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1 **A Statistical Evaluation of WRF-LES Trace Gas Dispersion**
2 **Using Project Prairie Grass Measurements**

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ABSTRACT

14 In recent years, new measurement systems have been deployed to monitor and quantify methane
15 emissions from the natural gas sector. Large-eddy simulation (LES) has complemented measure-
16 ment campaigns by serving as a controlled environment in which to study plume dynamics and
17 sampling strategies. However, with few comparisons to controlled-release experiments, the accu-
18 racy of LES for modeling natural gas emissions is poorly characterized. In this paper, we evaluate
19 LES from the Weather Research and Forecasting (WRF) model against measurements from the
20 Project Prairie Grass campaign, surface layer similarity theory, and the Gaussian Plume Model.
21 Using WRF-LES, we simulate continuous emissions from an ensemble of 30 near-surface trace
22 gas sources in two stability regimes: strong and weak convection. We examine the impact of grid
23 resolutions ranging from 6.25 m to 52 m in the horizontal dimension on model performance. We
24 evaluate performance in a statistical framework, calculating fractional bias and conducting Welch's
25 t -tests. WRF-LES accurately simulates observed surface concentrations at 100 m and beyond under
26 strong convection; the magnitude of fractional bias is less than 30% for the moderate- and fine-
27 resolution simulations. However, in weakly convective conditions with strong winds, WRF-LES
28 substantially overpredicts concentrations – the magnitude of fractional bias often exceeds 30%,
29 and all but one t -test fails. Despite the good performance of dispersion in the strongly convective
30 atmosphere, we find that both the strongly and weakly convective boundary layers disagree with
31 empirical wind and temperature Monin-Obukhov similarity theory profiles that are often used to
32 evaluate LES within the atmospheric surface layer.

33 **1. Introduction**

34 Natural gas production within the U.S. has surged in the past decade, increasing by more than
35 50% since 2010 (EIA 2020). Large emissions from routine operations (Thorpe et al. 2020) and
36 malfunctioning equipment (Conley et al. 2016) have spurred the development of new methane
37 emission monitoring instruments and platforms, including satellites, piloted aircraft, unmanned
38 aircraft, open-path lasers, and ground-based point sensors (Fox et al. 2019). Source estimation
39 techniques (SETs) are used to interpret source characteristics (e.g. emission rate) from the trace gas
40 concentration measurements collected via these systems (Harper et al. 2011). Operational source
41 estimation techniques (OSETs) are computationally low-cost and simple to use, and they vary
42 from instrument to instrument. Satellites and remote sensing aircraft often use the integrated mass
43 enhancement (IME) technique (Frankenberg et al. 2016; Varon et al. 2018; Jongaramrungruang
44 et al. 2019). In situ aircraft measurements often use mass balance techniques (Karion et al. 2013;
45 Conley et al. 2017). Many ground-based sensors employ techniques that rely on a transport and
46 dispersion model, such as the Gaussian Plume Model (Pasquill 1972; U.S. EPA 2014; Coburn et al.
47 2018).

48 To build trust, OSETs are often tested and calibrated against measurements in the field. Of all the
49 common OSETs used to quantify natural gas emissions, approaches based on the Gaussian Plume
50 Model have been the most extensively tested against measurements. The Gaussian Plume Model
51 has been evaluated and calibrated against hundreds of controlled releases through studies such as
52 Project Prairie Grass (PPG) (Barad 1958) and the EPA OTM 33A evaluation study (U.S. EPA 2014).
53 These studies have yielded better understanding of the accuracy and limitations of the Gaussian
54 Plume Model for studying emissions from the natural gas sector. However, OSET evaluation
55 studies that are based on measurements come with limitations, as they quantify performance in

56 the specific conditions that are encountered in the field (e.g. atmospheric stability, terrain). For
57 example, the OTM 33A evaluation study characterized performance in relatively flat terrain, but
58 the technique has since been applied in hilly terrain (Caulton et al. 2019). Additionally, OSETs that
59 rely on measurements from aircraft and satellites have been evaluated against fewer measurements.
60 These techniques are newer, and it can be more expensive and logistically complicated to make
61 measurements of controlled releases with these instruments. As a result, aircraft- and satellite-
62 based OSETs have relied more heavily on synthetic observations from models, namely large-eddy
63 simulation (LES). Overall, as new methods are developed to quantify methane emissions from the
64 natural gas sector, it is critical to ensure that their corresponding OSETs are accurate.

65 Recently, interest has grown in using LES as a tool for studying natural gas emissions. LES is
66 a computational approach that numerically solves the volume-averaged Navier-Stokes equations
67 for flow at large scales and parameterizes small-scale flow with subgrid-scale models. It has been
68 extensively applied in studies of the atmospheric boundary layer (ABL) (Deardorff 1972; Moeng
69 1984; Mason 1994; Beare et al. 2006). LES has been used as part of emission quantification
70 studies to improve measurement strategies (Conley et al. 2017), evaluate and improve OSETs
71 (Taylor et al. 2016; Varon et al. 2018), test new OSETs and their assumptions (Conley et al. 2017;
72 Jongaramrungruang et al. 2019), generate realistic synthetic measurements of methane (Saide et al.
73 2018), and act as a transport model for field campaign observations (Caulton et al. 2018). LES is
74 computationally expensive but offers several advantages over simpler gas transport and dispersion
75 models. LES models the dynamic behavior of plumes as driven by time-varying winds, thereby
76 circumventing the need to assume time-averaged fields or steady-state behavior, two assumptions
77 employed in many simpler models. LES provides meteorological and concentration fields at all
78 time steps and locations within a domain, whereas observations provide only a subset of these
79 fields. LES can be used to used to study plume dynamics under desired atmospheric forcing, and

80 furthermore, LES can simulate complicated physics encountered at real-world natural gas facilities,
81 such as complex terrain (Lundquist et al. 2012; Xue et al. 2018) and time-varying emissions (Saide
82 et al. 2018); therefore, in principle, LES could be used to accurately test OSETs or measurement
83 strategies under a wide variety of environmental conditions.

84 Unfortunately, LES of atmospheric trace gas dispersion has been statistically evaluated against
85 relatively few experimental measurements (Steinfeld et al. 2008; Ardeshiri et al. 2020), and thus
86 its accuracy for emission quantification studies is not extensively characterized. The most well-
87 known comparison studies focus on the strongly convective ABL in flat terrain. Convective
88 tank studies first done by Willis and Deardorff (1976) and improved upon by Weil et al. (2002)
89 provided a controlled environment to study tracer dispersion in strong convection. Additionally, the
90 CONDORS study (Eberhard et al. 1988) released tracers into a real convective ABL. Subsequent
91 LES studies have found good agreement with both sets of measurements in the mixed layer (Lamb
92 1978; Nieuwstadt and de Valk 1987; Weil et al. 2004, 2012; Nottrott et al. 2014).

