Advancing on the promises of techno-ecological nature-based solutions: A framework for green technology in water supply & treatment

Emma A. J. Blackburn[†], Monica B. Emelko^{*†}, Sarah Dickson-Anderson[‡], Micheal Stone[§]

⁺ Water Science, Technology & Policy Group, Department of Civil and Environmental

Engineering, University of Waterloo, Waterloo, Ontario, Canada.

[‡] Department of Civil Engineering, McMaster University, Hamilton, Ontario, Canada.

[§] Department of Geography and Environmental Management, University of Waterloo, Ontario,

Canada.

This manuscript has been published in *Blue-Green Systems*. Please refer to the following citation: Emma A. J. Blackburn, Monica B. Emelko, Sarah Dickson-Anderson, Micheal Stone; Advancing on the promises of techno-ecological nature-based solutions: A framework for green technology in water supply and treatment. *Blue-Green Systems* 1 January 2021; 3 (1): 81–94. doi: https://doi.org/10.2166/bgs.2021.008

1 ABSTRACT

- 2 Nature-based solutions (NBS) are increasingly proposed for effectively and adaptively addressing
- 3 societal challenges such as water security and natural disasters. However, NBS that are exclusively
- 4 reliant on natural processes are not fit-for-purpose for the provision of safe drinking water some range
- 5 of built technology is required. There is a wide spectrum of techno-ecological NBS 'green technologies'
- 6 that are fit-for-purpose in the treatment and distribution of safe drinking water. A framework was
- 7 developed to enable an accurate and transparent description of the 'green' attributes of technology –
- 8 including green infrastructure in the water industry. The framework differentiates technology
- 9 'greenness' by relatively examining key attributes that may cause environmental impacts across the
- 10 technology's life cycle through the lens of the environmental setting in which it is applied. In the water
- 11 industry, green technology can be described by four main attributes: natural-resource basis, energy
- 12 consumption, waste production, and footprint. These attributes are closely linked and must be
- 13 considered relative to the biophysical and human environments in which they are applied and the other
- 14 technologies to which they are being compared. The use of the framework can facilitate techno-
- ecological decision-making that strives to address diverse stakeholder priorities including the influence
- 16 of sociocultural factors on the green technology preferences of individuals, groups, or communities.

17 KEYWORDS

18 biofiltration, drinking water, green infrastructure, greenness, natural capital, source water protection

19 **GRAPHICAL ABSTRCT**

20 [Figure 2]

21 INTRODUCTION

22 Nature-based solutions (NBS) are increasingly proposed for effectively and adaptively addressing 23 societal challenges such as water security and natural disasters – they have been defined as 'actions to 24 protect, sustainably manage and restore natural or modified ecosystems ... while simultaneously 25 providing human well-being and biodiversity benefits' (Cohen-Shacham et al. 2016). NBS are growing in 26 popularity globally; however, they are not a panacea to water security, climate change, or any other of 27 society's grand challenges. The practical implementation of NBS can be challenging because of 28 differences in what should be prioritized and the relative importance associated with those priorities. 29 These challenges were recently highlighted by O'Sullivan et al. (2020) who cautioned that NBS have 30 sometimes been framed too idealistically, leading to undervaluation of biodiversity and unrealistic 31 expectations of the capacity of natural processes to provide the 'solutions' that are needed. Recognition 32 that the value and limits of NBS must be understood, so that they are robust and resilient is also growing 33 (Seddon et al. 2021). While rigid differentiation between nature- and technology-based approaches for 34 managing some challenges has been suggested (Mustafa et al. 2019), efforts to describe the synergies between technological and ecological systems are growing (Bakshi et al. 2015) and discussions of NBS 35 that are enhanced by or integrated with technology – 'techno-ecological NBS' – are emerging. 36 37 In the drinking water industry, the emergence of techno-ecological NBS is evident in industry-wide 38 prioritization of source water protection (SWP) (AWWA 2020) and increasing the promotion of 'green' 39 approaches, such as the use of forest management-based strategies and other NBS for source water 40 quality management and climate change adaptation (Ernst et al. 2004; Emelko et al. 2011; McLain et al. 41 2012; Robinne et al. 2019; Oral et al. 2020). Water managers are increasingly asked to integrate 'green' 42 approaches into water supply and treatment practices. Both 'green infrastructure' and 'green 43 technology' terminologies are used in the water industry. They are also frequently integrated to yield

44 techno-ecological concepts of natural resource-based treatment processes that reflect the technological

aspects of natural landscape processes, such as low-cost cascade aeration systems that enhance the air–
water transfer of atmospheric gases (e.g., oxygen and nitrogen) and volatile organic compounds (Figure
1).

48 The use of 'green infrastructure' in the water industry is consistent with its common broader use, which 49 reflects the practical application, preservation, and enhancement of natural capital using a management 50 approach that 'emphasizes the importance of environmental systems and networks for the direct 51 provision of ecosystem services to human populations' (Chenoweth et al. 2018). Here, the term 'natural 52 capital' is also consistent with its broader use and refers to environmental assets that provide people 53 with free goods and services that are often referred to as ecosystem services (Chenoweth et al. 2018). 54 Thus, in the water industry, 'green infrastructure' not only reflects natural capital, but also often 55 encompasses natural resource-based management approaches to achieve engineering (i.e., treatment) 56 targets – this inter-relationship between green infrastructure and natural capital directly aligns with the 57 recognition that there is a spectrum of degrees of 'naturalness' that ranges from environments with minimal human influence to those that have been built (Chenoweth et al. 2018). 58 59 In contrast, the use of 'green technology' in the water industry tends to reflect approaches that may be 60 linked to, but not necessarily reliant upon natural capital. Notably, while the 'green' descriptor is 61 frequently used interchangeably with 'sustainable' (Ngo et al. 2016), sustainability analysis typically 62 considers broad impacts on the environment, the economy, and society (Purvis et al. 2019). While life 63 cycle analysis is regularly included in technology evaluation and selection in the water industry, all of the 64 pillars of sustainability are not typically reflected in decision-making – even when they are discussed, trade-offs are of course required because of economic limitations. 65

The implementation of 'green technologies' in the water industry tends to focus on the treatment
processes themselves (Wu et al. 2015; Neoh et al. 2016) and reflects various engineering priorities such

68 as energy efficiency and low waste production, which can be described as 'green'. These technologies 69 are generally understood to complement and sometimes replace more traditional 'grey technologies', 70 which are human-engineered without reliance on the practical application, or prioritization of the 71 preservation or enhancement, of natural capital. This is because 'green technologies' are believed to 72 offer environmentally conscientious, energy-efficient, and/ or increasingly economically viable solutions 73 to address challenges such as the need to concurrently protect human health, adapt to climate change-74 exacerbated threats to water security, and reduce the environmental impacts of water treatment and 75 distribution (Gill et al. 2007; Emelko et al. 2011; Ngo et al. 2016).

