

1 **A data-driven approach for characterizing community scale air pollution exposure**
2 **disparities in inland Southern California**

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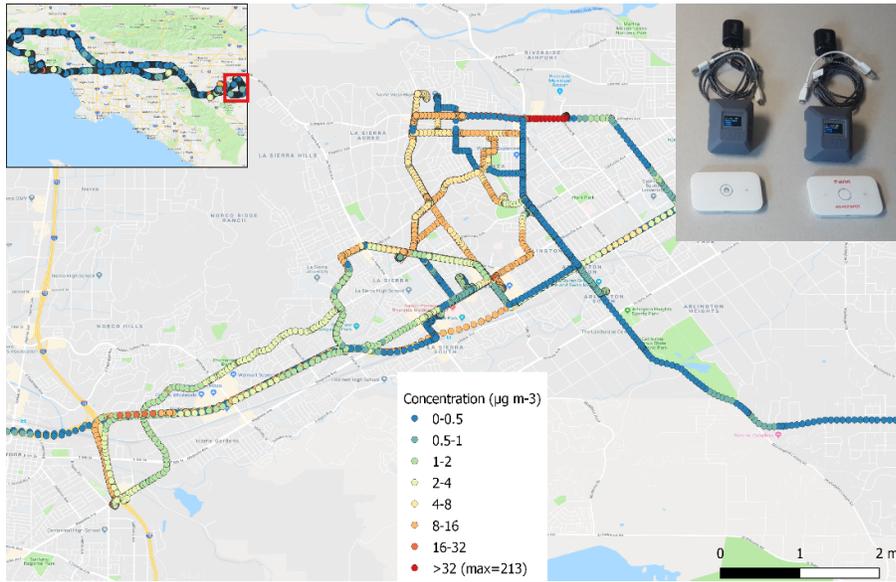
13 **Highlights**

- 14 • Wearable monitors enable high temporal resolution analysis of personal exposure
- 15 • Microenvironments were identified by GIS-based spatial clustering of measurements
- 16 • Most vulnerable community had highest observed personal-ambient ratios in the home
- 17 • High variability in personal PM_{2.5} despite low variability in ambient PM_{2.5}

18 **Abstract**

19 In 2017, Assembly Bill 617 was approved in state of California, which mandated the allocation of
20 resources for addressing air pollutant exposure disparities in underserved communities across the
21 state. The bill stipulated the implementation of community scale monitoring and the development
22 of local emissions reductions plans. We aimed to develop a streamlined, robust, and accessible
23 PM_{2.5} exposure assessment approach to support environmental justice analyses. We sought to
24 characterize individual PM_{2.5} exposure over multiple 24-hr periods in the inland Southern
25 California region, which includes the underserved community of San Bernardino, CA. Personal
26 sampling took place over five weeks in Spring of 2019, and personal PM_{2.5} exposure was
27 monitored for 18 adult participants for multiple, consecutive 24-hr periods. Exposure and location
28 data were available at five-second resolution, and participant data recovery was 50.8% on average.
29 A spatial clustering algorithm was used to classify data points as one of seven microenvironments.
30 Mean and median personal-ambient PM_{2.5} ratios were aggregated along SES lines for eligible
31 datasets. GIS-based spatial clustering facilitated efficient microenvironment classification for
32 more than 920,000 data points. Mean (median) personal-ambient ratios ranged from 0.02 (0.00) to
33 3.49 (0.55) for each microenvironment when aggregated along SES-lines. Aggregated ratios
34 indicated that participants from the lowest SES community experienced higher home exposures
35 compared to participants of all other communities over consecutive 24-hr monitoring periods,
36 despite high participant mobility and relatively low variability in ambient PM_{2.5} during the study.
37 The methods described here highlight the robust and accessible nature of the personal sampling
38 campaign, which was specifically designed to reduce participant fatigue and engage members of
39 the inland Southern California community who may experience barriers when engaging with the
40 scientific community. This approach is promising for larger-scale, community-focused, personal
41 exposure campaigns for direct and accurate analysis of environmental justice.

42 Keywords: particulate matter; wearable monitors; personal exposure; environmental justice



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1. Introduction

Ambient particulate matter (PM) has been widely studied, and researchers have carefully examined the impact of PM exposure on human health. Many air monitoring stations are operated in the U.S. to measure the trends and composition of ambient PM in support of the National Ambient Air Quality Standards (NAAQS). However, ambient PM concentrations may not reflect actual daily personal exposure (PE) (Koistinen et al., 2004). Further the sparseness of the monitoring network leads to low spatial resolution data and necessitates gap-filling, which affects the accuracy of PM exposure assessments that are based on ambient measurements (X. Yu et al., 2019).

