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29	Estimating Microbial Hydrogen Consumption in
30	Hydrogen Storage in Porous Media as a Basis for
31	Site Selection
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45 ABSTRACT

Subsurface storage of hydrogen, e.g. in depleted gas or oil fields (DOGF), is suggested 46 47 as a means to overcome imbalances between supply and demand in the renewable energy 48 sector. However, hydrogen is an electron donor for subsurface microbial processes, 49 which may have important implications for hydrogen recovery, gas injectivity and 50 corrosion. Here, we review the controls on the three major hydrogen consuming 51 processes in the subsurface, methanogenesis, homoacetogenesis, and sulfate reduction, as a basis to develop a hydrogen storage site selection tool. Testing our tool on 42 DOGF 52 53 showed that seven of the fields may be considered sterile with respect to hydrogen-54 consuming microbiota due to either temperatures >122 °C or salinities >5 M NaCl. Only 55 three fields can sustain all of the hydrogen consuming processes, due to either temperature, salinity or pressure constraints in the remaining fields. We calculated a 56 57 potential microbial growth in the order of 1-17*10⁷ cells ml⁻¹ for these fields. The associated hydrogen consumption is negligible to small (<0.01-3.2 % of the stored 58 59 hydrogen). Our results can help inform decisions about where hydrogen will be stored in the future. 60

61 Keywords: Hydrogen, underground storage, microbial hydrogen consumption,

- 62 homoacetogens, methanogens, sulfate reducers
- 63

64 Highlights

- Review of the most important hydrogen-utilizing microorganisms in the underground.
 Elucidation of the growth criteria for 480 strains of the mayor hydrogen-
- Elucidation of the growth criteria for 480 strains of the mayor hydrogen utilizers.

69		• Development of a site selection tool for sterile hydrogen storage.
70		• Evaluation of the site selection tool on 42 depleted oil and gas fields
71		(DOGF).
72		• Calculation of the microbial growth and hydrogen consumption in DOGF.
73		
74	Abbrev	iations and units
75	SSR	Sulfur species reduction
76	SSRM	Sulfur species reducing microorganisms
77	DOGF	Depleted oil and gas fields
78	EPS	Extracellular polymeric substances
79	М	Molarity (mol L ⁻¹)
80	MPa	Megapascal

81

82 **1. Introduction**

Zero-carbon energy generation from renewable sources can help mitigate carbon emissions and 83 84 abate climate change [1-3]. One of the most significant challenges for renewable energy is the 85 imbalance between supply and demand [3, 4]. The generation of hydrogen (H₂) via electrolysis of water during periods of renewable energy oversupply and subsequent H₂ storage is one way 86 of overcoming this imbalance, as H₂ can be recovered and used for electricity generation during 87 periods of renewable energy shortage [1, 5]. Subsurface storage of H₂ in salt caverns, depleted 88 gas or oil fields or saline aquifers is being considered as an alternative to expensive purpose-89 built storage containers [6]. However, the artificial elevation of the H₂ concentration in the 90 subsurface may stimulate the growth of H2-oxidizing (hydrogenotrophic) bacteria and archaea, 91 here collectively referred to as microorganisms, with possible adverse implications for gas 92

93 injectivity and withdrawal via permeability reduction, H₂ volume loss and corrosion of metal
94 infrastructure [4, 7]. Understanding the controls on microbial H₂ metabolism is therefore highly
95 important.

96 Much of the subsurface is characterized by combinations of elevated temperature [7], high salt 97 concentrations and high pressure [3], reduced void space [8], limited nutrient availability [9] 98 and typically highly reducing conditions [9-11]. The evidence for microbial life at depth is 99 plentiful (e.g. [12-16]). Most microorganisms in nature grow in biofilms attached to surfaces 100 (communities of aggregated microbial cells embedded in a secreted matrix of extracellular 101 polymeric substances (EPS)) [17, 18]. Even small amounts of biofilm can reduce pore throat 102 sizes and increase the flow-path tortuosity, resulting in dramatic decreases in permeability [19]. It has been postulated that biofilms may not form in the nutrient-limited underground when the 103 groundwater flow is low [19]. However, subsurface biofilms are commonly encountered during 104 geoenergy activities such as fracturing, hydrocarbon recovery or in geothermal plants [20-25]. 105 106 Biofilm formation may actually be enhanced under the harsh subsurface conditions as the EPS 107 layer acts as a protective clothing which ensures the normal reproduction and metabolism of 108 microorganisms [18].

109 Hydrogen plays a central role in the energy metabolism of subsurface life [9]. Yet, a 110 quantitative assessment of the consumption of H_2 by deep microbial communities in the context 111 of the global H_2 cycle is lacking [26]. In underground gas storage sites and oil reservoirs the 112 most abundant H_2 -oxidizers are hydrogenotrophic sulfate reducers, that couple H_2 -oxidation to 113 sulfate reduction to produce hydrogen sulfide (H_2 S); hydrogenotrophic methanogens that 114 reduce carbon dioxide (CO₂) to methane (CH₄) by oxidizing H_2 ; and homoacetogens that 115 couple H_2 oxidation to carbon dioxide (CO₂) reduction producing acetate [7, 24, 27]. These three groups of microorganisms are, amongst others, implicated in causing subsurfacecorrosion [7, 27, 28].

118 A recent review addressed the many possible abiotic and biotic H₂-producing and H₂-119 consuming processes in the subsurface [7]. However, it lacked a quantitative assessment of the 120 processes of microbial growth and H₂ consumption relevant for H₂ storage. Strobel et al. [29] 121 summarized the concept and potential of underground methanation using experimental data 122 from the Sun Storage project. These authors highlighted controls on the growth of methanogens and changes in gas composition due to methanogenesis, but did not quantify microbial growth. 123 124 Many studies report changes in gas composition, biofilm growth and clogging near injection 125 wells but hardly any studies report quantitative figures on microbial growth or on permeability changes [30]. 126

127 To date it remains unclear how subsurface microorganisms might react to elevated H₂ 128 concentrations [7] and hence whether microbial growth is a concern for H₂ storage. Even in natural, non-engineered subsurface environments, there is little information on microbial H₂ 129 130 turnover rates [31] and the behavior and population kinetics of microorganisms are not fully 131 understood [29]. The majority of the available data on microbial H₂ turnover rates come from 132 batch cultures at optimal growth conditions where the kinetics [31], the pace of life [32, 33], the physiological states and the prominent organisms may differ widely from the subsurface 133 134 environment [7, 32]. A further complication arises from the fact that many microorganisms in the deep subsurface are not culturable with modern enrichment techniques [12, 34]. 135

In this work, we review the state-of-the-art understanding of the controls of temperature, salinity, pH, pressure and nutrients and water on microbial growth on H_2 in the subsurface, with emphasis on the three major H_2 -consuming processes methanogenesis, sulfate reduction and homoacetogenesis, to determine what reservoir conditions will be unfavorable to microbial 140 activity and as such more suitable sites for long term gas storage operations of 30 years or141 longer, such as the UK Rough gas storage site.

Physicochemical data from 42 depleted or close to depleted oil and gas fields (DOGF) of the British and Norwegian North Sea and the Irish Sea as well as five H₂ storage test sites provide the base for an evaluation of the number of sites where microbial growth of methanogens, sulfate reducers and homoacetogens can be expected. Using average nutrient contents of the microbial cells and site-specific dissolved ion concentrations, we calculate significant growth and a small H₂ consumption for growth-permitting DOGF.

148 **2. State of the art understanding**

149 **2.1** Likely microbial hydrogen oxidation in hydrogen storage systems

Hydrogen oxidizing processes may be ranked according to the magnitude of their H₂ threshold and their standard free energy change ($\Delta G^{0'}$), two useful metrics to compare the likelihood of reactions to take place and the order at which they proceed (Table 1). The H₂ threshold defines the concentration of H₂ below which it is no longer consumed. Given all other factors are at optimum, the microbial population with the lowest H₂ threshold value is expected to be the most successful population in competing for H₂ [35].

156 The $\Delta G^{0'}$ marks the thermodynamic favorability of a reaction at ambient pressure and 157 temperature, pH 7 and 1 M of all reactants. In oligotrophic (nutrient poor) high pressure and 158 temperature environments, the order of the $\Delta G^{0'}$ may be used to determine which reaction is 159 more energetically favorable. As can be seen from Table 1, more negative $\Delta G^{0'}$ values (more 160 available free energy) are generally accompanied by lower H₂ thresholds. Not included in Table 161 1 are the kinetics which describe the rate of the electron transfer in the redox reaction.

Table 1. Biotic H₂-consuming processes ranked according to their free energy yield ($\Delta G^{0'}$) and measured H₂ threshold. Not included are

163	Vanadium, Cobalt,	Techneticum,	Uranium and Se	elenium reduction,	due their limited	l relevance for H ₂	2 storage. NA=	not available.
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H ₂ - oxidizing process	Reaction (number)		H ₂ threshold (nM)	ΔG ⁰ ' (KJ mol H2 ⁻¹)	Typical ambient [H2] (nmol L ⁻¹)	Relevance for H ₂ storage	
Chromate reduction	$\frac{1}{2}H_2 + \frac{1}{3}CrO_4^{2-} + \frac{5}{3}H^+ \to \frac{1}{3}Cr^{3+} + \frac{4}{3}H_2O$	(1)	<0.1 ^[36]	NA	NA	low	
Aerobic hydrogen oxidation (Knallgas)	$H_2 + \frac{1}{2}O_2 \to H_2O$	(2)	0.051 ^[7]	-237 ^[7, 36]	NA	low	
Denitrification	$H_2 + \frac{2}{5}H^+ + \frac{2}{5}NO_3^- \rightarrow \frac{1}{5}N_2 + \frac{6}{5}H_2O$	(3)	<0.05-0.5 ^[7]	-240.1 ^[7, 36] -224 ^[4, 37]	< 0.05 ^[4, 35, 36]	low	
Halorespiration	H_2 + halogenated compounds \rightarrow dehalogentated compounds + HCl	(4)	$\begin{array}{c} 0.05 \hbox{-} 0.27^{[36]} \\ < 0.3 \ ^{[38]} \\ 0.27 \hbox{-} 2^{[7]} \end{array}$	-230 to -187 ^[7]	NA	low	
Iron (III) reduction	$H_2 + ferric(oxy)hydroxides \rightarrow ferrous ir H_2O$	(5)	<0.11-0.8 ^[36, 38]	-228.3 ^[7, 38] -182.5 ^[36] -114 ^[4]	$\begin{array}{c} 0.2^{[4, 35]} \\ 0.2 \text{-} 1^{[36]} \end{array}$	intermediate	
Manganese (IV) reduction	$2H_2 + MnO_2 \rightarrow Mn(OH)_2 + 2H_2O$	(6)	< 0.05 ^[35]	-163 ^[4, 35]	< 0.05 ^[4, 35]	low	
Arsenate reduction	$H_2 + HAsO_4^{2-} + 2H^+ \rightarrow H_3AsO_3 + H_2O$	(7)	0.03-0.09 ^[36]	-162.4 ^[36]	0.4-0.7 ^[36]	low	
Ammonification	$4H_2 + 2H^+ + NO_3^- \to NH_4^+ + 3H_2O$	(8)	$0.015 - 0.025^{[38,}_{39]}$	-150 ^[4, 38]	< 0.05 ^[4, 35]	low	
Fumarate reduction	$H_2 + fumarate \rightarrow succinate$	(9)	0.015 ^[38, 39]	-86.2 ^[38]	NA	low	
Hydrogenotrophic sulfate reduction	$4H_2 + SO_4^{2-} + H^+ \to HS^- + 4H_2O$	(10)	1-15 ^[38, 39]	-38 ^[7, 38] -48 ^[36] -57 ^[4]	1-2 ^[4, 35]	high	
Hydrogenotrophic methanogenesis	$4H_2 + CO_2 \rightarrow CH_4 + 2H_2O$	(11)	0.4-95 ^[38-40]	-34 ^[4, 38] -43.9 ^[36]	5-10 ^[4, 35] 7-13 ^[36]	high	
Sulfur reduction	$H_2 + S \rightarrow HS^- + H^+$	(12)	2500 ^[7]	-33.1 ^[7]	NA	intermediate	
Homoacetogenesis	$4H_2 + 2CO_2 \rightarrow CH_3COOH + 2H_2$	(13)	328-3640 ^[38, 39]	-26 ^[4, 38] -36.1 ^[36]	100<[4], 117-150[36]	high	

Abiotically, most of the H₂-oxidizing reactions are very slow but mediated by microbial enzymes the processes are catalyzed [37, 41].

167 The three main microbial processes with implications for H₂ storage, hydrogenotrophic sulfate 168 reduction, hydrogenotrophic methanogenesis (for simplicity from now on just referred to as 169 sulfate reduction, and methanogenesis unless otherwise specified) and homoacetogenesis, 170 require the highest threshold [H₂] and are among the processes with lowest $\Delta G^{0^{\circ}}$ (Table 1). 171 Nevertheless, e.g. sulfate reduction is instantaneous in most geologic settings [42] possibly due 172 to fast kinetics [37] and/or a relatively high availability of sulfate.

Because sulfate reducers may use the same substrates as sulfur reducers (i.e. sulfide and thiosulfate [43, 44]), they are here collectively referred to as sulfur species reducing microorganisms (SSRM) performing sulfur species reduction (SSR). Direct respiration of sulfur is limited by its low solubility (1.6*10⁻⁷ M) and hence requires cell attachment to the sulfur particle [45]. However, sulfur readily reacts with sulfide formed during the reduction of sulfate to form easily metabolizable polysulphides [45, 46].

Iron (III) reduction relies on the availability of iron oxides and iron-bearing minerals such as 179 180 smectite and chlorite [47, 48], as well as the availability of organic carbon, since dissimilatory iron reducing bacteria (DIRB) are strict heterotrophs, i.e. synthesize cell carbon from organic 181 182 compounds [49]. Iron oxides are abundant in many sediments and aquifers [47] but are typically not available in the carbon-rich oil fields because they have been reduced over 183 184 millions of years and are not replenished [24]. Meanwhile, bacteria capable of reducing iron 185 are frequently isolated from hydrocarbon-contaminated or oil-associated sites (reviewed in [50]). However, the mere observation of iron reduction by bacteria, which are given a DIRB 186 enrichment medium in the laboratory, does not imply that these bacteria will reduce iron in 187 nature. In addition, cell counts are often low to intermediate (10-100 cells ml⁻¹) and may 188

include non-hydrogenotrophs (e.g. [51, 52]). In non-engineered environments rich in Fe oxides and organic carbon, IRB may have a great advantage over SSRM, methanogens and homoacetogens, due to a very high affinity for H₂ [47]. We evaluate this process as of intermediate relevance for H₂ storage in DOGF.

Many IRB and a few SSRM can also couple H_2 oxidation to reduction of a variety of other trace metal oxides, e.g. $MnO_4^{2^-}/MnO_2$, $CrO_4^{2^-}$, Co, $SeO_4^{2^-}$, UO_2^2 , TcO_4^- , AsO^{3^-} , and VO_4^- [43, 53, 54]. After Fe, the most abundant metal in sedimentary environments is Mn (~10 % of Fe abundance) [47, 53]. Due to the trace content of these compounds in the environment, their reduction has low relevance for H_2 storage.

