1	Effect of combined dry-wet irrigation and microbial dynamics on soil nutrient
2	bioavailability
3	Arnab Majumdar ¹ * (<u>arnabmajumdar891@gmail.com</u>), Munish Kumar Upadhyay ²
4	(<u>munish007up@gmail.com</u>), Pradeep Kumar Dubey ³ (<u>pradeep.dubey4@bhu.ac.in</u>), Biswajit
5	Giri ¹ (<u>giri.biswajit605@gmail.com</u>), Ashish Kumar Srivastava ⁴ (<u>ashishbarc@gmail.com</u>),
6	Sudhakar Srivastava ⁵ (<u>Sudhakar.srivastava@gmail.com</u>)
7	¹ Department of Earth Sciences, Indian Institute of Science Education and Research (IISER)
8	Kolkata, Mohanpur, West Bengal 741246, India
9	² Environmental Geochemistry Laboratory, Department of Civil Engineering, Indian Institute of
10	Technology Kanpur, Uttar Pradesh 208016, India
11	³ Institute of Environment & Sustainable Development, Banaras Hindu University, Varanasi
12	221005, India
13	⁴ Nuclear Agriculture and Biotechnology Division, Bhabha Atomic Research Centre, Mumbai
14	400085, India
15	⁵ Plant Stress Biology Laboratory, Institute of Environment & Sustainable Development, Banaras
16	Hindu University, Varanasi 221005, India
17	
18	*Correspondence- Dr. Arnab Majumdar (<u>arnabmajumdar891@gmail.com</u>)
19	
20	

21 The manuscript has submitted to 'Agronomy Journal', a member journal of the American Society 22 of Agronomy, and been revised based on one round of peer review, but has yet to be formally 23 accepted for publication. The authors are currently preparing additional revisions, and thus 24 subsequent versions of the manuscript will have different content.

25

26 Abstract

27 Alternate wetting-drying (AWD) cultivation with implication on soil microbiome, nutritional 28 dynamics and rice yield during pre-monsoon (boro) and monsoon (aman) season are not well 29 studied. In the present 4-year field study the impact of AWD (in pre-monsoon season) is compared 30 with conventional mode of irrigation (i.e. flooded field in monsoon season). The release of soil 31 nutrients into the soil-aqueous system, influencing microbial populations and modulating the redox 32 status were explored. Results indicated an increase in total content as well as bioavailability of 33 selected nutritional elements (N, P, K, Fe, Ca, Mg, Cu and Zn) by 16-54% in the pre-monsoon 34 cultivation relative to monsoon cultivation. Three plant growth phases (developing, milking and 35 harvest) were considered to check the nutrient modulations in soil and plant tissues along the 36 continuum plant growth and elemental uptake. Krona charts, relative abundance, rarefaction curve and multivariate analysis of metagenomics data showed that the pre-monsoon soil was more 37 38 enriched and maintained a balance between soil pH and microbial biomass than the monsoon soil. 39 Microbial community diversity associated with plant growth phases also found to be different 40 depending on the seasonal alterations. Bacillus sp., Acidothiobacillus sp., Pseudomonas sp., 41 *Rhizobium* sp., *Burkholderia* sp. were predominant in pre-monsoon soil releasing pulses of N, P, 42 K, Ca and Mg whereas Verrucomicrobia was found to be dominant in monsoon soil where Fe was 43 released. This study is a first of its kind that showed the combined effect of season and soil 44 microbes on macro-micro nutritional availability in soil and enhanced plant quality.

45 Keywords: Soil quality; Metagenomics; Nutrient availability; Pre-monsoon irrigation; Resource
46 conservation; Soil microbiome

53 Introduction

54 Upto 43% of global irrigation water (accounting nearly 27% freshwater available on the planet) is 55 directly consumed for rice production. The estimated average consumption of 2500 litres of water 56 for a kilogram of rice produced is huge water footprint for global rice production (Surendran et al., 57 2021). However, since rice is a staple food crop for nearly half of world population and source of 58 livelihood for hundreds of millions of farmers across globe, therefore, devising sustainable ways 59 of rice irrigation under changing climate scenario is highly imperative (Majumdar and Bose, 2017; 60 Dubey et al., 2020). These variable strategies adopted for rice cultivation depends on site specific 61 edaphic factors. soil physico-chemical properties and prevailing environmental 62 conditions(Majumdar et al., 2020a; b). The informal pump irrigation is largely a contingent water 63 source for a boro season rice cultivation in regions like West Bengal in India and in other South 64 Asian countries like Bangladesh etc. Overreliance on groundwater for irrigating paddy field in boro season (i.e. during pre-monsoon period) primarily causes groundwater depletion (Dangar et 65 66 al., 2021; Tulip et al., 2022), aquifer havoc further adding to the existing regional Arsenic toxicity challenges (Nickson et al., 2000; Johannesson and Neumann, 2013; Kumarathilaka et al., 2018), 67 68 as well as unsustainable usage of non-renewable forms of energy (as diesel/fuel) and power (electricity) (Mukherji, 2007). Hence, unravelling water resource conserving agronomic practices 69 for rice cultivation majorly during boro season holds paramount importance owing to its 70 environmental merits in response to changing climate, as well as safeguarding the energy-irrigation 71 nexus (Dubey et al., 2020; Surendran et al., 2021). In this backdrop, alternate wetting and drying 72 mode of irrigation, raised beds, system of rice intensification (SRI) techniques, or saturated soil 73 culture (SSC), ground cover system, or aerobic rice cultivation etc. has already gained scientific 74 momentum and is cited in several recent field-based studies done in West Bengal, Bangladesh and 75 76 adjacent regions (Baldwin and Mitchell, 2000; Venterink et al., 2002; Blackwell et al., 2010; Dodd et al., 2015; Surendran et al., 2021). Moreover, the agrarian population in the region are also partly 77 78 convinced in adopting irrigation method such as alternate wetting and drying of paddy soil, especially during boro season as an adaptive rice cultivation practice (Pearson et al., 2018; 79 80 Majumdar et al., 2021; Upadhyay et al., 2022). So far, the site-specific motives attached to adoption of AWD irrigation mode are the water resource crunch faced by large segment of 81 82 subsistence rice farmers during the peak season requirements and the erratic weather pattern 83 witnessed over preceding years due to changing climate (Pearson et al., 2018; Dubey et al., 2020).

84 However, in present study we have explored the implications of alternate wetting and drying of 85 paddy soil during boro season on variations in soil microbiome and their nutritional dynamics.

86 The drying of paddy field soil followed by subsequent re-watering enhances rice productivity 87 through with flushes of nutrient release to the soil influencing soil microbial populations (Baldwin and Mitchell, 2000; Dodd et al., 2015). The rotation of drying and re-watering process in pre-88 89 monsoonal dry season can influence the soil organic matter (SOM) mineralization and microbial 90 turnover; known as the 'Birch Effect' (Birch, 1958). This effect has been observed for available 91 C, N and P released in soils (Mikha et al., 2005; Blackwell et al., 2010; Mooshammer et al., 2017). 92 The re-watering practice influences the pre-monsoonal soil nutritional flux while affecting the 93 microbial load (Wu and Brookes, 2005; Xiang et al., 2008). However, the direct effect of such 94 seasonal variation and physico-chemical changes on microbial populations for other essential 95 nutrients like Fe, Ca, Mg, Cu and Zn had not been studied. This cycle of drying and rewatering soil during pre-monsoonal wintry-summer season induces the release of intra-cellular osmolytes 96 like amino acids, ammonium compounds, glycerol and other organic solutes to the extracellular 97 environment (Wang et al., 2017; Majumdar, 2021). In monsoon soil, a slower microbial 98 decomposition of organic matter was reportedly due to the lower requirement of N in anaerobic 99 100 metabolism (Nishio et al., 1994). During the re-watering process of wintry dry soil, water gets 101 entrapped within the soil aggregates leading to the better availability of soil organic matter (SOM) 102 for microbial decomposition which allows microbes to break down recalcitrant complexes and 103 release nutrient flushes to the soil (Miller et al., 2005; Wu and Brookes, 2005).

