

1 An operational framework for large-area
2 mapping of active cropland and short-term
3 fallows in smallholder landscapes using
4 PlanetScope data

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12 **Abstract**

13 Cropland mapping in complex smallholder landscapes is challenged by complex and
14 fragmented landscapes, labor-intensive and unmechanized land management causing high
15 within-field variability, rapid dynamics in shifting cultivation systems, and substantial
16 proportions of short-term fallows. To overcome these challenges, we here present a large-area
17 mapping framework to identify active cropland and short-term fallows in smallholder landscapes
18 for the 2020/2021 growing season at 4.77 m spatial resolution. Our study focuses on Northern
19 Mozambique, an area comprising 381,698 km². The approach is based on Google Earth Engine
20 and time series of openly available PlanetScope mosaics made available through the NICFI data
21 program. We conducted multi-temporal co-registration of the PlanetScope data using seasonal
22 Sentinel-2 base images and derived consistent and gap-free seasonal time series metrics to classify
23 active cropland and short-term fallows. An iterative active learning framework based on Random
24 Forest class probabilities was used for training rare classes and uncertain regions. The map was
25 accurate (area-adjusted overall accuracy 88.6% ± 1.5%), with the main error type being the
26 commission of active cropland. Error-adjusted area estimates of active cropland extent (61,799.5
27 km² ± 4,252.5 km²) revealed that existing global and regional land cover products tend to under-
28 , or over-estimate active cropland extent, respectively. Short-term fallows occupied 13% of the
29 mapped cropland, with consolidated agricultural regions showing the highest shares of short-
30 term fallows. Our approach relies on openly available PlanetScope data and cloud-based
31 processing in Google Earth Engine, which minimizes financial constraints and maximizes
32 replicability of the methods. All code and maps will be made available for further use.

33 **Keywords**

34 Mozambique; Sub-Saharan Africa; Shifting Cultivation; Agriculture; Land Use; Sentinel-2; Co-
35 Registration; Time Series; Google Earth Engine

36 **Highlights**

- 37 • PlanetScope mosaics used for mapping smallholder agriculture in Mozambique
- 38 • Multi-temporal co-registration improved with seasonal Sentinel-2 reference images
- 39 • Active learning based on Random Forest class probabilities
- 40 • Accurate map of active cropland and short-term fallows at 4.77 m resolution

41 **Introduction**

42 Smallholder crop production is a key livelihood in many tropical regions and has a
43 substantial relevance for food security. Simultaneously, agricultural expansion has been
44 identified as a principal driver of land use change, including the clearing of forests, savannas, and
45 other ecosystems. Timely and spatially detailed maps of cropland extent are essential for
46 assessing the productivity of agricultural land by evaluating crop types and yields (Jin et al., 2019;
47 Lambert et al., 2018), particularly in regions vulnerable to extreme climatic events (Nakalembe et
48 al., 2021). Likewise, data on the spatial distribution of croplands are essential for monitoring land
49 use change (Bey et al., 2020) and associated carbon emissions (Sy et al., 2019). Earth observation
50 technologies offer great potential for providing timely wall-to-wall cropland maps. However,
51 cropland mapping is not fully operational in complex smallholder landscapes, in contrast to
52 consolidated agricultural systems (d'Andrimont et al., 2021; Song et al., 2021). Consequently,
53 information on cropland distribution is either scarce or does not meet the particularly high
54 requirements for timeliness, spatial resolution, and thematic detail for smallholder-dominated
55 agricultural landscapes.

56 Key challenges for smallholder cropland mapping are 1) spatial fragmentation, 2) within-
57 field heterogeneity, 3) rapid dynamics, and 4) following as an integral component of the
58 agricultural system. These challenges are particularly complicating mapping in vast swaths of
59 Sub-Saharan Africa (SSA), where agricultural landscapes are highly fragmented and the majority
60 of crop fields are smaller than 0.64 ha (Lesiv et al., 2019). Furthermore, land management in
61 smallholder landscapes of SSA is unmechanized, labor-intensive, and nearly free of chemical
62 inputs. Burning, manual land clearing and preparation, heterogenous management skills and

63 labor inputs, as well as the presence of shading trees and shelters, lead to high within-field
64 heterogeneity in terms of vegetation types and cover density. Smallholder agriculture in SSA
65 often undergoes shifting cultivation, with frequent rotations between active cropland and short-
66 term fallows, resulting in dynamic, mosaic landscapes with substantial fractions of fallow land in
67 various stages and conditions (Tong et al., 2020). To be the most useful, monitoring thus requires
68 timely maps of high granularity allowing for untangling actively used cropland from fallows, in
69 order to precisely assess the land area used for agricultural production, as well as the overall land
70 footprint of agriculture in a landscape.

71 Approaches based on openly available Landsat and Sentinel-2 (S2) (30-10 m) imagery,
72 yield timely and high-resolution maps of cropland extent at the global (Karra et al., 2021 - 2021;
73 Zanaga et al., 2021), national (Estes et al., 2021; Kerner et al., 2020) and sub-national level (Bey et
74 al., 2020; Ibrahim et al., 2021). In the most fragmented landscapes of SSA, however, the spatial
75 resolution of these data still hampers the accurate identification of cropland, as indicated in
76 spatially explicit accuracy assessments (Tsendbazar et al., 2021) and regional comparisons of
77 mapped cropland area (Lambert et al., 2016; Wei et al., 2020). These spatial complexities highlight
78 the need for approaches based on image data with a spatial resolution below 10 m. Furthermore,
79 the high prevalence of cloud cover in the growing seasons of tropical and sub-tropical regions
80 often limits data availability for sensors with near-weekly revisit intervals. Pre-processing
81 approaches that aggregate images across multiple years may help to mitigate these limitations,
82 but these may obfuscate the spatio-temporal patterns of cropland distribution due to the rapid
83 changes. Lastly, fallows are often included in generic cropland definitions of existing map
84 products (Nabil et al., 2021), likely due to limitations in spatial resolution and minimum mapping
85 units, spectral-temporal similarities between fallows and other land covers, and challenges for

86 reference data collection. With a few exceptions, knowledge about the spatial distribution of
87 fallow land in smallholder systems of SSA is currently only available at the farm level.

88 The PlanetScope satellite constellation provides visible to near infrared imagery at 3-5 m
89 resolution at near-daily intervals (Roy et al., 2021) and thereby offers novel opportunities to
90 overcome persisting challenges for cropland mapping in smallholder agriculture. Recent studies
91 propose enhanced pre-processing routines to improve the quality of PlanetScope data (Scheffler
92 et al., 2017), and demonstrate the use of these data for assessments of vegetation phenology or
93 water use (Aragon et al., 2021; Cheng et al., 2020). These applications are promising, but mostly
94 have experimental character, focusing on local to regional scale study sites, partly because
95 PlanetScope image time series across large regions are costly to obtain. Due to constraints in
96 financial and technical resources in many smallholder-dominated countries, the computational
97 infrastructure, tools, and data and to produce such maps need to be accessible at a low cost. In
98 this regard, Norway's International Climate and Forest Initiative (NICFI) launched the NICFI
99 data program (Planet Labs Inc., 2020b), releasing 4.77 m 4-band PlanetScope mosaics across the
100 world's tropics at monthly intervals from September 2020 onwards. These data were made
101 available within Google Earth Engine cloud computing platform (Gorelick et al., 2017), which
102 offers novel opportunities for large-area mapping at high granularity. To date, however, the
103 potential of these data for characterizing complex smallholder landscapes has not been explored.

