

High resolution, annual maps of the characteristics of smallholder-dominated croplands at national scales

Lyndon D. Estes*¹, Su Ye^{1,2}, Lei Song¹, Boka Luo^{1,3}, J. Ronald Eastman^{1,3}, Zhenhua Meng¹, Qi Zhang¹, Dennis McRitchie⁴, Stephanie R. Debats⁴, Justus Muhando⁵, Angeline H. Amukoa⁵, Brian W. Kaloo⁵, Jackson Makuru⁵, Ben K. Mbatia⁵, Isaac M. Muasa⁵, Julius Mucha⁵, Adelide M. Mugami⁵, Judith M. Mugami⁵, Francis W. Muinde⁵, Fredrick M. Mwawaza⁵, Jeff Ochieng⁵, Charles J. Oduol⁵, Purent Oduor⁵, Thuo Wanjiku⁵, Joseph G. Wanyoike⁵, Ryan B. Avery⁶, Kelly K. Caylor^{6,7,8}

¹Graduate School of Geography, Clark University, Worcester, MA, USA

²Department of Natural Resources and the Environment, University of Connecticut, Storrs, CT, USA

³Clark Labs, Clark University, Worcester, MA, USA

⁴Independent contributor

⁵SpatialCollective, Nairobi, Kenya

⁶Department of Geography, University of California Santa Barbara, Santa Barbara, CA, USA

⁷Earth Research Institute, University of California Santa Barbara, Santa Barbara, CA, USA

⁸Bren School of Environmental Science and Management, University of California Santa Barbara, Santa Barbara, CA, USA

* corresponding author: lestes@clarku.edu

This pre-print not yet undergone peer review. It has been submitted to the *ISPRS Journal of Photogrammetry and Remote Sensing*. This version will be updated as it is revised, and the final published version will be accessible through its DOI link.

Abstract

Understanding agricultural change requires reliable, frequently updated maps that describe the characteristics of croplands. Such data are often unavailable for regions dominated by smallholder agricultural systems, which are particularly challenging for remote sensing. To overcome these challenges, we designed a system to minimize several sources of error that arise when mapping smallholder croplands. To overcome errors caused by mismatches between image resolution and cropland scales, as well as persistent cloud cover, the system converts daily, 3.7 m PlanetScope imagery into two seasonal composites within a single agricultural year. To reduce errors that occur when training classifiers, we built a labelling platform that rigorously assesses label accuracy, and creates more accurate consensus labels that train a Random Forests model. The labelling platform and model interact within an active learning process that boosts the accuracy of the resulting cropland probability map, which is used in a segmentation process to delineate individual field boundaries. We applied this system to map Ghana's croplands for the year 2018. We divided Ghana into 16 mapping regions (12,160-23,535 km²), training separate models for each using a total of 6,299 labels, plus 1,600 for validation. Using an independent map reference sample (n=1,207), we found that overall accuracies of the resulting cropland probability and field boundary maps were 88% and 86.7%, respectively, with User's accuracies for the cropland class of 61.2% and 78.9%, and Producer's accuracies of 67.3% and 58.2%. Croplands covered 16.1-23.2% of the mapped area, comprising 1,131,146 total fields with an average size of 3.92 ha. Estimates based on the map reference sample indicate the cropland percentage is 17.1% (15.4-18.9%) or 17.6% (15.6-19.6%), depending on the map used to estimate the standard error. Using the labellers' digitized field boundaries to estimate biases in field boundary statistics, we calculated an adjusted mean field size of 1.73 ha and total field count of 1,662,281. Although the cropland class contained substantial errors, the system was effective in mitigating error and quantifying resulting performance gains. By minimizing training errors, consensus labelling improved the model's F1 scores by up to 25%, while 3 iterations of active learning increased the F1 score by 9.1%, on average, which was 2.3% higher than training models with randomly selected labels. Map accuracy can be improved by replacing Random Forests with a convolutional neural network. These results demonstrate a readily adapted, transferrable framework for developing high resolution, annual, nation-scale maps that provide important details about smallholder-dominated croplands.

1 Introduction

Amidst all the challenges posed by global change, a particular concern is how agricultural systems will adapt to meet humanity's growing food demands, and the impacts that transforming and expanding food systems will have on societies, economies, and the environment (Searchinger et al. 2019). Significant efforts are being made to address the various aspects of this challenge, including work on diagnosing and closing yield gaps (Lobell et al. 2009, e.g. Licker et al. 2010, Mueller et al. 2012), expanding and commercializing production (Morris and Byerlee 2009), and to understand (Rulli and D'Odorico 2014, Kehoe et al. 2017, Davis et al. 2020) and mitigate (Estes et al. 2016b) agriculture's

40 ecological impacts. Answering many of the questions these efforts seek to address depends on reliable
41 data that describes the location and characteristics of cropland (Fritz et al. 2015), and how these are
42 changing over time. Unfortunately, the data that do exist are in many places inaccurate. Existing
43 estimates of how much global cropland there is tend to vary widely, and they often disagree about
44 where cropland is located (e.g. Fritz et al. 2011, 2013). Such errors can propagate in subsequent
45 analyses that rely on cropland data as inputs, resulting in potentially misleading answers (Estes et al.
46 2018). Beyond cropland distributions, few data are available on key cropland characteristics such as
47 field size, an important variable needed to estimate yield and other key food security variables (Carletto
48 et al. 2015), and as an indicator of farm size (Levin 2006, Samberg et al. 2016), a critical component of
49 rural livelihoods given increasing population densities and longstanding debates about the relationship
50 between farm size and productivity (Feder 1985, Carletto et al. 2013, Desiere and Jolliffe 2018).

51 These informational inadequacies are due to the fact that cropland data in much of the world are
52 derived from remotely sensed landcover maps, which can be notoriously high in error, particularly over
53 regions such as Africa (Fritz et al. 2010, Estes et al. 2018), where agricultural changes will be largest
54 and the need for accurate baseline data is thus greatest (Searchinger et al. 2015, Estes et al. 2016b,
55 Bullock et al. 2021). Cropland mapping over Africa is difficult for several reasons. The primary reason
56 relates to the characteristics of the continent's smallholder-dominated croplands, where half of all fields
57 are smaller than 1 ha (Lesiv et al. 2019). This size is small relative to the 30-250 m resolution of the
58 sensors typically used in many landcover mapping efforts (e.g. Chen et al. 2015, Sulla-Menashe et al.
59 2019), which results in errors due to mixed pixels and aspects of the modifiable area unit problem
60 (Openshaw and Taylor 1979). In the latter case, the pixel's shape may be poorly matched to that of
61 cropland, and is too coarse to aggregate to approximate that shape at the characteristic scales of crop
62 fields (Dark and Bram 2007, Estes et al. 2018). On top of the matter of scale is 1) high intra-class
63 variability of the cropland class, compounded by the fact that these particular croplands can be heavily
64 intergraded with surrounding vegetation (Debats et al. 2016, Estes et al. 2016a), and 2) the substantial
65 temporal variability within croplands, both within and between seasons. These latter two aspects pose
66 challenges for the classification algorithms that are applied to the imagery.

67 These problems arising from cropland characteristics are increasingly being addressed due to
68 technological advances. Recent advances in satellite technology have increased the coverage of high (<5

69 m) or near-high (10 m) resolution imagery with weekly to near-daily return intervals (Drusch et al.
70 2012, McCabe et al. 2017). This high spatial *and* temporal resolution addresses the sensor-field scale
71 mismatch, and more effectively captures the intra-seasonal dynamics of cropland, which helps classifiers
72 distinguish cropland from surrounding cover types (Debats et al. 2016, Defourny et al. 2019). On top
73 of this, advances in cloud computing (Gorelick et al. 2017), the opening of image archives (Wulder et al.
74 2016), and next generation machine learning approaches (Zhu et al. 2017, Maxwell et al. 2018) are
75 placing large volumes of these moderate to near-high resolution imagery together with the
76 computational and algorithmic resources necessary to classify them at scale. These capabilities are
77 already being used to create a new generation of higher resolution (10-30 m) cropland and landcover
78 maps for Africa and other regions (Xiong et al. 2017, Lesiv et al. 2017, ESA n.d.).

79 Despite these advances, the highest resolution (<5 m) image sources are still not used to map cropland
80 over very large extents, presumably because they are commercial and relatively high cost to acquire, in
81 addition to the greater computational demands. As such, map accuracy can still be a challenge,
82 particular for User's accuracy, which ranges between 46 and 76% for the cropland class (Xiong et al.
83 2017, e.g. Lesiv et al. 2017).

84 Accuracy may also suffer due to error-inducing factors that are becoming somewhat more pronounced
85 as a consequence of these technology advances, particularly with respect to algorithms. Advances in
86 machine learning are helping to greatly improve classification skill, but these algorithms generally
87 require large, high-quality training datasets (Maxwell et al. 2018, Ma et al. 2019, Elmes et al. 2020).
88 To satisfy this need for more training (and reference) samples, map-makers increasingly rely on visual
89 interpretation of high resolution satellite or aerial imagery to collect training (or validation) samples
90 (Chen et al. 2015, e.g. Xiong et al. 2017, Stehman and Foody 2019). Several web-based platforms have
91 been developed to facilitate such efforts, which provide convenient and highly scalable tools for training
92 data collection (Fritz et al. 2012, Estes et al. 2016a, e.g. Bey et al. 2016). Visually interpreted training
93 labels present two particular problems. The first is that such labels have inevitable interpretation errors
94 that can vary substantially according to the skill of the interpreter (Estes et al. 2016a, Waldner et al.
95 2019). These errors are typically not accounted for in reported accuracy metrics, despite the fact that
96 they can introduce substantial error into the resulting maps and subsequent analyses (Estes et al. 2018,
97 Elmes et al. 2020). The second problem is that visual interpretation depends on high resolution

98 imagery (<5 m), as lower resolutions make it difficult for a human analyst to discern cropland.
99 Typically the only practical source for such imagery are “virtual globe” basemaps provided by Microsoft
100 and Google, which are composed of mosaics of various high resolution satellite and aerial images that
101 typically span 3-5 years of time within a single country (Lesiv et al. 2018). This within-mosaic
102 temporal variation can set up a temporal mismatch between the imagery being interpreted and the
103 imagery being classified, which is usually from a different source (e.g. Landsat, Sentinel; Xiong et al.
104 (2017)). If a land change occurs in the interval between the two image sets (e.g. a new field was created),
105 this can introduce error into the training data that is then passed on to the classifier. This source of
106 error may be elevated in smallholder-dominated systems in the tropics, where swidden practices are
107 common (Van Vliet et al. 2013), or in rapidly developing agricultural frontiers (Zeng et al. 2018).

108 Improving the accuracy of cropland maps over smallholder-dominated systems thus requires an
109 approach that meets three requirements. First, it should be based on high spatial and temporal
110 resolution imagery, to be able to capture the fine grain and temporal variability of smallholders’ fields.
111 Second, an algorithm with suitable skill for classifying these images must be selected, and combined
112 with the computational resources needed to process large imagery volumes. Third, a method for
113 collecting large volumes of high quality training and validation data based on image interpretation is
114 essential. This method should quantify and minimize the errors associated with image interpretation.
115 It should also ensure that labels are collected either from the same imagery that is being classified, or
116 from contemporaneous imagery, in order to reduce errors introduced by land change processes.

117 We describe here a cropland mapping system that follows these requirements, with an emphasis on
118 delineating field boundaries. The first requirement is enabled by the recent availability of CubeSat data
119 that provides 3-4 m resolution imagery over large areas at near daily intervals (McCabe et al. 2017).
120 Although these data, currently collected by Planet’s CubeSat fleet, are of lower spectral depth and
121 quality than Landsat, Sentinel, or Worldview imagery, they enable country to continental scale image
122 mosaics to be created for multiple periods during the crop growing calendars, and capturing the
123 intra-annual variability can be more important for classifying cropland than spectral depth (Debats et
124 al. 2016). This daily revisit capacity is also important for developing seasonal composites in cloudy
125 regions, where satellites with longer return intervals may fail. Lastly, although this imagery being up to
126 ~16 times coarser than much of Bing or Google imagery, it is sufficiently resolved for humans to discern

127 most fields (Fourie 2009, e.g. see Estes et al. 2018). This allows labels to be generated on the same
128 imagery processed by the classifier, thereby addressing one of the two needs related to training data
129 (requirement 3).

130 The second requirement is addressed by a computer vision/machine learning classifier that is effective
131 for classifying smallholder croplands (Debats et al. 2016), re-engineered to run on high performance,
132 cloud-based computing clusters with a simplified feature set, and following recommended practices for
133 controlling for and measuring the errors that occur when training machine learning models for remote
134 sensing applications (Elmes et al. 2020). The classifier is tightly coupled to a front-end platform for
135 collecting label data, which includes rigorous accuracy assessment protocols and a novel approach for
136 merging multiple maps into consensus labels, thereby minimizing image interpretation error (Estes et al.
137 2016a, Elmes et al. 2020). The training and machine learning components are combined within an
138 active learning framework, wherein the machine learning process assesses classification uncertainty in
139 unlabelled areas after a training step, and selects sites from areas of highest uncertainty for additional
140 labelling (Cohn et al. 1994, Tuia et al. 2011). Our framework automates this interactive approach to
141 label selection, which is effective in boosting the performance of classification models while reducing the
142 overall number of training samples required to achieve a given level of performance (Debats et al. 2017,
143 e.g. Hamrouni et al. 2021). Finally, an unsupervised segmentation step is applied to the imagery and
144 merged with the pixel-wise classifications from the machine learning process, resulting in a vectorized
145 field boundary map that provides important information on field geometry.

146 We demonstrate this approach to map cropland in Ghana, a country where smallholder farming
147 predominates, and which has a broad mix of climates and agricultural systems, including large areas
148 where shifting agriculture is practiced (Samberg et al. 2016, Kansanga et al. 2019).

149 **2 Methods**

150 **2.1 System overview**

151 The mapping system design has four primary components focused on 1) image acquisition and
152 compositing, 2) training and validation data collection, 3) image classification, and 4) segmentation.
153 The first component is applied as a once-off step to generate an image catalog covering the mapping

154 geography. The second component provides tools for labelling imagery, and controls an interactive
155 model training and classification (component 3) pipeline that produces a map of cropland probabilities
156 for each image pixel (Figure 1). The final segmentation step (component 4) is then initiated and
157 applied to both the input image catalog and the posterior probability maps, resulting in vectorized field
158 boundaries. Each system component comprises an individual software module designed to be
159 implemented on cloud computing architecture, and available on a GitHub repository (see data and
160 software availability section for details).

