO over the Himalayan
 483 Zone (2021–2024)

484 **4. Conclusion**

485 The study integrates VIIRS and Sentinel-5P TROPOMI observations to characterize the
 486 spatiotemporal dynamics of pre-monsoon wildfires and their atmospheric impacts across Nepal
 487 during 2021–2024. By combining percentile-based hotspot detection with connected-component
 488 labeling and daily cluster-averaged trace gas analysis, the study provides a systematic and
 489 quantitative assessment of wildfire footprints and associated NO₂ and CO variability at high
 490 temporal resolution over Nepal.

491 Wildfire activity exhibited a consistent and pronounced seasonal cycle across all four years, with
 492 April accounting for approximately 62% of annual fire detections. Strong interannual variability

493 was observed in both the spatial extent and intensity of wildfire hotspot clusters, with 2022
494 recording the largest contiguous cluster (~6,718 km²) spanning western and mid-western Nepal,
495 while 2024 exhibited the most sustained fire season. Six districts -Bardiya, Banke, Dang, Surkhet,
496 Parsa, and Bara appeared persistently in the top wildfire hotspot clusters across all four years,
497 identifying these areas as regions of chronic and recurring fire risk that warrant prioritized
498 management attention.

499 Tropospheric NO₂ showed the most dramatic fire-season response, with mean April seasonal
500 enhancements of approximately 130% above the January–February pre-fire baseline, and intra-
501 monthly episodic amplitudes reaching up to ~1,250% in April 2023. These extreme episodic
502 spikes, driven by short-lived but intense fire plumes, highlight the acute air quality hazard posed
503 by Nepal's pre-monsoon wildfires, particularly for populations in the Terai lowlands. Total CO
504 column enhancements were more moderate (15–53% seasonal enhancement) and temporally
505 broader, consistent with CO's longer atmospheric lifetime, which maintains elevated background
506 concentrations throughout the fire season and produces a more gradual rather than episodic
507 atmospheric loading. Lagged correlation analysis confirmed that NO₂ responds most strongly to
508 fire activity at a 1-day lag, while CO peaks at a 2-day lag, findings that are physically consistent
509 with the differing atmospheric lifetimes. A statistically significant same-day correlation was
510 observed between lowland Terai fire cluster activity and trace gas anomalies over the Himalayan
511 zone

512 Several limitations of this study should be acknowledged. The observed trace gas enhancements
513 reflect contributions from all emission sources active during the fire season, including
514 anthropogenic agricultural crop residue burning, urban and industrial activity, and transboundary
515 pollution from northern India, and therefore cannot be fully attributed to forest fires alone without

516 dedicated atmospheric chemical transport modeling. Future research should employ models such
517 as WRF-Chem or HYSPLIT to isolate fire-specific emission contributions and quantify long-term
518 trends in transboundary pollutant transport across the Himalayan region during the pre-monsoon
519 season. Additionally, further studies could also assess population exposure to air pollution in
520 persistent wildfire hotspot districts and examine the situation of land-use change and forest
521 management practices in these areas.

522 Given the recurring nature of pre-monsoon wildfires over recent decades, it is likely that this
523 phenomenon will persist and potentially intensify under the influence of climate change. While
524 the Government of Nepal, in collaboration with research institutions, established the Forest Fire
525 Management System in 2021 to report fire incidents, governmental responses to wildfire dynamics
526 and their atmospheric impacts remain limited and have received insufficient policy attention
527 relative to their effects on biodiversity, atmospheric composition, and public health. Stronger
528 coordination among government agencies, policymakers, and research institutions is therefore
529 essential in a country like Nepal to formulate effective policies, strengthen enforcement
530 mechanisms, and address the growing challenges of wildfire activity and deteriorating air quality.

531 **Author contributions: CRediT**

532 **Dipson Bhandari:** Conceptualization, Methodology, Software, Validation, Formal Analysis,
533 Investigation, Data Curation, Writing-Original Draft, Writing-Review and Editing.

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538 **process**

539 During the preparation of this work the author used ChatGPT to assist with grammar correction,
540 sentence restructuring, and language editing. After using this tool, the authors reviewed and edited
541 the content as needed and take full responsibility for the content of the published article.