76 Despite the widespread use of the term 'green' across the broader water sector and within the drinking 77 water industry specifically, there is no consistently applied definition or framework for what constitutes 78 'green technology' or which aspects of 'greenness' are valued. A framework for describing the 'green' 79 attributes of the broad range of technologies – including natural capital – relevant to the water industry 80 is needed, as these attributes dictate how technologies are prioritized relative to others, and whether 81 they are considered 'green' at all. Such a framework will also enable stakeholders to better 82 communicate the technical and engineering aspects of technology approaches that best align with 83 community and individual sociocultural values, beliefs, and attitudes. In addition to the challenges 84 associated with the lack of a framework to describe the 'green' attributes of technologies or 85 infrastructure options for meeting broader water industry objectives, it is important to recognize that 86 'green technology' has not had much uptake in the drinking water industry, as compared to other 87 segments of the water sector.

88 [Figure 1]

The drinking water industry is necessarily conservative and somewhat averse to real or perceived risks
to public health that may be attributed to innovative technologies that are unproven, or require

91 operational shifts for control, relative to conventional technologies. These challenges have been 92 underscored for decades in the lack of widespread uptake of biological treatment processes because of 93 concerns regarding health risks that might be attributable to microbially mediated treatment, difficulties in operation, and unlikely regulatory approvals (Brown et al. 2015). While such concerns are misplaced 94 95 (Brown et al. 2015), well-known events such as the 1993 Milwaukee cryptosporidiosis outbreak, in 96 which more than 50 people died and more than 400,000 people became ill (EPA 1998), serve as stark 97 reminders of the importance of public health protection through the provision of safe drinking water as 98 the industry's paramount objective. Thus, any shifts in the fundamental way in which drinking water is 99 treated and distributed must be approached with clarity in purpose and confidence that public health 100 protection is not compromised.

101 Consistent with that recognition, it has been recently emphasized that the good science that is needed 102 for meaningful advancement of sustainability goals such as the development of NBS requires clearly 103 defined terminology rather than reliance on vague metaphors (Vos et al. 2007; Aronson 2011). 104 Fortunately, the promises of green technology can be advanced in the water supply and treatment 105 sector with sound initial foundations in scientific and engineering principles. These begin with the 106 foremost recognition that all drinking water treatment technologies must be effective for the protection 107 of public health – these targets must be achievable in regular practice, not only at idealized conditions. 108 Thus, any green technologies that would be considered for use within the drinking water industry must 109 be 'fit-for-purpose' for the protection of public health, meaning that they meet or exceed the drinking 110 water treatment performance expectations and regulatory criteria that they are intended to address. 111 For this reason, NBS that are exclusively reliant on natural processes are not fit-for-purpose for the 112 provision of safe drinking water – some range of built technology is required. For example, recent work 113 has demonstrated that viruses can be present in high-quality groundwater supplies and require 114 substantial treatment even in situations where it has been historically believed that no treatment is

required (Borchardt et al. 2012; Emelko et al. 2019). Additional built technologies would be required to
indicate water safety and ensure its safe distribution. In contrast, it will be demonstrated herein that
there is a wide spectrum of techno-ecological NBS – 'green technologies' – that are fit-for-purpose in the
treatment and distribution of safe drinking water.

119 Using the imperative fit-for-purpose criterion as a starting point, a framework is developed herein to 120 enable an accurate and transparent description of the 'green' attributes of technology – including green 121 infrastructure – used in the water industry. It differentiates technology 'greenness' by relatively examining key attributes that may cause environmental impacts across the technology's life cycle 122 123 through the lens of the environmental setting in which it is applied. It is proposed that the framework 124 developed herein can contribute to the development of more comprehensive techno-ecological NBS by 125 providing clear and accurate description of the 'green' attributes of technology options for the water 126 industry, as well as a framework for their relative comparison, thereby facilitating techno-ecological 127 decision-making that strives to address diverse stakeholder priorities. While a cost-benefit analysis 128 would be essential for the ultimate selection of a treatment technology, the associated analysis is 129 beyond the scope of the present work, which is focused on framework development. Microbiologically 130 mediated biofiltration technologies are presented as obvious and effective examples of underutilized 131 green technology opportunities in the drinking water industry. They are used to demonstrate that there 132 is a wide spectrum of techno-ecological NBS - 'green technologies' - that are fit-for-purpose in the 133 treatment and distribution of safe drinking water. Finally, two case studies are briefly presented to 134 highlight the benefits of green technologies in drinking water treatment, the use and limitations of the 135 developed framework, and the influence of sociocultural factors on green technology preferences of 136 individuals, groups, or communities.

A framework for evaluating technology greenness. The most widely recognized 'green' technologies in
 the broader water industry are likely found in stormwater management and include low-impact

139 development practices such as vegetated rooftops, roadside plantings, absorbent gardens, and other 140 measures. They are designed to mimic natural hydrological processes and landscape features to reduce 141 stormwater flows and improve stormwater quality by filtration, adsorption, or other means before 142 discharging to surface and groundwater supplies (Gill et al. 2007). In contrast, reductions in energy 143 consumption and waste production are common green foci of wastewater treatment (Wu et al. 2015; 144 Neoh et al. 2016). Here, many of the 'green' technologies include biological treatment processes that 145 remove or neutralize pollutants or other target compounds, often to yield less toxic or non-toxic 146 materials at a lower cost than technologies that are not biologically mediated (Delgadillo-Mirquez et al. 147 2016). Membrane bioreactors are one such example; they combine biological, secondary, and tertiary 148 wastewater treatment in one unit, thereby reducing carbon footprint relative to more conventional 149 processes (Smith et al. 2012; Neoh et al. 2016). Groundwater treatment at contaminated sites 150 increasingly involves the implementation of green in situ bioremediation technologies to reduce energy 151 costs and largely eliminate excavation and incineration costs common to ex situ 'pump and treat' 152 approaches (Haritash & Kaushik 2009; Wang & Chen 2009).

153 While the use of the term 'green technology' is less common in the drinking water industry, its broader 154 emergence is inevitable. For example, nature-based coagulants produced from renewable resources 155 (Teixeira et al. 2017) are regularly referred to as 'green' technologies. Reductions in energy consumption 156 and waste production are already common goals in the industry, and biological filtration processes that 157 'work for free' are referred to as either 'natural' or 'green' treatment technologies – their use in drinking 158 water treatment plants (DWTPs) is increasingly described as 'by design' rather than de facto (Basu et al. 159 2015; Brown et al. 2015; Petrescu-Mag et al. 2016; Kirisits et al. 2019). At the regional landscape scale, 160 sophisticated watershed management techniques focused on maintaining high-quality source water are 161 often relied upon to avoid the construction of costly filtration plants and are being increasingly 162 implemented for the mitigation of climate change-exacerbated landscape disturbances such as severe

wildfires (Emelko et al. 2011; Cristan et al. 2016; NAS 2018; Robinne et al. 2019). Indeed, interest in the
promise of 'green tech' is growing across the water industry and to the general public who increasingly
value it, and contribute to promoting it, as evident from public acceptance and willingness-to-pay for
green tech implementation for water resource management and treatment (Newburn & Alberini 2016;
Brent et al. 2017; The Water Institute 2017).

