In order to control the spread of the 2019 novel coronavirus (COVID-19) \cite{1,2}, China imposed nationwide restrictions on the movement of its population (lockdown) after the Chinese New Year in January of 2020, leading to large reductions in economic activities and associated emissions \cite{3,4}. However, despite such large decreases in primary pollution \cite{3}, there were nonetheless several periods of heavy haze pollution in East China during the COVID-19 lockdown, raising questions about the well-established relationship between human activities and air quality \cite{4}. Here, using comprehensive in-situ measurements and chemical transport modeling, we show the haze events during the COVID lockdown were driven by enhancements of secondary pollution. In particular, large decreases in NOx emissions from transportation increased ozone and nighttime NO3 radical formation, and these increases in atmospheric oxidizing capacity in turn facilitated the formation of secondary inorganic and organic particulate matter. Our results, afforded by the tragic natural experiment of the COVID-19 pandemic, indicate that mitigation of Chinese haze pollution may depend upon a coordinated and balanced strategy for controlling multiple pollutants.

Efforts to control the spread of the 2019 novel coronavirus (COVID-19) have drastically reduced human activities worldwide \cite{1,2}. As the epicenter of the pandemic, China was the first country to shut down commercial activities, restrict travel, and require its people to stay home beginning in late January 2020 \cite{3,4}. These restrictions
are believed to have drastically decreased air pollutant emissions. For example, the TROPOMI instrument on the Sentinel 5P satellite observed approximately 65% decrease in tropospheric NO\textsubscript{2} columns over eastern China region compared to the same period in 2019\textsuperscript{3}. Previous studies have shown that heavy haze pollution in eastern China has primarily been driven by accumulated anthropogenic emissions together with rapid secondary production\textsuperscript{5-11}. Indeed, strict policies reduced anthropogenic air pollution emissions in China between 2013-2017 have been the main driver of decreases in PM\textsubscript{2.5} pollution (concentration of particles with an aerodynamic diameter smaller than 2.5 µm) in the country\textsuperscript{12-14}.

Yet despite large reductions in primary pollutant emissions, there were several heavy haze events over eastern China during the COVID-19 lockdown, which have seeded doubt among the Chinese public and policymakers regarding the current scientific understanding of the mechanisms of haze pollution. Here, using comprehensive measurements of the mass and chemical compositions of PM\textsubscript{2.5} as well as related trace gases, together with numerical model simulations, we show that secondary pollution during the COVID lockdown was unexpectedly enhanced as the result of imbalances in the reduction of primary emissions. Our findings thus underscore the importance of a carefully tailored and balanced strategy of emission control to reduce haze pollution in China.

Fig. 1 and Extended Data Fig.1 show the large change in pollutant emissions during the three weeks of the Chinese New Year holiday and the COVID lockdown that immediately followed (the lock period) in comparison with three weeks before the Chinese New Year (pre-period), based on the real-time measurement by air quality monitoring network. Nitrogen dioxide (NO\textsubscript{2}) levels decline sharply (by >60%) during the lockdown, consistent with both the reduction in human activities reflected by the transportation index (Figs. 1a, 1e) and satellite observations (Extended Data Fig. 2)\textsuperscript{3}. Other primary gaseous pollutants such as carbon monoxide (CO) and sulfur dioxide (SO\textsubscript{2}) show patterns similar to NO\textsubscript{2} (Extended Data Fig. 1). Yet concentrations of PM\textsubscript{2.5}—perhaps the most important air pollutant from a public health perspective—do not show the same decrease over eastern China, but instead increase in the area of Beijing, Tianjin and Hebei (BTH) and in the northern part of China (Figs. 1b, 1e). Meanwhile, ozone (O\textsubscript{3}), which is an important secondary pollutant in warmer months but generally of less concern in winter\textsuperscript{15,16}, was also substantially enhanced (in some cases by >100%) over all of eastern China (Fig. 1e). The distribution of pollutant levels shows systematic shifts in O\textsubscript{3} and NO\textsubscript{2} in the heavily populated BTH and Yangtze River Delta (YRD) areas of eastern China during the COVID lockdown, with less change in PM\textsubscript{2.5}, particularly in the BTH region (Extended Data Fig. 3). The ratio of PM\textsubscript{2.5}/CO, an indicator of secondary formation to primary emission, reveals a large-scale enhancement throughout northern China (Fig. 1d). The anomalies in all these species diminish near the end of February, when COVID restrictions began to be eased and people outside of the Hubei province started back to work (Fig. 1e)\textsuperscript{1}. 

