

1 **Title: Metabolic retraction and virulence expansion in**
2 **a tropical urban river microbiome under dry-season**
3 **stress**

4 **Muhammad Ehteshamul Haque¹, Md. Shaminur Rahman², Munawar Sultana¹,**
5 **Anowara Beguma^{1*}**

6 **¹ Department of Microbiology, University of Dhaka, Dhaka-1000, Bangladesh**

7 **² Department of Microbiology, Jashore University of Science and Technology,**
8 **Jashore-7408, Bangladesh**

9 *** Corresponding author**

10 Email: anowara@du.ac.bd (AB)

11 Corresponding Co-Author: munawar@du.ac.bd (MS)

12 **Abstract**

13 Anthropogenic pollution and seasonal flow reduction concentrate untreated organic waste
14 in tropical rivers, yet the coupled response of microbial metabolism and virulence gene
15 repertoires remains poorly understood. We applied shotgun metagenomics to the
16 Shitalakshya River (Bangladesh) from early (December 2024) to peak (February 2025)
17 dry season. Dissolved oxygen fell to near-hypoxic levels, ammonia nearly doubled (5.1
18 → 9.95 mg L⁻¹), and total organic carbon surged. KEGG profiling revealed a contracting
19 metabolic core, with amino acid, carbohydrate, and energy metabolism declining by 59–
20 86%, while xenobiotic biodegradation rose 4.6-fold and secondary metabolite
21 biosynthesis 7.4-fold. Network analysis segregated the microbiome into a shrinking

22 growth-associated module and an expanding detoxification cluster. Parallel virulence
23 gene profiling (VFDB) uncovered a 39% increase in total abundance and a 29% richer
24 gene repertoire. Motility and chemotaxis genes nearly doubled, biofilm determinants
25 (*algD*, *alg8*, *mrk* operon) emerged de novo, and Type VI secretion system components
26 became markedly enriched. Conversely, siderophore-mediated iron acquisition and
27 immune evasion factors declined substantially. These coordinated shifts indicate a
28 transition from growth-oriented metabolism to persistence strategies based on
29 environmental sensing and surface colonization. The co-enrichment of motility, biofilm,
30 and T6SS traits alongside pollution-responsive pathways highlights a dual ecological filter
31 wherein dry-season stress simultaneously selects for xenobiotic-tolerant generalists and
32 opportunistic pathogens. Our findings establish a quantitative framework linking seasonal
33 pollution to enhanced pathogenicity potential in urban rivers, with direct implications for
34 drinking water abstraction and public health monitoring.

35 **Keywords:** metagenomics, virulence factors, KEGG (Kyoto Encyclopedia of Genes
36 and Genomes) pathways, Type VI secretion system, xenobiotic degradation.

37 **1. Introduction**

38 Urban rivers in rapidly developing tropical regions are increasingly strained by untreated
39 domestic and industrial wastewater, transforming them into conduits of concentrated
40 organic and chemical pollutants [1–4]. This anthropogenic burden is severely
41 compounded during the dry season, when diminished river flow reduces dilution capacity,
42 leading to sharp declines in dissolved oxygen, accumulation of ammonia, and surges in
43 total organic carbon [5–8]. This phenomenon is exacerbated by the resuspension of
44 organic-rich sediment, which can dramatically increase biological oxygen demand and

45 create or worsen hypoxic conditions [9]. Such physicochemical deterioration pushes
46 aquatic ecosystems toward hypoxic and eutrophic states that impose extreme energetic
47 and redox stress on resident microbial communities [10,11]. While metagenomic surveys
48 of polluted rivers have revealed extensive taxonomic shifts and enrichment of antibiotic
49 resistance genes [12–14], the coupled reprogramming of core metabolic pathways and
50 virulence-associated traits under intensifying dry-season stress remains poorly
51 understood.

52 Microbial communities adapt to environmental adversity through coordinated functional
53 transitions. In model systems, nutrient limitation and oxidative stress trigger stringent
54 responses that downregulate translation, amino acid biosynthesis, and central carbon
55 metabolism while upregulating stress-protective and detoxification modules [15–17]. In
56 freshwater environments, eutrophication and hypoxia have been associated with shifts
57 toward anaerobic energy metabolism and enhanced capacity for xenobiotic degradation
58 [18–20]. However, whether these metabolic adjustments occur in concert with changes
59 in virulence factor (VF) gene repertoires and what functional trade-offs govern such
60 transitions have not been systematically addressed. This gap is particularly pressing
61 because many fitness determinants annotated as virulence factors such as flagella,
62 biofilm components, and Type VI secretion systems (T6SS), also confer selective
63 advantages in extra-host environments, potentially enabling opportunistic pathogens to
64 thrive under polluted conditions [21–25].

65 Emerging evidence indicates that environmental pollution can inadvertently enrich
66 pathogen-associated traits. Wastewater-impacted rivers harbor elevated loads of motility,
67 biofilm, and secretion system genes, often linked to genera that include serious

68 opportunistic pathogens [26–28]. Co-selection of stress tolerance and virulence may
69 occur when the same regulatory networks govern responses to physicochemical
70 challenges and expression of colonization factors [29].

71 Recent spatially resolved metagenomic studies along pollution gradients, such as those
72 in the Kathajodi River, India, have demonstrated that catchment areas receiving untreated
73 urban wastewater exhibit distinct enrichment of fecal-associated bacteria
74 (*Bacteroides*, *Flavobacterium*, *Arcobacter*) and a broader repertoire of resistance and
75 virulence genes compared to upstream sites [30,31]. These spatial patterns suggest that
76 anthropogenic inputs function as strong ecological filters, reshaping community structure
77 and functional potential.

78 Yet, the extent to which dry-season pollution simultaneously reshapes metabolic
79 capabilities and virulence potential within a naturally assembled river microbiome has not
80 been quantified through integrated metagenomic analysis. Understanding this dual
81 reprogramming is essential to evaluate the public health implications of urban freshwater
82 degradation, particularly in settings where river water is used for drinking, bathing, and
83 irrigation [1,32]. Here, we applied shotgun metagenomics to the Shitalakshya River, a
84 tropical urban river in Bangladesh that receives substantial untreated sewage and organic
85 waste, to track the functional trajectory of the planktonic microbiome from the early to the
86 peak of the dry season (December 2024 to February 2025). Concomitant sharp declines
87 in dissolved oxygen, doubling of total organic carbon, and rising ammonia provided a
88 natural experimental setting to test the hypothesis that extreme dry-season stress drives
89 a coordinated metabolic retraction and virulence expansion in the microbial community.
90 By integrating KEGG pathway profiling with comprehensive virulence factor gene

91 cataloging using VFDB [33], we aimed to (i) delineate the global metabolic reprogramming
92 under pollution-induced stress, (ii) characterize shifts in virulence-associated gene
93 abundance and diversity, and (iii) examine functional trade-offs between core anabolism
94 and traits linked to environmental persistence and interbacterial competition. Our results
95 reveal a system-scale ecological transition toward a stress-adapted, pathogen-enriched
96 community state and introduce a compositional metric that may serve as a molecular
97 index of ecological stress in polluted surface water.

98 **2. Materials and methods**

99 **2.1 Study area**

100 The study was conducted at the Sarulia intake point (23°43'02" N, 90°30'00" E) on the
101 Shitalakshya River in Narayanganj, Bangladesh (Supplementary Fig. 1). This site is the
102 primary raw water source for the Saidabad Water Treatment Plant (Phases I and II), the
103 country's largest surface water treatment facility, which supplies drinking water to millions
104 of residents in Dhaka.

105 **2.2 Sample collection and rationale**

106 Water sampling was conducted during two hydrologically distinct dry-season periods,
107 early dry season (December 2024) and peak dry season (February 2025), to capture
108 contrasting environmental conditions associated with progressive flow reduction and
109 pollutant concentration in the river system. Rather than representing continuous seasonal
110 monitoring, these two targeted time points were selected to reflect hydrological extremes
111 within the dry season, when reduced river discharge and diminished dilution capacity
112 intensify anthropogenic pollution stress and alter microbial community dynamics. The

113 environmental relevance of this sampling window is substantiated by operational
114 evidence: during this period, deterioration of raw water quality at the intake of the
115 Saidabad Water Treatment Plant necessitated the implementation of additional biological
116 pre-treatment prior to conventional treatment processes [6].

117

118 At each time point, surface water samples (upper 0.5 m) were collected from the
119 midstream region at the drinking water abstraction zone using sterile 2 L high-density
120 polyethylene (HDPE) bottles. Triplicate water samples (n = 3 per site, 2 L each; 6 L total
121 per site) were collected independently on the same day to account for local spatial and
122 analytical variability within the sampling site. The midstream location was selected to
123 minimize bank-associated heterogeneity and to better represent the mixed water column
124 entering the treatment plant intake. Samples were immediately placed on ice, transported
125 to the laboratory, and stored at 4 °C before processing to minimize biological alteration.
126 Each 2 L replicate was partitioned as follows: 1 L for DNA extraction; 1 L for
127 physicochemical and nutrient analyses (500 mL unfiltered for physicochemical and TOC
128 measurements and 500 mL filtrate was used for nutrient determination and UV_{254}
129 absorbance measurement). This allocation ensured sufficient volume for all downstream
130 analyses while maintaining true field replication.

131 Although sediment can serve as a long-term reservoir and sink for pollutants, antibiotic
132 resistance genes, and microbial contaminants in flowing river systems, the present study
133 specifically focused on the water column because it directly represents the fraction
134 abstracted for drinking water treatment and therefore poses the most immediate public
135 health relevance.

136 By leveraging a targeted paired-sampling design, this approach generates a high-contrast
137 view of dry-season stress, pinpointing key microbial metabolic and virulence shifts under
138 acute pollution. The findings carry direct operational significance for safeguarding drinking
139 water quality at the point of abstraction.

140 **2.3 Physicochemical analysis**

141 Physicochemical parameters pH, temperature, turbidity, conductivity, total dissolved
142 solids (TDS), salinity, and dissolved oxygen (DO) were measured on a 500 mL unfiltered
143 aliquot using calibrated multiparameter probes (HQ30d/HQ40d), a conductivity meter
144 (MM150), and a portable turbidimeter (2100Q; all HACH, USA). Nutrient concentrations
145 ($\text{NH}_3\text{-N}$, $\text{NO}_2\text{-N}$, $\text{NO}_3\text{-N}$) and UV_{254} absorbance were determined from the non-acidified
146 filtrate (~500 mL) using a UV-VIS DR-6000 spectrophotometer (HACH, USA) following
147 standard APHA methods [34]. Total organic carbon (TOC) was quantified from a 40 mL
148 unfiltered aliquot using a Sievers InnovOx ES laboratory TOC analyzer with autosampler
149 (Veolia, USA) based on supercritical water oxidation. All instruments were calibrated per
150 the manufacturer's specifications.

151 **2.4 DNA extraction**

152 Microbial biomass from each 1 L replicate was collected on 0.22 μm membrane filters.
153 Total DNA was extracted with the PowerWater DNA Isolation Kit (Qiagen). Integrity and
154 purity were confirmed by agarose gel electrophoresis (0.8%) and A_{260}/A_{280}
155 spectrophotometry (NanoDrop 2000, Thermo Scientific), respectively. DNA was stored at
156 $-20\text{ }^\circ\text{C}$ until library preparation.

157 **2.5 Library preparation and shotgun metagenomic** 158 **sequencing**

159 Metagenomic libraries were prepared using a modified Nextera XT-based (iNextEra)
160 tagmentation protocol. Briefly, DNA (2.5 μ L) was tagmented with bead-linked
161 transposomes (Illumina DNA Library Prep) and 2 \times TMP buffer [35] in a 6 μ L reaction at
162 53 $^{\circ}$ C for 30 min. Adapter ligation and dual indexing were performed by limited-cycle PCR
163 using custom iNextEra primers with Q5 High-Fidelity DNA Polymerase (New England
164 Biolabs). Cycling conditions were: strand repair at 72 $^{\circ}$ C for 3 min; initial denaturation at
165 98 $^{\circ}$ C for 3 min; 12 cycles of 98 $^{\circ}$ C for 45 s, 62 $^{\circ}$ C for 30 s, and 72 $^{\circ}$ C for 2 min; and final
166 extension at 72 $^{\circ}$ C for 1 min. Amplified libraries (300–700 bp) were verified by agarose
167 gel electrophoresis, quantified using a Qubit Flex fluorometer (Thermo Fisher Scientific),
168 normalized, pooled, and purified with AMPure XP beads (Beckman Coulter). Paired-end
169 sequencing (2 \times 150 bp) was performed on an Illumina NovaSeq X Plus platform
170 (Novogene). The raw metagenomic sequence data have been submitted to the NCBI's
171 Sequence Read Archive (SRA) under the BioProject ID PRJNA1424261
172 (<https://www.ncbi.nlm.nih.gov/bioproject/PRJNA1424261>) (Accession: SAMN55380971
173 and SAMN55380975).