168 As highlighted by the examples above, green technologies in the field of drinking water supply and 169 treatment have been most frequently described as 'green' based on three key attributes or factors that are broadly associated with reducing environmental impacts: (1) nature- or natural resource-based 170 171 origin (Spatari et al. 2011; Keeley et al. 2013; Liu et al. 2017), (2) relatively low-energy consumption (Wu 172 et al. 2015; Ngo et al. 2016), and (3) relatively low waste production (Neoh et al. 2016; Ngo et al. 2016). 173 Physical footprint is further proposed as a fourth key factor that contributes to technology greenness in 174 the water supply and treatment field. The physical footprint of watershed management activities such 175 as forest harvesting, DWTP construction, and associated residuals management infrastructure has the 176 potential to adversely impact human health and ecosystems through fossil fuel emissions, destruction of 177 sensitive habitat, habitat fragmentation, and biodiversity decline, to name a few. The impacts of physical 178 footprint are generally understood to be linked to environmental impacts because they initiate a chain 179 reaction of environmental impacts that can be broadly characterized as human health and ecosystem 180 damage footprints. Thus, physical infrastructure footprints must be included in any evaluation of 181 greenness to reflect these cumulative environmental impacts. Accordingly, a framework for 182 characterizing water industry technology greenness based on four main key technology attributes is 183 proposed. As illustrated in Figure 2, they are (1) natural-resource basis, (2) energy consumption, (3) 184 waste production, and (4) footprint. Various fit-for-purpose drinking water treatment technology 185 examples considered for application in the same environmental setting are presented in Figure 2 to 186 demonstrate how the framework developed herein might be used. A more detailed description of the

technology attributes that contribute to greenness follows, and opportunities to link the framework to
 more comprehensive evaluations of trade-offs between technological NBS in the water sector are briefly
 discussed.

190 [Figure 2]

191 Natural resource-based technology incorporates renewable or non-depletable materials that are either 192 sourced from the surrounding environment or utilize natural processes to achieve treatment. Several of 193 these technologies, such as biofiltration and solar disinfection, are intrinsically passive and do not 194 require additional chemical inputs (McGuigan et al. 2012; Basu et al. 2015), which in turn contributes to 195 their low-energy consumption and waste production. Some natural coagulants, such as moringa seeds, 196 have been described as 'green' (Teixeira et al. 2017); however, despite being natural resource-based, 197 coagulants that are not sourced from the surrounding environment must still be transported to 198 treatment facilities for use. As such, proximity of the material source and site of use should be 199 considered, and those materials whose haulage has significant environmental costs should not be 200 considered green in this context. Beyond drinking water treatment, natural resource-based technologies 201 also include approaches such as forested watershed management practices that are applied for 202 managing drinking water source quality (i.e., SWP technologies) (Cristan et al. 2016; NAS 2018). 203 Energy consumption is often cited as an important and highly valued aspect of technology greenness

(Bolla et al. 2011; Ngo et al. 2016; Barcelos et al. 2018). Energy-efficient technologies often offer a cobenefit of reduced long-term operational costs; this is mainly attributed to their passive nature and
dependence on non-energy-intensive processes (e.g., naturally occurring biological activity) to achieve
treatment goals (Wu et al. 2015; Neoh et al. 2016). Processes that require high energy inputs to operate,
such as ozonation and UV disinfection, are relatively less green. High energy expenditures can also result
from water conveyance through pumping. Therefore, the elevation of a DWTP site is an important

210 design consideration and can impact overall energy consumption (Randtke & Horsley 2012). For 211 example, the need for pumping may be reduced if plant configuration follows natural topography. Even 212 less major design choices, such as the selection of flocculator type, can also result in energy 213 consumption changes. Although they offer substantively more operational control, mechanical 214 flocculators require higher energy inputs compared to hydraulic mixers and are therefore less green in 215 this respect (Crittenden et al. 2012). These types of decisions underscore the trade-offs that must be 216 clearly articulated and considered in the selection and design of water treatment technologies. 217 Waste produced during water treatment has the potential to cause adverse environmental impacts as a 218 result of its quantity and/or toxicity; thus, it is an important contributor to technology greenness. 219 Treatment processes that produce large amounts of waste products, such as coagulation (i.e., sludge) 220 and membrane technologies (i.e., brine, backwash, and residuals), can be generally considered as less 221 green. However, some chemical additions may reduce waste production, such as the addition of 222 polymers to alum or ferric chloride coagulants (Randtke & Horsley 2012). Membrane technologies 223 produce wastes in the form of backwash and cleaning-in-place residuals. Cleaning-in-place can increase 224 both waste quantity and toxicity because it involves chemicals such as hypochlorite, citric acid, and 225 caustic soda (Randtke & Horsley 2012). Additionally, waste in the form of emissions implies that air 226 stripping processes may be relatively less green due to exhaust fume emissions (Randtke & Horsley 227 2012).

The physical footprint of infrastructure contributes to water treatment technology greenness because it can also readily result in adverse environmental impacts. Processes that require a large footprint, such as horizontal flow basins and slow sand filters, will tend to be less green in this respect. Additional infrastructures – such as residuals management plants, chemical storage, and pumping infrastructure – also increase footprint. This highlights the interplay between green attributes; for example, high wasteproducing processes typically require the construction of a residuals management plant, which increases 234 the footprint and contributes to the reduction in greenness of the process. Additionally, chemically-235 assisted processes require chemical storage infrastructure on-site, which increases footprint and can 236 also increase energy consumption through the need for heating, ventilation, and air conditioning (HVAC) 237 systems and hydraulic lifting (Randtke & Horsley 2012). While this discussion generally suggests that 238 larger environmental footprints are more disruptive, infrastructure footprints cannot be considered in a 239 vacuum as they are intrinsically tied to the environmental setting in which they are to be applied. Thus, 240 the inclusion of physical footprint in an evaluation of technology greenness necessarily requires 241 consideration of the impacts to both the biophysical and human environments within that setting. For 242 example, the optimal location and extent of DWTP footprint is dependent on several factors including 243 distance from source water, elevation, and available space. Other environmental factors such as the 244 presence of important fish habitat in a natural waterway receiving discharge from the waste stream of 245 the DWTP also require consideration, however; as a result, limiting waste production may be ultimately 246 prioritized in this setting to limit adverse impacts to biodiversity in the natural waterway.

247 The four attributes of water industry technology that impact greenness (natural-resource basis, energy 248 consumption, waste production, and footprint) are closely linked and must be considered relative to 249 both the specific environmental settings in which they are applied and the other technologies to which 250 they are being compared. Thus, life cycles and supply chains should also be considered. Life cycle 251 analysis (LCA) involves the evaluation of the environmental impacts of a product, process, or service 252 over all of its stages of the life cycle; thus, it includes the environmental impacts of all relevant life cycle 253 aspects, which may include raw material extraction or processing, manufacturing, distribution, use, 254 regeneration, recycling, and final disposal (Ayres 1995). For example, processes using activated carbon 255 materials are generally less green since they require high energy inputs during the production and 256 regeneration stages. Rigorous LCA will thus reflect several aspects of supply chain analysis including how 257 risks can be reduced by bypassing certain suppliers and/or processes and reduce unnecessary