2
The observed changes in primary pollutants are proportionate to reductions in emissions during the COVID lockdown. Emissions estimates based on up-to-date activity levels suggest an overall reduction of NOx about 60-70% in eastern China, 70-80% of which was related to road traffic and 20-25% from industry and power plants (Extended Data Table 1). Model simulations with fixed emission inventory (to exclude the possible influence of synoptic weather on air pollution)\textsuperscript{1,17,18} show the same reduction in primary pollutants and elevated O\textsubscript{3} and PM\textsubscript{2.5}, especially in the BTH region (Extended Data Fig. 4). Importantly, studies of previous changes in pollution emissions have never found such a pronounced increase in secondary pollutants like O\textsubscript{3} and PM\textsubscript{2.5}\textsuperscript{19-23}. But the reduction in primary pollutants during the COVID lockdown are also more extreme than those observed previously (e.g., during past holidays or special events such as the 2008 Olympics or the 2014 meeting of the Asia-Pacific Economic Cooperation in Beijing)\textsuperscript{9,24-27}.

Prior studies have suggested many different processes of secondary PM formation in northern China\textsuperscript{5-11}. To investigate the causes for the increase in PM\textsubscript{2.5} in eastern China during the COVID lockdown, we examine the chemical compositions of PM\textsubscript{2.5} from a regional network in northern and eastern China. Because elemental carbon (EC) is mainly from primary emissions\textsuperscript{28,29}, the relative concentration of secondary species, e.g. sulfate, nitrate, ammonium, and organic matter, provides information on secondary PM formation. As shown in Fig. 2a, secondary aerosol production during the lockdown was especially high in eastern China in comparison to the pre-period, especially in the BTH region (with a ratio of $\Delta$SNAO/$\Delta$EC as high as 1.5). Such secondary formation is also consistent with the PM\textsubscript{2.5}/CO ratios from the air quality monitoring network (Fig. 1d). The scatter plots of the $\Delta$SNAO/$\Delta$EC during the lockdown indicate an overall reduction in EC for all station (the ratio of EC\textsubscript{lock}/pre lower than 1 with a minimum as low as 0.4), but a substantial enhancement in secondary PM formation, shown by the much higher ratio of $\Delta$SNAO/$\Delta$EC\textsubscript{lock}/pre (Figs. 2b-d). For sulfate and OM, the ratio is particularly high, about 20%~ 100% (Figs. 2b, 2d), but for nitrate, the ratio is relatively low (-30%~60%), with some cities in the YRD below the 1:1 line (Fig1. c), indicating weakened nitrate formation. However, for the BTH region all secondary inorganic aerosols, e.g. sulfate, nitrate (and ammonium), and organics show significantly higher net chemical production (see the ratio of SNAO\textsubscript{lock}/pre over 1.0, marked in red), thereby a pretty high concentration of PM\textsubscript{2.5} in the northern China in Fig. 1.

The diurnal cycle of the $\Delta$SNAO/$\Delta$EC\textsubscript{lock}/pre ratio further reveals the main processes that influence secondary PM. As that shown in Extended Data Fig. 5c, the secondary production of all aerosol species in the BTH area were enhanced (positive values) most of the day during the lockdown period, with the enhancement extending from midnight to around 9:00 LT(local time) in the morning. The enhancement of secondary aerosols coincides with the rise in ozone (over 100%) and also O\textsubscript{3}*NO\textsubscript{2}, a proxy of NO\textsubscript{3} radical, which is a vital oxidant for nighttime secondary aerosol formation\textsuperscript{30-34}. However, in other regions, such as YRD, the diurnal cycle of these ratios shows an overall positive value, except for nitrate, which was consistent with the change of O\textsubscript{3}*NO\textsubscript{2} proxy
These results imply that—although secondary haze formation was generally enhanced during the COVID lockdown—the magnitude of such secondary aerosol production was different in different regions, depending on O\textsubscript{3} and the NO\textsubscript{3} radical. This suggests the great importance of NOx on both daytime and nighttime atmospheric chemistry\textsuperscript{31,35-38}.

We further conduct model simulations using the Weather Research and Forecasting model coupled with chemistry (WRF-Chem), based on an up-to-date emissions inventory estimated by dynamically-adjusted human activity levels (Extended Data Table 1; Extended Data Fig. 6). As aforementioned, we find significant decreases in atmospheric NO\textsubscript{x}, due mainly to lower vehicle emissions, could cause a substantial increase in the availability of O\textsubscript{3} and nighttime NO\textsubscript{3} radical in eastern China (Fig. 3). Moreover, the simulations reproduce the enhancements of both O\textsubscript{3} and NO\textsubscript{3} radical in eastern China, a region with intensive anthropogenic emissions of NO\textsubscript{x} \textsuperscript{39,40}, and a decrease in southern and southwestern China (Fig. 3a and Extended Data Fig. 7). The diurnal cycles from both model simulations (Fig. 3b) and observation-based diagnose (Extended Data Fig. 5) show that although the daytime increase were stronger but because of a stronger NO-titration effect existed in high NO environment in winter\textsuperscript{37,38,15}, particular in the northern China, the relative change in O\textsubscript{3} are relatively substantial in nighttime (Fig. 3b, Extended Data Fig. 7). Accordingly, the NO\textsubscript{3} radical increased a lot below the altitude of 1 km during the night, even if its precursor NOx declined sharply. The enhanced atmospheric oxidizing capacity accelerated the formation of sulfuric acid (H\textsubscript{2}SO\textsubscript{4}) in the boundary layer\textsuperscript{41,42}. Comparatively, such kind of enhancement are much stronger in BTH region than the YRD region.

