

WILDFIRE SMOKE EXPOSURE WORSENS LEARNING OUTCOMES

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ABSTRACT

Wildfires have increased in frequency and severity over the past two decades, threatening to undo substantial air quality improvements. We investigate the effect of wildfire smoke exposure on learning outcomes across the US using standardized test scores from 2009-2016 for nearly 11,700 school districts and satellite-derived estimates of daily smoke exposure. Relative to a school year with no smoke, average cumulative smoke-attributable $PM_{2.5}$ exposure during the school year ($\sim 35 \mu g/m^3$) reduces test scores by $\sim 0.15\%$ of a standard deviation. These impacts are more pronounced among younger students and are observed across differing levels of economic disadvantage and racial-ethnic composition. Additionally, we project that smoke $PM_{2.5}$ exposure in 2016 reduced discounted future earnings by nearly \$1.7 billion (\$111 per student). Roughly 80% of these costs are borne by disadvantaged districts. Our findings quantify a previously unaccounted for social cost of wildfire that is likely to worsen under a warming climate.

1 THE frequency and severity of wildfires throughout the West-
2 tern United States have increased in recent decades and are
3 expected to worsen as the climate continues to warm [1]. Litera-
4 ture has linked these wildfire events and the smoke they generate
5 to a variety of social and economic impacts, in particular health
6 and infrastructure related damages [2, 3, 4]. Yet emerging evi-
7 dence from studies on non-wildfire air pollutants suggests that
8 wildfire smoke exposure could have much wider-ranging im-
9 pacts, including possible negative effects on human cognition
10 [5, 6, 7, 8, 9, 10, 11, 12, 13, 14, 15, 16, 17, 18, 19, 20]. Such
11 effects would have broader implications for human capital for-
12 mation and longer-term economic well-being, as well as for the
13 social costs of a warming climate, but have not been documented
14 in existing literature.

15 Recent studies have focused on the biological channels
16 through which air pollution exposure might affect human health
17 and have found that non-wildfire-related air pollution exposure
18 is associated with higher likelihood of neuroinflammation [5, 6],
19 and document significant associations between $PM_{2.5}$ exposure
20 and increased risks for Alzheimer’s, dementia, and Parkinson’s
21 disease [21, 22]. Epidemiological and social science studies
22 have begun to draw links between air pollution exposure and
23 cognitive performance on real-world tasks, including declining
24 performance in chess tournaments [8], stock trading [9], call
25 center productivity [10], umpire decisions [11], cognitive as-
26 sessments [15, 17, 18, 19, 23], and online brain games [12].
27 Similar to our setting, a handful of studies have assessed how
28 student test scores have responded to variation in exposure to
29 ambient (non-wildfire) air pollution [13, 16, 20, 7, 24]. A recent
30 study investigates the association between long-term ambient air
31 pollution exposure and student test performance in the US and
32 find negative impacts of increased ambient air pollution [24].
33 Other studies have used research designs that more plausibly
34 isolate variation in pollution from other factors that might also
35 affect student test performance. These studies have found that
36 short-term changes in air pollution on the day of the exam led to
37 declines in student performance [13, 16, 20, 7] and decreased
38 future earnings [13] and that installation of air filters in schools
39 in the vicinity of a gas leak improved student Math and English
40 test scores even into the following year [25]. While these studies
41 focus on the impact of air pollution on student test performance,
42 to our knowledge there are no studies that focus on wildfire
43 smoke particulate matter, which recent studies suggest could
44 potentially be more harmful to human health than other sources
45 of particulate matter [26, 27] and is likely the fastest growing
46 source of air pollution in the US [28].

47 As wildfire activity has dramatically increased in recent
48 decades due to a rapidly warming climate and a century of fire
49 suppression practices across the Western United States, wild-
50 fire smoke has become an increasingly important contributor
51 to surface particulate matter $<2.5\mu g/m^3$ ($PM_{2.5}$) concentrations
52 [28]. Increasing wildfire-derived $PM_{2.5}$ threatens to undermine
53 decades of progress in improving overall $PM_{2.5}$ concentrations
54 – improvements brought about by changes in manufacturing
55 practices, energy production, and legislation [29, 30, 28, 31].
56 Furthermore, while exposure to ambient smoke-derived $PM_{2.5}$
57 appears more evenly distributed across economic and racial-
58 ethnic groups than other sources of $PM_{2.5}$ [28], even similar am-
59 bient exposures may differentially impact communities due to a
60 variety of factors including differences in housing or school char-

acteristics [32, 33] or differences in knowledge of or ability to
undertake protective behaviors. Ultimately, the differences in re-
alized exposures could result in differential impacts across racial-
ethnic and socioeconomic groups, as has now been documented
for other environmental exposures [34, 35, 36, 37, 38, 39].

Here we quantify the impact of wildfire smoke exposure on
learning outcomes across the US, as measured by standardized
test scores, and estimate potential heterogeneous impacts of this
exposure across demographic and socioeconomic groups. We
first develop estimates of local-level wildfire-smoke-attributable
 $PM_{2.5}$ exposures across the US and over time, using a combina-
tion of high resolution predicted $PM_{2.5}$ data and satellite derived
wildfire smoke plumes [40, 41] (Methods). We then study the
effect of cumulative smoke exposure during the school year on
student learning outcomes, as measured through harmonized
national test score data for students from 3rd-8th grades col-
lected across nearly ~11,700 school districts between 2009-2016.
These comprehensive longitudinal data allow us to plausibly iso-
late the effect of wildfire-smoke-attributable $PM_{2.5}$ on student
learning outcomes.

We model the effect of smoke exposure on student test per-
formance using fixed-effects regression models that account for
unobserved time-invariant differences in smoke exposure and
test scores across districts as well as time-trending year-grade
specific differences common to all locations (Methods). As there
has been an upward trend in both wildfire smoke exposure and
test performance across our study period as well as large regional
differences in average smoke exposure (Figure 1), simple cross
sectional or time series regressions could conflate overall trends
or average differences in smoke with other factors that affect
learning outcomes. Rather than comparing across districts, our
approach compares particular districts to themselves over time as
smoke exposure fluctuates, after accounting for any differences
in grade-specific national averages between years. We incorpo-
rate gridded temperature and precipitation data [42] to flexibly
control for temperature and precipitation-related impacts on stu-
dent performance, especially as temperature has been shown
to affect learning and may be correlated with wildfire activity
[34, 35].

We then examine the heterogeneous impacts of wildfire smoke
exposure by estimating whether responses differ between school
and non-school days or by student age groups, levels of eco-
nomic disadvantage, and/or race and ethnicity – dimensions
along which earlier research has suggested environmental ex-
posures and impacts might differ. Finally, to quantify the eco-
nomic magnitude of smoke-related impacts, we explore how
learning outcomes differ between a less severe compared to a
more severe smoke year and provide estimates of the impact
of wildfire-smoke-attributable $PM_{2.5}$ in terms of students’ lost
future earnings, using literature-derived estimates of the rela-
tionship between test scores and earnings (Methods).

RESULTS

We find that smoke exposure in the year leading up to the
test has a statistically significant negative impact on learning
outcomes (Figure 2). An additional $\mu g/m^3$ of cumulative smoke
 $PM_{2.5}$ in the year leading up to the exam decreases average test
scores by 0.003% (95% confidence interval (CI): -0.005% to

