

# **A roadmap toward scalably quantifying field-level agricultural carbon outcome**

Kaiyu Guan<sup>†</sup>, Zhenong Jin<sup>†</sup>, Evan DeLucia, Paul West, Bin Peng, Jinyun Tang, Chongya Jiang, Sheng Wang, Taegon Kim, Wang Zhou, Tim Griffis, Licheng Liu, Ziqi Qin, Andrew Margenot, Vipin Kumar, Carl Bernacchi, Wendy H. Yang, DoKyoung Lee, Jonathan Coppess, James Gerber, Molly Jahn, Madhu Khanna, & Shang-Jen Yang

This manuscript is a non-peer reviewed preprint submitted to EarthArXiv to ensure rapid and free access to the work. Subsequent versions of the manuscript may have different content. If accepted in a peer-reviewed journal, the final manuscript will be available through the “peer-reviewed publication DOI” link on the EarthArXiv website.

Please contact the corresponding author if you have any questions or feedback about the manuscript, or if you would like early access to the dataset.

\*Corresponding to:

Kaiyu Guan (kaiyug@illinois.edu), Zhenong Jin (jinzn@umn.edu)

<sup>†</sup>Co-first authors

**Prior to publication in a peer-reviewed journal, please cite this manuscript as:**  
Kaiyu Guan<sup>†</sup>, Zhenong Jin<sup>†</sup>, Evan DeLucia, Paul West, Bin Peng, Jinyun Tang, Chongya Jiang, Sheng Wang, Taegon Kim, Wang Zhou, Tim Griffis, Licheng Liu, Ziqi Qin, Andrew Margenot, Vipin Kumar, Carl Bernacchi, Wendy H. Yang, DoKyoung Lee, Jonathan Coppess, James Gerber, Molly Jahn, Madhu Khanna, & Shang-Jen Yang (2021). A roadmap toward scalably quantifying field-level agricultural carbon outcome, EarthArXiv.

**Title: A framework for scalably quantifying field-level agricultural carbon outcomes**

Kaiyu Guan<sup>1,2,3\*†</sup>, Zhenong Jin<sup>4,5\*†</sup>, Evan DeLucia<sup>1,6</sup>, Paul West<sup>5</sup>, Bin Peng<sup>1,2</sup>, Jinyun Tang<sup>7</sup>, Chongya Jiang<sup>1,2</sup>, Sheng Wang<sup>1,2</sup>, Taegon Kim<sup>4,8</sup>, Wang Zhou<sup>1,2</sup>, Tim Griffis<sup>9</sup>, Licheng Liu<sup>4</sup>, Ziqi Qin<sup>1,2</sup>, Andrew Margenot<sup>1,10</sup>, Vipin Kumar<sup>11</sup>, Carl Bernacchi<sup>1,6,12</sup>, Wendy H. Yang<sup>1,6,13</sup>, DoKyoung Lee<sup>1,10</sup>, Jonathan Coppess<sup>14</sup>, James Gerber<sup>5</sup>, Molly Jahn<sup>15</sup>, Madhu Khanna<sup>12</sup>, & Shang-Jen Yang<sup>16</sup>

<sup>1</sup>Agroecosystem Sustainability Center, Institute for Sustainability, Energy, and Environment, University of Illinois at Urbana Champaign, Urbana, IL 61801, USA

<sup>2</sup>Department of Nature Resources and Environmental Sciences, College of Agricultural, Consumer and Environmental Sciences, University of Illinois at Urbana-Champaign, Urbana, IL 61801, USA

<sup>3</sup>National Center for Supercomputing Applications, University of Illinois at Urbana-Champaign, Urbana, IL 61801, USA

<sup>4</sup>Department of Bioproducts and Biosystems Engineering, University of Minnesota, St. Paul, MN 55108, USA

<sup>5</sup>Institute on the Environment, University of Minnesota, St. Paul, MN 55108, USA

<sup>6</sup>Department of Plant Biology, University of Illinois at Urbana-Champaign, Urbana, IL 61801, USA

<sup>7</sup>Climate Sciences Department, Lawrence Berkeley National Laboratory, Berkeley, CA\_94720, USA

<sup>8</sup>Department of Smart Farm, Jeonbuk National University, Jeonju, Jeollabuk-do 54896, Republic of Korea

<sup>9</sup>Department of Soil, Water, and Climate, University of Minnesota, St. Paul, MN 55108, USA

<sup>10</sup>Department of Crop Sciences, University of Illinois, Urbana, IL 61801, USA

<sup>11</sup>Department of Computer Science and Engineering, University of Minnesota, Minneapolis, MN 55455, USA

<sup>12</sup>USDA-ARS, Global Change and Photosynthesis Research Unit, Urbana, IL, USA

<sup>13</sup>Department of Geology, University of Illinois, Urbana, IL 61801, USA

<sup>14</sup>Department of Agricultural & Consumer Economics, University of Illinois, Urbana, IL 61801, USA

<sup>15</sup>College of Agricultural and Life Sciences, University of Wisconsin-Madison, Madison, WI 53706, USA

<sup>16</sup>Department of Agriculture and Biological Engineering Sciences, University of Illinois at Urbana-Champaign, Urbana, IL 61801, USA

\*Corresponding to: Kaiyu Guan (kaiyug@illinois.edu), Zhenong Jin (jinzn@umn.edu)

†Co-first authors

## 1. Introduction

Agriculture contributes about a quarter of global greenhouse gas (GHG) emissions, with approximately 14% directly from agricultural activities and 10% through clearing land to create new croplands and pastures<sup>1</sup>. In many countries with intensified crop production, such as the U.S., GHG emissions associated with soil and fertilizer management contribute to more than half of the total agricultural emissions<sup>2</sup>. Reducing these emissions is critical for limiting global warming to the Paris Agreement of 1.5 °C or 2.0 °C, but requires rapid adoption of multiple and coordinated solutions<sup>3-6</sup>. Certain farming practices have the potential to reduce GHG emissions and/or sequester carbon. These carbon-outcome-related practices, which strongly overlap with “conservation practices” and have recently been re-formulated as “regenerative agricultural practices” or “carbon farming practices”, include but are not limited to no-till, cover cropping, precision nitrogen (N) fertilizer management, biochar application, intercropping, etc<sup>5,7</sup>. The urgency in combating climate change and achieving sustainable development has spurred climate-pledges by individual companies to cut their carbon footprints and stimulate the growth of agricultural carbon markets to incentivize farmers to adopt these carbon-outcome-related practices. Accurate quantification of carbon emissions and carbon sequestration based on adopting various practices is the basis for emission reductions and carbon markets for industry and agriculture. However, existing scientific literature is not yet conclusive as to where, when, if and by how much these carbon-outcome-related practices might lead to genuine GHG reduction or carbon removal<sup>8-10</sup>.

Regardless of the debate over effectiveness of carbon-outcome-related practices for GHG reduction and carbon removal, agricultural carbon markets are around the corner given the strong political pushes in the European Union, the U.S., China, and other nations, as well as their real co-benefits for soil health, water quality and air quality<sup>8</sup>. It is thus more urgent than ever that the scientific community should develop a credible way to quantify the amount of carbon that is removed from the atmosphere or emissions to the atmosphere that are avoided. These estimates will form the basis for assessment of the climate mitigation potential of carbon-outcome-related practices, and perhaps more importantly, to ensure the market rewards mitigation actions fairly and accurately.