93 LES evaluation studies that examine atmospheric dispersion in the surface layer (less than
94 approximately 100 m above ground level) have often found worse performance. For example, Weil
95 et al. (2012) compared surface concentrations in the atmospheric surface layer from an LES-driven
96 Lagrangian particle dispersion model to observations from the PPG field campaign. The study
97 found good agreement between the two beyond approximately 500 m downwind of the source, but
98 LES underpredicted concentrations by as much as a factor of two 50 m downwind. Other studies
99 suggest that LES dispersion underperforms when forced by conditions other than strong convection.
100 In one neutral boundary layer, LES underpredicted horizontal trace gas dispersion (Nottrott et al.
101 2014). In a neutrally stratified field campaign study with multiple controlled releases, LES tended
102 to overpredict emissions (Caulton et al. 2018). As many methane monitoring technologies measure

103 within the atmospheric surface layer and in a range of atmospheric stabilities, understanding the
104 performance of LES in these scenarios is key.

105 In this paper, we evaluate the performance of LES from the Weather Research and Forecasting
106 model (WRF-LES) in the atmospheric surface layer under two types of forcing: strong convection
107 and weak convection. We compare simulated surface concentrations from WRF-LES to data from
108 the PPG field campaign, 50–800 m downwind of a passive tracer source. We assess the impact of
109 LES grid resolution on plumes. Additionally, we compare against two well-studied transport and
110 dispersion models that are often employed in ground-based OSETs: surface layer similarity (SLS)
111 theory and the Gaussian Plume Model. Recognizing the importance of stochastic uncertainty
112 caused by turbulence (Rao 2005), we evaluate performance in a statistical framework (Chang and
113 Hanna 2004) and simulate a 30-member ensemble of plumes. In doing so, we aim to better
114 understand the accuracy of WRF-LES under simple but realistic methane emission scenarios.

115 In Section 2, we describe the WRF-LES dispersion simulations, the PPG field campaign, the
116 transport and dispersion models, and the statistical metrics used in this study. In Section 3, we
117 evaluate the performance of WRF-LES in a strongly convective boundary layer, and we find good
118 agreement with both measurements as well as SLS theory. In Section 4, we find that WRF-LES
119 performance suffers in a weakly convective boundary layer. In Section 5, we discuss possible
120 paths to improve LES accuracy, and we evaluate wind and temperature profiles relative to Monin-
121 Obukhov similarity theory (MOST). In Section 6, we offer conclusions based on the study findings.

122 2. Methods

123 *a. WRF-LES Simulations*

124 We evaluate the performance of the LES code from Advanced Research WRF (WRF-ARW
125 Version 4.1.2) (Skamarock et al. 2019). WRF-ARW is a numerical weather prediction code that
126 uses the finite difference method to solve the compressible, nonhydrostatic Euler equations on a
127 mass-based grid. It is a popular community-driven code with more than 36,000 registered users,
128 and it serves as the foundation for several additional codes (Powers et al. 2017) with applications
129 ranging from fire modeling (WRF-FIRE) to renewable energy modeling.

130 To evaluate the performance of WRF-LES, we simulate dispersion in the atmospheric boundary
131 layer with six different configurations (Table 1). We model two types of convection—a strongly
132 convective boundary layer (SCBL) and a weakly convective boundary layer (WCBL)—and we
133 simulate each with a coarse-, moderate-, and fine-resolution grid. All cases incorporate flat terrain,
134 cyclic boundary conditions for meteorological fields, a surface roughness of $z_0 = 0.008$ m, and
135 homogeneous surface heating. Simulations are run without moisture, radiation, microphysics,
136 or other parameterizations commonly employed in mesoscale WRF runs. The simulations in
137 this study use third-order Runge-Kutta to step forward in time, as well as fifth-order horizontal
138 advection and third-order vertical advection. The nonlinear backscatter anisotropic turbulence
139 model captures subgrid effects (Kosović 1997; Mirocha et al. 2010), and MOST provides the lower
140 boundary condition via the MM5 surface layer model (Jiménez et al. 2012).

141 Both the SCBL and WCBL spin up for two model hours, after which WRF begins to save the
142 fields of interest. The SCBL is forced with constant 3.6 m s^{-1} geostrophic winds, $0.24 \text{ W K}^{-1} \text{ m}^{-1}$
143 surface heat flux, a $1 \times 10^{-4} \text{ s}^{-1}$ Coriolis parameter, and a 0.1-s time step. The SCBL horizontal
144 grid resolutions are $\Delta x = 52$ m, 26 m, and 10 m for the coarse, moderate, and fine simulations,

145 respectively. These forcings and the coarse grid resolution are consistent with Weil et al. (2012).
146 The WCBL is forced with constant 10 m s^{-1} geostrophic winds, $0.1 \text{ W K}^{-1} \text{ m}^{-1}$ surface heat flux,
147 a $1 \times 10^{-4} \text{ s}^{-1}$ Coriolis parameter, and a 0.05-s time step. The WCBL horizontal grid resolutions
148 are 31.25, 15.625, and 6.25 m for the coarse, moderate, and fine simulations, respectively. All
149 coarse and moderate simulations use constant vertical grid spacing, respectively 21 m and 10.5 m
150 in the SCBL and 10.5 m and 5.25 m in the WCBL. The fine-resolution simulations use vertical
151 grid resolutions that change. In the fine SCBL and WCBL simulations, the height of the first grid
152 cell is $z_1 = 3 \text{ m}$, and concentrations are output mid-cell height at 1.5 m. The near-surface grid cells
153 stretch at a rate of 3% until $\Delta z = 10$ or 6.25 m is reached for the fine SCBL and WCBL respectively.
154 Cells stretch again above the capping inversion at 3%, enabling higher resolution in the region area
155 of interest at reduced computational expense.

156 To address the highly stochastic nature of dispersion in the turbulent ABL, continuous emissions
157 are simulated from 30 different surface point sources in a grid with 500-m spacing, as in Weil
158 et al. (2012) (Figure 1). Each source experiences different local winds, so that each plume evolves
159 somewhat independently, circumventing the need for an ensemble of simulations for a single set
160 of conditions. Each plume is tagged so that concentrations from one source are distinguishable
161 from the other sources. Emissions are simulated from a point source at the lowest grid cell as in
162 Nunalee et al. (2014). Dispersion is modeled in an Eulerian framework. As a result, the height
163 of the emission source decreases as grid resolution is increased, which impacts concentrations
164 nearest the source. To nullify the impact of recirculating plumes resulting from periodic boundary
165 conditions, we include a trace gas absorbing plane 500 m upwind of each source.

166 After a two-hour spin-up, we sample trace gas fields and winds every second during a 10-minute
167 period, matching the PPG measurement period. From these concentration fields, we calculate
168 crosswind integrated concentration (CWIC) at a given radius as:

$$CWIC = \Delta s \left(\sum_i C_i \right), \quad (1)$$

169 where C_i is the concentration at a cell i and Δs is the arclength between cells. To account for the
 170 different release rates used in PPG, CWIC calculations throughout this study are normalized by
 171 emission rate Q , and this quantity is referred to as “concentration” though strictly speaking it is
 172 a “normalized crosswind integrated concentration”. In order to compare the medium and coarse
 173 simulations to the PPG horizontal array measurements collected at a height of 1.5 m, concentration
 174 profiles are extrapolated using a 5th-order polynomial fit to concentrations in the lowest 100 m.
 175 For each simulated emission source, we calculate 10-minute-averaged CWIC at 50, 100, 200, 400,
 176 and 800 m downwind.