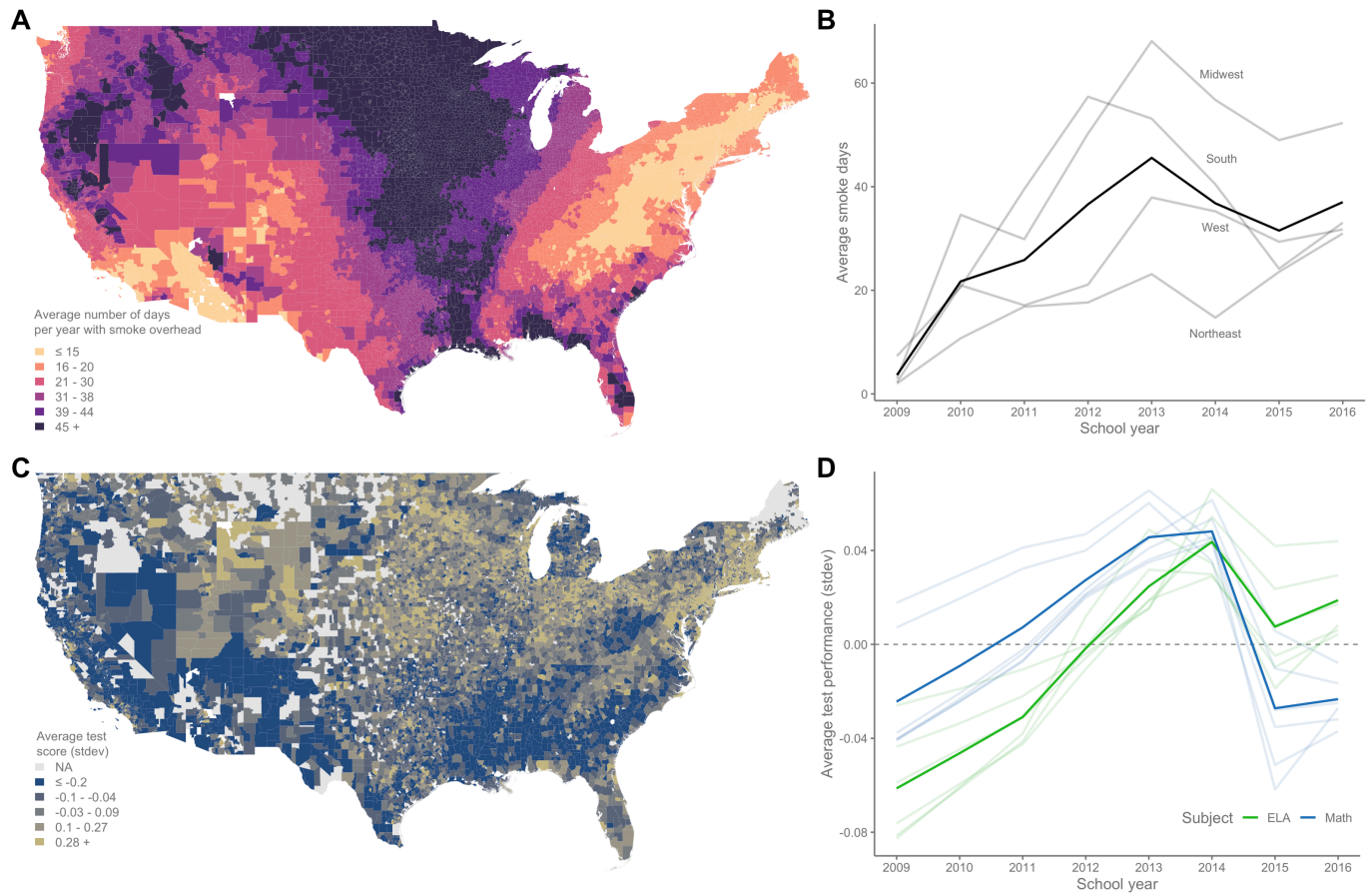


Figure 1: Spatiotemporal variation in wildfire smoke exposure and average test scores. **A.** Spatial distribution of the average number of days with smoke overhead from 2009-2016 for each school district in the continental US. **B.** Temporal variation in the average number of smoke days for various census regions. Black line represents the average over the entire US. **C.** Spatial distribution of test scores. Math and ELA scores are averaged across the study period from 2009-2016 for each district and are represented in standard deviations. **D.** Average test performance relative to the national reference cohort. Each state’s standardized test results are scaled to the nationally comparable (National Assessment of Educational Progress) test. Faded lines represent grade specific performance and darker lines represent the average over all grades.

118 -0.001%). The effect is similar across subjects with decreases
 119 in English language arts (ELA) test score by 0.003% (95% CI:
 120 -0.005% to -0.002%) and math test scores by 0.002% (95%
 121 CI: -0.004% to -0.001%) of a standard deviation for school
 122 districts across the US from 2009-2016 (Figure 2B). These
 123 results are robust to flexible functional forms such as higher-
 124 order polynomials of the smoke $PM_{2.5}$ response relationship and
 125 are fairly linear (Figure 2A).

126 Comparing school day vs. non-school day exposure, we find
 127 that smoke exposure on school days has a statistically significant
 128 negative effect on test performance where an additional $\mu g/m^3$
 129 of cumulative smoke $PM_{2.5}$ on school days decreases average test
 130 scores by 0.004% (95% CI: -0.008% to -0.001%) of a standard
 131 deviation. Exposure on non-school days results in a smaller
 132 negative effect that is also statistically significant (compared to
 133 no effect). Point estimates suggest the effect of exposure on
 134 a school day is nearly twice as harmful as exposure on a non-
 135 school day for both ELA and math, although these estimates are
 136 not statistically distinguishable from one another (Wald test for
 137 equivalence of coefficients: $F_{1,5092}=0.614, P=0.433$).

138 On school days with smoke in the air, the average ambient

smoke-attributable $PM_{2.5}$ concentration is $6 \mu g/m^3$. Given these
 139 averages, we estimate that exposure to an additional school week
 140 (five school days) of smoke in the year prior to the exam lowered
 141 scores by 0.131% (95% CI: -0.245% to -0.017%) of a standard
 142 deviation. We focus on cumulative smoke $PM_{2.5}$ exposure on
 143 school days in the year prior to the exam for the remainder of
 144 the analysis because our measurement of exposure is at school
 145 locations and exposure on non-school days is uncertain.
 146

147 While our main analysis clusters standard errors at the county
 148 level to account for correlation in errors across districts within
 149 the same county, we conduct additional analysis using a ran-
 150 domization inference approach to test the sharp null hypothesis
 151 of no effect for additional smoke $PM_{2.5}$ exposure on school
 152 days by randomly permuting test scores across districts within a
 153 county. This approach non-parametrically estimates statistical
 154 significance and is especially beneficial in the presence of fuzzy
 155 clustering where smoke exposure may be correlated within the
 156 cluster, but the correlation is imperfect [43, 44]. We find that
 157 the estimated effect of school day smoke $PM_{2.5}$ exposure is
 158 significantly different from the distribution of permuted effect
 159 estimates (Figure 2C).

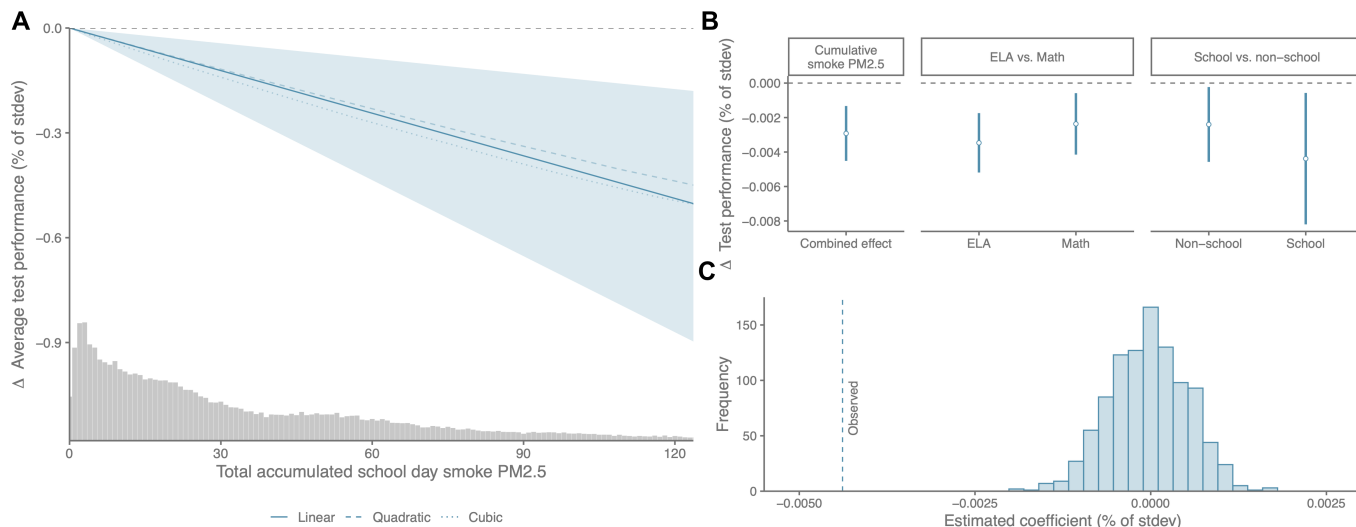


Figure 2: Effect of wildfire smoke exposure on student test scores. **A.** Test performance declines as a function of total accumulated daily smoke $PM_{2.5}$ during the school year prior to the test (only on school days); shaded area shows bootstrapped 95% confidence intervals. Δ Test performance is the change in test score relative to the national NAEP reference cohort, measured in percent of a standard deviation. **B.** Effect estimates of an additional $\mu g/m^3$ of smoke $PM_{2.5}$ in the year prior to the exam for school versus non-school day exposure, the combined average effect, and English language arts (ELA) and Math. **C.** Randomization inference test (1000 permutations) showing the estimated effect size of an additional $\mu g/m^3$ of smoke $PM_{2.5}$ on school days when outcomes are randomly permuted across districts within each county. The observed effects are significantly different from the randomization test effects.