In this perspective, we argue that field-level quantification of carbon outcomes is not only fundamental to a trustworthy, transparent, and cost-effective agricultural carbon market, but also critical to any other sustainability-oriented program for ecosystem services. Existing literature has illuminated the scientific and technical issues related to the rigor of these assessments of carbon sequestration in working soils<sup>9,11</sup> but does not lay out actionable roadmaps and pathways to quantify field-level carbon outcomes. Given market pressures and pressures on the social licenses to operate, there is a temptation to move ahead as if this problem has been solved and we have in hand methods to measure carbon sequestration in working soils as an essential prerequisite for large-scale government or private investment. To close the loop, we propose a foundational framework to scalably quantify field-level

carbon-related outcome, and propose an R&D agenda that can substantiate not only agricultural carbon markets but also sustainable indicators for agroecosystem management.

## **2. Foundational framework to scalably quantify field-level carbon-related outcomes at the agroecosystem scale**

### **2.1 Criteria for a successful quantification technology for field-level carbon outcome**

Effective carbon quantification technology applied at the field level must be accurate, scalable, and cost-effective. “Field-level accuracy” is needed if individual farmers’ carbon outcomes apparent as increased sequestered carbon in working soils is rewarded in the market; it is also required for traceability of any aggregated carbon outcome in carbon footprint quantification. “Scalable” here means that the quantification solution must have an independently verified and verifiable accuracy across all possible locations; in other words, showing that a solution works well at a few demonstration sites, as many existing measurement, reporting and verification (MRV) efforts do, is not enough. Instead, true “scalability” means one method must demonstrate an acceptable accuracy of the solution at randomly selected ‘real-world’ validation sites. Due to the challenges with scalability at the individual field scale, some practitioners argue that aggregated-level accuracy is sufficient because carbon markets have buyers who mostly purchase carbon credits in bulk. We argue that aggregated-level and field-level accuracy are complementary and both important and necessary. Aggregated-level accuracy, which is almost impossible to validate, must come from field-level accuracy. Finally, for any technology, there is a tradeoff between cost and accuracy, and the desired solution should be sufficiently cost-effective to achieve the needed accuracy <sup>12</sup>.

### **2.2 A proposed framework of field-level carbon outcome quantification**

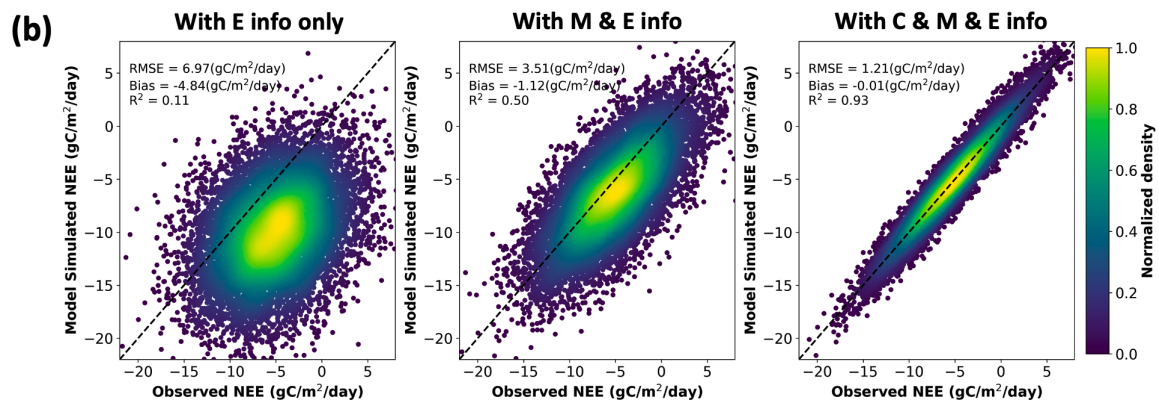
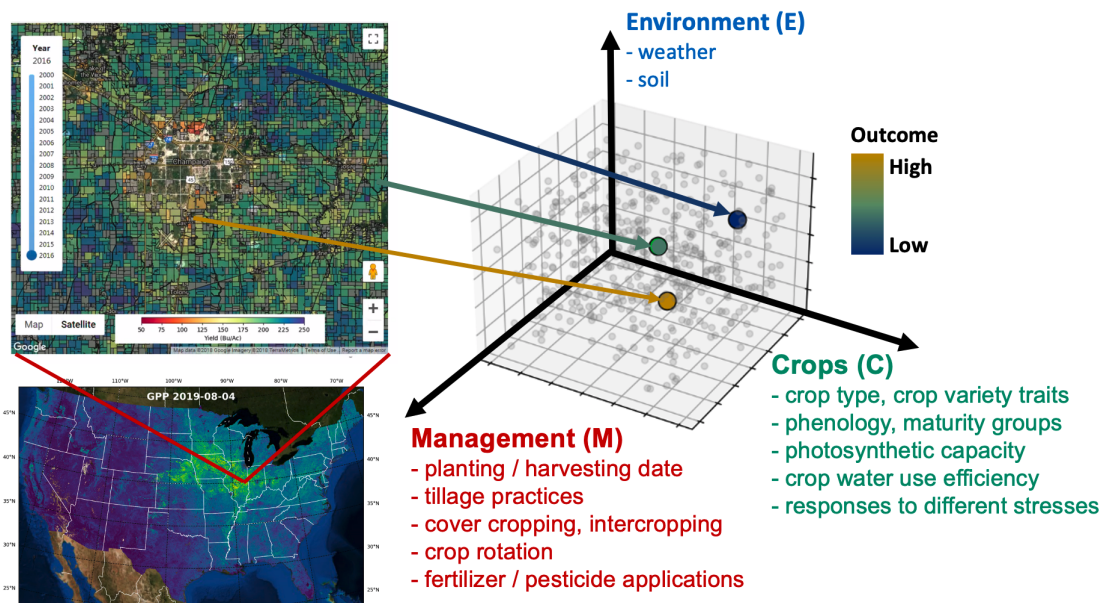
Here we propose a foundational framework for the quantification of field-level carbon-related outcomes for farmland, identify the scientific challenges in existing solutions, and discuss how to achieve scalable deployment. The foundational framework is proposed as below (Figure 1):

$$\text{Agroecosystem Outcome} = \text{Crops (C)} \times \text{Management (M)} \times \text{Environment (E)}.$$

Here, agroecosystem outcome includes crop productivity and various sustainability-related metrics (e.g. GHG emission, soil carbon sequestration, nutrient leaching etc). To calculate field-specific outcome, three dimensions of information (C, M, E) as well as their interactions (i.e. two “×” in the equation) must be well represented at the field level. Specifically, *E* primarily refers to weather and soil information, which is often available as public, gridded products. However, these datasets may contain significant uncertainty at the field level, and strategic soil sampling and local sensing may be needed to improve their accuracy. *M* primarily refers to farmers’ management practices. Since certain “actions” determine the carbon-related outcomes, both monitoring and auditing for *M* are needed. The default method to collect *M* information through farmer reporting is inefficient, error-prone, and leads to privacy concerns. Recent advancements in remote sensing and geospatial intelligence have unlocked

an opportunity to generate accurate, unbiased, and verifiable estimates for  $M$ .  $C$  refers to location-specific crop information such as crop variety and their interactions with  $M$  and  $E$ , which is manifested in pheno-stages, maturity group, photosynthetic capacity, crop water use strategy, crop responses to stresses, etc. Obtaining  $C$  information at the field-level is extremely challenging, but missing this information and especially how  $C$  interacts with  $E$  and  $M$ , can lead to the biggest uncertainties in quantifying agroecosystem carbon credits (Figure 1b). Without using  $C$  information in quantifying carbon outcome is a fundamental gap in the current modeling-based solutions. Finally, even when we have all the three types of information, the two “ $\times$ ” indicate the outcome quantification requires us to quantify the interactions among  $C$ ,  $E$ , and  $M$ . There is a rich scientific and technical literature and long history that serves as a foundation for this work, for the purposes of this paper binned as “crop models,” “ecosystem models,” or “soil biogeochemistry models” discussed below.