104 We here advance the current state-of-the-art by presenting an operational framework for
105 mapping active and fallow cropland across a large region using openly available PlanetScope
106 time series in conjunction with S2 to enhance multi-temporal co-registration consistency. We
107 focus our analysis on Northern Mozambique because the region is particularly heterogeneous

108 and dynamic, and because recent and accurate information on active and fallow cropland extent
109 is scarce. Our objectives are to:

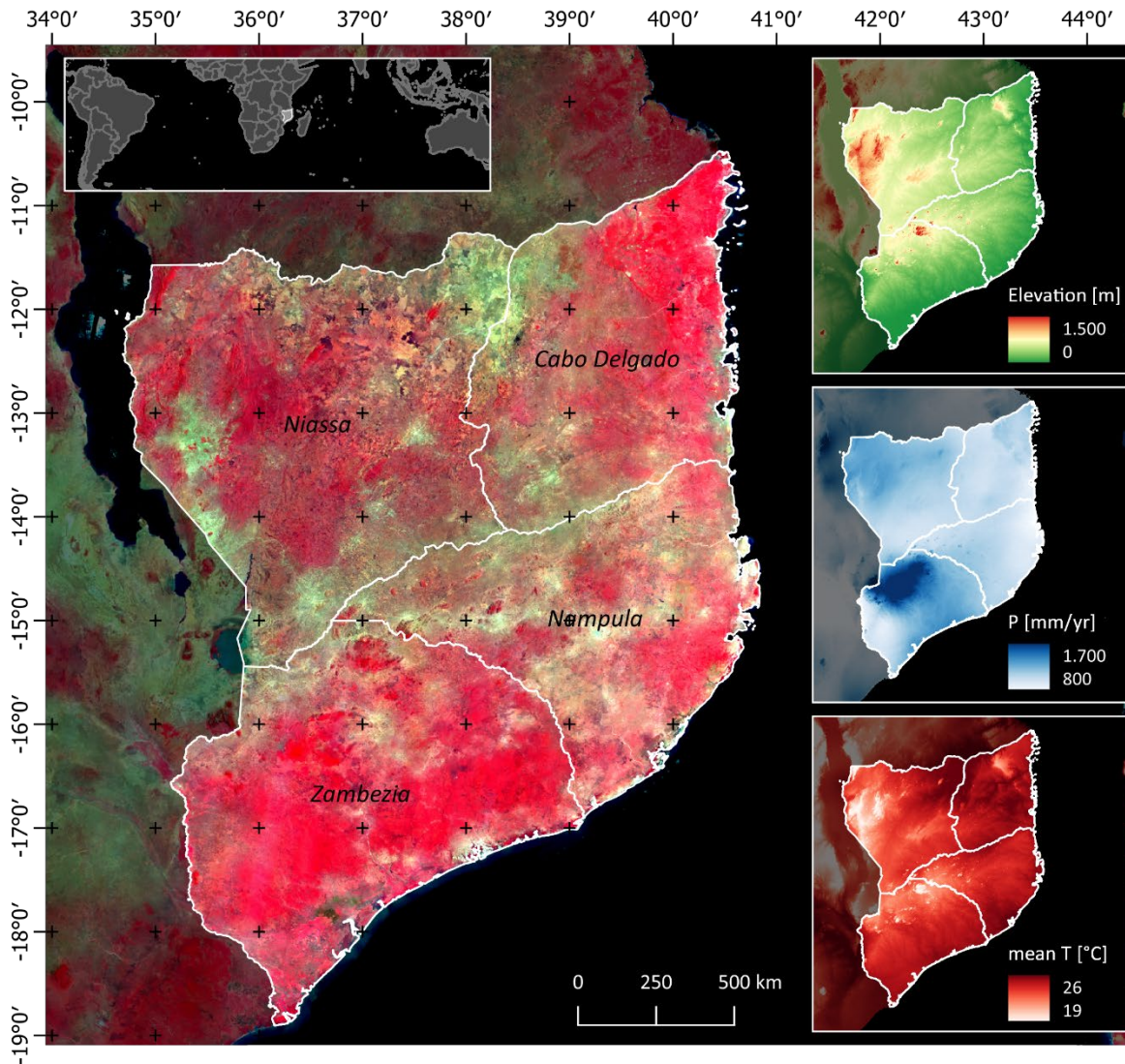
- 110 1) Present an operational framework based on openly available PlanetScope time series
111 for large area land cover mapping in Google Earth Engine.
- 112 2) Map and assess the distribution of active and short-term fallow cropland in Northern
113 Mozambique for the cropping season 2020/2021.
- 114 3) Assess differences in cropland extent derived from regional and global land cover
115 products with the presented map and unbiased area estimates.

116 **Data & Methods**

117 *Study area*

118 Our study area comprises four provinces in Northern Mozambique (Cabo Delgado,
119 Nampula, Niassa, and Zambezia), excluding lake Malawi, encompassing a total area of 381,698
120 km² (Figure 1). The region is dominated by Eastern Miombo woodlands covering vast parts of
121 the western and central study region. Total annual precipitation ranges between 800 and
122 1700mm/year, with the highest levels in the northern parts of Zambezia province and the
123 Lichinga plateau in the western parts of Niassa province. Ferrasols and Lixisols are the dominant
124 soil types. The west-to-east elevation gradient drops from 1,500 m on the Lichinga plateau
125 towards the eastern coast with some local topographical features. Agriculture in the region is
126 dominated by low-intensity smallholder agriculture with high labor demands, little to no
127 fertilizer or pesticide inputs, and frequent fallow rotations (Leonardo et al., 2018). Following a
128 vivid land use history, rapid land change dynamics occurred through smallholder expansion and

129 large-scale agricultural investments in the post-2000 era, particularly surrounding the Nacala
130 corridor, which links the major inland production regions with the Nacala harbor at the coast
131 (Bey et al., 2020; Kronenburg García et al., 2021).



133 Figure 1: Study region in Northern Mozambique. PlanetScope Mosaic for September 2020 in false-color
134 infrared (R: near infrared, G: red, B: green). Insets show elevation, total annual precipitation, and mean
135 temperature from BioClim data (Fick and Hijmans, 2017).

136 *Workflow*

137 The workflow of our study involves 1) a pre-processing framework to create consistent
138 seasonal input features from the PlanetScope surface reflectance mosaics provided through Level
139 1 access of the NICFI data program (hereafter PlanetScope mosaics), 2) an iterative scheme for
140 training data collection, model parametrization, and classification, 3) an area-adjusted accuracy
141 assessment and unbiased area estimation, and 4) a comparison of cropland extent and
142 distribution mapped here with regional and global LC products. All processing steps to create
143 the map were conducted using the Google Earth Engine Python API, post-processing was
144 performed on local machines with Python and R, reference data were labeled within QGIS.