160 Although we do not have access to evacuation orders related
 161 to each fire, we test whether the identified effects are driven by
 162 smoke $PM_{2.5}$ exposure or by direct wildfire effects by using a
 163 similar empirical strategy as before but drop districts that are
 164 certain distances from the nearest fire perimeter provided by the
 165 National Interagency Fire Center (NIFC) in that year (Methods).
 166 We find that the identified effect estimates remain fairly stable up
 167 to 6.2-miles (10-kilometers), which provides evidence that the
 168 effects identified are likely not driven by direct wildfire effects
 169 but rather by the smoke $PM_{2.5}$ impacts aligning with the research
 170 motivation of this paper (Supplemental Figure 3). We also test
 171 whether the effects of smoke exposure during previous school
 172 years carry over into test performance in the current year. While
 173 results are somewhat noisy, point estimates suggest that learning
 174 impacts can persist into future years (Supplemental Table 4).

175 We find strong evidence of heterogeneous effects of smoke
 176 $PM_{2.5}$ among sub-populations. In line with previous studies
 177 that find negative effects of air pollution exposure on younger
 178 children [45, 7, 46], we find that additional smoke $PM_{2.5}$ on
 179 school days is statistically significant and negative for primary
 180 school (3rd-5th grades) but does not appear to affect secondary
 181 school students (6th-8th grade) (Figure 3). Among primary
 182 school children, each additional $\mu g/m^3$ of cumulative smoke
 183 $PM_{2.5}$ on school days decreases scores by 0.012% (95% CI:
 184 -0.019% to -0.005%) of a standard deviation.

185 Consistent with previous work [28, 47], we find that exposure
 186 to ambient $PM_{2.5}$ from wildfire smoke is largely similar across
 187 racial-ethnic subgroups (Supplemental Table 1) and across dif-
 188 ferent levels of economic disadvantage (Supplemental Table
 189 2). However, similar ambient exposures could result in very
 190 different impacts across subgroups, due to potential differences
 191 in how pollutants infiltrate into indoor environments and/or dif-
 192 ferences in how a given increase in wildfire smoke exposure

interacts with baseline differences in other pollutant exposures
 or other determinants of learning outcomes. We thus consider the
 differential responses to a given exposure across districts with
 varying levels of economic disadvantage and proportions of non-
 White students. Each district’s level of economic disadvantage
 is measured by the Federal EdFacts data system and is typically
 defined at the state level as the proportion of students eligible for
 free or reduced-price lunch [48]. The proportion of non-White
 students in a district is calculated by subtracting the proportion
 of White students from 1. The proportion of White students is
 collected by the Common Core of Data and aggregated by the
 SEDA team to the district level [49]. We determine if districts
 have "High" versus "Low" levels of economic disadvantage and
 proportion of non-White students by thresholding at the median
 value of each variable. We emphasize that the estimated moder-
 ating effect of economic disadvantage or racial-ethnic categories
 in this analysis should be understood to reflect the possible effect
 of racist and/or discriminatory policies or attitudes on outcomes,
 rather than as reflecting inherent characteristics of individuals
 or communities that fall into these categories.

We find that districts with high economic disadvantage and
 high proportion of non-White student population as well as
 districts with low economic disadvantage and low proportion of
 non-White student population are more negatively affected by
 smoke $PM_{2.5}$ exposure compared to other subgroups (Figure 3).
 For students in districts with high economic disadvantage and a
 high proportion of non-White students, each additional $\mu g/m^3$ of
 cumulative smoke $PM_{2.5}$ on school days lowered test scores by
 0.008% (95% CI: -0.014% to -0.003%) of a standard deviation.
 Districts with low economic disadvantage and low proportion
 of non-White students also appeared negatively impacted by
 additional smoke $PM_{2.5}$ with decreases of 0.006% (95% CI: -
 0.012% to -0.001%) of a standard deviation. When we separate

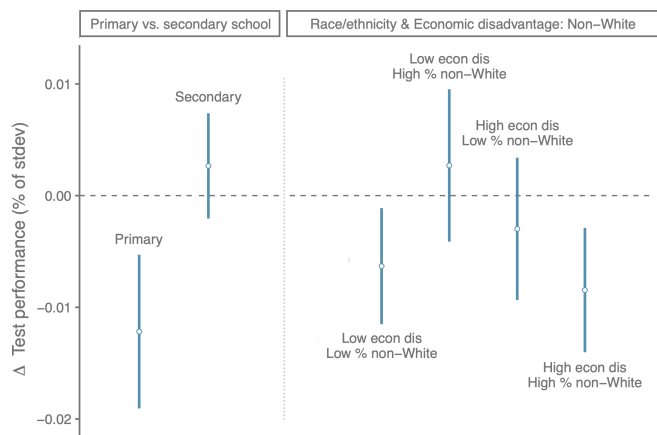


Figure 3: Heterogeneous effects of school day smoke $PM_{2.5}$ on test performance by grade, race/ethnicity, and level of economic disadvantage. Additional smoke $PM_{2.5}$ appear to impact primary school (grades 3-5) students more negatively than secondary school (grades 6-8) students. The right panel shows the point estimates and 95% confidence intervals for different intersecting levels of economic disadvantage and non-White racial-ethnic student population. The negative impacts of smoke $PM_{2.5}$ appear more pronounced for districts with high economic disadvantage and high non-White student population.

226 out the racial-ethnic subgroups we find that districts with a
 227 greater proportion of Asian, Black, or Hispanic students exhibit
 228 similar patterns of response to additional school day smoke
 229 $PM_{2.5}$ exposure as opposed to districts with a greater proportion
 230 of White students (Supplemental Figure 1).

231 To understand nation-wide impacts of less severe versus more
 232 severe average smoke years on learning, we compared the 2011
 233 versus 2016 school years, the former being the least smoky on
 234 average across districts and the latter being the most smoky year
 235 in our sample – albeit much less smoky than either 2018 or 2020,
 236 which are not in our sample. Taking into account heterogeneity
 237 in economic and racial-ethnic composition across school districts,
 238 we find substantial impacts of smoke exposure on learning
 239 across broad swaths of the US (Figure 4A), including large ex-
 240 tremes in the West, Midwest, and Northeast driven by fires in
 241 the Western United States and Canada. If smoke years continue
 242 to mirror the severity of 2011, we would expect students to expe-
 243 rience a decrease of 0.031% of a standard deviation in average
 244 test scores (median across districts), relative to a counterfactual
 245 of no smoke. On the other hand, if wildfire events are more
 246 similar to a severe smoke year, like the one in 2016, the median
 247 effect would be nearly an order of magnitude larger at 0.207%
 248 of a standard deviation decrease in average test scores. These
 249 impacts are again noticeable across all regions (Figure 4), but
 250 especially severe across the West.

251 As a rough estimate of the economic impact of cumulative
 252 smoke $PM_{2.5}$ exposure during the school year, we follow Park
 253 et al. (2020) [34] and calculate smoke impacts in terms of
 254 lost future earnings for students in our sample (Methods). We
 255 apply estimates from Chetty et al. (2014) [50], who found that
 256 a 1 standard deviation increase in teacher quality raised
 257 average tests scores by 0.13 standard deviations and resulted in
 258 a net present value of \$7,000 in future increased earnings for
 259 12 year-old students. Using this relationship, we estimate that

260 district-average smoke $PM_{2.5}$ exposure led to a reduction in the
 261 net present value of lost future earnings of \sim \$111 per student
 262 in 2016 compared to \sim \$17 in 2011. The lost earnings of \sim \$111 per
 263 student in 2016 totals nearly \$1.7 billion in potential lost future
 264 income from smoke $PM_{2.5}$ exposure when aggregating across
 265 all students in the US. We note that these impact estimates
 266 assume that increased future earnings due to teacher quality
 267 improvements are comparable to benefits of reducing smoke
 268 attributable $PM_{2.5}$ in the classroom. The estimated impacts could
 269 be overstated if teacher quality improvements result in other non-
 270 test performance related benefits that increase students’ future
 271 earnings. However, impacts of this magnitude illustrate the
 272 potential benefits of reducing wildfire smoke $PM_{2.5}$ exposure.

273 When we consider the cumulative losses over all study years
 274 and across subgroups (Figure 4B), we estimate net present value
 275 of lost future income of roughly \$544 million (95% CI: -\$999
 276 million to -\$100 million) from smoke $PM_{2.5}$ exposure in 2016
 277 for districts with low economic disadvantage and low propor-
 278 tion of non-White students. For districts with high economic
 279 disadvantage and high proportion of non-White students, we
 280 estimate cumulative impacts of \$1.4 billion (95% CI: -\$2.3 bil-
 281 lion to -\$477 million) from cumulative smoke $PM_{2.5}$ exposure in
 282 2016. Thus of the roughly \$1.7 billion in total costs during the
 283 smokiest year in our sample, 82% of the costs we estimate were
 284 borne by economically disadvantaged communities of color. The
 285 larger total burden in these communities is a function of both the
 286 more negative effect size and the relatively larger total number
 287 of students that attend schools in economically disadvantaged
 288 communities of color.