161 **2.2 System components**

162 **2.2.1 Image compositing**

163 The image processing components of our system were designed to work with PlanetScope Analytic
164 surface reflectance data (PlanetTeam 2018). PlanetScope provides three visual (red, green, blue) and a
165 near-infrared band at 3-4 m resolution at nominal daily frequency. Although these images are already
166 pre-processed and corrected for atmospheric effects, there are residual errors from inter-sensor
167 differences and the radiometric normalization process (Houborg and McCabe 2018), variation in the
168 orientation of scene footprints, as well as a high frequency of cloud cover over the study region (Wilson
169 and Jetz 2016). To correct for these factors, we developed a procedure for creating temporal composites
170 representing the primary growing and non-growing seasons within a single year.

171 PlanetScope imagery is accessed via the Planet API (PlanetTeam 2018), and an initial order is placed
172 for all imagery falling within the mapping geography and the date ranges for the two compositing
173 periods. The imagery is collected and transferred directly to a cloud storage platform (Amazon Web
174 Services [AWS] S3).

175 Individual images are then transformed into analysis ready data (ARD) (Dwyer et al. 2018), by
176 subsetting each downloaded image into 0.05 degree tiles, regardless of cloud cover. Tiles are organized
177 within a larger 1 degree resolution grid that covers the entire continent, which defines the minimum
178 mapping area of interest (AOI; Figure 2).

179 The temporal compositing process is applied to the tiled daily images for the time period of interest,
180 which in this case was one of two multi-month seasons, the primary growing and dry seasons for a

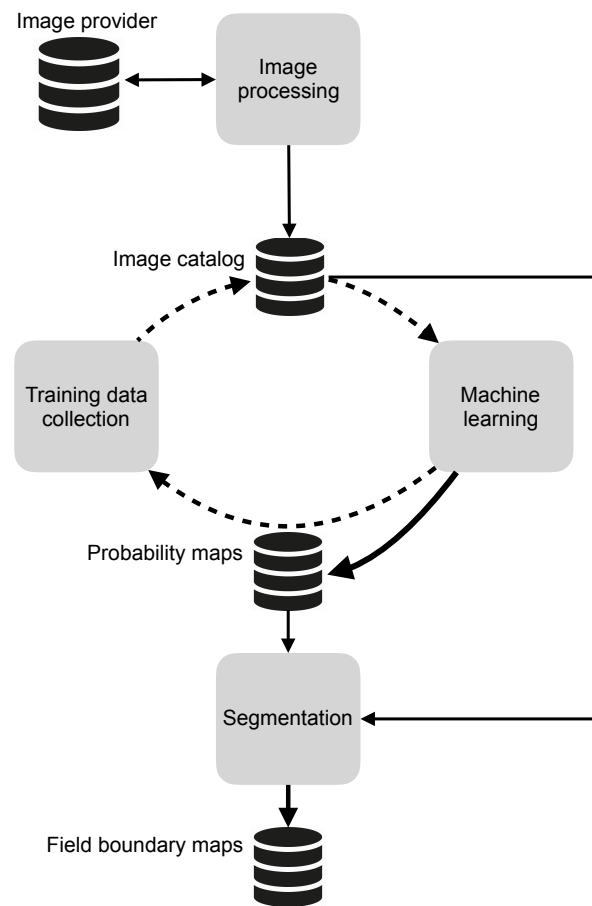


Figure 1: An overview of the primary system components, the data stores that hold the inputs and outputs from each component, and the direction of connections between them. The dashed line indicates iterative interactions, while solid lines indicate one-time or irregular connections.

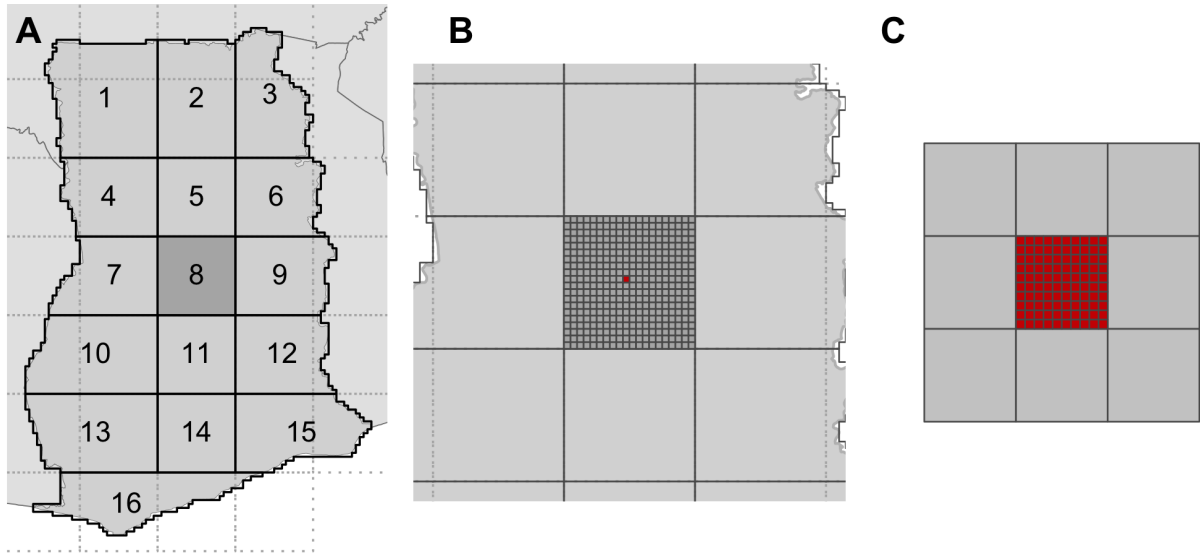


Figure 2: The reference system used in the mapping platform, including A) numbered areas of interest (AOIs) that define the minimum mapping geography (solid black lines; dotted lines indicate boundaries of 1 degree grid), B) the 0.05 degree tile used for compositing imagery, and C) the 0.005 degree resolution reference grid used for collecting training data and distributed computing.

181 single agricultural year. Imagery from two seasons helps to improve the performance of cropland
 182 classifiers (Debats et al. 2016), while having the seasons in the same year helps to minimize differences
 183 caused by land change. For each pixel in each image in each ARD temporal stack for a given season,
 184 two weights are calculated:

$$W1_t = \frac{1}{blue_t^2} \quad (1)$$

$$W2_t = \begin{cases} \frac{1}{NIR_t^4}, & \text{if } NIR_t < \text{median}\{NIR_{t1}, NIR_{t2}, \dots, NIR_{ti}\}. \\ 1, & \text{otherwise.} \end{cases} \quad (2)$$

185 Where t is a particular date in near-daily time series of PlanetScope images, which begins at date 1 for
 186 the given compositing period and ends on date i , $blue$ is the blue band, and NIR the near infrared band.
 187 Equation 1 assigns lower weights to hazy and clouded pixels as the blue band is sensitive to haze and
 188 cloud pixels (Zhang et al. 2002), while Equation 2 assigns low weights to pixels in cloud shadow
 189 considering the significant darkening effect of the cloud shadows in the Near Infrared band (Zhu and

190 Woodcock 2012, Qiu et al. 2020)

191 Once these two weights are calculated, the final composited pixel value for each of the four PlanetScope
192 bands is:

$$\bar{B} = \frac{\sum_{t=1}^T B_t * W_{1t} * W_{2t}}{\sum_{t=1}^T W_{1t} * W_{2t}} \quad (3)$$

193 Which is a weighted mean for each pixel for each band B for the particular compositing period. The
194 composited tiles were then added to the S3 store (Figure 1), where they are stored as cloud-optimized
195 geotiffs, and a “slippy map¹” rendering is created for each composite using Raster Foundry (Azavea
196 2020). The web-rendered imagery is presented within the training data platform (next section).

197 2.2.2 Labelling platform

198 Training and reference data are collected by a custom labelling platform, which was originally designed
199 for AWS’s Mechanical Turk job marketplace (Estes et al. 2016a). The basic structure of the system
200 remains the same, but we converted it into a standalone platform that allows us to enroll and pay
201 people directly for their labelling, and is designed to control and supervise the machine learning process.
202 The platform runs on a Linux virtual machine hosted on an AWS EC2 instance and is comprised of a
203 database (PostGIS/Postgres), a mapping interface (OpenLayers 3), an image server (Raster Foundry),
204 and a set of utilities for managing, assessing, and converting digitization work into rasterized labels for
205 training a machine learning algorithm. Each instance of the platform focuses on a specific AOI (Figure
206 2A)

207 The following sections provide an overview of the labelling platform’s architecture.

208 2.2.2.1 Mapping workflow

209 **2.2.2.1.1 Selecting training and reference sites** The labelling process begins with the random
210 selection of a subset (e.g. 500) of cells from a 0.005 degree grid, with the selection itself potentially split
211 into a training and validation sample, according to predetermined proportions. The grid, which is

¹https://wiki.openstreetmap.org/wiki/Slippy_Map

212 nested within the tiling and larger 1 degree grids (Figure 2C) defines the spatial unit for a labelling job.
213 The selected cells are placed into a queue within the platform’s database, and then converted into a
214 mapping *task* that has a specified number of *assignments* (boundaries drawn by an individual labeller)
215 that must be completed before the task is complete.

216 **2.2.2.1.2 Mapping assignments** Labellers registered in the system log in to the mapping
217 platform (built with Flask) and navigate to the OpenLayers-based field mapping interface (Figure 3),
218 where they are presented with a white target box representing the randomly selected grid cell, a set of
219 digitizing tools, and different image backdrops, including true and false color renderings of the growing
220 season and off-growing season PlanetScope composites, and several virtual globe basemaps. Following a
221 set of pre-defined digitizing rules (see SI), the labeller uses the polygon drawing tool to digitize the
222 boundaries of all crop fields intersecting the target grid cell that are visible within the PlanetScope
223 overlays. To aid with interpretation, the labeller can toggle between the PlanetScope renderings and
224 the basemaps to help form a judgement about what constitutes a field. The labeller assigns each
225 polygon a class category (e.g. annual cropland), and upon completing all fields submits the assignment
226 to the database. In cases where the target grid cell does not contain any fields, the labeller simply
227 submits the assignment to mark it complete. The labeller is then directed to the next available
228 assignment from a different labelling task.

229 **2.2.2.1.3 Processing completed assignments** All submitted polygons are cleaned to fix
230 topological irregularities that arose during digitization (see supporting information [SI]) and stored in a
231 PostGIS table. Each completed assignment represents one of two types of tasks: 1) accuracy
232 assessment, or 2) model training or validation. For the former type, an accuracy assessment routine is
233 invoked that executes a series of comparisons between the labeller’s results and a training reference
234 dataset, resulting in a assignment score:

$$\text{score}_i = \beta_0 I + \beta_1 O + \beta_2 F + \beta_3 E + \beta_4 C \quad (4)$$

235 Where i indicates the particular assignment, and β_{0-4} represent varying weights that sum to 1. I refers
236 to “inside the box” accuracy, O is the accuracy of those portions of the labeller’s polygons extending

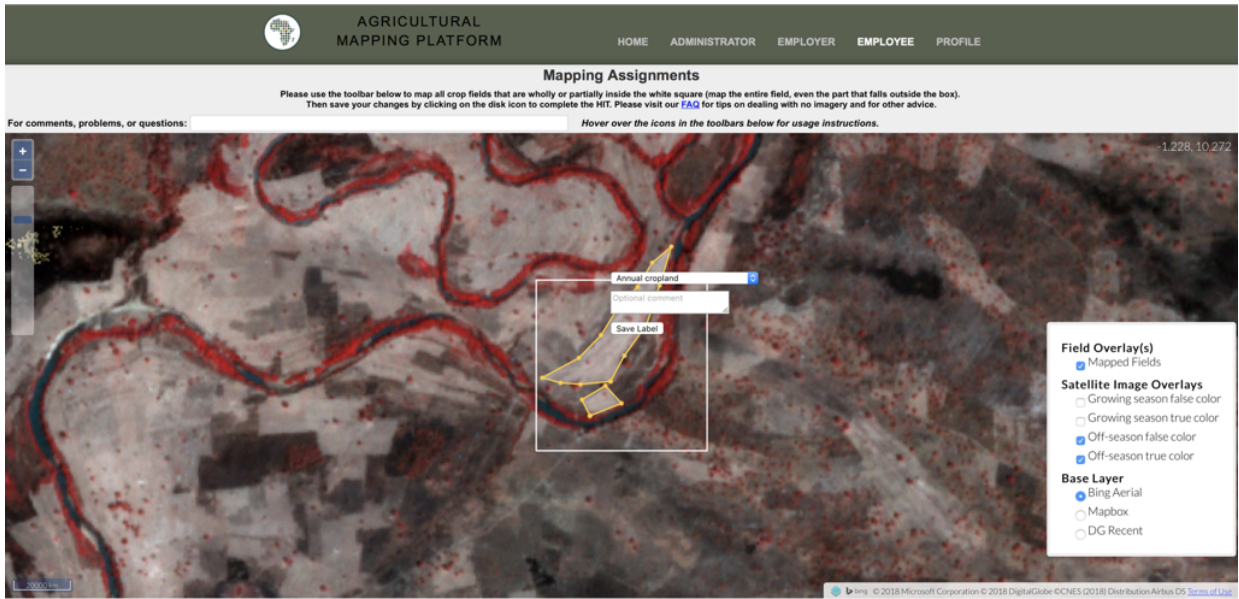


Figure 3: An overview of the labelling platform’s interface

237 beyond the target grid boundaries, F is fragmentation accuracy, a measure of how many individual
 238 polygons the labeller delineated relative to the reference, E measures how closely each polygon’s
 239 boundary matched its corresponding reference polygon boundary, and C assesses the accuracy of the
 240 labeller’s thematic labels (see SI for individual formulae). Equation 4 is an extension of the approach
 241 described by Estes et al. (2016).

242 Over time, labellers are assessed multiple times across a range of accuracy tasks, which are selected to
 243 represent the variability of the agricultural system being mapped. Each labeller’s score history is
 244 averaged to provide an overall accuracy measure, and this information is used for creating labels, the
 245 second task.