Oxygen and nitrate are scarce in the subsurface [11, 24, 55, 56] and aerobic hydrogen oxidation,
denitrification and ammonification hence only become significant when contamination of the
aquifer occurs, e.g. by drilling fluid [57-59].

Halogenated compounds are common in aquifers, and may arise from contamination or via natural processes in sediment [60, 61]. However, the concentrations of these compounds are extremely low: In aquifers of 167-1000 m depth, chloroflourocarbons reach maximum concentrations of $\leq 1.1 \ \mu g \ L^{-1}$ [61] and for pristine aquifers 0.003-0.007 $\mu g \ L^{-1}$ of chlorinated hydrocarbons were measured [60]. We evaluate the relevance of this process to H₂ storage as negligible.

Literature on the importance of anaerobic fumarate respiration using H₂ is scarce. Fumarate may be used as an alternative electron acceptor by SSRM [43, 62] and homoacetogens [63-66]. In the non-engineered subsurface, readily metabolizable organic matter, like fumarate, is rare [67]. Oil fields being rich in organic C compounds may contain more fumarate. Payler et al. [12] confirmed the presence of fumarate reductase, the key enzyme in fumarate reduction, in three out of five metagenomes from subsurface brines within sandstone. However, the metagenomes belonged primarily to non- H_2 utilizing bacteria (*Halorubrum*) and fumarate concentrations were not reported. Acknowledging the lack of data in this field, we evaluate this process as being of low relevance for H_2 storage.

216 **2.2 Factors governing microbial growth**

217 Microbial growth and H₂ consumption rates vary with nutrient availability and environmental 218 variables (e.g. [17, 68].) Each strain is adapted to an optimum set of nutrients and 219 environmental conditions where potentially the greatest growth rates occur. Beyond the optimum conditions, organisms may grow but at reduced rate or they become dormant. In this 220 221 section, we discuss the requirements for nutrients and water, and the overall impact of 222 temperature, salinity, pH and pressure on the growth of the major microbial H₂-oxidizers in DOGFs, in the ranges relevant to H₂ storage. The specific activity of microbial strains grown 223 at optimum conditions varies as well (reviewed in [69]) but the elucidation of differences 224 225 between strains is beyond the scope of this review.

226 2.2.1 Nutrients

227 Apart from water of sufficient thermodynamic activity (see Section 2.2.4), hydrogenotrophs 228 require H₂ as a source of electrons (energy), an electron acceptor and a carbon source for cell 229 division, together with a set of macro and trace elements as well as various organic nutrients [70]. Microorganisms can only access $H_{2(aq)}$ and hence the solubility of $H_{2(g)}$ is of direct 230 231 relevance for all H₂-consuming reactions. Given a gas phase of ~100 % H₂ in an H₂ storage system, the equilibrium solubility of H₂ exceeds the highest threshold value of an H₂-232 consuming microorganism of 3.6 µM (Table 1) by ~3 orders of magnitude at ambient pressure 233 234 and temperature and under static conditions (Fig. A.1a), with further increase at higher 235 pressures (Fig. A.1b). While under non-static conditions hydrogenotrophs will consume part

of the H₂, these figures suggest no limitation by the H₂ solubility on microbial growth under
H₂ storage conditions.

Elemental requirements include the macro elements C, N, H, P, Ca, Mg, S and Fe (>95 % of the microbial cell dry weight), and the trace elements Co, Mn, Ni, Mo, Cu, Zn, W as well as Se for some metabolic groups [71, 72]. For optimum growth, many microorganisms additionally require different vitamins (e.g. lipoic acid, biotin, riboflavin, folic acid, thiamine, etc.), yeast extract, coenzyme M, aromatic acids and phospholipids or a combination of these (e.g. [8, 65, 73-75]).

Nutrients may be assimilated from the solution or directly from minerals (e.g., [76-79]), the latter being of particular importance in oligotrophic environments [77]. Carbon, sulfur, phosphorous and iron are amongst the key elements released by mineral weathering [77]. The extent to which subsurface microbial communities depend on mineral weathering is unknown [77]. For soils, Huang et al. [80] analyzed that >50 % of the 1100 microbial strains were capable of mineral weathering, as tested by their ability to mineralize biotite.

Microbial cell carbon may be assimilated from CO₂ alone (autotrophy) or from organic carbon compounds (heterotrophy) [81]. Methanogens and homoacetogens can grow autotrophically or heterotrophically, and several can grow mixotrophically (e.g. [66, 82, 83]). SSRM typically grow heterotrophically but some grow autotrophically or mixotrophically [84, 85]. Nitrogen may be assimilated from ammonia and nitrate or by nitrogen-fixation (diazotrophy). Diazotrophy is common amongst SSRM, methanogens and homoacetogens [86-89], though homoacetogens often inhabit ammonia-rich environments [88].

Little is known about the differences in the nutrient requirements on the level of functional groups and the variation in nutrient requirement within a functional group. SSRM have a higher requirement for iron $(1.8*10^{-6} \text{ M})$ than is usually observed for microorganisms [90] while 260 methanogens have a higher requirement for sulfur with optimal levels ranging from 0.03 to
261 0.79 mM (reviewed in [91]).

Literature on when nutrients become limiting is also scarce. Sulfate reducing SSRM require a minimum sulfate and phosphorus concentrations of ~ 3 mM and $\sim 3.2-320*10^{-5}$ mM, respectively, for growth [92, 93]. Methanogens of the order *Methanosarcinae* require 29.6 mM Mg for optimum growth and growth ceases at 15.8 mM (reviewed in [91]). When grown under optimum conditions, the growth rate of autotrophs may be limited by the rate of transfer of H₂ and CO₂ from gas to liquid, as was shown for the methanogen *Methanobacterium thermoautotrophicum* [94] and for the sulfate reducers within *Desulfotomaculum sp* [95].

Carbon is unlikely to be limiting in the hydrocarbon-rich DOGF [56, 96, 97] but this is not a given in saline aquifers with no history of oil or gas. Sulfate is present in significant concentrations in most DOGF (Table 2) but H₂ injection can cause sulfate depletion due to accelerated growth of SSRM [98]. Nitrogen in the form of ammonia may be limiting in DOGF [51, 56, 90] but nitrate levels may be elevated [51], often due to contamination by drilling fluid [57-59].

275 **2.2.2 Temperature**

Temperatures of 22.5–80 °C or 20–100 °C have been suggested for H₂ storage based on a recommended depth range of 500- 2000 m for H₂ storage in DOGF and saline aquifers [99-101]. Microorganisms are classified according to their preferred growth temperature: psychrophiles grow optimally below 20 °C, psychotrophs grow optimally at or above 20 °C and may tolerate temperatures below 5 °C, mesophiles grow between 20 and 45 °C, thermophiles grow above 45-50 °C, and hyperthermophiles show optimal growth at temperatures of 80 °C or above [102, 103]. High temperatures alter the energetic properties (e.g., vibrational modes) of biomolecules in their aqueous solvent, change the substrate solubility or viscosity and the ionization of the aqueous medium [104]. Adverse effects of high temperature include DNA denaturing or damage, decreased protein stability, hydrolysis of ATP and ADP, amongst others [104, 105]. The metabolic strategies of thermophiles are highly diverse. For a discussion, the reader is referred to [106].

Thermophiles and hyperthermophiles are challenged by increased reaction rates at elevated temperature which can imply that abiotic reaction rates are so fast that there is no benefit to the microorganism if it catalyzes the reaction [41]. High-temperature-adapted microorganisms are therefore thought to produce enzymes with faster reaction rates [107].

Most cultivated hydrogenotrophic methanogens are mesophiles but known optimal growth temperatures for methanogens range from 15 to 98 °C (Fig. 1a). A considerable number of methanogens favor temperatures above 60 °C (Fig. 1a). The highest temperature that a methanogen was found to grow under is 122 °C (*Methanopyrus kandleri*) (Fig. 1b) [108].

297 Cultivated SSRM typically have optimum growth temperatures of 20-30 °C or 50-70 °C where 298 sulfur reducing archaea have higher optimum growth temperatures than sulfur and sulfate 299 reducing bacteria. The full range for optimum growth of SSRM spans 10-106 °C (Fig. 1i). The 300 critical temperature for growth of cultivated SSRM is 113 °C (*Pyrolobus fumarii*) [109].

Homoacetogens typically have optimum growth temperatures between 20-30 °C (85 % of the
here gathered cultivated strains; Fig. 1e). Thermophilic growth temperatures ≥60 °C have been
reported for eight strains, only (e.g. *Moorella mulderi*, *Thermoanaerobacter kivui*, *Acetogenium kivui*) [110-112]. Corresponding upper limits for growth are 70-72 °C (Fig. 1f)
[110-112].



Figure. 1. Distribution of optimum growth temperature, critical growth temperature, optimum pH values and critical salinity for 123-140 methanogens (a-d), 21-91 homoacetogens (e-h) and 151-255 sulfur species reducing microorganisms (SSRM) (i-l). Distributed between the graphs for the different groups of H₂-oxidizers are the temperatures, pH values and salinities of 42 depleted oil and gas fields (DOGF) and five test sites for H₂ injection. Where ranges of a parameter were given (see Table A.4), the lower end value was plotted.

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315 2.2.3 Salinity

The relevant salt concentration range for H_2 storage is 0-5 M NaCl [100], at which highly diverse prokaryote communities can be found [113]. Microorganisms are classified according to their salt tolerance: Non-halophilic microorganisms grow up to 0.2 M NaCl, slight halophile grow at 0.2–0.5 M NaCl, moderate halophile between 0.5–2.5 M NaCl, and extreme halophile that grow best in hypersaline media containing 2.5–5.2 M NaCl [113].

High salt concentrations exert osmotic stress [114], requiring any microorganism living at high salt concentrations to maintain its intracellular environment at least isosmotic with the environment [113]. Commonly, salt tolerance/requirement is enhanced at increased temperatures [113] but there are many examples of mesophilic halophiles.

Most hydrogenotrophic methanogens favor salt concentrations up to 0.77 M NaCl (the approximate salinity for seawater) but 16 known strains survive under more halophilic conditions. Two extremely halophilic mesophilic hydrogenotrophic methanogens, will tolerate salt concentrations of ~3.3- 3.4 M, *Methanocalculus halotolerans FR1T* [115] and *Methanocalculus natronophilus* [116] (Fig. 1d).

The large majority of SSRM grow optimally at low salinities between >0-0.4 M. However, fourteen SSRM (all mesophiles) have upper salinity limits for growth of \geq 1.7 M NaCl (Fig. 1h). *Desulfovibrio oxyclinae*, *Thiobacillus halophilus*, *Desulfohalobium utahense* and *Desulfohalobium retbaense*, have the highest upper salinity limits for growth of 4.0 to 4.2 M NaCl [117-120] (Fig. 1h).

The salt tolerance of homoacetogens is poorly investigated. The majority of homoacetogens have low optimum salinities of >0-0.4 M NaCl. However, a few strains, i.e. *Natroincola histidinovorans*, *Sporohalobacter lortii* and *Halanaaerobium praevalens*, grow optimally at >1.4 M NaCl and will tolerate salinities up to 2.6-5.1 M (Fig. 1h) [121-123].

339 2.2.4 Brine complexity

340 Natural brines contain dissolved ions whose interaction is extremely complex and may cause physicochemical stressors to brine habitability such as low water activity (a_w) , high ionic 341 342 strength, chaotropy (ability to disrupt the network of H₂-bonds between water molecules) or a combination of these [12]. Most bacteria grow well at an a_w around 0.98 (the approx. a_w for 343 sea water) but relatively few species can grow at a_w of 0.96 or lower [124]. Halophilic 344 345 microorganisms, including halophilic methanogens are one exception; several can grow at a_w as low as 0.75 [125] in [124]; [126]. Steinle et al. [127] challenged these limits by detecting 346 347 SSR in a nearly MgCl₂ saturated brine with a_w of ~0.4.

There are indications of a more important role of chaotropy over a_w in limiting microbial life 348 [126]. Chaotropic agents include MgCl₂, CaCl₂, FeCl₃, KI, LiBr, LiCl while examples of 349 kosmotropic agents are NaCl, KCl, Na₂SO₄, MgSO₄, K₂SO₄, FeSO₄ [128]. As such one may 350 speculate that most subsurface brines due to their dominance of NaCl and richness in sulfate 351 352 are kosmotropic and albeit also stress-inducing, more permissive of microbial growth [12, 126]. 353 Meanwhile, the interactions between chao- or kosmotropic agents, a_w and other physicochemical properties of brines may be very complex and hitherto not understood [12]. 354 355 The further elucidation of this topic is subject to more research and beyond the scope of this 356 paper.

357 *2.2.5 pH*

The brine pH may affect the growth of microorganisms via 1) a direct effect on the growth metabolism, and 2) an effect on the redox reaction. With respect to the former, most methanogens, homoacetogens and SSRM are adopted to a pH of 6.5-7.5 (Fig. 1c, g, k). Most methanogens and SSRM cannot grow outside the pH range 4–9.5 [29, 129, 130] (Fig. 1c, k). Ten known methanogens can endure a critical pH-value of 10 (e.g. the *Methanosalsum* *natronophilum* and *zhilinae*, and the *Methanocalculus natronophilus* and *alkaliphilus* [131].
At the other end of the spectrum, ten known methanogens can endure acidic conditions of pH
4, e.g. the *Methanoregula boonei*, the *Methanothermococcus okinawensis*, the *Methanosarcina spelaei* and the *Methanocaldococcus bathoardescens* [132-135].

Eighteen known SSRM are adapted to highly alkaline environment >pH 10, e.g. the Desulfonatronovibrio hydrogenovorans, the Desulforispira natronophila and the Desulfovibrio vietnamensis [136-138]. Thirteen known SSRM grow down to a pH of 4. Nine known SSRM, all of them sulfur reducers, grow down to a pH of 1, e.g. the Thiobacillus caldus, the Sulfolobus acidocaldarius, the Acidianus infernus and brierleyi, and the Stygiolobus azoricus [139-142].

Seven known homoacetogenic strains have high critical pH values up to 10.0-10.7, e.g. *Clostridium ultunense*, *Natrionella acetigena*, *Fuchsiella alkaliacetigena and ferrireducens*, *Natronoincola histidinovorans*, *Peptostreptococcus productus B-52* and *Moorella sp HUC22- I* [121, 122, 143-146]. The *Clostridium drakai*, *ljundahlii*, *scatologenes*, *coccoides* and *termoautrophicum* are the most acidophilic known strains; they can tolerate pH as low as 3.64.5 [147-151].