DISCUSSION

289 These results provide previously unaccounted for estimates
 290 of the negative impacts of smoke $PM_{2.5}$ exposure on test per-
 291 formance. Our study quantifies the impact of wildfire-smoke-
 292 attributable $PM_{2.5}$ exposure, a rapidly growing source of particulate
 293 exposure throughout much of the US and one which
 294 is expected to further increase as the climate warms [28]. We
 295 leverage a large time-series with test scores from school dis-
 296 tricts across the United States and a novel method for isolating
 297 smoke-attributable particulate matter and find that the negative
 298 impact of smoke exposure is present across test subjects, ap-
 299 pears stronger on days in which kids are in school, and affects
 300 communities with differing levels of economic disadvantage and
 301 racial-ethnic composition. While test scores are an imperfect
 302 measure of student cognition, they are a common metric for eval-
 303 uating student learning with relevance for long-term outcomes
 304 and opportunities [13, 50, 16]. The effects of smoke on school
 305 days suggests that smoke-related air pollution affects students’
 306 ability to learn in the classroom, and that these learning deficits
 307 ultimately affect their test performance.

309 Our findings add to a set of previous studies that have ex-
 310 amined the effect of other environmental exposures on student
 311 learning and test performance. In a study of heat on learning,
 312 an additional day with temperatures above 80°F (26.7°C) dur-
 313 ing the school year was found to decrease average test scores
 314 by 0.07% of a standard deviation [35], which was a little over
 315 twice our estimated impact of an average (6 $\mu g/m^3$) smoky day.
 316 Because there were on average more school days across the US
 317 with temperature above 80°F (32 in their sample) than average

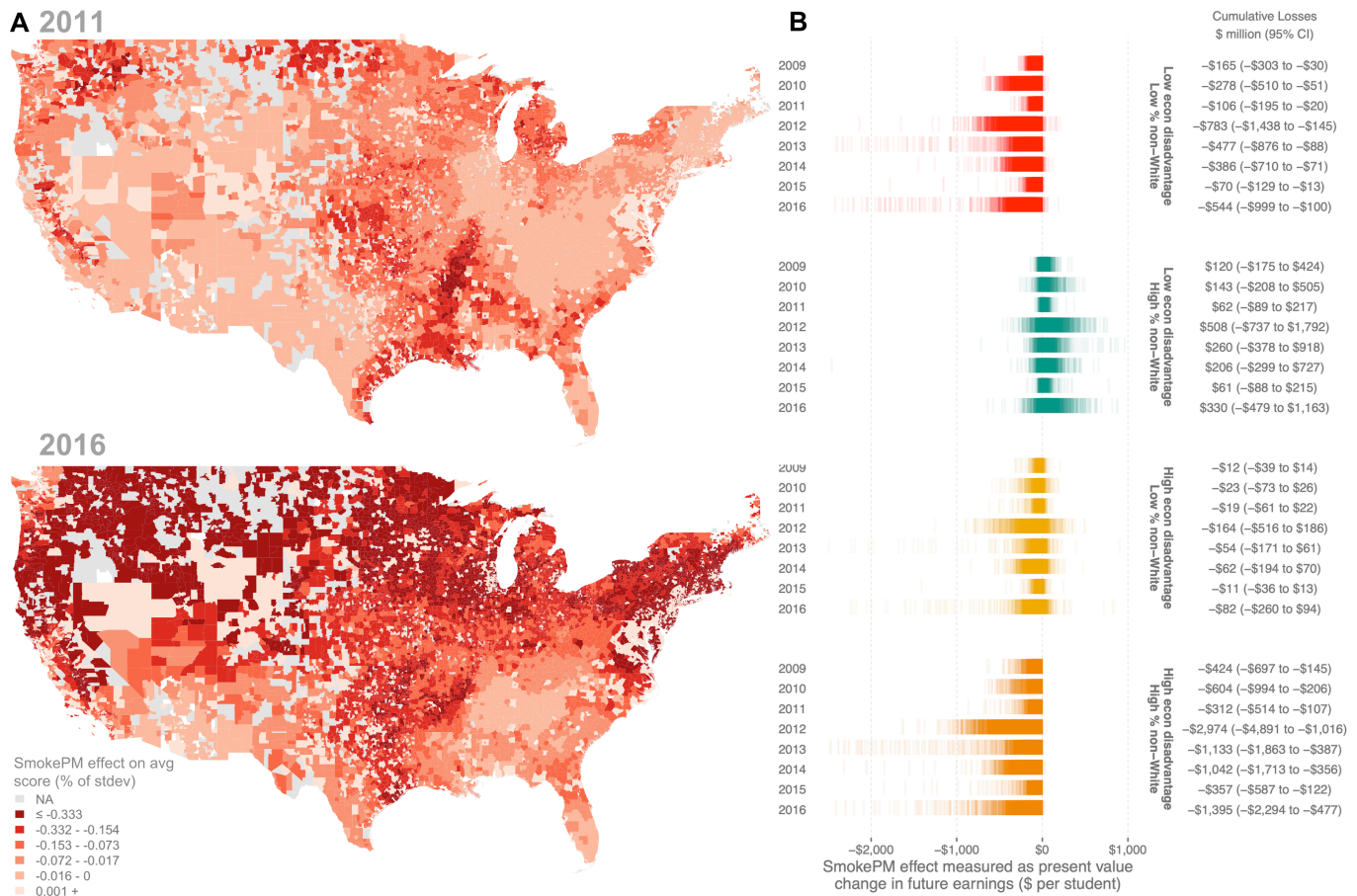


Figure 4: School day smoke PM_{2.5} effect on average test scores by district, and total effect on lost future earnings by economic disadvantage and racial-ethnic subgroup over time. **A.** Predicted effect of cumulative smoke PM_{2.5} exposure during school days on average student test scores in 2011 (top, a relatively low smoke year in our sample) and 2016 (bottom, a relatively high smoke year). In 2016, the West, Midwest, and Northeast experienced large effects from smoke PM_{2.5} exposure compared to in 2011. **B.** Effect of school day smoke PM_{2.5} on the net present value of future earnings separated by year, economic disadvantage, and racial-ethnic subgroup. Each tick mark represents a specific district in the matching year and subgroup (Methods). Cumulative net present value changes in future earnings are provided on the right and represent the total changes in future earnings across all students that fall into the matching year and subgroup, with 5th-95th percentile range across districts in parentheses.

318 days with smoke in the air in our sample (7 per school year), this
 319 suggests that the effects of heat are currently a more important
 320 determinant of learning outcomes than smoke. Nevertheless,
 321 the number of days with smoke in the air and the average con-
 322 centration of smoke PM_{2.5} on smoky days have both increased
 323 dramatically in the few years since the end of our study period
 324 [28, 47], suggesting a growing influence of smoke in more recent
 325 years.

326 In a study of the effect of air pollution exposure on the day of
 327 test taking on test performance in Israel, Ebenstein et al. (2016)
 328 found that a 1 standard deviation increase in PM_{2.5} (~16.7 (AQI))
 329 led to a 3.9% of a standard deviation decrease in test scores [13].
 330 Zivin et al. (2020) estimate that a 10µg/m³ increase in PM_{2.5} on
 331 the day of the National College Entrance Examination (NCEE)
 332 in China reduces test scores by 4.6% of a standard deviation [16].
 333 In our primary specification, we consider cumulative smoke
 334 PM_{2.5} exposure on school days in the year leading up to the
 335 exam period. A 1 standard deviation increase in the cumulative
 336 school day smoke PM_{2.5} (32.5µg/m³) would result in a decrease

of 0.14% (95% CI: -0.266% to -0.019%) a standard deviation. 337
 These results suggest that contemporaneous air pollution ex- 338
 339 posure has an order of magnitude larger effect on test scores
 340 compared to smoke PM_{2.5} exposure in the year prior to the exam.
 341 One explanation for this is that, for exposure during the school 342
 343 year, students can catch up on non-smoky days after suffering
 344 learning decrements on smoke days; such catch-up is not possi-
 345 ble when the exposure is on test day. Together these findings 346
 347 point to the desirability of executing "high stakes" cognitive
 348 tasks (e.g. standardized test taking) on days when air pollution
 349 from wildfires or other sources is likely to be low, although such
 350 avoidance behavior will be difficult for many tasks and increas-
 351 ingly difficult as the number of smoky days increases across the

country. 352
 One potential underlying biological mechanism for the 353
 354 observed negative effects of smoke PM_{2.5} exposure for primary
 355 school students compared to secondary school students is that
 air pollution is more harmful to younger children as their bodies
 are developing and their quicker breathing leads to increased

356 exposure which could ultimately affect their development [46].
357 Another possibility is that $PM_{2.5}$ affects students through more
358 absences due to potential health impacts, which ultimately re-
359 sults in reduced learning [51, 52, 14]. Better understanding the
360 mechanism by which younger students are affected will help
361 guide future intervention.