**(a) Agricultural Carbon Outcome = Crops (C)  $\times$  Management (M)  $\times$  Environment (E)**



**Figure 1.** Conceptual diagram of quantifying agroecosystem carbon outcomes at the field level for agroecosystems. (a) Carbon credit outcome is determined by three factors as well as their interactions. (b) Accuracy of the quantification methods improves significantly as more information is constrained at the field level; the given example is to quantify net ecosystem exchange (NEE), which is the net CO<sub>2</sub>

exchange between land and atmosphere, and the direct measurements are usually based on eddy-covariance flux tower sites in the U.S. Midwest<sup>13</sup>.

### 2.3 Issues in the existing quantification methods

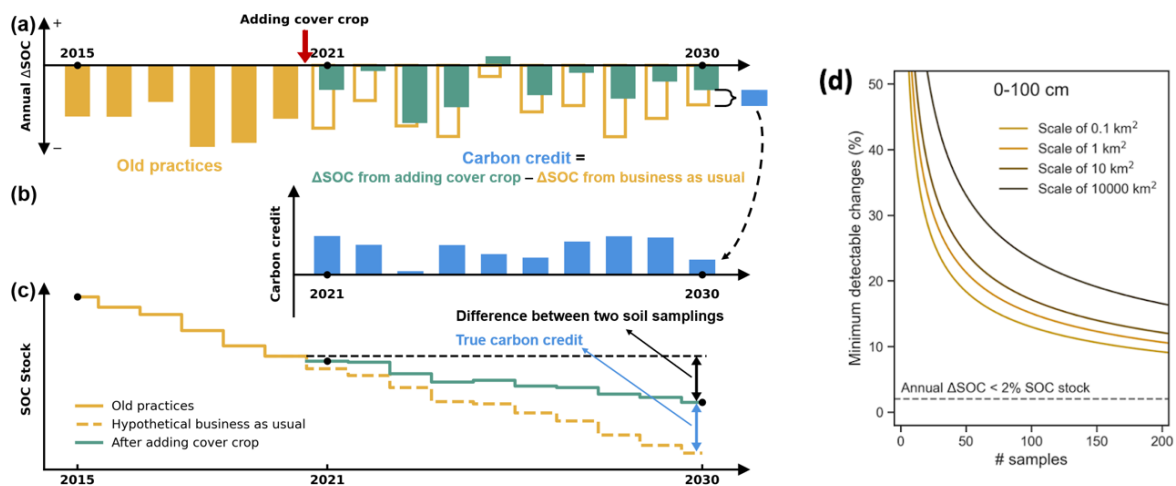
Based on the above framework, we can identify shortcomings of existing methods for carbon outcome quantification, including: (1) direct field measurements (such as soil sampling for SOC change<sup>14-16</sup>, and eddy-covariance sensors to measure GHG emissions<sup>17,18</sup>); (2) emission factor estimation, in which a fixed linear factor is used to approximate the “outcome” based on different management practices<sup>19</sup>; and (3) process-based modeling<sup>17,18, 20,21</sup>.

**Direct field measurements** have been widely viewed as the gold-standard solution for quantifying carbon outcomes, although they are in general cost-prohibitive and thus not scalable. However, direct measurements alone may not necessarily quantify the real climate benefits or genuine “carbon credit”, a fact that has not been treated with caution among practitioners and even scientists. As one example, we use a hypothetical corn-soybean rotation field in the U.S. Midwest to illustrate that soil sampling alone cannot measure the real carbon outcome of adopting cover crops with a ten-year commitment (Figure 2). In the “business-as-usual” scenario, this field is losing SOC over time as many other fields in the U.S. Midwest<sup>22</sup>. Adding cover cropping may not reverse the overall declining trend of SOC in most cases, but can slow down the rate of decline<sup>23</sup>. The cumulative difference of the  $\Delta$ SOC between the two scenarios is the real carbon benefit that the system generates in a period. Two important implications must be noted in this case. First, “additionality” requires us to know the SOC stock in the two scenarios, one with newly adopted cover cropping in which SOC stock can be directly measured, and another counterfactual scenario for “business-as-usual” in which SOC stock can no longer be measured but must be estimated through modeling. Second, because soil sampling cannot measure  $\Delta$ SOC that involves a hypothetical “business-as-usual” scenario, this “gold standard” method actually is not able to quantify the exact carbon benefits (e.g. carbon credit). Furthermore, soil sampling has its own inherent measurement uncertainties, which are found to be much larger than the detectable year-to-year changes in SOC stock<sup>24</sup> (Figure 2d), making soil sampling unfeasible as a short-term (i.e. annual) quantification tool but rather a tool to set the baseline (i.e. measure initial SOC stock) or periodic verification after 5+ years of practice changes.

**Emission factor methods**, the most widely used approaches in past IPCC reports<sup>19</sup> and also the easiest method to use, suffer from the inability to capture spatial and temporal heterogeneity of *E* and *C* and cannot comprehensively track the dynamics embedded in the interactions between *E*, *C* and *M*. The assumption of the same (or a linear scaling of) emission or sequestration outcome based on a particular “action” (*M*) across different fields is not only inaccurate, but also unfair to individual farmers participating in a carbon market.

**Process-based modeling** has been regarded as the most mechanistic method to quantify carbon outcome. Process-based models can simulate “business-as-usual” scenarios and other counterfactual

scenarios, thus this approach arguably addresses the soil sampling issues laid out above (Figure 2) and can allow direct calculation of the actual carbon benefit. However, the use of process-based modeling often suffers from “misconceptions” held by stakeholders. **First, “model denial”** stems from modeling uncertainty leading some to not believe any quantification through modeling-based approaches. We argue that models can be useful even with uncertainty, as long as they pass rigorous evaluation in a well-designed validation process (see more discussion in Section 4). **Second, “overconfidence in models”** exists among a large number of practitioners, who have unrealistic expectations that input data can be plugged in to generate reliable, unverified output <sup>25</sup>. This misuse of models could confuse the public, who in general cannot perceive modeling complexity and rely on practitioners to interpret the model output. When stakeholders realize many modeling results are contested and not inherently trustable, the “model denier” camp is reinforced. **Third, “infinite model improvement”** is common among academicians. We agree that theoretical advances in science should be ultimately incorporated into existing models to improve simulation of relevant processes, but models with more detailed mechanistic representations are not always better than simpler models in practice. Instead of devolving into a debate of “good” vs. “bad” models, we should focus on two fundamental questions: (1) Is a specific process indispensable for simulating the specific outcome and also achieving the desired accuracy? (2) Is there sufficient data to parameterize that specific process at both field and regional scales? If the answer to either question is no, then including the new process may not necessarily benefit the quantification of carbon outcome. The biggest challenges for the use of existing **process-based models** for carbon outcome quantification is the lack of spatially resolved information about C, M, and E that should be used to input and constrain model quantification, which leads to large uncertainty in the carbon outcome quantification (Figure 1b).