There is a dearth of study showcasing the effects of employing water resource conserving practices during one rice season (i.e. pre-monsoon boro rice season under the boro-aman rice cropping pattern) on the nutrient availability and soil microbial community together. Therefore, an exclusive long term four-year field study (starting from February-May 2016 and extended till July-October 2019) were conducted in West Bengal, India unrevealing the positive implications of alternate wetting drying irrigation method (employed during boro season) on soil microbial dynamics (applying metagenomic applications) and nutrients bioavailability improving the crop and system yield. Overall, the merits of aforementioned water resource conserving agronomic practice in comparison with conventional monsoonal waterlog system is comparatively evaluated. Moreover, saving critical natural resources in agriculture such as water and soil have direct and indirect 114 interlinkages with several United Nations Sustainable Development Goals (UN-SDGs) and their 115 target set for year 2030 (Dubey et al., 2021a; b, 2022; Huang et al., 2021).

116 Materials and methods

117 Study area and experimental field set up, sample processing and analysis

118 The study area, Katwa block (latitude 23°43'25.4"N, longitude 88°01'30.4"E) of Bardhhaman 119 district, West Bengal had been selected based on our earlier assessment and soil As tracing 120 (Majumdar, A., 2016-2019, unpublished data; (Laha et al., 2013). For a long-term study of the 121 effect of pre-monsoon cultivation, wintery-summer (boro) and rainy (aman) season was selected 122 for four consecutive years from 2016 to 2019. Three fields, 1800 square feet each, were selected 123 for cultivation and soil-plant sampling based on randomized block design to avoid any partiality. 124 The fields was selected based on the pattern of cultivation practiced locally for over two decades 125 using pond water only. Water depth was maintained at 30cm from the surface soil in the monsoon 126 field. In the pre-monsoon cycle, pond water was applied at the beginning and retained for 2 days 127 with subsequent release of the water from the field, drying for the next 10 days, making the soil 128 moisture reduced to 60% of saturated values, and the field was monsoon again. This drying-129 wetting cycle was continued till the rice grain milking stage followed by a semi-dried phase (50% 130 saturation) till harvesting. Soil and rice plant samples were collected at three phases- pre-mature 131 or developing, milking or grain filling stage and mature or harvesting stage.

The collection, processing of soils, sample preparation and analysis were conducted using wavelength dispersive X-ray fluorescence spectroscopy (WD-XRF, Tiger S8 Bruker) by following the method described in (Majumdar et al., 2018, 2021). Soil samples were analyzed for the physico-chemical properties like pH, redox potential (ORP) using Oakton Waterproof PCS Testr35 and ORP Testr 10 and total organic matter by following the Walkley and Black method(Walkley and Black, 1934). On-spot field observation of soil pH and ORP was also done to check for any changes in the collected samples. Four standard reference materials (SRMs) (SBC-1, GSP-2, MESS3 and NIST2711a) were used for optimized calibration with an 82-91% to check recovery rate. Instrumentation specifications have been mentioned in **Supplementary Table 3**. Total N content was measured by following the Kjeldahl method. Soil samples were also tested to the bioavailable fractions of the selected elements by using an inductively coupled plasma mass 143 spectrometer (ICP-MS, Parkin-Elmer)(Ray Sarkar et al., 2017; Shrivastava et al., 2020).
144 Instrument specifications are given in Supplementary Table 2. The detailed process and
145 chemicals used are mentioned in Supplementary Table 1.

146 Rice seeds (Pusa Basmati 1121) were kept moist for one week to germinate and then grown in the 147 nursery until the seedlings (around 20 days after germination) were transferred to the respective 148 fields. Plant samples were collected 18-20 days after the transfer to the field and were marked as 149 'developing'. After 40-45 days of first sampling, plants were collected during the milking phase 150 and after the seeds were ripened, final sampling was done at harvesting time. A destructive method 151 of di-acid digestion was followed after complete drying of plants at 60°C for 4-5 days, for the total 152 elemental content measurement in plant samples using ICP-MS (Majumdar et al., 2019, 2022).

153 DNA extraction from soil and metagenomics analysis

154 A small amount of soil was used for the DNA isolation using QiagenDNeasy Powersoil Kit 155 (Cat#12888). The study considered soils from inflorescence and harvesting phase, due to differed 156 water saturation and nutrient contents available. Region-specific targeting proprietary primers at 157 Genotypic Technology Pvt. Ltd., Bangalore, India, were used for all the metagenomics analysis 158 and V3-V4 metagenome library preparation. All the detailed processes have been mentioned in 159 the **Supplementary file**. V3-V4 primer sequences from the Illumina paired-end raw reads were 160 selected from the high-quality bases. RDP6 classifier was used to assign taxonomies with clustered 161 at \geq 97% sequence similarities while comparing to the Greengene database which created a biome 162 file. This biome file was further used for advanced analysis and visualization. Alpha and beta 163 diversity for microbial species richness was calculated for each sample. Krona charts were made 164 that are interactive html files consisting of the phylogenetic information at each taxonomic level. 165 Venn diagram, multivariate analysis, rarefaction plot, were made to identify the difference in 166 microbial diversity in different soil samples. Cytoscape (var. 3.8.2) was used for the microbial 167 network distribution based on the elemental content in soil, assigned as 'edges' and read counts of 168 classes were assigned to be 'target nodes' for generating this network.

169 Statistical analysis of soil-microbe data

170 All the experimental data of soil analysis for total and bioavailable elemental concentrations were 171 justified by analysis of variance (ANOVA) at P<0.05 significance level. Duncan multiple range 172 test (DMRT) was also used for the level of significance and inter-class variance identification.
173 Principal component analysis (PCA) and principal coordinate analysis (PCoA) were used to justify
174 the distribution pattern assessment in soil samples and inter-elemental relation along with the
175 microbial species distribution within the soil samples from altered seasons. Graphpad Prism 6,
176 SigmaPlot (ver. 12) and PAST (ver. 3.2.1) were used for the graphical and statistical analysis.

177 Results

178 Soil assessment for nutritional profiling and plant accumulation under altered irrigation

179 Soil samples from two altered seasons (monsoon and pre-monsoon) at three sampling time points 180 (developing, milking and harvest) were assessed for the total elemental concentration and obtained 181 results are presented in Fig. 1 as radar plots. This radar plot is a constrictive layered formation 182 with major and minor grids indicating the major concentration bands and sub-divisional 183 concentration bands within sampling phases which eases multi-parameter data comparison. The 184 blue line in the figure corresponds to the connecting pattern in changes of concentration in different 185 fields. In all the years, 2016-2019, soil samples from the harvest phase always showed the highest amounts of nutrients irrespective of cultivation processes. However, differences were evident 186 when monsoon soil was compared to the pre-monsoon soil where the deposition of elemental 187 content was highest (Fig. 1). Sub-divisions of the radar plot in each year showed a fluctuation in 188 189 the concentration pattern between the developing and harvest phase. From 2016 to 2019, the 190 increment in concentrations of total N, Ca, K, Mg, Fe, P, Cu and Zn were 38.6%, 38.0%, 4.4%, 191 37.3%, 12.4%, 24.3%, 54.7% and 36.9% respectively while comparing element contents in control 192 soil to pre-monsoon soil. Whereas the percent increments were 45.98%, 56.4%, 19.7, 38.9%, 193 16.1%, 30.4%, 64.3% and 54.4 % of N, Ca, K, Mg, Fe, P, Cu and Zn, respectively, in pre-monsoon 194 soil in comparison to the monsoon soil (Fig. 1a to 1d). The total content increased by the end of 195 2019 sampling. Fig. 2 shows the total nutrient enrichment in whole rice plants by summing the 196 root and shoot content from three different fields. Rice can alter the uptake of nutrients from the 197 soil dependent on the levels of supply and the demands of growth (Hasanuzzaman et al., 2018; 198 Afsal et al., 2020; Gupta et al., 2022). Fig. 2a-h shows a similar trend in elemental accumulation 199 in plants under altered seasonal cultivations. Groundwater application enhances the soil nutrient 200 concentrations but flooding flushed out nutrients from the field resulting in a low elemental profile 201 in rice grown under these conditions. An increase in elemental concentrations was observed in pre202 monsoon plants from 2016 to the end of 2019 at the harvesting phase. Except for Na (-28.95% in 203 monsoon to pre-monsoon), all other nutrients showed a positive increment. Comparing monsoon 204 soil to pre-monsoon soil the percent increases were 55.6%, 58.6%, 32.2%, 51.2%, 52.8%, 31.3%, 205 and 38.6% for Ca, K, Mg, Fe, P, Cu and Zn respectively at the harvesting phase. Plant's elemental 206 concentration at the pre-mature and milking phases had a similar accumulation trend as at the 207 harvesting stage.

