

1 Near-term forecasts of stream temperature
2 using process-guided deep learning and
3 data assimilation
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5 Jacob Zwart¹, Samantha Oliver¹, David Watkins¹, Jeffrey Sadler¹, Alison Appling¹, Hayley Corson-Dosch¹,
6 Xiaowei Jia², Vipin Kumar³, Jordan Read¹

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8 ¹U.S. Geological Survey, Water Mission Area

9 ²University of Pittsburgh, Department of Computer Science

10 ³University of Minnesota, Department of Computer Science and Engineering

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12 **Corresponding author:** Jacob Zwart (jzwart@usgs.gov)

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14 Abstract

15 Near-term forecasts of environmental outcomes can inform real-time decision making. Data assimilation
16 modeling techniques can be used for forecasts to leverage real-time data streams, where the difference
17 between model predictions and observations can be used to adjust the model to make better predictions
18 tomorrow. In this use case, we developed a process-guided deep learning and data assimilation approach
19 to make 7-day forecasts of daily maximum water temperature in the Delaware River Basin. Our modeling
20 system produced forecasts of daily maximum stream temperature with an average root mean squared
21 error (RMSE) from 1.2 to 1.6°C for 1-day lead time across all sites. The data assimilation algorithm
22 successfully adjusted the process-guided deep learning model states and marginally improved forecast
23 performance when compared to forecasts produced using the process-guided deep learning model alone
24 (7-13% lower RMSE with the data assimilation algorithm). Our model characterized forecast uncertainty
25 relatively well as 57-80% of observations were within 90% forecast confidence intervals across all sites
26 and lead times, and the uncertainty associated with our forecasts allow managers to anticipate probability
27 of exceedances of ecologically relevant thresholds and aid in decisions about releasing reservoir water
28 downstream. The flexibility of deep learning models to be applied to various prediction problems shows
29 promise for using these types of models to forecast many other important environmental variables and
30 aid in decision making.

31 Introduction

32 The intersection between human and environmental water needs can generate complex water allocation
33 decisions that must be made with imperfect information about the future (Vörösmarty et al. 2000).
34 Forecasts are predicted future conditions with associated uncertainty (Clark et al. 2001), and aquatic
35 forecasts are becoming critical management tools for balancing various uses of water to generate more
36 desirable outcomes (Viel et al. 2016; Turner et al. 2020). Parallel advances in predictive modeling, real-
37 time environmental observing systems, and decision analysis tools can be leveraged to create more useful
38 forecast products for water resources decisions.

39 To date, forecasts of aquatic systems have mostly relied on process-based models and statistical models.
40 These models have enabled managers and stakeholders to anticipate future changes in water
41 temperatures (Thomas et al. 2021), dissolved oxygen concentrations (Matos and de Sousa 1996; Abdi et
42 al. 2020), and streamflow (Block et al. 2009; Hansen et al., 2009; Turner et al. 2020). Process-based models
43 define relations between driver data and the variable being forecasted *a priori*, while statistical models
44 discover relations between driver data and the target variable during a model training phase assuming a
45 few simple assumptions about model structure and distributions. Process error can be a large source of
46 near-term ecological forecast error for process-based and statistical models (Dietze 2017; Massoud et al.
47 2018; Thomas et al. 2021). This indicates that these types of models may be misrepresenting system
48 dynamics by using rigid equations that simplify the natural world, require *a priori* parameterization, and
49 cannot adapt on the fly to mismatches between predictions and observations.

50 Deep learning (DL) models, in contrast to process-based and simple statistical models, learn a complex
51 mapping relation between drivers and output by exposing models to training data and an optimization
52 scheme (Shen 2018). DL models can be very accurate and have the potential to reduce error in forecasting
53 by reducing process error. Promising examples of DL-based aquatic forecasting efforts include a short-
54 term forecast of water demand using an artificial neural network by Jain et al. (2001), an accurate DL
55 approach to streamflow forecasting by Xiang and Demir (2020), and algal bloom prediction by Lee and Lee
56 (2018). However, none of these examples have assimilated real-time water observations – such as those

57 available for many sites in the United States (Hirsch and Fisher 2014) – to increase DL prediction accuracy.
58 Data assimilation (DA; Reichle 2008) is commonly used to improve process-based model performance
59 (Thomas et al. 2021), and a DL architecture with a similar ability to update model states with information
60 from real-time data could provide superior forecasts for water-use decisions.

61 In this paper, we describe a real-time forecasting system of stream water temperature that assimilates
62 observations into a DL model as they are collected and produces forecasts of stream temperature 7 days
63 into the future. We evaluated forecasts issued across 93 consecutive days and assess how the DL forecast
64 performance compares with and without data assimilation and compare this performance to a baseline
65 persistence forecast. These forecasts of stream temperature can be used by resource managers in the
66 Delaware River Basin to optimize releases of water from reservoirs to cool downstream segments while
67 retaining enough water to supply New York City and other municipalities with drinking water.

68 Methods

69 Study site

70 We generated real-time forecasts of stream water temperature in the Delaware River Basin in support of
71 drinking water reservoir management decisions. The Delaware River Basin is an ecologically diverse region
72 and a societally important watershed along the East Coast of the United States as it provides drinking
73 water to over 15 million people (Williamson et al. 2015). The extensive and expanding environmental
74 monitoring network in the Delaware River Basin provides a great opportunity to leverage both historical
75 and real-time aquatic information for forecasting stream water temperature at specific locations or basin
76 wide.

77 Drinking water reservoirs in the Delaware River Basin provide essential services to the public while also
78 potentially maintaining suitable aquatic habitat in stream segments directly downstream from the
79 reservoirs. The Flexible Flow Management Program between the State of Delaware, the State of New
80 Jersey, the State of New York, the Commonwealth of Pennsylvania, and the City of New York dictates
81 water management in the basin. The 2017 agreement includes provisions that aim to maintain maximum
82 daily stream water temperature below 23.9°C (75°F) in the upper Delaware River Basin (above Lordville,
83 New York) to ensure cold-water stream habitat. The Cannonsville, Pepacton, and Neversink Reservoirs
84 thermally stratify, which means that during the summer, the deep water release points of the reservoirs
85 discharge cold hypolimnetic reservoir water downstream. Reservoir releases can therefore be used to
86 mitigate anticipated temperature exceedances in streams below the dams. Water in the reservoirs is
87 highly regulated given competing demands of drinking water diversions, maintaining flows for habitat and
88 ensuring freshwater at drinking water intakes, and maintaining cold-water stream habitat. The volume of
89 water in the reservoir used to cool downstream river segments (“thermal banks”) is therefore limited in
90 volume, and reservoir managers need to anticipate when these downstream segments may exceed
91 thermal thresholds in order to use the thermal banks only when needed. Near-term forecasts of stream
92 water temperature with associated uncertainty can help inform managers of when stream temperature
93 may exceed these thermal thresholds.

94 There are five sites below three New York City reservoirs that are relevant to release decisions for
95 mitigating temperature exceedances and for which we forecasted water temperature (Figure 1). These
96 sites included two sites on the West Branch of the Delaware River that are affected by releases from the
97 Cannonsville Reservoir (at Hancock, New York and Hale Eddy, New York), one site on the East Branch of
98 the Delaware River that is impacted by releases from the Pepacton Reservoir (at Harvard, New York), one
99 site on the Delaware River (at Lordville, New York) that is affected by releases from both Cannonsville and

100 Pepacton Reservoirs, and one site on the Neversink River (at Bridgeville, New York) that is affected by
101 releases from the Neversink Reservoir.



102
103 **Figure 1.** Map of the target water temperature forecasting sites in the Delaware River Basin, with call-outs
104 detailing stream distances between upstream reservoirs (open circles) and target sites (closed circles).

105

106

107 **Datasets**

108 **Table 1.** Summary of the different datasets used to train the deep learning (DL) models and forecast stream
 109 temperature 7 days into the future. The DL models produced forecasts with and without data assimilation
 110 (DA), and their model performance was compared to a baseline persistence model that only utilized daily
 111 maximum water temperature to make forecasts. Dataset sources include the U.S. Geological Survey’s
 112 National Water Information System (NWIS), the Water Quality Portal (WQP), Spatial Hydro-Ecological
 113 Decision System (EcoSHEDS), New York City Department of Environmental Protection (NYCDEP), the Office
 114 of the Delaware River Master (ODRM), gridMET, National Oceanic and Atmospheric Administration’s
 115 Global Ensemble Forecasting System (GEFS), Precipitation Runoff Modeling System with a coupled Stream
 116 Temperature Network Model (PRMS-SNTemp), and the General Lake Model (GLM).