174 **2.6 Taxonomic classification and functional annotation of** 175 **metagenomic sequences**

176 Initial quality evaluation of raw FASTQ files was conducted using FastQC [36]. Adapter
177 trimming and quality filtering were carried out with Trimmomatic [37] employing strict
178 parameters: a 4 bp sliding window, a minimum Phred quality score of 20, and a minimum

179 read length of 50 bp. On average, 70.51 million read pairs per sample were retained post-
180 filtering (range: 52.98–81.91 million pairs). To enrich for microbial sequences, host-
181 derived reads were filtered out using BBduk from the BBtools package [38]. This step was
182 performed *in silico* by aligning the quality-trimmed reads against the human reference
183 genome (GRCh38, GCA_000001405.15_GRCh38_no_alt_analysis_set.fna) using a k-
184 mer length of 31. Reads that did not align to the host genome were retained as "host-
185 filtered" and used for downstream taxonomic profiling. Filtered reads were classified
186 taxonomically using Kraken2 [39] with a confidence threshold of 0.1 and a minimum of 2
187 hit groups required for classification. The analysis was performed against a custom
188 Kraken2 database, a prebuilt Kraken2 database, last updated in July 2025
189 (<https://benlangmead.github.io/aws-indexes/k2>). The database included archaeal,
190 bacterial, viral, and fungal genomes from RefSeq or other standard sources, though the
191 exact composition is user-defined. The Kraken2 classification reports were subsequently
192 processed with Bracken [40] to achieve more accurate species-level abundance
193 estimates. Taxonomic identification was enhanced using TaxonKit [41]. Detailed
194 sequencing quality-control metrics, including trimming efficiency, host-read removal, and
195 microbial-read classification statistics, are provided in S1 Dataset.

196 Functional profiling of antimicrobial resistance was performed using the AMR++ pipeline
197 [42]. High-quality reads were mapped to the MEGARes database (v2.0) [42,43]. Only
198 gene matches with $\geq 99\%$ sequence coverage were retained to ensure specificity. The
199 resulting high-confidence gene identifications were analyzed using ResistomeAnalyzer
200 (v2.1; <https://github.com/cdeanj/resistomeanalyzer>) to characterize antimicrobial
201 resistance profiles. This stringent filtering minimized partial matches and enhanced the

202 accuracy of resistome profiling. Reads Per Kilobase (RPK) gene family abundance tables
203 from HUMAnN3 [44] were processed to consolidate UniRef90 IDs and remove taxonomic
204 annotations. Gene abundances were summed for each UniRef90 identifier to eliminate
205 redundancy. This pipeline ensured comprehensive metabolic profiling while maintaining
206 compatibility with downstream statistical and comparative analyses.

207 **2.7 Data analysis**

208 All statistical analyses were performed in R (v4.2.0) using RStudio (2022.07.1).
209 Comparative analyses of water quality parameters and functional gene abundances
210 between sampling periods were conducted using appropriate parametric or non-
211 parametric tests following assessment of data normality and homogeneity of variance
212 [45–48]. Figures were generated using reproducible scripts within the R statistical
213 environment.

214 **3. Results**

215 **3.1 Water quality deterioration**

216 Physicochemical parameters captured a clear decline in water quality from the early dry
217 season (December 2024) to the peak dry season (February 2025) (**S1 Table**). Dissolved
218 oxygen dropped sharply from 1.55 to 0.54 mg/L, indicating severely hypoxic conditions,
219 while ammoniacal nitrogen increased more than threefold from 3.01 to 10.03 mg/L. Total
220 organic carbon (TOC) showed a dramatic 5.47-fold increase (3.2 to 17.5 mg/L), and
221 turbidity rose from 15.8 to 37.27 NTU (2.36-fold). These changes were accompanied by
222 increases in conductivity (452.94 to 748.36 $\mu\text{S}/\text{cm}$), UV_{254} absorbance (0.09 to 0.20

223 cm^{-1}), and color (15.63 to 52.2 Pt-Co), collectively indicating elevated organic, particulate,
224 and nutrient loading during the peak dry season.

225 **3.2 Global metabolic reprogramming revealed by KEGG** 226 **pathways**

227 At the highest functional level, the metagenome underwent a sweeping redistribution (**Fig**
228 **1**). Metabolism remained the largest category but contracted from 70.3% to 62.5% of
229 annotated reads. This decline was offset by substantial expansions in Human Diseases
230 (from 8.0% to 13.4%), Organismal Systems (1.5% to 7.6%), and Cellular Processes
231 (0.5% to 4.8%), while Genetic Information Processing halved from 16.6% to 8.0% and
232 Environmental Information Processing decreased modestly from 2.5% to 3.7%. Such a
233 sweeping redistribution of functional potential signals a systemic shift from growth-
234 oriented metabolism toward stress adaptation and defense.

235
236 **Fig 1.** Global KEGG pathway class distribution in December 2024 and February 2025.
237 Pie charts showing the relative abundance (%) of six major functional categories.
238 Metabolism contracts while Human Diseases, Organismal Systems, and Cellular
239 Processes expand.

240 **3.3 Pollution-Induced Shifts in Microbial Functional Potential**

241 To assess the impact of pollution stress on microbial community function, we analyzed
242 the relative abundance of metabolic pathways and functional subclasses between
243 December 2024 and February 2025. The data reveals a significant shift in functional

244 profiles, characterized by an upregulation of stress-response and detoxification pathways
245 and a marked downregulation of core growth and biosynthesis pathways.

246 **3.3.1 Shifts in Functional Subclasses**

247 The analysis of key functional subclasses (**Fig 2**) highlights a distinct divergence between
248 defense mechanisms and central metabolism. Pathways associated with stress tolerance
249 and environmental interaction showed robust increases in relative abundance. Most
250 notably, *Xenobiotics biodegradation and metabolism* exhibited the largest positive
251 change, increasing by +8.1% (from ~2.5% to ~10.5%). Other pathways that significantly
252 increased include *Biosynthesis of other secondary metabolites* (+5.5%), *Lipid*
253 *metabolism* (+3.6%), *Metabolism of terpenoids and polyketides* (+2.7%), and *Infectious*
254 *disease: bacterial* (+2.6%). Conversely, pathways fundamental to microbial growth and
255 energy production experienced substantial reductions. *Amino acid metabolism* showed
256 the steepest decline, dropping by -10.0% (from ~17.5% to ~7.5%). Significant decreases
257 were also observed in *Energy metabolism* (-6.9%), *Nucleotide metabolism* (-
258 6.6%), *Translation* (-6.1%), and *Carbohydrate metabolism* (-4.5%). These opposing
259 trends indicate a functional restructuring away from anabolic growth processes toward
260 survival, adaptation, and detoxification.

261
262 **Fig 2.** Functional restructuring under pollution stress. This dumbbell plot shows the shift
263 in microbial pathway abundance from December 2024 to February 2025. Core growth
264 and biosynthetic pathways (e.g., Amino acid and Energy metabolism) decreased
265 significantly (red lines), while stress-response and detoxification pathways (e.g.,

266 Xenobiotics biodegradation) increased (green lines), reflecting a community-wide shift
267 from growth to survival.

268 **3.3.2 Biological Importance and Magnitude of Change**

269 To further contextualize these shifts, we plotted the February 2025 abundance against
270 the December 2024 abundance, with point size indicating biological importance
271 (Abundance \times Change Magnitude) and shape denoting pathway type (**Fig 3**). Consistent
272 with the subclass analysis, Xenobiotics biodegradation and metabolism emerge as a
273 critical node, showing a high fold change (FC: 4.63x) and positioning significantly above
274 the dashed line of no change. Lipid metabolism also shows a strong positive shift (FC:
275 1.92x). In contrast, pathways such as Amino acid metabolism (FC: 0.41x)
276 and Carbohydrate metabolism (FC: 0.63x) fall below the baseline, confirming a reduction
277 in abundance.

278

279 **Fig 3.** Pathway Importance and Change Analysis. This scatter plot visualizes the shift in
280 microbial pathway abundance before (December 2024, x-axis) and after (February 2025,
281 y-axis) pollution stress. The dashed diagonal line represents no change; points above the
282 line increased in abundance, while points below decreased. Marker size indicates
283 biological importance (abundance \times fold change), and color/shape denotes functional
284 class. Key anabolic pathways (e.g., Amino acid and Energy metabolism) shifted below
285 the line, indicating decline, whereas detoxification and stress pathways (e.g., Xenobiotics
286 biodegradation and Infectious disease) surged above the line with high fold changes.

287

288 Despite its low absolute abundance, Infectious disease: bacterial exhibits a dramatic fold
289 change (FC: 22.33x), and Signaling molecules and interaction shows an infinite fold
290 change (emerging from near-zero levels), suggesting that even low-abundance stress-
291 response pathways are being heavily activated. Energy and translation-related pathways
292 remained clustered in the lower abundance range with fold changes near 0.3x, reinforcing
293 the conclusion that pollution stress suppresses microbial growth machinery in favor of
294 survival mechanisms.

295

296 **3.4 Integrated KEGG network analysis reveals ecological** 297 **resilience of the river microbiome**

298 KEGG pathway interaction network analysis demonstrated pronounced functional
299 reorganization of the Shitalakshya River microbiome during the dry-season period
300 (December 2024–February 2025). Central network modules (**Fig 4**) were dominated by
301 metabolism-associated pathways, including carbohydrate, amino acid, lipid, and
302 nucleotide metabolism, which exhibited high connectivity and interaction density,
303 indicating coordinated metabolic adaptation under environmentally stressed conditions.
304 Pathways related to genetic information processing, particularly translation and
305 replication/repair, were tightly integrated with core metabolic functions, suggesting
306 sustained microbial growth and active cellular maintenance. In contrast, several
307 peripheral functional clusters displayed directional shifts in abundance, reflecting
308 selective enrichment or suppression of specific ecological processes. Notably, xenobiotic
309 biodegradation pathways remained strongly connected within the network, consistent
310 with adaptive responses to anthropogenic contamination pressure. Collectively, the

311 network topology indicates that metabolic and information-processing pathways underpin
312 the ecological resilience and functional stability of the riverine microbiome during
313 seasonal environmental perturbation.

314

315 **Fig 4.** KEGG Pathway Interaction Network. Network diagram showing functional shifts in
316 the Shitalakshya River microbiome (Dec 2024 – Feb 2025). Nodes represent
317 pathways; node size reflects average abundance, fill color shows functional class,
318 and border color indicates change direction (blue = decreased, red = increased). The
319 network reveals a stark division: core biosynthetic and metabolic pathways (left cluster)
320 declined, while stress-related and disease pathways (right clusters) increased in
321 abundance.

322 **3.5 Dry-season stress enriches a functionally coherent** 323 **consortium of xenobiotic-degrading and disease-associated** 324 **microorganisms**

325 The 4.6-fold enrichment in community-level xenobiotic biodegradation capacity (from
326 2.22% to 10.27% of annotated KEGG reads) between December 2024 and February
327 2025 was accompanied by a profound restructuring of the Shitalakshya River
328 microbiome, characterised by the selective proliferation of two ecologically intertwined but
329 clinically distinct bacterial guilds: canonical xenobiotic degraders and opportunistic
330 human pathogens.

331

332

333 **3.5.1 Specialized xenobiotic-degrading consortium emerges.**

334 The functional surge in pollutant catabolism was underpinned by the disproportionate
335 expansion of a phylogenetically diverse but metabolically coherent assemblage of
336 degraders. The most extreme enrichment was observed for *Achromobacter aegrifaciens*,
337 an emerging degrader of organopollutants, which surged 1,945-fold from near-detection
338 limits (20 reads) to dominance (38,896 reads; Log_2 fold change = +10.93; adjusted $P <$
339 0.001; **Fig 5a**). This was accompanied by parallel expansions of *Brevundimonas*
340 *diminuta* (28.6-fold; 86,600 \rightarrow 2,481,257 reads), a hydrocarbon degrader,
341 and *Comamonas testosteroni* (9.5-fold; 110,297 \rightarrow 1,043,615 reads), renowned for
342 steroid and plastic catabolism (**S2 Table**). Unclassified *Pseudomonas* species expanded
343 8.3-fold (42,527 \rightarrow 352,601 reads), suggesting a substantial reservoir of uncharacterized
344 catabolic potential. Critically, this enrichment was sharply (Roges et al., 2020; Wang et
345 al., 2023selective: *Thauera humireducens*, an aromatic degrader, collapsed 25-fold
346 (19,869 \rightarrow 849 reads), while *Acidovorax temperans* declined 0.6-fold (117,948 \rightarrow 72,764
347 reads), indicating that the dry-season physicochemical regime characterized by elevated
348 pollutant concentrations, hypoxia, and reduced flow imposed a stringent ecological filter
349 favoring a specific subset of the degrading guild (**Fig 5b**). This filter favored a specific
350 cohort of organisms, predominantly within the *Pseudomonadota*, exhibiting a combination
351 of versatile catabolism, hypoxia-tolerance, and robust stress-response systems. The co-
352 enrichment of *Pseudomonas putida* and a vast pool of
353 uncharacterized *Pseudomonas* species (8.3-fold increase to 35% of the genus-level
354 reads) further indicates that much of this environment's xenobiotic degradation potential
355 resides in a flexible and rapidly responding functional guild, not a single taxonomic group.

356
357 **Fig 5.** Selective enrichment of a xenobiotic-degrading consortium under dry-season
358 conditions. (a): Volcano plot of differential bacterial abundance between December 2024
359 and February 2025. Each point represents a species. Significantly enriched (red; Log_2
360 fold change > 1 , adjusted $P < 0.05$) and depleted (blue; Log_2 fold change < -1 ,
361 adjusted $P < 0.05$) taxa are highlighted against non-significant background species
362 (grey). Dashed lines denote statistical thresholds. The most extreme responders,
363 *Achromobacter aegrifaciens* (1,945-fold enriched) and *Thauera humireducens* (25-fold
364 depleted) are labelled. (b) Dot plot heatmap showing the relative abundance (dot size)
365 and enrichment fold change (dot colour) of 16 key xenobiotic-degrading species, grouped
366 by substrate class (right sidebar). The heatmap highlights the selective nature of the
367 enrichment across different catabolic guilds.