542 **Data Availability**

543 Data will be made available on request.

544 **References:**

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Supplementary Information

Fig. S1. Illustration of Seven-day mean (1–7 April 2021) total CO column density ($\mu\text{mol}/\text{m}^2$) from TROPOMI overlaid with VIIRS-derived fire variables across Nepal: (a) fire count, (b) FRP sum, (c) FRP max, (d) daytime fire count, and (e) nighttime fire count

Fig. S2. Spearman correlation between monthly fire count and seasonal NO_2 enhancement (left), and mean lagged daily correlation at 0–3 day lags (right).

Fig. S3. Spearman correlation between daily fire count and NO_2 column density at lags 0–2 days for each year (2021–2024) and cluster (C1–C3) during March–May.

Fig. S4. Spearman correlation between monthly fire count and seasonal CO enhancement (left), and mean lagged daily correlation at 0–3 day lags (right).

Fig. S5. Spearman correlation between daily fire count and CO column density at lags 0–2 days for each year (2021–2024) and cluster (C1–C3) during March–May.

Fig. S6. Same-day Spearman correlation between the averaged NO_2 anomaly across three Terai fire clusters and the Himalayan zone NO_2 anomaly during March–May 2021–2024 .

Fig. S7. Year-by-year Spearman correlation between averaged Terai and Himalayan zone NO_2 anomalies during March–May; all years significant ($p < 0.05$). Dashed line indicates the pooled correlation across all four years combined ($\rho = 0.585$).