379 2.2.6 Pressure

Pressure ranges for H₂ storage of 5-20 MPa [99] or 1-50 MPa [100] have been reported. Life at high pressure requires homeostatic changes [103]. The high pressures encountered in pore spaces in the crust are generally less inhibitory to microbial cellular activity than the high temperatures, partly because of the relatively high osmotic pressure of cytoplasm [102], in particular in thermophiles and hyperthermophiles [41]. DNA synthesis and protein synthesis are among the most pressure-sensitive cellular processes [103, 152]. Protection against pressure includes biofilm [18] or spore formation [103]. 387 At 30-50 MPa, the growth of various mesophilic, atmospheric-pressure-adapted 388 microorganisms is inhibited [152] whereas pressure effects are generally favorable for the growth of hyperthermophiles; above 100 °C, elevated pressures are required to maintain a 389 390 liquid environment [105]. Microorganisms that grow optimally at 10 MPa or above are obligate 391 and facultative piezophiles, where the former do not tolerate ambient pressure and the latter do [103]. A recent publication listed all identified piezophiles and grouped them according to their 392 growth temperature optimum [103]. The list of species is rather short (and as we find 393 394 incomplete despite being published in 2020), possibly due to the fact that, to date, it has not 395 been possible to isolate genes associated with piezophily, so the effects of pressure on any 396 particular organism can only be determined empirically [103]. Empirical efforts however, do 397 not commonly include pressure tolerance in the description of the environmental growth 398 criteria of a microorganism. In addition, most mesophiles and thermophiles from habitats with 399 pressures of <50 MPa will grow in enrichment cultures incubated at atmospheric pressure [34]. The large majority of identified cultivated piezophiles are psychrophiles (27 strains) [103], the 400 401 relevance of which is low to our study. Only four mesophilic strains were reported, three of 402 them hydrogenotrophic sulfate reducers (the Desulfovibro profundus, piezophilus, and 403 hydrothermalis), growing optimally at 10-40 MPa [103]. Eight thermophiles were identified, including one hydrogenotrophic methanogen, Methanococcus thermolithrophicus, growing 404 405 optimally at 50 MPa. The hyperthermophilic group hosts the hydrogenotrophic Methanopyrus 406 kandleri and Methanocaldococcus jannaschii growing optimally at 20 to 75 MPa, respectively. 407 Examples of hydrogenotrophic piezophiles that are not included in [103] are the mesophilic 408 SSRM Parococcus pantrophus and Pseudodesulfovibrio indicus which growth optimally at 30 409 and 10 MPa, respectively [153, 154], and the thermophilic SSRM Piezobacter thermophiles and Archaeoglobus fulgidus TF2 which grow optimally at 30 and 42 MPa, respectively [75, 410 411 155].

A temperature dependence of the pressure response was reported for the SSRM *Desulfovibrio indonesiensis* which has similar growth rates at high and ambient pressure 45 °C but reduces its growth rate at 20 °C and 30 MPa relative to at 0.1 MPa [156]. Elevated pressure may increase the maximum growth temperature by 2-12 °C relative to lower pressure (0.1-3 MPa) [105, 108, 157].

417 *2.2.7 Inhibitors*

Exposure to hydrogen sulfide, H_2S , and its bisulfide ion, HS^- , causes damage to microbial proteins and coenzymes [91, 158]. It remains unclear whether H_2S or HS^- is responsible for the toxicity effect but there is general consensus that H_2S can penetrate the microbial cell membrane more easily than HS^- [158]. Hydrogen sulfide dissociates with a pK₁ of 6.99 at 10 MPa and 25 °C to form >99 % HS⁻ at pH 8.5 [159].

Growth of SSRM and methanogens is adversely affected at concentrations of $H_2S > 3.8-4.0$ mM [160-162]. At 5.0-6.3 mM H_2S growth is completely inhibited for SSRM [160, 162], without however stopping all metabolic activity [160]. For methanogens and homoacetogens 3.8-7.5 mM H_2S and total sulfide concentrations of 3.3 mM, respectively, stop the growth [158, 162]. In systems with circumneutral pH and ferric ion concentrations above 1 mM, the concentrations of H_2S are predicted to be kept below toxic levels due to its precipitation in makinawite [46].

Carbon dioxide pressure above 1 bar can be toxic for microorganisms as shown for the SSRM *Desulfotomaculum geothermicum* and the methanogen *Methanothermococcus thermolithotrophicus* [163]. For many anaerobes like methanogens and homoacetogens,
oxygen is toxic too [64, 102].

434 Nitrate inhibits homoacetogenesis [164], and ammonium [165] and sulfate inhibit 435 methanogenesis (reviewed in [166]), with minimum inhibitory concentrations varying depending on the environment [162, 165]. For instance, sulfate concentrations as low as $2*10^{-4}$ M were shown to inhibit methanogenesis for 10 hours in lake sediments, possibly by competition with SSRM for available H₂ and C-substrate [161] (see section 2.3.9). Under H₂ storage conditions however, sulfate is likely not to affect methanogenesis, because sulfate inhibition was shown to be reversed by addition of H₂[161]. For a discussion of an inhibitory effect of H₂, see section 2.5.

442 2.2.8 Summary of environmental growth constraints

443 Acknowledging the lack of data for the pressure sensitivity of many microorganisms [103], 444 and considering a general abundance of nutrients in DOGF (Table 2), we evaluate temperature and salinity as the most crucial environmental factors constraining the growth of 445 homoacetogens, methanogens and SSRM in DOGF. Pressures encountered in the crust are 446 documented to have less effect than temperature on microbial cellular activity, particularly in 447 thermophiles and hyperthermophiles [41, 102]. The pH does not pose a similar constraint to 448 449 the growth of homoacetogens, methanogens and SSRM because the pH ranges for growth typically span two to three pH units (not shown) and for most species they comprise the typical 450 451 aquifer pH values of 6-7 [167] (Table A.4). Brine complexity and inhibitors were not included 452 in this analysis due a lack of information on the brine composition of DOGF beyond a limited 453 set of dissolved ions.

Figure 2 shows the critical temperature versus critical salinity for 269 cultivated strains and reveals that salt tolerances up to 1-1.7 M are widely distributed over the entire temperature range while salt tolerances >1.7 M are mainly found at a critical temperature tolerances of 40-50 °C. Hence, from the point of view of minimizing microbial impacts on H₂ storage, sites with temperatures >50 °C and salinities >1.7 M are preferred.

Field name	Area (Km²)	P (MPa)	Temp (°C)	Salinity (M)	рН	HCO3 ⁻ (mM)	N2 (mM)	SO4 ²⁻ (mM)	K ⁺ (mM)	Ca ⁺² (mM)	Mg ⁺² (mM)	P (mM)	Na ⁺ (mM)	Cl ⁻ (mM)	Fe ⁺² (mM)	Orga acids (mM
Frigg	100	19.5	61	0.07-0.53	6.5-	16.3	0.4	NA	26.3-	0.4-2.0	1.9-7.1	NA	75.2-	58.7-	0.04-	<u>`</u>
88					7.4				31.2				534.8	490.3	0.27	1
Hamilton	15	9.6	30	1.59-4.18	5.8	4.8	2.1	0.6-	8.4-	72.8-	19.5-	0.012-	1354.8-	1453.3-	4.03	N
								7.4	29.7	720.0	37.6	0.028	2210.9	3700.7		
Barque	36	26.0	79	4.83	4.7	0.3	0.8	3.5	42.2	535.0	156.8	NA	2920.4	4405.4	2.15	Ν
Hamilton	8	10.5	30	2.93	7.9	11.0	2.3	23.1	18.8	13.6	13.6	NA	2640.9	2662.9	NA	Ν
North																
Miller	45	49.3	121	1.61	7.2	NA	0.6	0.0	41.6	30.0	NA	NA	1358.7	1471.9	0.02	Ν
Beryl	49	36.0	101	1.88	6.1	5.6	0.4	0.0	20.8	90.0	NA	NA	1469.6	1717.9	0.05	1
Judy	NA	46.9	137	0.14-0.15	6.8	8.4	0.6	6.4	2.9	4.5	NA	0.002	117.4	131.7	0.11	N
(Andrew 1)																
Amethyst	97	27.9	88	4.45	5.6	1.0	1.6	3.7	33.2	521.5	148.5	0.452	2673.9	4064.6	2.51	N
Rhyl	NA	14.9	36	5.80	5.5	13.5	2.8	14.0	62.4	147.0	21.2	0.031	4777.0	5297.9	0.81	>
Dalton	NA	28.8	91	0.26	5	0.9	1.0	1.8		15.6	5.5	NA	189.1	237.0	0.00	N
Davy	6	28.2	88	3.87	6.8	6.5	NA	7.0	219.2	15.6	10.7	0.155	818.3	1142.7	0.66	N
Veslefrikk ^{\$\$}	NA	29.8-	67-114	0.29-0.72	6.5	8.4-	NA	0.1-	NA	NA	NA	NA	298.0-	281.0-	NA	2.2
		35.0				17.2		0.15		1	10.0		666.0	745.0	0.06	
Average						7.9	1.1	5.2	44.7	166.8	42.2	0.113	14/3./	1857.4	0.967	

Table 2. Reservoir conditions for selected depleted, or soon to be depleted oil and gas fields. Except where otherwise indicated, the data are

from [168]. = reference [167]. The salinity was calculated from the chloride concentration and the concentrations of dissolved N₂ was

estimated from the mol percentage in the gas phase, neglecting any effect of salinity. NA= not analyzed. See Table A.4 for extended data.

463 Growth of all investigated microbial groups occurs up to 72 °C (Fig. 1). Above 72°C, known homoacetogens will not grow, and at 80-94°C sulfate reducers cease to grow. Thirty-six 464 cultivated SSRM and eleven methanogens have optimum growth temperature of ≥ 80 °C (Fig. 465 466 1a and g) and will still grow, albeit at reduced rate, beyond their optimum temperatures. The maximum growth temperature for known methanogens and sulfur reducers is 122 °C and 113 467 °C, respectively. The upper salinity limit that allows growth of all the major groups of 468 investigated H₂-oxidizing microorganisms is 3 M NaCl. The upper pH limit is 9.5 and the upper 469 470 pressure limit for most mesophiles is 30-50 MPa.



Figure. 2. Critical temperature (without salinity stress) versus critical salinity (without temperature stress) for methanogens, homoacetogens and SSRM.

471

472 **2.3 Microbial growth regulation by competition and syntrophy**

473 Homoacetogenic bacteria are ubiquitous in anaerobic sediments [65, 169] and often co-exist

474 with SSRM and methanogens [15, 170], as revealed by a combination of molecular (16S RNA

gene sequences) and culturing (e.g. metabolites, radiotracer) techniques. Few habitats have been identified in which homoacetogens compete with other H₂-consumers (culturing studies) [169, 171]. Exceptions include a low-temperature and low-salinity petroleum reservoir where homoacetogens dominated over methanogens and SSRM (molecular study) [62], a granite groundwater at 400 m depth where cell numbers of methanogens and homoacetogens were balanced (molecular study) [83], and subsurface marine sediments where mixotrophic homoacetogenesis outperformed methanogenesis (culturing study) [82].

Kinetic advantages of SSRM and methanogens (i.e. a higher affinity for H₂, expressed as a 482 483 low Michaelis-Menten constant, K_M , or Monod half saturation constant, K_S (H₂ concentration 484 at which growth rate reaches half maximum growth rate), and a higher maximum growth/reaction rate, V_{max} or μ_{max} for Michaelis-Menten kinetics and Monod kinetics, 485 respectively) were proposed as the underlying cause for the few examples of the poor 486 competitiveness of homoacetogens [172]. Very limited information on the H₂ consumption 487 488 kinetics of homoacetogenic bacteria is available in literature [173]. The available data show that μ_{max} differs by one order of magnitude between strains (0.02-0.5 h⁻¹) [4, 173]. This may or 489 may not be lower than the μ_{max} for SSRM 0.057-5.5 h⁻¹ [4, 40, 174] and methanogens 0.032-490 1.4 h⁻¹ [40, 174]. Krumholz et al. [172] showed that homoacetogens were not able to compete 491 effectively for H₂ in the presence of SSRM in a subsurface sandstone ecosystem at 30 °C 492 493 regardless of pH_2 , and despite significant homoacetogenesis at excess H_2 . Findings by Berta et al. [4] for a groundwater sediment held under excess pH₂ and 20 °C contrasts this as 494 495 homoacetogenesis rates were up to 21 times higher than SSR.

Environmental conditions may be a crucial determinant for the competitiveness of homoacetogens, as low temperatures (~15 °C) [175, 176] and low pH values [64, 93] favor their growth over methanogens. Under excess pH_2 , homoacetogenic strains with high u_{max} such as *Acetobacterium bakii* will outcompete methanogens [176]. The outstanding metabolic
flexibility of homoacetogens for utilizing a vast variety of substrates may additionally explain
why homoacetogens can compete with more specialized microorganisms like SSRM or
methanogens [65, 176, 177].

503 As for the competitiveness of methanogens and SSRM, the H₂ thresholds of methanogens 504 may be comparable (1-15 nM) or higher (>15-95) than for sulfate reducers and significantly lower than for sulfur reducers («2500 nM; Table 1), indicating an advantage of sulfate 505 506 reducers over methanogens and sulfur reducers in most non-engineered, low pH₂ environments. In line with this, Lackner [178] recently reviewed that sulfate reducers 507 508 outcompete methanogens for H₂ in most studies. However, at excess H₂, methanogens and 509 sulfate reducers would be expected to process equal shares of the in situ H₂ pool [174]. Also, 510 since concentrations of sulfate are much lower than bicarbonate in non-marine natural environments [40] (Table 2), the growth of sulfate reducers at excess H₂ will be limited by 511 512 availability of their electron acceptor, making it possible for methanogens to compete [40]. As 513 a general rule pH values below 7 favor the growth of methanogens over sulfate reducers [130]. Above pH 7.5, sulfate reducers grow faster than methanogens and would be expected to 514 515 outcompete them [130].

516 Syntrophic relationships between different functional groups have been documented frequently 517 (whereby the metabolic products of one group serve as substrates for the other). For example, 518 SSRM and homoacetogens were shown to participate cooperatively in microbial induced 519 corrosion of steel where SSRM grew on acetate produced by homoacetogenesis [70]. Substrate 520 provision by the co-culturing *Desulfovibrio vulgaris* enhanced growth of the dehalogen 521 *Dehalococcoides ethenogenes 195* by 24 % and caused three times higher dechlorination rates 522 [179]. Syntrophy may also explain the detection of a combination of the SSRMs *Desulfovibrio*

25

and the homoacetogens *Acetobacterium* in petroleum and subsurface CO_2 reservoirs [62, 180], and the presence of H₂-producing heterotrophs along with methanogens in petroleum reservoirs where the latter rely on H₂-transfer by the former [181].