By diagnosing the changes in HNO\textsubscript{3} and nitrate in WRF-Chem simulations (Extended Data Fig. 8), we can explain the differences in diurnal cycle of nitrate changes in BTH and other regions (Extended Data Fig. 5). In the BTH region, a substantial increase of HNO\textsubscript{3} occurs in the boundary layer almost the entire day with a positive nitrate formation occurring from night to early morning in the lower PBL. However, in the YRD region HNO\textsubscript{3} and nitrate show an overall decrease during the COVID lockdown (Extended data Fig. 8), consistent with observations (Fig. 2c, Extended Data Fig. 5f). As a result of the enhanced oxidizing capacity, the model shows a significant increase in sulfuric acid production during the daytime in eastern China (Fig. 3 and Extended Data Fig. 8). Observations at the SORPES station in Nanjing also show an increase in sulfuric acid concentration (~30 % increase) in both day and night (Extended Data Fig. 9). Consequently, vigorous new particle formation processes and stronger sulfate production were frequently observed at the station during the COVID lockdown. It also worth noting that a substantially higher OM fraction at night during the lockdown period was associated with stronger daytime NPF events (Extended Data Fig. 9), implying that the enhanced formation of secondary aerosols was due to a stronger nighttime oxidization by both O\textsubscript{3} and NO\textsubscript{3} radical\textsuperscript{31,32,43}, even under the condition of decreasing precursors (i.e., VOCs, SO\textsubscript{2}, and NOx). Extended Data Fig. 10 illustrates such non-linear relationship of reduced precursors causing enhanced oxidants and
secondary products.

The non-linear relationship between NOx reduction and oxidant enhancement is also evident in measurements from air quality monitoring networks. As shown in Extended Data Fig. 11, the relationship of O3 and night O3*NO2 proxy vs. NO2 in BTH and YRD regions shown distinct patterns during the COVID-lock and pre-COVID periods. A more substantial enhancement of oxidizing capacity, for both O3 and NO3 radical, took place in BTH region. By testing the sensitivity of our WRF-Chem simulations to different NOx reduction rate (from 10%-90%), we find a non-linear response in both O3 and NO3 radical in eastern China (Fig. 4). These results show that the sudden and precipitous reduction in NOx emissions could cause a substantial increase of O3, NO3 radical and hence a non-linear response of oxidation products like H2SO4, HNO3, and N2O5 and oxygen-containing organic compound. In BTH region, the enhancement is extremely significant because of the much higher NOx concentrations and weak incident solar radiation in northern China, which makes the NO-titration and O3-VOCs-NOx photochemistry more sensitive to the NOx reduction37. Fig. 4 also suggests that a further reduction of NOx (e.g., to 60%-70% in BTH and 50-60% in YRD, achieved from industrial or residuals sectors44) would reach a tipping-point for decreasing oxidizing capacity. Unfortunately, the emission reductions in both regions during the COVID lockdown were almost at the peak for secondary production. A concurrent reduction in VOCs could also have reduced the non-linear relationship. As that shown in Extended Data Fig. 12, synchronous VOCs reduction could have partly counterbalanced the enhancement of secondary pollution due to such non-linear response.