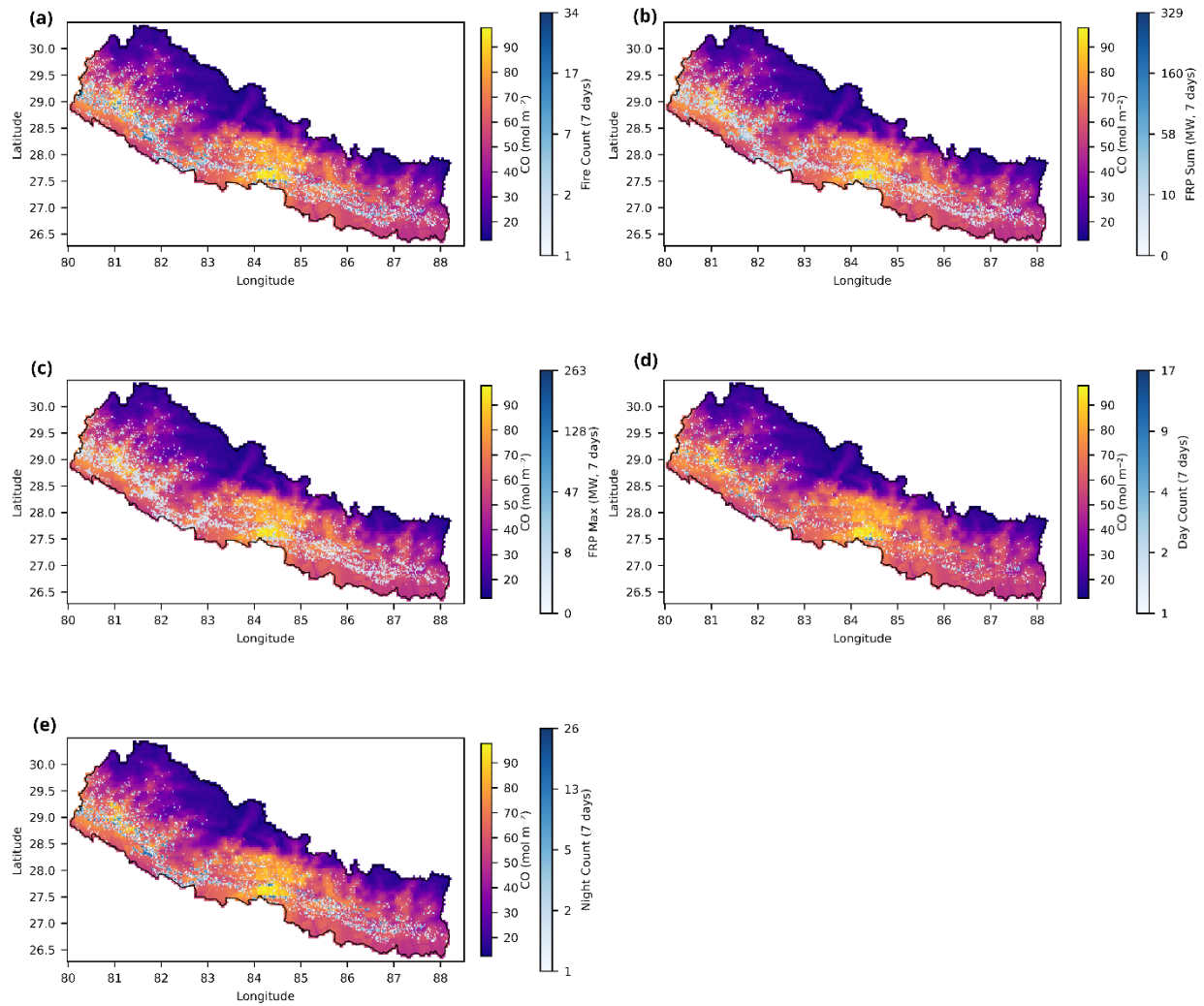


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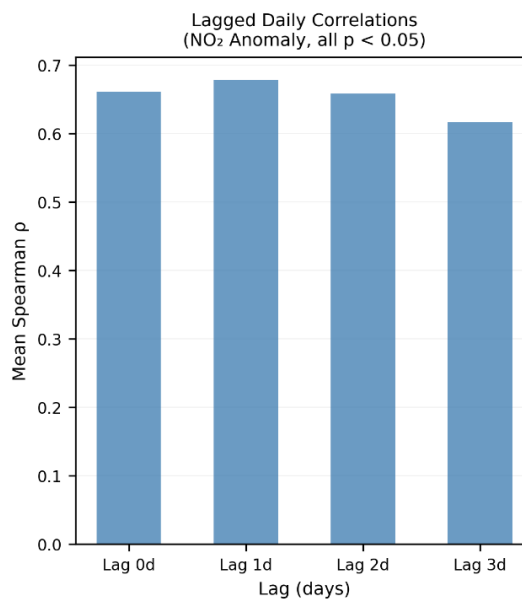
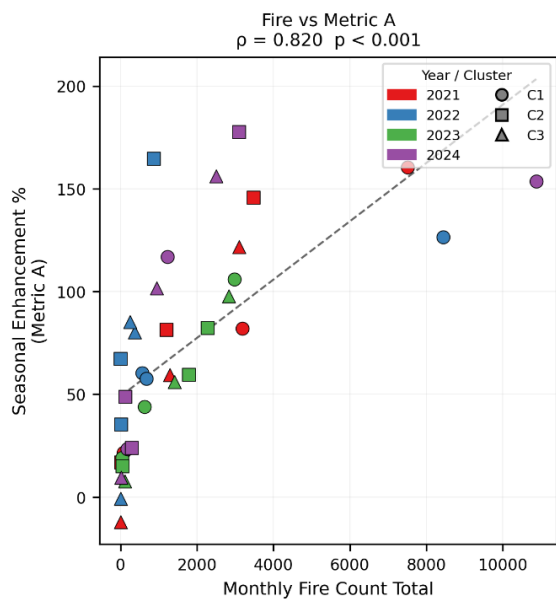


Fig. S2. Spearman correlation between monthly fire count and seasonal NO₂ enhancement (left), and mean lagged daily correlation at 0–3 day lags (right).