246 If the labeller’s completed assignment was a training/validation task, their maps remain stored in the
 247 database until the task’s outstanding assignments are completed by other labellers. Once complete,
 248 another routine is invoked, which combines the task’s completed assignments into a single consensus
 249 label using a Bayesian merging approach:

$$P(\theta|D) = \sum_{i=1}^n P(W_i|D)P(\theta|D, W_i) \quad (5)$$

250 Where θ represents the true cover type of a pixel (field or not field), D is the label assigned to that
 251 pixel by a labeller, and W_i is an individual labeller. $P(\theta|D)$ is therefore the probability that the actual
 252 cover type is what the labellers who mapped it says it is, while $P(W_i|D)$ is an individual labeller's
 253 average score over all the accuracy assessment assignments they have completed, and $P(W\theta|D, W_i)$ is
 254 the labeller's label for that pixel. This approach therefore uses the overall accuracy of each labeller to
 255 weight their labels when combined with those made by other labellers' for the same pixel (see SI for
 256 further details). As a further measure of confidence in the final consensus label, its average Bayesian
 257 Risk can be calculated (see SI). This measure ranges between 0 and 1, with 0 indicating full agreement
 258 between labellers for all pixels ($n = 40000$) in the label, and 1 indicating complete disagreement.

259 **2.2.3 Classification pipeline**

260 Upon completion of a batch of labels, the platform automatically launches an ephemeral Elastic Map
 261 Reduce² cluster consisting of tens of instances, depending on the size of the AOI.

262 **2.2.3.1 Feature extraction** The first step is the extraction of additional features from each
 263 seasonal image composite. Previous work showed that a large number of simple features that summarize
 264 the statistical properties of reflectance and vegetation indices in local neighborhoods are highly effective
 265 for classifying smallholder croplands (Debats et al. 2016). We followed this logic in this study, but were
 266 constrained to use a smaller feature space because the storage and memory requirements for our
 267 mapping geographies in this case were several orders of magnitude larger. For this implementation, we
 268 thus extract a set of 16 features, which are the mean and standard deviations calculated within an
 269 11X11 and 5X5 moving window, respectively (initial tests revealed these two window sizes to be most
 270 effective), resulting in 24 overall features, including the original bands (Table 1).

271 Table 1. List of image features.

Feature	Window Size	N Features
RGB-NIR	1X1	8
Mean	11X11	8
Standard deviation	5X5	8

²<https://docs.aws.amazon.com/emr/latest/APIReference/emr-api.pdf>

272 Feature extraction and the conversion of image features is handled by a combination of `GeoTrellis`³,
273 `rasterio`⁴, and `RasterFrames`⁵. These collectively extract subsets of imagery from the PlanetScope
274 temporal composites, derive the features, and convert these into Apache Spark DataFrames. Features
275 are extracted on the fly for each cell in the training and validation sets, a functionality enabled by
276 storing the image composites as Cloud-optimized Geotiffs⁶ (COGs).

277 **2.2.3.2 Classification** Once the features from the training sites are extracted into `RasterFrames`,
278 these are combined with their corresponding labels and passed to the machine learning classifier, a
279 `SparkMLlib` implementation of Random Forests (Breiman 2001). For this study, the model was trained
280 with a balanced sample and a tree depth of 15 and total tree number of 60, which initial testing showed
281 to provide a reasonable balance between computational time and performance.

282 After fitting, the model is applied to the features of the model validation set, and a set of performance
283 metrics is calculated, including binary accuracy, the F1 score (the geometric mean of precision and
284 recall), and the area under the curve of the Receiver Operating Characteristic (Pontius and Si 2014).

285 **2.2.4 The active learning loop**

286 After fitting and model evaluation a second prediction is undertaken to enable active learning. The
287 feature extraction process is repeated for the rest of the mapping geography that falls outside of the
288 training and validation sample, but applied to a subset of randomly drawn pixels from each cell in order
289 to reduce computational demand. The fitted model is applied to predict the cropland probability for
290 these selected pixels, and an uncertainty criterion (Debats et al. 2017) is calculated for each grid cell:

$$Q_I = \sum_{I(x,y) \in I} (p(x,y) - 0.5)^2 \quad (6)$$

291 Where Q is the uncertainty for grid cell I , calculated from the predicted probability p of the randomly
292 selected subset of pixels (x, y) drawn from it. Pixels with predicted probabilities closer to 0.5 are least

³<https://github.com/locationtech/geotrellis>

⁴<https://rasterio.readthedocs.io/en/latest/>

⁵<https://rasterframes.io/>

⁶<https://www.cogeo.org/>

293 certain as to their classification, thus images with the lowest values of Q represent sites posing the most
294 difficulty for the classifier.

295 After scoring with the uncertainty criterion, the top N most uncertain grid cells are selected and sent
296 back to the labelling platform, which are then digitized by the labellers. The resulting consensus labels
297 from the actively selected sample are added to the initial randomly selected sample, and a new cluster
298 is launched. The model is retrained, assesses uncertainty across the remaining unlabelled sites, and
299 selects the next most uncertain sites for labelling. This loop repeats until model performance gains
300 against the validation set show diminishing returns.

301 **2.2.5 Segmentation**

302 After the final iteration, the segmentation algorithm is invoked, which entails several steps. In the first
303 step, the meanshift algorithm (Yizong Cheng 1995) is applied to the original bands of the dry season
304 composite. A Sobel filter is then applied to the green, red, and near-infrared mean-shifted bands and
305 the probability map, and a combined edge image is computed using the sum of these four edge images
306 for the dry season only. A compact watershed algorithm (Neubert and Protzel 2014) is then run on the
307 weighted edge image, with a high level of segmentation specified. In this case, we specified **6400**
308 segments per tile.

309 Third, a region adjacency graph is constructed for each image tile, in which each node represents all
310 pixels within each polygon created in the previous step. The edge between two adjacent regions
311 (polygons) is calculated as the norm of the difference between the means of normalized colors of all
312 bands. Hierarchical merging is then applied, in which the most similar pairs of adjacent nodes are
313 merged until there are no edges remaining below a predetermined threshold of 0.05.

314 In the fourth step, the merged polygons are overlaid with the posterior probabilities resulting from the
315 final active learning loop, and polygons in which the average posterior probability is greater than a
316 predetermined threshold (here 0.5, but could vary locally) are retained as field polygons.

317 In the final step, the retained polygons are refined by removing holes and smoothing their boundaries
318 using the Visvalingam algorithm (Visvalingam and Whyatt 1993). Neighboring polygons that overlap
319 along tile boundaries are then merged.

320 To assess the accuracy of the final segmented boundaries, we used a two-step approach. First, we
321 assessed the overall thematic accuracy of the resulting classification against our map reference data.
322 Second, to assess the quality of the segmentation, we compared the mean area and relative frequencies
323 of the segmented polygons within different size classes against the same metrics derived from the
324 digitized fields of the most accurate worker to create the given map. We selected this relatively simple
325 procedure, as opposed to more complex measures of object accuracy (Ye et al. 2018), because, on the
326 one hand, both the automated segmentation algorithm and labellers are cueing in on the same
327 features—abrupt, physically detectable breaks within the imagery. On the other hand, no matter how
328 well the interpreted/segmented boundaries align with the boundaries of fields in the imagery, it is
329 logistically difficult to evaluate performance against real-world boundaries as the spectral distinction of
330 field boundaries will vary across different crop types and land use arrangements.

331 **2.3 Applying the system to map Ghana**

332 We applied the system to map Ghana’s croplands, excluding areas primarily cultivated with tree crops.
333 Ghana has several distinct agricultural regions, ranging from the primarily grain and vegetable crop
334 producing regions in the northern savannas to tree crop-dominated system in the forested southwest,
335 where cocoa and oil palm are among the dominant crops. For these latter regions, we did not attempt
336 to classify tree crops, and instead mapped clearings that potentially contain field crops or newly felled
337 or recently replanted tree crops. We made this decision because PlanetScope’s resolution is not high
338 enough for labellers to distinguish many tree crops from surrounding forest, and the boundaries of
339 many tree crops (e.g. cocoa) are often not visible.

340 To create the cropland maps, we divided the country into 16 distinct AOIs, which were developed by
341 grouping together each one degree cell fully contained within Ghana with the tiles belonging to any
342 adjacent degree cell that overlapped neighboring countries (Figure 2A). The exception was AOI 16,
343 which consisted of the four degree cells intersecting Ghana’s southern coast. The resulting AOIs ranged
344 from 12,160 to 23,535 km² in extent (average = 15,457 km²), A separate active learning and
345 segmentation process was run for each of these AOIs.

346 To collect the initial randomized samples for model training, we grouped the AOIs into three clusters: a
347 northern cluster comprising the 6 northernmost AOIs (Cluster 1), a central to southeastern cluster

348 (Cluster 2) consisting of the 3 middle (AOIs 7-9) and 2 southeastern AOIs (12 and 15), and a
349 southwestern cluster (Cluster 3) made up of the forest zone AOIs (10, 11, 13, 14, 16). Within each
350 cluster, we randomly selected and labelled 500 grid cells, which provided relatively large initial training
351 samples for these agro-ecologically similar regions, while helping to minimize the overall amount of
352 labelling effort. In addition to these samples, we randomly selected and labelled 100 grid cells within
353 each AOI to provide a validation sample.

354 After collecting the initial training and validation samples, we trained a starter model for each cluster
355 and applied it to each of the block's AOI. For each iteration, 100 samples were actively selected within
356 each AOI, and added to the training pool.

357 During the collection of training and validation samples, labellers were tasked to only digitize active or
358 recently active crop fields, avoiding tree crops, and fallow or potentially abandoned fields (see SI for the
359 digitizing rules).

360 To evaluate the performance of the system, we performed several analyses described in sections 2.3.1-4.

361 **2.3.1 Image quality**

362 We evaluated the overall quality of the resulting seasonal image composites by assessing a random
363 selection of 50 tiles. We graded both seasonal composites for each tile using a four category quality
364 score, which evaluated the degree of 1) residual cloud and 2) cloud shadow, 3) the number of visible
365 scene boundary artifacts, and 4) the proportion of the image that had its resolution degraded below the
366 typical 3-4 m PlanetScope resolution (e.g. because of between-date image mis-registrations). Each
367 category was qualitatively ranked from 0-3, with 0 being the lowest quality, and 3 the highest (see SI for
368 complete protocol), making the highest possible score 12. We rescaled scores to fall between 0 and 1.

369 **2.3.2 Model gains per iteration**

370 To assess the gain in model performance due to active learning, we measured the change in accuracy,
371 F1, and AUC (see 2.2.3.2) between each iteration and between the first and last iterations for each AOI.

372 To evaluate whether active learning improved model performance relative to a purely random approach
373 to selecting new training sites, we ran additional tests within a subset of AOIs (1, 8, and 15). We first

374 randomly selected and labelled 300-400 sites in each AOI. We then progressively added 100 of the
375 randomly selected samples to the relevant training pool and retrained the model, repeating the process
376 so that the number of iterations and samples matched those from the active learning process. We then
377 compared the difference in accuracy, AUC, and F1 between the randomly trained models and those
378 trained with active learning (Debats et al. 2017).

379 **2.3.3 Accounting for label error**

380 To quantify the potential impact of label error on classification results, we evaluated the performance
381 differences between models trained with three different sets of labels: 1) those from the lowest scoring
382 labeller to map each training site, 2) those from the highest scoring labeller, and 3) the consensus
383 labels. This assessment follows recommended Tier 1 (i.e. best practice) standards to account for
384 training data errors (Elmes et al. 2020).

385 **2.3.4 Accuracy assessment**

386 The model performance assessments described above (2.3.2-3) were not fully independent because they
387 used the same validation sites over multiple iterations (Elmes et al. 2020). To independently assess the
388 accuracy of our final map products, we followed recommended guidelines (Stehman and Foody 2019) to
389 create a separate map reference sample. We used a stratified design, randomly assigning square
390 polygons of ~0.1 ha extent into cropland and non-cropland strata, developed the map of segmented
391 field boundaries. Four classes were used for the map reference sample: cropland; non-cropland; unsure
392 but likely cropland; unsure but likely non-cropland. The latter two classes were used to provide insight
393 into the degree of uncertainty in the map reference sample. For efficiency, two separate supervisors
394 evaluated separate portions of the reference sample, but both jointly assessed a small subset of the
395 sample. We calculated their level of agreement on this subset to provide an additional assessment of
396 uncertainty in the map reference sample (Stehman and Foody 2019). The SI contains further details on
397 the design and collection of the map reference sample.

398 The map reference polygons were then intersected with both the probability images and the segmented
399 field boundaries, and confusion matrixes between the map reference labels and the extracted map
400 classes were constructed to assess the categorical accuracy of each map product. We calculated the

401 overall accuracy for each map, as well as the class-wise User’s and Producer’s accuracy, as well as the
402 95% confidence intervals for each accuracy measure (Olofsson et al. 2013, 2014, Stehman and Foody
403 2019).

404 To assess the accuracy of the segmented field boundaries, we compared the size class distributions of
405 the segmented field boundaries against those of the workers’ digitized polygons at map validation sites.
406 We chose this approach because of existing uncertainties in polygon-based accuracy assessment
407 methods (Ye et al. 2018), and because the map’s ability to represent field sizes was of greatest interest.
408 To undertake this comparison, we selected the field polygons from the most accurate labeller to digitize
409 each of the 100 validation sites in each AOI, and calculated the site-wise average area and number of
410 polygons. We then calculated the same statistics from the segmented boundaries that intersected each
411 validation grid. We compared the distributions and proximity of two measures of central tendency
412 (mean and median) calculated from the two datasets for each AOI, and across all AOIs.

413 **2.4 Assessing the characteristics of Ghanaian cropland**

414 Using the final mapped results, we calculated the estimated area of cropland in Ghana, as well as the
415 average size and total number of fields in the different AOIs. We used the map reference sample to
416 calculate adjusted area estimates and confidence intervals for each map class, and used the differences
417 between labellers’ polygons and segmented boundaries at validation sites to calculate bias-adjusted
418 estimates of mean field sizes and the total number of fields.

419 **3 Results**

420 We developed maps of Ghana’s cultivated croplands within an area of 247,299 km², which included
421 portions of neighboring countries overlapped by images tiles.

422 **3.1 Image catalog and quality**

423 To develop the maps, we first generated an image catalog for the 8,116 tiles covering Ghana. This
424 entailed processing all PlanetScope imagery intersecting these tiles between May-September, 2018 (the
425 growing season) and December, 2018 to February, 2019 (the subsequent dry season). The longer period
426 was necessary for the growing season because of the frequent cloud cover, which substantially limits the

427 number of clear scenes for any tile (Figure S3). For the cloudiest regions (AOIs 10, 11, 13, 14, 16) we
428 started the dry season window in November.

429 An assessment with two observers (see SI for observer details) found that average quality per growing
430 season composite tile was 0.88, with 70 percent having scores ≥ 0.85 , while the average quality of dry
431 season composites was 0.92 (74 percent ≥ 0.85). Composite quality in both seasons was highest in the
432 northern half of the country and lowest in the southwest (Figure 4A), where the substantially greater
433 cloud cover resulted in a much lower density of available PlanetScope imagery for each time period
434 (Figure S3).