People spend most of their time indoors (approximately 85-90%) and are most frequently exposed to indoor pollutants (Long et al., 2001). Home and workplace are the two most dominant indoor microenvironments. Indoor PM originates from cooking, smoking, cleaning products, vacuuming, and dusting; while in offices, PM is emitted from printing, mechanical grinding, consumer products, and dusting. The Environmental Protection Agency (EPA) carried out the particulate total exposure assessment methodology (PTEAM) study on 178 non-smoking randomly selected homes in Riverside, CA. The study showed that indoor $PM_{2.5}$ (PM with an aerodynamic diameter less than or equal to $2.5 \mu m$) levels were slightly lower than outdoor levels during the day. However, at night the differences were significant (Clayton et al., 1993; Özkaynak et al., 1996; Thomas et al., 1993). Although ambient $PM_{2.5}$ penetrates into indoor environments, individual behaviors and living conditions are found to be the most important factors that affect indoor concentrations of PM (Kulmala et al., 1999; Long et al., 2001; Wallace, 1996).

Further, human mobility must also be taken into account for accurate exposure assessment. Yu et al. compared call detail record and home-based methods to estimate biases in exposure methods. The study showed that the home-based method both over- and under-estimates air pollutant exposure levels (H. Yu et al., 2018). In addition, many studies have used outputs from chemical transport models to verify the misclassification when using central monitor concentrations (CMC) to represent the exposure near the monitoring sites. Hu et al. showed that the population weighted concentrations of primary $PM_{2.5}$ of the model differ from the CMC values by -40 to +60%. The misclassification could be significant when assuming the same representative distance across central monitoring sites for multiple pollutants in a large-scale, spatial and temporal epidemiology studies (Hu et al., 2019).

75 Advancements in low-cost environmental sensing technologies have enabled the development of
76 small, portable, and relatively precise PM sensors for personal exposure assessment. In a recent
77 study by Quinn et al., filter-based, wearable, automated microenvironmental aerosol samplers
78 (AMAS) were used to conduct a personal exposure study with 37 high school students from 25
79 high schools in Fresno, CA (Quinn et al., 2018). The wearable AMAS enabled the measurement
80 of black carbon and oxidative potential in targeted microenvironments, but the measurements were
81 coarsely-resolved in time. Further, low-cost optical PM sensors have very high sampling
82 frequencies, and low-cost sensing measurements are moderately accurate (Feenstra et al., 2019).
83 The Plantower PMS (v. 1003/3003) is a commonly used optical sensor, and has a correlation
84 coefficient of 0.88 with the federal reference method (FRM), which reflects the viability of the
85 sensor for exposure measurements (Kelly et al., 2017). Combined with Internet of Things (IoT)
86 technology, the Plantower PMS can be further integrated to deliver more functionalities to end
87 users. Data collected from a low-cost sensing device or IoT network can be uploaded to the cloud
88 and made available in near-real-time to users. Despite of all the conveniences of low-cost sensing,
89 there are still room for improvements of PM sensor accuracy. Sensors require consistent
90 calibration, and the measurements may require additional post-processing (Zheng et al., 2018).

91 In this paper, we detail a pilot-scale personal exposure campaign using wearable PM_{2.5} sensors
92 with real-time, remote monitoring capability. Our study engaged residents of five inland Southern
93 California cities and captured spatial and temporal variability of PM_{2.5} exposures over multiple,
94 consecutive 24-hour periods. The main objective of this pilot study was to develop and implement
95 a high-resolution monitoring and analysis framework for characterizing PM_{2.5} exposure variability
96 for individuals from different cities of residence and subsequently different socioeconomic status
97 (SES) neighborhoods. As Southern California historically has high ambient PM_{2.5} levels, we
98 sought to understand which microenvironments posed the greatest exposure risk in the region. Our
99 study elucidates the behavior-dependent patterns of PM_{2.5} exposure in a high-traffic, industrialized
100 region of Southern California.

101 **2. Materials and Methods**

102 ***2.1 Study Area***

103 Our personal exposure study was conducted in inland Southern California, better known as the
104 Inland Empire, covering an area of approximately 200 square miles (Figure S1). More specifically,
105 the study area includes the cities of Moreno Valley (2018 U.S. Census population of 209,050),
106 Redlands (71,596), Riverside (330,063), San Bernardino (215,941), and Yucaipa (53,682), CA
107 (Table S1). In 2018, median household income estimates were \$63,572, \$72,523, \$65,313,
108 \$43,136, and \$63,657; and poverty rates were 19.9%, 13.6%, 15.6%, 28.4%, and 12.3%,
109 respectively (U.S. Census Bureau). The major routes that service these cities include interstate
110 routes 10, 15, and 215, and U.S. highways 60, 66, 91, and 210. The major air pollution sources in
111 inland Southern California are on-road traffic, off-road mobile sources (e.g., railyard equipment),
112 industrial point sources (e.g., cement manufacturing and power generating facilities), and smaller
113 point sources (e.g., auto body shops, residential combustion, and restaurants). In recent years, the
114 logistics industry has expanded in the region, prompting the construction of large warehouses that
115 rely on heavy-duty vehicles for goods transport.