362 Perhaps surprisingly, we find that estimated effects of smoke
363 on learning are largest in both the least and the most disad-
364 vantaged communities. Similar effects at different ends of the
365 disadvantage spectrum could be a result of multiple sources of
366 heterogeneity that each have independent effects across groups.
367 For instance, in districts with high economic disadvantage and
368 high proportion of non-White students, differences in housing
369 or school characteristics – for example, a more permeable build-
370 ing envelope or differences in available filtration – could allow
371 more ambient pollution to infiltrate into and remain in indoor
372 environments [32, 33, 47]. Lower access to air conditioning
373 in disadvantaged schools [34] might also force classrooms to
374 keep windows open, increasing infiltration of wildfire smoke.
375 While limited work on infiltration of wildfire smoke does sug-
376 gest some role for factors such as income, race/ethnicity, and
377 housing quality in predicting infiltration into homes [47, 33],
378 more widespread measurement in schools will be needed to
379 understand whether differential infiltration can help explain the
380 heterogeneous results we find.

381 One potential explanation for the observed negative impacts
382 in low disadvantage communities is if the marginal effect of
383 additional exposure declines at higher baseline $PM_{2.5}$ exposure.
384 Such a non-linear relationship has been documented in health
385 impacts studies of wildfire specifically [53] and air pollution
386 more broadly [54, 55], and could be explained as the result of
387 adaptive investments in communities accustomed to higher aver-
388 age exposures, or alternatively as the relative importance of other
389 determinants of learning (e.g. school funding or teacher quality)
390 that happen to be correlated with baseline pollution exposure.
391 Indeed, we find that the effects of smoke $PM_{2.5}$ exposure are
392 more negative for districts with lower average baseline $PM_{2.5}$
393 levels (Supplemental Figure 2). Additionally, a majority of dis-
394 tricts with low economic disadvantage and low proportion of
395 non-White students within our sample are in the lowest baseline
396 $PM_{2.5}$ bin (Supplemental Table 3). However, because other sub-
397 groups also appear to have many districts in the lowest baseline
398 $PM_{2.5}$ bin, this explanation is also unlikely to fully explain the
399 heterogeneous effects we find.

400 Nevertheless, although both Whiter/wealthier and less White/
401 less wealthy communities experience similarly negative impacts
402 per student, we find that the overall burden in terms of total
403 lost earnings is borne mostly by disadvantaged districts of color
404 because these districts make up around 50% of the exposed stu-
405 dents in our sample. This suggests that additional increases in
406 future wildfire smoke exposure due to climate change will likely
407 disproportionately harm these communities, and that invest-
408 ments in wildfire risk reduction (e.g through fuels management)
409 could have large benefits in these communities.

410 We also find important regional differences that result from
411 where smoke travels, although these differences can change year
412 by year. For instance, while the Northeast experienced relatively
413 less smoke in the 2011 school year, smoke $PM_{2.5}$ exposure was
414 much higher in 2016. This could be due to wildfires in Canada,

415 which generate large amounts of smoke, and meteorological
416 conditions that transport smoke over great distances as in previ-
417 ously documented wildfire events [56, 57]. As wildfires continue
418 to increase due to climate change, regions that had previously
419 experienced relatively mild smoke events could begin to see
420 more wildfire smoke from regions far away even across national
421 boundaries [28].

422 Estimates of the present value of lost future earnings due to
423 decreased learning outcomes resulting from smoke exposure
424 suggest that in very smoky years, wildfire-attributable-smoke
425 $PM_{2.5}$ would effectively decrease the net present value of future
426 earnings by \$55,500 per school (for a school with 500 students)
427 and by nearly \$1.7 billion across the US, with ~80% of these
428 impacts in disadvantaged communities of color. These calcula-
429 tions assume that the relationship between test scores and future
430 earnings reported in Chetty et al. (2014) holds for smoke expo-
431 sure. Chetty et al. (2014) focuses on the effects of improved
432 teacher quality on test performance and eventual earnings [50],
433 which could be different from the effects studied here as im-
434 provement in teacher quality could lead to more than just test
435 score increases. If this is the case, then our estimates would over-
436 state the impact of smoke $PM_{2.5}$ on future income. Nevertheless,
437 while these are rough calculations, as wildfires and the asso-
438 ciated smoke events increasingly affect school districts across
439 the US, estimates like these can inform cost-benefit analyses of
440 investments aimed at reducing smoke $PM_{2.5}$ exposures.

441 Although the combined effect of smoke $PM_{2.5}$ exposure is
442 smaller than the identified school day smoke exposure effect, the
443 amount of smoke $PM_{2.5}$ on non-school days (includes summer
444 months) is much greater than the amount of smoke $PM_{2.5}$ on
445 school days. When we consider the combined effect of smoke
446 exposure (school and non-school days) and conduct a similar
447 calculation as shown in Figure 4B, we find much larger estimated
448 impacts across all years and subgroups (Supplemental Figure 4).
449 However, as we cannot be sure that students remain around their
450 school districts during summer months and other non-school
451 days, our analysis primarily focuses on school day exposures.

452 Compared to using satellite-derived smoke plume annotations
453 alone, our approach provides improved estimates of smoke-
454 attributable $PM_{2.5}$ by combining annotations with predicted
455 $PM_{2.5}$ data to separate smoke $PM_{2.5}$ from background $PM_{2.5}$.
456 However, the smoke plume annotations could be noisy because
457 they are drawn over multiple hours and usually only a couple
458 of times per day [57]. Future work to improve the precision
459 of the smoke annotations could lead to more precise estimates
460 of smoke attributable $PM_{2.5}$. Additionally, we currently do not
461 account for the specific district test taking dates and instead
462 remove any smoke observations between March - May. The
463 exposure calculation could be improved (where possible) by
464 compiling district specific testing dates, which would allow us
465 to more precisely measure the full period of exposure in the year
466 prior to the exam.

467 Our work contributes to a growing body of evidence demon-
468 strating the cognitive, health, and social harms of air pollution
469 in general, and wildfires specifically, and shows how dispari-
470 ties in these impacts across socioeconomic and racial-ethnic
471 groups can emerge even when there are negligible differences
472 across groups in ambient exposures. Our estimates also uncover
473 yet another substantial cost of a warming climate, with future

474 warming-driven increases in wildfire activity likely to worsen
475 learning outcomes.

476 METHODS

477 *Measuring wildfire-smoke-attributable PM_{2.5}*

478 To generate estimates of wildfire-smoke-attributable PM_{2.5} across all
479 school districts for all study years, we merge satellite derived smoke
480 plume data from the National Environmental Satellite, Data, and In-
481 formation Service (NESDIS) Hazard Mapping System (HMS) with
482 gridded estimates of daily PM_{2.5} concentrations from Di et al. (2021)
483 [40, 41]. We then estimate smoke-attributable PM_{2.5} as location- and
484 period-specific anomalous PM_{2.5} on days in which the plume data
485 indicated that smoke was overhead. Plume data derive from manual
486 annotations by trained analysts, using a variety of remote sensing products
487 including visible-band imagery from NOAA’s GOES satellites multi-
488 ple times per day across the US [57]. In total we use nearly 200,000
489 individual smoke plumes between 2008-2016.

490 The predicted PM_{2.5} data [40, 41] is provided as daily PM_{2.5} concen-
491 trations for all-source PM_{2.5} (not just wildfire PM_{2.5}) for the Contiguous
492 United States in a 1 kilometer grid from 2000 - 2016. The predictions
493 are made using an ensemble of 3 machine learning models including
494 neural networks, random forests, and gradient boosted trees. Each of
495 the models includes multiple explanatory variables including satellite
496 observations, land-use variables, chemical transport predictions, and
497 other variables. The authors note that the ensemble model achieved
498 performance of $r^2=0.86$ for daily PM_{2.5} predictions. To isolate PM_{2.5}
499 from wildfires, we follow [47] and calculate smoke-attributable PM_{2.5}
500 as the deviation from location-specific median PM_{2.5} on non-smoke
501 days in the same month, with the median calculated over a 3-year win-
502 dowed centered on the current year. Specifically, the smoke-attributable
503 PM_{2.5} anomaly is calculated by subtracting the month-specific 3-year
504 non-smoke day median estimated from the predicted PM_{2.5} at each
505 school district on days with a smoke plume overhead. After we obtain
506 the smoke PM_{2.5} anomalies, we set this smoke PM_{2.5} variable to 0
507 for non-smoke days and the positive anomaly for days with a plume
508 overhead. Smoke days with negative anomaly values were also set to 0.
509 The resulting measure of smoke PM_{2.5} isolates the smoke component
510 of overall PM_{2.5} so long as, on average, other PM_{2.5} sources are not
511 also anomalously high on days when smoke is in the air – a plausible
512 assumption given the large degree of temporal and spatial randomness
513 in when and where fires start and where plumes go.