Input category	Input	Source	Deep Learning Model				Persistence Model Forecasting
			Pretraining	Training	Forecasting		
					DL	DL+DA	
Observations	Aggregated daily maximum stream temperature	NWIS, WQP, EcoSHEDS		•	•	•	•
	Daily average reservoir release volume	NWIS, NYCDEP	•	•			
	Reservoir releases on forecast issue date	NWIS			•	•	
Scenarios	Reservoir release scenarios for days 1-7	ODRM			•	•	
Gridded weather drivers	Daily minimum air temperature	gridMET GEFS	•	•		•	•
	Daily maximum air temperature	gridMET GEFS	•	•		•	•
	Solar radiation	gridMET GEFS	•	•		•	•
Modeled stream temperature	Process-based average daily stream temperature predictions	PRMS-SNTemp, GLM	•				

117

118 To train our models and forecast stream water temperature at these sites, we pulled publicly available
 119 datasets from the U.S. Geological Survey’s (USGS) National Water Information System (NWIS) (USGS
 120 2021), the Water Quality Portal (WQP; Read et al. 2017), Spatial Hydro-Ecological Decision
 121 System (EcoSHEDS, <http://db.ecosheds.org/>), gridMET (Abatzoglou 2013), and the National Oceanic and
 122 Atmospheric Administration’s Global Ensemble Forecasting System (GEFS,
 123 <https://nomads.ncep.noaa.gov/>). We also pre-trained our DL models on process-based model output in
 124 order to guide the model towards more physically consistent predictions of water temperature (as in Jia
 125 et al. 2021). Below we briefly describe the training and forecasting datasets used to make near-term future
 126 predictions of stream temperature.

127 **Spatial fabric and stream temperature observation dataset**

128 We used the National Geospatial Fabric to define our stream segment physical characteristics, including
 129 segment length and location (Bock and Viger 2014; Viger 2014). We pulled sub-daily observations of
 130 stream water temperature from NWIS, WQP, and EcoSHEDS for both the training dataset and for
 131 assimilating in real-time when making forecasts. The sub-daily observations of stream temperature were
 132 aggregated to daily maximum temperature values for each site that were used as the target variable for

133 our DL models. We matched observations of maximum stream temperature collected at specific latitude
134 and longitude to the river segments by snapping observations to the nearest stream segment within a
135 tolerance of 250 m. Observations farther than 5,000 m along the river channel to the outlet of a segment
136 were omitted from our training and forecasting dataset. If a site was closer to the upstream segment
137 endpoint than the matched segment endpoint, the upstream segment was used as the location of the
138 observation. Segments with multiple observation sites were aggregated to a single maximum water
139 temperature value.

140 [Historical driver dataset](#)

141 We used five drivers (i.e. input features) to train our DL models from 1985 to 2021, including gridMET
142 daily minimum and maximum air temperature, daily average downward shortwave radiation, daily
143 average reservoir release volume from the reservoirs directly upstream of each site, and observations of
144 yesterday's maximum stream temperature. When observations of yesterday's maximum stream
145 temperature were not available, we used yesterday's predicted average stream temperature from the
146 process model pre-training dataset (described below). Observations of daily maximum water temperature
147 were available for 68-98% of the training period time steps depending on the stream segment.

148 [Process model pre-training dataset](#)

149 We used process-based model output from 1985 to 2020 to pre-train the DL models before fine-tuning
150 on observations of maximum water temperature. Pre-training DL models on process-based model output
151 can improve predictive performance, including for out-of-bounds predictions (Read et al. 2019, Jia et al.
152 2021). For water temperature prediction, process-based models use an energy budget approach to
153 estimating temperature change. Given the utility of learning these physical rules from an existing model
154 and the emphasis on predicting summertime maximum temperatures (which could be poorly represented
155 or out-of-bounds entirely from the training dataset), pre-training was used to generate initial weights and
156 biases for the Long Short-Term Memory (LSTM) that were then refined via training on actual observations
157 (following Jia et al. 2021). For models trained in the Delaware River Basin, we developed a pre-training
158 dataset that combines outputs from a reservoir-ignorant stream temperature model and a reservoir-only
159 temperature model to predict daily average stream temperature from 1980 to 2020. We started with the
160 Precipitation Runoff Modeling System with a coupled Stream Temperature Network Model (PRMS-
161 SNTemp) to make predictions of daily average stream temperature (Markstrom 2012, Sanders et al. 2017).
162 PRMS-SNTemp is a natural flow model, meaning that it does not represent reservoirs within the stream
163 network. This natural flow assumption routinely produces predictions of summer stream temperature
164 that are biased high for segments downstream from reservoirs.

165 Improved temperature predictions were generated for water emerging from two major reservoirs in the
166 basin, Cannonsville and Pepacton Reservoirs, using the General Lake Model (GLM v3.1; Hipsey et al. 2019).
167 These GLM models required lake structural information and were driven by daily inflows, outflows, and
168 inflow temperatures. Reservoir hypsographic curves were acquired from Nystrom (2018). Observed daily
169 inflows from major tributaries to each reservoir were obtained from the USGS National Water Information
170 System (USGS 2021) and further modified by a multiplier of 1.8 for Pepacton Reservoir and 1.4 for
171 Cannonsville Reservoir to represent total inflows from both gaged and ungaged tributaries; these
172 multipliers were necessary to achieve long-term water balance among inflows, outflows, precipitation,
173 runoff, and evaporation in each reservoir. Inflow temperatures were estimated as the flow-weighted
174 average of PRMS-SNTemp daily predictions of average stream temperature in the gaged tributaries.
175 Recorded outflows (releases and diversions) were obtained from USGS New York Water Data Reports
176 (USGS 2021). Given these input data, the GLM parameters cd (bulk aerodynamic transfer coefficient for
177 momentum), sw_factor (scaling factor for shortwave radiation), and Kw (light extinction coefficient) were
178 then calibrated to in-reservoir water temperature observations, which were provided by the New York

179 City Department of Environmental Protection (NYC DEP; Oliver et al. 2021). We used the calibrated GLM
180 models to predict the temperature of released water, computed as the flow-weighted average of reservoir
181 water temperatures at the two release depths (one surface, one deep) for each reservoir.

182 The pre-training dataset was computed as a weighted average of temperature predictions from PRMS-
183 SNTemp and GLM, where GLM predictions only influenced stream segments downstream from GLM-
184 modeled reservoirs. The weight given to GLM predictions in each reach was a function of distance, where
185 stream segments closer to upstream reservoirs had temperatures more similar to the reservoir releases.
186 The weight began at 1 at the reservoir outlet and declined with distance downstream according to an
187 exponential decay function that was fit to stream temperature observations; the decay rate was 0.036
188 km^{-1} for Cannonsville Reservoir and 0.040 km^{-1} for Pepacton Reservoir, such that the weight given to GLM
189 predictions was <0.1 after 64 stream kilometers.