435 **3.2 Active learning**

436 **3.2.1 Training data collection**

437 After training of models with the initial randomly selected label sets, the active learning process was
438 run for 3 iterations for 12 of the 16 AOIs, resulting in 800 labels per AOI. AOIs 10 and 14 stopped after
439 one and two iterations, respectively, as they started with high initial validation accuracies ($>83\%$) and
440 showed little subsequent improvement. The models for these two AOIs were thus trained with 600 - 700
441 samples. AOI 15 was run for 4 iterations (900 samples), while AOI 3 underwent a second active
442 learning cycle because the model produced during the first cycle was inaccurate (see SI). In this second
443 run, 300 initial training sites randomly selected within the AOI were used (Figure S4A), followed by 2
444 subsequent active learning iterations, resulting in a training sample of 500. Labels collected during the
445 active learning iterations showed distinct patterns in several AOIs, which often fell along ecotones, such
446 as the boundaries between agroecozones (see Figure S4A). The total number of unique training and
447 validation sites across the country were 6,299 and 1,600, respectively.

448 The distribution of training and validation sample collection effort was divided across 20 labellers, with
449 a core group of 13 who mapped more than 1,000 sites each. As each training site was mapped by 4
450 separate labellers, 34,014 sets of vector labels were made. Each labeller digitized an average of 2,001
451 (see Figure S5A for more details on labelling effort). Labeller accuracy was scored 9,389 times against
452 98 unique training reference sites (Figure S4A), with each labeller assessed an average of 552 times at a
453 rate of 1 training reference site for every 3.62 training site mapped. The mean of each labeller's average

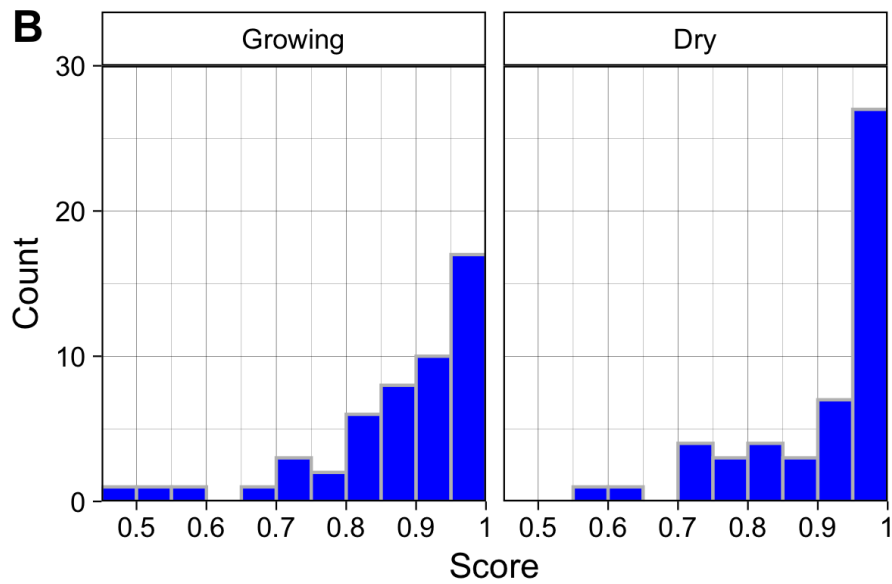
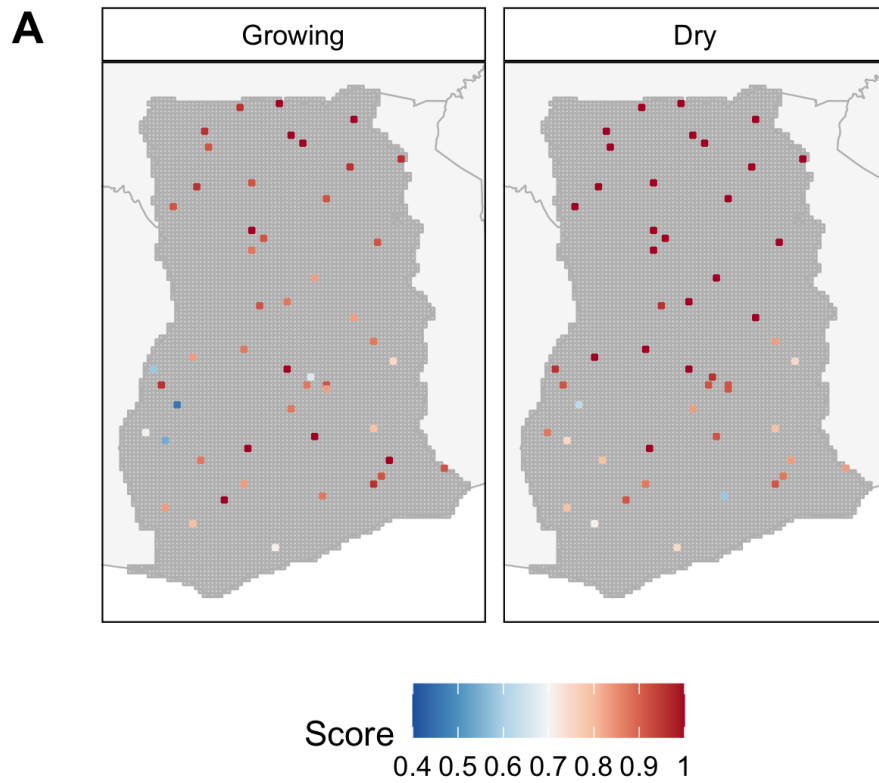


Figure 4: The location and quality scores of 100 randomly selected tiles for the growing (A) and off-growing season (B), and the corresponding distributions of the quality scores for each season, respectively (C and D).

454 accuracy score was 0.71 (range 0.6 to 0.85; see Figure S5B for detailed score distributions).

455 After each site was mapped by four labellers, consensus labels were generated. The Bayesian Risk (see
456 SI) of each training and validation label was calculated as an additional measure of label quality. The
457 average risk was 0.122 for training labels and 0.127 for validation labels. Risk was highest in the
458 northern AOIs (AOIs 1-6; Figures S6-7), falling between 0.157 for training and 0.173 for validation
459 labels (Figures S6-7), and lowest in the southwestern AOIs (AOIs 10, 11, 13, 14, 16; training risk =
460 0.079; validation risk = 0.065). Label risk in the central-southeastern AOIs (AOIs 7-9, 12, 15) was
461 slightly lower (training = 0.127; validation = 0.136) than in the north. Labeller experience also
462 appeared to reduce risk, which we observed during a relabelling of the 500 initial random site in this
463 cluster (see SI); the mean risk of the updated labels was 0.055, compared to 0.172 for original labels.

464 **3.2.2 Performance gains during active learning**

465 Model performance was calculated for each iteration within each AOI. The average accuracy, AUC, and
466 F1 at iteration 0 were 0.786, 0.809, and 0.464, respectively, increasing to 0.825, 0.818, and 0.507 by
467 iteration 3 (Figure 5). These differences represent respective gains of 4.9, 1.1, and 9.1 percent for the
468 three metrics. The largest gains for each metric occurred on iteration 1, averaging 2.9, 1, and 3.8
469 percent for accuracy, AUC, and F1, while the lowest gains were realized on iteration 3, with accuracy,
470 F1, and AUC respectively increasing by just 1.2%, 0.9%, and 0.3%. The scores achieved on the final
471 iteration varied substantially across AOIs and metrics. Accuracy ranged between 0.725 (AOI 15) and
472 0.948 (AOI 16), while AUC varied from 0.725 (AOI 4) and 0.93 (AOI 11), and F1 from 0.252 (AOI 13)
473 and 0.636 (AOI 8).

474 The comparison of active versus randomized training sample collection (in AOIs 1, 8, and 15) showed
475 that the former approach outperformed the latter. After three iterations, the accuracy, AUC, and F1
476 scores resulting from active learning were respectively 0.8, 0.6, and 2.3 percent higher than the scores
477 from a randomly trained model (Figure S8). However, there was more variability in earlier iterations,
478 with average score differences of -1.7 (accuracy), 0.6 (AUC), and 0.8 percent (F1) after iteration 1, and
479 -0.3 (accuracy), 0.4 (AUC), and 1.8 (F1) percent after iteration 2. The negative results for accuracy
480 was caused by results at AOI 15, where active learning accuracy was 8.37 percent lower than random
481 training after iteration 1 (see Figure 5). In comparison, iteration 1 active learning accuracies were 2.88

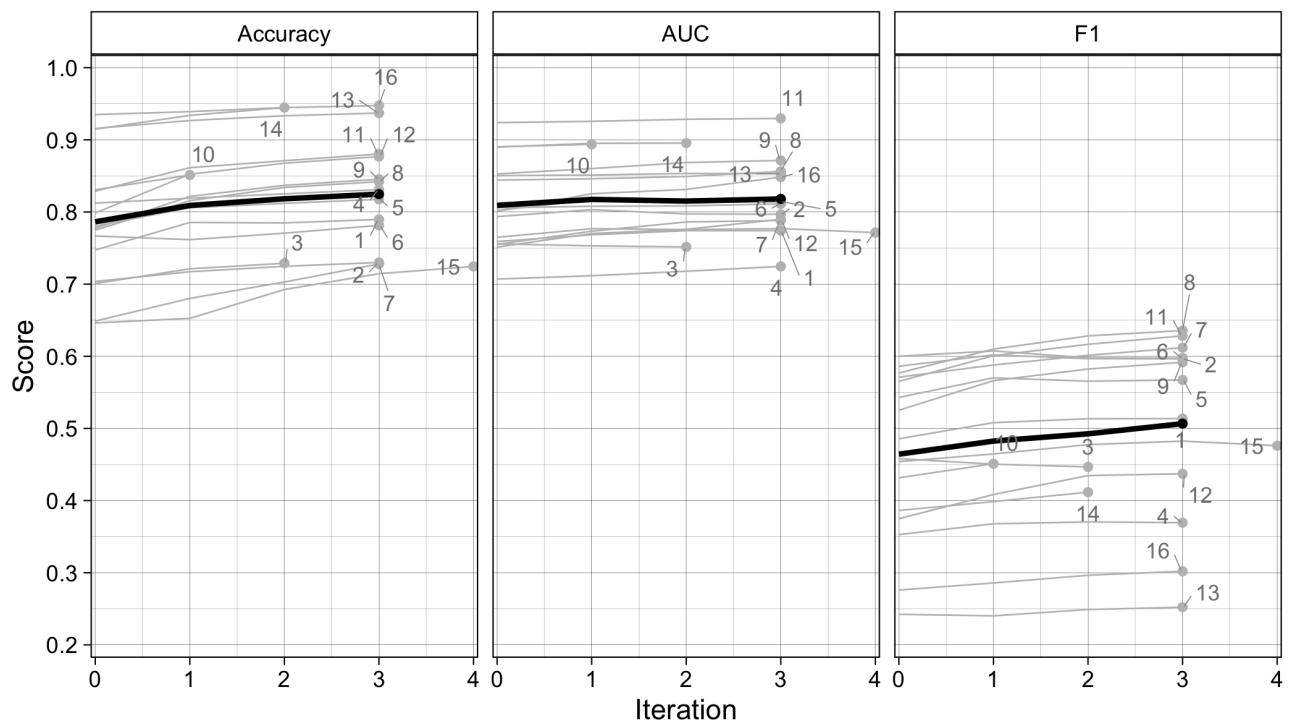


Figure 5: Scores for overall accuracy, area under the curve of the Receiver Operating Characteristic, and the F1 scores for the Random Forests model results after each iteration of the active learning loop for each AOI (gray lines), as well as the mean score per iteration across all AOIs (black lines).

482 and 0.45 percent higher than random training for AOIs 1 and 8, respectively. Accuracy under active
483 learning for AOI 15 exceeded randomized training after 4 iterations.

484 **3.2.3 The impact of training data error**

485 The potential impact of label errors on map quality was assessed in four AOIs (1, 2, 8, and 15). The
486 results of these tests showed that the average accuracy, AUC, and F1 scores for models trained with the
487 consensus labels were respectively 0.772, 0.8, and 0.555 (Figure 6). Performance metrics from
488 consensus-trained models were just 0.5 - 1.2 percent higher than those models trained with the most
489 accurate individuals' labels (accuracy = 0.762; AUC = 0.796; F1 = 0.55), but were 11.6 - 27.4 higher
490 than models trained with the least accurate individual labels (accuracy = 0.606; AUC = 0.716; F1 =
491 0.44).

492 A second measure of the impact of label error is found within the correlations between the mean label
493 risk per AOI and the model performance metrics (Table S3). Accuracy and AUC had strong
494 (Spearman's Rank Correlation = -0.824 to moderate ($r = -0.568$) negative correlations with label risk,
495 while F1 had a weaker but moderate positive association ($r = 0.456$). The positive sign of the latter
496 relationship is counter-intuitive, but is explained by risk's association with precision, one of two inputs
497 to F1, which was moderately positive (0.629), whereas risk had a negligible correlation with recall
498 (0.206), F1's other component. The correlation between risk and the false positive rate (0.688), another
499 important performance metric, shows that labelling uncertainty may increase model commission error.

500 **3.3 Map accuracy**

501 **3.3.1 Categorical accuracy**

502 We used a map reference sample of 1207 sites (487 cropland; 720 non-cropland) to evaluate the accuracy
503 of the per-pixel classifications (resulting from thresholding the Random Forests probability), as well as
504 the segmented field boundary maps. We first evaluated the uncertainty in the map reference classes by
505 assessing 1) the overall agreement between map reference labels collected by two separate supervisors at
506 23 sites, and 2) the confidence of the labels assigned by the supervisors (see SI for details). The first
507 measure showed that the two individual supervisors' labels agreed at 87% of common sites, while the
508 second showed that 15.7 of sites were labelled with the two classes that indicated a level of uncertainty.