514 *Assigning smoke PM_{2.5} exposure to school districts*

515 We calculate a student-population weighted average of school level
516 exposure to estimate aggregate exposure at the district level. We further
517 delineate school day exposure versus non-school day exposure, specify-
518 ing non-school days as weekends and federal bank holidays throughout
519 the year and all days from June 15 to August 15. Because standardized
520 testing in the US is conducted at various points throughout the Spring,
521 usually between March and May, our analysis focuses on exposures
522 from the previous June to February. For this analysis, we focus on
523 school years between 2009-2016 as the predicted PM_{2.5} data is only
524 available between 2000-2016.

525 *Outcome and covariate data*

526 Test score data were compiled by Stanford University and made
527 available through the Stanford Education Data Archive (SEDA) [58].
528 The SEDA team constructed the dataset by converting state-specific
529 proficiency data to a nationally comparable dataset by scaling the state
530 results using a nationally representative sample from the National As-
531 sessment of Educational Progress (NAEP). For specific details about

532 how the dataset was created and the calculations involved in scaling
533 the state scores to a national dataset, we direct interested readers to
534 [58]. The SEDA data contains nationally comparable test scores for
535 students in grades 3-8 from 2009-2018. These test scores are broken
536 down into district-level results for both math and English language arts
537 (ELA) subjects. Rather than represent an absolute score, the metric
538 provided in the dataset is a standardized score within subject and grade,
539 relative to representative cohorts who had taken the NAEP assessments
540 [49]. Therefore, a score of 0.25 for Math means that an average student
541 in that district performed 0.25 of a standard deviation higher than the
542 reference cohort that took the NAEP assessment. The primary outcome
543 we consider is the average test score between ELA and math at the
544 district level.

545 In addition to calculating an average treatment effect across aggre-
546 gated data in our main model specification, we also investigate hetero-
547 geneous effects using district-level racial-ethnic and economic disad-
548 vantage covariates. Specifically, the level of economic disadvantage is
549 measured by the Federal EdFacts data system and is typically defined
550 using the proportion of students eligible for free or reduced-price lunch
551 [48]. The proportion of non-White students in a district is calculated
552 by subtracting the proportion of White students from 1 where the pro-
553 portion of White students is collected by the Common Core of Data
554 (CCD) and aggregated by the SEDA team to the district level [49].

555 We use gridded (4km x 4km) temperature and precipitation data pro-
556 duced by the Parameter elevation Regressions on Independent Slopes
557 Model (PRISM) Climate Group at Oregon State University [42]. We
558 extract the max daily temperature at each school and take a weighted
559 average using the student population at the schools that belong to that
560 district. As with the weighting for smoke exposure, the student popu-
561 lation data was collected from the National Center for Education
562 Statistics (NCES). We then create bins by counting the number of days
563 in the year prior to the exam with max temperatures in 10°F (5.5°C)
564 increments from 0°F (-17.7°C) to 80°F (26.7°C). All days less than
565 or equal to 0°F (-17.7°C) are grouped into 1 bin and all days greater
566 than 80°F (26.7°C) are grouped into another bin. We process the daily
567 precipitation data similarly and extract the daily precipitation measure-
568 ments at each school then calculate the total annual precipitation at the
569 district level.

570 *Estimating the effect of smoke PM_{2.5} on student performance*

571 Our main regression specification is as follows (equation 1):

$$572 \text{Score}_{igy} = \beta_1 \text{SmokePM}_{iy}^{\text{school}} + \beta_2 \text{SmokePM}_{iy}^{\text{non-school}} \quad (1)$$

$$573 + f(\mathbf{X}_{iy}) + \eta_i + \gamma_{yg} + \epsilon_{igy}$$

574 Here, *Score* represents the scaled standardized score for each district
575 i in grade g , and year y . The district fixed effect η_i is a separate intercept
576 (dummy variable) for each district that accounts for any average differ-
577 ences in smoke exposure or test scores across districts. This empirical
578 approach ensures that we are not comparing districts that might inher-
579 ently be very different from each other. The grade-year fixed effects γ_{yg}
580 account for differences in year-grade specific exposures or outcomes
581 that affect all districts across the US, such as overall trends in test
582 scores or in average differences between test scores across grades. $f(\mathbf{X})$
583 represents a vector of controls including the number of days with max
584 temperatures in each of the 10°F (5.5°C) bins, which controls for po-
585 tential non-linear effects of temperature within the district for the year
586 preceding the test, and total annual precipitation in the year prior to the
587 exam. *SmokePM* is defined as the total amount of smoke-attributable
588 PM_{2.5} in the year prior to the exam in our primary specification. We
589 define $\text{SmokePM}_{iy}^{\text{school}}$ and $\text{SmokePM}_{iy}^{\text{non-school}}$ as the total cumulative
590 smoke PM_{2.5} in a preceding year y within district i between June and
591 February on school and non-school days respectively. β_1 represents the

590 average effect of an additional $\mu\text{g}/\text{m}^3$ of cumulative smoke $\text{PM}_{2.5}$ on
 591 school days on test performance. β_2 represents the average effect of
 592 an additional $\mu\text{g}/\text{m}^3$ of cumulative smoke $\text{PM}_{2.5}$ on non-school days
 593 on test performance. To estimate the subject specific coefficient esti-
 594 mates (Figure 2B), we replace mean score Score_{igy} with the subject
 595 specific scores and consider the cumulative SmokePM (both school and
 596 non-school) in the year prior to the exam. We cluster standard errors
 597 by county to account for arbitrary within-unit autocorrelation in ϵ_{igy}
 598 and weight districts by the total number of students who took the test
 599 provided in the SEDA dataset.

600 We also conduct secondary analyses (equation 2) looking at the
 601 heterogeneous effects of smoke exposure on test outcomes. To examine
 602 these effects, we study whether the effects of smoke $\text{PM}_{2.5}$ differ across
 603 different grade levels and a combination of economic disadvantage and
 604 race-ethnicity, using the following specification:

$$\begin{aligned} \text{Score}_{igy} = & \sum_n \beta_n (\mathbb{1}_n * \text{SmokePM}_{iy}^{\text{school}}) \\ & + \sum_n \beta_n (\mathbb{1}_n * \text{SmokePM}_{iy}^{\text{non-school}}) \\ & + f(\mathbf{X}_{iy}) + \eta_i + \gamma_{yg} + \epsilon_{igy} \end{aligned} \quad (2)$$

605 Here, $\mathbb{1}_n$ represents an indicator function for whether or not the ob-
 606 servation i falls into a specific bin n . To determine these bins, we divide
 607 districts into "High" or "Low" categories based on thresholding at the
 608 median value for race-ethnicity and economic disadvantage variables.
 609 The remainder of the equation is similar to equation 1.

610 Calculating the effect of smoke $\text{PM}_{2.5}$ in terms of lost future 611 income

612 To translate the effect estimates into the net present value of lost
 613 future earnings, we follow the approach used in Park et al. (2020)
 614 [34]. Specifically, we assume the relationship found in Chetty et al.
 615 (2014) [50] holds, which estimated that a 1 standard deviation increase
 616 in teacher quality raised average tests scores by 0.13 standard devia-
 617 tions and resulted in a net present value of \$7,000 in future increased
 618 earnings for 12 year-old students. Therefore, if the estimated effect
 619 of an additional $\mu\text{g}/\text{m}^3$ of smoke $\text{PM}_{2.5}$ is a decrease of 0.01% of a
 620 standard deviation and the average smoke $\text{PM}_{2.5}$ experienced in a year
 621 is $10 \mu\text{g}/\text{m}^3$, then we calculate the average effect as $0.01\% \times 10 = 0.1\%$.
 622 We can then apply the Chetty et al. (2014) conversion and calculate
 623 that $\frac{0.001 \times 7000}{0.13} = \53.85 on average per student for that year of smoke
 624 $\text{PM}_{2.5}$ exposure.