145 *Data & pre-processing*

146 We mitigated multitemporal co-registration issues in the time series by matching the
147 geometry of the PlanetScope mosaics for September 2020 through August 2021 (Planet Labs Inc.,
148 2020a) with seasonal S2 L2A reference images. For that, we performed cloud masking of the L2A
149 data based on the scene classification and the cloud displacement index (Frantz et al., 2018) with
150 a threshold of -0.8. We then created three seasonal near-infrared reference images, each covering
151 a total of four months (September-December, January-April, May-August), to reduce seasonality
152 effects in the co-registration procedure (Rufin et al., 2021a). We calculated displacement vectors
153 for the PlanetScope near-infrared bands based on the S2 reference using the displacement
154 function in Google Earth Engine. The algorithm performs a multi-scale rubber-sheeting correction
155 based on cross-correlation. After several tests, we defined a maximum offset as 100m and stiffness
156 as 5 and used the resulting displacement vectors to co-register all PlanetScope spectral bands. We

157 divided the study area into 0.3° grid tiles and processed each tile individually with a 0.05° buffer
 158 to avoid edge artifacts.

159 We assessed the effects of the co-registration procedure by calculating NDVI time series
 160 noise (Vermote et al., 2009) from triplets of measurements y_i , y_{i+1} , and y_{i+2} acquired at $month_i$,
 161 $month_{i+1}$, and $month_{i+2}$. We quantified the differences between the center NDVI and the linear
 162 interpolation between the two outer measurements as time series noise:

$$163 \quad TSNoise = \sqrt{\frac{\sum_{i=1}^{n-2} \left(y_{i+1} - \frac{y_{i+2} - y_i}{month_{i+2} - month_i} (month_{i+1} - month_i) - y_i \right)^2}{N-2}} \quad (1)$$

164 We created spatial representations of time series noise for selected sites and visually
 165 inspected the geometric consistency of time series metrics across 10% of the tiles during training
 166 data collection to assess the effect of the co-registration procedure.

167 Based on the co-registered mosaics, we generated time series metrics for three seasons
 168 ranging from September through December, January through April, and May through August
 169 (Bey et al., 2020). For every season, we calculated median (P50) values for the green, red, and
 170 near-infrared band, as well as NDVI, for which we added a 75th percentile metric (P75). In
 171 addition, we included texture and contrast based on the first season and third season median
 172 NDVI (Figure 2). Texture indices (TI) were derived by calculating median values in 25m, 100m,
 173 and 200m radial kernels, yielding three texture layers for two seasonal windows. The kernel sizes
 174 were determined by trial-and-error and are geared towards reflecting the distribution of field
 175 extents in the region. Contrast-enhancing indices (CI) were subsequently derived as normalized
 176 difference between pixel-level seasonal NDVI P50 and the corresponding TI:

177

$$CI_{P50NDVI} = \frac{(P50_{NDVI} - TI_{NDVI P50})}{(P50_{NDVI} + TI_{NDVI P50})} \quad (2)$$

178

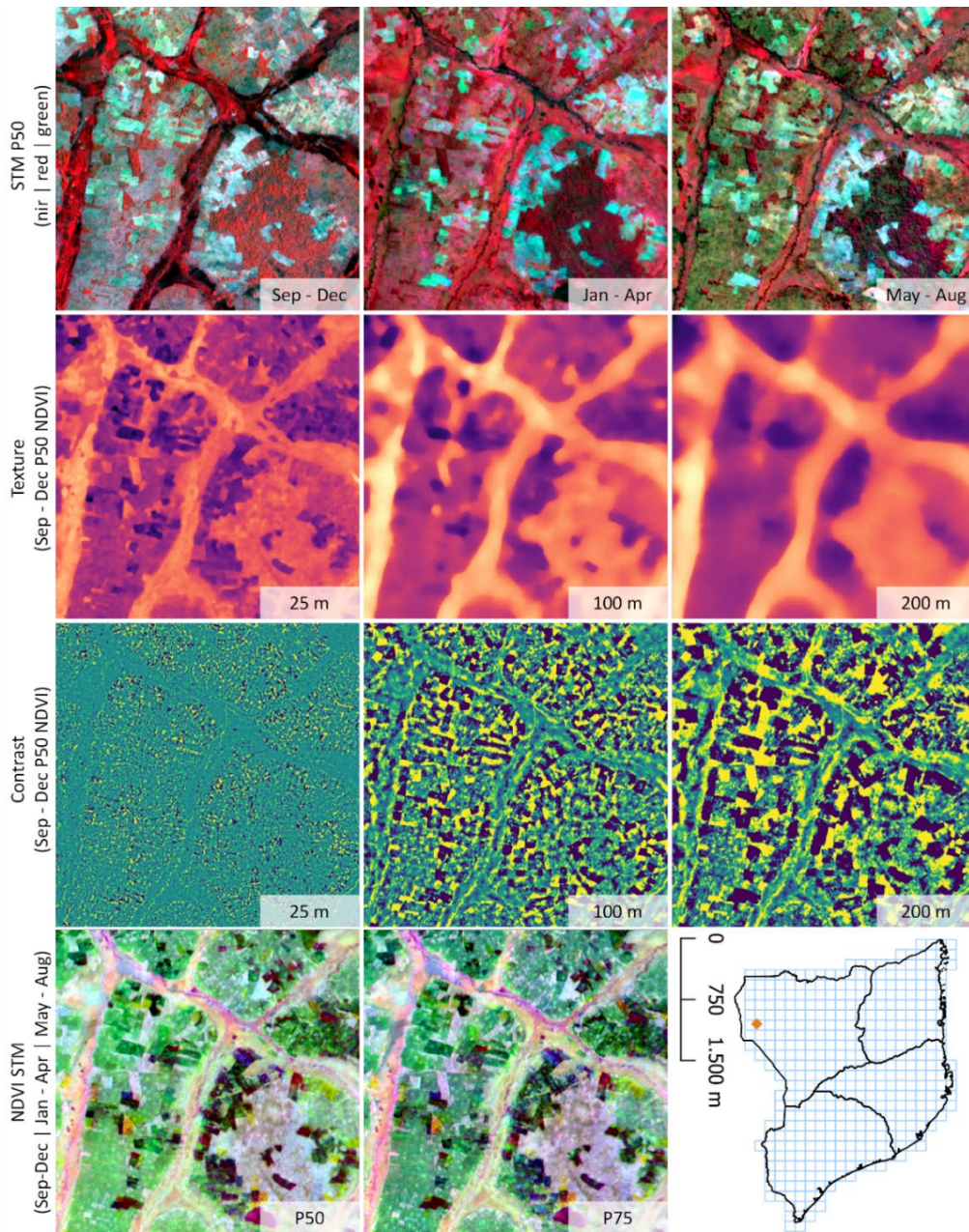
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These features normalize the NDVI values of a specific pixel with its surroundings, a technique that has been successful in large-area agriculture mapping studies (Deines et al., 2019). In addition, we included elevation and slope derived from the NASADEM (NASA JPL, 2020) as well as layers representing latitude and longitude, which have been shown to improve classification performance in large-area mapping studies (Pflugmacher et al., 2019).



183

185 Figure 2: Input features derived from the PlanetScope mosaics. Rows show false-color infrared
 186 visualization of the seasonal median metrics, the texture of the third season NDVI median across different
 187 kernel sizes, contrast-enhancing features based on different texture kernels, and RGB composite of seasonal
 188 NDVI metrics.

189 *Model training & classification*

190 We created an initial training dataset within a 10% random sample of the 422 0.3° tiles
191 (n=42). Within each tile, we randomly sampled 50 points with a minimum distance of 0.002° (~222
192 m) for initial model training. For each point, we screened the seasonal metrics and Google Earth
193 VHR imagery to record the class label according to our classification scheme (Table 1) and the
194 VHR acquisition date. If no appropriate label could be obtained for the target location, the sample
195 was moved into a nearby region to determine a class label.