116 The recently implemented California Assembly Bill 617 was designed partially to address
117 disproportionate impacts of air pollution in environmental justice communities, and San
118 Bernardino was selected as a Phase 1 community in 2018 (Garcia, 2017). Previous studies have
119 highlighted health disparities in the San Bernardino community due to its proximity to a large
120 railyard (Spencer-Hwang et al., 2015, 2016). Through our study, we sought to understand personal
121 exposure patterns as they relate to the unique environmental and socioeconomic characteristics of
122 inland Southern California.

123 **2.2 Sampling Campaign**

124 For the sampling campaign, we recruited 18 adult participants (18 years and older; 61% males;
125 55% Latinx) with varied occupations (50% identified as college students). All sampling activities
126 and interactions with participants were pre-approved by the University of California, Riverside
127 Institutional Review Board (protocol number: *HS 18-206*). The overall campaign took place over
128 a five-week period from 03-10-2019 to 04-14-2019. Each week on Sunday, we distributed a PM
129 monitoring pack to four participants, except for the first week which had two participants (Figure
130 1). Participants kept the packs for a duration of seven days, allowing the assessment of inter- and
131 intra-day exposure variability for each individual.

132



133

134 **Figure 1.** (Left) Wearable particulate matter monitors from Applied Particle Technology (St. Louis, MO).
135 Data were transmitted via Wi-Fi hotspots and were accessible online in real-time. (Right) PM sampling
136 pack used in the personal exposure study. The monitors were clipped outside of the pack, and the Wi-Fi
137 and GPS data loggers were housed inside of the pack.

138 We tracked participant locations with GPS data loggers. Participants were required to carry the
139 packs during the day, and packs were placed in their bedroom or living spaces at night. After the
140 seven-day deployment, the packs were returned to our research facility. We retrieved and removed
141 the GPS data from the data loggers before the next deployment for privacy. One participant's GPS
142 data were missing, so this dataset was removed, and subsequent analyses were carried out for 17
143 datasets. The participant breakdown by city was the following: two from Moreno Valley, two from
144 Redlands, five from Riverside, six from San Bernardino, and two from Yucaipa. We recognize the
145 uncertainty introduced by the sample size and city breakdown. However, our pilot study generated
146 useful insights that will be leveraged during our larger phase two sampling campaign.

147 ***2.3 Monitoring Equipment***

148 Each monitoring pack (total = 4) included a battery-powered PM monitor, a GlobalSat-DG-500
149 (New Taipei City, Taiwan) GPS module, a Huawei Wi-Fi hotspot, Elitech temperature log, and
150 necessary accessories. The PM monitors are developed by Applied Particle Technology (APT, St.
151 Louis, Missouri, USA) and utilize the Plantower PMS optical sensor (Figure 1). The monitors are
152 commercially available, and our research team was not directly involved with monitor
153 development. The dimensions of the PM monitors are 2 in. x 1 in. x 2.25 in. (L x W x H). The
154 APT monitor provided four PM₁, PM_{2.5}, and PM₁₀ measurements per minute, but we only analyzed

155 PM_{2.5} measurements due to the extensive literature and relevance of PM_{2.5} exposure and health,
156 and due to the availability of suitable reference measurements for monitor evaluation. The APT
157 monitors also provide measurements of relative humidity and temperature, and the data are
158 uploaded in real-time via the mobile hotspot to the vendor-hosted web interface. The size,
159 simplicity, mobility, and accessibility of the APT device was ideal for community engagement.
160 The sampling rate of the PM monitor was once every 15 seconds, totaling a maximum of
161 approximately 40,320 possible measurements at the end of the seven-day sampling period, plus or
162 minus a few hours of measurements depending on the scheduled pick-up and drop-off times.

163 ***2.4 Data Processing***

164 Although a uniform usage protocol was established for the study, datasets had varying degrees of
165 availability due to the operating habits of the participants. We assigned all missing PM
166 measurements as “-9999”, then PM data were synced with the GPS data by their dates and
167 timestamps. Since the GPS data logger’s sampling rate was once every five seconds, we performed
168 a linear interpolation on the PM data from 15 to five second intervals to obtain the highest
169 resolution for our datasets. The resulting combined datasets provide the date, time of day, PM_{2.5}
170 concentrations, relative humidity, temperature, and the corresponding latitude and longitude. As a
171 note, the GPS position was intermittently measured at times because the data logger stopped
172 recording if the no movement was detected after 30 seconds. To account for the idling periods, we
173 assigned the previous latitude and longitude to the missing timestamp if the distance between the
174 two intervals was less than 20 meters (Figure S2). When the distance was greater than 20 meters
175 and less than or equal 50 meters, we performed linear interpolation between the two points. A
176 distance greater than 50 meters was assigned “NaN” and considered an invalid data point due to
177 uncertainty in participant mobility during the idle period. The five-second syncing lends a
178 maximum of approximately 120,960 possible data points for each participant.