Our results show that the dramatic reductions in NOx and other air pollution emissions during China’s COVID lockdown led to substantial increases in O3, which in turn increased atmospheric oxidizing capacity and enhanced formation of secondary aerosols in eastern China. In China, winter haze and summer ozone pollution are the two major air quality challenges, but with different and separated control policies. To mitigate haze pollution, policies have focused on the reduction of primary emissions such as SO2, NOx, and NH312,45,46, while proposed efforts to reduce ozone pollution target VOCs16. Although regulation of SO2 and NOx emissions imposed since 2013 have successfully reduced haze pollution in eastern China12,13, our results suggest that the benefit of proposed further reductions in primary emissions44,47,48 might be offset by enhanced secondary formation of PM. Thus, the lockdown imposed to protect public health during the COVID pandemic has shown that efforts to further decrease PM2.5 pollution in eastern China may be more challenging than anticipated. Specifically, a non-linear tipping-point of NOx chemistry will require that future reductions in China’s haze pollution manage the balance of emitted species, with different consideration of VOCs and NOx ratios among different regions.
Fig. 1 | Changes in air quality before and during the COVID-19 lockdown in eastern China. a - d Differences in averaged concentrations of NO$_2$, PM$_{2.5}$, O$_3$ and PM$_{2.5}$/CO ratio before (1-24 January, 2020) and during (26 January - 17 February 2020) the COVID-19 lockdown periods at air quality monitoring stations in eastern China. e, Time series of normalized change in NO$_2$, O$_3$, PM$_{2.5}$ in eastern China (30 °N-40 °N, 110 °E-120 °E) and transportation index in China in January – February 2020. Note: The normalized ratio is calculated using time series with daily variation removed by EEMD filter and scaled with the first 10-day average.
Fig. 2 | Observational evidences of enhanced secondary PM production in eastern China during the COVID-19 period. a, Spatial distribution of the ratio of total secondary PM (sulfate, nitrate, ammonium and organic matters) verse element carbon (EC), i.e. (SNAO/EC)_{lock/pre}, between the COVID-lock and pre-COVID period. b-d, Scatter plots of the ratios of sulfate, nitrate, and organic matters during COVID-lock verse pre-COVID period as a function of the ratio of EC during the two periods. Note: Error bars present the standard deviations. Data of cities in the BTH and YRD regions are marked in red and blue edge line.
Fig. 3 | Modeling evidences of enhanced oxidizing capacity in eastern China due to the COVID-19 lockdown. 

a, Spatial distribution of averaged ozone enhancement by emission reduction due to the COVID-19 outbreak. 

b, Diurnal cycle of averaged vertical distribution of O\textsubscript{3}, NO\textsubscript{3} radical and sulfuric acid in eastern China (30 °N-40 °N, 110 °E-120 °E). Note: Black lines mean the contour of NO\textsubscript{3} radical (Unit: pptv), and red lines present that of sulfuric acid (Unit: pptv). The black and white bar near the x axis in b indicates the nighttime and daytime of a day.
Fig. 4 | Response of oxidizing capacity to NOx emission reduction in the BTH and YRD, China. a, b, Changes of NO$_2$, O$_3$, and NO$_3$ radical (upper panel) and gaseous products (proxy) (lower panel) as a function of emission reduction in the BTH and YRD regions, respectively. Note: BTH and YRD are defined as the domains of (36°N-40°N, 114°E-118°E) and (29-33°N, 118°E-112°E). Arom_op means total oxidization products of aromatics.
References


Liu, M. X. et al. Ammonia emission control in China would mitigate haze


Methods

Observational datasets and analysis method

Measurements on hourly concentrations of air pollutants are analyzed to understand air quality change due to the most extreme COVID-19 lockdown in China. Concentration of NO\textsubscript{2}, O\textsubscript{3}, CO, SO\textsubscript{2}, PM\textsubscript{2.5} and PM\textsubscript{10} at more than 1500 stations are recorded hourly mainly based on Thermo Scientific samplers and analyzers. The geographic locations of these station are marked in Fig. 1 (a-d). All these ground-based observations are archived at air monitoring data center of Ministry of Ecology and Environment of the People’s Republic of China (http://datacenter.mep.gov.cn), which are collected to derive the spatiotemporal variations in this study. As aforementioned, we divided the first two month of 2020 into two distinct time period according to the sharp drop of transportation activities due to the hit of COVID-19. The concentration level of multiple air pollutants during Pre-COVID and COVID-lock period. To clearly demonstrate the air quality response variations, ensemble empirical mode decomposition (EEMD) is applied to decompose observational data and to derive the low frequency signals (timescale larger than 2 days).

In addition, PM\textsubscript{2.5} chemical compositions are recorded at more than air pollution observation network developed by environment monitoring centers in developed provinces and cities in eastern China. In Nanjing, besides routinely measured aerosol and its chemical compositions, we also conducted measurements on particle size distribution, VOC and other trace gases at the Stations for Observing Regional Processes of the Earth System (SORPES). The SORPES station is a cross-disciplinary research and experiment platform that was established in 2011 to understand the impact of human activities in the rapidly urbanized and industrialized eastern China region\textsuperscript{13}. A scanning mobility particle sizer (SMPS), Aerodyne Soot Particle Aerosol Mass Spectrometer (SP-AMS), Time-of-Flight Aerosol Chemical Speciation Monitor (TOF-ACSM), compact IONICON PTF-TOF-MS trace VOC analyzer, and Aerodyne Research Long-Tof-CIMS were used to measure aerosol size distribution, fine particle chemical compositions, VOCs speciation, and sulfuric acid, respectively, at the station.

TROPOspheric Monitoring Instrument (TROPOMI) on board the Copernicus Sentinel-5 Precursor satellite provides retrievals of NO\textsubscript{2} column amount\textsuperscript{49}, which are also employed to illustrate the spatial pattern and temporal variation of pollution in China during COVID-19 lockdown. The retrieved column burden is then averaged and compared in different regions for Pre-COVID and COVID-lock period.

