

1 **Matters Arising**

2 **Unrealistic phytoplankton bloom trends in global lakes derived from Landsat**
3 **measurements**

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9 Given its advantages for synoptic and large-scale observations, satellite remote sensing
10 has been widely used to effectively monitor the water quality of inland and coastal
11 environments. Using satellite-derived reflectance data from the Landsat 5 Thematic
12 Mapper (L5TM) as a proxy for algal bloom intensity, Ho et al. ¹ showed an increase in
13 peak summertime bloom intensity in 68% of the 71 large lakes worldwide from 1982
14 to 2012. However, we question the veracity of their finding for at least two reasons: (1)
15 satellite-derived reflectance in a single near-infrared (NIR) band is not a reliable proxy
16 for bloom strength due to the strong impacts of suspended sediments and aquatic
17 vegetation, and (2) the infrequent satellite observations from L5TM (one cloud-free
18 image every 1-2 months) make it difficult to draw statistically meaningful conclusions.
19 Therefore, although it is natural to speculate that more blooms may be found in lakes
20 under changing climatic conditions, the work by Ho et al. ¹ needs to be revisited before
21 reaching any solid conclusions.

22 Ho et al. ¹ argued that the L5TM-estimated bloom intensity (B_{NIR}) (see Equation 2 in
23 Ho et al. ¹), which is basically the reflectance in the NIR band, is correlated with
24 chlorophyll-a (Chla) concentration or phytoplankton biomass. However, this argument
25 became questionable when we examined the correlations between in situ Chla and
26 water reflectance (in situ reflectance was aggregated into the NIR reflectance equivalent
27 of band 4 of L5TM ²) with data collected from 15 lakes in China and from waters with
28 varying eutrophic status (Chla ranging between 1.5 and 222.6 mg m⁻³) (see Extended
29 Data Fig. 1). We revealed nonsignificant relationships ($p>0.05$) between near surface
30 Chla and NIR reflectance in three different Chla ranges (full Chla range, Chla>50 mg
31 m⁻³ and Chla>10 mg m⁻³). Such complex relationships between spectral reflectance and
32 Chla concentrations were also demonstrated by Spyrakos et al. ³ when using in situ data
33 from various inland waters around the world. Theoretically, the signal in the NIR band
34 can be attributed to various water constituents in addition to algal blooms, and the
35 contributions from suspended sediments and the presence of aquatic plants could be
36 two of the most common perturbations in inland lakes. Ho et al. ¹ attempted to mask
37 out waters associated with high sediment loads with the use of hue, but as detailed later,
38 the hue defined in Ho et al. ¹ does not accurately reflect the color of a water body and
39 is thus not effective for distinguishing phytoplankton blooms from sediment-dominated
40 waters.

41 Bloom strength tends to be substantially overestimated in sediment-rich waters.
42 Examples from two of the lakes studied in Ho et al. ¹ (Songkhla Lake in Thailand and
43 Hongze Lake in China, see Fig. 1) show that the B_{NIR} of the high-turbidity, low-algae
44 pixels (yellowish in true-color images) was higher than that of the algae-present pixels
45 (greenish in true-color images) within the same images. The examination of true-color
46 and the corresponding B_{NIR} images shows that historical L5TM observations have
47 captured sediment plumes in at least 58 (82%) of the 71 studied lakes, and these plumes
48 could be incorrectly labeled as algal blooms due to their high B_{NIR} values (see some
49 examples in Extended Data Fig. 2). As well supported by previous studies using in situ
50 data from both of the studied lakes in Ho et al. ¹ and other global coastal/inland waters,
51 the NIR reflectance in turbid waters can be substantially enhanced by sediment-induced
52 strong backscattering signals (see Extended Data Table 1). In inland lakes, episodic
53 meteorological (e.g., wind, precipitation) and hydrological (e.g., riverine discharge)
54 events can strongly influence sediment concentrations ⁴, as exemplified by previous
55 studies in Lake Erie⁵ and Lake Okeechobee in the USA ⁶ and Hongze Lake in China ⁷
56 (three lakes examined in their study). As such, the impacts of water turbidity on B_{NIR}
57 should be evaluated carefully.