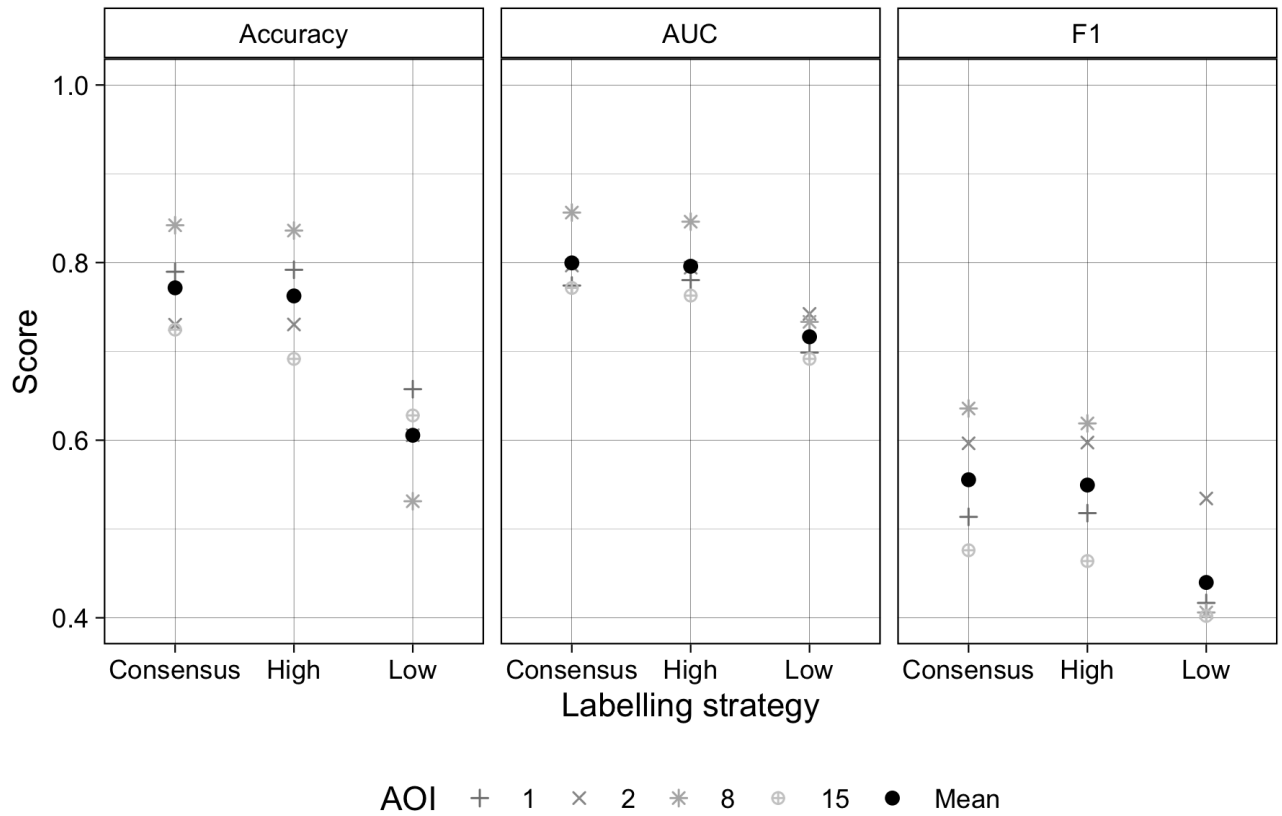


Figure 6: Scores for overall accuracy, area under the curve of the Receiver Operating Characteristic, and the F1 score resulting from models trained with consensus labels, and labels made by the most and least accurate labellers to map each site. Comparisons were made for AOIs 1, 2, 8, and 15, denoted by grey symbols, while the mean scores across these AOIs are shown for each metric.

509 We found that the overall accuracy of the pixel-wise classifications was 88% against this map reference
510 sample (Table 2). Confining the map reference sample to four distinct zones (Figure S10A) shows that
511 overall accuracy ranged from 83.3% in Zone 1 (AOIs 1-3) to 93.6% in Zone 3 (AOIs 10, 11, 13, 15, and
512 16). The Producer's accuracy of the cropland class was 61.7% across Ghana, ranging from 45.6% in
513 Zone 3 to 67.9% in Zone 1, while the User's accuracy for was 67.3% overall, ranging from 59.8% in
514 Zone 4 to 71.2% in Zone 1. Both measures of accuracy were substantially higher for the non-cropland
515 class across all zones, typically exceeding 90%. The lowest accuracies for the non-cropland class was in
516 Zone 1 (Producer's = 89.3%; User's = 87.7%).

517 The overall accuracies obtained from the segmented maps were generally 1-2 percentage points lower
518 than those of the per-pixel maps, while User's accuracies tended to be 8-10 percentage points less
519 (Table 2). In contrast, Producer's accuracies were 15-20 points higher than in the per-pixel map. The
520 segmentation step therefore helped to reduce omission error while substantially increasing commission
521 error.

522 3.3.2 Segmentation accuracy

523 The comparisons of digitized versus segmented field boundaries showed that the mean field size across
524 all validation sites averaged 4.97 ha (Median = 3.75; StDev = 6.04), which was 1.41 times larger than
525 the 2.06 ha (Median = 1.35; StDev = 3.26) mean area of labeller-digitized polygons. This discrepancy
526 was primarily caused by results in four AOIs (2, 3, 7, and 15; Figure S11), where segments averaged
527 between 7.76 and 10.76 ha, compared to 2.18 - 2.77 ha for the corresponding hand-digitized polygons.
528 The number of segmented fields per validation site averaged 3.08 (median = 2.66; StDev = 2.9)
529 compared to 4.4 (median = 3.38; StDev = 4.52) for digitized polygons (Figure S12).

530 3.4 Ghana's croplands

531 Two separate maps of cropland were produced for each AOI, a per-pixel map derived from the cropland
532 probabilities, and the vectorized map of field boundaries (Figure 7). The former provides the more
533 accurate picture of cropland distributions in Ghana, which are most concentrated in the Southeastern
534 corner (AOI 15), the central-western region (AOI 7, the northeastern and northwestern corners of AOIs
535 10 and 11, and the south of AOI 8), and the northeastern quadrant stretching from AOI 9 through

Table 2: Map accuracies and adjusted area estimates for the 3 m pixel-wise classifications (based on Random Forests predictions; top 5 rows) and the segmented map (bottom 5 rows). Results are provided for 4 zones (Zone 1 = AOIs 1-3; Zone 2 = AOIs 4-9; Zone 3 = AOIs 10, 11, 13, 14, 16; Zone 4 = AOIs 12, 15) plus the entire country. The error matrix (with reference values in columns) provides the areal percentage for each cell, and the Producer's (P), User's (U), and overall (O) map accuracies and their margins of error (in parenthesis) are provided, as well as the sample-adjusted area estimates (in km²) and margins of error.

		Non-crop	Crop	Total	U	O	n	Area	
Per-pixel classification	Zone 1	Non-crop	64.2	9	73.2	87.7 (5.5)	83.3 (4.3)	138	40992 (2468)
		Crop	7.7	19.1	26.8	71.2 (5.9)		226	16025 (2468)
		P	89.3 (5.5)	67.9 (5.9)					
		n	186	178					
	Zone 2	Non-crop	73.9	6.7	80.6	91.7 (4.2)	86.5 (3.6)	169	65123 (2866)
		Crop	6.8	12.6	19.4	64.8 (6.0)		247	15533 (2866)
		P	91.5 (4.2)	65.3 (6.0)					
		n	242	174					
	Zone 3	Non-crop	89.6	4.8	94.4	94.9 (3.2)	93.6 (3.1)	177	70885 (2413)
		Crop	1.6	4	5.6	71.4 (9.0)		98	6860 (2413)
		P	98.2 (3.2)	45.6 (9.0)					
		n	196	79					
	Zone 4	Non-crop	80.7	5.3	85.9	93.8 (5.9)	89.1 (5.3)	65	26473 (1615)
		Crop	5.7	8.4	14.1	59.8 (10.4)		87	4199 (1615)
		P	93.4 (5.9)	61.4 (10.4)					
		n	96	56					
	Ghana	Non-crop	77.2	6.7	83.9	92.0 (2.3)	88.0 (2.0)	549	202856 (4904)
		Crop	5.3	10.8	16.1	67.3 (3.6)		658	43233 (4904)
		P	93.6 (2.3)	61.7 (3.6)					
		n	720	487					
Segmentation	Zone 1	Non-crop	57.6	4.2	61.8	93.2 (5.3)	81.4 (3.9)	88	40890 (2236)
		Crop	14.4	23.8	38.2	62.3 (5.7)		276	15905 (2236)
		P	80.0 (5.3)	84.9 (5.7)					
		n	186	178					
	Zone 2	Non-crop	70.4	3.7	74.1	95.0 (3.9)	85.2 (3.2)	121	65642 (2599)
		Crop	11.2	14.8	25.9	56.9 (5.7)		295	14841 (2599)
		P	86.3 (3.9)	80.1 (5.7)					
		n	242	174					
	Zone 3	Non-crop	86.6	3	89.6	96.6 (2.9)	92.6 (2.8)	148	71695 (2181)
		Crop	4.3	6.1	10.4	58.3 (8.6)		127	7167 (2181)
		P	95.2 (2.9)	66.7 (8.6)					
		n	196	79					
	Zone 4	Non-crop	75.3	3.4	78.7	95.7 (6.0)	86.1 (5.1)	46	26712 (1593)
		Crop	10.4	10.8	21.3	50.9 (9.6)		106	4446 (1593)
		P	87.8 (6.0)	76.0 (9.6)					
		n	96	56					
	Ghana	Non-crop	73.2	3.6	76.8	95.3 (2.1)	86.7 (1.8)	403	204940 (4395)
		Crop	9.7	13.5	23.2	58.2 (3.4)		804	42359 (4395)
		P	88.3 (2.1)	78.9 (3.4)					
		n	720	487					

536 AOIs 5 and 6 and up to AOIs 2 and 3. The northern third of AOI 1 also has noticeable densities of
537 cropland. Several notable areas of low cropland density are also apparent, indicating the presence of
538 large protected areas, such as Mole National Park in the southeastern corner of AOI 1 and Digya
539 National Park in the northwestern corner of AOI 12. In contrast, the relative absence of cropland in
540 AOIs 13, 14, and 16 does not reflect the scarcity of agriculture in these areas, but rather the
541 predominance of tree crops, which we did not map.

542 Both the per-pixel and vectorized maps, when combined with the map reference sample, enable
543 separate estimates of the total extent of croplands in Ghana. The cropland extent estimated from the
544 vectorized map is 42,359 km² (with a margin of error of 4,395 km²), or 17.1 (15.4-18.9%) of the
545 mapped area. The estimate based on the per pixel map is 43,233 km² (margin of error = 4,904 km²),
546 or 17.6 (15.6-19.6%) of area.

547 The vectorized map provides additional information on how the characteristics of croplands can vary
548 geographically, ranging from narrow, strip-like fields in parts of AOI 15 (Figure 7's lower right inset) to
549 more densely packed, less distinctly shaped fields in AOI 5 (upper right inset in Figure 7). To explore
550 how field characteristics varied geographically, we mapped the average field size and total number of
551 fields within each 0.05 degree tile (Figure S13). These patterns generally correspond to those seen in
552 the cropland density map (Figure 7), with larger sizes and field counts generally occurring in areas of
553 higher field density, although the biases inherent in both measures (Figures S11-12) complicate the
554 interpretation of those variations. However, we can use the estimated biases to develop adjusted
555 estimates of field sizes and counts for each AOI, and for Ghana overall (Table 3). These adjusted
556 estimates show that the typical field size in Ghana is 1.73 ha, ranging from 0.96 in AOI 4 to 2.82 ha in
557 AOI 4, with fields in the forest zone AOIs (10, 11, 13, 14, 16) generally smaller than those in the
558 northern half of the country (Table 3). The total number of fields was estimated to be 1,662,281 overall,
559 or 205 fields per tile on average, ranging from 108/tile in AOI 4 to 399/tile in AOI 6.

560 4 Discussion

561 These results demonstrate a capability for mapping the characteristics of smallholder-dominated
562 cropping systems at high spatial resolution, annual time steps, and national scales. The resulting maps

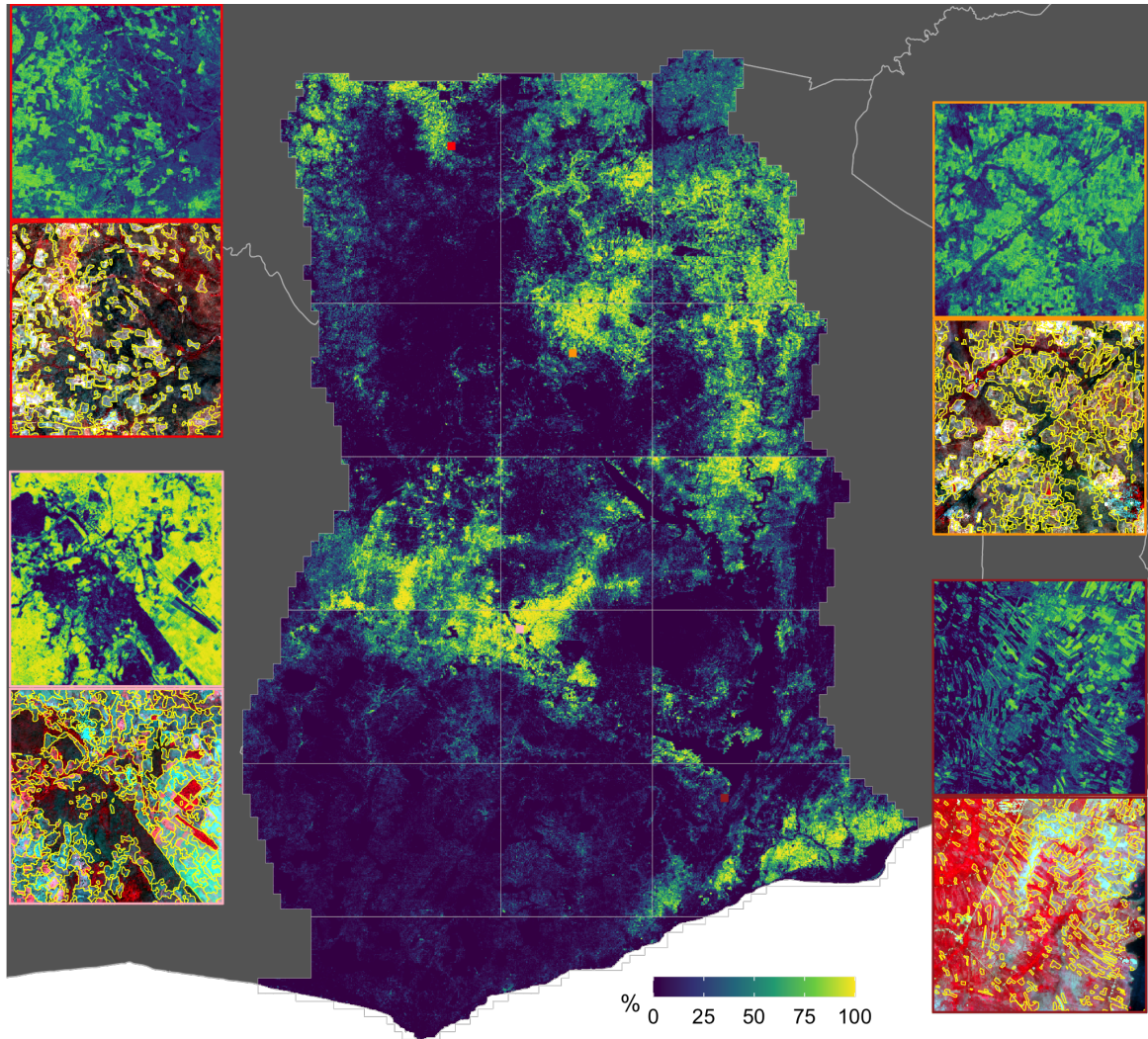


Figure 7: The distribution of croplands in Ghana. The main map shows the percentage of croplands in each 0.005 degree grid cell, derived from the predicted cropland probabilities. The insets on the margins illustrate predicted probabilities (top map in each couplet) at original image resolution (0.000025 degrees) and segmented field boundaries overlaid on the dry season PlanetScope composite, for four separate tiles. Each tile's position is shown on the main map, and is color-coded to the boundary lines around its corresponding inset.