258 inventories. Shipment of materials over long distances is a simple example of the importance of supply 259 chains in evaluating technology greenness because of associated indirect increases in energy 260 consumption and waste production via increased emissions. Co-benefits associated with certain 261 technologies should also be considered. For example, some of the waste products from water treatment 262 processes may be reused for various purposes such as land application, composting, cement 263 manufacturing, and road subgrade (Randtke & Horsley 2012; Márquez et al. 2019). While it could be 264 argued that an absolute, quantitative index could be developed to measure the 'greenness' of a given 265 technology, this is not proposed herein because such a metric would require assumptions regarding 266 both the relative value of the 'greenness' attributes and the impacts of the technology on the biophysical and human environments relevant to the setting where it is to be applied. 267 268 It is at this point of greenness evaluation that the inter-connectedness of the choice between 269 technology options and their relative greenness becomes iterative and complicated. The evaluation 270 becomes iterative because of the chain reaction of environmental impacts that is initiated by these 271 decisions, as demonstrated above. Approaches for characterizing these impacts are available, however. 272 For example, they can be broadly characterized as human health and ecosystem damage footprints. 273 Comprehensive damage assessments and LCAs have recently been applied to harmonized resource-274 based footprints (i.e., energy, material, land, and water) to demonstrate that resource footprints 275 provide good proxies for environmental (i.e., human health and ecosystem) damage (Steinmann et al. 276 2017). Evaluations of technology greenness and ultimate implementation are also complicated, 277 however, because of trade-offs between techno-ecological services. For example, the fail-safe provision 278 of safe water may conflict with other techno-ecological services such as waste minimization. Conflicts 279 may result from divergent sociocultural preferences among individuals, communities, or other 280 stakeholders that are differently impacted by the techno-ecological services that can be provided by the 281 technology that is ultimately implemented (King et al. 2015). Frameworks to characterize trade-offs in

ecosystem services that reflect biophysical constraints and divergent values have been developed
(Cavender-Bares et al. 2015; King et al. 2015) and offer further opportunities to advance on the
promises of techno-ecological NBS in the water sector. While the explicit recognition of differences
among stakeholder values and preferences is integral to ensuring that techno-ecological NBS achieve
intended impacts, strategies for navigating such conflicts and evaluating the implications of trade-offs
impacting biophysical and human environments are beyond the scope of the present work.

288 To illustrate the utility of the greenness framework shown in Figure 2 for identifying, naming, and 289 describing the 'green' attributes of treatment technology that may be valued in certain situations, the 290 relatively simple selection of fit-for-purpose surface water treatment systems can be explored in two 291 distinct environmental settings: remote and urban. Notably, technology typologies are excluded from 292 the discussion; only key green attributes are discussed. A remote community may be challenged by 293 accessibility and unreliable supply chains, unreliable power supplies, and institutional memory and staff 294 retention (Hall 2018; Chattha 2020) – these challenges may not be as significant in an urban 295 environment. In contrast, while available space and footprint may not be an issue in a rural or remote 296 area, an urban community may be constrained by the available space. Despite these differences, both 297 communities are likely challenged by competing demands between finances and treatment capacity, 298 resilience, and redundancy, as well as operational burden. The remote community may, therefore, value 299 technologies that are natural resource-based and easy to maintain, and reduce energy consumption and 300 waste production as compared to those that reduce physical footprint. Natural resource-based 301 technologies would address accessibility challenges as fewer components and chemicals would need to 302 be sourced externally for operation, maintenance, and repairs, thereby reducing often high 303 transportation costs. Additionally, natural resource-based technologies tend to be passive and therefore 304 typically have lower energy demands and are associated with lower operational burdens and capacities 305 than non-passive technologies. Thus, natural resource-based technologies may help to mitigate the

306 challenges presented by power supply reliability, institutional memory and staff retention, finances, and
307 operational burden and capacity. Technologies that generate relatively less waste might be prioritized,
308 as the management of waste and hazardous substances add to both the operational burden and
309 technical capacity requirements. Conversely, footprint may not be prioritized, as the small population
310 and remote location imply lower water demand and more available space, respectively.

311 In contrast, an urban centre may value footprint, energy conservation, and low waste production as 312 important green factors, with less importance placed on the passive quality of natural resource-based 313 technologies. Technologies designed to reduce the footprint may minimize the environmental impact 314 caused by the extent of infrastructure required to meet high production demands. Competition for 315 financial resources may encourage a focus on reducing energy consumption, as this often represents a 316 large fraction of a water utility's operational costs (Crittenden et al. 2012). Additionally, limiting waste 317 production reduces the need for additional waste management infrastructure, further reducing 318 footprint and energy demands.

319 It should be underscored that the framework illustrated in Figure 2 constitutes a simple organizational 320 structure to identify, name, and describe the 'green' attributes of the broad range of technologies -321 including natural capital – relevant to the water industry to enable stakeholders to clearly and 322 accurately communicate the technical and engineering aspects of technology approaches that best align 323 with their individual or community sociocultural values, beliefs, and attitudes. The framework 324 necessarily requires consideration of the environmental setting in which the technology is to be applied 325 and assessment of the technology's life cycle within that setting to provide structured discussion 326 regarding techno-ecological trade-offs as a first step in facilitating techno-ecological decision-making 327 that strives to address diverse stakeholder priorities.

328 **Biofiltration as a key example of green technology for drinking water treatment.** While minimizing 329 waste production and energy consumption are somewhat obvious strategies for increasing the 330 greenness of drinking water treatment and distribution approaches, the incorporation of natural 331 resource-based green technologies as techno-ecological NBS is at the precipice of a revolution in the 332 water industry. Biofiltration processes are arguably the most obvious and effective examples of 333 underutilized green technology opportunities in the drinking water industry. They have not yet 334 experienced as much uptake as conventional treatment technologies in some regions due to concerns 335 regarding the health risk attributable to microbially mediated treatment, difficulties in operation, and 336 unlikely regulatory approvals (Brown et al. 2015). However, such concerns are misplaced (Brown et al. 337 2015; Kirisits et al. 2019). Biofiltration technologies differ from conventional filtration in that biological 338 activity is promoted and maintained within and on filter media – in-built vessels or naturally in the 339 subsurface – to remove suspended particles (including pathogens) and dissolved organics from the 340 water phase (Basu et al. 2015; Kirisits et al. 2019). Biofiltration technologies harness natural microbial 341 processes, do not generally require additional energy inputs, and do not typically produce significant 342 waste relative to other treatment processes designed to achieve the same objectives (Fowler & Smets 343 2017). However, when biofilters are operated passively at low flow rates, they often require large 344 footprints to ensure targeted yields of drinking water. Notably, there are many types of biofiltration 345 technologies; although they can also be considered green, they fall along a spectrum of greenness. 346 Some common types of biofiltration used in drinking water treatment include:

Classical biofiltration: biofiltration in an otherwise conventional DWTP (preceded by
 coagulation/flocculation/sedimentation);

• Classical direct biofiltration: biofiltration preceded by coagulation/flocculation;

Biofiltration with pre-ozonation: biofiltration, either classical or classical direct, preceded by
 ozonation;

• Slow sand filtration (SSF): passively operated filtration through sand media; and

• Riverbank filtration (RBF): induced surface water infiltration to bankside abstraction wells.

354 [Figure 3]

355 The greener biofiltration technologies in this spectrum are generally operated passively and take 356 advantage of natural processes in the surrounding environment to achieve treatment goals; such 357 technologies include SSF and RBF. Combinations of biofiltration processes – such as roughing filters, 358 managed aquifer recharge and storage, and reservoir storage – may provide additional treatment and 359 can increase operational control, but increase footprint and energy requirements. As well, processes 360 such as classical biofiltration indirectly contribute to waste production due to pre-treatment by 361 coagulation and clarification processes prior to filtration; it is also more energy-intensive because it is 362 not passively operated and requires backwashing to remove accumulated solids. Biofiltration 363 technologies preceded by ozonation are especially effective in removing organics, but less green 364 because of the energy-intensive nature of ozonation.