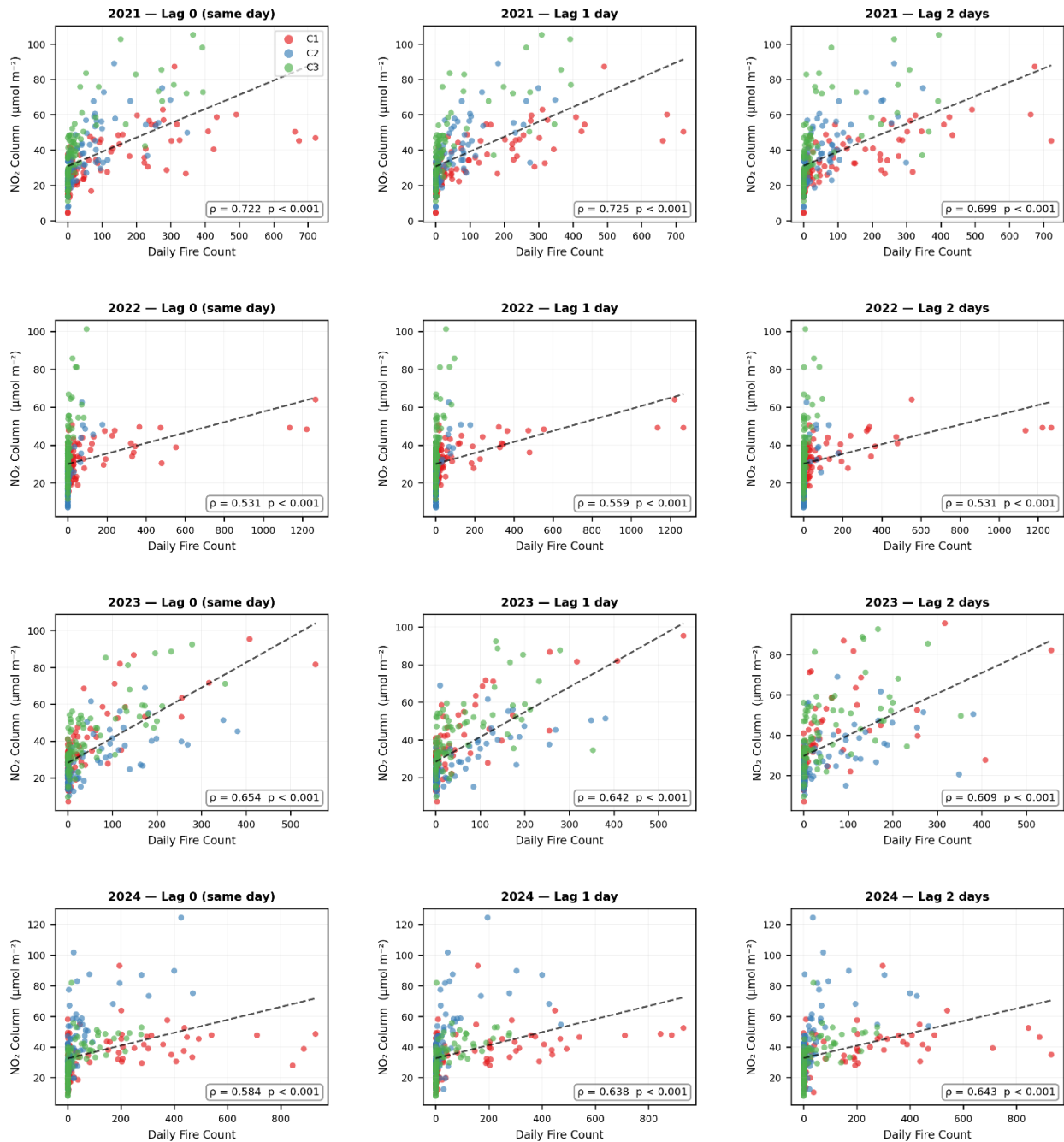


Fig. S3. Spearman correlation between daily fire count and NO₂ column density at lags 0–2 days for each year (2021–2024) and cluster (C1–C3) during March–May.