177 *b. Project Prairie Grass*

178 The PPG field campaign was conducted in 1956 in Kansas to study the near-surface behavior
 179 of passive tracer plumes during various meteorological conditions (Barad 1958). This campaign
 180 serves as a cornerstone for atmospheric dispersion models, informing key parameters in the Pasquill-
 181 Gifford stability classes for the Gaussian Plume Model (Venkatram 1996) and acting as a validation
 182 dataset for many regulatory dispersion models such as AERMOD (Cimorelli et al. 2005). Seventy
 183 controlled releases of SO_2 were carried out: six at 1.5 m above ground level and the remainder at
 184 0.48 m. For each controlled release, 10-minute average concentration measurements were collected
 185 at an array of 599 individual sampling points. Measurements were conducted in concentric arcs
 186 50, 100, 200, 400, and 800 m downwind of the release source. Along each arc, a horizontal
 187 array of point measurements was gathered at a height of 1.5 m, spaced 1° apart at 800 m and
 188 2° at all other downwind distances. A vertical array of measurements was also collected 100 m
 189 downwind at heights 0.5, 1.0, 1.5, 2.5, 4.5, 7.5, 10.5, 13.5, and 17.5 m. The overall concentration

190 uncertainties were reported as 1–2%. The roughness length of the site was estimated to be $z_0 =$
191 0.008 m (Sawford 2001). The winds employed in this study were measured with a cup anemometer
192 25 m west of the release source at a height of 2 m during a 10-minute period. Obukhov lengths
193 L and friction velocities u_* were not directly measured during the campaign but were estimated
194 from tower measurements in subsequent analysis (Horst et al. 1979). Normalized CWIC for the
195 horizontal array is taken from Horst et al. (1979), and normalized CWIC for the vertical array is
196 calculated using digitized data courtesy of www.harmon.org/jsirwin.

197 Measurements from a number of runs are either excluded in this analysis or not available. The
198 runs used here are listed in Table 2. Data was not reported for Run 63 and Run 64 because of
199 “extremely light and variable winds”. Vertical tower measurements were gathered only for Run
200 13 and beyond and were additionally not reported for runs 23, 28, 35, 53, 63, and 64; thus, fewer
201 vertical profiles are available for comparison. Winds speeds were not reported for Run 3 and Run
202 6, so those runs are excluded from this analysis.

203 We aim to compare as many observations to WRF-LES concentration simulations as possible.
204 In principle, this comparison would best be achieved by running one simulation for each controlled
205 release, because each release occurs in the presence of a different L and u_* ; however, running
206 one high-resolution simulation for each observation would be prohibitively expensive. As an
207 alternative, we assess the performance of WRF-LES by binning PPG runs with similar atmospheric
208 conditions into strongly convective and weakly convective categories. One common method to bin
209 data in atmospheric dispersion studies is the Pasquill-Gifford stability classes (De Visscher 2013).
210 These classes are traditionally delineated using wind speeds and solar radiation, but they can be
211 alternatively delineated using a roughness length and Obukhov length (Golder 1972). Class A
212 corresponds to $0 \geq L \geq -7$ m for the PPG roughness length. This range is used to bin PPG data for
213 comparison with the SCBL LES runs, which have L between -6.1 and -5.4 m. The WCBL LES

214 runs have L between -16 and -12.3 m, which falls on the border between Class B ($-7 \geq L \geq -15$
 215 m) and Class C stability ($-15 \geq L \geq -50$ m). Accordingly, we use intermediate values of the PPG
 216 runs, $-10 \geq L \geq -35$ m, for the LES WCBL comparison bin. To more closely resemble the WCBL
 217 LES, we additionally require $u_* \geq 0.4 \text{ m s}^{-1}$.

218 *c. Transport and Dispersion Models*

219 We use two transport and dispersion models for comparison with the LES results: SLS theory
 220 and the Gaussian Plume Model.

221 SLS theory (van Ulden 1978) is used to complement the PPG observations. Each observation
 222 has a different pair of u_* and L values, and none of these pairs precisely match the conditions in
 223 the LES; however, SLS theory can be used to calculate approximate CWIC under any desired u_*
 224 and L conditions. Normalized CWIC at a height z is calculated for the PPG runs as:

$$\frac{CWIC(z)}{Q} = \frac{0.73}{\bar{u} \bar{z}} \exp \left[- \left(\frac{0.66z}{\bar{z}} \right)^{1.5} \right], \quad (2)$$

225 where Q is the emission rate, \bar{z} is the plume centerline height, and \bar{u} is the wind speed at the
 226 plume centerline. The values of \bar{z} and \bar{u} are numerically computed based on MOST, and downwind
 227 distance x is implicitly a function of these variables. SLS theory is strictly valid for releases at a
 228 height of 0 m, but it agrees well with the observations in this study (Appendix). As such, we use
 229 SLS theory as a proxy for hypothetical observations, with u_* and L that match those of the LES.

230 Although SLS theory cannot be used to directly study the sensitivity of dispersion to source
 231 height and wind speed, the Gaussian Plume Model does approximate how dispersion responds to
 232 these two factors. Normalized CWIC at downwind distance x and height z is calculated for the
 233 SCBL with the Gaussian Plume Model (Arya 1999):

$$\frac{CWIC(x, z)}{Q} = \frac{1}{\sqrt{2\pi}u\sigma_z} \left[\exp\left(-\frac{(z-h)^2}{2\sigma_z^2}\right) + \exp\left(-\frac{(z+h)^2}{2\sigma_z^2}\right) \right], \quad (3)$$

234 where u is the wind speed at the source, σ_z is the vertical plume spread, and h is the emission
 235 height, assumed to be 0.48 m. The Briggs (1973) equations are used to calculate σ_z , where
 236 Pasquill-Gifford Class A is employed.

237 *d. Statistical Metrics*

238 Chang and Hanna (2004) summarize metrics for evaluating dispersion models by comparing an
 239 observation, C_o , to a model prediction, C_p . While there is no one optimal metric, they conclude
 240 that “good performing models” have predictions that fall within a factor 2 of observations (FAC2)
 241 at least 50% of the time, that the relative mean bias (here fractional bias, FB) is less than 30%, and
 242 that the relative scatter (here normalized mean square error, NMSE) is less than approximately a
 243 factor of two. FAC2 is calculated as the fraction of data within $0.5 \leq C_p/C_o \leq 2.0$. Fractional bias
 244 is calculated as:

$$FB = \frac{\overline{C_o} - \overline{C_p}}{2(\overline{C_o} + \overline{C_p})}, \quad (4)$$

245 where averages are taken over the set of measurements or simulations. NMSE is calculated as:

$$NMSE = \frac{\overline{(C_o - C_p)^2}}{\overline{C_o} \overline{C_p}}. \quad (5)$$

246 Here, only observations for the horizontal array are used for quantitative comparison, as the vertical
 247 array stability bins have only two or three observations.