While not reflected in Figure 3, filter media are also an important factor contributing to biofiltration technology greenness. Biofiltration technologies employing a form of granular-activated carbon are intrinsically less green because of the high energy required to manufacture adsorptive media. The physical and chemical manufacturing processes involve carbonization, or conversion of the raw material to a char, and activation or oxidation to develop the internal pore structure – temperatures of 800– 900°C are needed for the activation process (Edzwald 2011). Readily available filtration media, such as anthracite coal and sand, are more green options, especially when they can be locally sourced.

Greenness assessment of drinking water treatment systems. In addition to the relative greenness
 ranking of biofiltration technologies, common drinking water treatment systems may also be relatively
 ranked according to their greenness. Figure 4 presents a relative ranking of common drinking water

375 treatment system configurations; however, actual evaluation of technology greenness is case-specific, as 376 discussed previously. Generally, treatment systems using biofiltration, such as classical biofiltration, SSF, 377 or RBF (all followed by chlorine-based disinfection), are among the greenest treatment approaches 378 relative to conventional (i.e., coagulation, flocculation, sedimentation, non-biological filtration, and 379 chlorine-based disinfection) treatment because they are natural resource-based, require relatively lower 380 energy inputs, and produce relatively less waste. It is important to note, however, that some key trade-381 offs exist between less energy-intensive technologies and operational control. Although energy-efficient 382 technologies are generally more green, they often do not offer as much operational control as more 383 conventional treatment systems because of factors such as the lack of design and operational (i.e., 384 typically mechanical) controls over system components such as flow rates or microbially mediated 385 degradation of contaminants. As such, some green technologies are less able to respond to sudden 386 changes in source water quality, which can potentially compromise public health protection - this issue 387 requires further investigation to ensure resilient treatment, especially in environments vulnerable to 388 climate change-exacerbated landscape disturbances such as wildfires (Emelko et al. 2011; Stone et al. 389 2011).

390 [Figure 4]

391 Applying the green technology framework to case studies. Two DWTP design case studies presented 392 below highlight benefits of green technologies in drinking water treatment, use and limitations of the 393 developed framework, and influence of sociocultural factors on the green technology preferences of 394 individuals, groups, or communities.

395 CASE 1) Biofiltration to treat high ammonia groundwater for a small system (EPA 2014)
396 The implementation of an innovative biofiltration system for a small drinking water system in Iowa
397 highlights the promise of green tech to achieve a technologically fit-for-purpose treatment design. The

EPA conducted pilot-scale and full-scale studies for implementation of a novel biofiltration treatment technology in Palo, Iowa, which did not have centralized water treatment prior to 2008. Palo is a small town of just over 1,000 people, with limited technical capacity as the utility relies solely on one treatment plant operator who is also responsible for other municipal operations such as snow plowing and landscaping. Source water for the DWTP is groundwater characterized by high ammonia and iron concentrations and is low in dissolved oxygen.

404 Breakpoint chlorination is a common treatment option to address high ammonia concentrations 405 (Edzwald 2011). However, the chlorine dose required to adequately oxidize ammonia and nitrogen 406 species would be excessive for a small system. As an innovative alternative to breakpoint chlorination to 407 treat ammonia-rich groundwater, the EPA designed a novel biofiltration treatment system. The 408 treatment system, patented by the EPA, consists of aeration contactors, blowers, and dual media filters, 409 with added chemical feeds of phosphate, chlorine, and sodium hydroxide. An aeration contactor was 410 needed to ensure sufficient oxygen required for nitrification, as the groundwater source was low in 411 dissolved oxygen. The main goal of the treatment plant is to remove ammonia and iron, which was 412 consistently achieved in both the pilot- and full-scale systems.

413 An evaluation of all four green attributes discussed herein was not reported, as this is often not possible 414 due to limited time or resources. Nonetheless, the biofiltration system may be described as green 415 because it is natural resource-based and requires substantially less chemical input compared to 416 breakpoint chlorination, the alternative treatment option. Because of these green aspects, the 417 biofiltration system is operationally less demanding and thus also matches the operational (i.e., operator 418 training and treatment processes supervision) capacity of a smaller system. Most importantly, the 419 treatment system produces drinking water that consistently meets the regulatory targets set for 420 contaminants of concern, thereby ensuring a fit-for-purpose treatment design for the protection of 421 public health.

422 CASE 2) RBF for pre-treatment of municipally and industrially impacted surface water in 423 Louisville, Kentucky (Ball 2012)

424 Louisville Water Company in Louisville, Kentucky, implemented an RBF system as pre-treatment to 425 address concerns of microbial contamination possibly not addressed by the city's conventional 426 treatment system. The city is reliant upon the municipally and industrially impacted Ohio River for 427 drinking water. The Ohio River is consistently ranked as the most polluted in the United States, with an 428 estimated 30 million pounds of toxic chemicals illegally dumped into its waters each year (Kuhlman 429 2019). Louisville is a relatively large, established city and thus has limited available space. The Louisville 430 Water Company served a population of 764,769 in 2019 (EWG 2019) and has a high level of technical 431 capacity.

432 To address microbial contaminant concerns, the city launched a project to investigate the

433 implementation of an RBF system on the Ohio River. The RBF system would also address challenges with 434 water main breaks in the distribution system due to large variations in water temperature. As part of the 435 project, the city investigated drilling options for the tunnel and wells. Ultimately, the city decided on a 436 completely underground RBF system that includes a tunnel and collector wells. Although an above-437 ground system would have been much easier and less expensive to construct, the public did not want 438 any above-ground structures to impact the aesthetic value of the Ohio River. Additionally, while vertical 439 wells would be much easier to maintain than collector wells, collector wells were chosen due to the 440 possibility for construction complications with vertical wells. Additionally, the city's high technical 441 capacity was able to address the increased maintenance requirements associated with collector wells. 442 Similar to the previous case study in Palo, information detailing the green attributes of the treatment 443 process was not reported. Nonetheless, it is clear that Louisville's RBF system is relatively natural 444 resource-based, as it utilizes the natural subsurface to eliminate taste and odour compounds, provides

an additional barrier for waterborne pathogen removal, and creates a stable water temperature that
results in fewer main breaks in the distribution system. Despite this, the physical footprint of the RBF
system is relatively large due to the footprint needed during the construction of an underground
system.

449 This case study highlights the importance of discussing stakeholder priorities accurately and

450 transparently to achieve fit-for-purpose and socioculturally appropriate treatment design. Louisville

451 Water Company considered stakeholder priorities after ensuring treatment design met regulatory

452 requirements to uphold the protection of public health. While the public held sociocultural values that

453 aligned with preserving the aesthetic quality of the Ohio River, the Louisville Water Company sought to

454 minimize risk of construction complications. These needs were ultimately met by the selection of an

455 underground RBF system equipped with collector wells.

456 **CONCLUSIONS**

457 The main conclusions of the analysis presented herein are briefly summarized below. They are:

458 1. While the concept of green technology is widely recognized, its meaning varies considerably. In

459 the water industry, green technology can be described by four main attributes: natural

460 resource-basis, energy consumption, waste production, and footprint.