Emission reduction estimation due to the lockdown control

Emission reduction is estimated using the bottom-up inventory model of Multi-resolution Emission Inventory for China (MEIC), developed by Tsinghua University\textsuperscript{47}. The MEIC model uses the technology-based approach calculating air pollutant emissions from more than 700 anthropogenic sources in China, which can be categorized into five source sectors: power, industry, residential, transportation, and agriculture. The most up-to-date version provides China’s anthropogenic emissions
data until the year 2019.

To estimate emission reductions due to COVID-19 lockdown, we update China’s emissions data to Jan and Feb 2020 based on dynamic economic and industrial activity levels. The thermal power generation in the first two months of 2020 was 8.9% lower than that in 2019, while China generated 1.7% more thermal power in January and February 2019 than in 2018 (Natural Bureau of Statistics, http://www.stats.gov.cn). We then assume that the difference in the growth rates between 2019 and 2020 was the influence of COVID-19 lockdown on the power sector emissions. The same approach is applied to the industrial sector. For example, cement production in January and February 2020 was 29.5% lower than that in 2019, while China produced 0.5% more cement in January and February 2019 than in 2018. Cement emissions were then estimated to be reduced by 30% due to COVID-19. For the residential sector, emissions from the commercial use of boilers and stoves in the urban region were eliminated since the lockdown measure implemented, while emissions from residential heating and cooking in both urban and rural areas were assumed not affected. For the transport sector, decline in national traffic volume was estimated at 70% during the COVID-19 lockdown, according to the transportation index data. Activity level of off-road equipment such as construction machines was assumed approaching zero during the COVID-19 lockdown. As the result, on-road emissions declined by 70%, and the off-road emissions became zero during the lockdown. The detailed estimation of provincial emission reduction ratio of main trace gases and primary PM are presented in Extended Data Table 1. Evaluation of WRF-Chem simulations based on this inventory are given in Extended Data Fig. 6.

**Regional air quality modeling**

To quantitatively understand the response of air quality, regional coupled dynamical and chemical simulations are conducted based on WRF-Chem model (Weather Research and Forecasting model coupled with Chemistry, version 3.7.1). WRF-Chem model is a state-of-the-art meteorology/chemistry model developed that includes a variety of coupled physical and chemical processes such as emission and deposition of pollutants, advection and diffusion, gaseous and aqueous chemical transformation, aerosol chemistry and dynamics, etc. The model has been widely used and evaluated against measurements, which has been further improved by optimizing parametrization of aqueous and heterogeneous chemistry in Asia. The model domain with a grid resolution of 20 km covered the eastern China and its surrounding areas to get synoptic forcing. There are 30 vertical layers from the ground level to the top pressure of 50 hPa, in which more than ten layers are settled under 1 km to better describe boundary layer processes. The simulation is conducted from 1 December 2019 to 5 March 2020. The initial and boundary meteorological conditions were the NCEP global final analysis (FNL) data. NCEP Automated Data Processing (ADP) surface and global upper air observational weather data of wind, temperature and moisture is assimilated to better characterize the regional transport pattern of air pollution. The Yonsei University planet boundary layer (YSU PBL) scheme is applied to parameterize the boundary layer processes in this simulation. Other key parameterization options are the Noah land...
surface scheme to describe the land-atmosphere interactions, the Lin microphysics scheme with the Grell cumulus parameterization to reproduce the cloud and precipitation processes, and the RRTMG short- and long-wave radiation scheme. For the numerical representation of atmospheric chemistry, Carbon-Bond photochemical mechanism combined with MOSAIC (Model for Simulating Aerosol Interactions and Chemistry) aerosol module.

Both natural and anthropogenic emissions were included for the regional WRF-Chem modeling in the present work. MEGAN (Model of Emissions of Gases and Aerosols from Nature) module embedded in WRF-Chem model is used to calculate biogenic emissions online. It estimates the net emission rates of isoprene, monoterpenes and other biogenic VOCs from terrestrial ecosystems into the above-canopy atmosphere. Soil-derived dust emission is characterized by GOCART emission schemes. Anthropogenic emissions from power plants, residential combustion, industrial processes, on-road mobile sources and agricultural activities were derived from the MEIC, which is described above. Emissions outside China was obtained from the MIX Asian emission inventory \(^52\). We then use the up-to-date emission reduction ratio considering the activity level change due to COVID lockdown, as mentioned above, to quantitatively understand the emission-triggered perturbations in air pollution. Furthermore, a number of WRF-Chem simulations are carried out based on different emission scenarios with a series of NOx and VOC emission.