526 2.4 Microbial ecology in natural gas and petroleum reservoirs

527 Recent years have seen a considerable effort in describing deep subsurface microbial 528 communities, including those from gas and petroleum reservoirs. Isolated hydrogenotrophic 529 microbes from these habitats are from the SSRM families Archaeoglobaceae [182], Desulfomicrobiaceae, Desulfobulbaceae, *Peptococcaceae*, 530 Desulfobacteracceae, 531 Desulfovibrionaceae, Desulforobacteriaceae, Sulfurospirillaceae, Rhodobacteraceae. 532 Ectothiorhodospiraceae, *Hydrogenothermaceae* [27, 56. 97. 98. 183-187], the Eubacteriaceae and Sporomusaceae families which host homoacetogenic strains [97, 186, 533 188], and the methanogen families Methanosarcinaceae, Methanobacteriaceae, 534 535 Methanomicrobiaceae, Methanopyraceae, Methanococcaceae, Methanocalculaceae and 536 Methanosaetaceae [98, 115, 186] in addition to uncultured microbial taxa [56, 184, 185, 189]. 537 Our collection of hydrogenotrophs (Fig. 1) lists many examples of the above microbial families, including the strain that holds the highest critical temperature for a methanogen, 538 539 Methanopyrus kandleri. Sulphur reducing families that define the upper temperature limits for SSRM like Thermoproteaceae and Pyrodictiaceae were not reported. The cause for their 540 541 absence may be a predominance of mesophilic and thermophilic sites but may also reflect a generally stronger growth of sulfate reducers over sulfur reducers in oil and gas reservoirs. 542 543 Ranchou-Peyrouse et al. [98] showed that the microbial community in 35 out of 36 subsurface 544 wells from seven natural gas storage sites was dominated by sulfate reducers.

545 2.5 Effect of high hydrogen concentrations on the microbial metabolism and community 546 structure

26

547 A range of studies investigated the metabolism of methanogens at excess H₂ and ambient 548 pressure, with unambiguous results. Conrad et al. [190] demonstrated that excess H₂ stimulated methanogenesis and growth rates in a paddy soil (species not specified). Opposed to this, 549 550 results by Topcuoglu et al. [181] and Stewart et al. [191] suggest an inhibitory effect of high 551 partial pressures of H₂, pH₂, expressed as a ~10-fold drop in the growth yield (cells per mole CH₄) of *Methanocaldococcus jannaschii* and a slight drop of ~0.1-0.7 h⁻¹ in the growth rate. 552 Similar observations were made for Methanothermobacter thermoautotrophicus [192]. 553 554 However, within the excess H₂ experiment, higher H₂ concentrations stimulated growth [181], 555 suggesting a complex influence of pH_2 . Methanogens seem to express a pH_2 -dependent change 556 in their ecological strategy, i.e. maximum growth rate vs. maximum growth yield, as a means to cope with different environmental conditions [181]. Indeed, M. jannaschii is capable of 557 558 sensing subtle changes in dissolved H₂ concentration and restraining the energy-intensive growth of flagella to H₂-limiting conditions whereas at excess H₂ cells are mostly flagella 559 560 devoid [193].

561 Only few studies investigated microbial H₂ turnover at high pH₂ of up to 1.5-24.8 MPa [4, 194, 195]. Methanogens (M. jannaschii) showed a strong inhibitory effect at high pH₂ [194]. 562 563 However, the authors added CO₂ at a pressure of at least 0.2 MPa to the hydrogen gas mixture which at $pCO_2 > 0.1$ MPa can be toxic methanogens [163]. Hence it is not clear whether H₂ or 564 CO₂ performed the toxic action. For homoacetogens and SSRM, the H₂ consumption was 565 shown not to change in response to different pH₂ of 0.1-3.5 MPa [4, 195], indicating neither 566 567 stimulation nor toxicity at different levels of excess H₂. The comparison to limiting H₂ 568 conditions was not made.

Apart from microbial metabolism, the microbial community may also change in response to high pH_2 . Given a pertubation by H_2 injection it can be anticipated that other types of 571 microorganisms, e.g. the in hydrocarbon reservoirs, common fermenters [24, 96, 98, 183] will 572 decrease in abundance while hydrogenotrophs will increase [7], in line with the Baas Becking principle [196]. An increase in hydrogenotrophs in response to H₂ addition was recently 573 574 confirmed for soils, however H₂ consumption increased in only one of the investigated soils, suggesting a pronounced influence of the indigenous microbial community [197]. Bioreactor 575 experiments support a decrease in microbial diversity in response to high pH_2 as well [198, 576 199]. Puente-Sanchez et al. [200] were the first to report differences in the subsurface H₂-577 578 consuming community in response to varying pH_2 within the Iberian Pyrite Belt. Ranchou-579 Peyruse et al. [98] showed that town gas storage with more than 50 % H₂ changed the microbial 580 community from a predominantly sulfate reducing community to a dominance of methanogens, and this balance was active even decades after injection stopped, possibly via H₂ trapping in 581 582 the microporous system [98]. It was suspected that all sulfate was initially used up by SSRM 583 following increased growth of methanogens [98].

3. Evaluating the potential hydrogen consumption in DOGFs

585 **3.1 Calculation of the microbial growth**

We screened 42 DOGF in the North Sea and the Irish Sea and five H_2 storage test sites for temperature, salinity, pH and pressure data (Fig. 1, Table A.4). We discovered significant differences in the salinity for the DOGF reported by sources [201] and [168]. Because we relied on the solution compositions for the calculation of the potential microbial growth in the fields, which are available from [168], we chose to use the salinity data from the same source.

591 The environmental data from the DOGF and H_2 storage test sites were aligned with the 592 constraints for growth of methanogens, homoacetogens and SSRM (Fig. 1-2) to select in which 593 fields growth can be expected. For the few fields that fulfil the growth constraints of all investigated microorganisms, we calculated a first-order estimate of the microbial growth using
the elemental cell composition as a proxy for the nutrient requirement [202, 203] (Text A.1).

596 Our calculations assumed that the supply of N and C are covered by diazotrophic and 597 autotrophic growth, respectively. Requirements for trace elements were neglected in the 598 calculation due to a lack of information on the relevant trace element contents in the reservoirs. 599 Where a nutrient for a specific field was not available we used the average value from the fields 600 given in Table 2. Any effect of the pH_2 on microbial growth was neglected. We assumed that cells neither die nor are removed, and that nutrients are not replenished by inflow, re-601 602 mineralization from decaying biomass or mineral dissolution. Simultaneous growth by 603 different microorganisms was not considered.

Percentages of nutrients in the cells (Text A.1) were converted to mass using a wet cell mass of $1.77*10^{-12}$ g for methanogens [204], $3.2-6.2*10^{-13}$ g for homoacetogens and $7.81*10^{-13}$ g for SSRM. The cell wet weight of homoacetogens was calculated by dividing the cell volume of $1.62-3.14 \ \mu\text{m}^3$ for the subsurface mixotrophic homoacetogen *Acetobacterium psammolithicum* [172] with an assumed bacterial density of $1*10^{-12}$ g μm^{-3} [205]. The cell wet weight of SSRMs

was calculated using a cell dry weight of $3.125*10^{-13}$ g for *Desulfovibrio desulfuricans* [206] and dividing this with a general bacterial dry weight to wet weight ratio of 0.4 [207]. Subsequently, the concentrations of C, H, O, Ca, K, Na, S, Mg, P and Fe in the DOGF (Table 2) were divided by the mass of the respective cell nutrients per microbial cell calculated above. This resulted in the maximum cell count within each microbial group, *G*, that could potentially be created based on a single nutrient, where the lowest *G* indicated the limiting nutrient for cell growth. For an example of those calculations, see Text A.1.

616 **3.2 Estimation of the cell-specific hydrogen consumption**

Hydrogen may be consumed at rates of $0.02-5.0*10^5$ nM h⁻¹ for homoacetogens, $0.02-5.8*10^5$ nM h⁻¹ for methanogens and $0.005-130*10^5$ nM h⁻¹ for SSRM (Tables A.1-A.3), the latter considering sulfate concentrations in the range of $0-2.3*10^{-2}$ M in the DOGF (Table 2). In a few studies, the microbial H₂ consumption was related to growth (Tables A.1-A.3), enabling the calculation of the H₂ consumption per synthesized cell and the time for when the microbial cell count *G* would be reached (Text A.2).

623 **3.3 Calculation of the hydrogen consumption in a hydrogen storage system**

We calculated the minimum H_2 consumption for the DOGF Frigg and Hamilton by dividing the H_2 consumption per synthesized cell with the microbial cell count. The calculation of the moles of H_2 the in aquifer anticipated equal volumes of H_2 and water and used the ideal gas law and the field size, temperature and pressure data in Table 2 and Table A.3. The percentage of H_2 that was consumed as a function of growing and resting microbial cells was calculated by dividing the potential H_2 consumption with the H_2 concentration in the reservoir. Text A.3 shows our calculations for the Frigg reservoir and methanogens.

631 **4. Results and discussion**

632 **4.1 Characterization of the likelihood for growth in 42 DOGF**

Using the environmental limits constraining microbial growth on H₂, we analyzed the 633 634 physicochemical parameters for 42 DOGF in the British and Norwegian North Sea and the Irish Sea and five H₂ storage test sites (Fig. 1, Table A.4). Of the 47 fields, five fields have a 635 temperature of 122 °C or higher and may be considered sterile with respect to H2-consuming 636 637 microorganisms. Thirty-two fields have a temperature >72 °C, implying that homoacetogenesis cannot take place. Twenty fields have a temperature ≥90 °C implying that homoacetogenesis 638 and sulfate reduction cannot take place. Fourteen DOGF have a temperature >90 °C and <122 639 °C and pressures of 18.2-44 MPa where (piezophile) methanogens and SSRM will grow. 640

641 Of the fifteen sites with temperatures <72 °C where all investigated groups of microorganisms 642 will grow, only six fields (Frigg, Hamilton, Veslefrikk, Ketzin, Lehen and Lobodice) fulfill the remaining pressure and salinity requirements for growth. Two fields, Leman and Rhyl, have 643 644 salinities >5.8 M where no significant microbial growth can be expected. This finding is supported by stable gas compositions at the similarly saline H2-storage test sites of the 645 H2STORE project, Emsland and Altmark (Fig. 1, Table A.4), though a low microbial 646 population of $\sim 10^2$ cells ml⁻¹ was present [208]. Lennox, North Morecambe and South 647 Morecambe with temperatures of 30-33 °C and salinities of 4.6-5.1 M, could permit the growth 648 649 of the halophile homoacetogenic strain Halanaerobium praevalens, only. Hamilton North, Camelot and The V gas field complex with salinities of 2.9-5.0 M may permit the growth of 650 sulfate reducers and Halanaerobium praevalens. The Viking field has temperature of 65-80 °C 651 652 and a salinity of 3.8 M and so is likely to host only mesophilic SSRM, although pressures >30 653 MPa that could become growth inhibiting. The H₂-storage test site Ketzin has similar salinity 654 to the Viking field but a lower pressure (4.0 M NaCl, 35 °C, 6 MPa). Here SSRM were 655 suspected to cause a 2-4 % decrease in H₂ and a reduction in the concentration of sulfate from 22 to 8*10⁻³ M [208]. 656

4.2 Microbial growth estimates for three low-temperature and low-salinity DOGF

Our first order approach for calculating microbial growth in the Veslefrikk reservoir yielded a 658 maximum 1*10⁸ methanogenic cells mL⁻¹, 2*10⁸ SSRM cells mL⁻¹ or 5*10⁸ homoacetogenic 659 cells mL⁻¹. The Frigg reservoir a maximum of 1*10⁸ methanogenic cells mL⁻¹, 1*10⁸ SSRM 660 cells mL⁻¹ or 2*10⁸ homoacetogenic cells mL⁻¹. The Hamilton reservoir could host a maximum 661 of 1*10⁷ methanogenic cells mL⁻¹, 2*10⁷ SRCM cells mL⁻¹ or 6*10⁷ homoacetogenic cells mL⁻¹ 662 ¹. These cell counts describe a maximum cell growth for each hydrogenotrophic group because 663 simultaneous growth of hydrogenotrophs was not considered. The higher growth of 664 665 homoacetogens over SSRM and methanogens results from a lower wet cell mass that causes a

lower nutrient demand per cell (see Text A.1). Our calculations are in line with total cell concentrations of 10^{5} - 10^{15} cells/ mL⁻¹ in oil reservoirs [209], and equal to or up to four order of magnitudes higher than cell counts from gas reservoirs (0.001-1.2*10⁷ cells mL⁻¹)[51, 97, 185]. The range of the literature data reflects that our simple methodology to calculate microbial numbers may serve as a first approximation to estimating microbial growth in the subsurface, only.

672 Acknowledging that trace elements were not accounted for in our calculation, N and P are the 673 first limiting nutrients in the reservoirs Frigg, Hamilton and Veslefrikk. However, this does not 674 imply that microbial growth is N and P limited, as many microorganisms may use of 675 ammonium (not measured) as N-source, and in the Hamilton reservoir the C:P ratio was between 59:1 and 158:1, whereas the limiting C:P ratio for microbial growth is in the range of 676 400:1 to 800:1 (reported for the SSRM D. desulfuricans) [92]. At moderately acidic pH values 677 such as the pH of 5.8 in the Hamilton reservoir, P may further be continuously replenished by 678 679 mineral buffering with apatite.

680 **4.3 Hydrogen consumption in three low-temperature and low-salinity DOGF**

681 The H₂ consumption in the Frigg reservoir by homoacetogens constitutes <0.01- 3.2 % of the H₂ in the aquifer, <0.01- 1.3 % for methanogens and <0.01- 1.3 % for SSRM. In the Hamilton 682 reservoir, the rates are <0.01- 2.0 %, <0.01- 2.3 % and <0.01- 0.5 % for homoacetogens, 683 684 methanogens and SSRM, respectively. For actively growing cells these consumption rates may be reached after only 0.1-19.1 days, which is the time it takes for the microorganisms to grow 685 686 up to their maximum cell counts, based on the dissolved nutrient concentrations. Resting cells, 687 i.e. cells that undergo no or only very little cell division, need 2.5-3.5 months (SSRM) or up to 3.6-6.6 years (methanogens) to reach the maximum cell count and consume the given 688 percentage H₂. 689

690 In a real aquifer system, nutrients are likely to at least partly be replenished by decaying cells, 691 mineral weathering and inflowing brine, and cells will continue to consume H₂ beyond the time it takes to reach the maximum cell count (maintenance). As such our H₂ consumption estimates 692 693 may be regarded as minima. On the other hand, considering that, with the exception of one 694 study (Berta et al. [4]), our calculations employ laboratory H₂ consumption rates at optimal nutrient supply and optimal physicochemical conditions (Tables A.1-A.3), the H₂ consumption 695 in the oligotrophic subsurface is likely overpredicted. Comparing the employed laboratory H₂ 696 697 consumption rates to H₂ consumption rates by SSR and methanogenesis in oil and natural gas reservoirs of ~0.4-330 nM h⁻¹ and 0.02-1205 nM h⁻¹, respectively (SO₄²⁻: 8.3-805*10⁻⁵ M; 698 699 HCO_3 : 3.5-246*10⁻⁴ M) [51, 183], shows that the field H₂ consumption by SSR is 1.5 times to 700 five orders of magnitude lower, and 1.4 times to 7 orders of magnitude lower for 701 methanogenesis. Within the operation and injection wells of a natural gas reservoir, H₂ 702 consumption rates by SSR and methanogenesis were 2393 and 4475 nM h⁻¹, respectively, [51], 703 which falls within the lower range of the values reported from laboratory studies. 704 Acknowledging the unknown but presumably low pH_2 in above experiments, and that 705 maintenance requirements were not included in our H₂-consumption calculations, we expect 706 the actual H₂ consumption in a H₂ storage system to lie within the higher range of our calculated values. 707

Our lower-end results are in agreement with no H₂ consumption during storage operations of H₂ -rich town gas in Beynes, France [210]. Our upper end results are in agreement with a loss of \sim 3 %, presumably by methanogenesis, at the H₂ storage (SunStorage) test site in Lehen, Austria [211]. A H₂ consumption of 17 % by methanogens at the Lobodice town gas storage site over a time span of seven months [210, 212] seems exceptional in the light of our calculations and the reported SSR and methanogenesis rates from the field. With a very low salinity of 0.03 M, temperatures of 20-45 °C, a pH of 6.7 and 4 MPa pressure, Lobodice is among the few sites which has highly favorable conditions for microbial growth considering *all* of these parameters (Table A.4). The high H_2 consumption at Lobodice highlights the importance of our site selection tool, as H_2 storage may face serious economical and technical problems if a site with growth-favoring conditions is selected.