208 Soil physico-chemical influence on nutrient bioavailability and yield enhancement

209 In paddy fields, the soil pH and redox potential (ORP) show an opposite correlation(Kabata-210 Pendias, 2000) that influences the soil elemental concentration by modulating the inter-211 relationships of soil elements. Fig. 3 shows the correlation matrix plot for soil physico-chemical 212 parameters and nutrients. Fig. 3a, b and c correspond to the analysed soils from developing, 213 milking and harvest, justified based on the three algorithmic representations- shape (degree of an 214 oval to ellipsoid), color (dark red to violet) and directions (upward or downward) to show the 215 positive to negative correlation with statistically valid range marked as grey colored boxes. Higher 216 content of organic matter was observed in the developing phase, this also showed a strong positive 217 correlation with soil pH changes but no such relationship was observed with ORP. In the rice field, 218 the availability of P depends on both natural soil amounts and applied fertilizer levels and that 219 makes a strong association with soil pH and OM in the developing phase. Although both pH and 220 ORP revealed a negative relation with other nutrients, SOM influenced positively some of the 221 elements like P, Cu and Zn. Iron complexes are stable under high OM(Kabata-Pendias, 2000) and 222 total Fe showed a negative association along with Mg and K (see Fig. 3). In milking and harvest 223 phase soils, SOM was found to be lower than in the developing phase, but pH and ORP was highly 224 variable and a showed negative relationship. Water logging in the monsoon soils has lowered pH 225 and ORP values when compare with the pre-monsoon field. Redox potential changed positively 226 the available content of Mg, Fe, P and Cu whereas the rest of the elements were found negatively 227 correlated. Among inter-elemental dependences, Ca, Mg, Cu and Zn showed positive relationships 228 with all other elements. Soil nutritional status confers significant plant growth and final grain yield 229 after the harvesting phase and as shown in the Fig. 3d, total yield (calculated per hectare) was 230 found to be increasing yearly, although at marginal difference, but greater than the monsoonal 231 flooded fields. The yield percent increment was 6.46% in pre-monsoon boro cultivation and 5.2%232 in monsoonal aman cultivation.

233 Bioavailable elemental distribution in fields is influenced by pH and microbial biomass

234 Soil pH and microbial activity are very closely related as both depend on the soil moisture content, 235 SOM and under altered drying-wetting irrigation in pre-monsoon cultivation, higher content of 236 microbial biomass carbon (MBC) and alkaline soil pH was observed(Bickel and Or, 2020; 237 Majumdar et al., 2021). In this study, successive observations of soil pH and MBC in all four years 238 were assessed (Fig. 4). The contour map is a complex yet important plot to show the elemental 239 distribution in a field experiment with color coding from low to high. Each of the plots in Fig. 4 240 consisted of combined data from four years of harvesting time points from the three different 241 phases. The total microbial load was found to be the highest at this phase as shown in 242 Supplementary Fig. 1. The bioavailable concentration of each element was selected as the 'z-243 axis' or the distribution axis. A single plot contained soil pH, MBC data and bioavailable nutrient 244 contents for the three fields. In these plots the distribution of any element can be depicted from the 245 data range of MBC and pH from the different fields. For example, the distribution of N in the 246 figure can be seen to be higher near the contour region of MBC 3.3 to 3.8 mg, corresponding to 247 the pre-monsoon field soil, spanning to a wide range of soil pH. This indicated the higher N 248 availability at higher microbial biomass enriched soil of pre-monsoon fields. A comparatively 249 lower MBC corresponds with less available N in the soil within the same pH range in control or 250 monsoon fields. Similar results were found in the case of Ca, Mg and P distribution in three field 251 soils where higher MBC in pre-monsoon soil resulted in higher available elemental concentration. 252 The K availability was different to this trend possibly due to the greater requirement of K for 253 microbial growth and metabolic activity(Parmar and Sindhu, 2013). Hence, the lower MBC 254 content in soil resulted in higher available K in monsoon soils relative to the pre-monsoon soil. 255 With increasing pH, the availability of K also increased in pre-monsoon fields under low MBC 256 content. For Fe, the distribution was more or less similar to the K due to the redox changes and 257 utility of Fe by microbe's growth(Liesack et al., 2000; Jia et al., 2020). MBC within a range of 2.5 258 to 3 mg in monsoon soil had the highest range of available Fe compare to the control or pre-259 monsoon soils. The circulation of Cu and Zn was similar and only a higher range of MBC showed

260 the highest availability of these two micronutrients in pre-monsoon soil, irrespective of the soil 261 pH.

262 Principal components and statistical analysis of field soils

263 PCA plots with bi-plot distribution of selected elements within soils at three different sampling 264 phases have been shown in **Fig. 5**. In 2016, the pre-monsoon soil at the final phase of soil sampling 265 revealed the most variance in elemental content specifically K and Fe, whereas P, Ca and Mg were 266 more influenced during the beginning and middle of the cultivation in pre-monsoon field soil 267 (+PC1, +PC2). Compared to the pre-monsoon soils, monsoon soils were found to be less 268 statistically varied (-PC1, +PC2). The PCA analysis of 2017 data was different from the previous 269 plot in terms of elemental bi-plot distribution where most of the elements except P, showed high 270 variability at pre-monsoon harvest soils. Other monsoon soil data fall within either the (-PC1, 271 +PC2) or (-PC1, -PC2) quadrant. The propagation of selected elements in all the field soils in 2018 272 followed the same trend as in 2017. In the next year, the total elemental distribution bi-plot was 273 found to be the same but the PCA of field soils were variable. In 2019, the self-variance in the data 274 of pre-monsoon soil was observed as distributed in (+PC1, -PC2) to (+PC1, +PC2) quadrant. 275 Among the fields sampled, pre-monsoon field mainly constructed the PC1 and the discreet data 276 from control and monsoon makes PC2, depending on the concentration gradient and consecutive 277 distribution in those fields. Apart from the PCA, all the data from different fields and sampling 278 phases were statistically tested using one-way analysis of variance (ANOVA) at p < 0.05. All the 279 data were found significantly variable and independent to each other.