**Data availability**

All the observation and simulation data that support the findings of this study are available at Figshare data publisher. Daily satellite retrievals of NO\(_2\) column amount are openly accessible at http://www.temis.nl/airpollution/no2col/data/tropomi. The daily transportation index is provided by Baidu migration dataset (https://qianxi.baidu.com). The simulation data used in this study are stored on high performance computing center of Nanjing University and can be made available from the corresponding author upon request.

**Code availability**

Data processing techniques are available on request from the corresponding author. The source code of WRF-Chem model is archived on UCAR data repository (http://www2.mmm.ucar.edu/wrf/users/download).

**Acknowledgements**

This work was funded by the National Natural Science Foundation of China (41725020, 41922038, 41625020, 91744311, and 41921005), the Ministry of Science and Technology of the People's Republic of China (2016YFC0200500) and Jiangsu Provincial Fund on air pollution mitigation. We thank colleagues at Environmental
Monitoring Centers of cities in eastern China for their contributions on the field measurements.

Author Contributions.
A.D, X.H., Q. Z. and K.H. conceived the study and led the overall scientific questions. X. H and A.D. made the data analysis and modeling studies. J. G., F. C., Q. Fu and W. Q. collected and processes the chemical composition measurement data. B. Z. and D.T. developed the base year emission inventory and estimated emission reduction during the COVID-19 lockdown. X. H., A.D., Q. Z. and S.D. wrote the paper with contribution from all coauthors.

Competing interests
The authors declare no competing interests.

Correspondence and requests for materials should be addressed to A. D. (dingaj@nju.edu.cn) or Q. Z. (qiangzhang@tsinghua.edu.cn).
References:


Extended Data Table 1 | Estimation of provincial emission reduction ratio (%) of CO, NOx, SO2, VOC, PM$_{2.5}$, BC, OC due to COVID-19 lockdown in China.