625 In panel B of Figure 4, we plot the average net present value change in
 626 future earnings for each district as an individual tick mark, calculated as
 627 described above. For each of the four economic disadvantage and racial-
 628 ethnic subgroups, we draw 3000 samples from a normal distribution
 629 with mean centered at the matching subgroup coefficient (equation
 630 2) and standard deviation set to the estimated standard error. We
 631 then merge this with district information by matching on the districts'
 632 subgroup for each year. From this data, we estimate the district specific
 633 average impacts by year and we sample 1 observation out of the 3000 to
 634 show as a tick mark. Additionally, we use the sampled data to estimate
 635 95% interval estimates for the cumulative changes in net present value
 636 of future earnings.

637 Calculate distance to nearest fire perimeter

638 We calculate distances from schools to the nearest fire perimeter
 639 in each year provided by the National Interagency Fire Center [59].
 640 Then, we take an average of the minimum school-to-fire distances
 641 to get estimates of the average distance to the nearest fire for each

district. While evacuation zone distances vary, recent studies of wildfire
 evacuations in California suggest that short-distance evacuations are
 much more common than longer distance evacuations to destinations
 outside of the county of residence [60]. Additionally, a 1.5 mile (2.4
 kilometer) distance is often cited as the distance that forest fire embers
 can travel and ignite flammable materials at distant locations beyond the
 fire front [61]. Given this, we consider dropping school districts within
 a range of distances from 1-kilometer to 20-kilometers to the nearest
 fire perimeter and find that the identified effects are consistent up to
 dropping districts within 10-kilometers of the nearest fire perimeter
 (Supplemental Figure 3).

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1 **Supplementary Information for**
2 **Wildfire Smoke Exposure Worsens Learning Outcomes**

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6 *This is the supplementary information for a non-peer reviewed preprint submitted to EarthArXiv. It has been submitted for*
7 *publication in a peer reviewed journal, but has yet to be formally accepted for publication.*

8 **This PDF file includes:**

9 Figs. S1 to S4

10 Tables S1 to S4

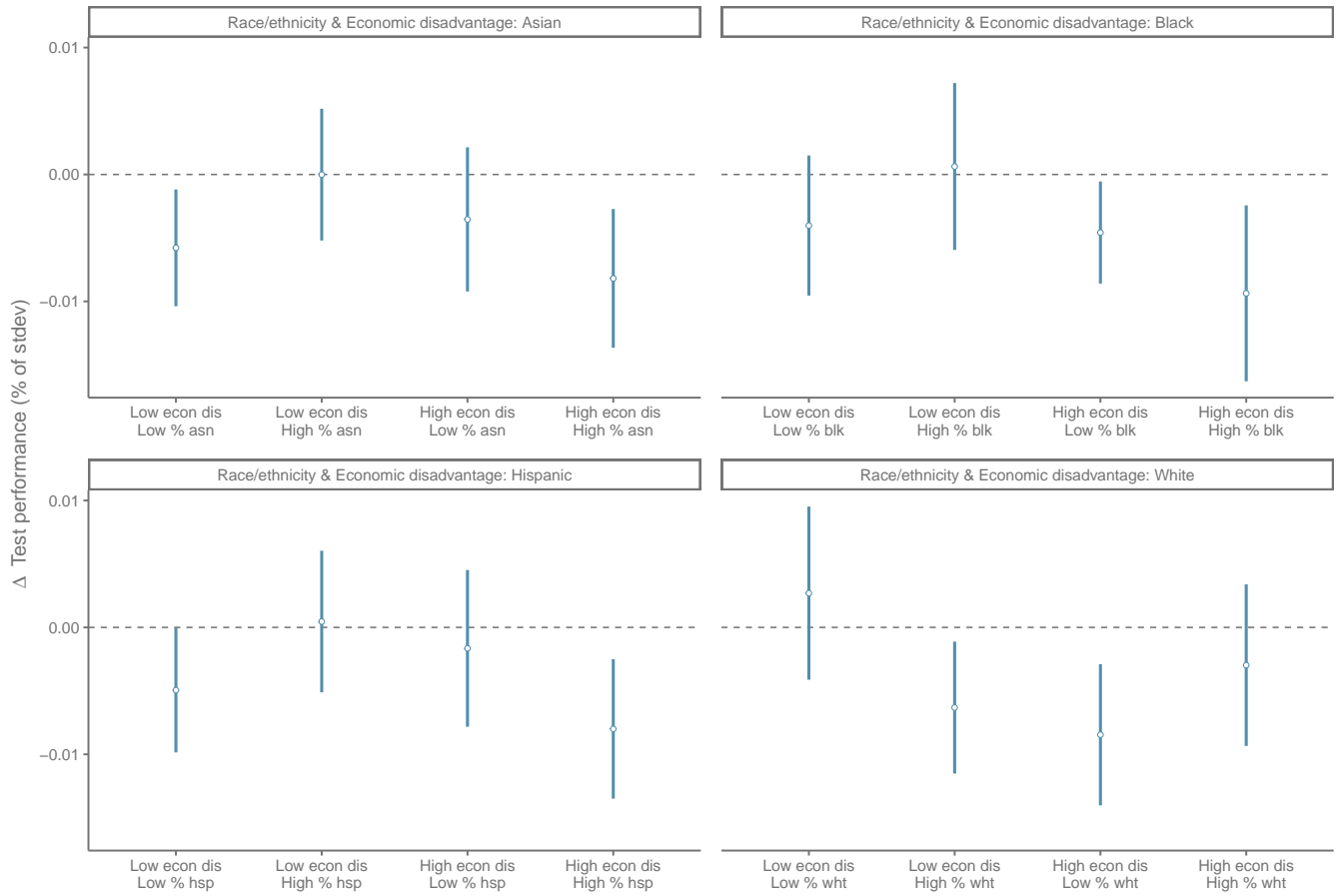


Fig. S1. Effect estimates of additional cumulative school-day smoke $PM_{2.5}$ exposure across different racial-ethnic groups and levels of economic disadvantage. The bottom right panel shows effect estimates across across different levels of % White students while the right panel in Figure 3 of the main text shows the complement and subtracts the % of White students from 1.

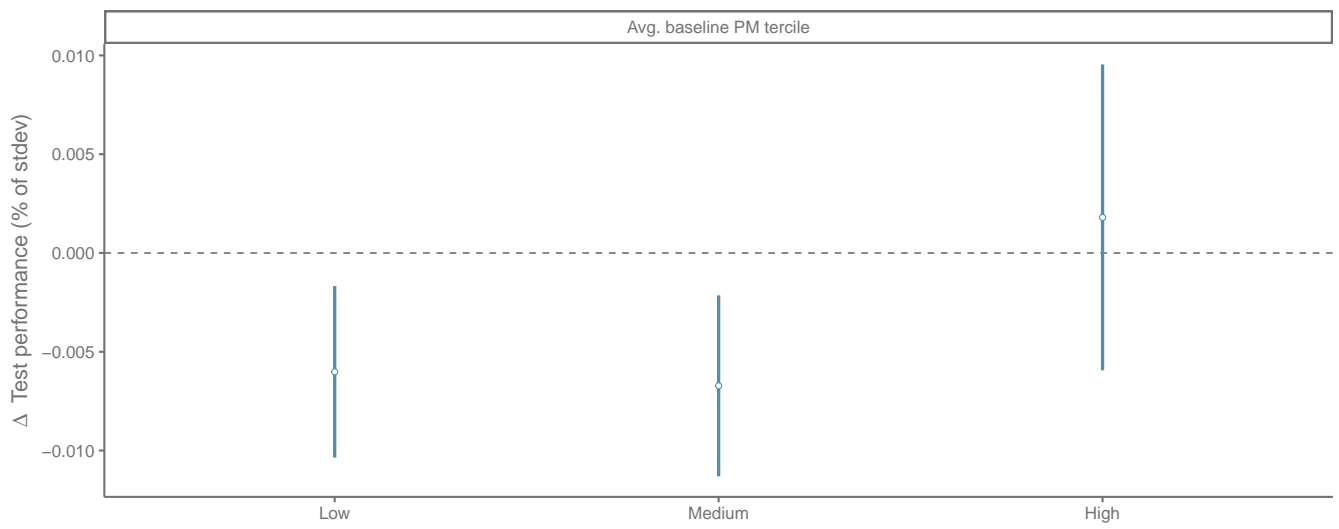


Fig. S2. Effect estimates of additional cumulative school-day smoke PM exposure across different levels of baseline $PM_{2.5}$. The baseline $PM_{2.5}$ bins were determined by calculating the average total $PM_{2.5}$ for each district and separating into bins based on terciles across our data sample.