280 Metagenomics analysis of soil and nutrition modulation

281 Changed microbial diversity abundance under altered water regime

The total phylum diversity in the soil after Illumina next generation sequencing metagenomics analysis has been shown in **Supplementary Fig. 1**. The pie chart showed the predominance of different phyla present in the three analyzed soils. In developing soil, the Proteobacteria accounted for 33.7% followed by Acidobacteria 20%, Planctomycetes 9.8%, Chloroflexi 8.3%, Nitrospirae 6.7%, Actinobacteria 5.4%, Bacteroidetes 5.2%, Gemmatimonadetes 3.4%, Verrucomicrobia 2.9% and Firmicutes 1.7%. Apart from these dominating phyla, some other key phyla like cyanobacteria were identified, although in lesser abundance (see **Supplementary Fig. 1a** with 289 respective read counts). Soil samples from milking and harvest soils had more varied phyla 290 compared to the developing soil, as presented in **Supplementary Fig. 1b** and **1c** respectively. 291 Proteobacteria were appraised as 37.0% and 31.0% in monsoon and pre-monsoon soils. 292 Chloroflexi 24.4% was the second most dominating phylum in pre-monsoon soil followed by 293 Acidobacteria 15.47%, Nitrospirae 7.44%, Actinobacteria 4.72%, Planctomycetes 4.06%, 294 Bacteroidetes 3.32% and Verruocomicrobia 1.16%. In the monsoon soil, the community differed 295 by Chloroflexi 16.8%, Acidobacteria 13.04%, Nitrospirae 7.33%, Bacteoidetes 6.36%, 296 Actinobacteria 5.11%, Planctomycetes 4.28%, Verrucomicrobia 2.19%, Cyanobacteria 1.63% and 297 Gemmatimonadetes 1.15%. **Supplementary Fig. 2** shows the microbial class distribution in three 298 soil samples with respective dominance and availability in the soil as counted in the metagenomics 299 assessment. From the color code, the maximum and minimum dominance can be seen and from 300 the size of the circle, the degree of availability can be understood. Alpha- and Betaproteobacteria 301 were dominant in harvest soil whereas Gamma- and Deltaproteobacteria were dominant 302 respectively in developing and milking soil phases.

303 Phylogenic interaction and nutrition modulation in soil

Microbial inter-relation in the soil environment represented by a krona chart for all the three soils (Fig. 6), that helps in the visualization and exploration of relative abundances in all the phylogenic levels of microbial classifications, generated in HTML format. This krona chart creates a radial space-filled multi-layered colored plot with interactive zooming to each of the phylogenic levels analyzed for the overall microbial community(Ondov et al., 2011). Fig. 6 represents a general graphical form where bacterial and archaeal communities are in different shades of color. Also, from these three kronas, it was observed that the presence of archaea was higher in the milking and harvesting phase soils, mostly related to anaerobic methanogenic species, compared to the developing phase soils. Influence of pre-monsoon (**a**, **b**, **c**) was also clearly found to be greater than the monsoonal soils (**a1**, **b1**, **c1**). Microbial count of some specific groups present in monsoon soil was higher than the other soil samples. But dominating this quantitative data, the nutritional release in the pre-monsoon soil was higher due to the combined effects of the irrigation process and heterogeneous microbial metabolic activities as it depicts the differences in alpha diversity by roomparing the library sizes for each of the samples, whereas beta diversity elucidates the species richness within an ecological community. **Supplementary Table 4** shows the analyzed alpha and 319 beta diversity indices for all three soil samples. The inter-ecological community comparison is 320 generally made by calculating the beta diversity and from the observed results, both alpha and beta 321 diversity showed that the pre-monsoon soil had higher species richness due to the changes in 322 irrigation practices and nutrient availability. Statistical approach justifies the metagenomics results 323 for the monsoon soil having variable soil conditions leading towards an unstable microbial 324 community compared to pre-monsoon soil. The species-specific dependency of each microbial 325 species had been analyzed by principal component (PCA) and principal coordinate analysis 326 (PCoA) as in **Supplementary Fig. 3**. PCA analysed in a bi-plot model and among the three-phase 327 soil microbiome, developing microbiota was placed in the first quadrant (+PC1, +PC2) whereas 328 harvest and milking biome placed into the fourth quadrant (+PC1, -PC2). PCoA suggests the 329 dissimilarities between samples analyzing the ordination space with a linear mapping of Euclidean 330 data and the observed variance within the original data (Ramette, 2007). The metagenomics data 331 presented here showed that most of the species' occurrence was centric to the plot and some of the 332 discrete points were dispersed and found to be unidentified species of microbes. Fig. 7 shows an 333 interactive microbial network analysis depending on the soil nutrient content and dispersion in the 334 field. During data loading and the network formation, microbial phylum and classes were 335 accounted with their specific encountered numeric values in the metagenomic sequence analysis 336 along with soil elemental concentrations. Microbial phyla were selected as source nodes while 337 microbial classes were selected as target nodes. Numeric values for each set of the field microbial 338 data were selected as target edge attribute followed by elemental concentration data as target node 339 attribute. The resulted network shows microbial classes interacting with each other while 340 modulating the nutrients content in soil.

341 Discussion

342 Plants were cultivated using stored pond water, that itself is not elementally enriched in 343 nutrients(Shrivastava et al., 2014) and application of such surface water did not increase the 344 elemental content any further except the natural redox de-coupling of elements from soil 345 complexes. The water application to dry soil releases nutrients to the soil-aqueous phase 346 (Venterink et al., 2002) that in turn can induce the microbial metabolic activity, exchange of 347 electrolytes in soil, changing the redox status of different elements and hence, the bioavailability 348 for crops changes (Miller et al., 2005; Wu and Brookes, 2005). Some reports suggest that the 349 mineralization of N increases during the drying phase after the water drainage with increasing soil 350 aeration and subsequent denitrification occurring in semi-dry soil(Fageria et al., 2010; Lu et al., 351 2020). The combination of these two processes leads to the increased bioavailability of soil N to 352 the plant system. The flushing out of surface soil elemental complexes increased the chance of 353 electrolyte release into the stagnant water (Majumdar et al., 2019; Upadhyay et al., 2021). A 354 prevailing soil aqueous phase with lower ORP values and reductive environment can alter the Fe-355 redox state(Majumdar and Bose, 2018) and this conversion will influence the available content of 356 Fe along with other nutritional elements in a monsoon field. In soil, the availability of P and K 357 depends on the drained water that controls the physical adsorption and chemical equilibria of these 358 elements on either soil particles or other elemental complexes (Parmar and Sindhu, 2013; Song et 359 al., 2021). Hence, the soluble content of K is more available to the plant systems under waterlogged 360 soil compared to the semi-arid soil. Although, soluble P (inorganic forms) are easily available to 361 the plant system and due to the consecutive re-watering of dry fields, this extractable P content 362 becomes around 40% higher than in longer-term monsoon fields(Bünemann et al., 2013). SOM 363 plays an important role in the association of Cu and Zn to other elemental complexes and controls 364 the bioavailability(Fageria et al., 2010). Hence, the bioavailability of Cu and Zn in control and pre-365 monsoon field soil was higher than the monsoon soil (see Fig. 3). The distribution of Cu and Zn, 366 hence, was higher in the region of high MBC (Fig. 4) which corresponds to the pre-monsoon field 367 throughout the pH ranges. In soil, nutrient dissolution from complexes and bioavailable nutrient 368 pools are crucial factors that influence soil microbial biomass and in turn, their heterogenic 369 metabolic activities under neutral pH and temperature(Leita et al., 1999; Wang et al., 2016; Zhou 370 et al., 2020). More availability of these macro-micronutrients in soil under pre-monsoonal wetting-371 drying conditions helps crop growth and yield better than the flooded condition at monsoonal 372 season (Fig. 3). Harvest index by calculating the plant leaf area, tiller numbers, root to shoot 373 lengths and panicle numbers were found to be greater under the AWD cultivation (data not shown). 374 Microbial biomass stoichiometry and community composition can variably perform soil nutrient 375 mineralization and immobilization to influence soil fertility(Heuck et al., 2015; Delgado-376 Baquerizo et al., 2016). Under pre-monsoonal drying-wetting cycles in fields, the intermittent 377 water application at certain times enhances the available soil C-N-P pool and help soil microbes 378 grow better than under continuously monsoon soil.