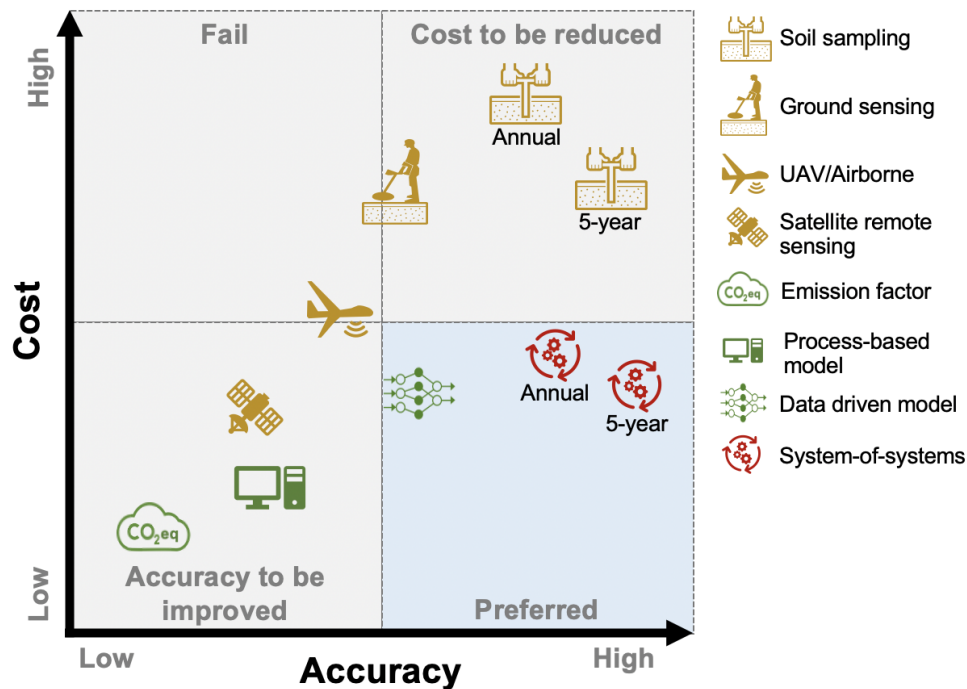


**Figure 2.** Illustration of the “additionality” concept for agricultural carbon credit, using a hypothetical corn-soybean rotation field in the U.S. Midwest as an example, assuming cover cropping is newly adopted in 2021 with a ten-year commitment. (a) Annual change in the SOC stock (i.e.  $\Delta$ SOC) since 2015, with hypothetical scenarios from 2021 to 2030. (b) Generated annual carbon credit from 2021 to

2030. (c) Change in SOC stock over time. (d) Soil sampling accuracy (i.e. minimum detectable change, in terms of relative change in the SOC stock) as a function of the number of soil samples and field sizes, which is much larger than the annual change of SOC stock in reality <sup>24</sup>.

### 3. “System-of-Systems” Solutions represent the most viable pathway

For any technology used for carbon quantification, there is a tradeoff between cost and accuracy (Figure 3). Although no clear criterion has been established so far to accept or reject a technology, we argue that for any quantification technology to be scalable, its per-acre operational cost must be meaningfully lower than expected monetized carbon values from adopting regenerative practices. In the U.S. agriculture carbon market with a carbon price of roughly US \$20/t CO<sub>2</sub>e, for example, this criterion, based on the DOE ARPA-E estimation <sup>12</sup>, means costs should be significantly lower than <\$10/acre/year for soil carbon and <\$50/acre/year for N<sub>2</sub>O quantification for large-scale deployment, including installation, calibration, operation, and hardware lifetime and at the same time, the technology should be able to achieve less than 20% error at the field level <sup>12</sup>. No single existing technology can meet both of these expectations. Instead, we argue the most viable path for quantification of field-level carbon outcome in agricultural soils is through an integration of sampling, sensing, and modeling, defined as the “system-of-systems” solution.

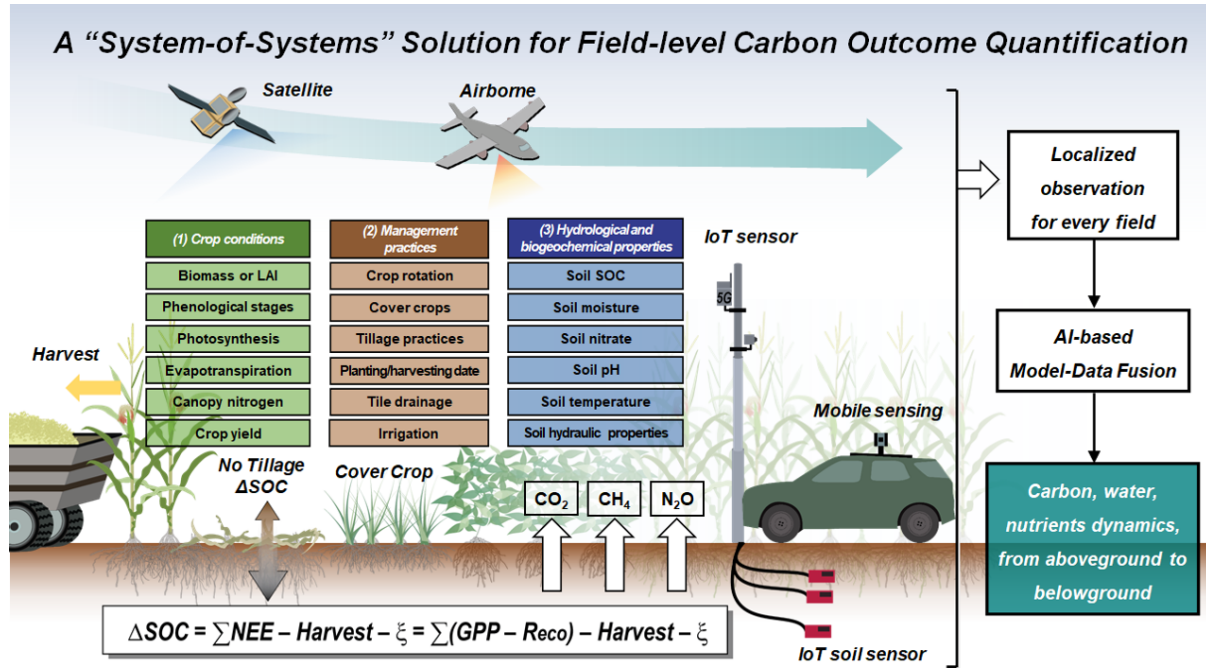


**Figure 3.** How different technological solutions for quantifying field-level carbon credit fit in the accuracy and cost diagram.

Such a “system-of-systems” solution should simultaneously provide the following three features (Figure 4): **(1)** scalable ground truth collection and cross-scale sensing of *C*, *M*, and *E* at the



local field level; **(2)** AI-assisted Model-Data Fusion, i.e. robust and efficient methods to integrate sensing data and models at each local farmland level; and **(3)** high computation efficiency to enable scaling to millions of individual fields with low cost. Thus the “system-of-systems” solution is a holistic system including multiple sub-systems for sensing, monitoring, modeling, and model-data fusion, targeting to assure field-level accuracy, scalability, and cost-effectiveness.



**Figure 4.** Illustration of a “system of systems” solution for quantifying field-level carbon outcome, including above and belowground processes. The “system of systems” solution includes sensing, monitoring, modeling, and model-data fusion, targeting to assure field-level accuracy, scalability, and cost-effectiveness.  $\xi$  represents carbon loss from leaching, which is usually very small (<0.5%) and thus can be neglected in most cases.