58 Similar to high sediment loads, the growth of aquatic vegetation can lead to the
59 overestimation of bloom severity. Pixels with high B_{NIR} values – in particular, vegetated
60 waters (darkish in true-color images) rather than bloom areas – were also found within
61 the same lakes (see Songkhla Lake in Fig. 1), where massive submerged plants have
62 previously been reported ⁸. The reason is that algal blooms and submerged vegetation
63 share similar spectral curvatures and comparable magnitudes of NIR reflectance values,
64 as demonstrated by the in situ hyperspectral measurements for Taihu Lake in China (a
65 shallow lake that is ~200 km from Hongze Lake) (see Extended Data Fig. 3). Moreover,
66 previous studies with datasets collected across various global regions and plant species
67 also showed markedly increased NIR reflectance due to the presence of submerged
68 vegetation (see Extended Data Table 2). Currently, challenges still exist when one
69 attempts to distinguish submerged plants from algal blooms with multispectral satellite
70 images, not to mention using a single NIR band ⁹. Indeed, a literature search revealed
71 that of the 71 studied lakes, 41 (58%) were found to contain abundant aquatic plants
72 (see Extended Data Table 3), and their impacts on B_{NIR} should have been considered.

73 A hue-based mask (Equations 3 & 4 in Ho et al. ¹) was designed to exclude potential
74 contamination from sediments. However, this approach has failed in numerous cases
75 (see examples in Extended Data Fig. 2). This is mainly due to the inclusion of the
76 atmospheric signals in the calculation of hue, i.e., the hue was estimated using the top-
77 of-atmosphere (TOA) reflectance. Thus, this hue reflects the color of the combined
78 signal of the atmosphere and the water, not the hue of the water itself. As shown in
79 Extended Data Fig. 4, atmospheric molecular scattering (or Rayleigh scattering) alone
80 could dominate the TOA reflectance for water bodies in the blue band ¹⁰. Even worse,
81 the method (i.e., F_{mask} ¹¹) used to determine lake surface area could lead to substantial
82 underestimations of bloom severity. As the examples in Fig. 1c-e and Extended Data

83 Fig. 5 show, when true-color images reveal in vivo bloom occurrences, such areas failed
84 to pass the Fmask and were excluded in further B_{NIR} calculations. Indeed, the
85 examination of their studied lakes showed that most of the severe blooms with surface
86 scum were missed due to the improper use of Fmask. This is because intense blooms
87 often cause high normalized difference vegetation index (NDVI) values that can exceed
88 the threshold used by Fmask (e.g., $NDVI < 0.1$) to identify water pixels¹¹. Since the
89 Fmask algorithm was originally designed for cloud and cloud-shadow detection¹¹,
90 further considerations are required when it is used for water area identification.

91 Furthermore, the infrequent L5TM observations are well known for their limitations in
92 terms of capturing the short- and long-term dynamics of lacustrine algal blooms. Such
93 limitations could be exacerbated by frequent cloud distributions, which also pose one
94 of the challenges associated with optical satellite remote sensing. Statistically, the
95 global mean daily cloud-free probability is 33%, with seasonal differences of $< 5\%$ ¹².
96 In other words, when L5TM overpasses 23 times within a year because of its 16-day
97 revisit period, the annual mean number of cloud-free observations for a given location
98 is only ~ 7.5 even without any other unfavorable observational conditions (such as sun-
99 glint). As a compromise between data availability and result fidelity, Ho et al¹ excluded
100 those years with fewer than 3 valid images in five summer months. We replotted a time
101 series of algal bloom areas in Taihu Lake that was produced by Hu et al.¹³ (see
102 Extended Data Fig. 6), which was obtained using cloud-free images from daily
103 Moderate-resolution Imaging Spectroradiometer (MODIS) satellite observations
104 (revisit period of ~ 1 image per day) between 2000 and 2008. Of the > 300 cloud-free
105 daily MODIS images within the 9-year period, only 24 shared the same overpassing
106 dates as L5TM. Furthermore, detecting a bloom on the basis of remote sensing imagery
107 depends strongly on wind, as the fraction of the satellite-observable surface bloom in
108 relation to the total phytoplankton biomass is also a function of wind speed^{14,15}. Due to
109 the unpredictable nature of cloud occurrence and wind speed, the temporal dynamics
110 of bloom features were difficult to characterize with L5TM datasets.

111 Our results have clearly demonstrated that the use of L5TM-based B_{NIR} by Ho et al.¹
112 as a proxy for algal bloom strength is questionable for the majority of the lakes
113 examined in their study. The incorrect use of a water mask algorithm (i.e., Fmask) also
114 leads to the omission of the most severe blooms with floating scum. The use of limited
115 Landsat observations (often one cloud-free image every 1-2 months) is problematic for
116 drawing statistically meaningful conclusions. Therefore, the trends in phytoplankton
117 blooms for the 71 global lakes derived by Ho et al.¹ appear unrealistic. In summary, a
118 significant amount of work, including the development of reliable algorithms for bloom
119 detection and the use of statistically meaningful observations, is still required to
120 estimate the multidecadal changes in bloom conditions before any attempt is made to
121 interpret such “changes.”