248 SLS model performance is compared against observations using FAC2, FB, and NMSE (Ap-
 249 pendix). In contrast, FAC2 and NMSE are not calculated for LES because these metrics require

250 each observation to be paired with a model prediction. Instead, we use the Welch's t -test to
251 compare the LES distribution and the observed distribution of concentration. Both distributions
252 are assumed to be Gaussian at each downwind location. The null hypothesis is that the mean
253 concentrations for these distributions are identical, and the test is conducted at the 95% confidence
254 interval with a two-sided tail. Mean LES concentrations are also evaluated using FB. However,
255 for the FB comparison, SLS theory serves as comparison—instead of observations—in order to
256 minimize error stemming from differences in L and u_* .

257 **3. Evaluation of LES in the Strongly Convective Boundary Layer**

258 *a. Horizontal Surface Concentrations*

259 LES of trace gas plumes in the SCBL performs well from the perspective of grid convergence
260 (Figure 2). Mean surface concentrations in the coarse-, moderate-, and fine-resolution simulations
261 collapse onto the same line beyond 200 m; however, LES surface concentrations upwind of 200
262 m increase as resolution is increased, suggesting that concentrations are grid-dependent close to
263 the source. For example, at 50 m downwind (Figure 2 inset), the fine simulation concentrations
264 exceed those from the coarse simulation by a factor of 1.6. This increase may be attributable to
265 two factors that change with resolution in the lowest grid cell: emission height and wind speed. As
266 vertical resolution increases, the simulated emission height decreases from 10.5 m in the coarse
267 simulation to 1.5 m in the fine simulation, as trace gas is released from the center of the lowest cell.
268 A lower emission height leads to higher surface concentrations near the source. This change in
269 resolution also leads to slower wind speeds in the lowest grid cell, due to the increased proximity
270 to the surface. The winds in the lowest grid cell of the fine-resolution simulation $u_{h=1.5, fine} = 1.16$
271 m s^{-1} are slower than those in the lowest grid cell of the coarse simulation $u_{h=10, coarse} = 1.92$ m

272 s^{-1} . Slower winds lead to less plume dispersion and therefore higher concentrations at the same
273 downwind distance.

274 We employ the Gaussian Plume Model to quantitatively estimate the impact of these two factors.
275 Using Equation 3, CWIC is calculated at 50 m for emission heights of 1.5 and 10.5 m, both
276 driven by fine winds $u_{h=1.5, fine}$. The concentration from the 1.5-m release exceeds that from the
277 10.5-m height release by a factor of 1.6. We also calculate 50-m CWIC for $u_{h=1.5, fine}$ winds and
278 $u_{h=10, coarse}$ winds at the same release height of 1.5 m. This change in wind speeds also leads to a
279 factor 1.6 increase in concentrations at higher wind speeds. Taken together, the Gaussian Plume
280 Model predicts that a change in source height and wind speed would lead to a factor 2.6 increase
281 in 50-m CWIC. This increase is larger than the observed factor 1.6 increase between the fine and
282 coarse LES. Nonetheless, we conclude that both factors contribute roughly equally to a near-source
283 increase in concentrations as grid resolution is refined.

284 LES of the SCBL also performs well relative to observations. Beyond 200 m downwind,
285 all LES resolutions show good fractional bias ($|FB| < 30\%$) relative to SLS (Table 3). This
286 behavior is consistent with Weil et al. (2012), who studied dispersion in identical SCBL conditions
287 with Lagrangian particle dispersion driven by a different LES code, NCAR-LES. As WRF-LES
288 resolution increases, performance improves close to the source. FB at 100 m decreases from 40%
289 to 15%, and FB at 50 m decreases from 78% to 38% when moving from the coarse to the fine
290 simulation.

291 A Welch's t -test at each downwind location is used to assess whether the average LES and average
292 measured concentrations differ significantly. The t -test results corroborate the FB findings. More
293 than 200 m downwind of the source, all LES resolutions produce concentration distributions whose
294 mean concentrations are statistically indistinguishable from those of PPG. Closer to the source,

295 resolution plays an increasingly important role. The coarse resolution simulation fails the t -test at
296 100 m, but the moderate- and fine-resolution cases succeed.

297 It is crucial that these comparisons are rooted in a statistical framework—LES ensemble members
298 in the SCBL display a significant amount of scatter. At 200 m and beyond, the minimum and
299 maximum concentrations differ by more than an order of magnitude. This scatter occurs even
300 though all plumes are subject to the same geostrophic winds and surface heating. In the SCBL,
301 individual plume behavior is strongly governed by the local presence of updrafts and downdrafts
302 (Weil et al. 2012).

303 *b. Vertical Concentration Profiles*

304 As with the horizontal array, we find that WRF-LES performs well against vertical profiles of
305 concentration (Figure 3). The average concentrations agree for the coarse-, moderate-, and fine-
306 resolution simulations at heights above 10.5 m, which is the height of the lowest concentration
307 measurement from the coarse simulation. The coarse simulation predicts relatively narrow vari-
308 ability between ensemble members, but the moderate- and fine-resolution simulation have similar
309 spread to each other. WRF-LES agrees well with SLS theory and shows only minor deviations at
310 17.5 m, which may be attributed to differing micrometeorological conditions. The PPG observa-
311 tions show a slightly stronger concentration gradient across the surface layer, but this difference
312 may also be attributable to different values of L and u_* . In the SCBL, LES qualitatively performs
313 well against surface concentrations as well as vertical profiles at 100 m.; thus in conjunction with
314 the analysis of the horizontal array, we conclude that WRF-LES accurately models realistic plume
315 behavior in the SCBL, provided sufficient resolution is used.

316 **4. Evaluation of LES in the Weakly Convective Boundary Layer**

317 *a. Horizontal Surface Concentrations*

318 Unlike the SCBL, the LES simulations of the WCBL perform poorly relative to SLS theory and
319 observations in the horizontal dimension (Table 4). Most comparisons show $|FB| > 30\%$, which
320 is outside the “good” performance threshold from Chang and Hanna (2004). While $|FB| < 30\%$
321 near 100 m, this downwind distance is simply the crossover point where LES transitions from
322 overprediction to underprediction. Additionally, every comparison aside from the 200-m coarse
323 resolution case fails the t -test. This single success case is dismissed as coincidental, because the
324 200-m results turn to “Reject” when grid resolution is increased.

325 *b. Vertical Concentration Profiles*

326 The WCBL similarly performs poorly relative to the vertical array of measurements (Figure 5).
327 Profiles of concentration do not converge as well across different resolutions in the WCBL as in the
328 SCBL. The mean LES concentrations for the moderate- and fine-resolution simulations agree above
329 10 m but show different behavior below. Interestingly, the moderate resolution simulations show
330 substantially less scatter than both the coarse and the fine simulations, further underscoring the lack
331 of grid convergence. Furthermore, LES substantially overpredicts concentrations relative to both
332 observations and SLS theory. Altogether, WRF-LES performs poorly in the weakly convective
333 case.