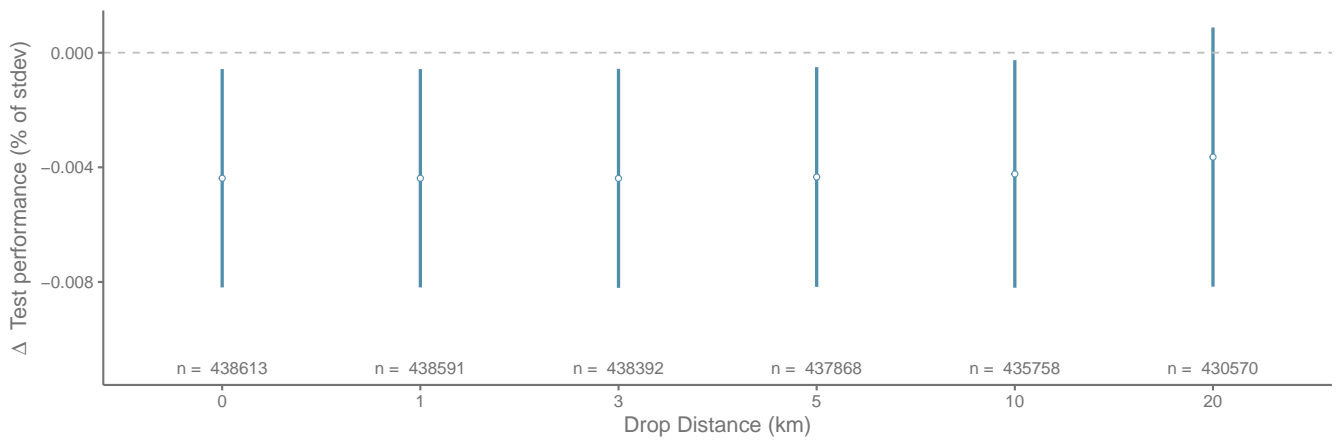


Fig. S3. To test that the identified effects are driven by exposure to wildfire-attributable smoke $PM_{2.5}$ rather than from the direct wildfire impacts, we drop school districts that within a certain distance to the nearest wildfire perimeter in that year. The estimated effects remain fairly stable even when dropping districts that are within 6.2 miles (10 kilometers) to the nearest fire perimeter. 95% confidence intervals are shown with standard errors clustered by county.

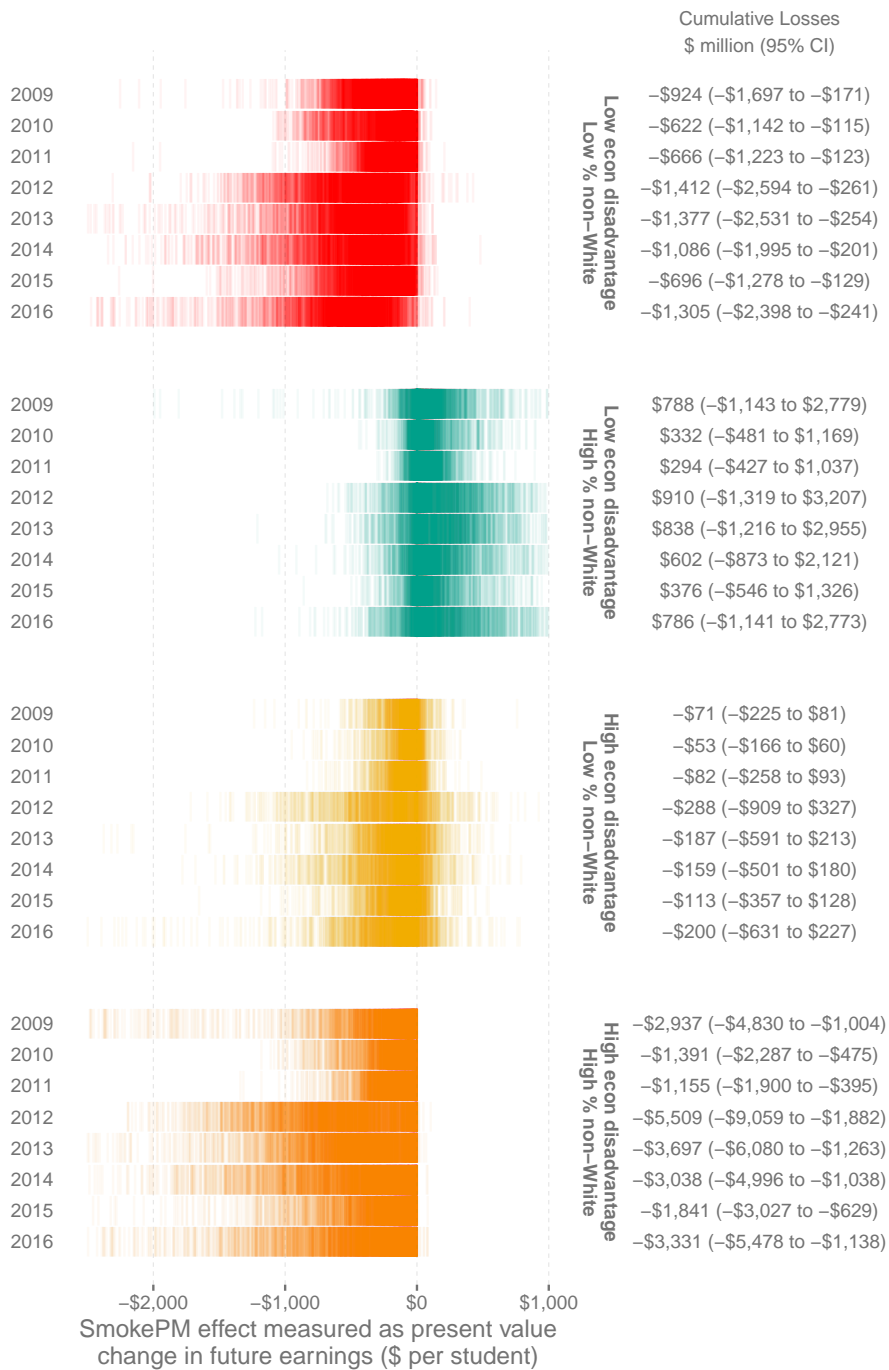


Fig. S4. Effect of total cumulative smoke $PM_{2.5}$ on the net present value of future earnings separated by year, economic disadvantage, and racial-ethnic subgroup. Figure differs from the Figure 4B in the main text as it uses estimates identified using total (school and non-school) day smoke $PM_{2.5}$ exposure rather than just school day smoke $PM_{2.5}$.

Table S1. Median cumulative smoke PM_{2.5} exposure by race/ethnicity on school and non-school days

Category	Asian		Black		Hispanic		Native Amer.		White	
	High	Low	High	Low	High	Low	High	Low	High	Low
Nonschool SmokePM	48.49	50.86	45.38	54.61	45.70	52.60	49.15	49.81	58.00	39.58
School SmokePM	19.96	20.16	18.26	22.20	19.79	20.28	19.97	20.13	23.11	16.80

Notes: Exposure to wildfire smoke PM_{2.5} is noticeably greater on non-school days as this includes the summer period before school begins.

Table S2. Median cumulative smoke PM_{2.5} exposure by economic disadvantage on school and non-school days

Category	% Economic Disadvantage	
	High	Low
Nonschool SmokePM	42.08	55.63
School SmokePM	17.41	22.45

Table S3. % of districts by subgroup within each average baseline PM_{2.5} bin.

Econ disadvantage & racial-ethnic subgroup	Avg. baseline PM _{2.5}		
	Low	Medium	High
High econ dis & High % non-White	37.60	33.02	29.38
High econ dis & Low % non-White	47.47	27.83	24.70
Low econ dis & High % non-White	37.61	32.12	30.27
Low econ dis & Low % non-White	43.95	28.70	27.35

Notes: A district's baseline PM_{2.5} is calculated as the average yearly PM_{2.5} across the sample and bins are created by splitting the data into terciles.

Table S4. Lagged impacts of school day smoke PM_{2.5} exposure

Model:	(1)	(2)	(3)
<i>Variables</i>			
Contemporaneous year school smoke PM _{2.5}	-0.004 (0.002)	-0.007 (0.003)	-0.010 (0.003)
1 year lagged school smoke PM _{2.5}		-0.004 (0.003)	-0.005 (0.003)
2 year lagged school smoke PM _{2.5}			-0.007 (0.003)
<i>Fixed-effects</i>			
District	✓	✓	✓
Year x Grade	✓	✓	✓
<i>Controls</i>			
Temperature	✓	✓	✓
Precipitation	✓	✓	✓
Observations	438,613	371,417	306,018

Notes: Standard errors clustered by county are shown in parentheses. Effect estimates are represented as a percent of a standard deviation change in average test score.