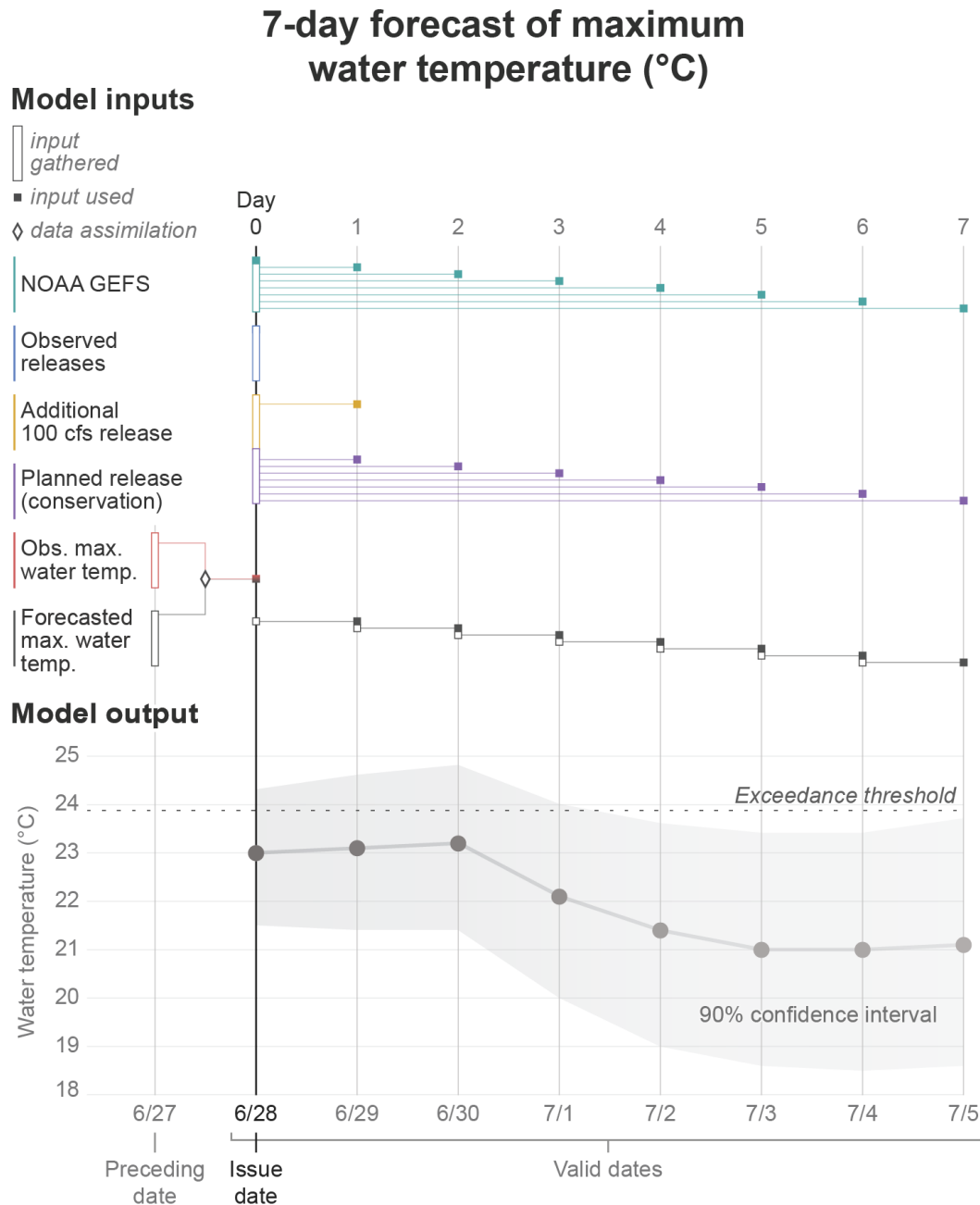
190 Forecasted driver dataset

191 When making near-term forecasts, we used the National Oceanic and Atmospheric Administration's
192 Global Ensemble Forecast System (GEFS, <https://nomads.ncep.noaa.gov/>) operational forecasts of
193 minimum and maximum air temperature, and daily average downward shortwave radiation to predict
194 daily maximum stream temperature for the forecast issue date and 7 days into the future. We also used
195 yesterday's maximum temperature and today's reservoir releases as additional drivers (Table 1). We
196 generated predictions one day at a time, building on predictions from days earlier in the prediction
197 sequence. For predictions made with a zero day lead time (i.e. nowcast), we used today's reservoir
198 releases and yesterday's maximum temperature analysis after assimilating observations (see data
199 assimilation section) as drivers. For predictions made with 1-7 day lead times, we used reservoir release
200 scenarios (see release scenario description below) and model predictions of yesterday's maximum
201 temperature as drivers.

202 We can anticipate realistic release scenarios for 7 days in the future because the Flexible Flow
203 Management Plan dictates release volumes that are mandated given the day of year and current reservoir
204 storage. These releases are referred to as "conservation releases" and are set once per month. For every
205 forecast generated, we made forecasts using two different reservoir release scenarios, one with no
206 additional reservoir release (current conservation release volume +0cfs) and one simulating a relatively
207 high reservoir release scenario (current conservation release volume +100cfs). The +0cfs scenario
208 simulates normal reservoir operations, whereas a +100cfs scenario simulates a thermal bank release that
209 would be used to mitigate expected temperature exceedances. The +100cfs scenario is typically a 20%
210 increase in outflow from the Cannonsville Reservoir. The additional reservoir release scenario was only
211 added on to the lead time day 1 reservoir release driver (Figure 2) and was intended to help the reservoir
212 manager anticipate what would happen to stream temperature tomorrow if a thermal release was made.

213 We downloaded GEFS V12.0 0.25-degree operational model (<https://nomads.ncep.noaa.gov/>) output
214 daily and used the 00 UTC (19:00 EDT) forecast cycle when making stream temperature forecasts. The
215 operational GEFS saves valid times at 3-hour intervals, i.e. 00:00, 03:00, etc. in UTC for 240 hours past the
216 forecast issue time. Starting with the 03:00 valid time, timesteps represent the average, minimum, or
217 maximum of the preceding 3 hours depending on the meteorological driver forecasted. To transform
218 these timesteps to daily values in average solar time in the Delaware River Basin (approximately UTC –
219 5:00), we treat the 09:00 through 30:00 UTC (4:00-25:00 in UTC –5:00) timesteps as day zero, 33:00 –
220 54:00 UTC (28:00 – 49:00 in UTC –5:00) as day one, and so on. This provides the closest possible alignment
221 of GEFS timesteps with average solar time in the Delaware River Basin. Minimum, maximum, and average
222 daily values for these timesteps were then calculated for each GEFS grid cell for minimum air temperature,
223 maximum air temperature, and downward shortwave radiation, respectively. To map GEFS values to

224 individual stream segments, we matched a 0.25-degree NOAA GEFS grid cell with the centroid of the
 225 target stream segment and used the meteorological drivers of that grid cell for the given segment. All
 226 stream segments were almost entirely contained within a single NOAA GEFS grid cell. All 31 ensemble
 227 members were used to make ensemble forecasts of stream temperature; see the model forecasts section
 228 below for how the GEFS ensembles are incorporated into the DL forecasts.



229
 230 **Figure 2.** Schematic of forecast inputs and outputs for the deep learning with data assimilation model,
 231 including when observations or driver data are being generated and used in the model. The model output
 232 panel shows forecast ensemble averages and 90% confidence interval for the forecast issued on June 28,
 233 2021, for the Delaware River at Lordville site.

234

235 Deep learning model

236 We used a Long Short-Term Memory (LSTM) network to forecast daily maximum stream temperature. An
 237 LSTM is a type of Recurrent Neural Network (RNN) that captures temporal relations between system
 238 states and has been used in several other hydrologic modeling applications with great success (Kratzert
 239 et al. 2018; Rahmani et al. 2020). The cell states of the system, \mathbf{c}_t , evolve through time and are modified
 240 at each time t by a filtered, transformed version of the model inputs at that time, \mathbf{x}_t (e.g. meteorological
 241 drivers). The LSTM also inherits information from previous model timesteps through a second set of states
 242 (hidden states, \mathbf{h}_{t-1}) that are a function of \mathbf{c}_{t-1} . At each timestep, the LSTM first generates a candidate cell
 243 state $\tilde{\mathbf{c}}_t$ as:

$$244 \quad \tilde{\mathbf{c}}_t = \tanh(\mathbf{W}_h^c \mathbf{h}_{t-1} + \mathbf{W}_x^c \mathbf{x}_t + \mathbf{b}_c) \quad (1)$$

245 where $\tanh(\cdot)$ is the hyperbolic tangent function, \mathbf{W}_h^c and \mathbf{W}_x^c are learnable weight matrices of the hidden
 246 state and input features, respectively, and \mathbf{b}_c is a learnable bias vector. $\tilde{\mathbf{c}}_t$ represents new ‘memories’ that
 247 could be added at timestep t . Then the LSTM generates a forget gate, \mathbf{f}_t , an input gate, \mathbf{i}_t , and an output
 248 gate, \mathbf{o}_t , as:

$$249 \quad \mathbf{f}_t = \sigma(\mathbf{W}_h^f \mathbf{h}_{t-1} + \mathbf{W}_x^f \mathbf{x}_t + \mathbf{b}_f) \quad (2)$$

$$250 \quad \mathbf{i}_t = \sigma(\mathbf{W}_h^i \mathbf{h}_{t-1} + \mathbf{W}_x^i \mathbf{x}_t + \mathbf{b}_i) \quad (3)$$

$$251 \quad \mathbf{o}_t = \sigma(\mathbf{W}_h^o \mathbf{h}_{t-1} + \mathbf{W}_x^o \mathbf{x}_t + \mathbf{b}_o) \quad (4)$$

252 where $\sigma(\cdot)$ is the sigmoid function, and \mathbf{W} and \mathbf{b} are again matrices and vectors, respectively, of learnable
 253 model parameters. The forget gate is used to filter the information inherited from \mathbf{c}_{t-1} , and the input gate
 254 is used to filter the candidate cell state at time t , in essence determining how much old and new memory
 255 will be passed to the next timestep. The new cell state, \mathbf{c}_t , and the hidden representation, \mathbf{h}_t , are computed
 256 as:

$$257 \quad \mathbf{c}_t = \mathbf{f}_t \otimes \mathbf{c}_{t-1} + \mathbf{i}_t \otimes \tilde{\mathbf{c}}_t \quad (5)$$

$$258 \quad \mathbf{h}_t = \mathbf{o}_t \otimes \tanh(\mathbf{c}_t) \quad (6)$$

259 where \otimes represents element-wise multiplication. The predicted target variables, $\hat{\mathbf{y}}_t$ (in this case just
 260 maximum stream temperature), are calculated as:

$$261 \quad \hat{\mathbf{y}}_t = \mathbf{W}_h^y \mathbf{h}_t + \mathbf{b}_y \quad (7)$$

262 During the model training phase, all weights, \mathbf{W} , and biases, \mathbf{b} , of the LSTM are optimized via iterative
 263 minimization of a loss function, \mathcal{L} , via the back-propagation algorithm (Rumelhart et al. 1986). Our loss
 264 function was the root mean squared error (RMSE) across all timesteps:

$$265 \quad \mathcal{L} = RMSE = \sqrt{\frac{1}{T} \sum_{t=1}^T (\hat{\mathbf{y}}_t - \mathbf{z}_t)^2} \quad (8)$$

266 where T is the total number of training timesteps and \mathbf{z}_t are the observations at time t .

267 We estimated uncertainty in our DL model predictions using Monte Carlo Dropout (MCD) as described in
 268 Gal and Gharhramani (2016), where the model randomly removes a proportion of the network’s recurrent
 269 and input elements (by setting their weights to 0) during each training iteration or prediction activity.
 270 When making many predictions using MCD, this produces an ensemble of LSTM model structures, such

271 that the distribution of their predictions approximates the model uncertainty. We used MCD during
 272 training and for an ensemble of predictions; specifics of these procedures, model dimensions, and other
 273 model hyperparameters are described in the Model Training section below.

274 Data assimilation

275 During the forecasting period, we used data assimilation (DA) to adjust the LSTM states based on recently
 276 collected maximum temperature observations. Previous studies have shown promise for data assimilation
 277 improving DL model predictions (Brajard et al. 2020, Fang and Shen 2020). Additionally, trained LSTM
 278 memory cells can represent dynamic system storages that are similar to our understanding of
 279 environmental processes (Kratzert et al. 2019a), indicating that certain LSTM cells states could be updated
 280 via model-data fusion techniques under the paradigm of using the LSTM as a “scaffold” to connect across
 281 different datasets (Dietze et al. 2013).