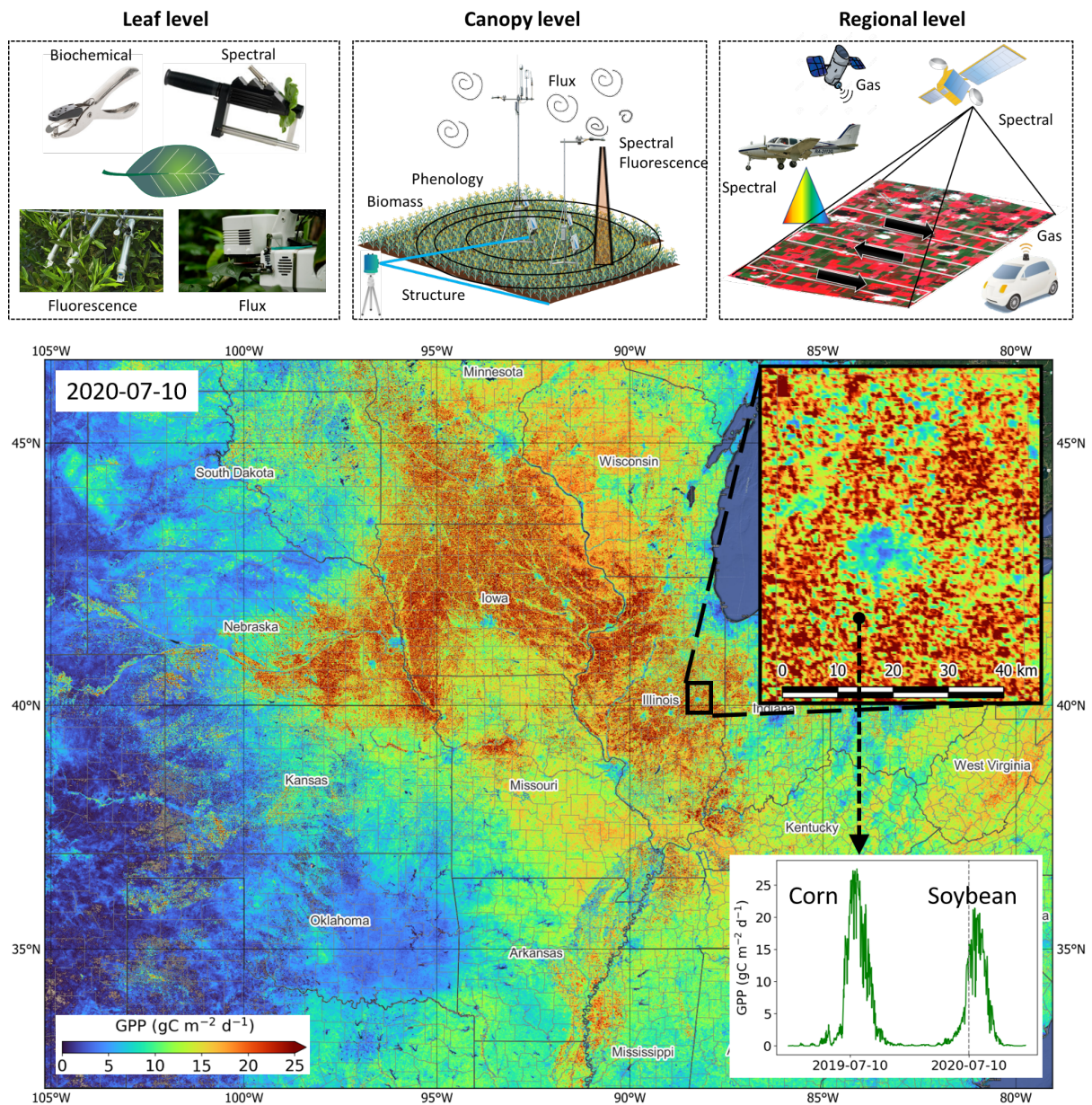
### 3.1 Scalable ground truth collection and cross-scale sensing of field-level information

Scalable sensing/estimating local information of  $C$ ,  $E$ , and  $M$  at the field level is the first step, which involves two seemingly different but inherently connected tasks: (1) ground truth collection, and (2) cross-scale sensing. Ground truth here is broadly defined as information that is collected on the ground to train, constrain and/or validate models. Agricultural ground truth is scarce and expensive to collect. For example, the gold-standard carbon flux data requires eddy-covariance flux towers (e.g. SMARTFARM Phase 1 sites), which are generally costly to set up (~\$100K needed to set up) and operate. We argue that the need for ground truth data is non-negotiable and should be a major investment with the public funding (see Section 4). However, even with low-cost sensing technology

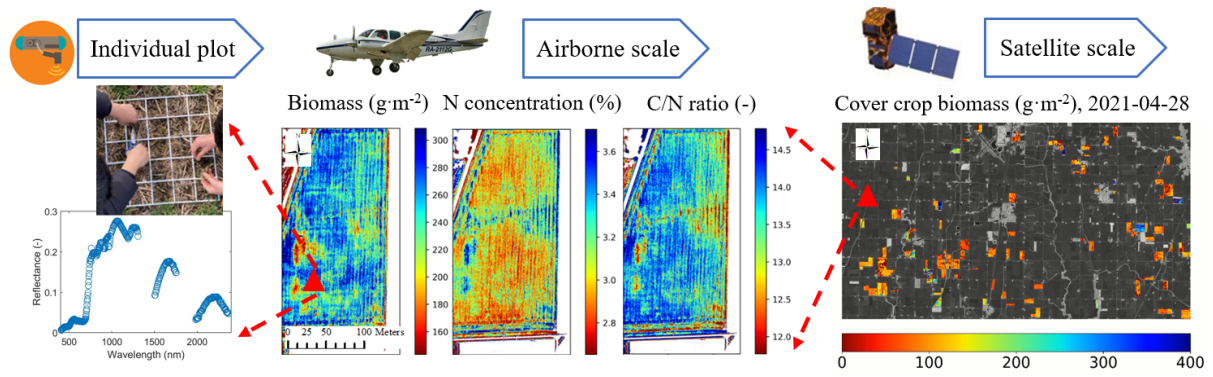
or crowdsourcing efforts, one cannot collect ground truth for every field. Instead, we will need cross-scale sensing approaches, especially those enabled by remote sensing, to scale-up “ground truth” collection to large scales.

Cross-scale sensing can be demonstrated by the most recent development of deriving field-level photosynthesis information. Photosynthesis is the only term for land carbon input and also the largest carbon budget term<sup>26</sup>. Correctly quantifying photosynthesis at the field level puts significant constraint and reduces uncertainty on simulating crop carbon dynamics, crop residues and soil carbon dynamics<sup>13,27,28</sup>. The recent breakthrough in the remote sensing of photosynthesis was made possible by full integration of leaf-level chamber/sensor measurements, canopy-level hyperspectral sensing (especially solar-induced fluorescence, SIF)<sup>29</sup>, and regional-scale mapping through satellite fusion data (Figure 5)<sup>30</sup>. The cross-scale sensing here is guided by the domain knowledge of plant physiology, radiative transfer modeling, and hyperspectral theories; the ground truth data - in particular, leaf-level samples and eddy-covariance flux tower data - are extensively used in the model development stage, but once the translation from ground-truth data to satellite-scale signals can be robustly developed, satellite fusion data can expand the photosynthesis information for every single field every day since 2000 to present<sup>31</sup>.

Another advance in cross-scale sensing is the use of intermediate sensing to augment traditional ground truth collection, and enable the scaling from leaf-level or plot-level ground measurements with coarse satellite pixel size - a classic problem in the area of remote sensing. A typical example is airborne hyperspectral imaging (AHI). Hyperspectral imaging can provide estimates of soil and plant traits with very high accuracy<sup>32</sup>, although its application for scalable mapping has been limited by high cost. The novel use of AHI is to treat AHI data as an intermediate bridge between ground truth collection and satellite scale-up. A general procedure is to first develop robust methods to translate AHI signals with targeted estimates (i.e. surface SOC, cover crop biomass) based on data from intensive lab and field experiments; and then to use AHI as a strategic sampler to selectively “sample” over space and time; and finally, to use satellite data overlaid with the AHI sampled area to translate satellite multispectral signals along with environmental variables to the plant and soil related estimation, thus deriving targeted C, M, E variables ubiquitously using satellite data. Though similar approaches have achieved impressive success in mapping forests canopy biogeochemistry<sup>33,34</sup>, they have rarely been used in agroecosystems. Once advanced and automated pipelines are established to conduct AHI collection and data processing<sup>35,36</sup>, AHI can be applied to estimate crop canopy nitrogen content, cover crop biomass, and crop residue fraction and tillage practices. Figure 6 shows a demonstration of how AHI is used as a way to scale up the estimation of cover crop adoption and biomass at the regional scale. Other sensing solutions, such as mobile vehicle sensing<sup>37</sup>, IoT sensing network and robotics<sup>38,39</sup>, could also achieve a similar function to augment ground truth collection and enable satellite scaling-up to regional scales.