Table 3: The average size and total number of crop fields for each AOI and for Ghana overall. The original and bias-adjusted values for each measure are provided, as well as the total number of 0.05° degree tiles in each AOI.

AOI	N tiles	Size	Size (adj)	N	N / tile	N (adj)	N (adj) / tile
1	777	3.71	1.26	97,822	126	127,580	164
2	597	7.66	1.96	87,666	147	120,651	202
3	501	8.24	2.18	108,819	217	104,422	208
4	465	2.44	2.82	26,276	57	50,163	108
5	400	4.24	2.09	43,290	108	53,756	134
6	429	5.10	2.15	81,363	190	145,347	339
7	471	5.64	1.49	93,282	198	123,005	261
8	400	4.89	1.98	55,500	139	78,868	197
9	479	4.10	1.82	72,081	150	89,840	188
10	630	2.24	1.04	119,019	189	170,907	271
11	400	3.65	1.52	52,510	131	94,709	237
12	471	3.44	1.77	44,667	95	52,947	112
13	627	0.84	0.96	67,996	108	125,368	200
14	400	1.09	2.72	56,006	140	101,767	254
15	548	4.95	1.54	75,752	138	105,681	193
16	521	0.95	1.41	49,097	94	117,268	225
Ghana	8,116	3.92	1.73	1,131,146	139	1,662,281	205

563 provide an updated and more granular view of the distribution and extent of croplands in Ghana,
564 complementing existing national to regional land cover maps derived from moderate resolution sensors
565 (Hackman et al. 2017, Xiong et al. 2017, ESA n.d.). This prior work found that cropland covered 19.4
566 (Xiong et al. 2017) to 32% (Hackman et al. 2017) of Ghana in 2015, whereas our 2018 maps have
567 cropland cover of 16.1-23.2% (Table 2), and our map reference sample-based estimates finds 17.1-17.6%
568 cover. Our results thus suggest that Ghana’s cropland is less than previously estimated, but the
569 difference is perhaps attributable to our use of a cropland definition that excluded longer fallows and
570 abandoned fields, which in some regions can account for over half of the total area that could be
571 counted as cropland (Tong et al. 2020).

572 In addition to the more detailed update of cropland extent, our maps also provide new information on
573 the size and number of fields in Ghana (Figures 7, S11-12). Previous work to estimate such agricultural
574 characteristics have often focused on farm, rather than field, size using census data (Von Braun 2004,
575 Samberg et al. 2016, Jayne et al. 2016, Lowder et al. 2016). Efforts to map field boundaries in

576 smallholder-dominated agricultural systems have either used *in situ* data collection (Carletto et al.
577 2013, 2015) or remote sensing studies over relatively small (e.g. Forkuor et al. 2014, Persello et al.
578 2019) or discontinuous (Estes et al. 2016a) areas. The most extensive studies to date used
579 crowdsourced volunteers to classify fields into broad size classes, based on their interpretations of
580 imagery sampled from high resolution virtual globes (Fritz et al. 2015, Lesiv et al. 2019). Those efforts
581 included country-specific results for Ghana (n = 263), which can be converted into an average field size
582 estimate of 5.33 ha⁷. This estimate exceeds our Ghana-wide average segment size (3.92 ha; Table 3),
583 but is closer to the mean (4.97 ha) within AOIs 1-9, 12, and 15, which is where most of the
584 crowdsourced sample appears to have been collected. However, our bias-corrected estimates of 1.73
585 (Ghana-wide) and 1.87 (AOIs 1-9, 12, and 15) ha were much smaller.

586 4.1 Map accuracy and key sources of error

587 Although the maps generated by our system provide valuable new information, they nevertheless
588 contain substantial errors. The overall map accuracies (86.7-88%, Table 2) are near the boundary of
589 what might be considered *achievable* map accuracy (Elmes et al. 2020), given the inherent uncertainty
590 in the map reference sample, our best estimate of the “truth,” in which we have roughly 85%
591 confidence. However, accuracies for the cropland class were much lower, falling between 62 (Producer’s)
592 to 67 (User’s) percent country-wide for the per-pixel map (Table 2), meaning the model produced
593 substantial commission and omission errors for this class. The segmented boundary maps had fewer
594 omission errors (Producer’s accuracy = 79%), but higher false positive errors (User’s accuracy =
595 58.2%). These accuracies are near the middle to upper ranges of those reported for the cropland class
596 in other large-area mapping studies (Hackman et al. 2017, Xiong et al. 2017, Lesiv et al. 2017).

597 The patterns of cropland-class accuracies varied by zone. These zones largely align, albeit with some
598 discrepancies, with the country’s agroecozones, thus the accuracy patterns may be partially attributed
599 to some regions being harder to map than others. Producer’s accuracies for both maps were highest in
600 the two northern zones (1 and 2), which are primarily savannas (Figure S10), and lowest in zones 3 and
601 4, which are comprised of forest or coastal savannas. User’s accuracies followed a similar pattern, with

⁷Obtained by calculating the weighted mean from the count of the five size classes and the mean of the hectare range provided for the four smallest size classes, and the lower bound of the size range provided the largest size class. Data sourced from Table S3 in Lesiv et al. 2019.

602 the exception of Zone 3, which had the highest User’s accuracy, albeit from a very small sample.
603 Aligning the reference samples more precisely with agroecozone boundaries (Figure S10B) provides
604 further insight into error patterns (Table S4). Coastal savannas in the southeast had the highest
605 Producer’s accuracy but lowest User’s accuracy for the per-pixel map, presumably because this region’s
606 numerous areas of high density cropland, combined with low woody cover in surrounding uncultivated
607 areas, helped to promote commission error. Maps in the two northern savanna agroecozones had the
608 best balance between omission and commission error, and had the highest overall User’s accuracy. The
609 transitional (between forest and savanna) agroecozone had a very low Producer’s accuracy (21%),
610 which likely reflects the fact that it was divided between several AOIs for mapping (Figure S4), within
611 which it typically covered a smaller share of area relative to the other agroecozones. This likely caused
612 insufficient representation of this AEZ in training samples, particularly in AOIs 10 and 11 (Figure S4B).

613 Beyond the errors linked to regional differences, several other important factors contributed to reducing
614 the accuracies in the cropland class. The first of these stems from the overall mapping extent and the
615 high resolution of the data. Given the goal of developing a country-scale map at high resolution, the
616 attendant data volume required us to use a relatively small set of image features and less than the
617 recommended tree number and depth (Maxwell et al. 2018) in our Random Forests implementation, in
618 order to limit computational costs. Previous work found that Random Forests achieves much better
619 performance on small-scale croplands when trained on a much larger number of features (Debats et al.
620 2016). However, applying such a large feature set within the extent of our AOIs was intractable, as the
621 computing time would have been several-fold larger than the ~4-8 hours of runtime on 800 CPUs
622 required for a single active learning iteration, followed by ~10-14 hours for prediction. This reduced the
623 skill of the model, particularly when it came to differentiating cropland from adjacent bare patches or
624 natural vegetation with sparse herbaceous cover, which were common in many AOIs.

625 The inherent difficulty of the labelling task was another major limiting factor. Our system was designed
626 to minimize the error inherent in labelling, but determining croplands from non-croplands in these
627 agricultural systems can be a difficult task. Labellers have to evaluate multiple image sources and to
628 rely heavily on judgement, which inevitably leads to errors. Interpretation is particularly hard where
629 the background savanna vegetation and croplands have similar reflectance during the dry season, which
630 is a particular problem in AOIs 2 and 3. Smaller field sizes also complicate labelling, as these become

631 increasingly indistinct in the ~4 m PlanetScope composites. The difficulty of labelling is reflected in the
632 magnitude of the Bayesian Risk metrics (Figure S6), and by the average score achieved by each labeller
633 against our training reference dataset (71%; Figure S5B). Although prior work (Rodriguez-Galiano et
634 al. 2012, Mellor et al. 2015) found that Random Forests are robust to label errors, we found that they
635 have substantial impact (Figure 6), which suggest that simply improving label quality may be one of
636 the single most important investments towards improved model accuracy.

637 Image quality was another issue, although primarily in the forested AOIs, where frequent cloud cover
638 reduced the number of available images in all seasons, resulting in composites with more brightness
639 artifacts and blur (Figure 4). This impacted labellers' abilities to discern fields, and doubtless affected
640 model predictions. There is little to be done to mitigate these errors, short of confining imagery to the
641 less cloudy dry season, which may further undermine model performance, given the importance of
642 multi-temporal imagery for cropland classification (Debats et al. 2016, Defourny et al. 2019).

643 Composite quality could be improved by using imagery from the same seasons over multiple years, but
644 this would undermine the goal of developing annual maps, while the dynamism of the croplands would
645 blur field boundaries within the imagery.

646 The final major source of error arose from the segmentation process. The vectorized maps had high
647 commission errors caused by uncertainties in the Random Forests predictions. Model uncertainty led to
648 many pixels over non-crop areas with probabilities straddling the 0.5 classification threshold. Segments
649 that intersected such areas were retained as fields when the average probability of intersecting pixels
650 exceeded 0.5. A more accurate classifier would reduce such errors, or the application of a locally
651 varying classification threshold (e.g. Waldner and Diakogiannis 2020). Over-merging was another
652 source of error in the segmentation algorithm, which in some areas led to overestimated field sizes and
653 unrealistic shapes, particularly in high density croplands (e.g. in AOIs 2 and 8; Figure 7) where the
654 boundaries between adjacent fields are often indistinct in the PlanetScope imagery. Minimizing or
655 preventing merging would help in such cases, although could result in the opposite problem,
656 over-segmentation, and thereby underestimate field size.

657 4.2 Error mitigation features

658 Despite the error sources mentioned above, several features of the system proved effective in mitigating
659 error, leading to a higher overall accuracy than would have otherwise been possible. Label accuracy
660 assessment and consensus labelling appeared to be the most effective error mitigation tools. Label
661 accuracy measures allowed us to quantify the substantial impact of label error on model performance
662 (Figure 6), while consensus labels substantially reduced individual labelling errors, resulting in maps
663 that were more accurate than they would have been had we used individually generated labels.
664 Labeller-specific accuracy measures also helped to improve the quality of the consensus labels, by
665 placing higher weight on labels more likely to be accurate during the merging process, rather than
666 giving equal weight to potentially less accurate labels. The ability to select the most accurate
667 individual labels for a site also allowed us to develop independent estimates of field size to which
668 measures of confidence can be attached (Figure S5B), which we were in turn able to correct estimates
669 of field sizes and numbers (Table 3).

670 The active learning approach helped to improve overall model performance relative to randomized
671 training site selection, in line with findings from two recent efforts (Debats et al. 2017, Hamrouni et al.
672 2021). Although the performance gains relative to randomized model training that we observed were
673 smaller (e.g. Debats et al. (2017) 29% higher model performance after one iteration, and 8% higher on
674 the final iterations), those comparisons were made from lower initial bases, with initial training samples
675 that were less than 1/10 the size, in terms of pixels, of our initial training sample. Our large initial
676 randomly selected sample (500 grid cells) meant that our models were substantially trained before they
677 were exposed to actively selected labels, thereby diluting their impact on performance. Nevertheless,
678 the higher average performance of the active approach across three performance metrics demonstrated
679 its effectiveness. Most notable were the larger improvements seen in the F1 score (Figure S8), a
680 balanced performance metric. Gains in accuracy and AUC were smaller. For accuracy, the reduced
681 advantage was primarily due to active learning being outperformed by randomized training after the
682 first iteration in AOI 15, which proved one of the hardest AOIs to both map and label. Active learning
683 likely resulted in the selection of sites that were harder to label than randomly selected ones, leading to
684 more label error, and thus initially lower model accuracy. However, this deficit was overcome by the 5th
685 iteration. The plateau in AUC gains at 0.5% better than randomized training reflects the findings that

686 active learning reduced both the false and true positive rates, the two inputs to AUC. Although the
687 decline of the false positive rate (30.7% between Iterations 0 and 3) was nearly three times larger than
688 that of the true positive rate (10.9%), AUC should be quite sensitive to the reduction in the latter, as it
689 assesses how the tradeoff between the two rates varies across a full range of possible classification
690 thresholds (Pontius and Si 2014).

691 The detail, temporal precision, and large extent of these maps was enabled by the system’s ability to
692 process PlanetScope data, which is currently the only source of sub-5 meter imagery with daily
693 coverage (McCabe et al. 2017). Daily revisits are important for creating seasonal composites within a
694 single year over cloudy areas. The compositing technique we developed allowed us to develop a
695 complete image catalog for the country representing the two seasons for 2018 agricultural year.
696 Although Sentinel-2 is free, has better radiometric quality, and has sufficiently high resolution (10 m) to
697 accurately classify small-scale agricultural systems (e.g. Defourny et al. 2019, Kerner et al. 2020), its
698 5-day interval may be too infrequent to generate sufficiently cloud-free composites during the growing
699 season over southern Ghana. Sentinel-1 is not affected by the same problem, but labelling fields in
700 more coarsely resolved radar images is challenging.