368 **3.5.2 Concurrent enrichment of clinically relevant pathogens**

369 Strikingly, this niche filter also selected for a broad array of disease-associated
370 microorganisms, revealing a convergent ecological response where traits conferring
371 fitness under environmental stress, such as multidrug efflux, biofilm formation, and broad
372 substrate utilization, simultaneously potentiate opportunistic pathogenicity. Reads
373 assigned to the major opportunistic pathogen *Pseudomonas aeruginosa* increased 7.1-
374 fold (8,236 \rightarrow 58,220), while the multidrug-resistant nosocomial
375 pathogen *Stenotrophomonas maltophilia* rose 9.5-fold (2,554 \rightarrow 24,346). The recently
376 described bloodstream infection-associated *Stenotrophomonas pigmentata* exhibited a
377 325-fold expansion (107 \rightarrow 34,775 reads). Several clinically significant species
378 emerged *de novo* in the February 2025 metagenome, having been undetectable in

379 December 2024: *Clostridioides difficile* (0 → 277 reads), *Enterococcus faecalis* (0 → 137
380 reads), *Listeria monocytogenes* (0 → 244 reads), and *Pasteurella multocida* (0 → 236
381 reads) (**Fig 6; S3 Table**). The most dramatic proliferation was observed for *Tissierella*
382 *carlieri*, an emerging anaerobic opportunistic pathogen associated with soft tissue
383 infections and bacteremia, which appeared *de novo* and reached 1,995,804 reads,
384 becoming one of the most abundant single taxa in the February community.
385 Similarly, *Aeromonas media*, an enteric pathogen causing diarrheal disease, surged
386 116.8-fold (12,596 → 1,470,792 reads), while *Clostridium tetani*, the causative agent of
387 tetanus, expanded 171.5-fold (67 → 11,490 reads). The anaerobic
388 opportunist *Bacteroides thetaiotaomicron* increased 248-fold (65 → 16,126 reads),
389 and *Empedobacter felsenii*, a hospital-acquired bloodstream pathogen, rose 50.3-fold
390 (14,723 → 739,932 reads). These expansions occurred alongside the persistence of high-
391 risk pathogens at clinically concerning abundances: *Klebsiella pneumoniae* maintained a
392 persistent high load (~48,000 reads at both time points), while *Morganella*
393 *morganii* expanded 4.2-fold to 871,649 reads.

394

395 **Fig 6.** Lollipop chart of differentially abundant disease-associated species. Point colour
396 indicates enrichment status; diamond shapes denote *de novo* emergence (absent in
397 December 2024)

398 **3.5.3 Ecological coupling of degradation and pathogenicity traits.**

399 The co-enrichment of xenobiotic degraders and opportunistic pathogens was not
400 taxonomically independent. Several of the most highly enriched taxa occupy both
401 functional categories simultaneously. *Achromobacter xylosoxidans*, a known degrader of

402 aromatic compounds and herbicides (36.9-fold enrichment; 625 → 23,038 reads), is also
403 a recognised opportunistic pathogen causing bacteremia and pneumonia in cystic fibrosis
404 patients. *Brevundimonas diminuta*, which expanded 28.6-fold to dominate the community,
405 couples hydrocarbon degradation with the capacity to cause catheter-related bacteremia
406 and peritonitis. *Pseudomonas aeruginosa*, enriched 7.1-fold, is both a model xenobiotic
407 degrader and a paradigmatic multidrug-resistant nosocomial pathogen. This dual
408 functionality suggests that the genomic determinants of environmental stress tolerance,
409 including broad-specificity efflux pumps, mobile genetic elements, and catabolic
410 plasmids, may simultaneously encode or facilitate the acquisition of virulence-associated
411 traits, blurring the boundary between environmental generalist and opportunistic
412 pathogenicity.

413 **3.5.4 Selective depletion of waterborne and zoonotic pathogens**

414 In marked contrast to the enrichment of nosocomial and wound-associated opportunists,
415 several classical waterborne pathogens declined precipitously. *Vibrio cholerae*, the
416 causative agent of cholera, decreased from 9,970 to 632 reads, while *Vibrio*
417 *parahaemolyticus* and *Vibrio fluvialis* fell to 656 and 95 reads, respectively. The
418 foodborne pathogen *Yersinia enterocolitica* declined from 584 to 76 reads. These parallel
419 declines suggest that the intensified dry-season conditions, particularly elevated salinity
420 and reduced dissolved oxygen may have exceeded the physiological tolerance ranges of
421 these typically estuarine and freshwater pathogens, even as they selected for the
422 hypoxia-tolerant, biofilm-forming opportunists that dominated the enriched fraction.

423

424

425 **3.5.5 Synthesis**

426 Together, these taxonomic shifts reveal a deterministic ecological response to intensifying
427 dry-season stress: the selective enrichment of a dual-functional microbial assemblage
428 capable of both xenobiotic catabolism and opportunistic infection. The convergence of
429 these traits in a single, co-enriched consortium dominated
430 by *Achromobacter*, *Pseudomonas*, *Brevundimonas*, *Comamonas*, *Stenotrophomonas*,
431 and anaerobic *Clostridium/Tissierella* lineages indicates that chronic pollution and
432 hydrological stress may inadvertently cultivate a latent public health reservoir within urban
433 riverine ecosystems. This coupling of biodegradative capacity and pathogenic potential
434 represents an underexplored dimension of water quality deterioration in
435 anthropogenically impacted tropical rivers.

436 **3.6 Virulence factor dynamics**

437 The comparative analysis of virulence factor (VF) profiles between December 2024 and
438 February 2025 revealed a pronounced restructuring of functional traits under dry-season
439 stress (**Fig 7**). While the overall composition remained structurally consistent across ten
440 functional categories, substantial shifts in relative abundance indicate strong ecological
441 reprogramming.

442

443 **Fig 7.** Stacked bar chart showing the relative abundance of major virulence factor
444 categories in December 2024 and February 2025. A pronounced increase in motility-
445 related factors is observed in February, accompanied by enrichment of biofilm formation,
446 indicating enhanced dispersal and persistence under dry-season stress. In contrast,

447 nutritional/metabolic, immune modulation, and invasion-related functions declined,
448 reflecting reduced investment in host-associated and energy-intensive processes.

449 **3.7 Expansion and restructuring of virulence potential**

450 Temporal metagenomic analysis revealed a significant expansion of virulence-associated
451 genetic potential from early (December 2024) to peak dry season (February 2025). Total
452 virulence factor (VF) abundance increased by 38.9% (from 46.02 to 63.91), accompanied
453 by a 28.6% expansion in the number of detected VF genes (from 294 to 378; **Table 1**).
454 This coordinated increase suggests a net gain in community-wide virulence capacity
455 rather than a simple redistribution of existing genes.

456 **Table 1. Temporal dynamics of virulence factor abundance and gene counts**
457 **(December 2024 – February 2025).**

Category	Abundance (Dec)	Abundance (Feb)	Δ Abundance	Genes (Dec)	Genes (Feb)	Δ Genes	Trend
Motility	19.35	33.58	+14.23	55	108	+53	↑↑
Biofilm formation	2.41	4.63	+2.22	8	13	+5	↑

Category	Abundance (Dec)	Abundance (Feb)	Δ Abundance	Genes (Dec)	Genes (Feb)	Δ Genes	Trend
Effector delivery system	22.68	22.88	+0.20	68	98	+30	↑
Adherence	22.89	21.37	-1.52	81	76	-5	↓
Nutritional/Metabolic factor	13.78	8.06	-5.72	51	55	+4	⇓
Immune modulation	6.79	3.38	-3.41	18	16	-2	↓
Antimicrobial activity/Competitive advantage	6.47	4.07	-2.40	2	2	0	↓
Invasion	3.01	1.08	-1.93	4	6	+2	↓

Category	Abundance (Dec)	Abundance (Feb)	Δ Abundance	Genes (Dec)	Genes (Feb)	Δ Genes	Trend
Regulation	2.56	0.94	-1.62	4	3	-1	↓
Exotoxin	0.09	0.03	-0.06	3	1	-2	↓
TOTAL	46.02	63.91	+17.89	294	378	+84	↑

458

459 3.8 Motility and biofilm expansion

460 The most dramatic restructuring occurred within motility-associated genes, which surged
461 from 19.35 to 33.58 in cumulative abundance (+14.23), making it the top category in
462 February (**Fig 8**). Gene count in this class nearly doubled from 55 to 108, reflecting
463 massive mobilization of flagellar and chemotaxis systems (**S1 Dataset**). Specific flagellar
464 motor and switch components *fliM* (0.65 → 1.92), *flgG* (0.49 → 1.16), *fliG*, and *flgI*
465 showed substantial increases, while chemotaxis genes *cheW*, *cheB*, *cheY*,
466 and *pomA2* became more abundant or appeared *de novo*.

467

468 **Fig 8.** Virulence factor abundance across functional categories during dry-season
469 progression. Relative abundance of VF categories in December 2024 and February 2025.
470 A pronounced increase in motility ($\Delta = +14.00$) and biofilm formation ($\Delta = +2.00$) was
471 observed in February, indicating enhanced microbial dispersal and surface colonization

472 potential. In contrast, nutritional/metabolic functions ($\Delta = -7.00$), immune modulation (Δ
473 $= -3.00$), and antimicrobial activity ($\Delta = -2.00$) declined.

474
475 Concomitantly, biofilm pathway genes displayed a marked upswing (+2.22 absolute). The
476 alginate biosynthesis cluster (*algD*, *alg8*, *algE*, *algQ*, *algR*) was absent in December but
477 reached notable levels in February, while type 3 fimbrial genes (*mrkA-F*) increased
478 steadily, indicating enhanced potential for surface attachment and matrix production. The
479 co-enrichment of motility and biofilm traits indicates a dynamic ecological balance
480 between dispersal and surface-associated survival, potentially reflecting biofilm
481 formation–dispersal cycles rather than mutually exclusive behaviors

482 **3.9 Effector delivery systems remain robust with nuanced** 483 **changes**

484 Effector delivery systems, predominantly Type II, Type III, and Type VI secretion systems,
485 remained the second most abundant class (**S1 Dataset**), with a modest overall increase
486 (22.68 \rightarrow 22.88). The Type VI secretion system (T6SS) expanded robustly. The T6SS
487 ATPase *clpB* emerged *de novo* (0 \rightarrow 1.10), together with multiple *ats* genes
488 (*atsC*, *atsH*, *atsJ*), contractile sheath subunits *vipA/tssB* and *vipB/tssC*, and the
489 effector *vgrG2* (0 \rightarrow 0.55). In contrast, some Type III secretion system effectors
490 (*espL1*, *espX1*) declined, indicating a strategic shift toward T6SS-mediated interbacterial
491 competition rather than eukaryotic cell manipulation.

492

493 **3.10 Attenuation of iron acquisition systems and immune** 494 **evasion mechanisms**

495 Categories linked to nutrient acquisition and immune subversion decreased substantially
496 **(S1 Dataset)**. Nutritional/metabolic factors decreased by 5.72 units (13.78 → 8.06),
497 driven largely by reductions in siderophore biosynthesis and transport genes. The
498 enterobactin system (*entA-F*, *fepA-G*) and aerobactin receptor *iutA* all exhibited marked
499 attenuation, as did yersiniabactin-associated genes (*irp2*, *ybtS*, *ybtU*), many of which fell
500 below detection. Immune modulation capacity declined by 3.41 units (6.79 → 3.38).
501 Capsule assembly proteins (e.g., *KP1_RS17345*, *Wzi*) and lipopolysaccharide
502 modification enzymes (*lpxC*, *waaC*, *waaG*) were strongly reduced or lost. Similarly,
503 invasion-associated genes (*ibeB*, *ompA* variants) and exotoxins contracted. The
504 attenuation of these traits likely reflects relaxed selection for host-associated functions
505 and reduced iron limitation in hypoxic, organic-rich waters.

506 **4. Discussion**

507 The parallel application of KEGG pathway profiling and virulence factor (VF) gene
508 cataloging to the Shitalakshya River, a tropical system under acute organic pollution,
509 reveals a coherent, multidimensional microbiome stress response as dry-season
510 conditions intensify. Over only two months, the metagenome shifted decisively away from
511 biosynthesis and growth toward detoxification, motility, biofilm formation, and
512 interbacterial competition, a transformation that simultaneously curtails core anabolism
513 and enriches traits with direct implications for public health. This temporal functional
514 trajectory is consistent with spatial patterns observed along pollution gradients in other

515 urban rivers, where catchment areas receiving untreated wastewater serve as hotspots
516 for enrichment of resistance and virulence genes [12,49,50,20,47,30,31].