179 ***2.5 Co-location and Adjustments***

180 We co-located the personal PM monitors at the Mira Loma Van Buren (MLVB, AQS ID:
181 060658005) air monitoring site to evaluate the hourly performance of the monitors. We housed
182 the wearable monitors in a home-built enclosure and positioned the enclosure near the site’s federal
183 equivalent method (FEM) PM_{2.5} samplers. The enclosure was built using steel mesh panels to

184 maximize the air flow over the monitors. The monitors were kept on-site for two weeks, and we
185 continuously monitored the activities of each sensor through the web server to ensure that each
186 device was operating optimally. At the end of the co-location period, we obtained PM_{2.5} reference
187 data for the performance analysis. For our study, we used polynomial fitting to adjust the raw data
188 to the FEM reference data. Our measurements were determined to be uninfluenced by relative
189 humidity and temperature, hence the polynomial fittings were solely based on two parameters:
190 reference measurements and raw measurements (Note S1). The fitting method is well described in
191 a paper by Zheng et al. (Zheng et al., 2018). We also explored one other approach to adjust the
192 raw data, for which we utilized using machine learning with random forest regression (RFR) to
193 construct a pattern-based relationship between the reference and raw data. See Note S2 for further
194 discussion of the calibration model testing.

195 **2.6 Data Analysis**

196 We classified microenvironments of all data points based on the GPS measurements. We used the
197 density-based spatial clustering of applications with noise (DBSCAN) algorithm in the QGIS
198 (<https://www.qgis.org/>) open source GIS platform, and DBSCAN clusters points based on a two-
199 dimensional implementation.(QGIS Development Team, 2019) We then defined each spatial
200 cluster by mandating a minimum size of 120 PM_{2.5}/GPS measurements within a maximum distance
201 of 0.0005 degrees (~55 meters). The clusters were manually evaluated and assigned a
202 microenvironment class and activity by overlaying the clusters onto Google Maps.
203 Microenvironment classes included home (H), work (or university, W), restaurant (R), retail (RE),
204 leisure indoor (LI), leisure outdoor (LO), and transient (T); and microenvironment was classified
205 and assigned to the cluster based on the proximity of the cluster center to labels available in Google
206 Maps. The “transient” classification indicates that the speed measurement was greater than 10
207 kilometers per hour, regardless of prior cluster classification. The “unclassified” classification was
208 given to non-clustered, non-transient data points. We make no assumptions about participant
209 mobility within the microenvironment.

210 **2.7 Ambient PM_{2.5} Contour Fields**

211 We constructed a PM_{2.5} contour mesh over Southern California to compare the personal exposure
212 of PM_{2.5} to ambient PM_{2.5}. Participant mobility varied, and measurement locations were up to 100

213 miles away from the main study location. The input data for the ambient PM_{2.5} spatial fields were
214 accessed from the regulatory monitoring network of the South Coast Air Quality Management
215 District. To construct hourly contour fields, we performed cubic interpolation on hourly PM_{2.5}
216 measurements from 18 monitoring stations. Participant coordinates were paired to the
217 corresponding contour location, resulting in corresponding ambient and personal PM_{2.5} data points
218 for all participants.

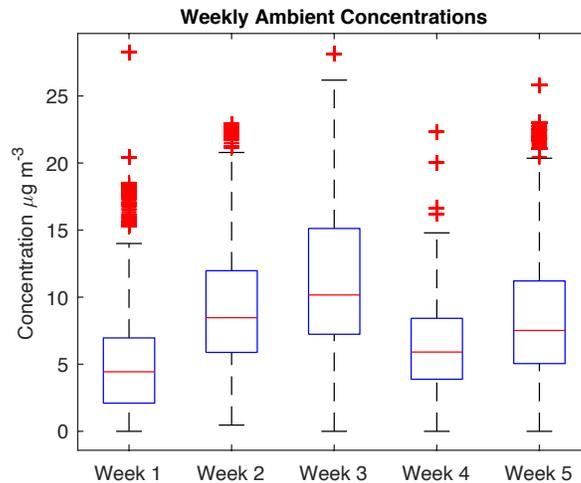
219 **3. Results**

220 *3.1 Personal and Ambient Data Overview*

221 Calibration of PM monitors using the polynomial fittings resulted in good agreement between the
222 adjusted personal measurements and reference PM_{2.5} measurements. The mean bias for the four
223 monitors ranged from -0.11 to 0.61, slopes ranged from 0.99 to 1.10, intercepts ranged from 0.012
224 to 0.75, and R² ranged from 0.41-0.45 (Note S2).