701 **4.3 Lingerin**g questions

702 Several potential issues not addressed in our assessment merit further exploration. One of these is the
703 degree of correspondence between image- and ground-collected labels. However, such comparisons may
704 reveal unresolvable differences between the two perspectives. The highly dynamic nature of many
705 agricultural systems means that relatively narrow differences between the dates of ground- and
706 image-based digitizing campaigns can lead to substantial disagreement between the resulting field
707 boundaries, simply because the fields themselves may have shifted during the interval (Elmes et al.
708 2020). These discrepancies could be further exacerbated by differences in the definition of what
709 constitutes a field, which might vary on the ground depending on who is being asked, or who is doing
710 the collecting. These factors suggest that ground versus image label differences would not necessarily
711 indicate how far image-based labellers were from the “truth.” Nevertheless, a comparison against
712 ground data would help to assess how accurately image-collected labels capture the typical size of fields,
713 and thus merits further investigation for this reason.

714 The temporal discrepancies mentioned above (and discussed in Elmes et al. 2020) are another reason
715 why we chose not to label on basemap imagery (in addition to restrictive usage terms), which is
716 typically several years old (Lesiv et al. 2018). However, we did not assess whether the higher label
717 accuracy one might achieve by digitizing on a <1-2 m resolution basemap would offset model errors
718 caused by temporal mismatches.

719 Another potential issue is the degree to which our assessment of the impact of label error on model
720 performance (Figure 6) was influenced by the validation labels we used, which were generated using the
721 consensus method. This could have confounded the assessment, particularly when comparing models
722 trained with the most accurate individual label and those trained with consensus labels. However, the
723 visual assessment of their resulting probability maps confirm the differences in scores: consensus and
724 most accurate individual labels produce nearly identical maps with relatively high certainty, while low
725 quality labels led to a markedly less certain map (Figure S9).

726 **4.4 Next steps**

727 The maps presented here represent a version 1 product that is freely available to use, along with its
728 underlying code (see SI for details). These data were developed according to the recommended best
729 practices for training and assessing error in machine learning models (Elmes et al. 2020). In their
730 current form, the maps may be useful for a variety of research applications. For example, analyzing the
731 distributions of values in the probability maps may provide additional insight into the relative extents
732 of active versus fallow croplands (Tong et al. 2020). However, use of these data, particularly for
733 decision-making processes (e.g. cropped area estimates), should be careful to account for the reported
734 errors (Olofsson et al. 2014, Stehman and Foody 2019).

735 To facilitate the next step, generating more accurate version 2 maps, several improvements will be
736 made. The first is to replace Random Forests with a more advanced convolutional neural network
737 (CNN), which can generate and learn from a large number of features representing a variety of spatial
738 scales (Ma et al. 2019). Recent work suggests that a common architecture such as U-Net, when trained
739 to distinguish field edges from interiors and combined with a post-hoc segmentation routine, is effective
740 in delineating field boundaries (Waldner and Diakogiannis 2020). Our system can readily incorporate
741 such a model. The labelling platform already provides the methods needed to develop and assess the

742 quality of labels that include field edges and interior classes, while active learning has proven to be
743 effective for optimizing training datasets for deep learning models (Liu et al. 2017, Cao et al. 2020).
744 Our current framework can be adjusted so that it starts by training a CNN from scratch with a large
745 initial random sample, and then uses a transfer learning approach (Pan and Yang 2010) to update the
746 model with the most informative samples from different AOIs or agroecozones.

747 **4.5 Conclusion**

748 This work demonstrates a proof of concept for developing high resolution, annual maps of
749 smallholder-dominated croplands at national to regional scales, using a framework that can be readily
750 updated to improve map accuracy as technologies improve. Maps that include information on field
751 boundaries can help improve remote estimation of crop planted area and yield, and provide deeper
752 insights into important socioeconomic aspects of agricultural systems, such as the relationships between
753 agricultural productivity and livelihoods. Such maps will be important for developing an understanding
754 of the rapid agricultural change that is currently unfolding throughout much of the continent.

755 **5 Acknowledgements**

756 The primary support for this work was provided by Omidyar Network’s Property Rights Initiative, now
757 PLACE. Additional support was provided by NASA (80NSSC18K0158), the National Science
758 Foundation (SES-1801251; SES-1832393), and Princeton University. Computing support was provided
759 by the AWS Cloud Credits for Research program and the Amazon Sustainability Data Initiative.
760 Azavea provided significant contributions in engineering the machine learning pipeline. We thank
761 Meridia for providing information about local cropping systems and the characteristics of fields, and
762 Radiant Earth Foundation for advice and guidance regarding machine learning best practices. We
763 thank Manushi Trivedi, Sitian Xiong, and Tammy Woodard for their contributions to the underlying
764 datasets and methods, and Michelle Gathigi, Omar Shehe, and Primoz Kovacic for support and
765 management of the labelling efforts.

766 **6 References**

767 Azavea. 2020. Raster Foundry. <https://github.com/raster-foundry/raster-foundry>.

- 768 Bey, A., A. Sánchez-Paus Díaz, D. Maniatis, G. Marchi, D. Mollicone, S. Ricci, J.-F. Bastin, R. Moore,
769 S. Federici, M. Rezende, C. Patriarca, R. Turia, G. Gamoga, H. Abe, E. Kaidong, and G. Miceli.
770 2016. Collect Earth: Land Use and Land Cover Assessment through Augmented Visual
771 Interpretation. *Remote Sensing* 8:807.
- 772 Breiman, L. 2001. Random Forests. *Machine Learning* 45:5–32.
- 773 Bullock, E. L., S. P. Healey, Z. Yang, P. Oduor, N. Gorelick, S. Omondi, E. Ouko, and W. B. Cohen.
774 2021. Three Decades of Land Cover Change in East Africa. *Land* 10:150.
- 775 Cao, X., J. Yao, Z. Xu, and D. Meng. 2020. Hyperspectral Image Classification With Convolutional
776 Neural Network and Active Learning. *IEEE Transactions on Geoscience and Remote Sensing*
777 58:4604–4616.
- 778 Carletto, C., S. Gourlay, and P. Winters. 2015. From Guesstimates to GPStimates: Land Area
779 Measurement and Implications for Agricultural Analysis. *Journal of African Economies* 24:593–628.
- 780 Carletto, C., S. Savastano, and A. Zezza. 2013. Fact or artifact: The impact of measurement errors on
781 the farm sizeproductivity relationship. *Journal of Development Economics* 103:254–261.
- 782 Chen, J., J. Chen, A. Liao, X. Cao, L. Chen, X. Chen, C. He, G. Han, S. Peng, M. Lu, W. Zhang, X.
783 Tong, and J. Mills. 2015. Global land cover mapping at 30 m resolution: A POK-based operational
784 approach. *ISPRS Journal of Photogrammetry and Remote Sensing* 103:7–27.
- 785 Cohn, D., L. Atlas, and R. Ladner. 1994. Improving generalization with active learning. *Machine*
786 *Learning* 15:201–221.
- 787 Dark, S. J., and D. Bram. 2007. The modifiable areal unit problem (MAUP) in physical geography.
788 *Progress in Physical Geography* 31:471–479.
- 789 Davis, K. F., H. I. Koo, J. Dell’Angelo, P. D’Odorico, L. Estes, L. J. Kehoe, M. Kharratzadeh, T.
790 Kuemmerle, D. Machava, A. de J. R. Pais, N. Ribeiro, M. C. Rulli, and M. Tatlhego. 2020. Tropical
791 forest loss enhanced by large-scale land acquisitions. *Nature Geoscience*:1–7.
- 792 Debats, S. R., L. D. Estes, D. R. Thompson, and K. K. Caylor. 2017. Integrating active learning and
793 crowdsourcing into large-scale supervised landcover mapping algorithms. *PeerJ Preprints*.
- 794 Debats, S. R., D. Luo, L. D. Estes, T. J. Fuchs, and K. K. Caylor. 2016. A generalized computer vision
795 approach to mapping crop fields in heterogeneous agricultural landscapes. *Remote Sensing of*
796 *Environment* 179:210–221.
- 797 Defourny, P., S. Bontemps, N. Bellemans, C. Cara, G. Dedieu, E. Guzzonato, O. Hagolle, J. Inglada, L.
798 Nicola, T. Rabaute, M. Savinaud, C. Udrouiu, S. Valero, A. Bégué, J.-F. Dejoux, A. El Harti, J.
799 Ezzahar, N. Kussul, K. Labbassi, V. Lebourgeois, Z. Miao, T. Newby, A. Nyamugama, N. Salh, A.
800 Shelestov, V. Simonneaux, P. S. Traore, S. S. Traore, and B. Koetz. 2019. Near real-time
801 agriculture monitoring at national scale at parcel resolution: Performance assessment of the
802 Sen2-Agri automated system in various cropping systems around the world. *Remote Sensing of*
803 *Environment* 221:551–568.

- 804 Desiere, S., and D. Jolliffe. 2018. Land productivity and plot size: Is measurement error driving the
805 inverse relationship? *Journal of Development Economics* 130:84–98.
- 806 Drusch, M., U. Del Bello, S. Carlier, O. Colin, V. Fernandez, F. Gascon, B. Hoersch, C. Isola, P.
807 Laberinti, P. Martimort, A. Meygret, F. Spoto, O. Sy, F. Marchese, and P. Bargellini. 2012.
808 Sentinel-2: ESA’s Optical High-Resolution Mission for GMES Operational Services. *Remote
809 Sensing of Environment* 120:25–36.
- 810 Dwyer, J. L., D. P. Roy, B. Sauer, C. B. Jenkerson, H. K. Zhang, and L. Lymburner. 2018. Analysis
811 Ready Data: Enabling Analysis of the Landsat Archive. *Remote Sensing* 10:1363.
- 812 Elmes, A., H. Alemohammad, R. Avery, K. Caylor, J. R. Eastman, L. Fishgold, M. A. Friedl, M. Jain,
813 D. Kohli, J. C. Laso Bayas, D. Lunga, J. L. McCarty, R. G. Pontius, A. B. Reinmann, J. Rogan, L.
814 Song, H. Stoyanova, S. Ye, Z.-F. Yi, and L. Estes. 2020. Accounting for training data error in
815 machine learning applied to Earth Observations. *Remote Sensing* 12:1034.
- 816 ESA. (n.d.). ESA CCI LAND COVER S2 prototype Land Cover 20m map of Africa 2016.
817 <http://2016africalandcover20m.esrin.esa.int/>.
- 818 Estes, L., P. Chen, S. Debats, T. Evans, S. Ferreira, T. Kuemmerle, G. Ragazzo, J. Sheffield, A. Wolf,
819 E. Wood, and K. Caylor. 2018. A large-area, spatially continuous assessment of land cover map
820 error and its impact on downstream analyses. *Global Change Biology* 24:322–337.
- 821 Estes, L. D., D. McRitchie, J. Choi, S. Debats, T. Evans, W. Guthe, D. Luo, G. Ragazzo, R. Zempleni,
822 and K. K. Caylor. 2016a. A platform for crowdsourcing the creation of representative, accurate
823 landcover maps. *Environmental Modelling & Software* 80:41–53.
- 824 Estes, L. D., T. Searchinger, M. Spiegel, D. Tian, S. Sickinga, M. Mwale, L. Kehoe, T. Kuemmerle, A.
825 Berven, N. Chaney, J. Sheffield, E. F. Wood, and K. K. Caylor. 2016b. Reconciling agriculture,
826 carbon and biodiversity in a savannah transformation frontier. *Phil. Trans. R. Soc. B* 371:20150316.
- 827 Feder, G. 1985. The relation between farm size and farm productivity: The role of family labor,
828 supervision and credit constraints. *Journal of Development Economics* 18:297–313.
- 829 Forkuor, G., C. Conrad, M. Thiel, T. Ullmann, and E. Zoungrana. 2014. Integration of Optical and
830 Synthetic Aperture Radar Imagery for Improving Crop Mapping in Northwestern Benin, West
831 Africa. *Remote Sensing* 6:6472–6499.
- 832 Fourie, A. 2009. Better Crop Estimates in South Africa. *ArcUser Online*.
- 833 Fritz, S., I. McCallum, C. Schill, C. Perger, L. See, D. Schepaschenko, M. van der Velde, F. Kraxner,
834 and M. Obersteiner. 2012. Geo-Wiki: An online platform for improving global land cover.
835 *Environmental Modelling & Software* 31:110–123.
- 836 Fritz, S., L. See, I. McCallum, C. Schill, M. Obersteiner, M. van der Velde, H. Boettcher, P. Havlik,
837 and F. Achard. 2011. Highlighting continued uncertainty in global land cover maps for the user
838 community. *Environmental Research Letters* 6:044005.
- 839 Fritz, S., L. See, I. McCallum, L. You, A. Bun, E. Moltchanova, M. Duerauer, F. Albrecht, C. Schill, C.
840 Perger, P. Havlik, A. Mosnier, P. Thornton, U. Wood-Sichra, M. Herrero, I. Becker-Reshef, C.