517

518 **4.1 Ecological filtering of core metabolism**

519 The collapse of central metabolism is the most conspicuous signature of environmental
520 deterioration. Amino acid, carbohydrate, and energy metabolism, together with translation
521 and nucleotide biosynthesis, contracted by 60–85%, forming a tightly connected module
522 of suppressed functions. This coordinated downregulation mirrors the stringent response
523 and stationary phase physiology known from model organisms, in which resource scarcity
524 and redox stress halt ribosome production and anabolic pathways [51,52,17,24]. The
525 sharp decline in dissolved oxygen to near-hypoxic levels, the doubling of total organic
526 carbon, and the accumulation of toxic ammonia impose simultaneous energy limitation,
527 carbon catabolite repression, and nitrogen stress [7,8,4]. The link between elevated
528 organic matter in resuspended sediment and subsequent oxygen depletion, as modeled
529 by Morote-Sánchez et al. [9], provides a mechanistic explanation for the severe hypoxia
530 observed in the late dry season. Under such combined pressures, shutting down
531 expensive anabolic machinery is a conserved survival strategy, as evidenced across
532 aquatic metagenomes under eutrophic and hypoxic conditions [53,11,54,30]. The halving
533 of genetic information processing capacity further indicates that the late dry-season
534 community has entered a regime of profoundly reduced growth, likely reflecting a
535 systemic metabolic shutdown in planktonic autochthonous bacteria.

536

537

538 **4.2 Recruitment of detoxification modules**

539 In place of core metabolism, we observed a massive induction of xenobiotic
540 biodegradation and secondary metabolite biosynthesis. These pathways expanded as a
541 distinct, loosely connected network module, decoupled from central carbon metabolism.
542 This architecture suggests the recruitment of peripheral, often plasmid- or island-encoded
543 catabolic pathways that operate with minimal interference with core processes. The 4.6-
544 fold surge in xenobiotic degradation capacity, the enrichment of secondary metabolite
545 biosynthesis, and the expansion of lipid metabolism point to the activation of detoxification
546 and membrane remodeling mechanisms characteristic of pollutant-degrading specialists
547 and biofilm-associated communities [18,19,50]. The concomitant rise in
548 terpenoid/polyketide pathways further indicates extensive production of stress-protective
549 compounds [29]. Thus, the community transitions from a growth-oriented metabolic
550 economy to one centered on detoxification and chemical defense, a functional shift
551 analogous to the enrichment of carbohydrate and xenobiotic metabolism in wastewater-
552 impacted river sections [30].

553 **4.3 Expansion and restructuring of virulence potential**

554 A particularly consequential dimension of this reprogramming is the expansion and
555 restructuring of virulence-associated genetic potential. Total virulence factor abundance
556 increased by 39%, accompanied by a 29% increase in detected genes, indicating a net
557 gain in community-wide pathogenic capacity [53,20,24,47,48,27].
558 Flagellar and chemotaxis gene counts nearly doubled, indicating enhanced capacity for
559 directional movement toward nutrients and away from toxins, a process tightly regulated
560 by the flagellar assembly machinery [65]. The *de novo* appearance of the full alginate

561 biosynthesis cluster, together with enhanced fimbrial genes, signals a heightened
562 potential for surface colonization and matrix production [55,33,31].

563 Motility and biofilm are often viewed as opposing lifestyles. Still, their co-enrichment here
564 supports the concept of biofilm formation dispersal cycles, where motile cells seed new
565 surfaces and biofilm cells later release propagules, a dynamic bet-hedging strategy that
566 maximizes persistence under fluctuating stress [56]. Elevated turbidity and TOC increase
567 suspended organic particles that act as nutrient-rich microsites and attachment surfaces,
568 promoting biofilm formation [57–59].

569 The robust expansion of the Type VI secretion system (T6SS) repertoire, including the
570 emergence of key structural components (*clpB*, *vgrG2*) alongside an increase in pivotal
571 sheath proteins, indicates a strategic pivot toward interbacterial competition [60,61]. T6SS
572 functions as a contact-dependent weapon that injects toxic effectors into neighboring
573 competitors; its upregulation is frequently associated with high cell density, surface-
574 associated growth, and interspecies warfare. The parallel decline of Type III secreted
575 effectors suggests a departure from eukaryotic cell-targeted virulence [62], consistent
576 with the reduced immunological pressure in an environmental setting. In the resource-
577 limited, high-density communities of the late dry season, T6SS-mediated killing can
578 confer a strong fitness advantage, potentially favoring pathogens that double as effective
579 microbial competitors.

580 **4.4 Attenuation of iron acquisition and immune evasion**

581 Remarkably, this heightened competitive and dispersal potential was accompanied by a
582 strong attenuation of iron acquisition systems and immune evasion factors. Siderophore
583 biosynthesis (enterobactin, aerobactin, yersiniabactin) and capsule/LPS modification

584 genes receded sharply [23,29]. In hypoxic, organic-rich sediments, iron bioavailability
585 typically increases due to reductive dissolution and complexation with humic substances,
586 thereby diminishing the selective advantage of high-affinity iron chelators.
587 Simultaneously, the loss of immune-modulatory and invasion traits may reflect relaxed
588 selection for host-associated functions in an extra-host environment, or a trade-off with
589 the energy demands of stress survival [63]. The net effect is a community that combines
590 enhanced environmental persistence and interbacterial competitiveness with a
591 preserved, albeit altered, arsenal of virulence determinants [12].

592 **4.5 Ecological and public health implications**

593 The elevation of KEGG "Infectious disease: bacterial" pathways (a 23-fold increase) and
594 the 38.9% rise in total VF content should not be interpreted as a simple increase in
595 obligate pathogen load. Many VFDB-annotated genes—flagella, T6SS, and biofilm
596 polysaccharides—are also fundamental ecological fitness factors. However, their
597 amplification in a water body that serves as a source of drinking, bathing, and irrigation
598 water raises legitimate concerns about the dissemination of opportunistic pathogens such
599 as *Pseudomonas aeruginosa*, *Aeromonas* spp., and *E. coli*, which exploit these very
600 traits for environmental survival and human infection [1,23]. The *de novo* appearance of
601 quorum-sensing and signaling pathways hints at the activation of cell–cell communication
602 networks that could synchronize virulence factor expression at the community scale,
603 further amplifying risk [26]. These findings align with evidence that severe organic
604 enrichment and hypoxia enhance the persistence of enteric pathogens [32,28].

605

606

607 **4.6 A proposed ecological index of stress**

608 By integrating KEGG functional profiling with VF enumeration, we propose a composite
609 "xenobiotic degradation to translation ratio" (XD/T) as a potential molecular index of
610 ecosystem stress. In December, XD/T was ~ 0.27 (2.22/8.23); by February, it jumped to
611 4.80 (10.27/2.14), an 18-fold increase. This ratio captures the fundamental trade-off
612 between anabolic capacity (translation) and detoxification potential (xenobiotic
613 degradation). Similar composite indices have been proposed in soil ecology (e.g., carbon
614 quality indexes), and the XD/T ratio could be developed as a sensitive biomonitoring
615 metric for assessing functional deterioration in polluted surface water systems.
616 Metatranscriptomic validation will be required to confirm the active transcription of these
617 pathways and to link them directly to pollutant concentrations.

618 **4.7 Conceptual framework**

619 Conceptual framework (Fig 9) illustrating the integrated dual-filter model of dry-season
620 stress on the river microbiome. Seven hierarchical levels depict the cascade from
621 environmental drivers to public health outcomes: (1) Environmental drivers (low flow,
622 declining DO, rising NH_3 and TOC); (2) Metabolic retraction (suppression of amino acid,
623 energy, and translation pathways); (3) Microbial community restructuring (enrichment of
624 xenobiotic degraders and opportunistic pathogens); (4) Virulence factor dynamics
625 (expansion of motility, biofilm, T6SS; contraction of iron acquisition and immune
626 modulation); (5) Ecological mechanisms (energy trade-off, co-selection, regulatory cross-
627 talk, environmental cues); (6) Ecosystem and public health outcomes; and (7) The XD/T
628 ratio as a composite stress index ($0.27 \rightarrow 4.80$, 18-fold increase). DO, dissolved oxygen;

629 NH₃, ammonia; TOC, total organic carbon; T6SS, Type VI secretion system; XD/T,
630 xenobiotic degradation to translation ratio.

631

632 **Fig 9:** Conceptual framework of dry-season stress-induced microbiome
633 restructuring. The cascade from environmental drivers (↓DO, ↑NH₃, ↑TOC) to metabolic
634 retraction (↓86% amino acids, ↓69% energy, ↓74% translation), microbial enrichment
635 (xenobiotic degraders and pathogens), virulence shifts (↑motility +14.2, ↑biofilm +2.2,
636 ↑T6SS; ↓iron -5.7, ↓immune -3.4), ecological mechanisms, outcomes, and the XD/T
637 stress index (18-fold increase).

638 **4.8 Limitations and future directions**

639 This study is limited by its reliance on DNA-level metagenomic potential, which may not
640 directly correspond to expressed phenotypes. We examined only two time points and did
641 not resolve taxonomic identities in detail. Future metatranscriptomic and isolation-based
642 studies are needed to confirm that the functional shifts are realized at the protein level
643 and to identify the specific bacterial lineages driving these changes. Additionally, while
644 our analysis focused on planktonic communities, sediment and epilithic biofilm
645 communities may harbor distinct functional profiles [64,31]. Nevertheless, the dramatic,
646 coherent functional restructuring we document unveils a fundamental ecological principle:
647 as dry-season stress intensifies, the river microbiome transitions from a growth-oriented,
648 anabolic community to one configured for survival, competition, and chemical defense
649 with the unintended consequence of elevated pathogenicity potential. These findings
650 highlight an urgent need to incorporate functional metagenomic monitoring into water

651 quality assessment, especially in tropical river systems facing escalating anthropogenic
652 and climatic pressures.

653 **5. Conclusions**

654 Our time-series metagenomic analysis of the Shitalakshya River uncovered a coordinated
655 functional shift from primary metabolism to pollutant degradation and stress adaptation,
656 accompanied by a profound remodelling of the virulence gene landscape. Motility, biofilm
657 formation, and T6SS-mediated competition were strongly selected, while iron acquisition
658 and immune evasion traits declined. These changes paralleled a severe deterioration in
659 water quality, linking the microbiome functional signature to the magnitude of organic
660 pollution. The results provide a systems-level understanding of how anthropogenic inputs
661 re-engineer the river microbiome through strong ecological filtering, with direct
662 implications for water quality monitoring and public health risk assessment. Future work
663 integrating metatranscriptomics, isolate genomics, and spatial-temporal sampling will
664 further resolve the active players and mechanisms underlying this dynamic response.

665 **Acknowledgments**

666 The authors thank Dhaka WASA for their valuable assistance with sample collection and
667 logistical support during the fieldwork. This study was conducted as part of the Ph.D.
668 research of M.E.H., who also gratefully acknowledges the University Grants Commission
669 (UGC) of Bangladesh for providing a Ph.D. fellowship.

670 **Funding**

671 The authors received no specific funding for this work.

672

673 **Conflict of interest**

674 All authors declare that they have no conflicts of interest.

675 **Author contributions**

676 M.E.H.: Conceptualization, Investigation, Formal analysis, Writing—original draft.

677 M.S.R.: Methodology, Bioinformatics analysis, Writing—review & editing.

678 M.S.: Validation, Supervision, Writing—review & editing.

679 A.B.: Conceptualization, Supervision, Writing—review & editing.

680 **Ethics statement**

681 This article does not contain any studies with human participants or animals performed
682 by any of the authors.

683 **Data availability**

684 The datasets generated during this study are available in the NCBI SRA repository under
685 BioProject PRJNA1424261 (<https://www.ncbi.nlm.nih.gov/bioproject/PRJNA1424261>)
686 (Accession: SAMN55380971 and SAMN55380975); S1 Dataset includes KEGG
687 pathways, VFGs Abundance, and sequencing quality metrics (trimming, host filtering, and
688 Kraken2-based microbial classification). All other relevant data are within the manuscript
689 and its supporting information files.

690 **ORCID**

691 Muhammad Ehteshamul Haque; ID: <https://orcid.org/0009-0001-0307-3243>

692 Md. Shaminur Rahman; ID: <https://orcid.org/0000-0002-3039-9337>

693 Dr. Munawar Sultana; ID: <https://orcid.org/0000-0002-8563-3661>

694 Dr. Anowara Begum; ID: <https://orcid.org/0000-0001-9915-3311>

695 **Correspondence** and requests for materials should be addressed to **A. Begum**.