196 The short-term fallow class was particularly difficult to train. We labeled a location as
197 short-term fallow only if Google Earth VHR data revealed a recent fallow stage and previous
198 cultivation in the last five years. Due to the high requirements for suitable pairs of VHR imagery,
199 the fallow class was underrepresented in the initial random training dataset. We, therefore,
200 trained an initial Random Forest model (Breiman, 2001) and predicted class probabilities across
201 the training tiles. The probability layers were used in conjunction with Google Earth VHR data
202 to amend the number of samples for the short-term fallow class from 66 to 209.

203 Table 1: Class catalog, definitions, and training sample size for both iterations.

Class	Definition	Final training samples	(Initial)	Validation samples
Active cropland	Actively used cropland with signs of recent land management	677	(582)	325
Short-term fallow cropland	Fallow croplands with active use in or after 2015	301	(209)	232
Herbaceous vegetation	Natural grasslands and wetlands	370	(239)	245
Open woodland	Open woody canopy with 10-75% cover fraction	650	(532)	383
Closed woodland	Closed woody canopy with >75% cover fraction, including forestry plantations.	601	(514)	336
Unvegetated	Open soil, built-up, rock	226	(168)	213
Water	Perennial water bodies	53	(51)	206

204

205 We performed an initial classification and then further enhanced the training dataset by
206 sampling eight additional tiles (2%), selected to match the distribution of cropland fractions in
207 our training tiles with that of the study region. We labeled 50 random points per tile. Next, we
208 calculated probability margins M_{prob} at the pixel level as the difference between the probability of
209 the most likely and the second most likely class. We calculated the 25th percentile of M_{prob} for each
210 class and sampled fifty locations per class in regions with M_{prob} scores below the 25th percentile,
211 amending the initial sample by 350 locations.

212 We deliberately designated one remote sensing expert with field experience as interpreter
213 for both training and validation, to avoid misinterpretations by inexperienced interpreters and
214 resulting inconsistencies in the reference data, which would require further interaction by
215 experts. The resulting reference dataset is expected to be of highest consistency and quality, which
216 is crucial for RF-based classification of smallholder cropland (Estes et al., 2021).

217 *Validation and unbiased area estimation*

218 We created a stratified random sample to validate the map results and generate an
219 unbiased estimation of class areas following recommended procedures (Olofsson et al., 2014). We
220 calculated sample size ($n=1,875$) targeting a standard error of the overall accuracy of 1% and
221 assuming User's accuracies of 0.75 for all classes (Cochran, 1977). We allocated 200 samples per
222 class and the remainder ($n = 475$) according to class proportions. Each sampling unit (pixel) was
223 labeled based on false-color infrared representations of the time series metrics at three zoom
224 levels, Google Earth VHR images (for cropland and fallows we constrained the use to those with
225 acquisition dates after 2016), and S2 NDVI time series profiles obtained using the Google Earth
226 Engine TimeSeries Explorer plugin for QGIS (Rufin et al., 2021b). We could determine a class

227 label for 89.8% of the samples. We derived area-adjusted accuracy scores, 95% confidence
228 intervals, as well as error-adjusted area estimates from the reference data, implemented in the
229 mapac package for R v0.11 (Pflugmacher, 2022).

230 *Assessing cropland distribution and comparing land cover products*

231 We assessed the spatial patterns of cropland by calculating the share of active cropland,
232 short-term fallow, and total cropland (including both) per 0.1° grid cell. We then investigated the
233 fraction of short-term fallows relative to total cropland along gradients of accessibility expressed
234 in travel times to the next city (Weiss et al., 2018) and the degree of agricultural consolidation
235 expressed as the total cropland proportion at the 0.1° grid cell level.

236 For comparison, we compiled several global and regional land cover and cropland
237 products (Table 2), from which we calculated the extent and spatial distribution of cropland at
238 the 0.1° grid cell level. The two regional land cover products included here are one map produced
239 by the Fundo Nacional de Desenvolvimento Sustentável (FNDS, 2019), and a second map
240 developed in the context of tree plantation monitoring (Bey and Meyfroidt, 2021). Importantly,
241 both regional products were principally produced for mapping forests and tree plantations and
242 thus have a generic cropland definition involving both active cropland and short fallows.

243 Table 2: Key characteristics of land cover products used for comparison in this study.

Name	Cropland definition	Scope	Res. (m)	Year	Reference
ESA WorldCover	Land covered with annual cropland that is sowed/planted and harvestable at least once within the 12 months after the sowing/planting date. The annual cropland produces an herbaceous cover and is sometimes combined with some tree or woody vegetation. Excludes perennial woody crops.	Global LC	10	2020	(Zanaga et al., 2021)
ESRI Land Cover	Human planted/plotted cereals, grasses, and crops not at tree height; examples: corn, wheat, soy, fallow plots of structured land.	Global LC	10	2020	(Karra et al., 2021 - 2021)

MOD12Q1 V006	Cropland where at least 60% of the area is cultivated cropland, and mosaics of small-scale cultivation 40-60% with natural tree, shrub, or herbaceous vegetation.	Global LC	500	2019	(Sulla-Menashe et al., 2019)
COPERNICUS LC 100m C3	Lands covered with temporary crops followed by harvest and a bare soil period (e.g., single and multiple cropping systems). Note that perennial woody crops will be classified as the appropriate forest or shrubland cover type.	Global LC	100	2019	(Buchhorn et al., 2020)
GLAD Cropland	Land used for annual and perennial herbaceous crops for human consumption, forage (including hay), and biofuel. Perennial woody crops, permanent pastures, and shifting cultivation are excluded from the	Global Cropland	30	2019	(Potapov et al., 2021)