282 We used the ensemble Kalman Filter (EnKF) as our data assimilation algorithm to update maximum stream
 283 temperature predictions and the cell states of the LSTM (Evensen 1994). We use the EnKF given the ease
 284 of adding multiple sources of uncertainty via the ensemble members including meteorological driver
 285 uncertainty (e.g. NOAA GEFS ensembles), model process uncertainty (e.g. MCD), and initial condition
 286 uncertainty (e.g. initial LSTM states).

287 We concatenated LSTM state estimates (stream temperature and cell states) into one vector \mathbf{Y} , and we
 288 used this vector for storing model predictions and updating in the EnKF. The \mathbf{Y} vector is propagated
 289 through the forecasting period using the trained LSTM weights and biases, the LSTM state estimates, and
 290 model drivers. The forecasted \mathbf{Y} vector is denoted by \mathbf{Y}^f :

$$291 \mathbf{Y}_{m,t}^f = \begin{bmatrix} \hat{y}_{m,t} \\ \mathbf{c}_{m,t} \end{bmatrix} \quad (9)$$

292 where m indicates the m^{th} ensemble member of the model and the model states are estimated as
 293 described in equations 5-7. $\hat{y}_{m,t}$ was an s by 1 vector where s was the number of stream segments in the
 294 forecasting period for which stream temperature was predicted, and $\mathbf{c}_{m,t}$ was an $[s \times u]$ by 1 vector where
 295 u was the number of hidden units of the LSTM model. When observations were available at a given time
 296 step t , the \mathbf{Y}^f vector was updated using the Kalman gain. The updated \mathbf{Y} vector (\mathbf{Y}^u) was expressed as:

$$297 \mathbf{Y}_{m,t}^u = \mathbf{Y}_{m,t}^f + \mathbf{K}_t \left(\mathbf{z}_t - \mathbf{H}_t \mathbf{Y}_{m,t}^f \right) \quad (10)$$

298 where m was the ensemble member ($m = 1, 2, \dots, N_m$; $N_m = 3100$). \mathbf{z}_t was an s by 1 vector of
 299 observation data from each stream segment being modeled in the forecasting period as $\mathbf{z}_t = \mathbf{z}_t + \boldsymbol{\varepsilon}$,
 300 where $\boldsymbol{\varepsilon}$ was the observation error randomly drawn from a normal distribution with average of 0 and
 301 variance of 0.25°C (see description of observation variance below). \mathbf{H}_t was an s by $[s + s \times u]$ measurement
 302 operator matrix with ones when observations of stream temperature $[I(\mathbf{z}_t)]$ were available and zeros
 303 otherwise. Equation 11 shows what \mathbf{H}_t would look like if there were two stream segments, a and b , and
 304 three LSTM hidden units:

$$305 \mathbf{H}_t = \begin{bmatrix} I(\mathbf{z}_t^a) & 0 & 0 & 0 & 0 & 0 & 0 & 0 \\ 0 & I(\mathbf{z}_t^b) & 0 & 0 & 0 & 0 & 0 & 0 \end{bmatrix} \quad (11)$$

306 \mathbf{K}_t was the Kalman gain weighting matrix, expressed as:

$$307 \mathbf{K}_t = \frac{1}{N_m - 1} \Delta \mathbf{Y}_t \Delta \mathbf{Y}_t^\top \mathbf{H}_t^\top \left(\frac{1}{N_m - 1} \mathbf{H}_t \Delta \mathbf{Y}_t \Delta \mathbf{Y}_t^\top \mathbf{H}_t^\top + \mathbf{R}_t \right)^{-1} \quad (12)$$

308 where \mathbf{R}_t was a s by s observation error covariance matrix. If two stream segments, a and b , were predicted
 309 in the forecasting phase, \mathbf{R}_t would be expressed as:

$$310 \quad \mathbf{R}_t = \begin{bmatrix} \text{Variance } z_t^a & 0 \\ 0 & \text{Variance } z_t^b \end{bmatrix} \quad (13)$$

311 where *Variance* was the estimated variance around observations of stream temperature in each stream
 312 segment at time t . Variance of maximum temperature observations was set to 0.25 C for all observation
 313 times and locations based on reported error in thermistors and temperature variation within stream
 314 segments when there were multiple sampling locations. More analyses are needed to determine if
 315 observation error varies significantly by location, day-of-year, or stream conditions (e.g. streamflow).

316 $\Delta \mathbf{Y}_t$ in equation 14 was an $[s + s \times u]$ by N_m matrix of all ensemble deviations from the average of the
 317 estimated states at time t ($\bar{\mathbf{y}}_t$), expressed as:

$$318 \quad \Delta \mathbf{Y}_t = [\Delta \mathbf{y}_{1,t} \dots \Delta \mathbf{y}_{m,t} \dots \Delta \mathbf{y}_{N_m,t}] \quad (14)$$

319 where the m th column of $\Delta \mathbf{Y}_t$ was:

$$320 \quad \Delta \mathbf{y}_{m,t} = \hat{\mathbf{y}}_{m,t} - \bar{\mathbf{y}}_t \quad (15)$$

321 After states were updated using equation 10, the c states of the LSTM were set to the updated states from
 322 $\mathbf{Y}_{m,t}^u$, and we then proceeded to the next time step to make predictions as time $t+1$. We did not update
 323 the LSTM h states because maximum temperature and c states were updated and the h state is calculated
 324 from the updated c state.

325 Model training

326 We trained separate LSTM models on five management-relevant and well-monitored segments within the
 327 Delaware River Basin. All five stream segments are downstream from reservoirs ranging from 14.6 km to
 328 70.1 km downstream (Figure 1). Each model was first pre-trained for 50 epochs at a learning rate of 0.05
 329 with data from a single segment from 1985 to 2020, using the pre-training dataset as the target features
 330 and the historical drivers as the input features. Next, pre-trained model weights and biases were fine-
 331 tuned with observations of daily maximum temperature for each stream segment, for 350 epochs, at a
 332 learning rate of 0.05, using the same historical drivers from 1985-05-01 to 2021-04-14. We used 6 hidden
 333 units and recurrent and elemental dropout rates of 0.4 for both the pre-train and fine-tune phases. The
 334 model hyperparameters (e.g. number of epochs, learning rates, hidden units, dropout rates) were tuned
 335 manually to achieve reasonable model performance. Although we did not perform a thorough
 336 hyperparameter search, we found these hyperparameters to perform well for our DL forecasting models.
 337 We used TensorFlow v2.5.0 (TensorFlow Developers, 2021) to train and forecast with the LSTM models.

338 An ensemble of LSTM instances is required in the forecasting phase not just for uncertainty reporting but
 339 also for the data assimilation routine described above. We used 31 identical batches for each stream
 340 segment during the training phases, corresponding to the 31 ensemble members from GEFS
 341 meteorological drivers. The ending LSTM states from the fine-tune training phase were used to initialize
 342 the LSTM states during the forecasting phase.

343

344

345 [Model forecasts](#)

346 At each forecast issue time step, we made predictions 7 days into the future using the starting conditions
347 of the previous time step's LSTM model (i.e. stream temperature, hidden, and cell states). We made
348 predictions using trained LSTM models with data assimilation (DL-DA) and without data assimilation (DL)
349 as well as a deterministic persistence forecast as our baseline model (Table 1). The persistence model
350 simply forecasts the same maximum water temperature that was observed yesterday for all 7 days that
351 follow. The DL model forecasted maximum water temperature 7 days into the future using the drivers
352 described in Table 1. And for the DL-DA model, if there were observations available at a given time step,
353 we assimilated the observations into the LSTM model as described in equations 10-15 and updated the
354 stream temperature estimates and LSTM cell states (Figure 2). The DL-DA model for the next forecast
355 issue date was initialized with the updated states and was used to make predictions 7 days into the future.

356 We evaluated forecasts generated from 2021-04-16 to 2021-07-16 because that was the extent of the
357 GEFS archive that we had available. Our LSTM models used the ending states from the fine-tune training
358 phase (fine-tune training ended on 2021-04-14) to initialize the hidden and cell states during the
359 forecasting phase. The DL and DL-DA models incorporated three different sources of uncertainty via the
360 ensemble predictions: driver uncertainty (e.g. GEFS ensembles), DL model uncertainty (e.g. MCD), and
361 initial condition uncertainty (e.g. initial hidden and cell states and yesterday's maximum temperature
362 predictions). Each DL and DL-DA model had a total of 3100 ensemble members where each of the 31 GEFS
363 ensembles were repeated 100 times and used as batch inputs to the DL and DL-DA models. We repeated
364 the ensemble members 100 times to produce a distribution of model predictions for each GEFS ensemble
365 using MCD. We produced 100 predictions for each GEFS ensemble to balance the representativeness of
366 the prediction distribution and computational speed and storage.

367 All model training was conducted on USGS Advance Research Computing resources (USGS Tallgrass; U.S.
368 Geological Survey n.d.) and the operational forecasts were generated using Amazon Elastic Compute
369 Cloud instances. Operational forecasts were run each day immediately following a scheduled observation
370 data pull that occurred at 9:00 AM EDT. Forecasts were summarized to the average of the ensemble
371 predictions, 90% confidence intervals (CI) of the ensembles, and the probability of exceeding 23.9°C (75°F)
372 over the next 7 days for each of the sites of interest, and this summarized forecast output could be sent
373 to the reservoir managers to aid in thermal release decision making (Table 2).