379 From the metagenomics data, it can be seen that the encountered reads of occurrence of a single 380 species or genus were higher in monsoon soil but the overall species diversity was greater in pre-381 monsoon soil. The species richness has confirmed that the harvesting phase had a greater 382 community diversity which can modulate the nutrient pool to a greater extent as shown in the Fig. 383 4. Chloroflexi is such a metabolically heterogeneous group of phyla that consists of both 384 anoxygenic-photoautotrophy as well as oxygenic chemoheterotrophy(Gupta, 2013). The high 385 percentage of Chloroflexi in the pre-monsoon soils indicate the semi-aerobic cultivation with 386 possibly abundant nutritional pulses enhanced by the Chloroflexi growth. In pre-monsoon drying-387 wetting, a subtle change of pH and redox potential during the water content alteration resulted in 388 the release of available N, P, K, Ca and Mg that could have triggered the predominance of another 389 metabolically diverse microbial phylum Acidobacterium. Like Proteobacteria, this phylum can 390 also thrive under the differential supply of macro-micronutrients in soils, stimulating the 391 conversion of soil nitrite and nitrate, various types of organic matter and carbohydrate 392 utilization(Kielak et al., 2016). These two phyla are also present in the monsoon soils although at 393 the lower abundance. The higher predominance of Cyanobacteria, was found in the monsoon soil 394 due to continuous flooding in the field. Cyanobacteria participate in soil N-cycle dynamics and that allows the growth of another phylum, the Nitrospirae. Both of these phyla were in high 395 396 abundance in the monsoon soils as found from the metagenomics analysis. Ammonia production 397 under water logging promotes the dominance of the Nitrospirae(Vlek and Craswell, 1981) while further conversion to nitrite enhances Cyanobacterial growth. Alpha-proteobacteria can participate 398 399 in diverse soil activities due to their variable metabolic abilities like mineralization, metal 400 mobilization by changing the soil redox status, methanogenesis etc. The plethora of *Gemmatas*p. 401 and Pirllula sp. that comes under the phylum Plactomycetes, were higher in the pre-monsoon 402 milking soil phase due to their aerobic chemoheterotrophic metabolism. During this phase, water 403 application was ceased resulting aerobic soil status. The highly abundant genus DA101 of phylum 404 Verrucomicrobia was found to be more dominant in developing soil compare to the milking or 405 harvesting soils, although its proper assessment and role in environmental ecology is still unclear 406 except that these microbes participate in geochemical cycling in soils. The microbial distribution 407 and inter-connectivity were found to be influenced by or linked to the differential distribution of 408 elements in soil. Chloflexi, Proteobacteria, Planctomycetes, Acidobacteria, Cyanobacteria, 409 Verrucomicrobia, Actinobacteria and Bacteroidetes are the main influencers in the observed

410 microbial network analysis where these microbial groups are in the centre of the distribution, their 411 influence branching out to other microbial groups. The network pattern was similar in the 412 developing and milking phase soils but was changed in the harvesting phase. A large part of this 413 metagenomics data revealed the crucial role of unidentified microbes in nutrient cycling as well as 414 inter-microbial community interactions, but their probable mechanism remains uncertain.

415 Some earlier studies reported that the very high water-holding capacity of soil (i.e. monsoon 416 condition) can result in a continual decrease in the total N mineralization compared to the 50-70% 417 water content soil or the re-watering process(Denef et al., 2001; Chen et al., 2012). After re-418 watering of the paddy soil, aerobic microbes consume the oxygen present and the stagnant anoxic 419 condition triggers other anaerobic or facultative microbes to use other electron acceptors present 420 in the soil like nitrate, Mn(IV), Fe(III), SO42- and CO2, respectively(Liesack et al., 2000). 421 Verrucomicrobia and rice field archaea are predominant in the monsoon soil, influencing the 422 polysaccharolysis and methanogenesis in soil(Schütz et al., 1989; Janssen et al., 1997). In anoxic 423 soil, oxygen-independent Fe oxidation process occurs by nitrate utilizing Fe-oxidizer microbes 424 (Ratering and Schnell, 2000). In soil, the bioavailable content of P is also less (around 0.05% of 425 the total P available in soil) and many species of free-living bacteria, endophytic bacteria and 426 actinobacteria convert that unavailable fraction of organic P to phyto-available forms(Adhya et al., 2015). Microbial dissolution of K from mineral complexes is generally performed by several 427 428 rhizobacterial groups including several species of Bacillus sp., Acidothobacillus sp., Pseudomonas sp., Rhizobium sp., Burkholderia sp.(Sattar et al., 2019), that are mostly predominate in the pre-429 430 monsoon cycling soil. Incremental release of soil Ca and Mg under pre-monsoon soil compared to 431 the monsoon soil has been observed in this study, which also might result from additional 432 microbial mineralization. Microbial mineralization of Ca in soil was found to increase in the 433 presence of more available Mg(Chen et al., 2018). Based on previous reports and the findings from 434 this present study, it must be emphasized that the combined effect of drying-wetting cycles of 435 irrigation can help maintain a thriving soil microbial community that in turn can influence the 436 availability of soil nutrients in the plant rhizosphere. This field study showed the increased total 437 and bioavailable concentrations of nutrients in pre-monsoon soil was due to the frequent change 438 in soil physico-chemical parameters when compared with conventional stagnant monsoon irrigated 439 soil. This correlative research has established the role of different microbial phyla in the release of 440 nutrient pulses under the pre-monsoon soil cycles providing higher bioavailability to the crop.

441 Therefore, the pre-monsoon drying-wetting cycle of field management and microbial activities can442 release nutrients in a more bioavailable form while compared to the conventional monsoon443 irrigation process.

444 Concluding remarks

In conclusion it can be stated that rice cultivation under pre-monsoon drying-wetting 445 446 irrigation has been proven to be efficient which further found to be effective in modulating soil 447 biota. This field study showed the increased total and bioavailable concentrations of nutrients in 448 pre-monsoon soil at all the three plant growth phases, was due to the frequent change in soil 449 physico-chemical parameters compared to the conventional monsoon soil. Cultivated rice plant 450 uptake of total nutrients was measured higher in pre-monsoonal cultivation at all the three growth 451 phases- developing, milking and harvest than the monsoonal cultivations. Cycling of water 452 stagnancy, release and re-watering during the pre-monsoon season was found useful in loosing the 453 bulk nutrients from the soil minerals, increasing the plant nutrient quality as we as the microbial 454 biomass load in the soil than the prolonged waterlog during the monsoonal season causing flushed 455 out the surface nutrient content and less microbial biomass load. Natural soil microbial diversity 456 has been counted with their species richness in these respective soils with greater activity and 457 diversity observed by metagenomics analysis in the pre-monsoon soil with highest richness at the 458 harvest phase. With further correlative research, the role of different microbial phyla in the release 459 of nutritional pulses under the pre-monsoon soil with higher bioavailability to the plant system has 460 been established. Hence, a combinatory effect of pre-monsoon field management and microbial 461 activities in releasing nutritional elements in a more bioavailable form compared to the 462 conventional monsoon cultivation process has been proven.

463 Acknowledgment

464 The authors are thankful to IISER-K for providing infrastructure as well as core facilities (XRF 465 DST-FIST laboratory). The authors are also thankful to BHU lab facilities. AM is thankful to the 466 Ministry of Earth Sciences, India (MoES/P.O. (Geosci)/56/2015) for providing JRF and IISER 467 Kolkata for providing SRF. Bulk of the fellowship amount to AM has been used in research 468 expenses. MKU is thankful to the Council of Scientific and Industrial Research (CSIR) for CSIR-469 SRF (09/013(0853)/2018-EMR-I). PKD is thankful for UGC-SRF. Authors are also thankful to 470 Genotypic Technology [P] Ltd., Bangalore, India for the metagenomics analysis. Authors are
471 cordially thankful to Prof. Anthony J. Miller, John Innes Centre, Norwich UK for his unconditional
472 support in form of thorough revision of this manuscript.

473 Authors Credit

474 AM conceptualized the idea; AM performed field works, lab experiments and manuscript writing; 475 MKU performed partial lab work and manuscript writing; PKD interpreted analysed dataset and 476 contributed in revising manuscript; BG performed XRF analysis and manuscript writing; SS, and 477 AKS revised and finalized the manuscript.