225 For interpolated personal measurements, we define data recovery as the percentage of five-second
226 data points available out of the total possible data points for each participant's sampling period
227 (range: 0.5 – 95.6%). Mean data recovery was 50.8%, corresponding to 54,120 valid data points
228 per participant; and median data recovery was 51.8%, corresponding to 53,921 valid data points
229 per participant (Table S2). In comparison to prior studies our approach was successful in collecting
230 an exceptionally large amount of data, where valid personal data points from all 17 participants
231 totaled 920,045 (Bekö et al., 2015; Li et al., 2017; Minet et al., 2018; Piedrahita et al., 2017; Quinn
232 et al., 2018; Thomas et al., 1993).

233 Personal PM_{2.5} measurements were compared to corresponding ambient PM_{2.5} measurement, and
234 ambient data were extracted from contours of hourly measurements from regulatory monitoring
235 stations (Figure S3). Median ambient PM_{2.5} concentrations for each sampling week ranged from
236 4.4 to 10.2 µg m⁻³, and maximum concentrations ranged from 22.3 to 28.2 µg m⁻³ (Figure 2).
237 Ambient PM_{2.5} concentrations are lowest in the spring season in southern California.



238

239 **Figure 2.** Distributions of ambient PM_{2.5} concentrations (µg m⁻³) corresponding to participant locations
 240 during each week of the study. Median concentrations were 4.4, 8.5, 10.2, 5.9, and 7.5, for weeks 1-5,
 241 respectively.