374

375

376 **Table 2.** Example of summarized forecast output that can be delivered daily to the New York City reservoir
 377 managers.

7-day water temperature forecast for Delaware River Basin, issued on 2021-06-29, using conservation release + 0cfs as the release volume.

This information is preliminary or provisional and is subject to revision. It is being provided to meet the need for timely best science. The information has not received final approval by the U.S. Geological Survey (USGS) and is provided on the condition that neither the USGS nor the U.S. Government shall be held liable for any damages resulting from the authorized or unauthorized use of the information.

	EBDR @ Harvard	WBDR @ Hale Eddy	WBDR @ Hancock	DR @ Lordville	NR @ Bridgeville
Forecasted maximum water temperature (deg F), with 90% confidence interval					
Date					
Tuesday, June 29	71.1 (68.7-73.0)	59.4 (57.2-61.3)	64.6 (62.6-66.4)	73.9 (71.2-76.1)	72.3 (70.0-73.9)
Wednesday, June 30	70.7 (67.8-73.2)	59.0 (56.5-61.3)	64.8 (62.2-67.1)	73.9 (70.9-76.6)	71.8 (69.1-73.8)
Thursday, July 01	68.2 (64.4-72.0)	56.7 (53.8-59.5)	63.1 (59.9-66.0)	71.4 (67.8-75.0)	67.8 (63.9-71.8)
Friday, July 02	67.3 (62.8-71.2)	57.0 (53.8-60.3)	62.8 (59.4-65.8)	70.3 (65.5-74.3)	66.7 (61.9-71.1)
Saturday, July 03	66.7 (62.4-70.9)	57.6 (54.1-60.6)	62.6 (59.4-65.7)	69.4 (65.1-73.6)	66.2 (61.7-70.7)
Sunday, July 04	66.7 (62.4-70.7)	57.9 (54.9-60.8)	62.8 (59.5-65.8)	69.1 (64.9-73.6)	67.1 (62.6-71.2)
Monday, July 05	66.9 (62.6-70.9)	58.3 (55.2-61.2)	63.0 (60.1-66.0)	69.3 (64.9-73.6)	67.6 (63.1-71.4)
Tuesday, July 06	67.1 (63.0-70.9)	58.5 (55.4-61.2)	63.3 (60.3-66.4)	69.8 (65.5-73.8)	68.5 (64.6-71.8)
Probability of exceeding 75 degrees F					
Date					
Tuesday, June 29	0	0	0	0.24	0
Wednesday, June 30	0	0	0	0.32	0
Thursday, July 01	0	0	0	0.05	0
Friday, July 02	0	0	0	0.02	0
Saturday, July 03	0	0	0	0.01	0
Sunday, July 04	0	0	0	0.01	0
Monday, July 05	0	0	0	0.01	0
Tuesday, July 06	0	0	0	0.01	0

378

379 **Model evaluation**

380 We evaluated model forecast performance with root mean squared error (RMSE), continuous ranked
 381 probability score (CRPS; Thomas et al. 2021), and bias by comparing valid date predictions to observations
 382 of maximum temperature that occurred on the same dates. Bias and RMSE reflect accuracy of the average
 383 of the ensembles relative to the observed maximum temperature. We also evaluated the distribution of
 384 ensemble predictions using CRPS, which measures both the accuracy and precision of a probabilistic
 385 forecast, where lower values indicate a better model. Because multiple forecasts were valid for any given
 386 day, this allowed us to analyze how model forecast accuracy changed with different lead times and how
 387 it may change across the entire forecast period.

388 We evaluated how well the model characterizes forecast uncertainty with reliability plots where we
389 calculate the proportion of observations that fall within confidence intervals calculated from the
390 ensemble predictions as in Thomas et al. (2021). A well-calibrated forecasting model would have 10% of
391 observations within the 10% forecast confidence interval (i.e., the 45th-55th quantiles of the forecast
392 probability distribution), 20% of observations in the 20% forecast confidence interval (the 40th-60th
393 quantiles), and so on. If a higher or lower percentage of the observations fall within a given forecast
394 confidence interval, then the model is considered underconfident or overconfident, respectively.

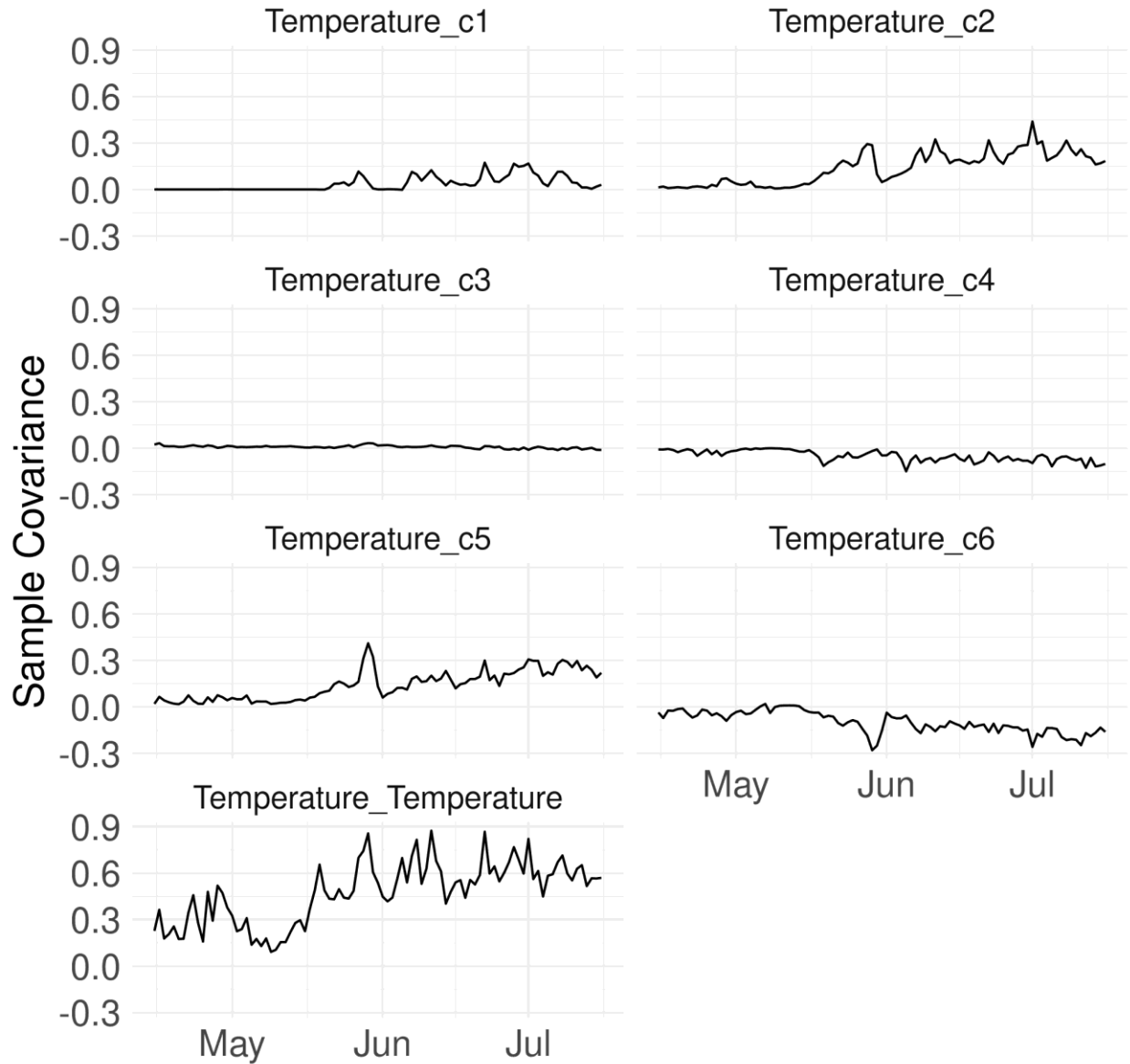
395 Model code used to create pre-training process-based datasets, train the DL models, and forecast with
396 the trained DL models can be found at Zwart et al. (2021). All model drivers and predictions used in this
397 analysis can be found at Oliver et al. (2021).