696 **References**

- 697 1. Abraham WR. Megacities as sources for pathogenic bacteria in rivers and their
698 fate downstream. *Int J Microbiol.* 2011;2011:798292. doi: 10.1155/2011/798292
- 699 2. Mittal P, Prasoodanan PKV, Dhakan DB, Kumar S, Sharma VK. Metagenome of a
700 polluted river reveals a reservoir of metabolic and antibiotic resistance genes.
701 *Environ Microbiome.* 2019;14(1):5. doi: 10.1186/s40793-019-0345-3
- 702 3. Reddy B, Dubey SK. River Ganges water as reservoir of microbes with antibiotic
703 and metal ion resistance genes: High throughput metagenomic approach. *Environ*
704 *Pollut.* 2019;246:443-451. doi: 10.1016/j.envpol.2018.12.022
- 705 4. Serajuddin M, Chowdhury MAI, Haque ME. Assessing seasonal water quality
706 variations using water quality index (WQI): A case study. *Scholars J Eng Technol.*
707 2019;7(8):245-251. doi: 10.36347/sjet.2019.v07i08.007
- 708 5. Bussi G, Shawal S, Hossain MA, Whitehead PG, Jin L. Multibranch modelling of
709 flow and water quality in the Dhaka River system, Bangladesh: Impacts of future
710 development plans and climate change. *Water.* 2023;15(17):3027. doi:
711 10.3390/w15173027
- 712 6. Haque ME, Kibria MG, Beauty SA, Sultana M, Akter S. Moving bed biofilm reactor
713 (MBBR) technology for removal of emerging contaminants. In: Proceedings of the
714 International Water Conference; 2025 Mar 19-20; Colombo, Sri Lanka.
- 715 7. Jaiswal D, Pandey J. Anthropogenically enhanced sediment oxygen demand
716 creates mosaic of oxygen deficient zones in the Ganga River: Implications for river

- 717 health. *Ecotoxicol Environ Saf.* 2019;171:709-720. doi:
718 10.1016/j.ecoenv.2019.01.039
- 719 8. Parida PK, Behera BK, Dehury B, Rout AK, Sarkar DJ, Rai A, et al. Community
720 structure and function of microbiomes in polluted stretches of river Yamuna in New
721 Delhi, India, using shotgun metagenomics. *Environ Sci Pollut Res.*
722 2022;29(47):71311-71325. doi: 10.1007/s11356-022-20766-1
- 723 9. Morote-Sánchez B, Colomer J, Soler M, Gilabert J, Rodellas V, Garcia-Orellana J,
724 et al. Aquatic deoxygenation associated with resuspension of anthropogenic
725 organic matter. *Water Res.* 2025;278:123327. doi: 10.1016/j.watres.2025.123327
- 726 10. Behera BK, Sahu P, Rout AK, Parida PK, Sarkar DJ, Kaushik NK, et al. Exploring
727 microbiome from sediments of River Ganga using a metagenomic approach. *Aquat*
728 *Ecosyst Health Manag.* 2021;24(4):12-22. doi: 10.14321/ae hm.024.04.04
- 729 11. Cheng H, Cheng L, Wang L, Zhu T, Cai W, Hua Z, et al. Changes of bacterial
730 communities in response to prolonged hydrodynamic disturbances in the eutrophic
731 water-sediment systems. *Int J Environ Res Public Health.* 2019;16(20):3868. doi:
732 10.3390/ijerph16203868
- 733 12. Behera BK, Chakraborty HJ, Patra B, Rout AK, Dehury B, Das BK, et al.
734 Metagenomic analysis reveals bacterial and fungal diversity and their
735 bioremediation potential from sediments of River Ganga and Yamuna in India.
736 *Front Microbiol.* 2020;11:556136. doi: 10.3389/fmicb.2020.556136
- 737 13. Marathe NP, Pal C, Gaikwad SS, Jonsson V, Kristiansson E, Larsson DGJ.
738 Untreated urban waste contaminates Indian river sediments with resistance genes

- 739 to last-resort antibiotics. *Water Res.* 2017;124:388-397. doi:
740 10.1016/j.watres.2017.07.060
- 741 14. Yadav R, Rajput V, Dharne M. Functional metagenomic landscape of polluted river
742 reveals potential genes involved in degradation of xenobiotic pollutants. *Environ*
743 *Res.* 2021;192:110332. doi: 10.1016/j.envres.2020.110332
- 744 15. Edirisinghe JN, Weisenhorn P, Conrad N, Xia F, Overbeek R, Stevens RL, et al.
745 Modeling central metabolism and energy biosynthesis across microbial life. *BMC*
746 *Genomics.* 2016;17:568. doi: 10.1186/s12864-016-2887-8
- 747 16. Brooks AN, Turkarslan S, Beer KD, Lo FY, Baliga NS. Adaptation of cells to new
748 environments. *Wiley Interdiscip Rev Syst Biol Med.* 2011;3(5):544-561. doi:
749 10.1002/wsbm.136
- 750 17. Mitchell A, Romano GH, Groisman B, Yona A, Dekel E, Kupiec M, et al. Adaptive
751 prediction of environmental changes by microorganisms. *Nature.*
752 2009;460(7252):220-224. doi: 10.1038/nature08112
- 753 18. Srivastava A, Verma D. Ganga River sediments of India predominate with aerobic
754 and chemo-heterotrophic bacteria majorly engaged in the degradation of
755 xenobiotic compounds. *Environ Sci Pollut Res.* 2023;30(3):752-772. doi:
756 10.1007/s11356-022-22198-3
- 757 19. Johnson J, Jain KR, Patel A, Parmar N, Joshi C, Madamwar D. Chronic industrial
758 perturbation and seasonal change induces shift in the bacterial community from
759 Gammaproteobacteria to Betaproteobacteria having catabolic potential for
760 aromatic compounds at Amlakhadi canal. *World J Microbiol Biotechnol.*
761 2024;40(2):52. doi: 10.1007/s11274-023-03848-1

- 762 20. Rout AK, Dehury B, Parida PK, Sarkar DJ, Behera B, Das BK, et al. Taxonomic
763 profiling and functional gene annotation of microbial communities in sediment of
764 River Ganga at Kanpur, India: Insights from whole-genome metagenomics study.
765 Environ Sci Pollut Res. 2022;29(54):82309-82323. doi: 10.1007/s11356-022-
766 21644-6
- 767 21. Coulthurst SJ. The Type VI secretion system—A widespread and versatile cell
768 targeting system. Res Microbiol. 2013;164(6):640-654. doi:
769 10.1016/j.resmic.2013.03.017
- 770 22. Ho BT, Dong TG, Mekalanos JJ. A view to a kill: The bacterial type VI secretion
771 system. Cell Host Microbe. 2014;15(1):9-21. doi: 10.1016/j.chom.2013.11.008
- 772 23. Roges EM, Gonçalves VD, Cardoso MD, Festivo ML, Siciliano S, Berto LH, et al.
773 Virulence-associated genes and antimicrobial resistance of *Aeromonas hydrophila*
774 isolates from animal, food, and human sources in Brazil. Biomed Res Int.
775 2020;2020:1052607. doi: 10.1155/2020/1052607
- 776 24. Rout AK, Tripathy PS, Dixit S, Behera DU, Behera B, Das BK, et al. Unveiling the
777 microbiome landscape: A metagenomic study of bacterial diversity, antibiotic
778 resistance, and virulence factors in the sediments of the River Ganga, India.
779 Antibiotics. 2023;12(12):1735. doi: 10.3390/antibiotics12121735
- 780 25. Shayan MNM, Tanaka Y, Shinji A, Nakayama T, Hirano R, Kuroda K, et al.
781 Resuscitation, regrowth, and release of bacteria in biofilm in a river receiving
782 chlorinated reclaimed water. Water Res X. 2026;31:100524. doi:
783 10.1016/j.wroa.2026.100524

- 784 26. Mao G, Liang J, Wang Q, Zhao C, Bai Y, Liu R, et al. Epilithic biofilm as a reservoir
785 for functional virulence factors in wastewater-dominant rivers after WWTP
786 upgrade. *J Environ Sci (China)*. 2021;101:27-35. doi: 10.1016/j.jes.2020.05.014
- 787 27. Zou Y, Xiao Z, Wang L, Wang Y, Yin H, Li Y. Prevalence of antibiotic resistance
788 genes and virulence factors in the sediment of WWTP effluent-dominated rivers.
789 *Sci Total Environ*. 2023;897:165441. doi: 10.1016/j.scitotenv.2023.165441
- 790 28. Sahoo S, Sahoo RK, Gaur M, Behera DU, Sahu A, Das A, et al. Urban wastewater
791 contributes to the emergence of carbapenem-resistant *Klebsiella pneumoniae*
792 (CRKP) in an urban receiving river in eastern India. *Lett Appl Microbiol*.
793 2023;76(1):ovac005. doi: 10.1093/lambio/ovac005
- 794 29. Wang J, Wang Y, Lou H, Wang W. AlgU controls environmental stress adaptation,
795 biofilm formation, motility, pyochelin synthesis and antagonism potential in
796 *Pseudomonas protegens* SN15-2. *Microbiol Res*. 2023;272:127396. doi:
797 10.1016/j.micres.2023.127396
- 798 30. Parwin N, Dixit S, Das S, Sahoo RK, Subudhi E. Metagenomic analysis of
799 microbiome spatial dynamics in urban river confluence affected by city wastewater.
800 *Genomics Inform*. 2025;23:27. doi: 10.1186/s44342-025-00054-3
- 801 31. Parwin N, Dixit S, Subudhi E. From city waste to river risk: tracking AMR and
802 pathogenic potential in urban water gradient of the Kathajodi river. *Environ Sci Eur*.
803 2026;38:64. doi: 10.1186/s12302-026-01347-2
- 804 32. Xie Y, Liu X, Wei H, Chen X, Gong N, Ahmad S, et al. Insight into impact of sewage
805 discharge on microbial dynamics and pathogenicity in river ecosystem. *Sci Rep*.
806 2022;12(1):6894. doi: 10.1038/s41598-022-09579-x

- 807 33. Liu B, Zheng D, Jin Q, Chen L, Yang J. VFDB 2019: A comparative pathogenomic
808 platform with an interactive web interface. *Nucleic Acids Res.* 2019;47(D1):D687-
809 D692. doi: 10.1093/nar/gky1080
- 810 34. Rice EW, Baird RB, Eaton AD, Clesceri LS, editors. Standard methods for the
811 examination of water and wastewater. 22nd ed. Washington (DC): American Public
812 Health Association, American Water Works Association, Water Environment
813 Federation; 2012.
- 814 35. Jones A, Stanley D, Ferguson S, Schwessinger B, Borevitz J, Warthmann N. Cost-
815 conscious generation of multiplexed short-read DNA libraries for whole-genome
816 sequencing. *PLoS ONE.* 2023;18(1):e0280004. doi:
817 10.1371/journal.pone.0280004
- 818 36. Andrews S. FastQC: a quality control tool for high-throughput sequence data
819 [software]. Cambridge (UK): Babraham Institute; 2010. Available
820 from: <https://www.bioinformatics.babraham.ac.uk/projects/fastqc/>
- 821 37. Bolger AM, Lohse M, Usadel B. Trimmomatic: a flexible trimmer for Illumina
822 sequence data. *Bioinformatics.* 2014;30(15):2114-2120. doi:
823 10.1093/bioinformatics/btu170
- 824 38. Bushnell B. BBMap: a fast, accurate, splice-aware aligner [software].
825 SourceForge; 2014.
- 826 39. Wood DE, Lu J, Langmead B. Improved metagenomic analysis with Kraken 2.
827 *Genome Biol.* 2019;20(1):257. doi: 10.1186/s13059-019-1891-0

- 828 40. Lu J, Breitwieser FP, Thielen P, Salzberg SL. Bracken: estimating species
829 abundance in metagenomics data. PeerJ Comput Sci. 2017;3:e104. doi:
830 10.7717/peerj-cs.104
- 831 41. Shen W, Ren H. TaxonKit: a practical and efficient NCBI taxonomy toolkit. J Genet
832 Genomics. 2021;48(9):844-850. doi: 10.1016/j.jgg.2021.03.006
- 833 42. Bonin N, Doster E, Worley H, Pinnell LJ, Bravo JE, Ferm P, et al. MEGARes and
834 AMR++, v3.0: an updated comprehensive database of antimicrobial resistance
835 determinants. Nucleic Acids Res. 2023;51(D1):D744-D752. doi:
836 10.1093/nar/gkac1043
- 837 43. Chen L, Yang J, Yu J, Yao Z, Sun L, Shen Y, et al. VFDB: a reference database
838 for bacterial virulence factors. Nucleic Acids Res. 2005;33(Database issue):D325-
839 D328. doi: 10.1093/nar/gki008
- 840 44. Beghini F, McIver LJ, Blanco-Míguez A, Dubois L, Asnicar F, Maharjan S, et al.
841 Integrating taxonomic, functional, and strain-level profiling of diverse microbial
842 communities with bioBakery 3. eLife. 2021;10:e65088. doi: 10.7554/eLife.65088
- 843 45. Oksanen J, Blanchet FG, Kindt R, Legendre P, Minchin PR, O'Hara RB, et al.
844 vegan: community ecology package (Version 2.0-10) [R package]. CRAN. 2013.
845 Available from: <https://CRAN.R-project.org/package=vegan>
- 846 46. Gu Z, Gu L, Eils R, Schlesner M, Brors B. circlize: circular visualization in R.
847 Bioinformatics. 2014;30(19):2811-2812. doi: 10.1093/bioinformatics/btu393
- 848 47. Rout AK, Tripathy PS, Dixit S, Behera DU, Behera B, Das BK, et al. Metagenomics
849 analysis of sediments of river Ganga, India for bacterial diversity, functional

- 850 genomics, antibiotic resistant genes and virulence factors. *Curr Res Biotechnol.*
851 2024;7:100187. doi: 10.1016/j.crbiot.2024.100187
- 852 48. Rout AK, Rout SS, Panda A, Tripathy PS, Kumar N, Parida SN, et al. Potential
853 applications and future prospects of metagenomics in aquatic ecosystems. *Gene.*
854 2025;967:149720. doi: 10.1016/j.gene.2025.149720
- 855 49. Behera BK, Patra B, Chakraborty HJ, Sahu P, Rout AK, Sarkar DJ, et al.
856 Metagenome analysis from the sediment of river Ganga and Yamuna: In search of
857 beneficial microbiome. *PLoS ONE.* 2020;15(10):e0239594. doi:
858 10.1371/journal.pone.0239594
- 859 50. Rout AK, Dixit S, Tripathy PS, Rout SS, Parida SN, Parida PK, et al. Metagenomic
860 landscape of sediments of river Ganga reveals microbial diversity, potential plastic
861 and xenobiotic degradation enzymes. *J Hazard Mater.* 2024;471:134377. doi:
862 10.1016/j.jhazmat.2024.134377
- 863 51. Potrykus K, Cashel M. (p)ppGpp: still magical? *Annu Rev Microbiol.* 2008;62:35-
864 51. doi: 10.1146/annurev.micro.62.081307.162903
- 865 52. Battesti A, Majdalani N, Gottesman S. The RpoS-mediated general stress
866 response in *Escherichia coli*. *Annu Rev Microbiol.* 2011;65:189-213. doi:
867 10.1146/annurev-micro-090110-102946
- 868 53. Behera BK, Patra B, Chakraborty HJ, Rout AK, Dixit S, Rai A, et al. Bacteriophages
869 diversity in India's major river Ganga: A repository to regulate pathogenic bacteria
870 in the aquatic environment. *Environ Sci Pollut Res.* 2023;30(12):34101-34114. doi:
871 10.1007/s11356-022-24637-7