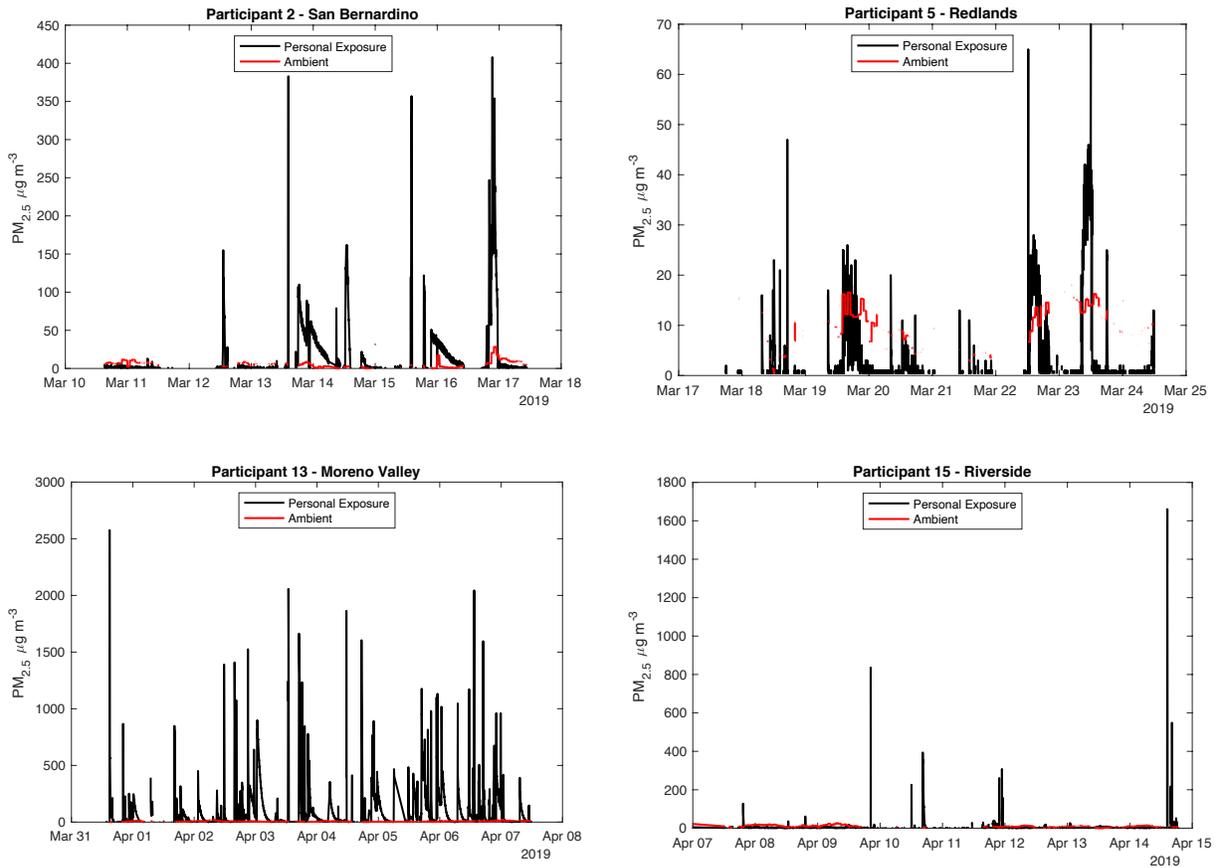
242 3.2 Exposure and Activity

243 Time series of individual personal exposure measurements identify acute PM_{2.5} exposure episodes
 244 (less than one hour, > 35 µg m⁻³), and acute exposures were highly variable for all participants.
 245 We highlight time series of consecutive, 24-hour personal measurements at 5-seconds resolution
 246 along with the corresponding ambient hourly measurement for four participants. Maximum acute
 247 exposures ranged from approximately 70 (Redlands) to 2500 (Moreno Valley) µg m⁻³, further
 248 justifying the need for individual level analysis of exposure risk (Figure 3). Participant 2 (San
 249 Bernardino) experienced the highest exposures in the home microenvironment in the late
 250 afternoons and early evening, as well as in an indoor residential microenvironment that was not
 251 classified as home. Participant 5 (Redlands) experienced all acute episodes in the work/university
 252 microenvironment, and the residential location university housing. Participant 5 exposures were
 253 not as severe as the other highlighted exposures.

254 Participant 13 (Moreno Valley) experienced frequent, extreme exposures with consistently high
 255 measurements greater than 500 µg m⁻³ in the home and leisure indoor microenvironments. High
 256 measurements were observed in short intervals in the restaurant microenvironments, specifically
 257 a popular burger and coffee chain. High measurements were also infrequently observed in the

258 transient and work microenvironments. Based on the short duration (< 10 minutes) of the extreme
259 exposures and the occurrence in the majority of microenvironments, it is suspected that the
260 participant is a smoker. Participant 15 (Riverside) experienced exposures greater than $100 \mu\text{g m}^{-3}$
261 in the home microenvironment, and consistently elevated $\text{PM}_{2.5}$ was observed during time spent
262 in a restaurant microenvironment (range 20–50 $\mu\text{g m}^{-3}$). Time series for all participants can be
263 found in Note S3 in the Supplementary Material.

264



265 **Figure 3.** Sample time series of 5-second personal (black) and hourly ambient (red) monitoring data for
266 four participants from San Bernardino (*top-left*), Redlands (*top-right*), Moreno Valley (*bottom-left*), and
267 Riverside (*bottom-right*).

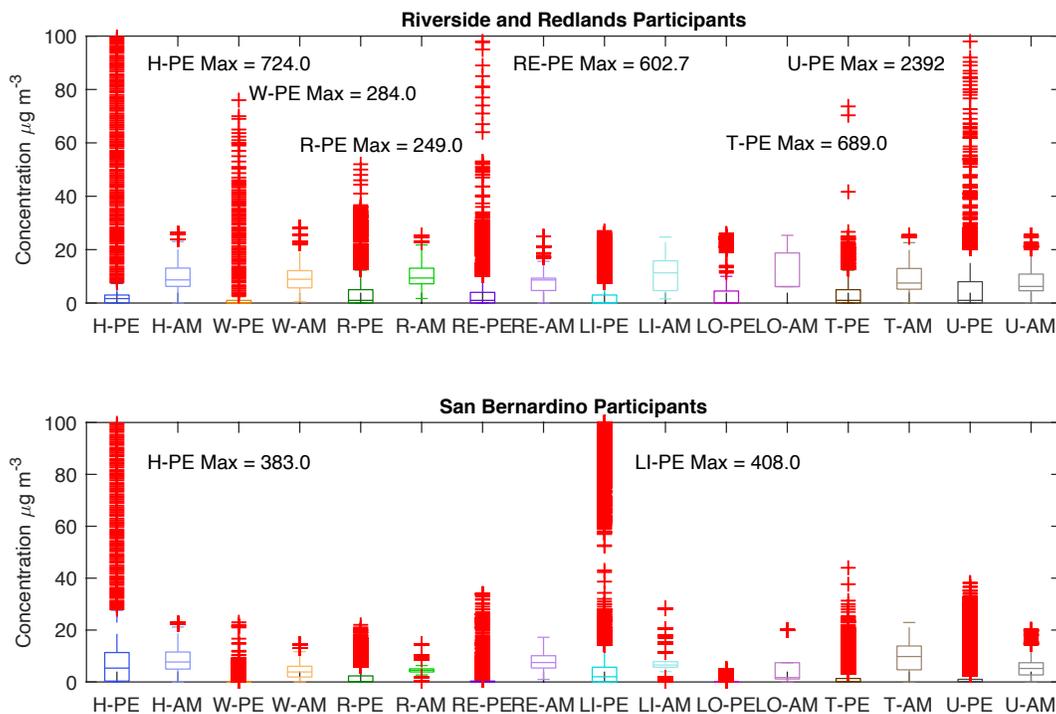
268 **3.3 Inter-City Comparative Analysis**