398 Results

399 Data assimilation and deep learning

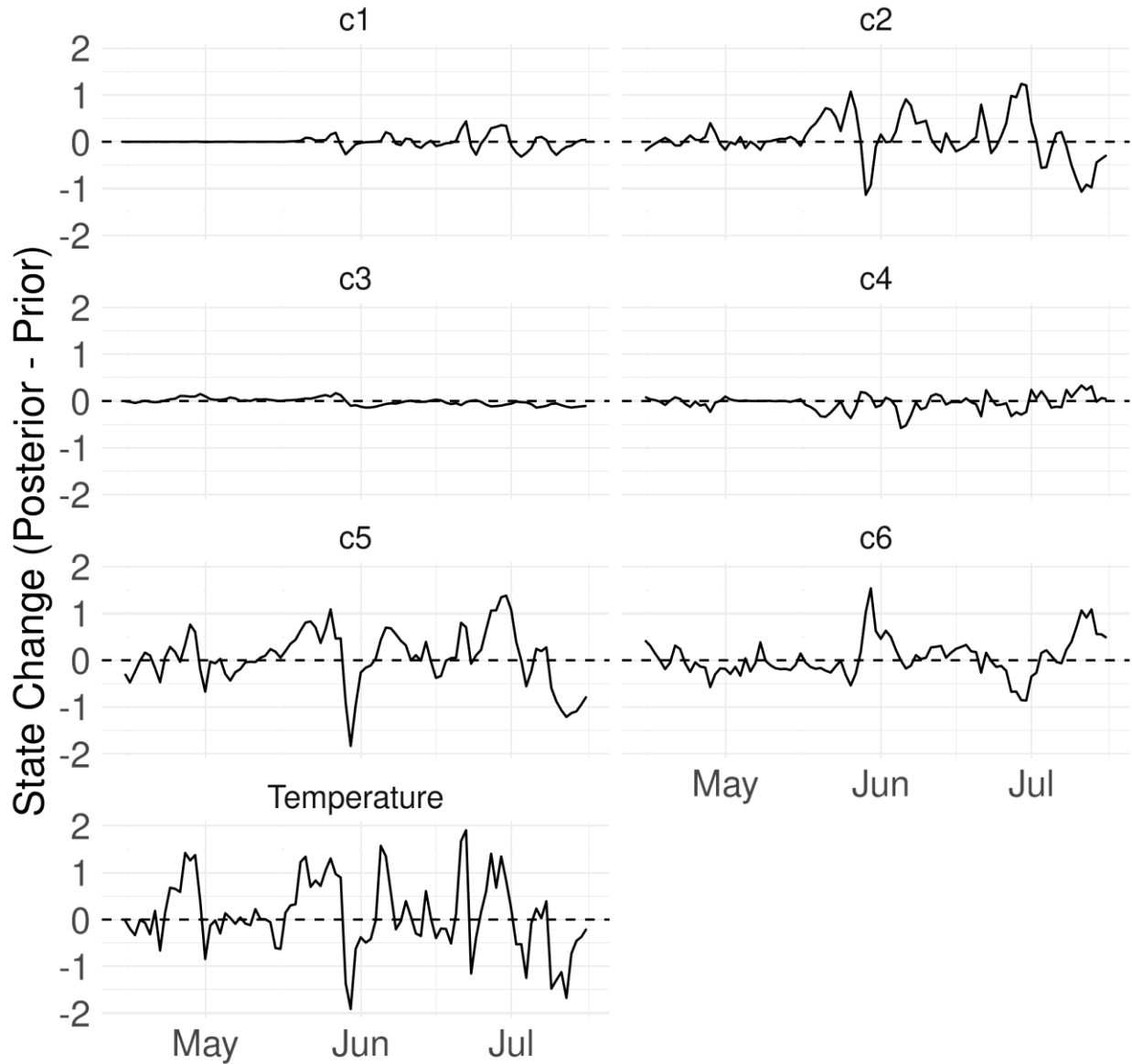
400 Our data assimilation routine successfully updated LSTM cell states using equations 9-15. As an example,
401 we used the Delaware River at Lordville DL-DA model to show relations between temperature predictions,
402 LSTM cell states, and adjustment of these states after assimilating observations of daily maximum stream
403 temperature. The sample error covariance showed that some cell states covaried more strongly with
404 temperature than others, as cell state 2, 5, and 6 had the strongest covariance with maximum stream
405 temperature predictions for the Lordville DL-DA model (Figure 3). Assimilating daily maximum
406 temperature observations into the LSTM adjusted the average ensemble temperature and cell states
407 (Figure 4). Daily maximum stream temperature predictions and cell state 2, 5, and 6 showed the greatest
408 adjustment after assimilating daily maximum stream temperature observations for the Lordville DL-DA
409 model.

410 After assimilating daily maximum stream temperature observations into the LSTM, bias in the average of
411 the ensemble daily maximum stream temperature predictions compared to observations was reduced
412 (Figure 5). Daily maximum stream temperature observations were only available for lead time day -1
413 (Figure 2), so Figure 5 represents the bias in maximum stream temperature predictions for lead time day
414 -1 for both the DL (prior) and DL-DA (posterior) models.



415

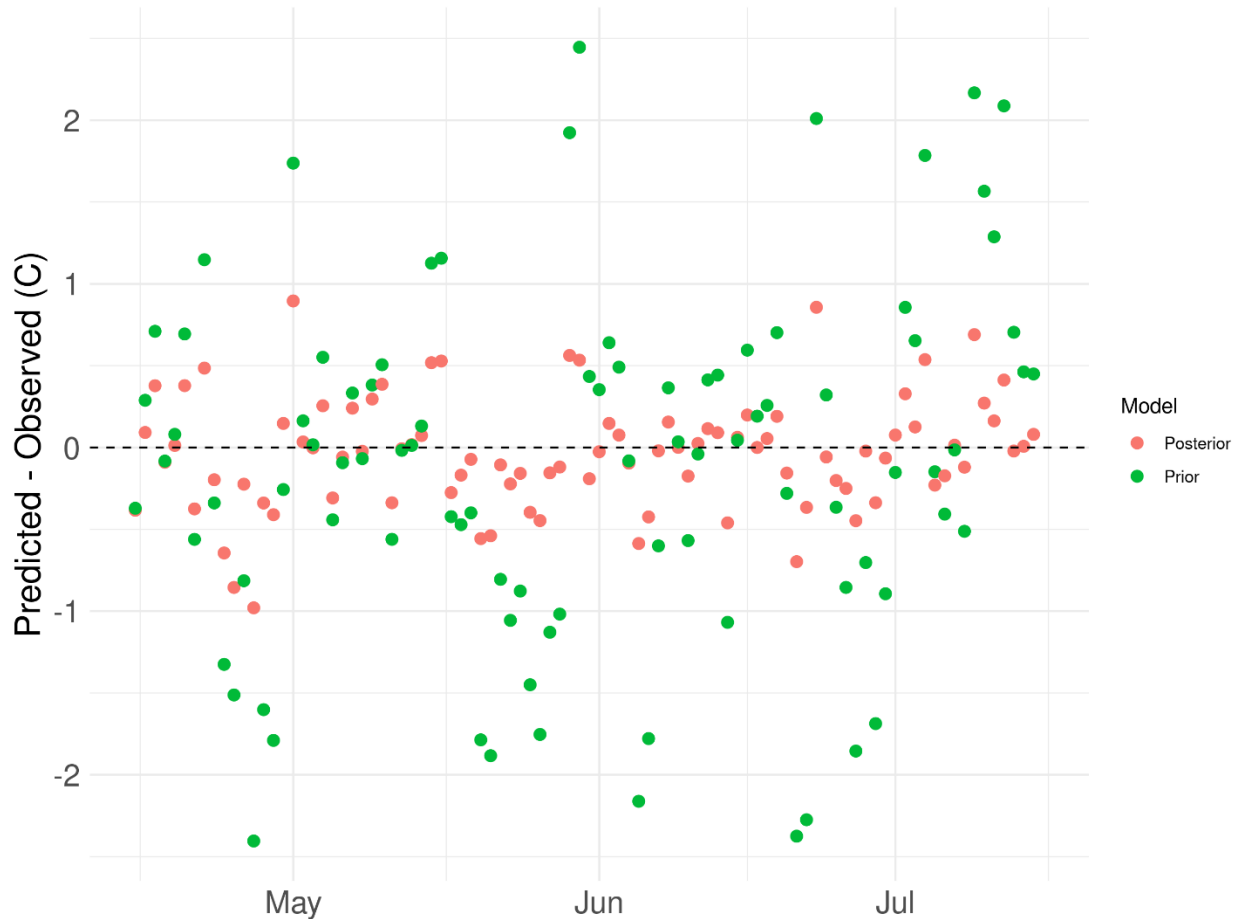
416 **Figure 3.** Sample covariance between daily maximum stream temperature predictions (*Temperature*) and
 417 LSTM cell states 1-6 (*c1-c6*) for the Delaware River at Lordville DL-DA model from 2021-04-15 to 2021-07-
 418 16. Sample covariance is calculated as $\frac{1}{N_m-1} \Delta Y_t \Delta Y_t^T$ as shown in equation 12.



419

420 **Figure 4.** Average ensemble state change for the six LSTM states (c1-c6) and the daily maximum stream
421 temperature prediction (Temperature) for the Delaware River at Lordville DL-DA model from 2021-04-15
422 to 2021-07-16. The average state change is calculated as the difference between the average of the prior
423 state distribution and the average of the posterior state distribution after assimilating daily maximum
424 temperature observations.

425



426

427 **Figure 5.** Difference between predicted daily maximum stream temperature and observed daily maximum
 428 stream temperature for Delaware River at Lordville site from 2021-04-15 to 2021-07-16 for both the
 429 average of the prior distribution and the average of the posterior distribution after assimilating daily
 430 maximum stream temperature observations.