<table>
<thead>
<tr>
<th>Province</th>
<th>CO</th>
<th>NOx</th>
<th>SO2</th>
<th>VOC</th>
<th>PM$_{2.5}$</th>
<th>BC</th>
<th>OC</th>
</tr>
</thead>
<tbody>
<tr>
<td>Beijing</td>
<td>22%</td>
<td>45%</td>
<td>26%</td>
<td>45%</td>
<td>18%</td>
<td>46%</td>
<td>8%</td>
</tr>
<tr>
<td>Tianjin</td>
<td>21%</td>
<td>38%</td>
<td>20%</td>
<td>41%</td>
<td>14%</td>
<td>22%</td>
<td>6%</td>
</tr>
<tr>
<td>Hebei</td>
<td>15%</td>
<td>45%</td>
<td>16%</td>
<td>36%</td>
<td>12%</td>
<td>17%</td>
<td>5%</td>
</tr>
<tr>
<td>Shanxi</td>
<td>18%</td>
<td>40%</td>
<td>20%</td>
<td>33%</td>
<td>16%</td>
<td>19%</td>
<td>10%</td>
</tr>
<tr>
<td>Inner Mongolia</td>
<td>14%</td>
<td>29%</td>
<td>15%</td>
<td>34%</td>
<td>13%</td>
<td>16%</td>
<td>6%</td>
</tr>
<tr>
<td>Liaoning</td>
<td>21%</td>
<td>40%</td>
<td>28%</td>
<td>36%</td>
<td>16%</td>
<td>28%</td>
<td>8%</td>
</tr>
<tr>
<td>Jilin</td>
<td>16%</td>
<td>39%</td>
<td>23%</td>
<td>34%</td>
<td>13%</td>
<td>18%</td>
<td>5%</td>
</tr>
<tr>
<td>Heilongjiang</td>
<td>17%</td>
<td>37%</td>
<td>27%</td>
<td>28%</td>
<td>13%</td>
<td>15%</td>
<td>7%</td>
</tr>
<tr>
<td>Shanghai</td>
<td>35%</td>
<td>48%</td>
<td>42%</td>
<td>45%</td>
<td>34%</td>
<td>54%</td>
<td>42%</td>
</tr>
<tr>
<td>Jiangsu</td>
<td>23%</td>
<td>50%</td>
<td>26%</td>
<td>41%</td>
<td>16%</td>
<td>35%</td>
<td>7%</td>
</tr>
<tr>
<td>Zhejiang</td>
<td>41%</td>
<td>50%</td>
<td>29%</td>
<td>45%</td>
<td>30%</td>
<td>49%</td>
<td>20%</td>
</tr>
<tr>
<td>Anhui</td>
<td>14%</td>
<td>56%</td>
<td>22%</td>
<td>31%</td>
<td>11%</td>
<td>22%</td>
<td>4%</td>
</tr>
<tr>
<td>Fujian</td>
<td>29%</td>
<td>51%</td>
<td>30%</td>
<td>42%</td>
<td>19%</td>
<td>31%</td>
<td>7%</td>
</tr>
<tr>
<td>Jiangxi</td>
<td>24%</td>
<td>53%</td>
<td>21%</td>
<td>43%</td>
<td>19%</td>
<td>30%</td>
<td>9%</td>
</tr>
<tr>
<td>Shandong</td>
<td>23%</td>
<td>50%</td>
<td>25%</td>
<td>39%</td>
<td>19%</td>
<td>35%</td>
<td>9%</td>
</tr>
<tr>
<td>Henan</td>
<td>23%</td>
<td>57%</td>
<td>22%</td>
<td>41%</td>
<td>18%</td>
<td>35%</td>
<td>8%</td>
</tr>
<tr>
<td>Hubei</td>
<td>19%</td>
<td>55%</td>
<td>23%</td>
<td>35%</td>
<td>16%</td>
<td>23%</td>
<td>10%</td>
</tr>
<tr>
<td>Hunan</td>
<td>22%</td>
<td>51%</td>
<td>25%</td>
<td>36%</td>
<td>20%</td>
<td>24%</td>
<td>15%</td>
</tr>
<tr>
<td>Guangdong</td>
<td>38%</td>
<td>50%</td>
<td>33%</td>
<td>46%</td>
<td>27%</td>
<td>42%</td>
<td>13%</td>
</tr>
<tr>
<td>Guangxi</td>
<td>24%</td>
<td>50%</td>
<td>28%</td>
<td>39%</td>
<td>17%</td>
<td>27%</td>
<td>5%</td>
</tr>
<tr>
<td>Hainan</td>
<td>24%</td>
<td>44%</td>
<td>25%</td>
<td>36%</td>
<td>14%</td>
<td>25%</td>
<td>4%</td>
</tr>
<tr>
<td>Chongqing</td>
<td>18%</td>
<td>53%</td>
<td>32%</td>
<td>37%</td>
<td>14%</td>
<td>20%</td>
<td>4%</td>
</tr>
<tr>
<td>Sichuan</td>
<td>16%</td>
<td>50%</td>
<td>27%</td>
<td>33%</td>
<td>9%</td>
<td>15%</td>
<td>3%</td>
</tr>
<tr>
<td>Guizhou</td>
<td>24%</td>
<td>39%</td>
<td>25%</td>
<td>30%</td>
<td>22%</td>
<td>25%</td>
<td>20%</td>
</tr>
<tr>
<td>Yunnan</td>
<td>24%</td>
<td>51%</td>
<td>25%</td>
<td>41%</td>
<td>18%</td>
<td>21%</td>
<td>8%</td>
</tr>
<tr>
<td>Tibet</td>
<td>16%</td>
<td>35%</td>
<td>15%</td>
<td>35%</td>
<td>14%</td>
<td>14%</td>
<td>5%</td>
</tr>
<tr>
<td>Shaanxi</td>
<td>19%</td>
<td>45%</td>
<td>18%</td>
<td>34%</td>
<td>13%</td>
<td>22%</td>
<td>5%</td>
</tr>
<tr>
<td>Gansu</td>
<td>13%</td>
<td>47%</td>
<td>16%</td>
<td>29%</td>
<td>9%</td>
<td>13%</td>
<td>3%</td>
</tr>
<tr>
<td>Qinghai</td>
<td>23%</td>
<td>46%</td>
<td>22%</td>
<td>39%</td>
<td>20%</td>
<td>20%</td>
<td>7%</td>
</tr>
<tr>
<td>Ningxia</td>
<td>24%</td>
<td>36%</td>
<td>24%</td>
<td>39%</td>
<td>20%</td>
<td>23%</td>
<td>8%</td>
</tr>
<tr>
<td>Xinjiang</td>
<td>16%</td>
<td>35%</td>
<td>15%</td>
<td>35%</td>
<td>14%</td>
<td>14%</td>
<td>5%</td>
</tr>
</tbody>
</table>
Extended Data Fig. 1 | Air quality before and during the COVID-19 epidemic in eastern China. **a,** Time series of daily PM$_{2.5}$, NO$_2$ and transportation index in the BTH (upper panel) and the YRD (lower panel) during January-February 2020. **b,** Time series of daily O$_3$, SO$_2$, and CO in the BTH (upper panel) and the YRD (lower panel) during January-February 2020.
Extended Data Fig. 2 | Satellite retrievals of NO\textsubscript{2} in eastern China before and during the COVID-19 epidemic. a-b, Spatial distribution of averaged Trop-OMI tropospheric NO\textsubscript{2} column in eastern China during the pre-COVID and COVID-lock periods. c-d, changes in Trop-OMI tropospheric NO\textsubscript{2} column in the BTH and YRD regions.
Extended Data Fig. 3 | Comparison of frequency distribution of air pollutants in eastern China due to COVID-19 lockdown. **a-c**, Normalized frequency distributions of NO$_2$, O$_3$, and PM$_{2.5}$ in the BTH region during the COVID-lock (red lines) and pre-COVID (black lines) periods. **d-f**, Normalized frequency distributions of NO$_2$, O$_3$, and PM$_{2.5}$ in the YRD region.
Extended Data Fig. 4 | Changes in air pollution indicated by comparison between observation and simulations. a, b, Time series of daily ratio of observation verse WRF-Chem simulated O$_3$ with fixed emission inventory in the BTH and YRD region. c,d, same as a and b but for NO$_2$. e,f, same as a and b but for PM$_{2.5}$. 
Extended Data Fig. 5 | Diurnal cycle of secondary verse primary PM$_{2.5}$ ratio and changes in O$_3$ and NO$_3$ proxy in BTH and YRD. a, b, Changes in O$_3$ and O$_3$*NO$_2$ proxy between the COVID-lock and pre-COVID periods in BTH and YRD. c, d, changes in SNAO/EC ratio between the COVID-lock and pre-COVID periods in BTH and YRD. Note: N, S, A and OM means nitrate, sulfate, ammonium and organic matter, respectively.
Extended Data Fig. 6 | Evaluation of WRF-Chem simulations with ambient measurements by the national air quality monitoring networks. a-d, Scatter plots of simulation verse observation for O$_3$, NO$_2$, CO and PM$_{2.5}$. Note that the blue and red markers presents simulations and observations during the COVID-lock and pre-COVID periods, respectively. Mean bias of modeled concentrations are labeled in the bottom right corner of each subplot.
Extended Data Fig. 7 | Modeling evidences of enhanced oxidization capacity during the COVID-19 lockdown period in BTH and YRD, China. 