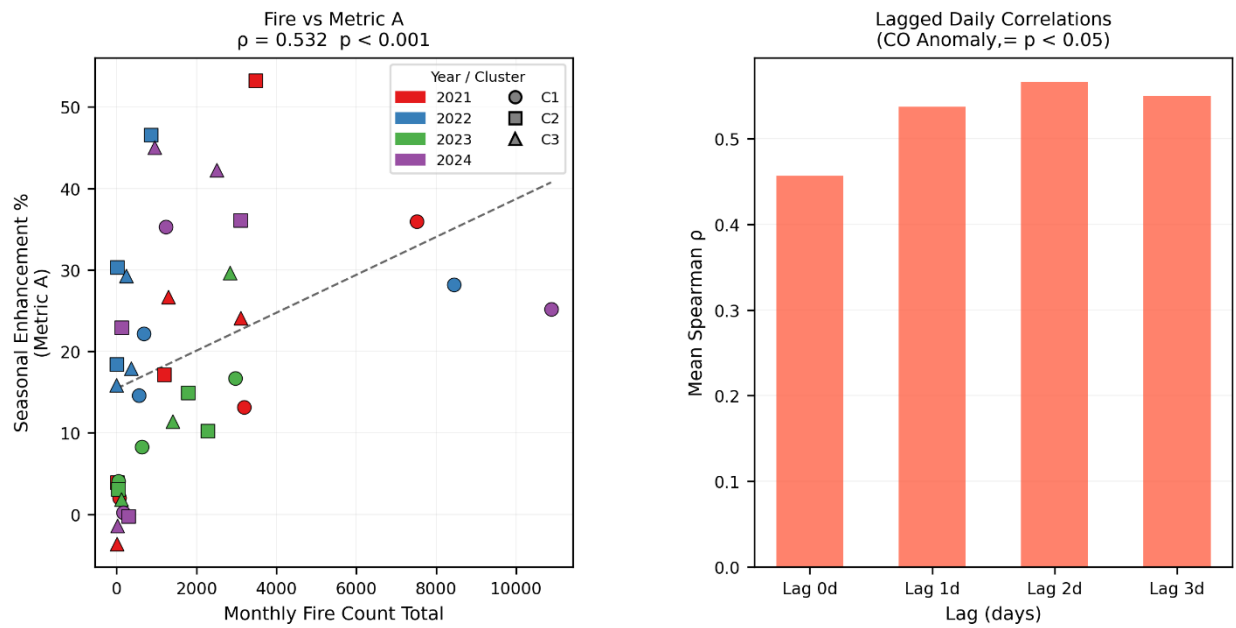


Fig. S4. Spearman correlation between monthly fire count and seasonal CO enhancement (left), and mean lagged daily correlation at 0–3 day lags (right).