719 As mentioned, Berta et al. [4] measured high H₂ consumption rates under excess H₂ and oligotrophic conditions (P< $9.7*10^{-7}$ M; SO₄² \leq $9.5*10^{-4}$ M; DOC= $2.6*10^{-4}$ M), indicating that 720 721 nutrient scarcity does not imply low H₂ consumption. A comparison to the nutrient concentrations in the DOGF reveals that many of them have a higher nutrient status (P=0.002-722 $0.452*10^{-3}$ M; SO₄²⁻= up to 23.1*10⁻³ M; organic acids= 1.2-8.1*10⁻³ M, Table 2), implying 723 that H₂ consumption in DOGF under excess H₂ conditions may be even higher than reported in 724 [4]. The experiment by Berta et al. [4] is further highly relevant because cells were at steady 725 state, i.e. at the predominant growth stage in nature, but still consumed vast amounts of H₂. 726 Indeed the H₂ consumption of cells at steady state or resting may be just as high or higher than 727 728 for growing cells but growth is low or absent (Tables A.1-A.3).

729 **4.4 Knowledge gaps and future research**

730 More work is needed to predict the magnitude of microbial growth, H₂ consumption rates, and (not least) the mutual interaction of the microbial processes in DOGFs. The list of unknowns 731 and uncertainties is long. To begin with are the poorly elucidated nutrient requirements of the 732 733 microorganisms, especially in mixed cultures (e.g., [71]). Adding to this are the missing or incomplete datasets on the physical environment of certain reservoirs along with their gas phase 734 and brine compositions, including chaotropy and kosmotropy characteristics. A better 735 736 elucidation of the latter would allow the calculation of the dominating microbial processes via 737 their free energies of the reaction. Combined with an analysis of the microbial community and

metabolism this could give new insights into whether or not we can theoretically predict which
 microbial processes occur in DOGF and to which extend.

740 A further complication is the non-cultivability of many microorganisms in the deep subsurface, 741 including DOGF [12, 34, 56, 98, 183]. Considering tiny culturabilities of $\leq 0.1\%$ of the total 742 viable cell count in many subsurface environments [34], any attempts to assign sterile habitats or quantify microbial H₂ consumption via cultivated microorganisms may seem in vain. In gas 743 744 reservoirs, the percentage of cultured bacteria may be higher, ranging between 86-95% within 745 each phylum [98]. Field-based metabolic activity measurements could circumvent any non-746 cultivability issues observed in laboratory experiments. Field studies should also be prioritized 747 considering that microbial cell sizes and masses in nature are only 4-21 % of the laboratorial cell masses [213] which reduces the nutrient requirement per cell, thereby allowing more cells 748 749 to proliferate on any given amount of nutrients.

750 The lack of knowledge about the changes in microbial ecology as a response to increased H₂ 751 concentrations beyond the level of functional groups is one of the major hurdles in our attempt 752 to understand of the effect of high H₂ concentrations on the subsurface microbiology. Emerging 753 evidence on the subject highlights species-specific responses to high pH₂ [98, 198, 200], and 754 that H₂ injection may leave its fingerprint on the subsurface microbial community for decades [98]. Knowledge about the initial effect of a drastic increase in pH_2 in the subsurface is lacking. 755 756 One possibility is that more EPS will be produced as a response to the perturbation with increased H₂, as has been shown for other types of perturbation [18, 92, 214], and considering 757 the toxicity of high pH_2 on methanogens [181, 192, 194], with possible adverse effects on gas 758 759 injectivity and withdrawal.

Future research should address the effect of high pH_2 on the metabolisms of different functional groups and the EPS production in different geological settings and under changing nutritional 762 supply and physicochemical conditions. Mixed culture studies at low and high pH₂ can give 763 insight into competitive and syntrophic relations under these conditions and reveal changes in the microbial community structure due to the pertubation with elevated H₂. Protocols for the 764 765 careful cultivation of nutrient-deprived deep subsurface cells need to be developed. More base-766 line research includes determinations of the critical salinities and pressure tolerances that to date are missing for many cultivated strains, as well as the study of the brine compositional 767 768 effects on the microbial community and metabolism. Research employing already cultivated 769 species can make use of the fact that the large majority of the cultivated species isolated from 770 subsurface environments can be found in other near-surface marine and terrestrial geothermal 771 environments [9, 183], and should employ chemostat studies that mimic the natural 772 environment.

773 **5. Conclusion**

774 Here we presented a novel site selection tool for H₂ storage and demonstrated its application 775 for 42 DOGF in the British and Norwegian North Sea and the Irish Sea and five H₂ storage test sites. Our results highlight the hard limits to the cultivable microbial life on H₂ and can –with 776 777 some certainty- exclude life in several high-salinity or high-temperature, i.e. deeper reservoirs. 778 For low-salinity and low-temperature reservoirs our calculations indicate significant microbial growth and a small but not insignificant H₂ consumption, both of which may further increase 779 during repeated storage cycles, giving replenishment of nutrients by mineral weathering, 780 781 decaying microbial cells and inflowing water. Hence, from the point of view of minimizing H₂ 782 loss, clogging and corrosion, sites with more extreme conditions may be chosen over low-783 temperature and low-salinity reservoirs where the majority of microorganisms can proliferate. 784 Yet, any storage operation will have to consider increased operational difficulties and costs with increased depth. Additional investigation on subsurface life on H_2 is encouraged to help manifest whether H_2 consumption in low-temperature aquifers is a threat to H_2 storage.

787

788 ASSOCIATED CONTENT

789 Appendix. Figure A.1 shows the solubility of hydrogen as a function of temperature and

790 pressure. Laboratories studies investigating homoacetogenesis, methanogenesis and SSR are

⁷⁹¹ listed in Tables A.1, A.2 and A.3, respectively. Table A.4 provides the reservoir conditions

for 42 DOGF and five H₂ storage test sites. Text A.1 holds a discussion of the importance of

other hydrogen oxidizing processes for hydrogen storage. A detailed calculation of the

number SSRM cells that could grow based on the N content in the Frigg reservoir can be

found in Text A.1. Text A.2 and A.3 hold an example calculation of the hydrogen

consumption per synthesized cell and the calculation of the potential hydrogen consumption

in a hydrogen storage system, respectively.

798 Author Contributions

799 The manuscript was written through contributions of all authors. All authors have given

approval to the final version of the manuscript.

801 **Declaration of interest**

802 The authors declare no competing financial interest.

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809 REFERENCES

Kharel S, Shabani B. Hydrogen as a long-term large-scale energy storage solution to
 support renewables. Energies 2018;11(10):1-17.

B12 2. Duan HX. The public perspective of carbon capture and storage for CO₂ emission
reductions in China. Energ Policy 2010;38(9):5281-9.

Beckingham LE, Winningham L. Critical knowledge gaps for understanding water rock-working phase interactions for compressed energy storage in porous formations. Acs

816 Sustain Chem Eng 2020;8(1):2-11.

817 4. Berta M, Dethlefsen F, Ebert M, Schafer D, Dahmke A. Geochemical effects of
818 millimolar hydrogen concentrations in groundwater: An experimental study in the context of
819 subsurface hydrogen storage. Environ Sci Technol 2018;52(8):4937-49.

5. Heinemann N, Booth MG, Haszeldine RS, Wilkinson M, Scafidi J, Edlmann K.

821 Hydrogen storage in porous geological formations - onshore play opportunities in the

Midland Valley (Scotland, UK). Int J Hydrogen Energy 2018;43(45):20861-74.

823 6. Zivar D, Kumar S, Foroozesh J. Underground hydrogen storage: A comprehensive
824 review. Int J Hydrogen Energy 2020; in press.

7. Gregory SP, Barnett MJ, Field LP, Milodowski AE. Subsurface microbial hydrogen

cycling: Natural occurrence and implications for industry. Microorganisms 2019;7(53):1-27.

827 8. Fredrickson JK, McKinley JP, Bjornstad BN, Long PE, Ringelberg DB, White DC, et

al. Pore-size constraints on the activity and survival of subsurface bacteria in a late cretaceous

shale-sandstone sequence, northwestern New Mexico. Geomicrobio J 1997;14(3):183-202.

830 9. Colman DR, Poudel S, Stamps BW, Boyd ES, Spear JR. The deep, hot biosphere:

831 Twenty-five years of retrospection. PNAS Perspective 2017;114(3):6895-903.

38

10. Lovley D, Chapelle FH. Deep subsurface microbial processes Rev Geophys
1995;33(3):365-81.

Hallbeck L, Pedersen K. Characterization of microbial processes in deep aquifers of
the Fennoscandian Shield. J Appl Geochem 2008;23:1796-819.

- 12. Payler SJ, Biddle JF, Lollar BS, Fox-Powell MG, Edwards T, Ngwenya BT, et al. An
- ionic limit to life in the deep subsurface. Front Microbiol 2019;10:1-16.
- Krumholz LR, McKinley JP, Ulrich GA, Suflita JM. Confined subsurface microbial
 communities in cretaceous rock. Nature 1997;386(6):64-6.
- 840 14. Methe BA, nelson KE, Eisen JA, Paulsen IT, Nelson W, Heidelberg JF, et al. Genome

841 of Geobacter sulfurreducens: Metal reduction in subsurface environments. Science

842 2003;302(5652):1967-9.

Aüllo T, Ranchou-Peyruse A, Ollivier B, Magot M. Desulfotomaculum spp. and
related gram-positive sulfate-reducing bacteria in deep subsurface environments. Front
Microbiol 2013;4:1-12.

Roh Y. Isolation and characterization of metal-reducing *Thermoanaerobacter* strains
from deep subsurface environments of the Piceance Basin, Colorado. Appl Environ

848 Microbiol 2002;68(2):6013-20.

849 17. Pedersen K. Microbial processes in radioactive waste disposal. Stockholm, Sweden;
850 2000.

18. Yin W, Wang Y, Liu L, He J. Biofilms: The microbial "protective clothing" in

- extreme environment. Int J Mol Sci 2019;20(3423):1-18.
- 19. Coombs P, Wagner D, Bateman K, Harrison H, Milodowski AE, Noy D, et al. The
- role of biofilms in subsurface transport processes Q J Eng Geol 2010;43:131-9.
- 855 20. Escudero C, Vera M, Oggerin M, Amils R. Active microbial biofilms in deep poor
- porous continental subsurface rocks. Nature Sci Rep 2018;8(1538):1-9.

Wu X, Pedersen K, Edlund J, Eriksson L, Aastroem M, Andersson AF, et al. Potential
for hydrogen oxidizing chemolithoautotrophic and diazotrophic populations to iniate biofilm
formation in oligotrophic, deep terrestrial subsurface waters. Microbiome 2017;5(37):1-13.

22. Daly RA, Roux S, Borton MA, Morgan DM, Johnston MD, Booker AE, et al. Viruses
control dominant bacteria colonizing the terrestrial deep biosphere after hydraulic fracturing.
Nat Microbiol 2019;4:352-61.

863 23. Frank YA, Kadnikov VV, Gavrilov SN, Banks D, Gerasimchuk AL, Mardanov AV,
864 et al. Stable and variable parts of microbial community in Siberian deep subsurface thermal

aquifer system revealed in a long-term monitoring study. Front Microbiol 2016;7:1-15.

Pannekens M, Kroll L, Mueller H, Mbow FT, Meckenstock RU. Oil reservoirs, an
exceptional habitat for microorganisms. N Biotechnol 2019;49:1-9.

25. Lerm S, Westphal A, Miethling-Graff R, Alawi M, Seibt A, Wolfgramm M, et al.

869 Thermal effects on microbial composition and microbiologically induced corrosion and

870 mineral precipitation affecting operation of a geothermal plant in a deep saline aquifer.

871 Extremophiles 2013;17(2):311-27.

Zgonnik V. The occurrence and geoscience of natural hydrogen: A comprehensive
review. Earth-Sci Rev 2020;203(103140):1-50.

874 27. Kleinitz W, Boehling E. Underground gas storage in porous media- operating

875 experience with bacteria on gas quality 67th EAGE Conference and Exhibition SPE

EUROPEC 13-16 June Madrid, Spain Society of Petroleum Engineers 2005 p. 1-6.

28. Loto CA. Microbiological corrosion: mechanism, control and impact—a review. Int J
Adv Manuf Tech 2017;92:4241-52.

Strobel G, Hagemann B, Huppertz TM, Ganzer L. Underground bio-methanation:
Concept and potential. Renew Sust Energ Rev 2020;123(109747):1-11.

30. Taylor SW, Jaffe PR. Biofilm growth and the related changes in the physical
properties of a porous medium 1. Experimental investigation. Water Resour Res
1990;26(9):2153-9.

884 31. Harris SH, Smith RL, Suflita JM. In situ hydrogen consumption kinetics as an
885 indicator of subsurface microbial activity. Fems Microbiol Ecol 2007;60(2):220-8.

32. Hoehler TM, Barker Joergensen B. Microbial life under extreme energy limitation.
Nat Rev 2013;11:83-94.

33. Maier RM, Pepper IL, Gerba CP. Environmental Microbiology Second Edition ed:
Academic Press 2009.

890 34. Parkes RJ, Sass H. Deep sub-surface In: Schaechter M, editor. Encyclopaedia of
891 Microbiology Elsevier Academic Press 2009.

35. Lovley D, Goodwin S. Hydrogen concentrations as an indicator of the predominant
terminal electron-accepting reactions in aquatic sediments. Geochim Cosmochim Acta
1988;52:2993-3003.

36. Heimann A, Jakobsen R, Blodau C. Energetic constraints on H₂-dependent terminal
electron accepting processes in anoxic environments: A review of observations and model
approaches. Environ Sci Technol 2010;44:24-33.