478 Authors would also like to thank Mr. Kedarnath Majumdar, landowner who selflessly helped with479 the field set up and management during this long-term experiment.

480 Conflict of interest

481 Authors declare that there is no conflict of interest.

482

483 References

484 Adhya, T.K., N. Kumar, G. Reddy, A.R. Podile, H. Bee, et al. 2015. Microbial mobilization of

soil phosphorus and sustainable P management in agricultural soils. Curr. Sci. 108(7):

486 1280–1287. doi: 10.18520/cs/v108/i7/1280-1287.

- 487 Afsal, F., A. Majumdar, J.S. Kumar, and S. Bose. 2020. Microbial Inoculation to Alleviate the
- 488 Metal Toxicity in Crop Plants and Subsequent Growth Promotion. Sustain. Solut. Elem.
- 489 Defic. Excess Crop Plants: 451–479. doi: 10.1007/978-981-15-8636-1_17.

490 Baldwin, D.S., and A.M. Mitchell. 2000. The effects of drying and re-flooding on the sediment

- 491 and soil nutrient dynamics of lowland river-floodplain systems: A synthesis. River Res.
- 492 Appl. 16(5): 457–467. doi: 10.1002/1099-1646(200009/10)16:5<457::aid-rrr597>3.0.co;2-
- 493 b.
- 494 Bickel, S., and D. Or. 2020. Soil bacterial diversity mediated by microscale aqueous-phase
 495 processes across biomes. Nat. Commun. 11(1): 1–9.

496 Birch, H.F. 1958. The effect of soil drying on humus decomposition and nitrogen availability.
497 Plant Soil 10(1): 9–31. doi: 10.1007/BF01343734.

498 Blackwell, M.S.A., P.C. Brookes, N. de la Fuente-Martinez, H. Gordon, P.J. Murray, et al. 2010.

499 Phosphorus Solubilization and Potential Transfer to Surface Waters from the Soil Microbial

500 Biomass Following Drying–Rewetting and Freezing–Thawing. 1st ed. Elsevier Inc.

501 Bünemann, E.K., B. Keller, D. Hoop, K. Jud, P. Boivin, et al. 2013. Increased availability of

phosphorus after drying and rewetting of a grassland soil: processes and plant use. Plant
Soil 370(1): 511–526.

504 Chen, Y., R. Liu, C. Sun, P. Zhang, C. Feng, et al. 2012. Spatial and temporal variations in

nitrogen and phosphorous nutrients in the Yangtze River Estuary. Mar. Pollut. Bull. 64(10):
2083–2089.

507 Chen, T., P. Shi, Y. Li, T. Duan, Y. Yu, et al. 2018. Biomineralization of varied calcium

carbonate crystals by the synergistic effect of silk fibroin/magnesium ions in a microbial
system. CrystEngComm 20(17): 2366–2373.

510 Dangar, S., A. Asoka, and V. Mishra. 2021. Causes and implications of groundwater depletion in
511 India: A review. J. Hydrol. 596: 126103.

512 Delgado-Baquerizo, M., F.T. Maestre, P.B. Reich, T.C. Jeffries, J.J. Gaitan, et al. 2016.

513 Microbial diversity drives multifunctionality in terrestrial ecosystems. Nat. Commun. 7(1):
514 1–8.

515 Denef, K., J. Six, H. Bossuyt, S.D. Frey, E.T. Elliott, et al. 2001. Influence of dry-wet cycles on

the interrelationship between aggregate, particulate organic matter, and microbial

517 community dynamics. Soil Biol. Biochem. 33(12-13): 1599–1611.

518 Dodd, I.C., J. Puértolas, K. Huber, J.G. Pérez-Pérez, H.R. Wright, et al. 2015. The importance of

soil drying and re-wetting in crop phytohormonal and nutritional responses to deficit

520 irrigation. J. Exp. Bot. 66(8): 2239–2252. doi: 10.1093/jxb/eru532.

521 Dubey, P.K., G.S. Singh, and P.C. Abhilash. 2020. Agriculture in a changing climate. Adaptive
522 Agricultural Practices. Springer. p. 1–10

523 Dubey, P.K., A. Singh, R. Chaurasia, K.K. Pandey, A.K. Bundela, et al. 2021a. Planet friendly
524 agriculture: Farming for people and the planet. Curr. Res. Environ. Sustain. 3: 100041.

Dubey, P.K., A. Singh, R. Chaurasia, K.K. Pandey, A.K. Bundela, et al. 2022. Animal manures
and plant residue-based amendments for sustainable rice-wheat production and soil fertility
improvement in eastern Uttar Pradesh, North India. Ecol. Eng. 177: 106551.

528 Dubey, P.K., A. Singh, R. Chaurasia, K.K. Pandey, and S. Singh. 2021b. Fostering Nature-Based

529 Solutions for Ecorestoration and Attaining United Nations Sustainable Development Goals.

Fageria, N.K., V.C. Baligar, and C.A. Jones. 2010. Growth and mineral nutrition of field crops.
CRC Press.

532 Gupta, R.S. 2013. Molecular markers for photosynthetic bacteria and insights into the origin and
533 spread of photosynthesis. Adv. Bot. Res. 66: 37–66.

Gupta, A., A. Majumdar, and S. Srivastava. 2022. Approaches for assisted phytoremediation of
arsenic contaminated sites. Assisted Phytoremediation. Elsevier. p. 221–242

Hasanuzzaman, M., K. Nahar, and M. Fujita. 2018. Mechanisms of arsenic toxicity and tolerance
in plants. Mech. Arsen. Toxic. Toler. Plants (November): 1–508. doi: 10.1007/978-981-131292-2.

Heuck, C., A. Weig, and M. Spohn. 2015. Soil microbial biomass C: N: P stoichiometry and
microbial use of organic phosphorus. Soil Biol. Biochem. 85: 119–129.

Huang, Z., E.L. Nya, M.A. Rahman, T.B. Mwamila, V. Cao, et al. 2021. Integrated water
resource management: Rethinking the contribution of rainwater harvesting. Sustainability
13(15): 8338.

544 Janssen, P.H., A. Schuhmann, E. Mörschel, and F.A. Rainey. 1997. Novel anaerobic

545 ultramicrobacteria belonging to the Verrucomicrobiales lineage of bacterial descent isolated

by dilution culture from anoxic rice paddy soil. Appl. Environ. Microbiol. 63(4): 1382–
1388.

548 Jia, R., K. Wang, L. Li, Z. Qu, W. Shen, et al. 2020. Abundance and community succession of

19

- nitrogen-fixing bacteria in ferrihydrite enriched cultures of paddy soils is closely related to
 Fe (III)-reduction. Sci. Total Environ. 720: 137633.
- 551 Johannesson, K.H., and K. Neumann. 2013. Geochemical cycling of mercury in a deep, confined
- aquifer: Insights from biogeochemical reactive transport modeling. Geochim. Cosmochim.

553 Acta 106: 25–43.

- 554 Kabata-Pendias, A. 2000. Trace elements in soils and plants. CRC press.
- 555 Kielak, A.M., C.C. Barreto, G.A. Kowalchuk, J.A. van Veen, and E.E. Kuramae. 2016. The
- ecology of Acidobacteria: moving beyond genes and genomes. Front. Microbiol. 7: 744.
- 557 Kumarathilaka, P., S. Seneweera, A. Meharg, and J. Bundschuh. 2018. Arsenic speciation
- dynamics in paddy rice soil-water environment: sources, physico-chemical, and biological
- 559 factors-a review. Water Res. 140: 403–414.
- Laha, M., H. Arambagh, and W. Bengal. 2013. Social Implication of Arsenic Pollution in
 Eastern Barddhaman. Transactions 35(2): 173–184.
- 562 Leita, L., M. De Nobili, C. Mondini, G. Muhlbachova, L. Marchiol, et al. 1999. Influence of
- inorganic and organic fertilization on soil microbial biomass, metabolic quotient and heavy

metal bioavailability. Biol. Fertil. Soils 28(4): 371–376.