**a-b,** Spatial distribution of averaged changes in NO\textsubscript{3} radical and H\textsubscript{2}SO\textsubscript{4} between the COVID-lock and pre-COVID periods (Ctrl minus Pre) in BTH and YRD, China. 

**c, d,** Diurnal cycle of averaged vertical distribution of O\textsubscript{3} and boundary layer (0-1 km) O\textsubscript{3} column between the two period in BTH and YRD. 

**e, f,** Diurnal cycle of NO\textsubscript{3} radical and H\textsubscript{2}SO\textsubscript{4} between the two period in BTH and YRD. Note: The black and white bar near the x axis in **c-f** indicates the nighttime and daytime of a day.
Extended Data Fig. 8 | Comparison of simulated diurnal cycles of vertical distribution of \( \text{HNO}_3 \) and nitrate in BTH and YRD. a, b, WRF-chem simulated changes in diurnal cycle of \( \text{HNO}_3 \) vertical distribution in BTH and YRD. c, d, same as a and b but for nitrate.
Extended Data Fig. 9 | Observations of trace gases and aerosols at the SORPES station in Nanjing during the pre-COVID and COVID-lock periods. a, Diurnal cycle of O₃ and sulfuric acid (SA) during the two periods. b-c, Diurnal cycle of aerosol size during pre-COVID and COVID-lock periods. d, Diurnal cycles of VOCs species at the SORPES station. (Ben- Benzene, Tol-Toluene, Arom- C8-C10 aromatics). e-f, Diurnal cycles of percentage of PM₁ speciation the pre-COVID and COVID-lock periods.
Extended Data Fig. 10 | Conceptual model showing how secondary formation offsets the reduction of primary pollutions in China with different processes during the day and night. Note: The upper and lower panels, indicated by moon and sun, represent nighttime and daytime processes, respectively. Blue downward arrows mean the reduction of emission or decrease of concentrations in the atmosphere, while the red upward arrows mean enhancement of production in the atmosphere. The length of these arrows indicated the degree of change.
Extended Data Fig. 11 | Observed non-linear relationship between main oxidants and NO$_2$ caused by the COVID-19 lockdown in China. **a, b,** Scatter plots of O$_3$ and proxy of NO$_3$ radical (NO$_2$*O$_3$) verse NO$_2$ between the COVID-lock and pre-COVID periods in BTH of China. **c, d,** Same as **a** and **b** but for the YRD region.
Extended Data Fig. 12 | Response of oxidization capacity to VOCs emission reductions in the BTH and YRD, China. a, b, Changes of NO\textsubscript{2}, O\textsubscript{3}, and NO\textsubscript{3} radical (upper panel) and gaseous products (proxy) (lower panel) as a function of emission reductions (NO\textsubscript{x}: 50%, VOCs: 10-90%) in the BTH and YRD regions, respectively. Note: BTH and YRD are defined as the domains of (36˚N-40˚N, 114˚E-118˚E) and (29-33˚N, 118˚E-112˚E). Arom\textsubscript{op} means total oxidization products of aromatics.