461 2. The greenness of a technology can be evaluated with respect to each of the above-mentioned

462 attributes and is therefore relative to both the environmental setting and the other technologies

to which it is being compared.

The paramount objective of treatment is public health protection and thus technologies must be
fit-for-purpose with respect to their use and meet regulated performance targets regardless of
their greenness.

467 4. Operational control is often reduced as the greenness of a technology is increased.

- 468 5. In the water sector, environmental setting (i.e., location-specific factors including hydroclimate,
- sensitive habitat(s), water quality, temperature, etc.) is a critical consideration that can limit the
 practical application of some technologies.
- 471 6. Biofiltration processes are arguably the most obvious and effective examples of underutilized
- 472 green technology opportunities in the drinking water industry. These technologies can be
- 473 differentiated along a spectrum of greenness.
- Prioritization of the factors contributing to technology greenness varies based on sociocultural
 considerations of individuals, groups, and communities, as identified based on their collective
 knowledge, values, attitudes, beliefs, feelings, and behaviours.
- 477 8. The framework developed herein enables an accurate and transparent description of the 'green'
- 478 attributes of technology including green infrastructure used in the water industry. It
- differentiates technology 'greenness' by relatively examining key attributes that may cause
- 480 environmental impacts across the technology's life cycle through the lens of the environmental
- 481 setting in which it is applied. It can contribute to the development of more comprehensive
- 482 techno-ecological NBS by providing a clear and accurate description of the 'green' attributes of
- 483 technology options for the water industry, as well as a framework for their relative comparison,
- 484 thereby facilitating techno-ecological decision-making that strives to address diverse
- 485 stakeholder priorities.

486



Figure 1: Low-cost cascade aeration system that enhances the air-water transfer of atmospheric gases
(e.g. oxygen, nitrogen) and volatile organic compounds. The term 'green technology' commonly
invokes images of such technologies; however, green technologies span a broad spectrum of
treatment typologies.

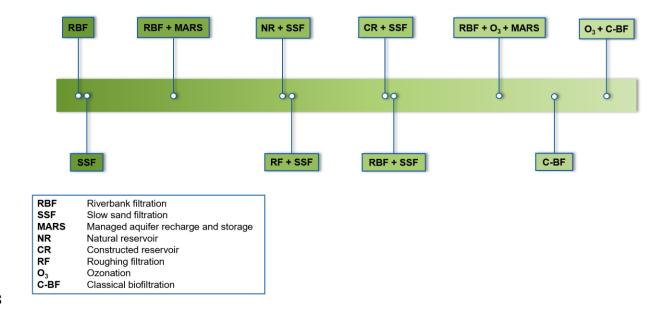


*Technologies are assumed to be fit-for-purpose. Whether or not technologies are green is not absolute; they are more or less green relative to one another.

493

494	Figure 2: Framework for the evaluation of green attributes of water supply, treatment, and distribution
495	technologies. (Photo credits bottom row from left to right: Humboldt Bay Municipal Water
496	District; Reprinted from Nalwanga et al. (2014), with permission from Elsevier; Mount Carmel

497 Ltd; DVGW, Water Technology Center, Karlsruhe).



498

499 Figure 3: Greenness spectrum of biofiltration technologies for drinking water treatment.

Treatment system	Natural resources-basis	Energy consumption	Waste production	Footprint		
High-rate clarification plant*	Low	High	High	High	Less green	More operational control
Dissolved air flotation plant*	Low	High	High	High	Less green	More operational control
Direct filtration plant*	Moderate	Moderate to high	High	Moderate	Moderately green	Moderate operational control
Classical biofiltration plant*	Moderate	Moderate to high	Moderate to high	Moderate	Moderately green	Moderate operational control
Slow sand filtration & chlorine disinfection	High	Low	Low	Low to moderate	More green	Less operational control
Riverbank filtration & chlorine disinfection	High	Low	Low	Low	More green	Less operational control

* plant refers to an otherwise conventional treatment setting

500

501 Figure 4: General greenness assessment of common drinking water treatment typologies.

502

503 **REFERENCES**

- 504 Aronson, J. 2011 Sustainability science demands that we define our terms across diverse disciplines.
- 505 *Landscape Ecology 26*(4), 457–460. https://doi.org/10.1007/s10980-011-9586-2. AWWA 2020
- 506 State of the Water Industry Report (pp. 1–41, Rep.).
- 507 Ayres, R. U. 1995 Life cycle analysis: a critique. *Resources, Conservation and Recycling* 14(3–4), 199–
- 508 223. https://doi.org/10. 1016/0921-3449(95)00017-D. Bakshi, B. R., Ziv, G. & Lepech, M. D. 2015
- 509 Techno-ecological synergy: a framework for sustainable engineering. Environmental Science &

510 *Technology 49*(3), 1752–1760. https://doi.org/10.1021/es5041442.

- 511 Ball, K. 2012 Louisville launches innovative RBF project. Journal American Water Works Association
- 512 *104*(3), 60–67. https:// doi-org.proxy.lib.uwaterloo.ca/10.5942/jawwa.2012.104.0043.
- 513 Barcelos, M. C., Lupki, F. B., Campolina, G. A., Nelson, D. L. & Molina, G. 2018 The colors of
- 514 biotechnology: general overview and developments of white, green and blue areas. *FEMS*

515 *Microbiology Letters 365*(21), fny239. https://doi.org/10.1093/ femsle/fny239.

- 516 Basu, O. D., Dhawan, S. & Black, K. 2015 Applications of biofiltration in drinking water treatment a
- 517 review. Journal of Chemical Technology & Biotechnology 91(3), 585–595. https://doi-
- 518 org.proxy.lib.uwaterloo.ca/10.1002/jctb.4860.
- 519 Bolla, R., Davoli, F., Bruschi, R., Christensen, K., Cucchietti, F. & Singh, S. 2011 The potential impact of
- 520 green technologies in next-generation wireline networks: is there room for energy saving
- 521 optimization? *IEEE Communications Magazine 49*(8), 80–86.
- 522 https://doi.org/10.1109/MCOM.2011.5978419.
- 523 Borchardt, M. A., Spencer, S. K., Kieke, B. A., Lambertini, E. & Loge, F. J. 2012 Viruses in nondisinfected
- 524 drinking water from municipal wells and community incidence of acute gastrointestinal illness.
- 525 *Environmental Health Perspectives 120*(9), 1272–1279. https://doi.org/10.1289/ehp.1104499.
- 526 Brent, D. A., Gangadharan, L., Lassiter, A., Leroux, A. & Raschky, P. A. 2017 Valuing environmental