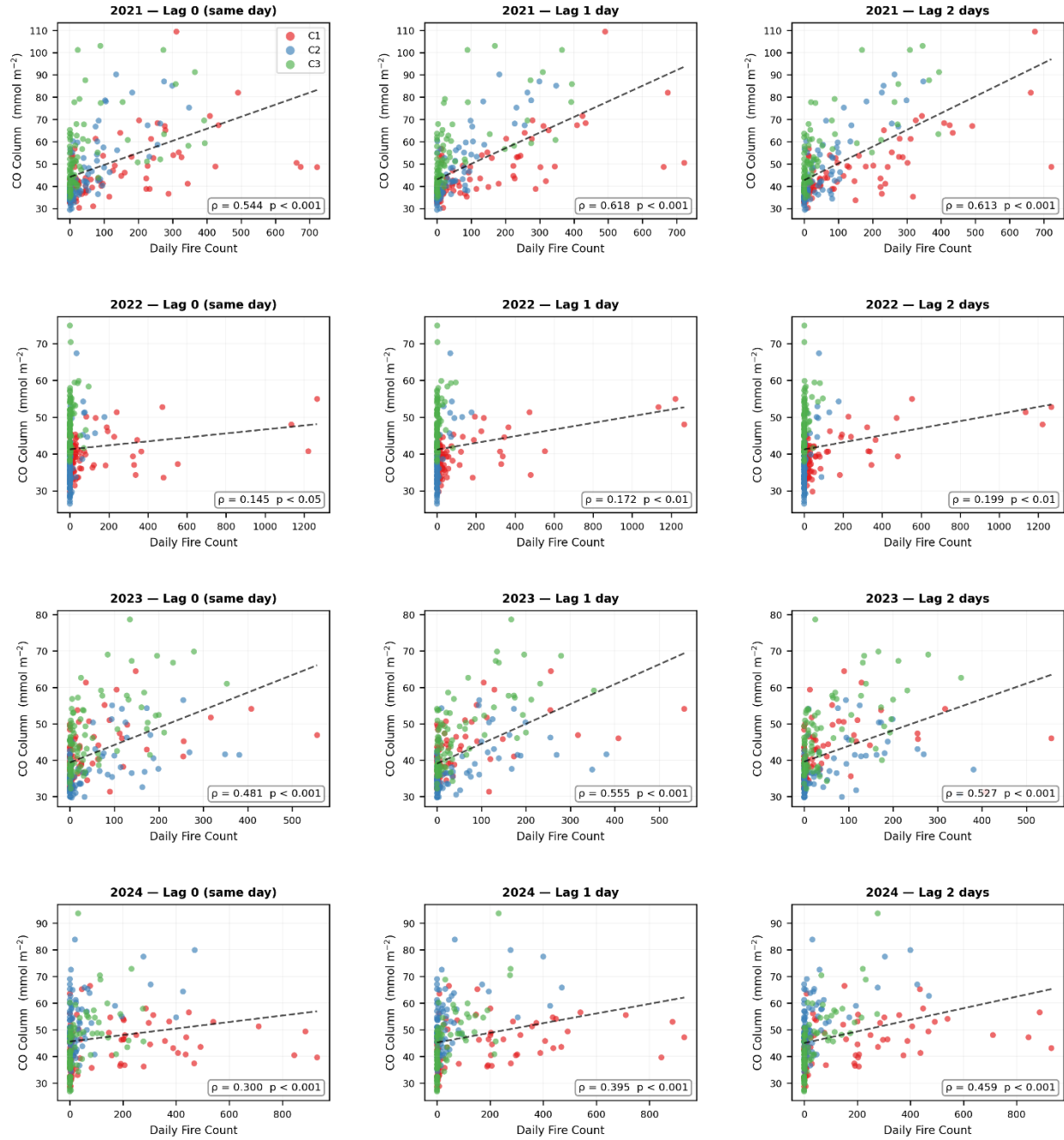


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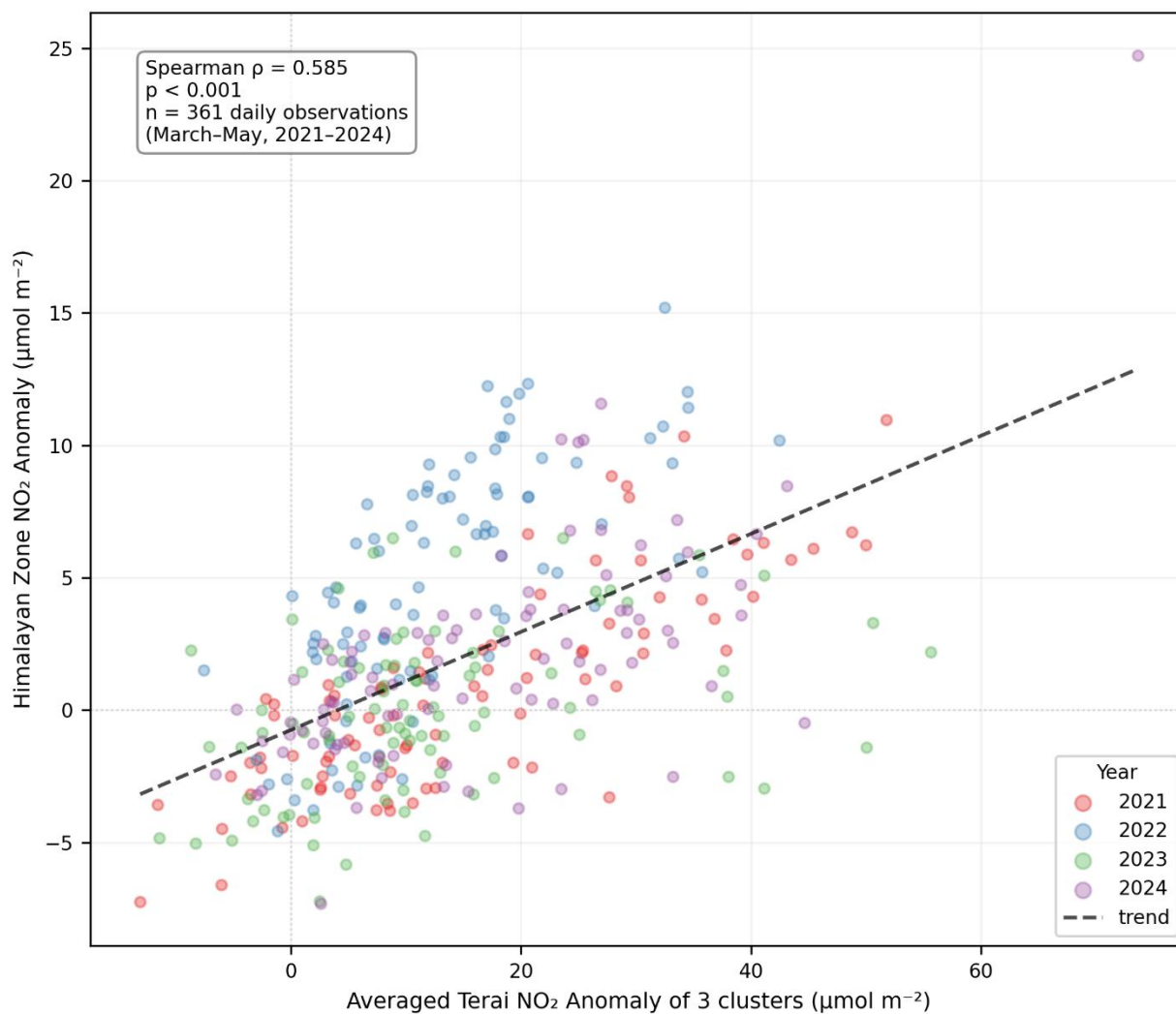


Fig. S6. Same-day Spearman correlation between the averaged NO₂ anomaly across three Terai fire clusters and the Himalayan zone NO₂ anomaly during March–May 2021–2024 .