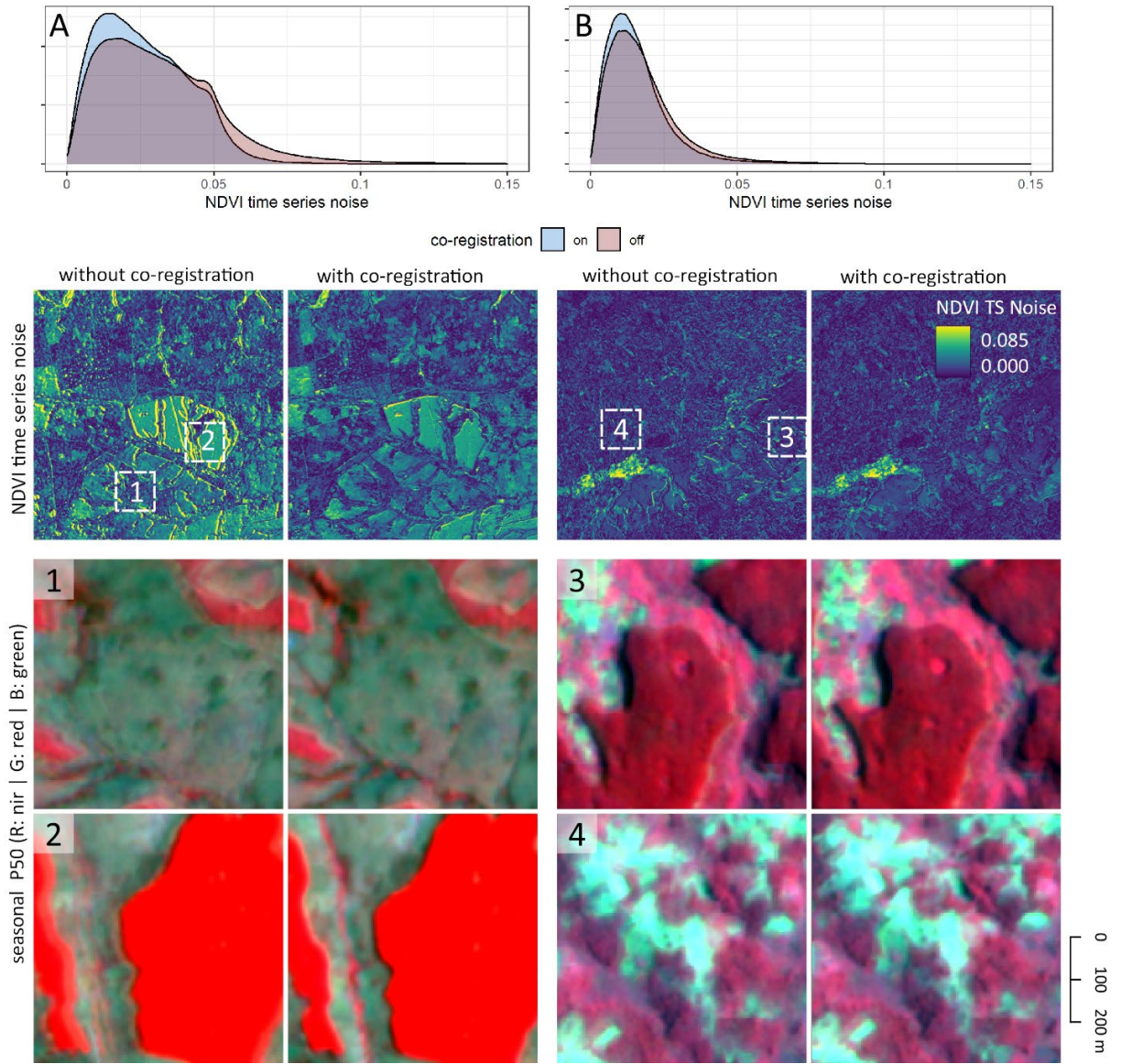
definition. The fallow length is limited to 4 years for the cropland class.

GFSAD30	All cultivated plants harvested for food, feed, and fiber, including plantations (e.g., orchards, vineyards, coffee, tea, rubber), and fallow areas, but excluding pastures.	Global	30	2015	(Xiong et al., 2017)
FNDS Land Cover	Non-woody crops with at least 20% of use and a minimum mapping unit of 1 ha.	National LC	10 – 30 m	2016	(FNDS, 2019)
Bey and Meyfroidt (2021)	Cultivated landscapes including active and fallow cropland at the 30m level.	National LC	30	2017	(Bey and Meyfroidt, 2021)

245 **Results**

246 *Pre-processing*

247 Our geometric co-registration approach based on S2 reference images successfully
248 mitigated multi-temporal co-registration issues in most cases (Figure 3). The effect of the co-
249 registration was most notable at sharp transitions of land cover classes, such as the edges of
250 forestry plantations (Example A). More subtle effects were observed in fragmented landscapes
251 (Example B). Importantly, high noise is not per se an indication of low geometric consistency but
252 can also represent highly dynamic land surface characteristics, such as floodplains with multiple
253 cropping cycles within a year. Remaining issues were observed in forested hillslopes and
254 topographically challenging terrain, often in the seasonal window overlapping with the wet
255 season. Aggregating the individual mosaics into seasonal median or 75th percentile metrics
256 further eradicated mosaicking artifacts and cloud contamination. The code for creating the
257 seasonal S2 reference images, co-registering the PlanetScope mosaics, and creating input metrics
258 will be made available in a separate repository after acceptance of the article.



259

260 Figure 3: Time series noise as density and maps, with and without co-registration for two regions. Image

261 chips represent subsets of seasonal median metrics within the presented regions.

262 *Accuracy assessment and area estimates*

263 The area-adjusted overall accuracy of the final map was 88.6% ($\pm 1.5\%$), or 91.6% ($\pm 1.2\%$)
264 when considering only the classes active cropland, fallow cropland, and non-cropland. Class-
265 wise user's accuracies ranged between 71.1% ($\pm 4.9\%$) for the active cropland class to 99.0% (\pm
266 1.3%) for water, producer's accuracies ranged between 61.4% ($\pm 5.8\%$) for herbaceous vegetation
267 to 99.7% ($\pm 0.4\%$) for closed woodland. The iterative model training led to a 12% improvement in
268 map accuracy as compared to the initial prediction (overall accuracy 76.0% ($\pm 2.1\%$)). The
269 confusion matrix (Table 3) revealed that commission errors of the active cropland class were
270 dominant, affecting herbaceous vegetation, open woodland, and short-term fallows. Omission
271 errors of the active cropland class were negligible. Commission errors of short-term fallow
272 occurred mostly on herbaceous vegetation, whereas omission errors mostly occurred in active
273 cropland. Error-adjusted area estimates based on the reference sample revealed the overall class
274 proportions and related confidence intervals. In the growing season 2020/2021, active cropland
275 accounted for 16.2% ($\pm 1.1\%$), and fallow for 6.6% ($\pm 0.8\%$) of the study area. Area estimates for
276 both cropland classes combined are 22.8% ($\pm 1.2\%$).

277 Table 3: Confusion matrix populated with probabilities, class-wise area-adjusted user's (UA) and
 278 producer's accuracy (PA) with 95% confidence intervals (CI), as well as sample-based area estimate with
 279 95% confidence intervals (CI).