- 527 services provided by local stormwater management. *Water Resources Research 53*(6), 4907–
 528 4921. https://doi.org/10.1002/2016WR019776.
- 529 Brown, J., Summers, R. S., Lechevallier, M., Collins, H., Roberson, J. A., Hubbs, S. & Dickenson, E. 2015
- 530 Biological drinking water treatment? Naturally. *Journal American Water Works Association*
- 531 *107*(12), 20–30. https://doi.org/10.5942/jawwa. 2015.107.0183.
- Cavender-Bares, J., Polasky, S., King, E. & Balvanera, P. 2015 A sustainability framework for assessing
 trade-offs in ecosystem services. *Ecology and Society 20*(1). https://doi.org/10.5751/es-06917 200117.
- 535 Chattha, S. 2020 Streamlining water treatment solutions for indigenous and non-urban communities.
 536 Water Canada 20(3), 10–13.
- 537 Chenoweth, J., Anderson, A. R., Kumar, P., Hunt, W., Chimbwandira, S. J. & Moore, T. L. 2018 The
- interrelationship of green infrastructure and natural capital. *Land Use Policy 75*, 137–144.
 https://doi.org/10.1016/j.landusepol.2018.03.021.
- 540 Cohen-Shacham, E., Walters, E., Janzen, C. & Maginnis, S. 2016 *Nature-Based Solutions to Address Global*

541 Societal Challenges. IUCN, Gland. https://doi.org/10.2305/IUCN.CH.2016.13.en.

- 542 Cristan, R., Aust, W. M., Bolding, M. C., Barrett, S. M., Munsell, J. F. & Schilling, E. 2016 Effectiveness of
- 543 forestry best management practices in the United States: literature review. *Forest Ecology and*544 *Management 360*, 133–151. https://doi. org/10.1016/j.foreco.2015.10.025.
- 545 Crittenden, J. C., Trussell, R. R., Hand, D. W., Howe, K. J. & Tchobanoglous, G. 2012 *MWH's Water*

546 Treatment: Principles and Design. John Wiley & Sons, Hoboken, NJ.

- 547 Delgadillo-Mirquez, L., Lopes, F., Taidi, B. & Pareau, D. 2016 Nitrogen and phosphate removal from
- 548 wastewater with a mixed microalgae and bacteria culture. *Biotechnology Reports* 11, 18–26.
- 549 https://doi.org/10.1016/j.btre.2016.04.003.

- Edzwald, J. K. 2011 *Water Quality & Treatment: A Handbook on Drinking Water*, 6th edn. McGraw-Hill,
 Denver.
- 552 Emelko, M. B., Silins, U., Bladon, K. D. & Stone, M. 2011 Implications of land disturbance on drinking
- 553 water treatability in a changing climate: demonstrating the need for 'source water supply and
- 554 protection' strategies. *Water Research* 45(2), 461–472.
- 555 https://doi.org/10.1016/j.watres.2010.08.051.
- 556 Emelko, M., Schmidt, P. & Borchardt, M. 2019 Confirming the need for virus disinfection in municipal
- 557 subsurface drinking water supplies. *Water Research* 157, 356–364.
- 558 https://doi.org/10.1016/j.watres.2019.03.057.
- 559 EPA 1998 Regulatory Impact Analysis for the Interim Enhanced Surface Water Treatment Rule. Available
- from: https://nepis. epa.gov/Exe/ZyPDF.cgi/910209DL.PDF?Dockey=910209DL.PDF. Accessed 1
 May 2021.
- 562 EPA 2014 Engineering Design and Operation Report: Biological Treatment Process for the Removal of
- 563 Ammonia from a Small Drinking Water System in Iowa: Pilot to Full-Scale (EPA/600/R-14/336).
- 564 Office of Research and Development. Available from:
- 565 https://cfpub.epa.gov/si/si_public_record_report.cfm?Lab=NRMRL&dirEntryId=302694.
- 566 Accessed 1 May 2021.
- 567 Ernst, C., Gullick, R. & Nixon, K. 2004 Conserving forests to protect water. *Opflow 30*(5), 1–7.
- 568 https://doi.org/10.1002/j.1551- 8701.2004.tb01752.x.
- 569 EWG 2019 EWG's Tap Water Database: What's in Your Drinking Water? Available from:
- 570 https://www.ewg.org/tapwater/ search-results.php?stab=KY&searchtype=largesys (accessed 6
 571 May 2021).
- 572 Fowler, S. J. & Smets, B. F. 2017 Microbial biotechnologies for potable water production. *Microbial*
- 573 Biotechnology 10(5), 1094–1097. https://dx.doi.org/10.1111%2F1751-7915.12837.

574	Gill, S. E., Handley, J. F., Ennos, A. R. & Pauleit, S. 2007 Adapting cities for climate change: the role of the
575	green infrastructure. Built Environment 33(1). 115–133. https://doi.org/10.2148/benv.33.1.115.

576 Hall, N. L. 2018 Australian Indigenous remote communities and water, sanitation, and hygiene. *Water*

577 Source. Available from: https://watersource.awa.asn.au/publications/technical-

- 578 papers/australian-indigenous-remote-communities-and-watersanitation-and-hygiene/.
- 579 Accessed 1 May 2021.
- 580 Haritash, A. K. & Kaushik, C. P. 2009 Biodegradation aspects of polycyclic aromatic hydrocarbons (PAHs):

581 a review. Journal of Hazardous Materials 169(1–3), 1–15.

582 https://doi.org/10.1016/j.jhazmat.2009.03.137.

583 Keeley, M., Koburger, A., Dolowitz, D. P., Medearis, D., Nickel, D. & Shuster, W. 2013 Perspectives on the

584 use of green infrastructure for stormwater management in Cleveland and Milwaukee.

585 *Environmental Management 51*(6), 1093–1108. https://doi.org/10.1007/s00267-013-0032-x.

586 King, E., Cavender-Bares, J., Balvanera, P., Mwampamba, T. H. & Polasky, S. 2015 Trade-offs in

587 ecosystem services and varying stakeholder preferences: evaluating conflicts, obstacles, and

588 opportunities. *Ecology and Society 20*(3). https://doi.org/10. 5751/es-07822-200325.

589 Kirisits, M. J., Emelko, M. B. & Pinto, A. J. 2019 Applying biotechnology for drinking water biofiltration:

advancing science and practice. *Current Opinion in Biotechnology* 57, 197–204.

591 https://doi.org/10.1016/j.copbio.2019.05.009.

592 Kuhlman, M. 2019 Kentuckians defend Ohio River quality standards. *Public News Service*. Available from:

- 593 https://www.publicnewsservice.org/2019-02-11/water/kentuckians-defend-ohio-river-quality594 standards/a65493-1. Accessed 1 May 2021.
- Liu, Y., Engel, B. A., Flanagan, D. C., Gitau, M. W., McMillan, S. K. & Chaubey, I. 2017 A review on

596 effectiveness of best management practices in improving hydrology and water quality: needs