37. Appelo CAJ, Postma D. Geochemistry, groundwater and pollution second ed. Leiden
A.A.Balkema Publishers 2007.

900 38. Loeffler FE, Tiedje JM, Sanford RA. Fraction of electrons consumed in electron

901 acceptor reduction and hydrogen thresholds as indicators of halorespiratory physiology. Appl

902 Environ Microbiol 1999;65(9):4049-56.

903 39. Cord-Ruwisch R, Seitz HJ, Conrad R. The capacity of hydrogenotrophic anaerobic

bacteria to compete for traces of hydrogen depends on the redox potential of the terminal

electron acceptor. Arch Microbiol 1988;149:350-7.

41

- 40. Karadagli F, Rittmann BE. Kinetic characterization of *Methanobacterium bryantii*M.o.H. Environ Sci Technol 2005;39:4900-5.
- 41. Amend JP, Shock EL. Energetics of overall metabolic reactions of thermophilic and
 hyperthermophilic archaea and bacteria. FEMS Microbiol Rev 2001;25:175-243.
- 910 42. Machel HG. Bacterial and thermochemical sulfate reduction in diagenetic settings:
- old and new insights. Sediment Geol 2001;140:143-75.
- 912 43. Muyzer G, Stams AJM. The ecology and biotechnology of sulphate-reducing bacteria.
 913 Nat Rev 2008;6:441-54.
- 44. Findlay AJ. Microbial impact on polysulfide dynamics in the environment. Fems
- 915 Microbiol Lett 2016;363:1-12.
- 916 45. Hedderich R, Klimmek O, Kroeger A, Dirmeier R, Keller M, Stetter KO. Anaerobic
- respiration with elemental sulfur and with disulfides. FEMS Microbiol Rev 1999;22:353-81.
- 918 46. Rickard D, Luther GW. Chemistry of iron sulfides. Chem Rev 2007;107:514-62.
- 919 47. Lovley D. Dissimilatory Fe(III) and Mn(IV) reduction. Microbiol Rev
- 920 1991;55(2):259-87.
- 921 48. Hernsdorf AW, Amano Y, Miyakawa K, Ise K, Suzuki Y, Ananharaman K, et al.
- 922 Potential for microbial H₂ and metal transformations associated with novel bacteria and
- archaea in deep terrestrial subsurface sediments. Nature 2017;11:1915-29.
- 49. Javaherdashti R. Microbially Influenced Corrosion . An Engineering Insight second
 ed. Switzerland Springer; 2008.
- 926 50. Wiegel J, Hanel J, Aygen K. Chemolithoautotrophic thermophilic iron(III)-reducer.
- 927 In: Ljungdahl LG, Adams MW, Barton LL, Ferry JG, Johnson MK, editors. Biochemistry
- and physiology of anaerobic bacteria. New York: Springer; 2003. p. 235-51.

- 929 51. Ivanova AE, Borzenkov IA, Tarasov AL, Milekhina EI, Belyaev SS. A
- 930 microbiological study of an underground gas storage in the process of gas extraction.
- 931 Microbiology 2007;76:461-8.
- 932 52. Slobodkin AI, Jeanthnon C, L'Haridon S, Nazina T, Miroshnichenko M, Bonch-
- 933 Osmoloskaya EA. Dissimilatory reduction of Fe(III) by thermophilic bacteria and archaea in
- deep subsurface petroleum reservoirs of Western Siberia. Curr Microbiol 1999.
- 935 53. Kashefi K, Lovley D. Reduction of Fe(III), Mn(IV), and toxic metals at 100°C by
- 936 *Pyrobaculum islandicum*. Appl Environ Microbiol 2000;66(3):1050-6.
- 937 54. Jiang Y, Zhang B, He C, Shi J, Borthwick AGL, Huang X. Synchronous microbial
- 938 vanadium (V) reduction and denitrification in groundwater using hydrogen as the sole
- electron donor. Water Resour Res 2018;141:289-96.
- 940 55. Wisotzksy F, Eckert P. Sulfat-dominierter BTEX Abbau im Grundwasser eines
 941 ehemaligen Gaswerks-standortes. Grundwasser 1997;2:11-20.
- 942 56. Ranchou-Peyruse A, Gasc C, Guignard M, Auello T, Sequidt D, Peyret P, et al. The
- 943 sequence capture by hybridization: a new approach for revealing the potential of mono-
- 944 aromatic hydrocarbons bioattenuation in a deep oligotrophic aquifer. Microb Biotechnol
- 945 2017;10(2):469-79.
- 946 57. Zettlitzer M, Moeller F, Morozova D, Lokay P, Würdemann H. Re-establishment of
- 947 the proper injectivity of the CO₂-injection well Ktzi 201 in Ketzin, Germany. Int J Greenh
- 948 Gas Control 2010;4(6):952-9.
- 58. Bath A. Drilling fluid tracers Review and update of industry experience and issues
 for RWMD site characterisation programme. Loughborough, UK: Nuclear Decommissioning
 Authority, Radioactive Waste Management Directorate; 2011.

952 59. Gittel A, Soerensen KB, Skovhus KI, Schramm A. Prokaryotic community structure
953 and sulfate reducer activity in water from high-temperature oil reservoirs with and without
954 nitrate treatment. Appl Environ Microbiol 2009;75(22):7086-96.

- 60. Laturnus F, Lauritsen FR, Groen C. Chloroform in a pristine aquifer system: Toward
 an evidence of biogenic origin. Water Resour Res 2000;36(10):2999-3009.
- 957 61. Hoehener P, Werner D, Balsiger C, Pasteris G. Worldwide occurrence and fate of
 958 chlorofluorocarbons in groundwater. Crit Rev Environ Sci Technol 2003;33(1):1-29.
- 959 62. Grabowski A, Nercessian O, Fayolle F, Blanchet D, Jeanthon C. Microbial diversity
- 960 in production waters of a low-temperature biodegraded oil reservoir. Fems Microbiol Ecol
 961 2005;54:427-43.
- 962 63. Liu F, Conrad R. Chemolithotrophic acetogenic H₂/CO₂ utilization in Italian rice field
 963 soil. Isme J 2011;5:1526-39.
- 964 64. Kuesel K, Drake HL. Acetogens In: Thiel JRV, editor. Encyclopedia of Geobiology:
- 965 Springer Science+Business Media B.V; 2011.
- 966 65. Schuchmann K, Mueller V. Energetics and Application of Heterotrophy in
- 967 Acetogenic Bacteria. Appl Environ Microbiol 2016;82(14):4056-69.
- 968 66. Bengelsdorf FR, Beck MH, Erz C, Hoffmeister S, Karl MM, Riegler P, et al. Chapter
- 969 four- Bacterial anaerobic synthesis gas (syngas) and CO₂ + H₂ fermentation. Adv Appl
- 970 Microbiol 2018;103:143-221.
- 971 67. Esteve-Nunez A, Nunez C, Lovley DR. Preferential reduction of Fe(III) over fumarate
- 972 by Geobacter sulfurreducens. J Bacteriol 2004;186(9):2897-9.
- 973 68. Eecke HCV, Akerman NH, Huber JA, Butterfield DA, Holden JF. Growth kinetics
- and energetics of a deep-sea hyperthermophilic methanogen under varying environmental
- 975 conditions. Environ Microbiol Rep 2013;5(5):665-71.

69. Freitag TE, Prosser JI. Correlation of methane production and functional gene
transcriptional activity in a peat soil. Appl Environ Microbiol 2009;75(21).
70. Usher K, Kaksonen A, Bouquet D, Cheng KY, Geste Y, Chapman PG, et al. The role
of bacterial communities and carbon dioxide on the corrosion of steel. Corros Sci

980 2015;98:354-65.

981 71. Choong YY, Norli I, Abdullah AZ, Yhaya MF. Impacts of trace element

982 supplementation on the performance of anaerobic digestion process: A critical review.

983 Bioresour Technol 2016;209:369-79.

984 72. Pedersen K, Karlsson F. Investigations of subterranean microorganisms. Their
985 importance for performance assessment of radioactive waste disposal. Swedish Nuclear Fuel
986 and Waste Management Co; 1995.

987 73. Moench TT, Zeikus JG. Nutritional growth requirements for Butyribacterium

988 *methylotrophicum* on single carbon substrates and glucose. Curr Microbiol 1983;9:151-4.

989 74. Magot M, Basso O, Tardy-Jacquenod C, Caumette P. Desulfovibrio bastinii sp. nov.

and Desulfovibrio gracilis sp. nov., moderately halophilic, sulfate reducing bacteria isolated

from deep subsurface oilfield water. Int J Syst Evol Microbiol 2004;54:1693-7.

992 75. Steinsbu BO, Thorseth IH, Nagakawa S, Inagaki F, Lever MA, Engelen B, et al.

993 Archaeoglobus sulfaticallidus sp. nov., a thermophilic and facultatively lithoautotrophic

994 sulfate-reducer isolated from black rust exposed to hot ridge flank crustal fluids. Int J Syst

995 Evol Microbiol 2010;60:2745-52.

996 76. Casar CP, Kruger BR, Flynn TM, Masterson AL, Momper LM, Osburn MR. Mineral-

hosted biofilm communities in the continental deep subsurface, Deep Mine Microbial

998 Observatory, SD, USA. Geobiology 2020;18:508-22.

999 77. Samuels T, Bryce C, Landenmark H, Marie-Loudon C, Nicholson N, Stevens AH, et

1000 al. Microbial weathering of minerals and rocks in natural environments. In: Dontsova K,

1001	Balogh-Brunstad Z, Le Roux G, editors. Biogeochemical cycles: Ecological drivers and
1002	environmental impact: American Geophysical Union. John Wiley & Sons, Inc.; 2020.
1003	78. Wlodarczyk A, Lirski M, Fogtman A, Koblowska M, Bidzinski G, Matlakowska R
1004	The oxidative metabolism of fossil hydrocarbons and sulfide minerals by the lithobiontic
1005	microbial community inhabiting deep subterrestrial kupferschiefer black shale. Front
1006	Microbiol 2018;9(972):1-14.

1007 79. Napieralski S, Buss HL, Brantley SL, Lee S, Xu H, Roden EE. Microbial

1008 chemolithotrophy mediates oxidative weathering of granitic bedrock. PNAS

1009 2019;116(52):26394-401.

1010 80. Huang J, Sheng X-F, Xi J, Lin-Yan H, Huang Z, Wang Q, et al. Depth-related

1011 changes in community structure of culturable mineral weathering bacteria and in weathering

patterns caused by them along two contrasting soil profiles. Appl Environ Microbiol
2014;80(1):29-42.

1014 81. Alber BE. Autotrophic CO₂ metabolism. In: Schaechter M, editor. Encyclopedia of
1015 Microbiology Elsevier Academic Press; 2009.

1016 82. Liu S, Suflita JM. H₂-CO₂-Dependent anaerobic O-demethylation activity in

1017 subsurface sediments and by an isolated bacterium. Appl Environ Microbiol

1018 1993;59(5):1325-31.

1019 83. Kotelnikova S, Pedersen K. Evidence for methanogenic archaea and homoacetogenic
1020 bacteria in deep granitic rock aquifers. FEMS Microbiol Rev 1997;20:339-49.

1021 84. Londry KL, Jahnke LL, Des Marais DJ. Stable carbon isotope rations of lipid

1022 biomarkers and biomass for sulfate-reducing bacteria grown with different substrates.

1023 Goldschmidt Conference2001.

1024 85. Camacho A. Sulfur bacteria In: Likens GE, editor. Encyclopedia of inland waters 1:
1025 Elsevier Science 2009.

1026 86. Welsh DT, Bourges S, de Wit R, Herbert RA. Seasonal variations in nitrogen-fixation
1027 (acetylene reduction) and sulphate-reduction rates in the rhizosphere of Zostera noltii:

1028 Nitrogen fixation by sulphate-reducing bacteria Mar Biol 1996;125:619-28.

1029 87. Whitman WB, Bowen TL, Boone DR. The methanogenic bacteria. In: Balows A,

1030 Truper HG, Dworkin M, Harder W, Schleifer K-H, editors. The Prokaryotes New York:

1031 Springer-Verlag; 1992. p. 719-67.

1032 88. Drake HL. Acetogenesis London, United Kingdom Chapman & Hall; 2012.

1033 89. Kapili BJ, Barnett SE, Buckley DH, Dekas AE. Evidence for phylogenetically and

1034 catabolically diverse active diazotrophs in deep-sea sediment Isme J 2020;14:971-83.

1035 90. Herbert BN, Gilber PD, Stockdsle H, Watkinson RJ. Factors controlling the activity

1036 of sulphate-reducing bacteria In reservoirs during water injection. Society of Petroleum

1037 Engineers; 1985. Report No.: SPE-13978-MS Contract No.: SPE-13978-MS.

1038 91. Chen Y, Cheng JJ, Creamer KS. Inhibition of anaerobic digestion process: A review.
1039 Bioresour Technol 2008;99:4044-64.

1040 92. Okabe S. Rate and stoichiometry of sulfate reducing bacteria in suspended and

1041 biolfilm cultures. Montana, USA: Montana State University; 1992.

1042 93. Van Verseveld HW, Duine JA, editors. Proceedings of the 5th international

symposium on microbial growth on C₁ compounds. International Symposium on microbial

1044 growth on C₁ compounds; 1986; University of Groningen, The Netherlands.

1045 94. Taylor GT, Pirt SJ. Nutrition and factors limiting the growth of a methanogenic

- 1046 bacterium (Methanobacterium thermoautotrophicum) Arch Microbiol 1977;113:17-22.
- 1047 95. van Houten RT, Yun SY, Lettinga G. Thermophilic sulphate and sulphite reduction in

lab-scale gas-lift reactors using H₂ and CO₂ as energy and carbon source. Biotechnol Bioeng
1049 1997;55:807-14.

- 1050 96. Slobodkin AI, Slobodkina GB. Thermophilic prokaryotes from deep subterranean
 1051 habitat. Mikrobiologiya 2014;83(3):255-70.
- 1052 97. Basso O, Lascourreges JF, Le Borgne F, Le Goff C, Magot M. Characterization by
- 1053 culture and molecular analysis of the microbial diversity of a deep subsurface gas storage1054 aquifer. Res Microbiol 2009;160:107-9.
- 1055 98. Ranchou-Peyruse M, Auguet J-C, Maziere C, Restrepo-Ortiz CX, Guignard M,
- Dequidt D, et al. Geological gas-storage shapes deep life. J Environ Biol 2019;21(10):395364.
- 1058 99. Matos CR, Carneiro JF, Silva PP. Overview of large-scale underground energy
- 1059 storage technologies for integration of renewable energies and criteria for reservoir
- 1060 identification. J Energy Storage 2019;21:241-58.
- 1061 100. Shi Z, Jessen K, Tsotsis TT. Impacts of the subsurface storage of natural gas and
 1062 hydrogen mixtures Int J Hydrogen Energy 2020;45(15):f8757-73.
- 1063 101. Hassanpouryouzband A, Joonaki E, Edlmann K, Heinemann N, Yang J.
- 1064 Thermodynamic and transport properties of hydrogen containing streams. Sci Data
- 1065 2020;1(1):1-14.
- 1066 102. Yen TF. Microbial enhanced oil recovery: Principle and practice Boca Raton, Florida:
 1067 CRC Press; 1990.
- 1068 103. Salwan R, Sharma V. Physiological and biotechnological aspects of extremophiles.
- 1069 London, United Kingdom Elsevier 2020.
- 1070 104. Jaenicke R, Sterner R. Life at high temperatures. Prokaryotes 2006;2:167-209.
- 1071 105. Holden JF. Extremophiles: Hot Environments In: Schaechter M, editor.
- 1072 Encyclopaedia of Microbiology Elsevier Academic Press; 2009.