**Figure 5.** Cross-scale sensing to generate photosynthesis information at the field level. (Top) The cross-scale sensing from leaf to canopy, and to regional levels for estimating photosynthesis. (Bottom) A snapshot of field-level estimation of photosynthesis on 07-10-2020, derived from the large-scale SLOPE photosynthesis data at daily frequency<sup>31</sup>, showing field-level Champaign County pattern and a field-level daily time series of photosynthesis.



**Figure 6.** Cross-scale sensing to generate regional high-resolution cover crop information. Ground truth of cover crop growth variables (aboveground biomass, nitrogen concentration and carbon/nitrogen ratio) was collected from individual ground sampling plots. Then, airborne hyperspectral imaging along with machine learning and soil-vegetation radiative transfer modeling was applied to upscale plot level measurements to airborne scale. Finally, massive airborne hyperspectral survey derived cover crop variables were integrated with STAIR multi-source satellite fusion data <sup>40</sup> to derive regional cover crop growth information.

### 3.2 AI-assisted Model-Data Fusion with efficiency and robustness at individual fields

Model-Data Fusion (MDF) here refers to a set of techniques that constrains the uncertainty of states and parameters of process-based models or fine tunes data-driven models (e.g. statistical model or neural networks) with local information (i.e. field-level C-M-E data) to generate improved estimation of “carbon outcomes”. When implemented properly, MDF can effectively reduce uncertainties in observations, model inputs, model parameters and model processes <sup>41</sup>. MDF also has the ability to evolve by incorporating new sensors/sensing data or new model developments to this framework. A drawback of conventional MDF methods, such as Bayesian Inference and Data Assimilation, is that they are computationally too expensive to run even at a few sites, making it impossible to scale to millions of individual fields.

We believe the application of AI to these challenges could significantly advance the MDF approach through multiple avenues. For example, AI can speed up Bayesian Inference methods by providing computationally cheap surrogate models (also known as emulators) for calibrating parameters <sup>41,42</sup> or parameter learning that can effectively exploit the value of all available observations <sup>43,44</sup>. When coupling physical models with neural networks, Tsai et al. (2021)<sup>45</sup> showed that parameter learning can be several orders of magnitude faster than traditional parameter calibration algorithms while obtaining physically more sensible parameter sets. Moreover, modern probabilistic distribution learning methods such as normalizing flows offer opportunities to represent more general and empirical distributions that are better-suited for describing complex systems <sup>46</sup>, and hence more accurate and efficient uncertainty

estimation. As the integration of AI and MDF become an active research area <sup>44</sup>, we expect more innovative methods will rise on the horizon.

### **3.3 High computation efficiency to enable scaling to millions of individual fields**

Scaling a system-of-systems solution to all the individual fields with similar accuracy and at a low cost is a twofold problem: (1) scalable sensing to generate rich C-M-E information for constraining various aspects of agricultural carbon cycles <sup>47</sup> (as was discussed in Section 3.1); and (2) scalable application of MDF over millions of individual fields. The latter, in particular, requires a transition from CPU-heavy to GPU-heavy models on supercomputing platforms for massive deployments. Fully upgrading existing agroecosystem models to GPU-accelerated systems would require intensive code redesign and rewrite, thus requiring longer coordinated efforts with dedicated funding support <sup>48</sup>. A more mid-term (~5-10 years) solution is hybrid modeling. This could be achieved either by replacing the most time-consuming part of the process-based model (e.g. the ODE and PDE solvers) with deep learning modules for acceleration, or by developing deep learning models to learn complex patterns from data while incorporating domain-specific knowledge, such as physical rules (e.g. mass conservation), causality (e.g. dependency structure between variables) and nature of variables (e.g. states vs fluxes), informed by process-based models <sup>44</sup>. Knowledge Guided Machine Learning (KGML) is one such hybrid modeling approach that integrates scientific knowledge embodied in process-based models with machine learning, and thus can go much beyond the black-box use of data-centric deep learning and achieve better predictive generalization across space and time <sup>49</sup>. Although the early success of KGML mainly comes from hydrology and climate science in which physical rules are better described, some recent studies have demonstrated the huge potential of this method in boosting prediction accuracy for soil and crop dynamics, such as yield and soil carbon change <sup>50</sup> and nitrous oxide emissions <sup>51</sup>. Further research could make greater impacts by developing hybrid AI models that capture previously underrepresented physics, biogeochemistry, and missing dynamics that can be generalized (e.g. from local eddy-covariance observations to larger areas with sparse data through transfer learning), as well as investigating spatiotemporally optimal management practices for agroecosystem sustainability.

## **4. Financial investment for substantiating agricultural carbon market and sustainable agroecosystems**

Looking forward, we argue that the “system-of-systems” solution will be the only viable technology for field-level carbon outcome quantification. This integrated system consists of several components that are still at their nascent stages, thus requiring considerable R&D investment by government and industry. Coincidentally, these investments will build the foundation for the next generation of precision agriculture whose scope has been expanded from site-specific management following spatial variability<sup>60</sup> to big data-driven integration of sensing, analytics and automation for

guiding farming activities at various scales<sup>61</sup>. However, technical advances alone are insufficient for substantiating the agricultural carbon market or agricultural sustainability more broadly; success will also rely on synergies among citizens, researchers, corporations, and governments to remove scientific and practical hurdles.

First and foremost, unified protocols that provide guidance on measurements and modeling schemes are essential that address standard approaches to uncertainty and bias. Such protocols must be established through community effort to achieve scientific rigor and transparency. Existing efforts led by certification organizations such as Verra<sup>52</sup> and Climate Action Reserve<sup>53</sup> are important and valued, but tend to be simplistic and conservative given the limited empirical data and insufficient MRV tools<sup>54</sup>. To successfully establish carbon markets, a concerted campaign of more advanced field work, data collection, and modeling assessment will be necessary. It is anticipated that debate will intensify as more disciplines and stakeholders become involved in the new phase of protocol development and validation, especially when the necessary rigor requires technical sophistication beyond traditional quantification approaches. To foster open and constructive conversations that increase credibility and market confidence, three principles must be emphasized. First, **the quantification uncertainty of field-level carbon outcome must be specifically emphasized, and especially for the carbon credit market the uncertainty of the calculated carbon credit should be reflected in its price** to ensure that the incentivized impact is not over- or under-compensated. For example, the standard deviation of a MRV system can be used to discount the value of credits generated<sup>55</sup>. This is an essential requirement for the protocol to be usable, not just a subjective technical preference. Second, **validation is the only way to report system-wide uncertainty**. No exemption should be made for any quantification tool, even if the tool is widely used or peer-reviewed. There are some academic-based model intercomparison MIP efforts<sup>56,57</sup> that can shed light on how to set up such validation, but given the transaction purpose of carbon credits, a high bar must be set for acceptable model performance. **Third, demonstrating performance at the scale of an individual field is obligatory**. Due to the challenges of achieving scalability, some practitioners suggest compromise by focusing on the aggregated accuracy of quantified carbon credit. We argue that aggregated accuracy, which is almost impossible to validate, must come from field-level accuracy.