- 841 Justice, M. Hansen, P. Gong, S. Abdel Aziz, A. Cipriani, R. Cumani, G. Cecchi, G. Conchedda, S.
842 Ferreira, A. Gomez, M. Haffani, F. Kayitakire, J. Malanding, R. Mueller, T. Newby, A. Nonguierma,
843 A. Olusegun, S. Ortner, D. R. Rajak, J. Rocha, D. Schepaschenko, M. Schepaschenko, A. Terekhov,
844 A. Tiangwa, C. Vancutsem, E. Vintrou, W. Wenbin, M. van der Velde, A. Dunwoody, F. Kraxner,
845 and M. Obersteiner. 2015. Mapping global cropland and field size. *Global Change Biology*
846 21:1980–1992.
- 847 Fritz, S., L. See, and F. Rembold. 2010. Comparison of global and regional land cover maps with
848 statistical information for the agricultural domain in Africa. *International Journal of Remote*
849 *Sensing* 31:2237–2256.
- 850 Fritz, S., L. See, L. You, C. Justice, I. Becker-Reshef, L. Bydekerke, R. Cumani, P. Defourny, K. Erb, J.
851 Foley, S. Gilliams, P. Gong, M. Hansen, T. Hertel, M. Herold, M. Herrero, F. Kayitakire, J. Latham,
852 O. Leo, I. McCallum, M. Obersteiner, N. Ramankutty, J. Rocha, H. Tang, P. Thornton, C.
853 Vancutsem, M. van der Velde, S. Wood, and C. Woodcock. 2013. The need for improved maps of
854 global cropland. *Eos, Transactions American Geophysical Union* 94:31–32.
- 855 Gorelick, N., M. Hancher, M. Dixon, S. Ilyushchenko, D. Thau, and R. Moore. 2017. Google Earth
856 Engine: Planetary-scale geospatial analysis for everyone. *Remote Sensing of Environment* 202:18–27.
- 857 Hackman, K. O., P. Gong, and J. Wang. 2017. New land-cover maps of Ghana for 2015 using Landsat
858 8 and three popular classifiers for biodiversity assessment. *International Journal of Remote Sensing*
859 38:4008–4021.
- 860 Hamrouni, Y., E. Paillassa, V. Chéret, C. Monteil, and D. Sheeren. 2021. From local to global: A
861 transfer learning-based approach for mapping poplar plantations at national scale using Sentinel-2.
862 *ISPRS Journal of Photogrammetry and Remote Sensing* 171:76–100.
- 863 Houborg, R., and M. McCabe. 2018. Daily Retrieval of NDVI and LAI at 3 m Resolution via the
864 Fusion of CubeSat, Landsat, and MODIS Data. *Remote Sensing* 10:890.
- 865 Jayne, T. s., J. Chamberlin, L. Traub, N. Sitko, M. Muyanga, F. K. Yeboah, W. Anseeuw, A. Chapoto,
866 A. Wineman, C. Nkonde, and R. Kachule. 2016. Africa’s changing farm size distribution patterns:
867 The rise of medium-scale farms. *Agricultural Economics* 47:197–214.
- 868 Kansanga, M., P. Andersen, D. Kpienbaareh, S. Mason-Renton, K. Atuoye, Y. Sano, R. Antabe, and I.
869 Luginaah. 2019. Traditional agriculture in transition: Examining the impacts of agricultural
870 modernization on smallholder farming in Ghana under the new Green Revolution. *International*
871 *Journal of Sustainable Development & World Ecology* 26:11–24.
- 872 Kehoe, L., A. Romero-Muñoz, L. Estes, H. Kreft, E. Polaina, and T. Kuemmerle. 2017. Nature at risk
873 under future agricultural expansion and intensification. *Nature Ecology and Evolution* 1:1129–1135.
- 874 Kerner, H., G. Tseng, I. Becker-Reshef, C. Nakalembe, B. Barker, B. Munshell, M. Paliyam, and M.
875 Hosseini. 2020. Rapid Response Crop Maps in Data Sparse Regions. arXiv:2006.16866 [cs, eess].
- 876 Lesiv, M., S. Fritz, I. McCallum, N. Tsendbazar, M. Herold, J.-F. Pekel, M. Buchhorn, B. Smets, and
877 R. Van De Kerchove. 2017, November. Evaluation of ESA CCI prototype land cover map at 20m.
878 Monograph, <http://pure.iiasa.ac.at/id/eprint/14979/>.

- 879 Lesiv, M., J. C. Laso Bayas, L. See, M. Duerauer, D. Dahlia, N. Durando, R. Hazarika, P. Kumar
880 Sahariah, M. Vakolyuk, V. Blyshchyk, A. Bilous, A. Perez-Hoyos, S. Gengler, R. Prestele, S. Bilous,
881 I. ul H. Akhtar, K. Singha, S. B. Choudhury, T. Chetri, Ž. Malek, K. Bungnamei, A. Saikia, D.
882 Sahariah, W. Narzary, O. Danylo, T. Sturn, M. Karner, I. McCallum, D. Schepaschenko, E.
883 Moltchanova, D. Fraisl, I. Moorthy, and S. Fritz. 2019. Estimating the global distribution of field
884 size using crowdsourcing. *Global Change Biology* 25:174–186.
- 885 Lesiv, M., L. See, J. Laso Bayas, T. Sturn, D. Schepaschenko, M. Karner, I. Moorthy, I. McCallum, and
886 S. Fritz. 2018. Characterizing the spatial and temporal availability of very high resolution satellite
887 imagery in Google Earth and Microsoft Bing maps as a source of reference data. *Land* 7:118.
- 888 Levin, G. 2006. Farm size and landscape composition in relation to landscape changes in Denmark.
889 *Geografisk Tidsskrift-Danish Journal of Geography* 106:45–59.
- 890 Licker, R., M. Johnston, J. A. Foley, C. Barford, C. J. Kucharik, C. Monfreda, and N. Ramankutty.
891 2010. Mind the gap: How do climate and agricultural management explain the ‘yield gap’ of
892 croplands around the world? *Global Ecology and Biogeography* 19:769–782.
- 893 Liu, P., H. Zhang, and K. B. Eom. 2017. Active Deep Learning for Classification of Hyperspectral
894 Images. *IEEE Journal of Selected Topics in Applied Earth Observations and Remote Sensing*
895 10:712–724.
- 896 Lobell, D. B., K. G. Cassman, and C. B. Field. 2009. Crop Yield Gaps: Their Importance, Magnitudes,
897 and Causes. *Annual Review of Environment and Resources* 34:179–204.
- 898 Lowder, S. K., J. Scoet, and T. Raney. 2016. The Number, Size, and Distribution of Farms,
899 Smallholder Farms, and Family Farms Worldwide. *World Development* 87:16–29.
- 900 Ma, L., Y. Liu, X. Zhang, Y. Ye, G. Yin, and B. A. Johnson. 2019. Deep learning in remote sensing
901 applications: A meta-analysis and review. *ISPRS Journal of Photogrammetry and Remote Sensing*
902 152:166–177.
- 903 Maxwell, A. E., T. A. Warner, and F. Fang. 2018. Implementation of machine-learning classification in
904 remote sensing: An applied review. *International Journal of Remote Sensing* 39:2784–2817.
- 905 McCabe, M. F., M. Rodell, D. E. Alsdorf, D. G. Miralles, R. Uijlenhoet, W. Wagner, A. Lucieer, R.
906 Houborg, N. E. C. Verhoest, T. E. Franz, J. Shi, H. Gao, and E. F. Wood. 2017. The future of
907 Earth observation in hydrology. *Hydrology and Earth System Sciences* 21:3879–3914.
- 908 Mellor, A., S. Boukir, A. Haywood, and S. Jones. 2015. Exploring issues of training data imbalance
909 and mislabelling on random forest performance for large area land cover classification using the
910 ensemble margin. *ISPRS Journal of Photogrammetry and Remote Sensing* 105:155–168.
- 911 Morris, M., and D. Byerlee. 2009. *Awakening Africa’s Sleeping Giant*. World Bank and FAO,
912 Washington, DC.
- 913 Mueller, N. D., J. S. Gerber, M. Johnston, D. K. Ray, N. Ramankutty, and J. A. Foley. 2012. Closing
914 yield gaps through nutrient and water management. *Nature* 490:254–257.

- 915 Neubert, P., and P. Protzel. 2014. Compact Watershed and Preemptive SLIC: On Improving Trade-offs
916 of Superpixel Segmentation Algorithms. Pages 996–1001 2014 22nd International Conference on
917 Pattern Recognition. IEEE, Stockholm, Sweden.
- 918 Olofsson, P., G. M. Foody, M. Herold, S. V. Stehman, C. E. Woodcock, and M. A. Wulder. 2014. Good
919 practices for estimating area and assessing accuracy of land change. *Remote Sensing of*
920 *Environment* 148:42–57.
- 921 Olofsson, P., G. M. Foody, S. V. Stehman, and C. E. Woodcock. 2013. Making better use of accuracy
922 data in land change studies: Estimating accuracy and area and quantifying uncertainty using
923 stratified estimation. *Remote Sensing of Environment* 129:122–131.
- 924 Openshaw, S., and P. J. Taylor. 1979. A million or so correlation coefficients: Three experiments on
925 the modifiable areal unit problem. *Statistical applications in the spatial sciences* 21:127–144.
- 926 Pan, S. J., and Q. Yang. 2010. A Survey on Transfer Learning. *IEEE Transactions on Knowledge and*
927 *Data Engineering* 22:1345–1359.
- 928 Persello, C., V. A. Tolpekin, J. R. Bergado, and R. A. de By. 2019. Delineation of agricultural fields in
929 smallholder farms from satellite images using fully convolutional networks and combinatorial
930 grouping. *Remote Sensing of Environment* 231:111253.
- 931 PlanetTeam. 2018. Planet application program interface: In space for life on Earth.
932 <https://api.planet.com>, San Francisco, CA.
- 933 Pontius, R. G., and K. Si. 2014. The total operating characteristic to measure diagnostic ability for
934 multiple thresholds. *International Journal of Geographical Information Science* 28:570–583.
- 935 Qiu, S., Z. Zhu, and C. E. Woodcock. 2020. Cirrus clouds that adversely affect Landsat 8 images:
936 What are they and how to detect them? *Remote Sensing of Environment* 246:111884.
- 937 Rodriguez-Galiano, V. F., B. Ghimire, J. Rogan, M. Chica-Olmo, and J. P. Rigol-Sanchez. 2012. An
938 assessment of the effectiveness of a random forest classifier for land-cover classification. *ISPRS*
939 *Journal of Photogrammetry and Remote Sensing* 67:93–104.
- 940 Rulli, M. C., and P. D’Odorico. 2014. Food appropriation through large scale land acquisitions.
941 *Environmental Research Letters* 9:064030.
- 942 Samberg, L. H., J. S. Gerber, N. Ramankutty, M. Herrero, and P. C. West. 2016. Subnational
943 distribution of average farm size and smallholder contributions to global food production.
944 *Environmental Research Letters* 11:124010.
- 945 Searchinger, T. D., L. Estes, P. K. Thornton, T. Beringer, A. Notenbaert, D. Rubenstein, R. Heimlich,
946 R. Licker, and M. Herrero. 2015. High carbon and biodiversity costs from converting Africa’s wet
947 savannahs to cropland. *Nature Climate Change* 5:481–486.
- 948 Searchinger, T., R. Waite, C. Hanson, J. Ranganathan, P. Dumas, E. Matthews, and C. Klirs. 2019.
949 Creating a sustainable food future: A menu of solutions to feed nearly 10 billion people by 2050.
950 Final report. WRI.

- 951 Stehman, S. V., and G. M. Foody. 2019. Key issues in rigorous accuracy assessment of land cover
952 products. *Remote Sensing of Environment* 231:111199.
- 953 Sulla-Menashe, D., J. M. Gray, S. P. Abercrombie, and M. A. Friedl. 2019. Hierarchical mapping of
954 annual global land cover 2001 to present: The MODIS Collection 6 Land Cover product. *Remote*
955 *Sensing of Environment* 222:183–194.
- 956 Tong, X., M. Brandt, P. Hiernaux, S. Herrmann, L. V. Rasmussen, K. Rasmussen, F. Tian, T.
957 Tagesson, W. Zhang, and R. Fensholt. 2020. The forgotten land use class: Mapping of fallow fields
958 across the Sahel using Sentinel-2. *Remote Sensing of Environment* 239:111598.
- 959 Tuia, D., M. Volpi, L. Copa, M. Kanevski, and J. Munoz-Mari. 2011. A Survey of Active Learning
960 Algorithms for Supervised Remote Sensing Image Classification. *IEEE Journal of Selected Topics in*
961 *Signal Processing* 5:606–617.
- 962 Van Vliet, N., O. Mertz, T. Birch-Thomsen, and B. Schmook. 2013. Is There a Continuing Rationale
963 for Swidden Cultivation in the 21st Century? *Human Ecology* 41:1–5.
- 964 Visvalingam, M., and J. D. Whyatt. 1993. Line generalisation by repeated elimination of points. *The*
965 *Cartographic Journal* 30:46–51.
- 966 Von Braun, J. 2004. Small-scale farmers in a liberalized trade environment. Page 21.
- 967 Waldner, F., and F. I. Diakogiannis. 2020. Deep learning on edge: Extracting field boundaries from
968 satellite images with a convolutional neural network. *Remote Sensing of Environment* 245:111741.
- 969 Waldner, F., A. Schucknecht, M. Lesiv, J. Gallego, L. See, A. Pérez-Hoyos, R. d’Andrimont, T. de
970 Maet, J. C. L. Bayas, S. Fritz, O. Leo, H. Kerdiles, M. Díez, K. Van Tricht, S. Gilliams, A.
971 Shelestov, M. Lavreniuk, M. Simões, R. Ferraz, B. Bellón, A. Bégué, G. Hazeu, V. Stonacek, J.
972 Kolomaznik, J. Misurec, S. R. Verón, D. de Abelleira, D. Plotnikov, L. Mingyong, M. Singha, P.
973 Patil, M. Zhang, and P. Defourny. 2019. Conflation of expert and crowd reference data to validate
974 global binary thematic maps. *Remote Sensing of Environment* 221:235–246.
- 975 Wilson, A. M., and W. Jetz. 2016. Remotely Sensed High-Resolution Global Cloud Dynamics for
976 Predicting Ecosystem and Biodiversity Distributions. *PLOS Biology* 14:e1002415.
- 977 Wulder, M. A., J. C. White, T. R. Loveland, C. E. Woodcock, A. S. Belward, W. B. Cohen, E. A.
978 Fosnight, J. Shaw, J. G. Masek, and D. P. Roy. 2016. The global Landsat archive: Status,
979 consolidation, and direction. *Remote Sensing of Environment* 185:271–283.
- 980 Xiong, J., P. S. Thenkabail, J. C. Tilton, M. K. Gumma, P. Teluguntla, A. Oliphant, R. G. Congalton,
981 K. Yadav, and N. Gorelick. 2017. Nominal 30-m Cropland Extent Map of Continental Africa by
982 Integrating Pixel-Based and Object-Based Algorithms Using Sentinel-2 and Landsat-8 Data on
983 Google Earth Engine. *Remote Sensing* 9:1065.
- 984 Ye, S., R. G. Pontius, and R. Rakshit. 2018. A review of accuracy assessment for object-based image
985 analysis: From per-pixel to per-polygon approaches. *ISPRS Journal of Photogrammetry and*
986 *Remote Sensing* 141:137–147.

- 987 Yizong Cheng. 1995. Mean shift, mode seeking, and clustering. *IEEE Transactions on Pattern Analysis*
988 *and Machine Intelligence* 17:790–799.
- 989 Zeng, Z., L. Estes, A. D. Ziegler, A. Chen, T. Searchinger, F. Hua, K. Guan, A. Jintrawet, and E. F.
990 Wood. 2018. Highland cropland expansion and forest loss in Southeast Asia in the twenty-first
991 century. *Nature Geoscience* 11:556–562.
- 992 Zhang, Y., B. Guindon, and J. Cihlar. 2002. An image transform to characterize and compensate for
993 spatial variations in thin cloud contamination of Landsat images. *Remote Sensing of Environment*
994 82:173–187.
- 995 Zhu, X. X., D. Tuia, L. Mou, G.-S. Xia, L. Zhang, F. Xu, and F. Fraundorfer. 2017. Deep learning in
996 remote sensing: A review. *IEEE Geoscience and Remote Sensing Magazine* 5:8–36.
- 997 Zhu, Z., and C. E. Woodcock. 2012. Object-based cloud and cloud shadow detection in Landsat
998 imagery. *Remote Sensing of Environment* 118:83–94.