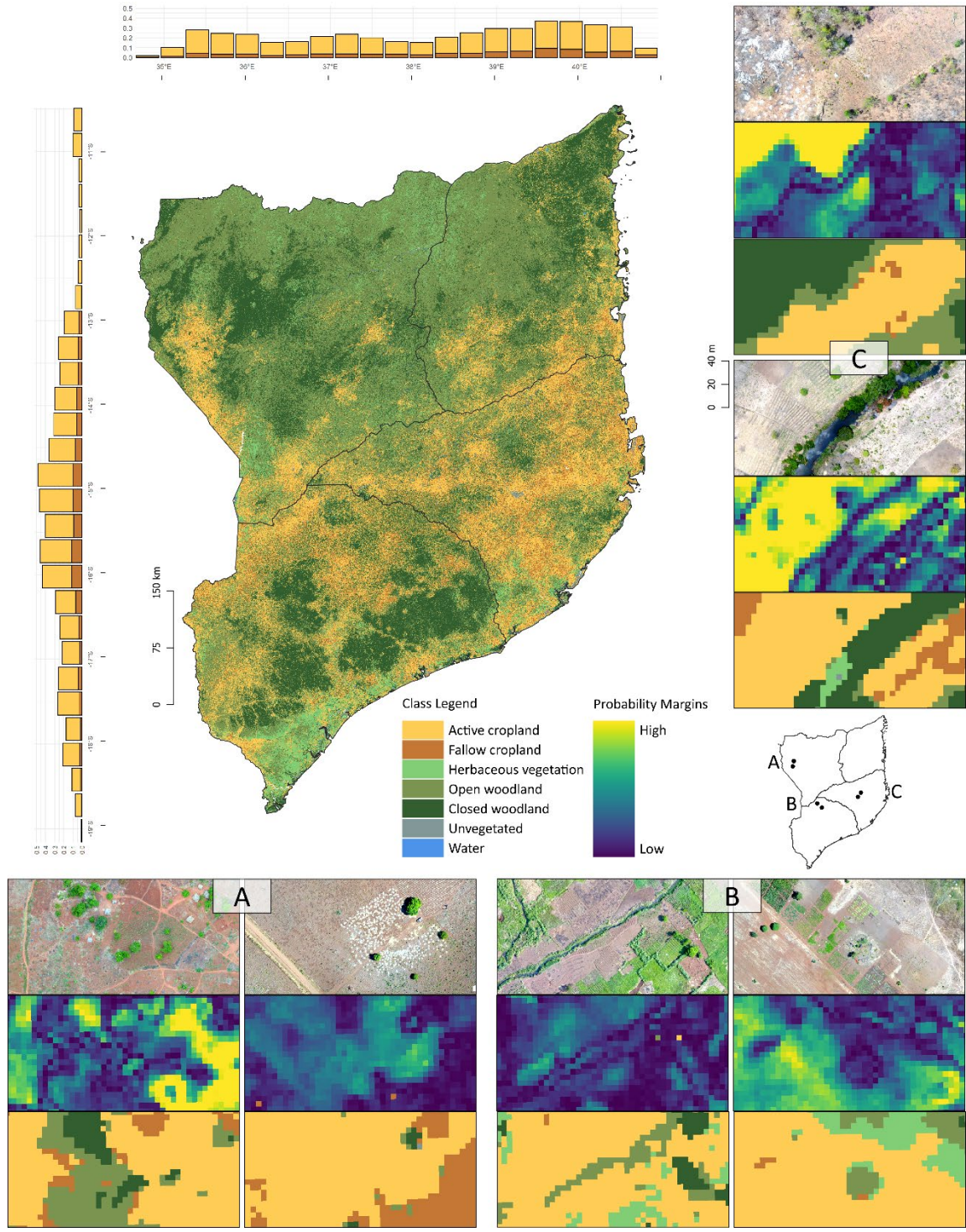
	Active Cropl.	Short Fallow	Herb. Veg.	Opn. Wdl.	Cld. Wdl.	Non-Veg.	Water	UA	95% CI.
Active Cropl.	0.1578	0.0116	0.0266	0.0205	0	0.0055	0	71.1%	4.9%
Short Fallow	0.002	0.0437	0.0035	0.0013	0	0	0	86.7%	4.4%
Herb. Veg.	0.0019	0.0058	0.0691	0.0006	0.0006	0.001	0	87.3%	4.2%
Opn Wdl.	0	0.0039	0.0097	0.3558	0	0.001	0	96.1%	1.9%
Cld Wdl.	0	0.0008	0.003	0.0129	0.2389	0	0	93.5%	2.6%
Non-Veg.	0.0002	0.0007	0.0004	0	0	0.0143	0.0003	89.7%	4.1%
Water	0	0	0.0001	0	0	0	0.0066	99.0%	1.3%
PA	97.5%	65.8%	61.4%	91.0%	99.7%	65.8%	95.6%		
95% CI	1.2%	7.3%	5.8%	2.2%	0.4%	13.2%	4.1%		
Sample-based area	16.2%	6.6%	11.2%	39.1%	24.0%	2.1%	0.7%		
95% CI	1.1%	0.8%	1.1%	1.2%	0.7%	0.4%	0.0%		

280

281 *Spatial patterns of active and fallow cropland*

282 The province Nampula had the highest share of active cropland (40.3%), followed by
283 Zambezia (26.8%), Cabo Delgado (15.4%), and Niassa (10.3%), with the Niassa reserve – a large
284 protected area in the northeastern part of Niassa province and northwestern part of Cabo
285 Delgado province – being essentially void of cropland (Figure 4). The largest fractions of active
286 cropland were mapped in the vicinity of larger agglomerations such as Mocuba and Pebane in
287 Zambezia, Pemba in Cabo Delgado, Lichinga in Niassa province, and Nampula city in Nampula
288 province. Besides these, the spatial patterns of active cropland follow the major transportation
289 corridors along the coast and the east-west Nacala Corridor. The map will be made available
290 online for further use in compliance with the custom license of the NICFI data program.

291 According to the maps, short-term fallows accounted for 13.2% of the total cropland area,
292 though this remains an underestimate as our error-adjusted estimate indicates 28.9%. The most
293 densely populated province Nampula had the highest shares of fallows (12.8%), followed by
294 Zambezia (6.0%), Cabo Delgado (2.2%), and Niassa (1.4%). Fallow fractions (relative to total
295 cropland) rarely exceeded 40% at the 0.1° grid level. Fallow fractions increased non-linearly with
296 increasing cropland fractions and accessibility, indicating higher shares of short-term fallows in
297 consolidated production regions (Figure 5).

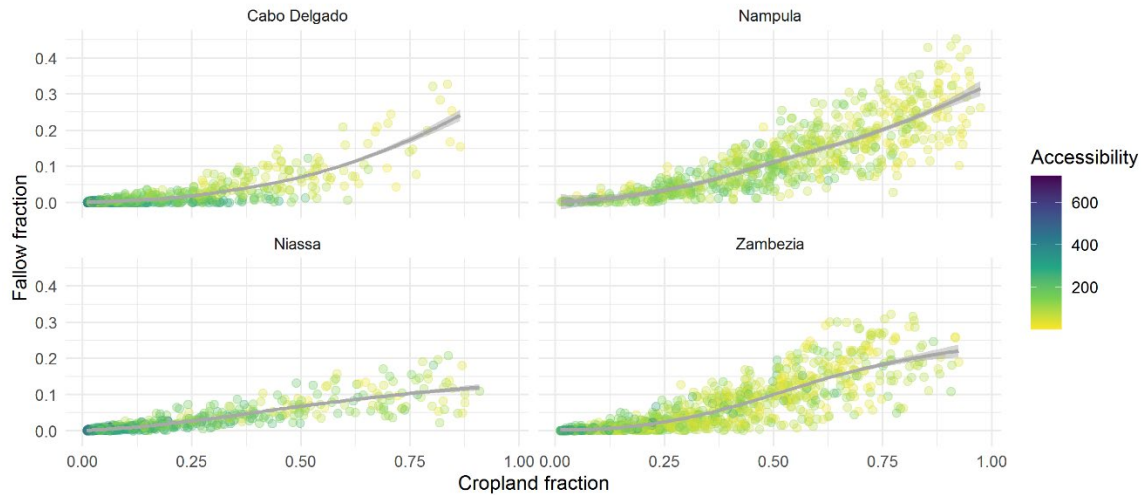


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299 Figure 4: Map overview with histograms depicting cropland and fallow distribution by latitude and

300 longitude. Zoom-ins show in-situ drone data obtained during fieldwork in mid-November 2021,

301 probability margins, and classification outputs.