269 Personal and ambient PM_{2.5} data were aggregated for cities with two or more participants with
 270 50% or greater data recovery, which was the criteria for inclusion in the inter-city analysis (Table
 271 1). Results from those participants were then stratified along SES lines: Redlands/Riverside (N =
 272 5, high SES) and San Bernardino (N = 4, low SES); there were no datasets from Moreno Valley
 273 and Yucaipa that met the aggregation criteria. Average data recovery for these participants was
 274 73% (Redlands/Riverside) and 72% (San Bernardino). Aggregated median ambient concentrations
 275 were consistently higher than median personal concentrations, and the highest median personal
 276 concentrations were observed in home microenvironment for both SES groups. San Bernardino
 277 personal medians in the home microenvironment were higher despite having slightly lower
 278 ambient medians than Redlands/Riverside. Short-term personal exposures were higher than 20 µg
 279 m⁻³ in work, university, restaurant, retail, leisure indoor, and transient microenvironments for
 280 aggregated datasets (Figure 4).

281 **Table 1.** Summary of the total number of valid data points, average data recovery, and median personal
 282 (ambient) PM_{2.5} concentrations (µg m⁻³) for Redlands and Riverside (N = 5), and San Bernardino (N = 4)
 283 participants with data recovery greater than 50%.

City	Redlands and Riverside	San Bernardino
Number of Data Points (Average Data Recovery)	387,781 (73%)	302,305 (72%)
Home	1.67 (8.66)	5.33 (7.69)
Work or University	0.00 (8.91)	0.00 (3.85)
Restaurant	1.00 (9.36)	0.00 (4.50)
Retail	1.00 (8.64)	0.00 (7.48)
Leisure Indoor	0.00 (11.3)	2.00 (6.52)
Leisure Outdoor	0.00 (6.17)	0.00 (1.68)
Transient	1.00 (7.49)	0.00 (9.79)
Unclassified	1.00 (6.22)	0.00 (5.20)

284



285

286 **Figure 4.** Distributions of personal and ambient $PM_{2.5}$ measurements for Redlands and Riverside ($N = 5$),
 287 and San Bernardino ($N = 4$) participants with data recovery greater than 50%. The labels indicate the
 288 microenvironment classifications: home (H), work (or university, W), restaurant (R), retail (RE), leisure
 289 indoor (LI), leisure outdoor (LO), transient (T), and unclassified (U). Personal exposure measurements are
 290 labeled “-PE,” and ambient data are labeled as “-AM.”

291 For SES-aggregated datasets, mean personal-ambient (P-A) ratios for each microenvironment
 292 ranged from 0.02 to 3.49, and median ratios ranged from 0.00 to 0.55 (Table 2). Higher mean ratios
 293 compared to median ratios reflect the influence of the outliers in the personal measurements. Ratios
 294 less than one indicate that personal environments had lower $PM_{2.5}$ levels than those derived from
 295 ambient data. For classified microenvironment clusters, the highest mean P-A ratios were observed
 296 in the retail 1.45 (0.60, Redlands/Riverside) and home (3.49, San Bernardino) microenvironments
 297 (Table 2). Redlands/Riverside had ratios greater than one for transient (1.17) and unclassified data
 298 points (2.81), while the mean home ratio was 0.76. San Bernardino retail ratio was 2.47. The
 299 highest median P-A ratios were observed in the home microenvironments for both
 300 Redlands/Riverside (0.16) and San Bernardino (0.55) for classified clusters. Wilcoxon rank sum

301 tests indicated significant ($p < 0.05$) differences between non-outlier personal-ambient data pairs
 302 for all microenvironments and for every participant with the exception of the leisure indoor and
 303 restaurant microenvironments for Participants 5 and 8, respectively. Outlier personal data and
 304 corresponding ambient data were excluded from the Wilcoxon tests. Mean and median ratios for
 305 all participants can be found Tables S4 and S5 in the Supplementary Material.

306 **Table 2.** Mean (median) personal-ambient ratios by city of residence for Redlands and Riverside (N = 5)
 307 and San Bernardino (N = 4) participants with data recovery greater than 50%. Bold indicates higher personal
 308 PM_{2.5} concentrations than the corresponding ambient concentrations.