122 **Data availability** The Landsat data can be obtained from the U.S. Geological Survey
123 at <https://glovis.usg.gov>. The in situ spectral and Chla data will be provided to the public

124 upon acceptance of this manuscript.

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164 **Author contributions** L.F. initiated the project and wrote an initial draft of the
165 manuscript, and Y.D., X.H., and Y.X. performed the data processing and analysis. All
166 authors participated in interpreting the results and revising the manuscript.

167 **Competing interests** Declared none.

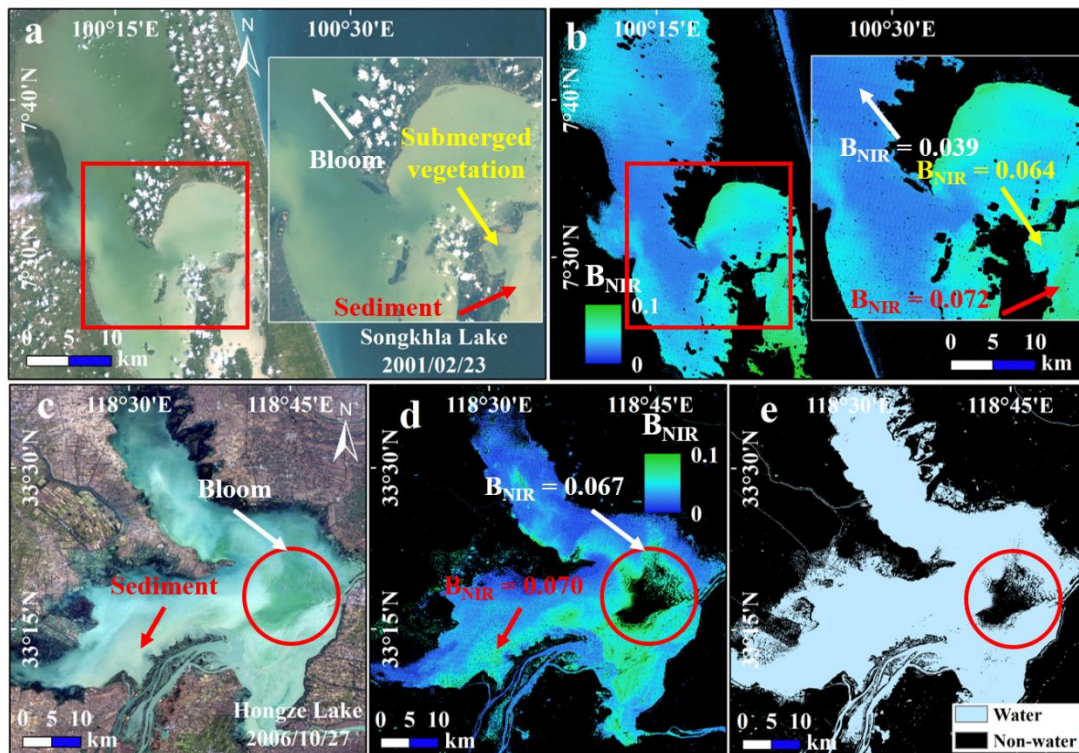
168 **Additional information**

169 **Supplementary information** accompanies this Comment.

170 **Correspondence and requests for materials** should be addressed to L.F.

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176 **Figure 1 | Examples showing the problems associated with L5TM-based bloom**
177 **intensity (B_{NIR} , estimated with Equation 2 in Ho et al. ¹) in global lacustrine**
178 **phytoplankton bloom detection. L5TM true-color composites and corresponding**
179 B_{NIR} map for Songkhla Lake in Thailand (a-b) and Hongze Lake in China (c-d). (e)
180 Water mask determined by F_{mask}^{11} for Hongze Lake using the same image in c.
181 Areas with either high sediment loads (yellowish in true-color images, indicated by
182 red arrows) or the presence of submerged vegetation (darkish in true-color images,
183 indicated by yellow arrows) exhibit higher B_{NIR} values than the bloom-occurring
184 pixels (greenish in true-color images, indicated by white arrows), leading to erroneous
185 interpretation of algal blooms. An intense bloom in Hongze Lake (within the red
186 red circle) has been erroneously classified as non-water by F_{mask} and excluded in the
187 B_{NIR} map (d). More examples of these problems in many other lakes studied in Ho et
188 al. ¹ are available in the Extended Data Figs. 2&5. The red squares within panels a & b
189 indicate inset location.