- Liesack, W., S. Schnell, and N.P. Revsbech. 2000. Microbiology of flooded rice paddies. FEMS
 Microbiol. Rev. 24(5): 625–645. doi: 10.1016/S0168-6445(00)00050-4.
- 567 Lu, T., Y. Wang, H. Zhu, X. Wei, and M. Shao. 2020. Drying-wetting cycles consistently
- increase net nitrogen mineralization in 25 agricultural soils across intensity and number of
- drying-wetting cycles. Sci. Total Environ. 710: 135574.
- 570 Majumdar, A. 2021. Arsenic Bio-regulation Management and Altered Plant Physiology under
 571 Different Irrigation Regime.
- 572 Majumdar, A., A. Barla, M.K. Upadhyay, D. Ghosh, P. Chaudhuri, et al. 2018.
- 573 Vermiremediation of metal(loid)s via Eichornia crassipes phytomass extraction: A
- sustainable technique for plant amelioration. J. Environ. Manage. 220(May): 118–125. doi:

576 Majumdar, A., and S. Bose. 2017. Toxicogenesis and metabolism of arsenic in rice and wheat
577 plants with probable mitigation strategies. Arsen. Risks Expo. Behav. Environ. Toxicol.:
578 149–166.

579 Majumdar, A., and S. Bose. 2018. A glimpse on uptake kinetics and molecular responses of

arsenic tolerance in Rice plants. Mechanisms of arsenic toxicity and tolerance in plants.

581 Springer. p. 299–315

582 Majumdar, A., J.S. Kumar, Sheena, and S. Bose. 2020a. Agricultural Water Management

583 Practices and Environmental Influences on Arsenic Dynamics in Rice Field. Arsen. Drink.

584 Water Food: 425–443. doi: 10.1007/978-981-13-8587-2_17.

585 Majumdar, A., A. Shrivastava, S.R. Sarkar, S. Sathyavelu, A. Barla, et al. 2020b. Hydrology,

sedimentation and mineralisation: A wetland ecology perspective. Clim. Chang. Environ.
Sustain. 8(2): 134–151.

588 Majumdar, A., M.K. Upadhyay, B. Giri, S. Srivastava, A.K. Srivastava, et al. 2021. Arsenic

589 dynamics and flux assessment under drying-wetting irrigation and enhanced microbial

590 diversity in paddy soils: A four year study in Bengal delta plain. J. Hazard. Mater.

591 409(November): 124443. doi: 10.1016/j.jhazmat.2020.124443.

592 Majumdar, A., M.K. Upadhyay, J.S. Kumar, Sheena, A. Barla, et al. 2019. Ultra-structure

solution alteration via enhanced silicon uptake in arsenic stressed rice cultivars under intermittent

594 irrigation practices in Bengal delta basin. Ecotoxicol. Environ. Saf. 180(May): 770–779.

595 doi: 10.1016/j.ecoenv.2019.05.028.

596 Majumdar, A., M.K. Upadhyay, M. Ojha, F. Afsal, B. Giri, et al. 2022. Enhanced

597 phytoremediation of Metal (loid) s via spiked ZVI nanoparticles: An urban clean-up

strategy with ornamental plants. Chemosphere 288: 132588.

599 Mikha, M.M., C.W. Rice, and G.A. Milliken. 2005. Carbon and nitrogen mineralization as

affected by drying and wetting cycles. Soil Biol. Biochem. 37(2): 339–347. doi:

601 10.1016/j.soilbio.2004.08.003.

21

Miller, A.E., J.P. Schimel, T. Meixner, J.O. Sickman, and J.M. Melack. 2005. Episodic rewetting
enhances carbon and nitrogen release from chaparral soils. Soil Biol. Biochem. 37(12):
2195–2204.

Mooshammer, M., F. Hofhansl, A.H. Frank, W. Wanek, I. Hämmerle, et al. 2017. Decoupling of
microbial carbon, nitrogen, and phosphorus cycling in response to extreme temperature
events. Sci. Adv. 3(5): 1–14. doi: 10.1126/sciadv.1602781.

Mukherji, A. 2007. The energy-irrigation nexus and its impact on groundwater markets in
eastern Indo-Gangetic basin: Evidence from West Bengal, India. Energy Policy 35(12):
6413–6430.

611 Nickson, R.T., J.M. McArthur, P. Ravenscroft, W.G. Burgess, and K.M. Ahmed. 2000.

612 Mechanism of arsenic release to groundwater, Bangladesh and West Bengal. Appl.

613 geochemistry 15(4): 403–413.

Nishio, T., H. Sekiya, K. Toriyama, and K. Kogano. 1994. Changes in gross rates of nitrogen
transformations in soil caused by conversion of paddy fields to upland fields. Soil Sci. Plant
Nutr. 40(2): 301–309.

617 Ondov, B.D., N.H. Bergman, and A.M. Phillippy. 2011. Interactive metagenomic visualization in
a Web browser. BMC Bioinformatics 12(1): 385. doi: 10.1186/1471-2105-12-385.

619 Parmar, P., and S.S. Sindhu. 2013. Potassium Solubilization by Rhizosphere Bacteria: Influence

of Nutritional and Environmental Conditions. J. Microbiol. Res. 3(1): 25–31. doi:

621 10.5923/j.microbiology.20130301.04.

622 Pearson, K.A., G.M. Millar, G.J. Norton, and A.H. Price. 2018. Alternate wetting and drying in

Bangladesh: Water-saving farming practice and the socioeconomic barriers to its adoption.

624 Food Energy Secur. 7(4): 1–12. doi: 10.1002/fes3.149.

Ramette, A. 2007. Multivariate analyses in microbial ecology. FEMS Microbiol. Ecol. 62(2):
142–160. doi: 10.1111/j.1574-6941.2007.00375.x.

Ratering, S., and S. Schnell. 2000. Localization of iron-reducing activity in paddy soilby profile
studies. Biogeochemistry 48(3): 341–365.

- 629 Ray Sarkar, S., A. Majumdar, A. Barla, N. Pradhan, S. Singh, et al. 2017. A conjugative study of
- Typha latifolia for expunge of phyto-available heavy metals in fly ash ameliorated soil.
- 631 Geoderma 305(November): 354–362. doi: 10.1016/j.geoderma.2017.06.022.
- 632 Sattar, A., M. Naveed, M. Ali, Z.A. Zahir, S.M. Nadeem, et al. 2019. Perspectives of potassium
- 633 solubilizing microbes in sustainable food production system: A review. Appl. Soil Ecol.
- 634 133(July): 146–159. doi: 10.1016/j.apsoil.2018.09.012.
- 635 Schütz, H., W. Seiler, and R. Conrad. 1989. Processes involved in formation and emission of
 636 methane in rice paddies. Biogeochemistry 7(1): 33–53.
- 637 Shrivastava, A., A. Barla, A. Majumdar, S. Singh, and S. Bose. 2020. Arsenic mitigation in rice
- 638 grain loading via alternative irrigation by proposed water management practices.
- 639 Chemosphere 238: 124988. doi: 10.1016/j.chemosphere.2019.124988.
- 640 Shrivastava, A., A. Barla, H. Yadav, and S. Bose. 2014. Arsenic contamination in shallow
- groundwater and agricultural soil of Chakdaha block, West Bengal, India. Front. Environ.Sci. 2: 50.
- 643 Song, T., D. Das, Q. Hu, F. Yang, and J. Zhang. 2021. Alternate wetting and drying irrigation
- and phosphorus rates affect grain yield and quality and heavy metal accumulation in rice.
- 645 Sci. Total Environ. 752: 141862.
- 646 Surendran, U., P. Raja, M. Jayakumar, and S.R. Subramoniam. 2021. Use of efficient water
- saving techniques for production of rice in India under climate change scenario: A critical
 review. J. Clean. Prod. 309: 127272.
- 649 Tulip, S.S., M.S. Siddik, M.N. Islam, A. Rahman, A.T. Haghighi, et al. 2022. The impact of
- 650 irrigation return flow on seasonal groundwater recharge in northwestern Bangladesh. Agric.
- 651 Water Manag. 266: 107593.
- 652 Upadhyay, M.K., A. Majumdar, A. Barla, S. Bose, and S. Srivastava. 2021. Thiourea
- supplementation mediated reduction of grain arsenic in rice (Oryza sativa L.) cultivars: A
- two year field study. J. Hazard. Mater. 407(September 2020): 124368. doi:
- 655 10.1016/j.jhazmat.2020.124368.