Next, **establishing a gold standard dataset for developing, calibrating and validating MRV systems** is essential to building stakeholder trust in these technologies. The gold standard dataset should ensure site representativeness to include different soil, weather, crop, and management types, and be open-source but compiled under a protocol of community-wide acceptance. An analogy is the ImageNet database<sup>58</sup> for computer vision research, with which new algorithms will be benchmarked to show their progress in visual object recognition. Modeling carbon emission and sequestration, however, is more complicated, and hence more challenging to establish an “ImageNet for Agriculture”. Due to the often large uncertainty associated with agricultural measurements, protocols for standardized data collection, and processing techniques must be carefully evaluated and imposed. Some long-term experiment and

observation networks that have collected a complete suite of *C, M, E* variables have the great potential to be gold standard sites. Examples include the USDA Long-Term Agroecosystem Research (LTAR) network, some National Ecological Observatory Network (NEON) sites, and AmeriFlux sites on cropland and pastureland<sup>18</sup>. Further, the recently launched U.S. Department of Energy ARPA-E SMARTFARM sites have been collecting soil, crop, and GHG fluxes data with even greater spatial and temporal resolutions<sup>59</sup>, enabling a new generation of R&D development such as high-resolution remote sensing monitoring, or novel modeling methods that can capture granular dynamics such as hot-spot and hot-moment patterns of GHG emissions. Lastly, a large number of controlled experiment sites can be used to test the model and method scalability. These sites often have limited amounts of ground measurements but represent the real-world conditions for operational use. In all dimensions, however, **significant efforts are needed to harmonize the data that are measured by a wide range of methods and instruments**. Further investment in gold standard data collection should prioritize experiments that can help understand the carbon outcomes associated with different bundles of carbon-outcome-related practices, such as the combination of no-till and cover crop, as well as measurements that can disentangle the opaque “black box” of complex plant-soil-microbe interactions<sup>60</sup>. In addition, deep sampling of soils beyond the typical surface sampling depths (e.g. 0-30cm) is necessary to accurately quantify and monitor the extent of SOC changes<sup>61</sup> and to corroborate estimates by models.

While our discussion mainly focused on agricultural carbon outcomes, it is important to note the myriad environmental and economic co-benefits (e.g. improving soil health, reducing water and air pollution, and increasing climate resilience), which in turn can bring further benefits to carbon mitigation programs *per se*. Some recent case studies have demonstrated that, given the relatively low carbon credit price, participation of farmers may be primarily driven by these co-benefits<sup>58,59</sup>. The “system of systems” framework proposed in this perspective can be extended to assist the accounting of these co-benefits, and inform sustainable agroecosystem management by holistically studying the often coupled carbon, water, and nutrient cycles and human activities, a topic itself at the frontier of the earth system science.

#### **Acknowledgment:**

The authors acknowledge support from DOE ARPA-E SMARTFARM program and NSF Signal in Soil Program. K.G. also acknowledges the NSF CAREER Program. We thank David Tilman for helpful suggestions and comments throughout the manuscript development.

## References:

1. IPCC. Climate Change 2014: Mitigation of Climate Change. Contribution of Working Group III to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change. *Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA* (2014).
2. Clark, M. A. *et al.* Global food system emissions could preclude achieving the 1.5° and 2°C climate change targets. *Science* vol. 370 705–708 (2020).
3. Wollenberg, E. *et al.* Reducing emissions from agriculture to meet the 2 °C target. *Glob. Chang. Biol.* **22**, 3859–3864 (2016).
4. Bossio, D. A. *et al.* The role of soil carbon in natural climate solutions. *Nature Sustainability* **3**, 391–398 (2020).
5. Fargione, J. E. *et al.* Natural climate solutions for the United States. *Sci Adv* **4**, eaat1869 (2018).
6. Searchinger, T. *et al.* *Creating a sustainable food future: A menu of solutions to feed nearly 10 billion people by 2050. Final report.* (WRI, 2019).
7. Paustian, K. *et al.* Climate-smart soils. *Nature* **532**, 49–57 (2016).
8. Bradford, M. A. *et al.* Soil carbon science for policy and practice. *Nature Sustainability* **2**, 1070–1072 (2019).
9. Smith, P. *et al.* How to measure, report and verify soil carbon change to realize the potential of soil carbon sequestration for atmospheric greenhouse gas removal. *Glob. Chang. Biol.* **26**, 219–241 (2020).
10. Ranganathan, J., Waite, R., Searchinger, T. & Zions, J. Regenerative agriculture: Good for soil health, but limited potential to mitigate climate change. (2020).
11. Paustian, K. *et al.* Quantifying carbon for agricultural soil management: from the current status toward a global soil information system. *Carbon Management* **10**, 567–587 (2019).
12. DOE ARPA-E: DE-FOA-0002250: *Systems For Monitoring And Analytics For Renewable Transportation Fuels From Agricultural Resources And Management (Smartfarm)* <https://arpa-e-foa.energy.gov/Default.aspx?foaId=e7465e6e-cb9c-455c-9625-4c3db4386f00>. (2020).
13. Zhou, W. *et al.* Quantifying carbon budget, crop yields and their responses to environmental



- variability using the ecosys model for U.S. Midwestern agroecosystems. *Agric. For. Meteorol.* **307**, 108521 (2021).
14. Norman, A. G. & Allison, L. E. Organic Carbon. in *Methods of Soil Analysis. Part 2. Chemical and Microbiological Properties* (American Society of Agronomy, Soil Science Society of America, 1965).
  15. Smith, P. Monitoring and verification of soil carbon changes under Article 3.4 of the Kyoto Protocol. *Soil Use Manage.* **20**, 264–270 (2006).
  16. Wendt, J. W. & Hauser, S. An equivalent soil mass procedure for monitoring soil organic carbon in multiple soil layers. *Eur. J. Soil Sci.* **64**, 58–65 (2013).
  17. Baldocchi, D. D., Hincks, B. B. & Meyers, T. P. Measuring Biosphere-Atmosphere Exchanges of Biologically Related Gases with Micrometeorological Methods. *Ecology* vol. 69 1331–1340 (1988).
  18. Baldocchi, D. *et al.* FLUXNET: A New Tool to Study the Temporal and Spatial Variability of Ecosystem-Scale Carbon Dioxide, Water Vapor, and Energy Flux Densities. *Bull. Am. Meteorol. Soc.* **82**, 2415–2434 (2001).
  19. IPCC. *2019 Refinement to the 2006 IPCC Guidelines for National Greenhouse Gas Inventories.* (2019).
  20. Ogle, S. M. *et al.* Scale and uncertainty in modeled soil organic carbon stock changes for US croplands using a process-based model. *Glob. Chang. Biol.* **16**, 810–822 (2010).
  21. Sándor, R. *et al.* The use of biogeochemical models to evaluate mitigation of greenhouse gas emissions from managed grasslands. *Sci. Total Environ.* **642**, 292–306 (2018).
  22. Thaler, E. A., Larsen, I. J. & Yu, Q. The extent of soil loss across the US Corn Belt. *Proceedings of the National Academy of Sciences* vol. 118 e1922375118 (2021).
  23. Qin, Z. *et al.* Assessing the impacts of cover crops on maize and soybean yield in the U.S. Midwestern agroecosystems. *Field Crops Res.* **273**, 108264 (2021).
  24. Maillard, É., McConkey, B. G. & Angers, D. A. Increased uncertainty in soil carbon stock measurement with spatial scale and sampling profile depth in world grasslands: A systematic analysis. *Agric. Ecosyst. Environ.* **236**, 268–276 (2017).