- 872 54. Li Y, Jing H, Kao SJ, Zhang W, Liu H. Metabolic response of prokaryotic microbes
873 to sporadic hypoxia in a eutrophic subtropical estuary. *Mar Pollut Bull.*
874 2020;154:111064. doi: 10.1016/j.marpolbul.2020.111064
- 875 55. Colin R, Ni B, Laganenka L, Sourjik V. Multiple functions of flagellar motility and
876 chemotaxis in bacterial physiology. *FEMS Microbiol Rev.* 2021;45(6):fuab038. doi:
877 10.1093/femsre/fuab038
- 878 56. Yin W, Wang Y, Liu L, He J. Biofilms: the microbial "protective clothing" in extreme
879 environments. *Int J Mol Sci.* 2019;20(14):3423. doi: 10.3390/ijms20143423
- 880 57. Serajuddin M, Chowdhury MAI, Haque MM, Haque ME. Using turbidity to
881 determine total suspended solids in an urban stream: a case study. *Int J Eng*
882 *Trends Technol.* 2019;67(9):83-88. doi: 10.14445/22315381/IJETT-V67I9P214
- 883 58. Sivadon P, Barnier C, Urios L, Grimaud R. Biofilm formation as a microbial strategy
884 to assimilate particulate substrates. *Environ Microbiol Rep.* 2019;11(6):749-764.
885 doi: 10.1111/1758-2229.12785
- 886 59. Zhang H, Chen G, Sun Z, Zhang X, Jiang Z, Zhang X. The microstructures of
887 *Tetrademus obliquus* biofilms formed under different nutrient supplies. *Appl Sci.*
888 2025;15(9):5104. doi: 10.3390/app15095104
- 889 60. Gallique M, Bouteiller M, Merieau A. The Type VI Secretion System: a dynamic
890 system for bacterial communication? *Front Microbiol.* 2017;8:1454. doi:
891 10.3389/fmicb.2017.01454
- 892 61. Gonzalez Moreno PJ, Nishiguchi MK. The competitive edge: T6SS-mediated
893 interference competition by *Vibrionaceae* across marine ecological niches.
894 *Microorganisms.* 2025;13(6):1370. doi: 10.3390/microorganisms13061370

- 895 62. Galán JE, Lara-Tejero M, Marlovits TC, Wagner S. Bacterial type III secretion
896 systems: Specialized nanomachines for protein delivery into target cells. *Annu Rev*
897 *Microbiol.* 2014;68:415-438. doi: 10.1146/annurev-micro-092412-155725
- 898 63. Webb SA, Kahler CM. Bench-to-bedside review: Bacterial virulence and
899 subversion of host defences. *Crit Care.* 2008;12(6):234. doi: 10.1186/cc7091
- 900 64. Lawrence JR, Chenier MR, Roy R, Beaumier D, Fortin N, Swerhone GD, et al.
901 Microscale and molecular assessment of impacts of nickel, nutrients, and oxygen
902 level on structure and function of river biofilm communities. *Appl Environ Microbiol.*
903 2004;70(7):4326-4339. doi: 10.1128/AEM.70.7.4326-4339.2004
- 904 65. Nakamura S, Minamino T. Flagella-driven motility of bacteria. *Biomolecules.*
905 2019;9(7):279. doi: 10.3390/biom9070279

906 **Supporting information**

907 S1 Dataset. Sequencing quality control metrics and KEGG pathway abundances.
908 Includes trimming efficiency, host-read removal statistics, Kraken2-based
909 microbial classification, and complete KEGG pathway relative abundances for both
910 time points. (Provide filename: S1 Dataset.xlsx)

911

912 S1 Table. Physicochemical parameters at the Sarulia intake point during the early
913 dry season (December 2024) and peak dry season (February 2025). (Provide
914 filename: S1 Table.docx)

915

916 S2 Table. Key Xenobiotic-Degrading Bacteria Enriched in February 2025 (Provide
917 filename: S2 Table.docx)

918

919 S3 Table. Disease-Associated Microorganisms Enriched from December 2024 to
920 February 2025. (Provide filename: S3 Table.docx)

KEGG Functional Category Relative Abundance Shitalakshya River – December 2024 vs February 2025

Sampling Period ■ Dec. 2024 ■ Feb. 2025

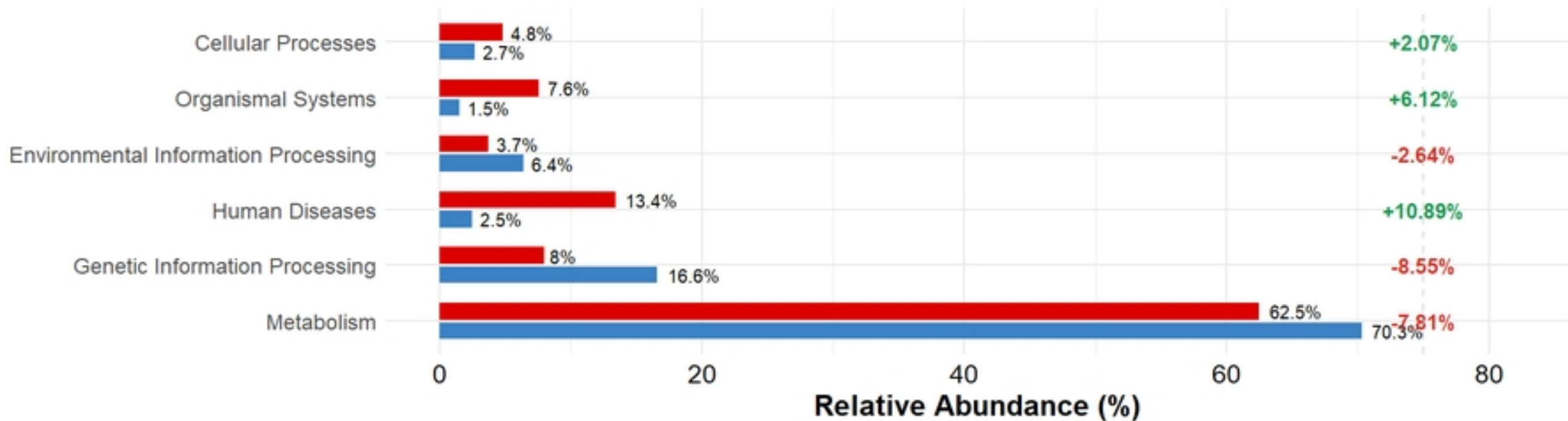


Fig 1

Key Subclass Changes (Dec 2024 → Feb 2025)

Functional shift from growth metabolism to stress adaptation

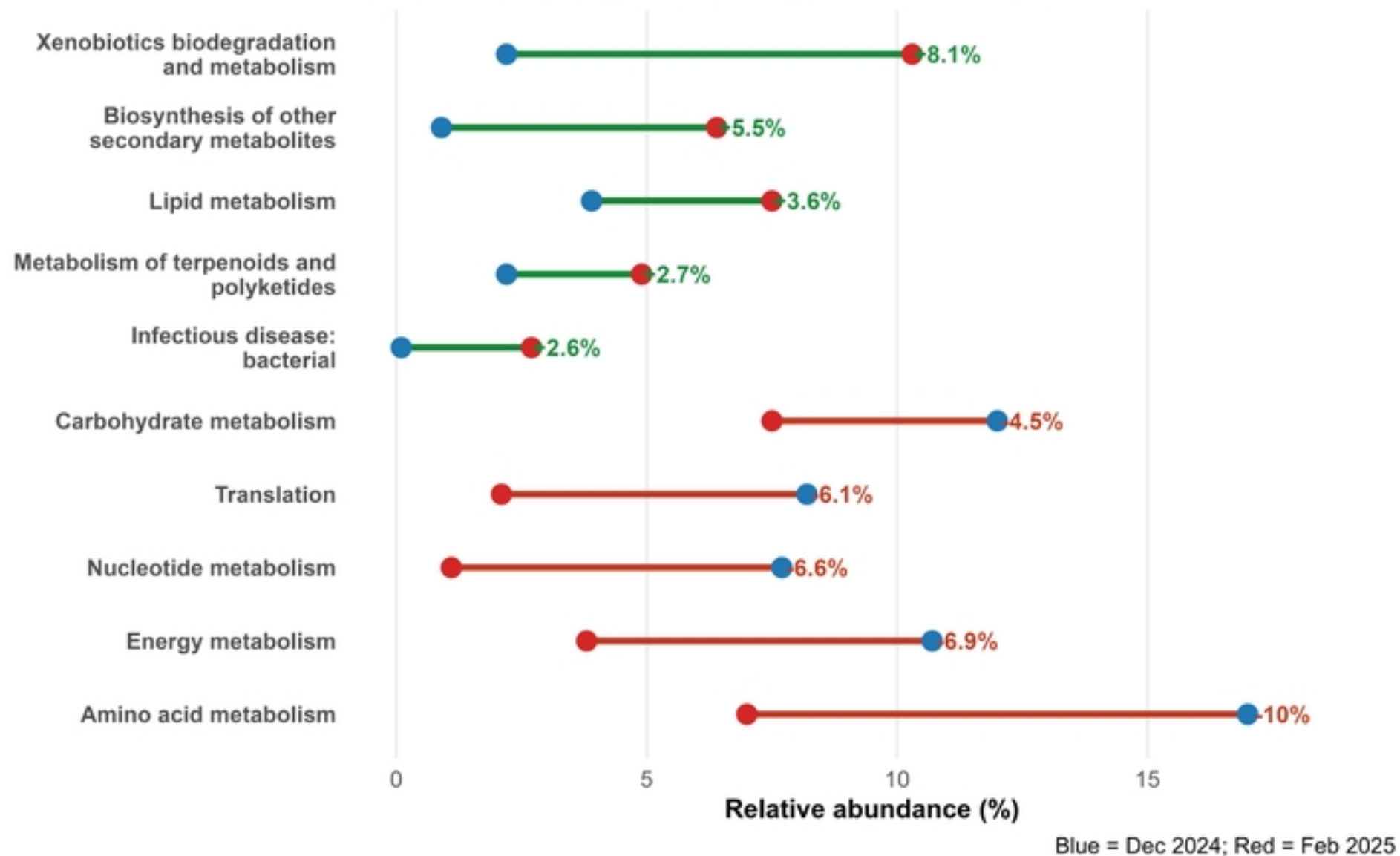


Fig 2

Pathway Importance and Change Analysis

Size indicates biological importance (abundance × change magnitude)