- 1073 106. Hosh S, Lepcha K, Basak A, Mahanty AK. Thermophiles and thermophilic
- hydrolases In: Salwan R, Sharma V, editors. Physiological and biotechnological aspects of
 extremophiles. London, United Kingdom Elsevier Acadamic Press; 2020.
- 1076 107. Miller JF, Nelson CM, Ludlow JM, Shah NN, Clark DS. High pressure-temperature
- 1077 bioreactor: assays of thermostable hydrogenase with fiber optics. Biotechnol Bioeng

1078 1989;34:1015-21.

- 1079 108. Takai K, Nakamura K, Toki T, Tsunogai U, Miyazaki M, Miyazaki J, et al. Cell
- 1080 proliferation at 122°C and isotopically heavy CH₄ production by a hyperthermophilic
- 1081 methanogen under high-pressure cultivation. PNAS 2008;105(31):10949-54.
- 1082 109. Pley U, Schipka J, Gambacorta A, Jannasch HW, Fricke H, Rachel R, et al.
- 1083 Pyrodictium abyssi sp. nov. represents a novel heterotrophic marine archaeal
- 1084 hyperthermophile growing at 110°C Syst Appl Microbiol 1991;14:245-53.
- 1085 110. Basen M, Geiger I, henke L, Mueller V. A genetic system for the thermophilic
- 1086 acetogenic bacterium *Thermoanaerobacter kivui*. Appl Environ Microbiol 2018;84(3):1-11.
- 1087 111. Leigh JA, Mayer F, Wolfe RS. Acetogenium kivui, a new thermophilic hydrogen-
- 1088 oxidizing, acetogenic bacterium. Arch Microbiol 1981;129:275-80.
- 1089 112. Balk M, Weijma J, Friedrich MW, Stams AJM. Methanol utilization by a novel
- 1090 thermophilic homoacetogenic bacterium, Moorella mulderi sp. nov., isolated from a
- 1091 bioreactor. Arch Microbiol 2003;179:315-20.
- 1092 113. Oren A. Life at high salt concentrations In: Dworkin M, Falkow S, Rosenberg E,
- 1093 Schleifer K-H, Stackebrandt E, editors. The Prokaryotes A handbook on the biology of
- 1094 bacteria Ecophysiology and Biochemistry. Volume 2. Singapore: Springer; 2006.
- 1095 114. Oren A. The bioenergetic basis for the decrease in metabolic diversity at increasing
- 1096 salt concentrations: implications for the functioning of salt lake ecosystems. Hydrobiologia

1097 2001;466(1-3):61-72.

- 1098 115. Ollivier B, Fardeau M-L, Cayol J-L, Magot M, Patel BKC, Prensier G, et al.
- 1099 *Methanocalculus halotolerans* gen. nov., sp. nov., isolated from an oil-producing well. Int J
- 1100 Syst Bacteriol 1998;48:821-8.
- 1101 116. Zhilina TN, Zavarzina DG, Kevbrin VV, Kolganova TV. Methanocalculus
- 1102 natronophilus sp. nov., a new alkaliphilic hydrogenotrophic methanogenic archaeon from a
- soda lake, and proposal of the new family Methanocalculaceae. Microbiology

1104 2013;82(6):698-706.

- 1105 117. Krekeler D, Sigalevich P, Teske A, Cypionka H, Cohen Y. A sulfate-reducing
- 1106 bacterium from the oxic layer of a microbial mat from Solar Lake (Sinai), Desulfovibrio
- 1107 oxyclinae sp. nov. Arch Microbiol 1997;167:369-75.
- 1108 118. Ollivier B, Hatchikian G, Guezennec J, Garcia J-L. Desulfohalobium retbaense gen.
- 1109 nov. sp. nov. a halophilic sulfate-reducing bacterium from sediments of a hypersaline lake in
- 1110 Senegal. Int J Syst Bacteriol 1991;41(1):74-81.
- 1111 119. Wood AP, Kelly DP. Isolation and characterisation of Thiobacillus halophilus sp.
- 1112 nov., a sulphur-oxidising autotrophic eubacterium from a Western Australian hypersaline
- 1113 lake. Arch Microbiol 1991;156:277-80.
- 1114 120. Jakobsen RF, Kjeldsen KU, Ingvordsen K. Desulfohalobium utahense sp. nov., a
- 1115 moderately halophilic, sulfate-reducing bacterium isolated from Great Salt Lake. Int J Syst
- 1116 Evol Microbiol 2006;56:2063-9.
- 1117 121. Rosenberg E, DeLong EF, Lory S, Stackebrankt E, Thompson F. The Prokaryotes.
- 1118 Firmicutes and Tenericutes fourth edition ed. Heidelberg: Springer Science and Business

1119 Media 2014.

- 1120 122. Zhilina TN, Detkova EN, Rainey FA, Osipov GA, Lysenko AM, Kostrikina NA, et al.
- 1121 *Natronoincola histidinovorans* gen. nov., sp. nov., a new alkaliphilic acetogenic anaerobe.
- 1122 Curr Microbiol 1998;37:177-85.

- 1123 123. Zeikus JG, Hegge PW, Thompson TE, Phelps TJ, Langworthy TA. Isolation and
- description of Haloanaerobium praevalens gen. nov. and sp. nov., an obligately anaerobic
- halophile common to Great Salt Lake sediments. Curr Microbiol 1983;9:225-34.
- 1126 124. Motamedi M, Karland O, Pedersen K. Survival of sulfate reducing bacteria at
- different water activities in compacted bentonite. Fems Microbiol Lett 1996;141(1):83-7.
- 1128 125. Kushner DJ. Microbial life in extreme environments. London: Academic Press 1978.
- 1129 126. Hallsworth JE, Yakimov MM, Golyshin PN, Gillion JLM, D'Auria G, de Lima ALves
- 1130 F, et al. Limits of life in MgCl₂-containing environments: chaotropicity defines the window.
- 1131 Environ Microbiol 2007;9(3):801-13.
- 1132 127. Steinle L, Knittel K, Felber N, Casalino C, de Lange G, Tessarolo C, et al. Life on the
- 1133 edge: active microbial communities in the Kryos MgCl₂- brine basin at very low water
- 1134 activity. Isme J 2018;12:1414-26.
- 1135 128. Cray JA, Russell JT, Timson DJ, Singhai RS, Hallsworth JE. A universal measure of
 1136 chaotropicity and kosmotropicity. Environ Microbiol 2013;15(1):287-96.
- 1137 129. Yuan H, Chen Y, Zhang H, Jiang S, Zhou Q, Gu G. Improved Bioproduction of
- 1138 Short-Chain Fatty Acids (SCFAs) from Excess Sludge under Alkaline Conditions. Environ
- 1139 Sci Technol 2006;40:2025-9.
- 1140 130. O'Flatherty V, Mahony T, O'Kennedy R, Colleran E. Effect of pH on growth kinetics
- 1141 and sulphide toxicity of a range of methanogenic, synthrophic and sulphate-reducing bacteria.
- 1142 Process Biochem 1998;33(5):555-69.
- 1143 131. Sorokin DY, Abbas B, Merkel AY, Riipstra EIC, Sinninghe Samste JS, Sukhacheva
- 1144 MV, et al. *Methanosalsum natronophilum* sp. nov., and *Methanocalculus alkaliphilus* sp.
- 1145 nov.,haloalkaliphilic methanogens from hypersaline soda lakes. Int J Syst Evol Microbiol
 1146 2015;65:3739-45.

- 1147 132. Stewart LC, Jung J-H, Kim Y-T, Kwon S-W, Park C-S, Holden JF.
- 1148 Methanocaldococcus bathoardescens sp. nov., a hyperthermophilic methanogen isolated
- 1149 from a volcanically active deep-sea hydrothermal vent. Int J Syst Evol Microbiol
- 1150 2015;65:1280-3.
- 1151 133. Takai K, Inoue A, Horikoshi K. Methanothermococcus okinawensis sp. nov., a
- 1152 thermophilic, methane-producing archaeon isolated from a Western Pacific deep-sea
- 1153 hydrothermal vent system. Int J Syst Evol Microbiol 2002;51:1089-95.
- 1154 134. Ganzert L, Schirmack J, Alawi M, Mangelsdorf K, Sand W, Hillebrand-Voiculescu
- 1155 A, et al. *Methanosarcina spelaei* sp. nov., a methanogenic archaeon isolated from a floating
- biofilm of a subsurface sulphurous lake. Int J Syst Evol Microbiol 2014;64:3478–84.
- 1157 135. Braeuer SL, Cadillo-Quiroz H, Ward RJ, Yavitt JB, Zinder SH. Methanoregula
- 1158 boonei gen. nov., sp. nov., an acidiphilic methanogen isolated from an acidic peat bog. Int J
- 1159 Syst Evol Microbiol 2011;61:45-52.
- 1160 136. Zhilina TN, Zavarzin GA, Rainey FA, Pikuta EN, Osipov GA, Kostrikina NA.
- 1161 Desulfonatronovibrio hydrogenovorans gen. nov., sp. nov., an alkaliphilic, sulfate-reducing
- 1162 bacterium. Int J Syst Bacteriol 1997;47(1):144-9.
- 1163 137. Sorokin DY, Muyzer G. *Desulfurispira natronophila* gen. nov. sp. nov.: an obligately
- anaerobic dissimilatory sulfur-reducing bacterium from soda lakes. Extremophiles
- 1165 2010;14:349-55.
- 1166 138. Nga DP, Ha DTC, Hien LT, Stan-Lotter H. Desulfovibrio vietnamensis sp.nov., a
- 1167 halophilic sulfate-reducing bacterium from Vietnamese oil fields. Anaerobe 1996;2:385-92.
- 1168 139. Hallberg KB, Lindstroem EB. Characterization of *Thiobacillus caldus* sp. nov., a
- 1169 moderately thermophilic acidophile. Microbiology 1994;140:3451-1456.

- 1170 140. Segerer A, Neuner A, Kristjansson JK, Stetter KO. Acidianus infernus gen. nov. sp.
- 1171 nov. and Acidianus brierleyi comb. nov.: facultatively aerobic, extremely acidophilic
- thermophilic sulfur-metabolizing archaebacteria. Int J Syst Bacteriol 1986;36(4):559-64.
- 1173 141. Segerer A, Trincone A, Gahrtz M, Stetter KO. Stygiolobus azoricus gen. nov., sp.
- 1174 nov. represents a novel genus of anaerobic, extremely thermoacidophilic archaebacteria of
- 1175 the order *Sulfolobales*. Int J Syst Bacteriol 1991;41(4):495-501.
- 1176 142. Fliermans CB, Brock TD. Ecology of sulfur-oxidizing bacteria in hot acid soils. J
 1177 Bacteriol 1972;111(2):343-50.
- 1178 143. Schnuerer A, Schink B, Svensson BH. Clostridium ultunense sp. nov., a mesophilic
- 1179 bacterium oxidizing acetate in syntrophic association with a hydrogenotrophic methanogenic
- 1180 bacterium. Int J Syst Bacteriol 1996;46(4):1145-52.
- 1181 144. Inokuma K, Nakashimada Y, Akahoshi T, Nishio N. Characterization of enzymes
- involved in the ethanol production of Moorella sp. HUC22-1. Arch Microbiol 2007;188:37-45.
- 1184 145. Hirano S, Masuda N. Characterization of NADP-Dependent 7, B-Hydroxysteroid
- 1185 Dehydrogenases from Peptostreptococcus productus and *Eubacterium aerofaciens*. Appl
- 1186 Environ Microbiol 1982;43(5):1057-63.
- 1187 146. Zhilina TN, Zavarzina DG, Detkova EN, Patutina EO, Kuznetsov BB. Fuchsiella
- 1188 ferrireducens sp. nov., a novel haloalkaliphilic, lithoautotrophic homoacetogen capable of
- 1189 iron reduction, and emendation of the description of the genus *Fuchsiella*. Int J Syst Evol
- 1190 Microbiol 2015;85:2432-40.
- 1191 147. Wiegel J, Braun M, Gottschalk G. *Clostridium thermoautotrophicum species novum*,
- 1192 a thermophile producing acetate from molecular hydrogen and carbon dioxide. Curr
- 1193 Microbiol 1981;5:255-60.

- 1194 148. Tanner RS, Miller LM, Yang D. Clostridium ljungdahlii sp. nov., an acetogenic
- species in clostridial rRNA homology group I Int J Syst Bacteriol 1993;43(2):232-6.

1196 149. Kaneuchi C, Benno Y, Mitsuoka T. Clostridium coccoides, a new species from the

1197 feaces of mice. Int J Syst Bacteriol 1976;26(4):482-6.

- 1198 150. Kuesel K, Dorsch T, Acker G, Stackebrandt E, Drake HL. Clostridium scatologenes
- 1199 strain SL1 isolated as an acetogenic bacterium from acidic sediments. Int J Syst Evol
- 1200 Microbiol 2000;50:537-46.
- 1201 151. Gößner AS, Picardal F, Tanner RS, Drake HL. Carbon metabolism of the moderatey
- 1202 acid-tolerant acetogen Clostridium drakai isolated from peat. Fems Microbiol Lett
- 1203 2008;287:236-42.
- 1204 152. Abe F, Kato C, Horikoshi K. Pressure-regulated metabolism in microorganisms.
 1205 Trends Microbiol 1999;7(11):447-53.
- 1206 153. Vikromvarasiri N, S. B, Pisutpaisal N. Comparative performance of Halothiobacillus

Neapolitanus and *Paracoccus Pantotrophus* in sulphur oxidation. Energy Procedia
2015;79:885-9.

- 1209 154. Cao J, Gayet N, Zeng X, Shao Z, Jebbar M, Alain K. Pseudodesulfovibrio indicus
- 1210 gen. nov., sp. nov., a piezophilic sulfate-reducing bacterium from the Indian Ocean and
- 1211 reclassification of four species of the genus Desulfovibrio. Int J Syst Evol Microbiol
- 1212 2016;66:3904-11.
- 1213 155. Takai K, Miyazaki M, Hirayama H, Nakagawa S, Querellou J, Godfroy A. Isolation
- 1214 and physiological characterization of two novel, piezophilic, thermophilic
- 1215 chemolithoautotrophs from a deep-sea hydrothermal vent chimney. J Environ Biol

1216 2009;11(8):1983-97.