- 597 and opportunities. *Science of the Total Environment* 601–602, 580–593.
- 598 https://doi.org/10.1016/j.scitotenv.2017.05.212.
- 599 Márquez, A. J., Filho, P. C., Rutkowski, E. W. & Isaac, R. D. 2019 Landfill mining as a strategic tool
- 600 towards global sustainable development. *Journal of Cleaner Production 226*, 1102–1115.
- 601 https://doi.org/10.1016/j.jclepro.2019.04.057.
- 602 McGuigan, K. G., Conroy, R. M., Mosler, H., Preez, M. D., Ubomba-Jaswa, E. & Fernandez-Ibañez, P. 2012
- Solar water disinfection (SODIS): a review from bench-top to roof-top. *Journal of Hazardous Materials 235–236*, 29–46. https://doi. org/10.1016/j.jhazmat.2012.07.053.
- 605 McLain, R., Poe, M., Hurley, P. T., Lecompte-Mastenbrook, J. & Emery, M. R. 2012 Producing edible
- 606 landscapes in Seattle's urban forest. *Urban Forestry & Urban Greening 11*(2), 187–194.
- 607 https://doi.org/10.1016/j.ufug.2011.12.002.
- Mustafa, S., Estim, A., Tuzan, A. D., Ann, C. C., Seng, L. L. & Shaleh, S. R. 2019 Nature-based and
- 609 technology-based solutions for sustainable blue growth and climate change mitigation in marine
- 610 biodiversity hotspots. *Environmental Biotechnology* 15(1), 1–7.
- 611 https://doi.org/10.14799/ebms302.
- 612 Nalwanga, R., Quilty, B., Muyanja, C., Fernandez-Ibañez, P. & McGuigan, K. G. 2014 Evaluation of solar
- 613 disinfection of E. coli under Sub-Saharan field conditions using a 25 L borosilicate glass batch
- 614 reactor fitted with a compound parabolic collector. *Solar Energy 100*, 195–202.
- 615 NAS 2018 Review of the New York City Department of Environmental Protection Operations Support Tool
- 616 for Water Supply (Rep.). The National Academies Press, Washington, DC.
- 617 https://doi.org/10.17226/25218.
- 618 Neoh, C. H., Noor, Z. Z., Mutamim, N. S. & Lim, C. K. 2016 Green technology in wastewater treatment
- 619 technologies: integration of membrane bioreactor with various wastewater treatment systems.
- 620 Chemical Engineering Journal 283, 582–594. https://doi.org/10.1016/j.cej.2015.07.060.

- 621 Newburn, D. A. & Alberini, A. 2016 Household response to environmental incentives for rain garden
- adoption. *Water Resources Research 52*(2), 1345–1357.
- 623 https://doi.org/10.1002/2015WR018063.
- Ngo, H. H., Guo, W., Chen, Z., Surampalli, R. Y. & Zhang, T. C. 2016 Green Technologies for Sustainable
- 625 Water Management. American Society of Civil Engineers, Reston, VA.
- 626 https://doi.org/10.1061/9780784414422.
- 627 Oral, H. V., Carvalho, P., Gajewska, M., Ursino, N., Masi, F., Hullebusch, E. D., Kazak, J. K., Exposito, A.,
- 628 Cipolletta, G., Andersen, T. R., Finger, D. C., Simperler, L., Regelsberger, M., Rous, V., Radinja, M.,
- 629 Buttiglieri, G., Krzeminski, P., Rizzo, A., Dehghanian, K., Nikolova, M. & Zimmermann, M. 2020 A
- 630 review of nature-based solutions for urban water management in European circular cities: a
- 631 critical assessment based on case studies and literature. *Blue-Green Systems 2*(1), 112–136.
- 632 https://doi.org/10.2166/bgs.2020.932.
- 633 O'Sullivan, F., Mell, I. & Clement, S. 2020 Novel solutions or rebranded approaches: evaluating the use of
- 634 nature-based solutions (NBS) in Europe. *Frontiers in Sustainable Cities 2*, 1–15.
- 635 https://doi.org/10.3389/frsc.2020.572527.
- 636 Petrescu-Mag, R., Petrescu, D., Safirescu, O., Hetvary, M., Oroian, I. & Vâju, D. 2016 Developing public
- 637 policy options for access to drinking water in peripheral, disaster and polluted rural areas: a case
- 638 study on environment-friendly and conventional technologies. *Water 8*(3), 80.
- 639 http://dx.doi.org/10.3390/w8030080.
- 640 Purvis, B., Mao, Y. & Robinson, D. 2019 Three pillars of sustainability: in search of conceptual origins.
- 641 *Sustainability Science 14,* 681–695. https://doi.org/10.1007/s11625-018-0627-5.
- 642 Randtke, S. J. & Horsley, M. B. 2012 *Water Treatment Plant Design,* 5th edn. McGraw-Hill, New York.
- 643 Robinne, F., Bladon, K. D., Silins, U., Emelko, M. B., Flannigan, M. D., Parisien, M., Wang, X., Kienzie, S.

644	W. & Dupont, D. P. 2019 A regional-scale index for assessing the exposure of drinking-water
645	sources to wildfires. <i>Forests 10</i> (5), 384. https://doi. org/10.3390/f10050384.
646	Seddon, N., Smith, A., Smith, P., Key, I., Chausson, A., Girardin, C., House, J., Srivastava, S. & Turner, B.
647	2021 Getting the message right on nature-based solutions to climate change. Global Change
648	<i>Biology 27</i> (8), 1518–1546. https://doi.org/10. 1111/gcb.15513.
649	Smith, A. L., Stadler, L. B., Love, N. G., Skerlos, S. J. & Raskin, L. 2012 Perspectives on anaerobic
650	membrane bioreactor treatment of domestic wastewater: a critical review. Bioresource
651	Technology 122, 149–159. https://doi.org/10.1016/j. biortech.2012.04.055.
652	Spatari, S., Yu, Z. & Montalto, F. A. 2011 Life cycle implications of urban green infrastructure.
653	Environmental Pollution 159(8–9), 2174–2179. https://doi.org/10.1016/j.envpol.2011.01.015.
654	Steinmann, Z. J. N., Schipper, A. M., Hauck, M., Giljum, S., Wernet, G. & Huijbregts, M. A. J. 2017
655	Resource footprints are good proxies of environmental damage. Environmental Science &
656	Technology 51(11), 6360–6366. https://doi.org/10.1021/ acs.est.7b00698.
657	Stone, M., Emelko, M., Droppo, I. & Silins, U. 2011 Biostabilization and erodibility of cohesive sediment
658	deposits in wildfire-affected streams. Water Research 45(2), 521–534.
659	https://doi.org/10.1016/j.watres.2010.09.016.
660	Teixeira, M. R., Camacho, F. P., Sousa, V. S. & Bergamasco, R. 2017 Green technologies for cyanobacteria
661	and natural organic matter water treatment using natural based products. Journal of Cleaner
662	Production 162, 484–490. https://doi.org/10. 1016/j.jclepro.2017.06.004.
663	The Water Institute 2017 WaterTalk: Diane Dupont [Video]. YouTube. Available from:
664	https://www.youtube.com/watch?v=uqGuZRevbw. Accessed 1 May 2021.
665	Vos, J., Opdam, P., Steingrover, E. G. & Reijnen, R. 2007 Landscape ecology: The state of the science. In:
666	Key Topics in Landscape Ecology J. Wu & R. J. Hobbs (eds). Cambridge University Press,
667	Cambridge, pp. 271–287. https://doi.org/ 10.1017/CBO9780511618581.014.

- 668 Wang, J. & Chen, C. 2009 Biosorbents for heavy metals removal and their future. *Biotechnology*
- 669 *Advances 27*(2), 195–226. https://doi.org/10.1016/j.biotechadv.2008.11.002.
- 670 Wu, H., Zhang, J., Ngo, H. H., Guo, W., Hu, Z., Liang, S., Fan, J. & Liu, H. 2015 A review on the
- 671 sustainability of constructed wetlands for wastewater treatment: design and operation.
- 672 *Bioresource Technology 175*, 594–601. http://dx.doi.org/10. 1016/j.biortech.2014.10.068.