334 **5. Discussion on Disagreement in the WCBL**

335 Near-surface turbulence within the atmospheric surface layer is characterized by anisotropy, a
336 small outer length scale, a strong dependence on atmospheric stability, and a “reverse turbulent

337 cascade” where small scales transfer energy to larger scales (Sullivan et al. 2003). These character-
 338 istics make it challenging for LES to accurately model flow in this region, and the inability of our
 339 LES to capture all of these features likely drives the overpredicted concentrations in the WCBL.
 340 Modelers are actively researching methods to improve LES accuracy near solid surfaces. Within
 341 the atmospheric surface layer, these techniques include improving subgrid-scale models (Porté-
 342 Agel et al. 2000; Bou-Zeid et al. 2005; Chung and Matheou 2014; Mokhtarpoor and Heinz 2017),
 343 improving wall models (Maronga et al. 2019), and refining grid size and aspect ratio (Brasseur and
 344 Wei 2010; Daniels et al. 2016).

345 During their development, LES techniques for the surface layer are typically evaluated against
 346 MOST. This theory is derived for flat terrain under homogeneous forcing, as is the case in this LES
 347 study, and it has been shown to agree well with observations in these conditions (Businger et al.
 348 1971; Dyer 1974). MOST describes wind and temperature profiles in the atmospheric surface layer
 349 based on L ; a non-dimensional wind shear, ϕ_m ; and a non-dimensional temperature gradient, ϕ_h
 350 (Stull 1988). This non-dimensional function takes one of many similar empirical forms (Maronga
 351 and Reuder 2017), and it is calculated from either observations or LES as

$$\phi_m \left(\frac{z}{L} \right) = \frac{d\bar{u}_h}{dz} \frac{\kappa z}{u_*}, \text{ and} \quad (6)$$

$$\phi_h \left(\frac{z}{L} \right) = \frac{d\bar{\theta}}{dz} \frac{\kappa z}{\theta_*}, \quad (7)$$

352 where \bar{u}_h is the mean horizontal wind speed, $\bar{\theta}$ is the average potential temperature, θ_* is the
 353 kinematic heat flux divided by friction velocity, and κ is the von Kármán constant, taken to be 0.4.

354 We calculate ϕ_m directly from LES wind fields and compare it to empirical profiles based on the
 355 LES values of u_* and L using the Dyer (1974) equations (Figure 6 a,b). At all three resolutions,

356 the LES-based non-dimensional wind shear profiles in the SCBL agree well with one another.
357 These profiles are larger than the empirical MOST profile by about a factor of two, but they all
358 qualitatively show similar behavior. On the other hand, the ϕ_m profiles in the WCBL behave
359 differently. A large peak (“overshoot”) is observed in the LES-based profiles, and the height of this
360 overshoot decreases as resolution increases, as in Brasseur and Wei (2010). We similarly calculate
361 ϕ_h profiles (Figure 6 c,d). All LES-based profiles show dependence on grid resolution as well as
362 an overshoot both within the SCBL and in the WCBL. Interestingly, this overshoot is larger within
363 the SCBL even though no overshoot was observed in its corresponding wind shear profiles.

364 These profiles illuminate an interesting discrepancy between using PPG observations and using
365 MOST to diagnose LES performance. LES of the WCBL agrees poorly with the PPG trace gas
366 observations; thus, the LES-based MOST profiles unsurprisingly agree poorly with their anticipated
367 form. At the same time however, LES of the SCBL agrees well with PPG while simultaneously
368 disagreeing with MOST profiles. This inconsistency suggests one of two scenarios: that either the
369 geostrophic wind and the heat flux selected for the SCBL LES were coincidentally good choices or
370 that LES can accurately resolve near-surface dispersion under certain conditions even if it disagrees
371 with MOST. For example, perhaps the lack of a wind shear overshoot in the SCBL explains its
372 good dispersion performance. Future LES studies of near-surface dispersion will clarify which
373 case is true.

374 **6. Conclusion**

375 In this study, we assess the accuracy of WRF-LES for simulating trace gas dispersion in three
376 strongly convective and three weakly convective boundary layers where grid resolution is varied.
377 We compare 30 plumes within each simulation to horizontal and vertical measurements from
378 the Project Prairie Grass campaign (50–800 m downwind of a source, with measurements at

379 heights of 0.5–17.5 m). We also compare WRF-LES simulations to surface layer similarity
380 theory and the Gaussian Plume Model. We evaluate the performance of WRF-LES dispersion
381 using a statistical framework, relying on the fractional bias metric and Welch’s *t*-tests to compare
382 distributions. In strongly convective conditions with weak winds, WRF-LES, the Project Prairie
383 Grass measurements, and the SLS theory tend to agree well. Furthermore, WRF-LES performs
384 better as grid resolution is increased. In contrast, during weak convection and stronger winds,
385 WRF-LES substantially overpredicts concentrations.

386 To shed more light on the performance of LES within the lower atmospheric surface layer, we
387 evaluate wind and temperature profiles against Monin-Obukhov similarity theory (MOST). We find
388 that the weakly convective LES poorly agrees with MOST, which may justify the poor performance
389 of dispersion under this forcing; however, we simultaneously find that LES of the strongly convective
390 boundary layer also disagrees with MOST, even though the simulated concentrations agree with
391 Project Prairie Grass measurements. We suggest further study on the relationship between wind,
392 temperature, and trace gas concentration for LES of the atmospheric surface layer.

393 The results of this study caution that WRF-LES, and atmospheric LES codes in general, should be
394 evaluated in a statistical framework to available empirical datasets when possible. By simulating 30
395 plumes under identical large-scale forcing, we consider the stochastic nature of turbulent diffusion.
396 At times we observe order-of-magnitude differences in 10-minute averaged concentrations. This
397 study examined the simple case of flat terrain and homogeneous forcing, but the conclusions are
398 broadly applicable to studies examining dispersion in more challenging scenarios, such as complex
399 terrain or urban environments.

400 LES has many unique features, which makes it an invaluable tool for modeling emissions of trace
401 gases. LES can be, and has been used, to improve measurement strategies for field campaigns. It can
402 simulate dispersion in complex environments, which is valuable as regulators seek to characterize

403 real-world emissions in industrial environments with complex terrain and time-varying emissions.
404 Through further comparisons against controlled releases, trust in LES dispersion can be fostered,
405 and it can begin to take a more central role in the emission quantification challenge.

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421 and plume fields used in this study can be found at <https://doi.org/10.5281/zenodo.3909881>.
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TABLE 1. Key input parameters and observed values for each simulation.

Case	SCBL – Coarse	SCBL – Moderate	SCBL – Fine	WCBL – Coarse	WCBL – Moderate	WCBL – Fine
Domain Size (Lx, Ly, Lz) [km]	(5, 5, 2)	(5, 5, 2)	(5, 5, 2)	(3, 3, 1)	(3, 3, 1)	(3, 3, 1)
Cell Count (Nx, Ny, Nz)	(96, 96, 96)	(192, 192, 192)	(500, 500, 200)	(96, 96, 96)	(192, 192, 192)	(500, 500, 160)
Horizontal Resolution [m]	52	26	10	31.25	15.625	6.25
First Cell Height [m]	20.8	10.4	3	10.4	5.2	3
Geostrophic Wind (Ug, Vg) [m/s]	3.6	3.6	3.6	10	10	10
Surface Heating [Km/s]	0.24	0.24	0.24	0.1	0.1	0.1
Obukhov Length [m]	-6.1	-5.9	-5.4	-16.0	-14.9	-12.3
Friction Velocity [m/s]	0.29	0.29	0.29	0.46	0.47	0.45
Bottom of Capping Inversion [m]	1050	1050	1050	525	525	525

TABLE 2. PPG observations used in this study.