- 1217 156. Fichtel K, Logemann J, Fichtel J, Rullkoetter J, Cypionka H, Engelen B. Temperature
 1218 and pressure adaptation of a sulfate reducer from the deep subsurface. Front Microbiol 2015
 1219 6:1-13.
- 1220 157. Kurr M, Huber R, Koenig H, Jannasch HW, Fricke H, Trincone A, et al.
- 1221 Methanopyrus kandleri, gen. and sp. nov. represents a novel group of hyperthermophilic
- 1222 methanogens, growing at 110 ° C*. Arch Microbiol 1991;156:239-47.
- 1223 158. Ntagia E, Chatzigiannidou I, Williamson AJ, Arends JBA, Rabaey K.
- 1224 Homoacetogenesis and microbial community composition are shaped by pH and total sulfide
- 1225 concentration. Microb Biotechnol 2020;13(4):1026-38.
- 1226 159. Suleimenov OM, Seward TM. A spectrophotometric study of hydrogen sulphide
- ionisation in aqueous solutions to 350°C. Geochim Cosmochim Acta 1997;61(24):5187-98.
- 1228 160. Kushkeyvych I, Dordevic D, Vitezova M. Toxicity of hydrogen sulfide toward
- sulfate-reducing bacteria *Desulfovibrio piger Vib-7*. Arch Microbiol 2019;201(3):389-97.
- 1230 161. Winfrey MR, Zeikus JG. Effect of sulfate on carbon and electron flow during
- 1231 microbial methanogenesis in freshwater sediments. Appl Environ Microbiol 1977;33(2):275-
- 1232 81.
- 1233 162. Choi E, Rim JM. Competition and inhibition of sulfate reducers and methane
- 1234 producers in anaerobic treatment. Water Sci Technol 1991;23:1259-64.
- 1235 163. Dupraz S, Fabbri A, Joulian C, Dictor M-C, Battaglia-Brunet F, Menez B, et al.
- 1236 Impact of CO₂ concentration on autotrophic metabolisms and carbon fate in saline aquifers –
- 1237 A case study. Geochim Cosmochim Acta 2013;119:61-76.
- 1238 164. Froestl JM, Seifritz C, Drake HL. Effect of nitrate on the autotrophic metabolism of
- 1239 the acetogens *Clostridium thermoautotrophicum* and *Clostrium thermoaceticum* J Bacteriol
- 1240 1996;178(15):4597–603.

- 1241 165. Wang H, Zhang Y, Angelidakis I. Ammonia inhibition on hydrogen enriched
- 1242 anaerobic digestion of manure under mesophilic and thermophilic conditions. Water Resour1243 Res 2016;105:314-9.
- 1244 166. Conrad R. Contribution of hydrogen to methane production and control of hydrogen
- 1245 concentrations in methanogenic soils and sediments. Fems Microbiol Ecol 1999;28:193-202.
- 1246 167. Barth T, Riis M. Interactions between organic acid anions in formation waters and
- 1247 reservoir mineral phases. Org Geochem 1992;19(4-6):455-82.
- 1248 168. Oil and gas field data from the North Sea [Internet]. Oil and Gas Authority 2020
- 1249 [cited 19.5.2020]. Available from: <u>https://www.ogauthority.co.uk/data-centre/</u>.
- 1250 169. Hoehler TM, Albert DB, Alperin MJ, Martens CS. Acetogenesis from CO₂ in an
- anoxic marine sediment. Limnol Oceanogr 1999;44(3):662-7.
- 1252 170. Pedersen K. Microbial life in deep granitic rock. Fems Microbiol Ecol 1997;20:3991253 414.
- 1254 171. Breznak JA. Acetogenesis from carbon dioxide in termite guts. In: H.L. D, editor.
- 1255 Acetogenesis Chapman & Hall Microbiology Series (Physiology / Ecology / Molecular
- 1256 Biology / Biotechnology). Boston, MA: Springer; 1994.
- 1257 172. Krumholz LR, Harris SH, Tay ST, Suflita JM. Characterization of two subsurface H₂-
- 1258 utilizing bacteria, Desulfomicrobium hypogeium sp. nov. and Acetobacterium
- 1259 *psammolithicum* sp. nov., and their ecological role. Appl Environ Microbiol
- 1260 1999;65(6):2300-6.
- 1261 173. Phillips J. Extracellular electron uptake by acetogenic bacteria: Does H₂ consumption
- 1262 favor the H₂ evolution reaction on a cathode or metallic iron? Front Microbiol 2020;10:1-13.
- 1263 174. Robinson JA, Tiedje JM. Competition between sulfate-reducing and methanogenic
- bacteria for H₂ under resting and growing conditions. Arch Microbiol 1984;137:26-32.

- 1265 175. Fu B, Jin X, Conrad R, Liu H, Liu H. Competition between chemolithotrophic
- acetogenesis and hydrogenotrophic methanogenesis for exogenous H₂/CO₂ in anaerobically
 digested sludge. Front Microbiol 2019;10:1-9.
- 1268 176. Kotsyurbenko OR, Glagolev MV, Nozhevnikova AN, Conrad R. Competition
- 1269 between homoacetogenic bacteria and methanogenic archaea for hydrogen at low
- 1270 temperature. Fems Microbiol Ecol 2001;38:153-9.
- 1271 177. Lever MA. Acetogenesis in the energy-starved deep biosphere–a paradox? Front
 1272 Microbiol 2012;2:1-18.
- 1273 178. Lackner N, Wagner AO, Ilmer P. Effect of sulfate addition on carbon flow and
- 1274 microbial community composition during thermophilic digestion of cellulose. Appl Microbiol
- 1275 Biotechnol 2020;104:4605-15.
- 1276 179. Men Y, Feil H, VerBerkmoes NC, Shah MB, Johnson DR, Lee PKH, et al.
- 1277 Sustainable syntrophic growth of Dehalococcoides ethenogenes strain 195 with Desulfovibrio
- 1278 vulgaris Hildenborough and Methanobacterium congolense: global transcriptomic and
- 1279 proteomic analyses. Isme J 2012;6:410-2.
- 1280 180. Freedman AJE, BoonFei T, Thompson JR. Microbial potential for carbon and nutrient
- 1281 cycling in a geogenic supercritical carbon dioxide reservoir. Environ Microbiol
- 1282 2017;19:2228-45.
- 1283 181. Topcuoglu BD, Meydan C, Nguyen TB, Lang SQ, Holden JF. Growth kinetics,
- 1284 carbon isotope fractionation, and gene expression in the hyperthermophile
- 1285 Methanocaldococcus jannaschii during hydrogen-limited growth and interspecies hydrogen
- 1286 transfer. Appl Environ Microbiol 2019;85(9):1-14.
- 1287 182. Stetter KO, Huber R, Bloechl E, Kurr M, Eden RD, Fielder M, et al.
- 1288 Hyperthermophilic archaea are thriving in deep North Sea and Alaskan oil reservoirs Nature
- 1289 1993;365:743-5.

1290 183. Bonch-Osmoloskaya EA, Miroshnichenko ML, Lebedinsky AV, Chernyh TN, Nazina

1291 TN, Ivoilov VS, et al. Radioisotopic, culture-based, and oligonucleotide microchip analyses

1292 of thermophilic microbial communities in a continental high-temperature petroleum reservoir.

1293 Appl Environ Microbiol 2003;69(10):6143–51.

1294 184. Tian H, Gao P, Chen Z, Li Y, Li Y, Wang Y, et al. Compositions and abundances of

sulfate-reducing and sulfur-oxidizing microorganisms in water-flooded petroleum reservoirs

1296 with different temperatures in China. Front Microbiol 2017;8(143):1-14.

1297 185. Auello T, Berlendis S, Lascourreges JF, Dessort D, Duclerc D, Saint-Laurent S, et al.

New bio-indicators for long term natural attenuation of monoaromatic compounds in deep
terrestrial aquifers. Front Microbiol 2016;7(122):1-16.

1300 186. Mori K, Tsurumaru H, Harayama S. Iron corrosion activity of anaerobic hydrogen-

1301 consuming microorganisms isolated from oil facilities. J Biosci Bioeng 2010;110(4):426-30.

1302 187. Okpala GN, Chen C, Fida T, Voordouw G. Effect of thermophilic nitrate reduction on

1303 sulfide production in high temperature oil reservoir samples. Front Microbiol 2017;8:1-13.

1304 188. Balk M, Mehboob F, Gelder A, Riipstra WIC, Sinninghe Damste JS, Stams AJM.

1305 (Per)chlorate reduction by an acetogenic bacterium, Sporomusa sp., isolated from an

1306 underground gas storage. Appl Microbiol Biotechnol 2010;88:595-603.

1307 189. Berlendis S, Lascourreges JF, Schraauwers B, Sivadon P, Magot M. Anaerobic

1308 biodegradation of BTEX by original bacterial communities from an underground gas storage

1309 aquifer. Environ Sci Technol 2010;44:3621-8.

1310 190. Conrad R, Schuetz H, Babbel M. Temperature limitation of hydrogen turnover and
1311 methanogenesis in anoxic paddy soil. Fems Microbiol Ecol 1987;45:281-9.

1312 191. Stewart LC, Algar CK, Fortunato CS, Larson BI, Vallino JJ, Huber JA, et al. Fluid

1313 geochemistry, local hydrology, and metabolic activity define methanogen community size

and composition in deep-sea hydrothermal vents. Isme J 2019;13(7):1711-21.

- 1315 192. Enoki M, Shinzato N, Sato H, Nakamura K, Y. K. Comparative proteomic analysis of
- 1316 Methanothermobacter themautotrophicus DH in pure culture and in co-culture with a
- 1317 butyrate-oxidizing bacterium. Plos One 2011;6(8):1-10.
- 1318 193. Mukhopadhyay B, Johnson EF, Wolfe RS. A novel pH₂ control on the expression of
- 1319 flagella in the hyperthermophilic strictly hydrogenotrophic methanarchaeaon Methanococcus
- 1320 jannaschii. PNAS 2000;97(21):11522-7.
- 1321 194. Miller JF, Shah NN, Nelson CM, Ludlow JM, Clark DS. Pressure and temperature
- 1322 effects on growth and methane production of the extreme thermophile *Methanococcus*
- 1323 *jannaschi*. Appl Environ Microbiol 1988;54(12):3039-42.
- 1324 195. Schieche D, Murty MVS, Kermode RI, Bhattacharyya D. Biohydrogenation of
- 1325 fumarate using Desulfovibrio desulfuricans: Experimental results and kinetic rate modelling.
- 1326 J Chem Technol Biotechnol 1997;70(3):316-22.
- 1327 196. Baas Becking LGM. Geobiologie of inleiding tot de milieukunde. Den Hague, the
- 1328 Netherlands: W.P. Van Stockum & Zoon; 1934.
- 1329 197. Xu Y, Teng Y, Wang XB, Li R, Christie P. Exploring bacterial community structure
- and function associated with ploychlorinated biphenyl biodegradation in two hydrogen-
- 1331 amended soils Sci Total Environ 2020;745(140839):1-12.
- 1332 198. Braga Nan L, Trably E, Santa-Catalina G, Bernet N, Delgenes J-P, Escudie R.
- 1333 Biomethanation processes: new insights on the effect of a high H₂ partial pressure on
- 1334 microbial communities. Biotechnol Biofuels 2020;13(141):1-17.
- 1335 199. Treu L, Kogias PG, de Diego-Diaz B, Campanaro S, Bassani I, Fernandez-Rodriguez
- 1336 J, et al. Two-year microbial adaptation during hydrogen-mediated biogas upgradingprocess in
- a serial reactor configuration. Bioresour Technol 2018;264(140-147).

- 1338 200. Puente-Sánchez F, Arce-Rodríguez A, Oggerind M, García-Villadangosa M, Moreno-
- Paza M, Blanco Y, et al. Viable cyanobacteria in the deep continental subsurface. PNAS
 2018;115(42):10702-7.
- 1341 201. Gluyas JG, Hichens HM. The United Kingdom oil and gas fields commemorative
- 1342 millennium volume Gluyas JG, Hichens HM, editors: Memoirs of the Geological Society of

1343 London; 2003.

- 1344 202. Zhang Y, Zhang Z, Suzuki K, Maekawa T. Uptake and mass balance of trace metals
 1345 for methane producing bacteria. Biomass Bioenerg 2003;25:427–33.
- 1346 203. Scherer P, Lippert H, Wolff G. Composition of the major elements and trace-elements
- 1347 of 10 methanogenic bacteria determined by inductively coupled plasma emission-
- 1348 spectrometry. Biol Trace Elem Res 1983;5(3):149-63.
- 1349 204. Amid A, Mignard D, Wilkinson M. Seasonal storage of hydrogen in a depleted
 1350 natural gas reservoir. Int J Hydrogen Energy 2016;41(12):5549-58.
- 1351 205. Kettle H, Louis P, Holtrop G, Duncan SH, Flint HJ. Modelling the emergent
- dynamics and major metabolites of the human colonic microbiota. Environ Microbiol
 2015;17(5):1615-30.
- 1354 206. Littlewood D, Postgate JR. On the osmotic behaviour of Desulphovibrio
- 1355 *desulphuricans* J Gen Microbiol 1957;17:378-89.
- 1356 207. Bratbak G, Dundas I. Bacterial dry matter content and biomass estimations. Appl
 1357 Environ Microbiol 1984;48(4):755-7.
- 1358 208. Würdemann H, Halm H, Lerm S, Kleyböcker A. Verbund-Forschungsvorhaben
- 1359 H2STORE: Untersuchung der geohydraulischen, mineralogischen, geochemischen und
- 1360 biogenen Wechselwirkungen bei der Untertage-Speicherung von H2 in konvertierten
- 1361 Gaslagerstätten: Teilprojekt 4- Mikrobiologie: Abschlussbericht: Berichtszeitraum:

1362 01.08.2012 bis 31.12.2015. Potsdam: Helmholtz-Zentrum Potsdam Deutsches

1363 GeoForschungsZentrum GFZ; 2016.

- 1364 209. Nazina TN, Pavlova NK, Tatarkin YV, Shestakova NM, Babich TL, Sokolova DS, et
- al. Microorganisms of the carbonate petroleum reservoir 302 of the Romashkinskoe oilfield
- and their biotechnological potential. Microbiology 2013;82(2):190-200.
- 1367 210. Stolten D, Emonts B. Hydrogen Science and Engineering: Materials, Processes,
- 1368 Systems, and Technology, 2 Volume Set: Wiley-VCH; 2016.
- 1369 211. Bauer S. Underground Sun Storage. Final Report Vienna, Austria; 2017.
- 1370 212. Smigan P, Greksak M, Kozankova J, Buzek F, Onderka V, Wolf I. Methanogenic
- 1371 bacteria as a key factor involved in changes of town gas stored in an underground reservoir.
- 1372 Fems Microbiol Ecol 1990;73(3):221-4.
- 1373 213. Fagerbakke KM, Heldal M, Norland S. Content of carbon, nitrogen, oxygen, sulfur
- and phosphorus in native aquatic and cultured bacteria. Aquat Microb Ecol 1996;10 15-27.
- 1375 214. Mitchell AC, Phillips AJ, Hiebert R, Gerlach R, Spangler LH, Cunningham AB.
- 1376 Biofilm enhanced geologic sequestration of supercritical CO₂. Int J Greenh Gas Control
- 1377 2009;3:90-9.

1378

1379