25. McNunn, G. *et al.* Climate smart agriculture opportunities for mitigating soil greenhouse gas emissions across the U.S. Corn-Belt. *Journal of Cleaner Production* vol. 268 122240 (2020).
26. Beer, C. *et al.* Terrestrial gross carbon dioxide uptake: global distribution and covariation with climate. *Science* **329**, 834–838 (2010).
27. Li, X., Xiao, J., Fisher, J. B. & Baldocchi, D. D. ECOSTRESS estimates gross primary production with fine spatial resolution for different times of day from the International Space Station. *Remote Sensing of Environment* vol. 258 112360 (2021).
28. Peng, B. *et al.* Improving maize growth processes in the community land model: Implementation and evaluation. *Agric. For. Meteorol.* **250-251**, 64–89 (2018).
29. Kimm, H. *et al.* A physiological signal derived from sun-induced chlorophyll fluorescence quantifies crop physiological response to environmental stresses in the US Corn Belt. *Environ. Res. Lett.* **16**, 124051 (2021).
30. Porcar-Castell, A. *et al.* Chlorophyll a fluorescence illuminates a path connecting plant molecular biology to Earth-system science. *Nat Plants* **7**, 998–1009 (2021).
31. Jiang, C., Guan, K., Wu, G., Peng, B. & Wang, S. A daily, 250 m and real-time gross primary productivity product (2000–present) covering the contiguous United States. *Earth System Science Data* vol. 13 281–298 (2021).
32. Wang, S., Guan, K., Zhang, K., Lee, D., Margenot, A., Ge, Y., Peng, J., Zhou, W., Zhou, Q., Huang, Y. Using soil library hyperspectral reflectance and machine learning to predict soil organic carbon: assessing potential of airborne and spaceborne optical soil sensing. *Remote Sens. Environ.* (2022).
33. Asner, G. P. *et al.* Progressive forest canopy water loss during the 2012–2015 California drought. *Proceedings of the National Academy of Sciences* vol. 113 E249–E255 (2016).
34. Asner, G. P. *et al.* Airborne laser-guided imaging spectroscopy to map forest trait diversity and guide conservation. *Science* **355**, 385–389 (2017).
35. Wang, S. *et al.* Airborne hyperspectral imaging of nitrogen deficiency on crop traits and yield of maize by machine learning and radiative transfer modeling. *Int. J. Appl. Earth Obs. Geoinf.* **105**, 102617 (2021).

36. Zhou Q, Wang S, Liu N, Townsend P, Jiang C, Bin P, Guan K. High-performance atmospheric correction of airborne hyperspectral imaging spectroscopy: model intercomparison, parameter retrieval, and machine learning surrogates. (*Under review*).
37. Yan, Y. & Ryu, Y. Exploring Google Street View with deep learning for crop type mapping. *ISPRS J. Photogramm. Remote Sens.* **171**, 278–296 (2021).
38. Tzounis, A., Katsoulas, N., Bartzanas, T. & Kittas, C. Internet of Things in agriculture, recent advances and future challenges. *Biosystems Eng.* **164**, 31–48 (2017).
39. Elijah, O., Rahman, T. A., Orikumhi, I., Leow, C. Y. & Hindia, M. N. An Overview of Internet of Things (IoT) and Data Analytics in Agriculture: Benefits and Challenges. *IEEE Internet of Things Journal* **5**, 3758–3773 (2018).
40. Luo, Y., Guan, K. & Peng, J. STAIR: A generic and fully-automated method to fuse multiple sources of optical satellite data to generate a high-resolution, daily and cloud-/gap-free surface reflectance product. *Remote Sens. Environ.* **214**, 87–99 (2018).
41. Fer, I. *et al.* Linking big models to big data: efficient ecosystem model calibration through Bayesian model emulation. *Biogeosci. Discuss.* 1–30 (2018).
42. Brajard, J., Carrassi, A., Bocquet, M. & Bertino, L. Combining data assimilation and machine learning to infer unresolved scale parametrization. *Philos. Trans. A Math. Phys. Eng. Sci.* **379**, 20200086 (2021).
43. Schneider, T., Lan, S., Stuart, A. & Teixeira, J. Earth system modeling 2.0: A blueprint for models that learn from observations and targeted high-resolution simulations. *Geophys. Res. Lett.* **44**, 12,396–12,417 (2017).
44. Reichstein, M. *et al.* Deep learning and process understanding for data-driven Earth system science. *Nature* **566**, 195–204 (2019).
45. Tsai, W.-P. *et al.* From calibration to parameter learning: Harnessing the scaling effects of big data in geoscientific modeling. *Nat. Commun.* **12**, 5988 (2021).
46. Kobyzev, I., Prince, S. J. D. & Brubaker, M. A. Normalizing Flows: An Introduction and Review of Current Methods. *IEEE Trans. Pattern Anal. Mach. Intell.* **43**, 3964–3979 (2021).
47. Peng, B. *et al.* Towards a multiscale crop modelling framework for climate change adaptation

- assessment. *Nat Plants* **6**, 338–348 (2020).
48. Bauer, P. *et al.* The digital revolution of Earth-system science. *Nature Computational Science* **1**, 104–113 (2021).
  49. Willard, J., Jia, X., Xu, S., Steinbach, M. & Kumar, V. Integrating physics-based modeling with machine learning: A survey. *arXiv preprint arXiv:2003.04919* **1**, 1–34 (2020).
  50. Wang Zhou, Licheng Liu, Kaiyu Guan, Zhenong Jin, Bin Peng, Chongya Jiang, Jinyun Tang, Sheng Wang, Robert Grant, Symon Mezbahuddin, Xiaowei Jia, Shaoming Xu, Vipin Kumar. Knowledge-based artificial intelligence for agroecosystem carbon budget and crop yield estimation. in (2021).
  51. Liu, L. *et al.* KGML-ag: A Modeling Framework of Knowledge-Guided Machine Learning to Simulate Agroecosystems: A Case Study of Estimating N<sub>2</sub>O Emission using Data from Mesocosm Experiments. doi:10.5194/gmd-2021-317.
  52. Verra. *VM0042 methodology for improved agricultural land management, v1.0*. <https://verra.org/methodology/vm0042-methodology-for-improved-agricultural-land-management-v1-0/> (2020).
  53. Climate Action Reserve. *Requirements and Guidance for Model Calibration, Validation, Uncertainty, and Verification For Soil Enrichment Projects Version 1.0a*. <https://www.climateactionreserve.org/wp-content/uploads/2020/10/SEP-Model-Requirements-and-Guidance-v1.0a.pdf> (2020).
  54. Oldfield, E.E., A.J. Eagle, R.L. Rubin, J. Rudek, J. Sanderman, D.R. Gordon. Agricultural soil carbon credits: Making sense of protocols for carbon sequestration and net greenhouse gas removals. *Environmental Defense Fund, New York, New York*. (2021).
  55. Kim, M.-K. & McCarl, B. A. Uncertainty discounting for land-based carbon sequestration. *J. Appl. Agric. Econ.* **41**, 1–11 (2009).
  56. Rosenzweig, C. *et al.* The Agricultural Model Intercomparison and Improvement Project (AgMIP): Protocols and pilot studies. *Agric. For. Meteorol.* **170**, 166–182 (2013).
  57. Eyring, V. *et al.* Overview of the Coupled Model Intercomparison Project Phase 6 (CMIP6) experimental design and organization. *Geoscientific Model Development* vol. 9 1937–1958 (2016).

58. Deng, J. *et al.* ImageNet: A large-scale hierarchical image database. in *2009 IEEE Conference on Computer Vision and Pattern Recognition* 248–255 (ieeexplore.ieee.org, 2009).
59. ARPA-E. ARPA-E Announces \$20 Million in Funding to Develop Feedstock Monitoring and Carbon Storage Technology. *ARPA-E website, News & Media, Press Releases* (2019).
60. Yan, M., Pan, G., Lavallee, J. M. & Conant, R. T. Rethinking sources of nitrogen to cereal crops. *Glob. Chang. Biol.* **26**, 191–199 (2020).
61. Tautges, N. E. *et al.* Deep soil inventories reveal that impacts of cover crops and compost on soil carbon sequestration differ in surface and subsurface soils. *Glob. Chang. Biol.* **25**, 3753–3766 (2019).