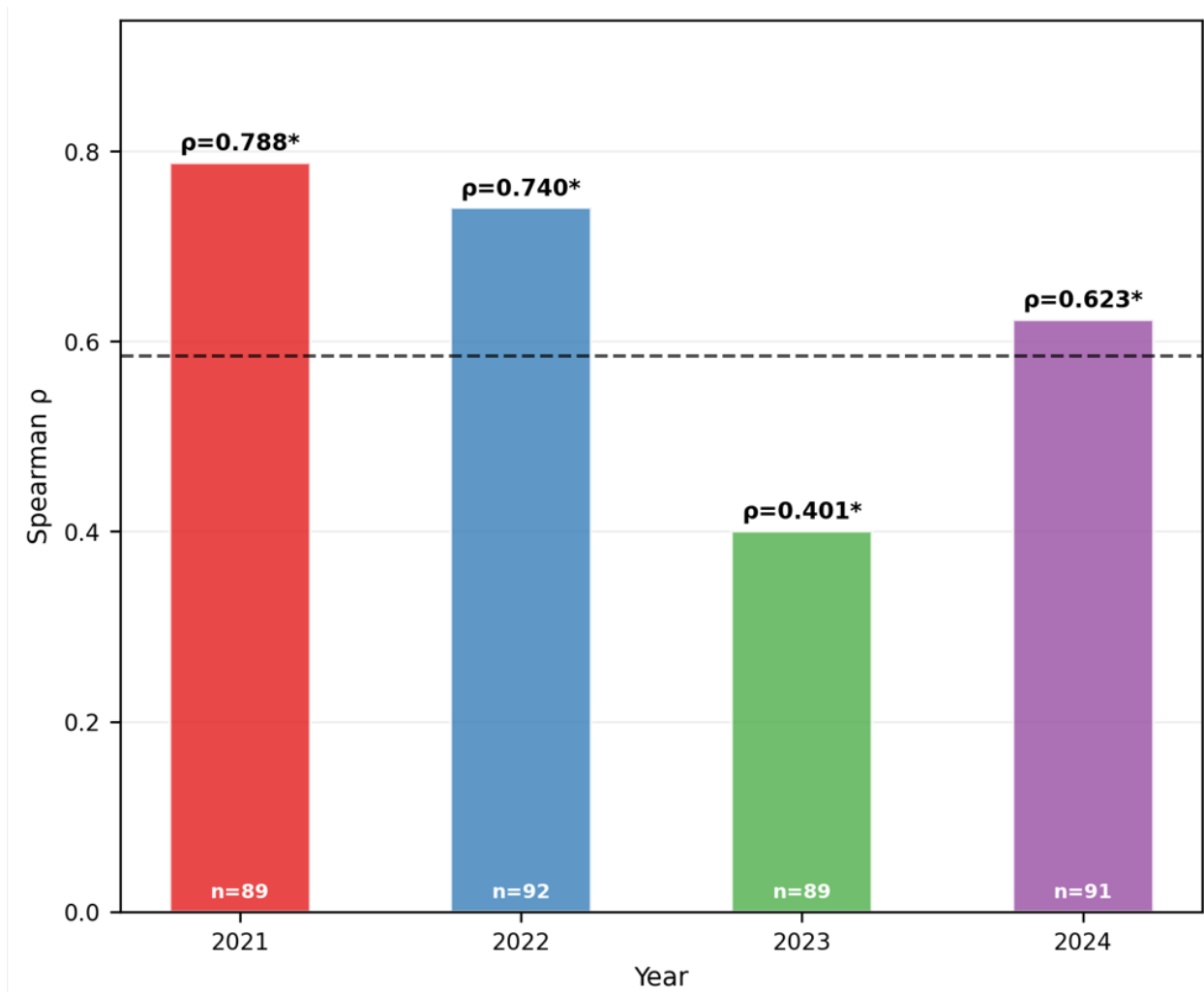


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