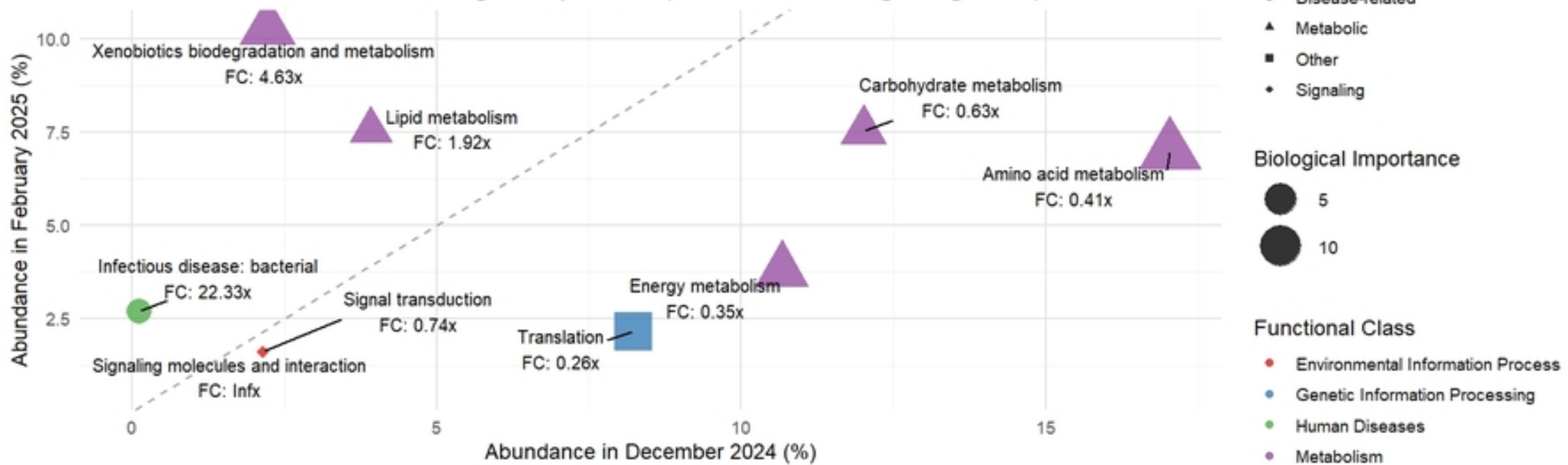
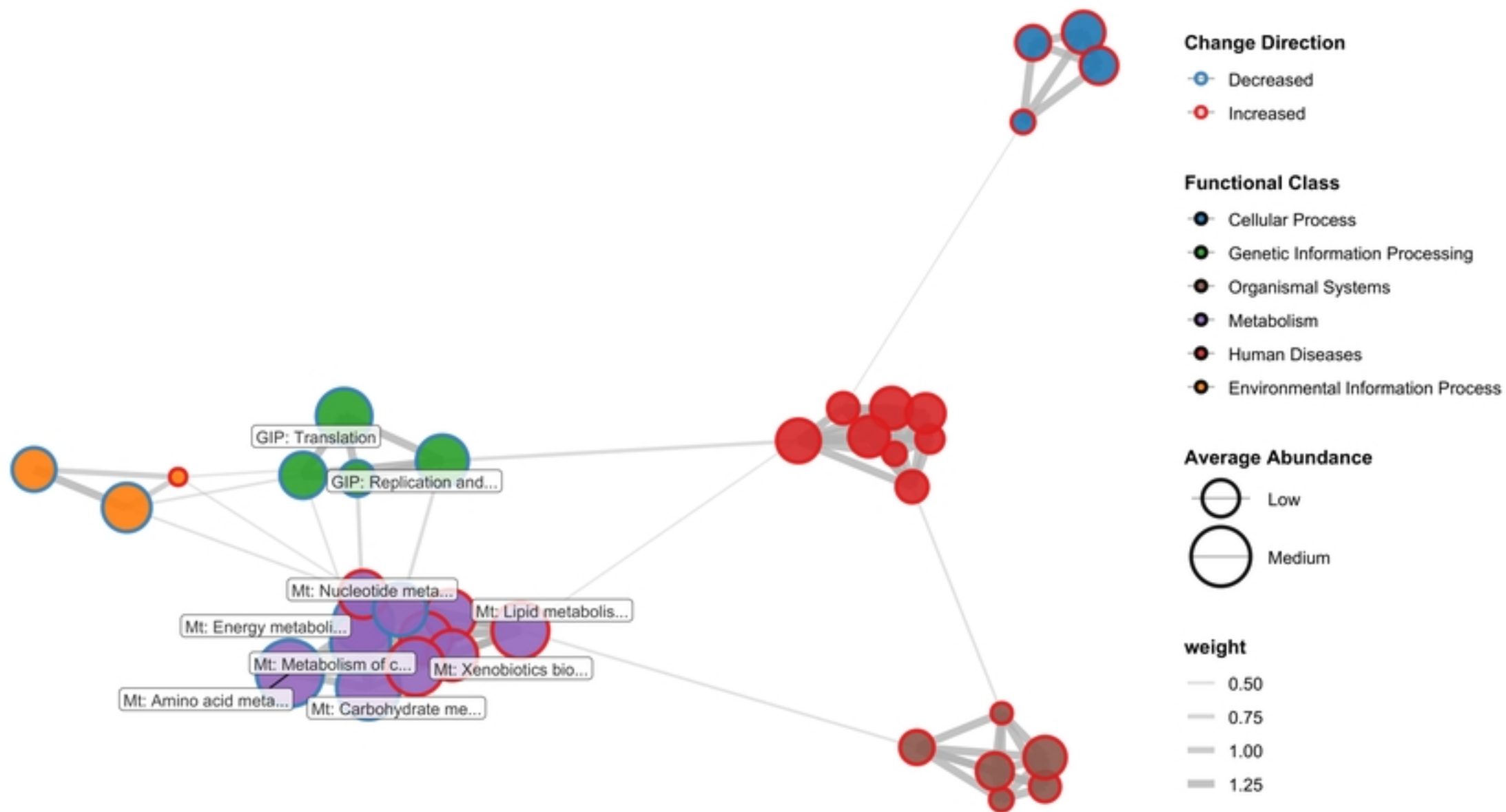


Fig 3

KEGG Pathway Interaction Network

Shitalakshya River Microbiome: December 2024 - February 2025



Node size: Average pathway abundance
Node fill: Functional class
Node border: Change direction (red=increased, blue=decreased)
Edge thickness/transparency: Connection strength

Fig 4

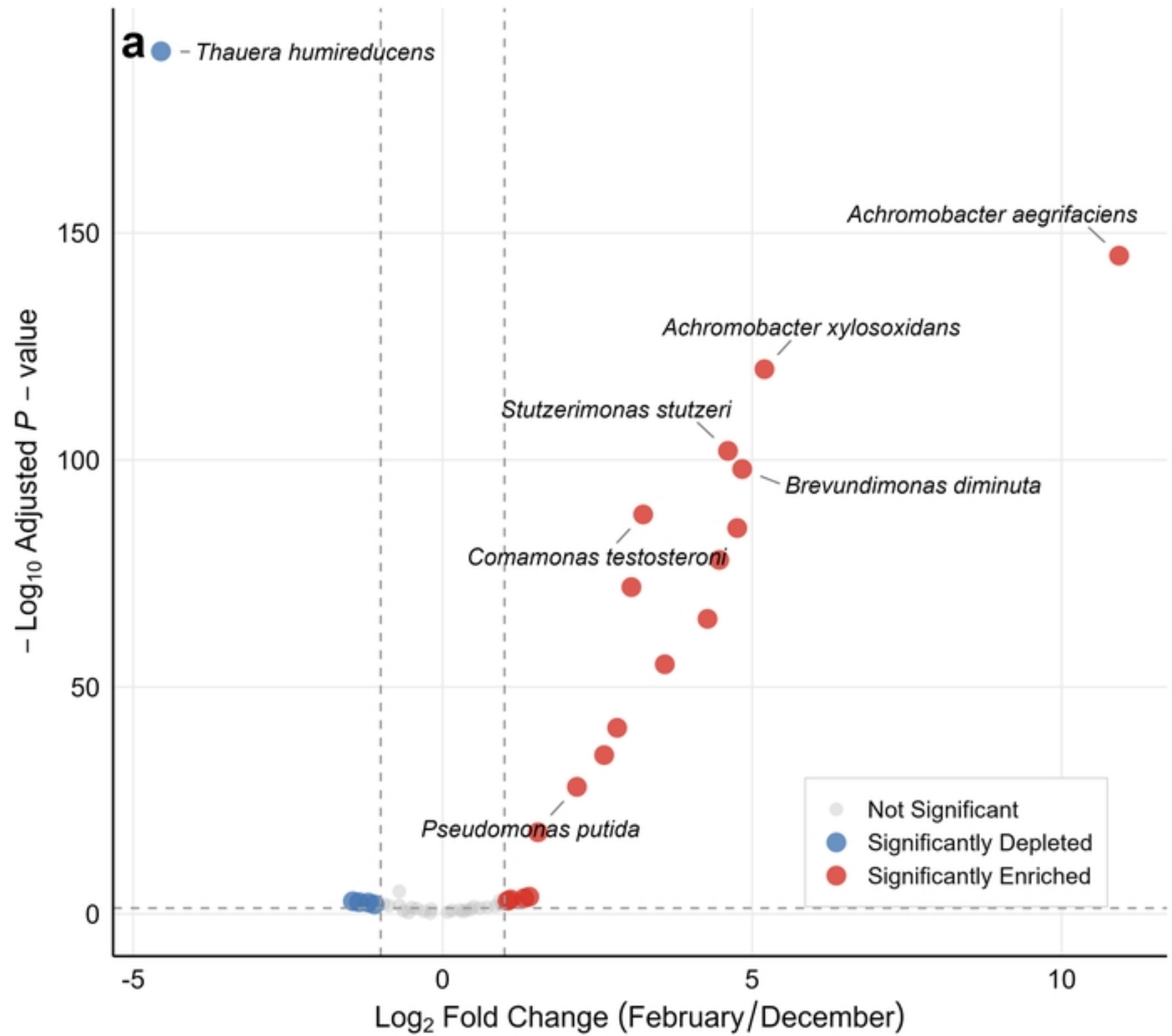


Fig 5a

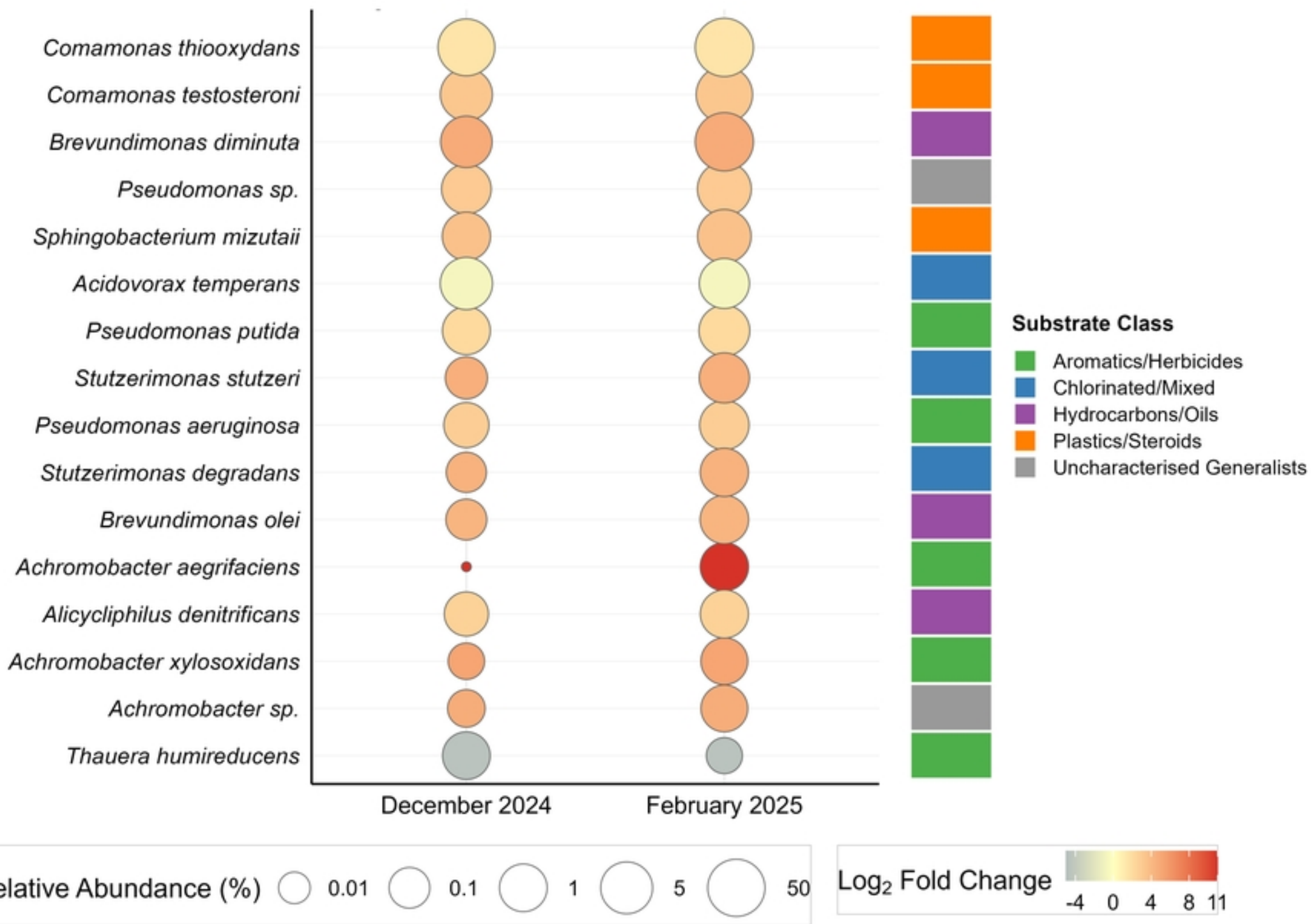
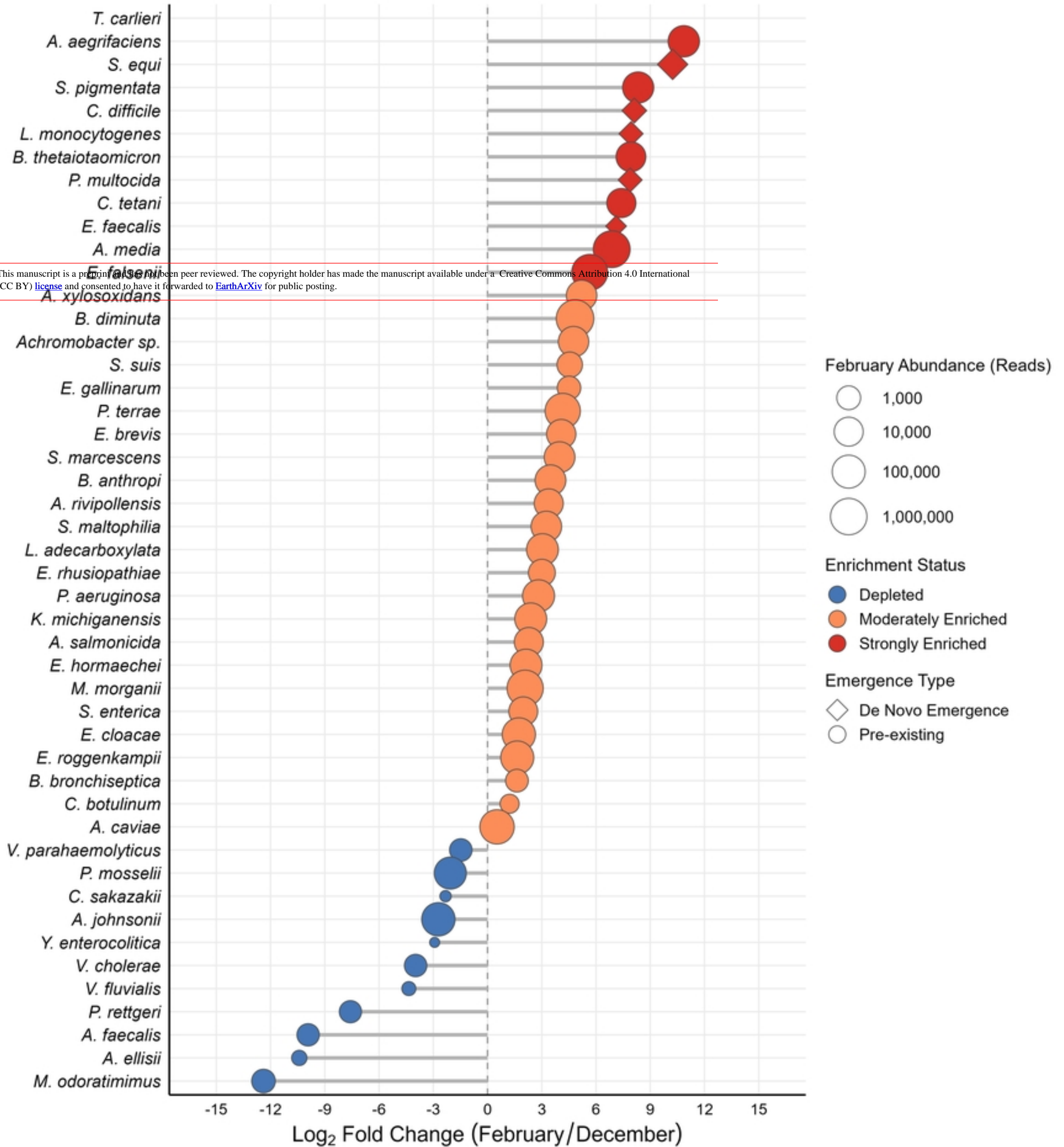