SCBL			WCBL		
Run	u^* [m/s]	L [m]	Run	u^* [m/s]	L [m]
15	0.22	-6.6	2	0.12	-18
16	0.23	-3.3	5	0.37	-29
25	0.19	-5.4	8	0.29	-19
47	0.22	-5.3	9	0.43	-34
48S	0.21	-5.2	12	0.5	-48
			19	0.37	-25

TABLE 3. LES performance in the SCBL.

	Coarse LES		Moderate LES		Fine LES	
	FB (%)	t-Test	FB (%)	t-Test	FB (%)	t-Test
50 m	78	Reject	56	Reject	38	Reject
100 m	40	Reject	2	Not Reject	15	Not Reject
200 m	-27	Not Reject	4	Not Reject	-3	Not Reject
400 m	-16	Not Reject	8	Not Reject	-7	Not Reject
800 m	4	Not Reject	5	Not Reject	17	Not Reject

TABLE 4. LES performance in the WCBL.

	Coarse LES		Moderate LES		Fine LES	
	FB (%)	t-Test	FB (%)	t-Test	FB (%)	t-Test
50 m	92	Reject	75	Reject	55	Reject
100 m	25	Reject	12	Reject	-26	Reject
200 m	-17	Not Reject	-50	Reject	-58	Reject
400 m	-48	Reject	-93	Reject	-115	Reject
800 m	-65	Reject	-107	Reject	-138	Reject

	SCBL			WCBL		
	FAC2 (%)	FB (%)	NMSE (%)	FAC2 (%)	FB (%)	NMSE (%)
50 m	100	3	2	100	4	1
100 m	100	6	5	100	26	7
200 m	100	3	10	100	19	5
400 m	100	-2	7	100	15	5
800 m	60	18	3	83	23	14

Table A1. Performance of SLS theory relative to PPG observations.