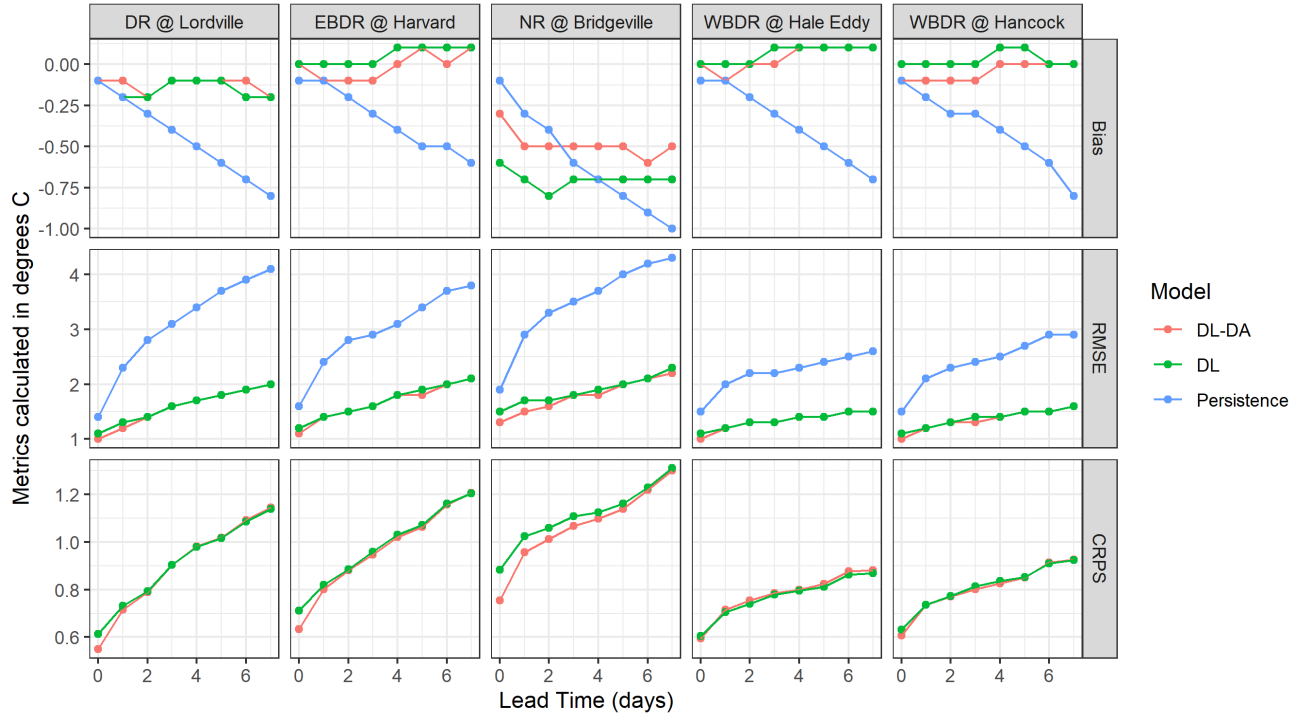
431 [Model accuracy](#)

432 The addition of our data assimilation algorithm (DL-DA model, Figure 6) marginally improved DL forecast
 433 performance for lead times 0 and 1 across all stream segments (7-13% lower RMSE) but had a greater
 434 increase in performance for the sites below Pepacton Reservoir (East Branch of the Delaware River [EBDR]
 435 at Harvard and the Delaware River [DR] at Lordville) and Neversink Reservoir (Neversink River [NR] at
 436 Bridgeville) reservoirs. In some cases, data assimilation reduced forecast accuracy (e.g. CRPS for West
 437 Branch of the Delaware River [WBDR] at Hale Eddy site) and increased model absolute bias; however,
 438 differences in RMSE, CRPS, and biases were minimal (>2% difference) when data assimilation reduced DL
 439 model performance.

440 Our models produced predictions of maximum stream temperature from 1-7 day lead times, with RMSEs
 441 ranging from 1.2 to 1.6°C across all segments for day-ahead forecasts (lead time day 1) for the DL-DA
 442 model (Figure 6). Performance of the models generally worsened as lead time increased; RMSE ranged
 443 from 1.5 to 2.2°C across segments for 7-day ahead DL-DA forecasts. Temperature at the site below the
 444 Neversink Reservoir (NR @ Bridgeville) was the hardest to predict as NR @ Bridgeville site had the highest
 445 RMSE, greatest bias (underpredicted), and highest CRPS for all lead times and both across all models. DL

446 and DL-DA model absolute bias was within 0.2°C for all sites and lead times except for the NR @ Bridgeville
 447 site. Persistence forecast bias became more negative at longer lead times, which is expected when
 448 evaluating stream temperature forecasts issued during the spring to early summertime periods.

449

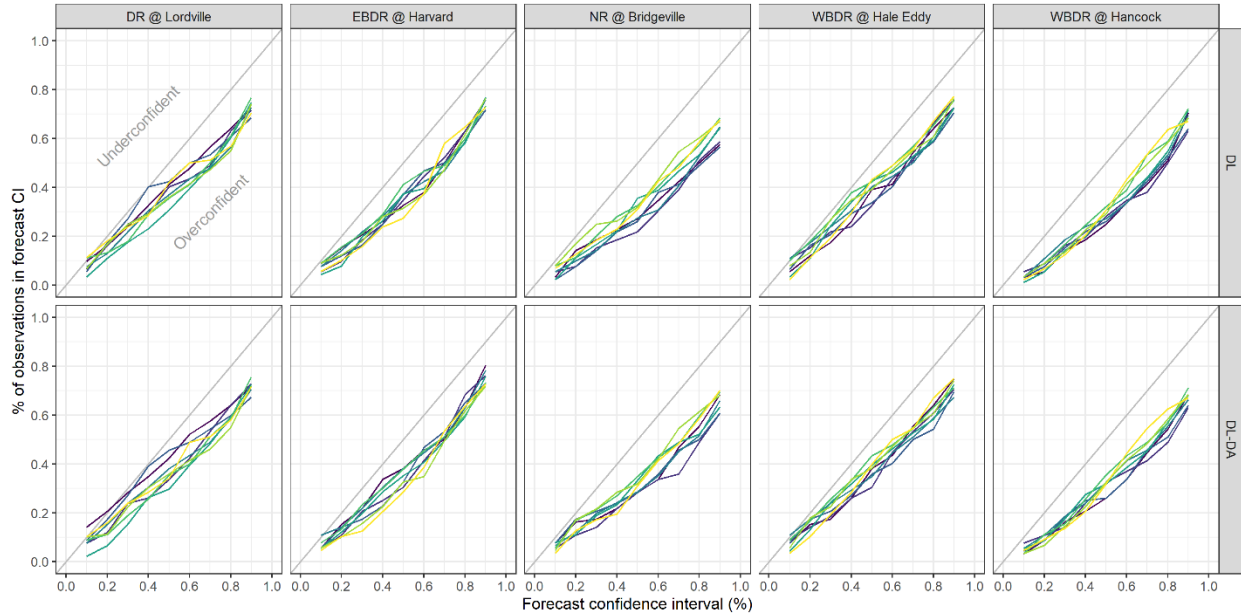


450

451 **Figure 6.** Model forecast accuracy as a function of forecast lead time for each site in the Delaware River
 452 Basin, including bias, root mean squared error (RMSE), and continuous ranked probability score (CRPS).
 453 Forecasts were produced for the Delaware River at Lordville (DR @ Lordville), East Branch of the Delaware
 454 River at Harvard (EBDR @ Harvard), Neversink River at Bridgeville (NR @ Bridgeville), West Branch of the
 455 Delaware River at Hale Eddy (WBDR @ Hale Eddy), the West Branch of the Delaware River at Hancock
 456 (WBDR @ Hancock).

457 **Model Reliability**

458 Our DL and DL-DA models characterized uncertainty relatively well as 57 to 80% of observations were
 459 within 90% confidence intervals (CIs) across all sites, models, and lead times (Figure 7). Overall, there was
 460 very little difference in reliability between the DL and DL-DA models. Uncertainty estimates were mildly
 461 overconfident, with more reliable uncertainty estimates for the central 10% of the CI distribution and
 462 greater overconfidence for larger CIs. There was not a clear pattern in forecast reliability for differing lead
 463 times across sites, as lower lead times were generally more reliable for some sites (e.g. DR @ Lordville),
 464 while longer lead times were more reliable for other sites (e.g. NR @ Bridgeville).



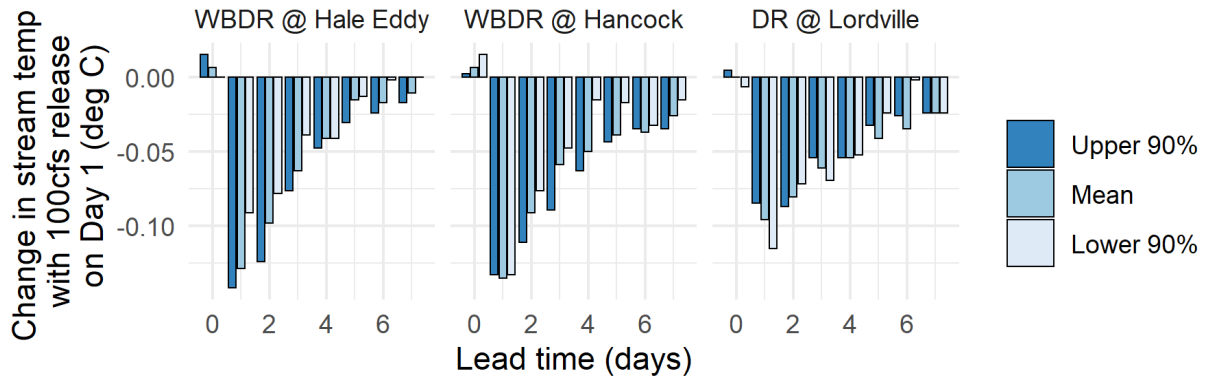
465

466 **Figure 7.** Reliability plots showing the percentage of observations occurring in each of nine forecast
 467 confidence intervals, for deep learning (DL) forecasts (top row) and deep learning with data assimilation
 468 (DL-DA) forecasts (bottom row). A well-calibrated forecast model would fall along the 1:1 line. Forecasts
 469 were produced for the Delaware River at Lordville (DR @ Lordville), East Branch of the Delaware River at
 470 Harvard (EBDR @ Harvard), Neversink River at Bridgeville (NR @ Bridgeville), West Branch of the Delaware
 471 River at Hale Eddy (WBDR @ Hale Eddy), the West Branch of the Delaware River at Hancock (WBDR @
 472 Hancock).