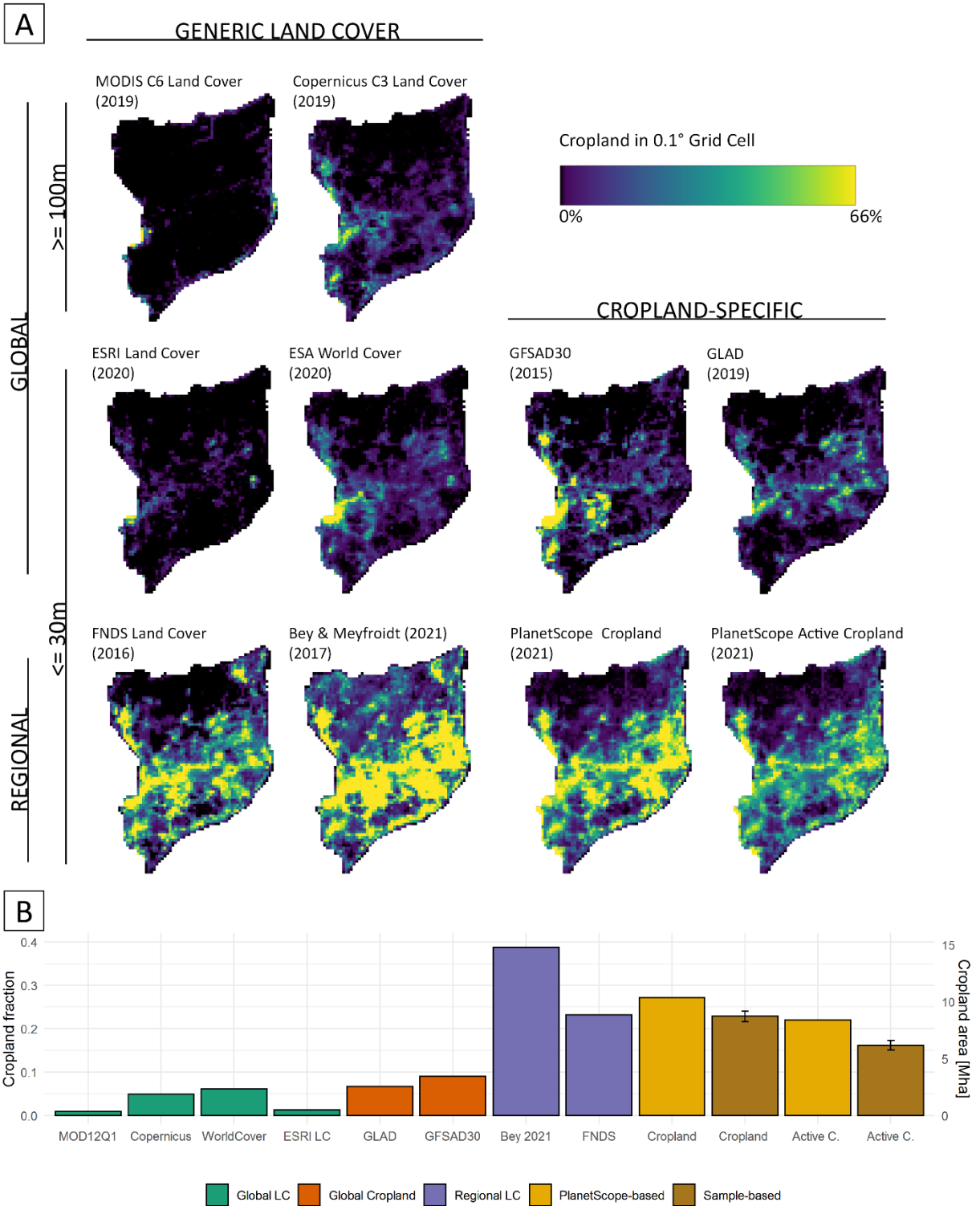


302

303 Figure 5: Relationship between cropland fraction, fallow fraction, and accessibility (in minutes travel time)
 304 per 0.1° grid cell.

305 *Comparing cropland extent across products*

306 A comparison at the grid-cell level revealed the similarities and differences in cropland
 307 distribution derived from the eight LC products considered here (Figure 6 A). Two of three global
 308 land cover products with a generic class catalog (ESRI LC, MODIS LC) show cropland fractions
 309 close to 1% in the study region (Figure 6 B), and the WorldCover product indicated 6%. The
 310 cropland-specific products (GFSAD30, GLAD) showed shares of 7%, and 9%, respectively. The
 311 regional products show substantially higher shares of cropland, with 23% and 39%.



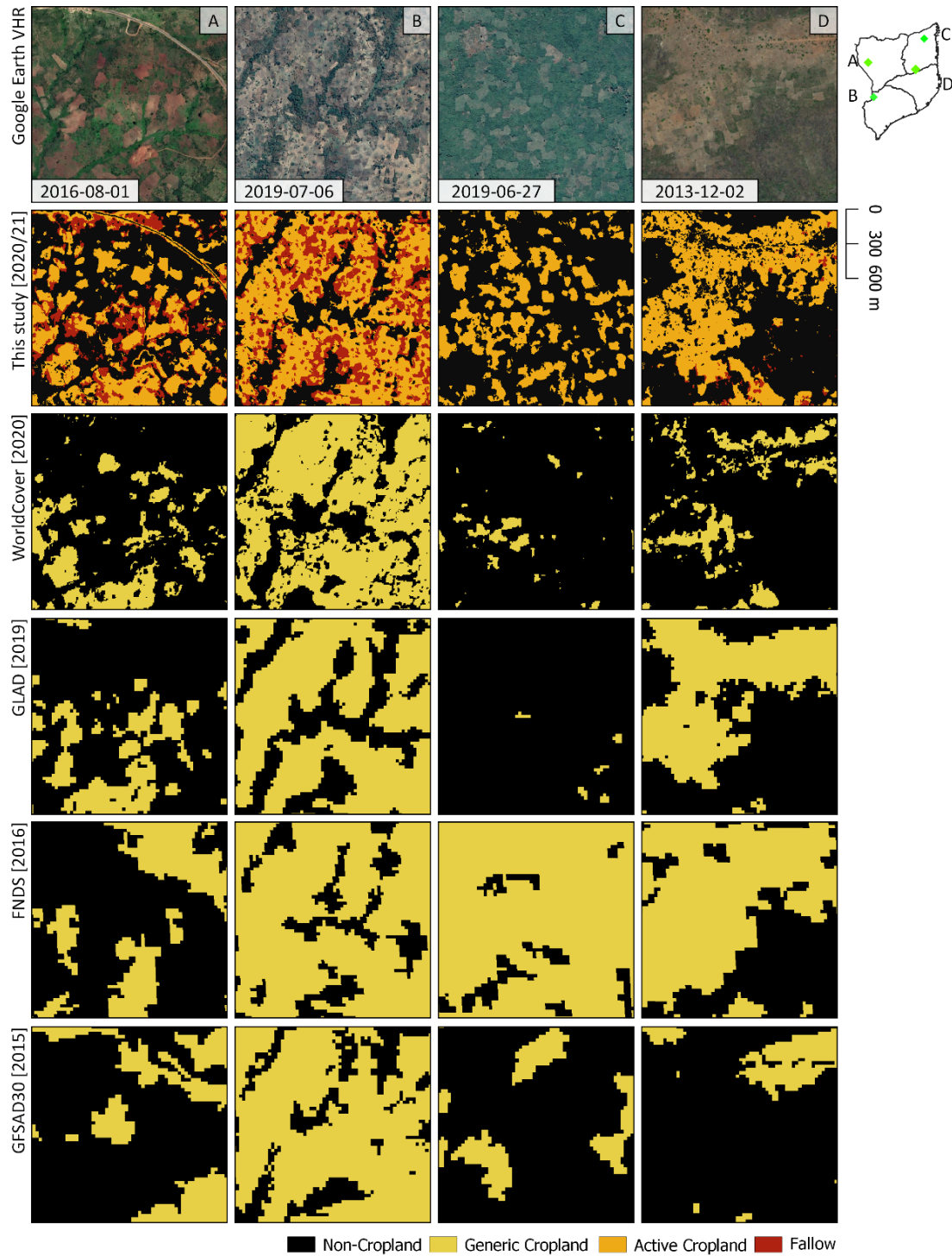
312
313

Figure 6: Cropland fractions for different land cover products (see Table 2 for details) per 0.1° grid cell (A)

314

and aggregated for the entire study region (B).