City	Redlands and Riverside	San Bernardino
Home	0.76 (0.16)	3.49 (0.55)
Work or University	0.30 (0.00)	0.06 (0.00)
Restaurant	0.35 (0.12)	0.48 (0.22)
Retail	1.45 (0.15)	0.09 (0.00)
Leisure Indoor	0.28 (0.00)	2.47 (0.29)
Leisure Outdoor	0.22 (0.00)	0.02 (0.00)
Transient	1.17 (0.08)	0.14 (0.00)
Unclassified	2.81 (0.21)	0.23 (0.00)

309

310 4. Discussion

311 The majority of data points were classified as home for the highlighted participants (mean: 65%,
 312 median: 69%) (Table S2). This is slightly higher, but consistent with previous personal exposure
 313 studies (Bekö et al., 2015; Hsu et al., 2020; Quinn et al., 2018). Data points were classified in these
 314 microenvironments at an average of 31% (median: 16%) of the time, therefore non-home
 315 exposures may be significant in the long-term (Table S2). Transient PM_{2.5} measurements were
 316 within range of a previous personal exposure study conducted in California (Ham et al., 2017).
 317 Microenvironment distributions of personal and ambient measurements can be found in Note S4
 318 in the Supplementary Material.

319 Calculations of time spent in each microenvironment are impacted by data recovery, and charging
320 protocols were best adhered to in the home environments near a convenient supply of electricity.
321 There were compliance issues during sampling that affected data recovery, which is common in
322 human subjects research (Chenail, 2011; Mehra, 2001). Monitor mobility and real-time data
323 transfer of PM monitors enabled the high-resolution personal sampling of our study. However,
324 data collection was impeded when component batteries drained, although a charging schedule was
325 provided but not always adhered to. At times the hardware stalled, or data transfer was limited by
326 availability of Wi-Fi signal. Participant accidents with the monitors, while rare, also interrupted
327 sampling; minor damages to the protective casings were mended before redeployment.

328 Our monitoring approach intuitively identifies participants that may be actively or passively
329 exposed to cigarette or vaping smoke, as very high personal measurements ($> 100 \mu\text{g m}^{-3}$) are
330 classified as outliers in a five-second resolution dataset (Figure 4) (Götschi et al., 2002; Koistinen
331 et al., 2004; Salmon et al., 2018; Slezakova et al., 2009). Suspected smoking events occur at
332 relatively shorter time scales throughout the day and are easily identified in the time series and
333 boxplots of personal measurements. Consequently, median P-A ratios derived from high temporal
334 resolution data are useful for evaluating non-smoking related $\text{PM}_{2.5}$ exposures when smoking
335 status is undisclosed. Therefore, when comparing the bulk (non-outliers) of personal and ambient
336 measurements for Redlands/Riverside microenvironments, personal $\text{PM}_{2.5}$ measurements are
337 much less than ambient $\text{PM}_{2.5}$. Conversely, the San Bernardino median home microenvironment
338 exposure was most similar to the corresponding median ambient exposure (Table 1).

339 Considering the relatively small number of participants in the study, definitive generalizations
340 cannot be made regarding influences of residential location. However, the large amount of
341 measurements analyzed here provides a preliminary, yet robust, investigation of exposure
342 disparities. San Bernardino (highest poverty rates, lowest median household income) participants
343 with greater than 50% data recovery experienced higher home exposures compared with
344 participants from other cities. Redlands/Riverside (second/third lowest poverty rate,
345 highest/second-highest household incomes) participants overall had lower home personal
346 exposures and experienced higher personal exposures outside of the home. Since most time was
347 spent in the home microenvironment for the majority of participants, San Bernardino participants
348 were more likely to be exposed to higher $\text{PM}_{2.5}$ concentrations, even when taking into account the

349 high degree of mobility of participants which is reflected in the diversity of classified
350 microenvironments.

351 **5. Conclusions**

352 Our pilot study highlights the variability in community-scale exposure in a socioeconomically
353 diverse air basin that is heavily burdened by air pollution. A novel spatial clustering approach was
354 applied to classify the microenvironments of more than 900,000 high temporal resolution personal
355 exposure data points. Results from the study indicate that participants from the lowest
356 socioeconomic status community experienced overall higher personal exposures over consecutive
357 24-hr monitoring periods, despite high participant mobility and low variability in ambient PM_{2.5}
358 during the study. Our inclusive monitoring protocol minimizes participant fatigue and is well-
359 suited for real-time, long-term characterization of PM_{2.5} exposure disparities in underserved
360 communities. PM_{2.5} serves as a useful surrogate species for many other air pollutants that may
361 influence disproportionate exposures. The application of our streamlined, data-driven methods in
362 a larger-scale exposure study will further elucidate personal exposure disparities along racial and
363 socioeconomic lines.

364 **Data Availability**

365 In accordance with the University of California, Riverside Institutional Review Board, personal
366 data may only be distributed in an aggregated form to preserve participant privacy. All aggregated
367 and anonymized data are summarized in the Supplementary Material.

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369

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