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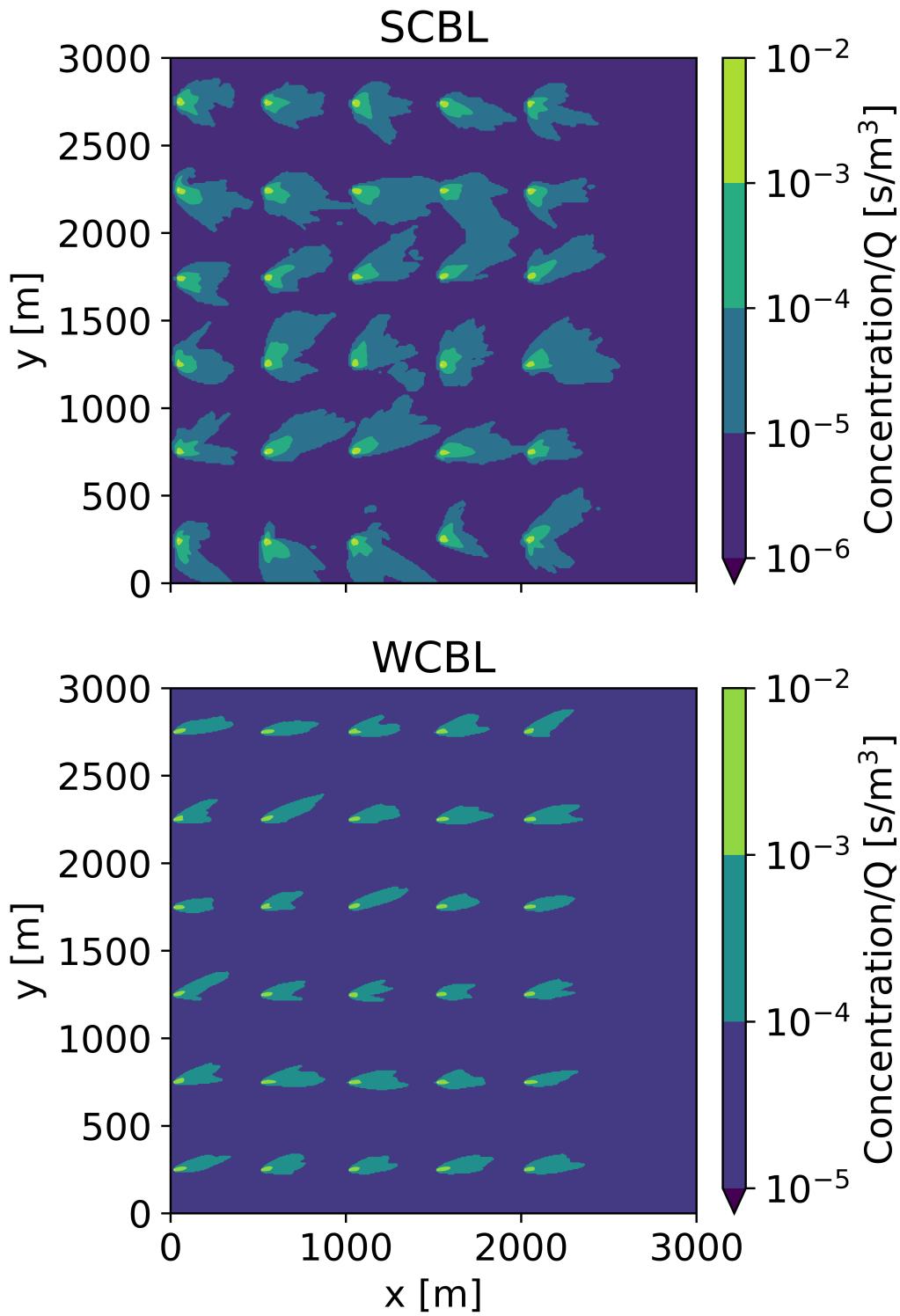
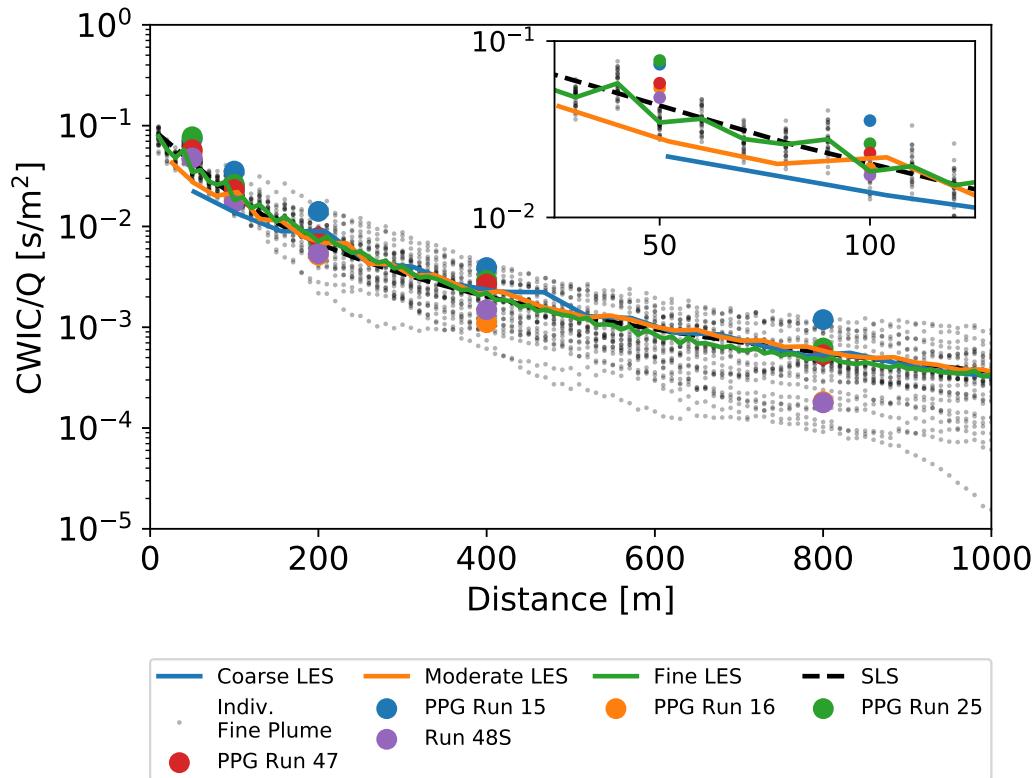
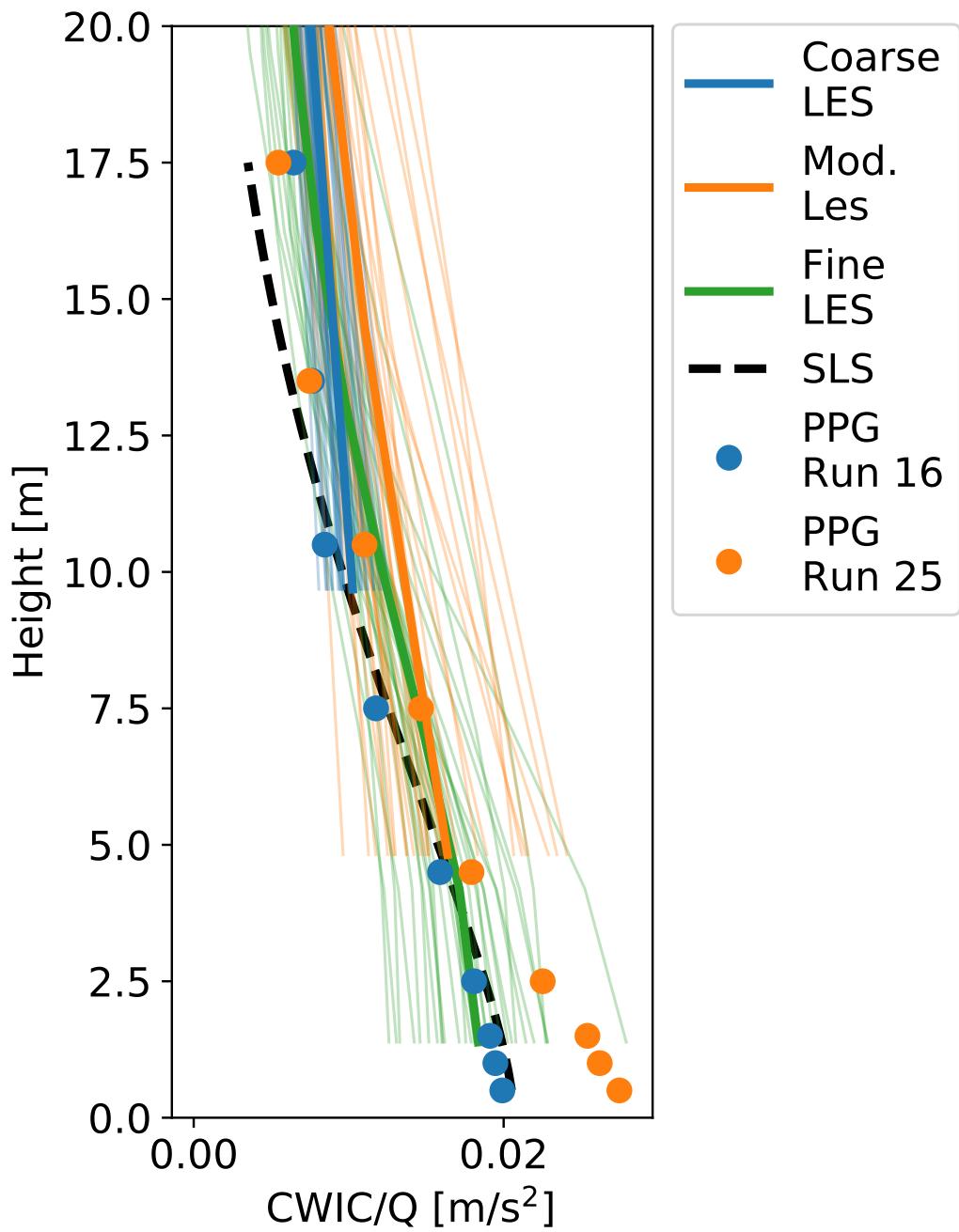


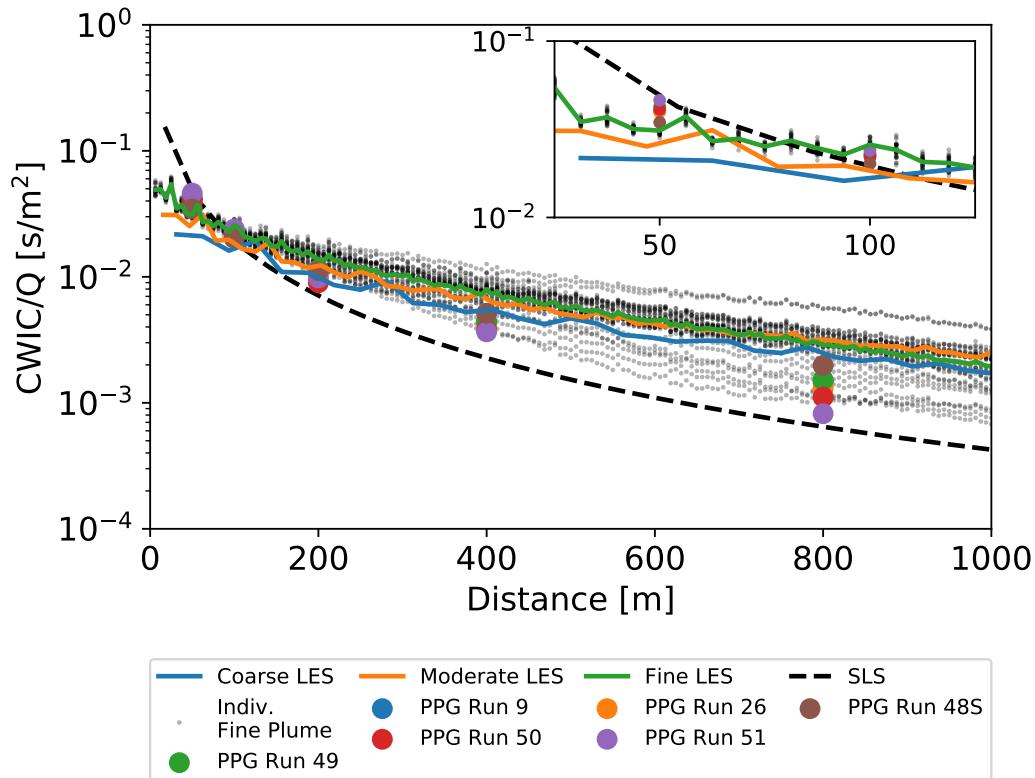
FIG. 1. Grid of normalized 10-minute averaged plume concentrations at 1.5 m within the SCBL and WCBL.