Fig 5b

a**Fig 6**

Stacked Bar Chart of Virulence Factor Abundance

December 2024 vs February 2025

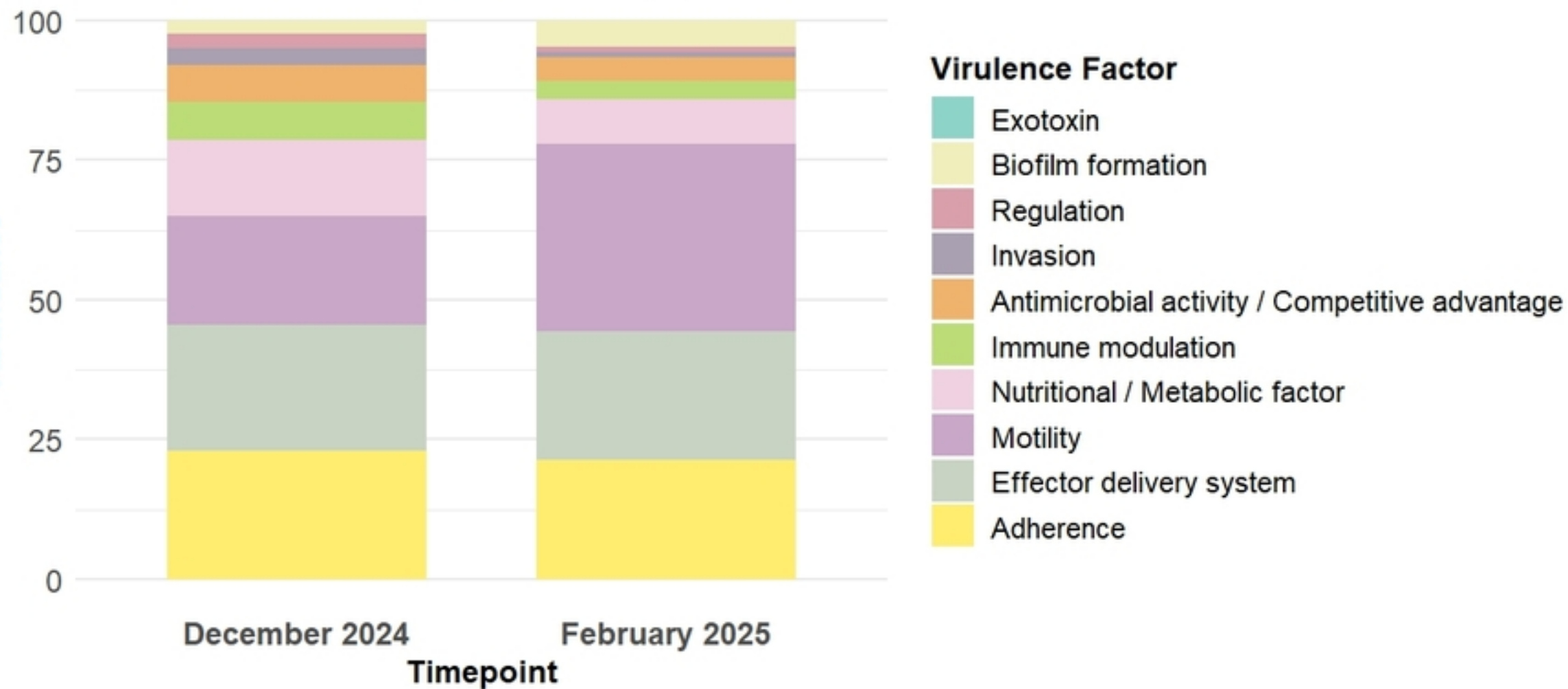


Fig 7

Virulence factor abundance by functional category (Δ change)

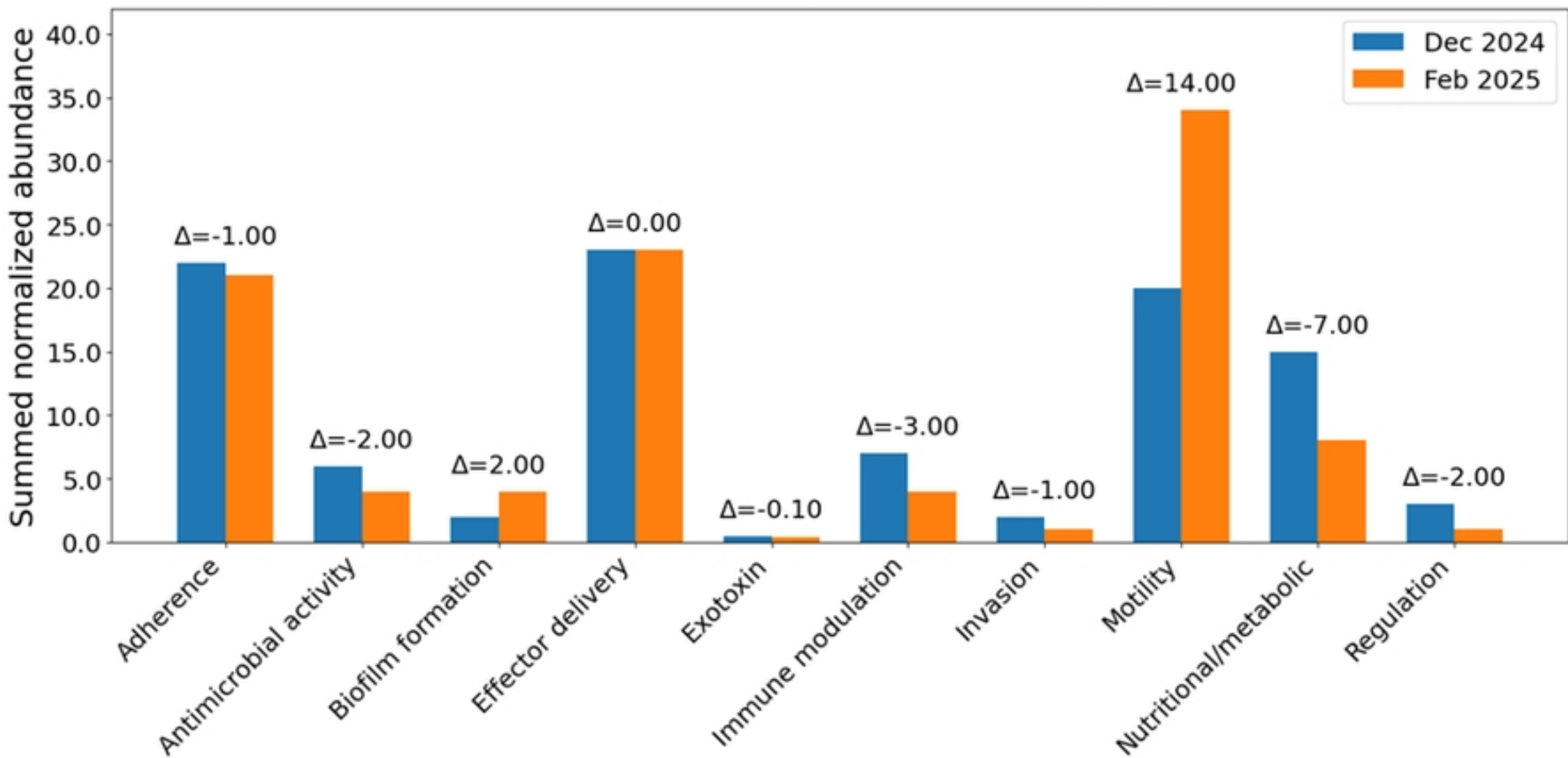


Fig 8

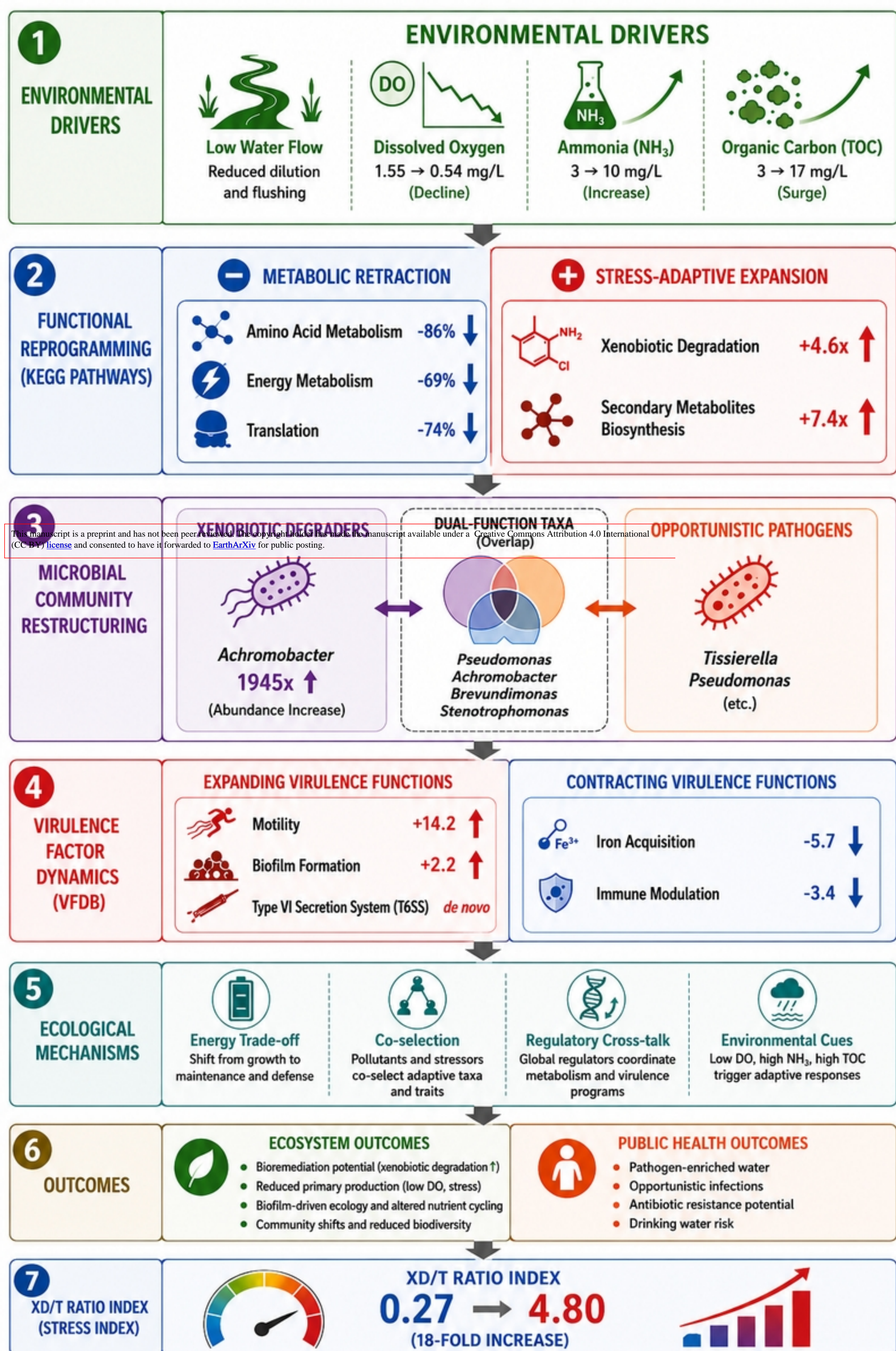


Fig 9