315 Snapshots of exemplary regions reveal the strengths and caveats of our PlanetScope-based
316 maps in comparison with existing products mapping generic cropland (Figure 7). First (A), the
317 spatial detail of the PlanetScope data in conjunction with our pixel-based classification allows for
318 the identification of very small fields, which may go unnoticed in other products. However, some
319 commission errors are visible at land cover transitions, e.g. next to roads. Second (B), while
320 existing products show good agreement in consolidated regions, only the PlanetScope-based map
321 can be used to disentangle active from the fallow components of cropland. Third (C), our map
322 accurately depicts active cropland in particularly dynamic landscapes, whereas large
323 disagreement was found in existing products. Here, minimum mapping units (FNDS), object-
324 based approaches (GFSAD30), multi-year aggregation of imagery, and cropland definitions
325 involving fallows or excluding shifting cultivation (GLAD) may lead to highly differing
326 representations of cropland extent. The omission of short-term fallows in the PlanetScope-based
327 map here likely results from the high rates of vegetation growth, causing a high resemblance with
328 open woodlands after short periods. Lastly (D), the PlanetScope-based map captures the
329 landscape complexity, as it for instance includes small cropland parcels next to housing, and
330 excludes trees on crop fields.



331

332 Figure 7: Image subsets of Google Earth VHR data, the PlanetScope based map, and selected land cover

333 products, with target years in square brackets.

334 **Discussion**

335 This study presents an operational framework for mapping complex smallholder
336 landscapes at 4.77 m resolution using PlanetScope data provided through the NICFI data
337 program and Google Earth Engine. We combine state-of-the-art pre-processing techniques to
338 create maps of active cropland and short-term fallows, besides five other land cover classes, for a
339 large area in Northern Mozambique. Our maps are accurate and allow for assessments of the
340 distribution of active cropland and short-term fallow cropland at unparalleled timeliness, spatial
341 and thematic detail across more than 380,000 km².

342 Our pre-processing framework for the PlanetScope mosaics produced consistent, spatially
343 detailed, and temporally precise analysis-ready image features in a cloud-prone and
344 topographically challenging region. The S2-based co-registration procedure in Google Earth
345 Engine could not resolve all geometry-related artifacts, but nevertheless led to substantial
346 improvements in the quality of time series metrics used for classification. The creation of seasonal
347 median reflectance and index values further mitigated the presence of cloud and cloud shadow
348 remnants. The iterative training procedure to amend our training dataset improved the
349 classification substantially (+12.6% compared to the initial prediction). The inclusion of Random
350 Forest probability layers was particularly valuable to identify training locations of short-term
351 fallows, as its narrow definition imposed high requirements on VHR data availability. This
352 approach may be particularly useful for studies focusing on rare and hard-to-train classes, such
353 as change classes, in the absence of reference data.

354 The resulting maps substantially improve on existing approaches in terms of spatial
355 resolution and thematic detail, with accuracies in line with or exceeding those of previous

356 approaches on active cropland mapping in complex smallholder regions (Estes et al., 2021;
357 Ibrahim et al., 2021). Similar to previous works, commission errors of cropland were a key error
358 type (Estes et al., 2021; FNDS, 2019; Xiong et al., 2017), which should be a focus of further research.
359 We observed this error in land cover transition zones, such as between moist herbaceous
360 vegetation and dry, bright sandy soils. Moreover, we observed the omission of fallows in regions
361 with higher soil moisture due to confusion with herbaceous vegetation. Integrating information
362 on moisture availability, e.g. through optical measurements in the shortwave infrared domain,
363 could potentially mitigate these errors.

364 Comparing global and regional land cover products revealed a substantial variation in
365 overall cropland extent. While mismatches between the target year of our study and those of
366 existing products may contribute to the observed differences, the interplay of spatial resolution
367 or minimum mapping units, production methods, spatial coverage, and, as a result, differing
368 class definitions certainly play a key role in the observed differences. Generic cropland definitions
369 (including fallow) are commonly used due to mixtures of non-crop vegetation, active and fallow
370 cropland at the sub-pixel level. However, they limit the usefulness of products for estimating the
371 area of active cropland and fallows, which are key for assessments of agricultural productivity,
372 food security, livelihoods, and land change.

373 Knowledge about the fractions and spatial distribution of fallows in SSA is scarce. We
374 found that short-term fallows occupy a substantial part of the cropland area, with a high spatial
375 heterogeneity. Our short-term fallow fractions (map-based 13%, sample-based 29%) are in line
376 with fallow fractions of 25% reported at the farm-level in other parts of Mozambique (Leonardo
377 et al., 2018), whereas farm-level decisions may lead to varying fractions and lengths of fallows

378 (Temudo and Silva, 2012). Large differences compared to a remote-sensing-based estimate
379 documenting fallow fractions of 63% in the Sahel belt (Tong et al., 2020) can be explained by
380 regional differences as well as through our narrow fallow definition, which excludes long-term
381 fallows of more than 5 years of length. Additionally, our map entails omission errors of short-
382 term fallows in regions with smaller agricultural footprints and high vegetation growth rates,
383 where fallows were identified as woody cover. Integrating PlanetScope imagery from past
384 growing seasons may help to delineate a broader range of fallow types and stages, including
385 those with high fractions of woody cover. However, the reduced temporal density of the
386 PlanetScope mosaics in the pre-2020 era and the limited availability of S2 L2A products in Google
387 Earth Engine was a major constraint for including pre-2020 PlanetScope mosaics to broaden the
388 definition of fallow land in this study. The continuation of the NICFI data program will allow for
389 the enhancement of the methods presented here in terms of disentangling short-term and long-
390 term fallows and providing detailed insights into year-to-year change processes.

391 **Conclusion**

392 Spatially detailed and timely maps on active cropland and short-term fallow in
393 smallholder landscapes are pivotal for assessments of food security, livelihoods of local
394 communities, land use change, and carbon budgeting, but are commonly not available. This work
395 provides a framework for mapping active cropland and short-term fallows in highly fragmented
396 smallholder landscapes. Our approach relies on PlanetScope mosaics made available through the
397 NICFI data program. We derived seasonal analysis-ready datasets from co-registered time series
398 of PlanetScope mosaics and used iterative learning to map active and fallow cropland using
399 Google Earth Engine. The resulting maps cover 380,000 km² at <5 m spatial resolution, separate

400 active cropland from short-term fallows in the growing season 2020/2021, and are highly accurate
401 (88.6% \pm 1.5%). The PlanetScope-based cropland map presented here enables more precise
402 estimates of actively used cropland through its high granularity, temporal precision, and thematic
403 depth untangling active cropland and short-term fallows. Our approach is operational and thus
404 suitable to tackle persistent constraints related to spatial complexity and dynamics for mapping
405 complex smallholder landscapes in the tropics.

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