473 **Impact of Release Scenarios**

474 Our DL-DA model predicted a decrease in stream water temperature in response to the additional 100cfs
 475 release added to lead time day 1 (Figure 8). The average decrease in stream temperature across all
 476 forecast issue times in June and July with a 1-day lead time was 0.13, 0.14, and 0.10°C at the WBDR @
 477 Hale Eddy, WBDR @ Hancock, and DR @ Lordville sites, respectively. The effect of the simulated reservoir
 478 release on the simulated temperatures at WBDR @ Hale Eddy and Hancock was greater than that DR @
 479 Lordville. Across all sites, the effect of the simulated release decreased at longer lead times. For nearly all
 480 lead times at the WBDR @ Hale Eddy and Hancock sites, the upper 90th percentile prediction was affected
 481 more by the simulated reservoir release compared to the average prediction and the lower 90th percentile
 482 prediction.

483



484

485 **Figure 8.** Change in simulated maximum stream temperature from a simulated 100cfs release from the
 486 Cannonsville Reservoir added to lead time day 1. Data shown are from June and July and averaged across
 487 all forecast issue dates. The EBDR @ Harvard and NR @ Bridgeville sites are not shown because they are
 488 not affected by releases from the Cannonsville Reservoir.

489 Discussion

490 Accurate and reliable forecasts of environmental variables are becoming important tools for
 491 environmental resource management (Dietze et al., 2018; Bradford et al. 2020). Using a novel integration
 492 of DL and DA, our models produced accurate predictions of maximum stream water temperature 7 days
 493 into the future for forecasts issued between 2021-04-16 and 2021-07-16, with an average RMSE of 1.5°C
 494 across all sites and forecast lead times. Our models also characterized uncertainty relatively well and can
 495 aid reservoir managers in decisions about when to release water from reservoirs to cool stream segments.
 496 Below we discuss the forecast performance and further development that could improve upon the DL-DA
 497 forecast accuracy and reliability.

498 Forecast accuracy

499 Although our models produced accurate predictions overall, there was variation in performance across
 500 sites and lead times (Figure 6). The persistence forecast performance can give insight into the
 501 predictability across sites as better performing persistence forecasts often indicated better performing DL
 502 and DL-DA model performance. For example, WBDR @ Hale Eddy and Hancock sites had the best
 503 performing persistence, DL, and DL-DA forecasts at longer lead times, indicating that there is likely
 504 stronger autocorrelation in maximum stream temperature at these sites and higher predictability (Figure
 505 3). Stream temperature at these sites are heavily influenced by the upstream Cannonsville Reservoir and
 506 water released from the reservoir likely mutes variability in maximum stream temperature thereby
 507 increasing autocorrelation. This contrasts with the NR @ Bridgeville site which generally had the lowest
 508 forecast performance for the persistence, DL, and DL-DA models across all lead times. The NR @
 509 Bridgeville DL model was potentially missing important drivers that influence maximum stream
 510 temperature such as streamflow or precipitation. We did not model reservoir temperature in the
 511 Neversink Reservoir above Bridgeville, and the lack of a pre-trainer that simulated reservoir outlet
 512 dynamics may have partially accounted for decreased performance at this site. To improve upon the
 513 accuracy of DL model forecasts for these sites, DL models could be trained across multiple sites (Kratzert
 514 et al. 2019b), additional upstream information or meteorological variables could be added as drivers
 515 (Rahmani et al. 2020), or the DL model structure could be modified to represent network connections
 516 across stream segments (Jia et al. 2021).

517 [Data assimilation and deep learning](#)

518 We found that combining data assimilation with our DL models modestly improved forecast performance
519 of maximum stream temperature, typically at 0- and 1-day lead times, compared to our DL model without
520 data assimilation. Data assimilation is heavily relied upon in forecasting applications to adjust model
521 states, parameters, and drivers prior to issuing forecasts, and there are increasingly more examples of
522 assimilating data into DL models to make improved predictions. For example, Brajard et al. (2020) use an
523 iterative DL and data assimilation algorithm to emulate the Lorenz 96 model using sparse and noisy
524 observations, which can be a challenging task for DL models alone, and Fang and Shen (2020) show that
525 a data integration kernel within a DL model improves near-term forecasts of soil moisture. Given these
526 past successes for improving DL predictions with data assimilation, we might expect our DL-DA model to
527 show even more improvement beyond our DL model without data assimilation. However, the modest
528 improvement is likely due to our base DL model still having access to the same observation information
529 as the DL-DA model because yesterday's maximum temperature observation was used as a model driver
530 for both the DL and DL-DA models. Additionally, loss of information occurs using any data assimilation
531 algorithm (Nearing et al. 2018) and the choice of the data assimilation algorithm will dictate how
532 efficiently the algorithm combines these pieces of information together. Given these losses of information
533 from data assimilation algorithms with rigid assumptions, it may be beneficial to allow the DL model itself
534 decide how to use the new observation information; however, more research is needed to assess which
535 data assimilation algorithms optimally combine new observations with DL model predictions to improve
536 forecast performance.

537 [Forecast uncertainty](#)

538 Our DL-DA model characterized forecast uncertainty well; however, the model tended to be overconfident
539 in predictions for the time period we analyzed. The overconfidence is likely due to overfitting models to
540 the training dataset, which is a common problem in all types of modeling, especially machine learning. To
541 improve upon the forecast uncertainty estimates, we may want to train global DL models on many stream
542 segments and fine-tune to individual segments, add in noise regularization to the drivers during training,
543 and / or increase the dropout rate for the recurrent and input elements during model training.
544 Additionally, Klotz et al. (2021) show that mixture density networks have better uncertainty
545 characterization compared to MCD for streamflow predictions, indicating that adoption of these relatively
546 newer methods may be promising for improving DL forecast reliability.

547 [Reservoir release scenarios](#)

548 Some forecast products can trigger decisions that alter the conditions being forecasted; in our use case, a
549 forecast that predicts water temperature will exceed 23.9°C (75°F) in the future could trigger a manager to
550 call for larger reservoirs releases that would cool downstream waters to maintain coolwater fish habitat.
551 Generating a forecast with feedbacks that can be exposed to the decision maker can make the forecasts
552 more actionable. Our implementation included two water release scenarios, which provided a range of
553 expected outcomes for taking no action (conservation releases +0cfs thermal bank release) and
554 alternatively, issuing a single-day preventative action (conservation releases +100cfs release; see Figure
555 5) designed to keep future downstream water temperatures from exceeding a biologically relevant
556 temperature threshold. We observed a decrease in stream temperature following the reservoir release
557 scenario, with the DR @ Lordville site having the smallest response, which was expected because DR @
558 Lordville was the farthest downstream from the Cannonsville Reservoir of the three sites. Although we
559 observed an expected decrease in stream water temperature following the +100cfs release scenario, the
560 resulting average decrease was small for the three segments downstream from the Cannonsville Reservoir
561 (decrease ~0.1°C). Currently, the model does not explicitly have information that would help it distinguish
562 between increases in release volumes that are coming from the surface of the reservoir (through the

563 spillway) versus the bottom of the reservoir (via conservation or directed releases), nor is it exposed to
564 flow changes related to precipitation events. We speculate that adding in stream discharge and/or
565 precipitation as an additional model driver may help the DL model learn about the effects of reservoir
566 releases under various release conditions. Additionally, we could expose the DL model to more pre-
567 training scenarios that could help the model learn about the effect of reservoir releases on downstream
568 temperature, potentially increasing the effect of reservoir release on changes in stream temperature.

569 Future efforts could extend this simple +0cfs/+100cfs case by exposing a greater variety of scenarios to
570 reservoir release decision makers, including additional release volumes and their future timing (e.g.,
571 beyond the day-of-issue). Alternatively, the complete modeling system could be provided to the end-users
572 where model inputs could be fully controlled by the decision makers. The combination of real-time data
573 and modern deep learning modeling advances described here can provide timely and accurate
574 information for critical water resources decisions.

575 [Summary](#)

576 We demonstrate the utility of combining data assimilation with DL models for forecasting management-
577 relevant environmental variables. Our DL-DA model produced accurate and reasonably reliable forecasts
578 of maximum stream temperature 7 days into the future. Despite the reliability and accuracy of the
579 forecasts, future improvements could include increasing the complexity of the underlying models or the
580 modes of DA, in addition to providing sub-daily inputs to the model that could help better resolve daily
581 maximum temperatures. Given the flexibility of DL models, we expect future extensions and
582 improvements on this type of framework could be used for many types of environmental forecasting
583 applications.

584

585

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