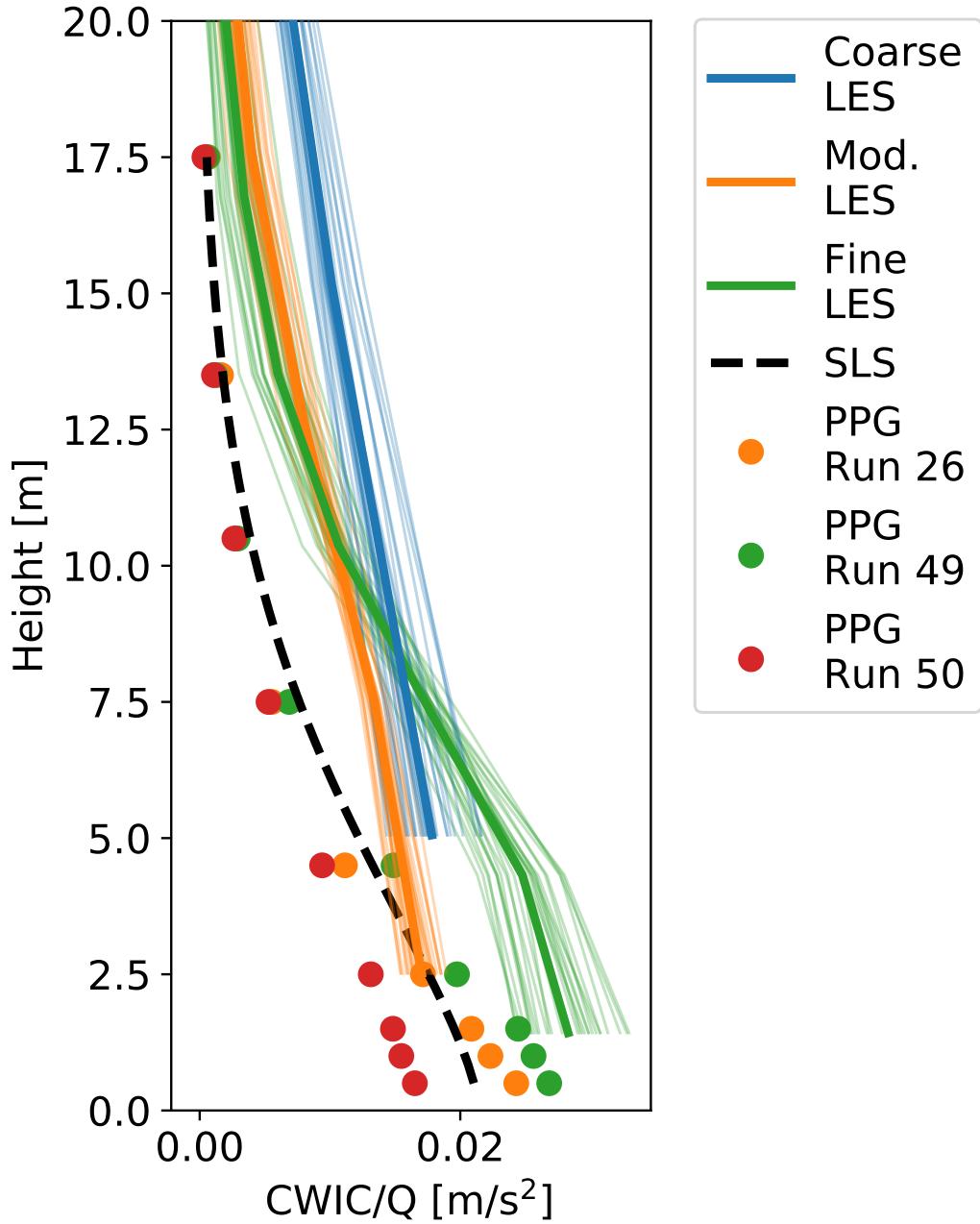
673 FIG. 2. SCBL observations and model predictions for the horizontal array. Ensemble average LES concentra-
 674 tions are shown as solid lines. SLS concentrations are calculated for $L=-6$ m, $u_* = 0.29$ m s⁻¹.



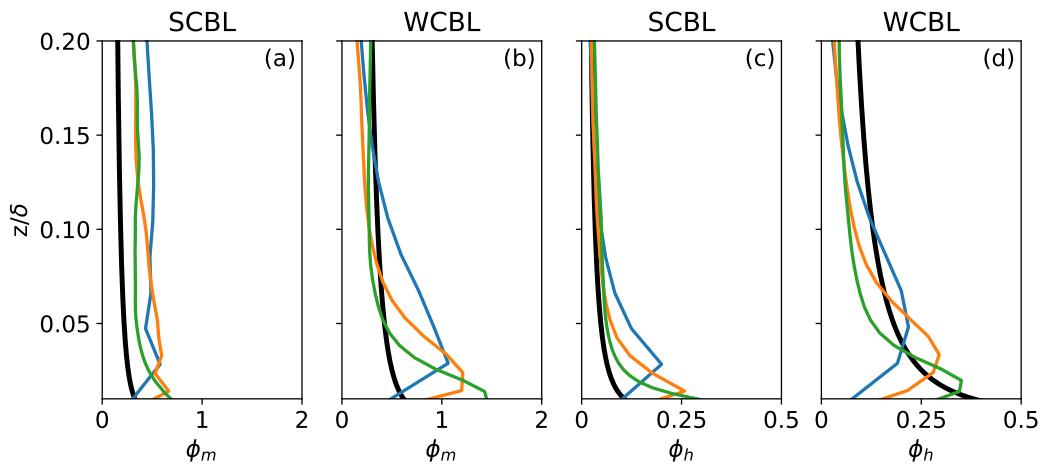
675 FIG. 3. SCBL observations and model predictions for the vertical array at 100-m downwind distance. Ensemble
 676 average LES concentrations are shown as solid lines, and individual plumes are shown as thin lines. SLS
 677 concentrations are calculated for $L=-6$ m, $u_* = 0.2$ m s⁻¹.



678 FIG. 4. WCBL observations and model predictions for the horizontal array. Ensemble average LES concen-
 679 trations are shown as solid lines. SLS concentrations are calculated for $L=-15$ m, $u_*=0.45$ m s⁻¹.



680 FIG. 5. WCBL observations and model predictions for the vertical array at 100-m downwind distance.
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