- 656 Upadhyay, M.K., A. Majumdar, A.K. Srivastava, S. Bose, P. Suprasanna, et al. 2022.
- 657 Antioxidant enzymes and transporter genes mediate arsenic stress reduction in rice (Oryza
- sativa L.) upon thiourea supplementation. Chemosphere 292: 133482.
- 659 Venterink, H.O., T.E. Davidsson, K. Kiehl, and L. Leonardson. 2002. Impact of drying and re660 wetting on NPK dynamics in a wetland soil.pdf. : 119–130.
- Vlek, P.L., and E.T. Craswell. 1981. Ammonia volatilization from flooded soils. Fertil. Res.
 2(4): 227–245.
- Walkley, A., and I.A. Black. 1934. An examination of the Degtjareff method for determining soil
 organic matter, and a proposed modification of the chromic acid titration method. Soil Sci.
 37(1): 29–38.
- 666 Wang, Y., C.R. Jensen, and F. Liu. 2017. Nutritional responses to soil drying and rewetting
- 667 cycles under partial root-zone drying irrigation. Agric. Water Manag. 179: 254–259. doi:
 668 10.1016/j.agwat.2016.04.015.
- 669 Wang, J., F. Pan, J. Soininen, J. Heino, and J. Shen. 2016. Nutrient enrichment modifies
 670 temperature-biodiversity relationships in large-scale field experiments. Nat. Commun. 7(1):
 671 1–9.
- Wu, J., and P.C. Brookes. 2005. The proportional mineralisation of microbial biomass and
 organic matter caused by air-drying and rewetting of a grassland soil. Soil Biol. Biochem.
 37(3): 507–515.
- 675 Xiang, S.-R., A. Doyle, P.A. Holden, and J.P. Schimel. 2008. Drying and rewetting effects on C
- and N mineralization and microbial activity in surface and subsurface California grassland
- 677 soils. Soil Biol. Biochem. 40(9): 2281–2289.
- 678 Zhou, Z., C. Wang, and Y. Luo. 2020. Meta-analysis of the impacts of global change factors on
 679 soil microbial diversity and functionality. Nat. Commun. 11(1): 1–10.
- 680
- 681

Fig 1. Radar plot distribution of total soil elemental concentration in four consecutive years 2016-2019.





This radar plot shows a layered representation of element concentrations in two seasons with three sampling phases in four consecutive years from 2016 to 2019 under a same paddy field (**a-d**). This plot uses the average data of triplicated results. Radar plot shows the result in a centre-out form with lowest data at the core and the highest data at the verge. Dark grey colored lines inside the plot represents the minor grid that is sub-set of each data point. Major data range are mentioned with a gap of 10000 unit (unit = mg kg⁻¹) and sub set data gap represents 2000 unit each. This plot shows the concentration change of selected elements under field conditions with corresponding sub-set lines. From the graph, it can be seen that the pre-monsoonal soil contained highest amount of nutrients compared to the monsoonal soils. Data has been validated using one-way ANOVA (p < 0.05) while plotting.





Rice plants grown in two different seasons accumulated differential concentrations of nutritional elements at three growth phases in four consecutive years from 2016 to 2019. Summed results of root and shoot analysis has been presented for a better assessment of element contents in whole rice plants. Selected elements concentration has been presented as a mean of triplicate data with standard deviation. One-way ANOVA (p<0.05) was used for data validation and for statistical differences in each data set, DMRT was used.

Fig 3. Correlation matrix plot of soil physico-chemical parameters and bioavailable nutritional elements in three phases and relative year-wise yield data.



Soil physico-chemical parameters like pH, redox potential (ORP) and organic matter (OM) can influence bioavailability of soil elements. This figure represents the correlation matrix of bioavailable content of nutritional elements in soil with pH, ORP and OM under three different plant growth phases (a) developing, (b) milking and (c) harvest. All the data from four years are compared together to check the effect of plant growth phases on soil nutritional availability. Color code here represents a minimum (dark red) to maximum (violet) correlation whereas the sharpness in shape and direction of ovals also shows the degree of correlation. The more it is sharp, the more strongly correlated and upwards to downwards indicates either positive or negative relations. Grey colored shaded boxes indicates the p<0.05 significance results. The nutritional availability was found to be effective in yield enhancement as shown in (d) with yearly increment in the pre-monsoonal season. DMRT was used for statistical difference assessment.

Fig 4. Bioavailable concentration distribution of selected elements in field conditions under the influence of soil pH and microbial biomass.



Bioavailability of crop nutritional elements are dispersive in nature and their distribution in the soil varies with active microbial biomass that in turn depends on soil pH. This contour map shows the soil pH and MBC of all the three phase soils together showing the pattern of elemental content distribution, getting changed with the other two. Due to the differential pH and MBC content in these soils, distribution of elements also gets changed. Each element has a different range of bioavailability and that has been represented in a colour code to understand the pattern of concentration distribution.

Fig 5. Principal component analysis of selected elements during three plant growth phases in four years (2016-2019).



In four consecutive years 2016-2019 (**a-d**), the soil elements are differentially measured and found that all the soils have diverse content of elements and getting influenced under field conditions along with the plant growth phases. Among the analyzed components, first two components are presented here with variance % mentioned in parenthesis. Green lines are showing the bi-plot model in this multi-component analysis.

Fig 6. Krona chart of microbial diversity with relative abundance percent analysis of two seasonal soils at three growth phases.







Krona interactive plot of microbial phylogenic classifications within each sample with occurrence percentages has been shown here. Developing (**a-a1**), milking (**b-b1**) and harvest (**c-c1**) phases soil samples were analysed using this krona plot where predominant bacterial and archaeal communities are represented in different shades of color. Samples from both pre-monsoonal (**a**, **b**, **c**) and monsoonal (**a1**, **b1**, **c1**) seasons have been considered for microbial diversity and abundance cross-check. The relative abundance of predominant microbial classes is presented here with mentioned percent obtained in the metagenomics analysis. 1 Fig 7. Microbial network analysis of three growth phase soils under influence of elemental content in pre-monsoonal and 2 monsoonal seasons.





5 Microbial network analysis of three different phase soils- developing, milking and harvest (a-a1, b-b1, c-c1), shows different distribution
6 and connection pattern depending on the selected elemental distribution and content in the field at two different seasonal point of rice
7 cultivation- pre-monsoon (a, b, c) and monsoon (a1, b1, c1). Nutrient change in the soil system modulates microbial community
8 distribution and this graph shows such interactive patterns. Nodes in these graphs are representing microbial classes and edges are
9 numeric values of element's concentration. These graphs are made using Cytoscape (version 3.8.2).

Effect of combined dry-wet irrigation and microbial dynamics on soil nutrient bioavailability

Contents-

Methodology- Details process of metagenomics analysis.

Supplementary Figure 1. Soil microbial phylum diversity under three field conditions (a-c).

Supplementary Figure 2. Microbial class distribution in three field soils with availability and dominance.

Supplementary Figure 3. Principal component analysis (PCA) and principal coordinate analysis (PCoA) in three different phase soil microbiomes.

Supplementary Table 1. Details of sequential extraction analysis method with reactant application dosages.

Supplementary Table 2. Instrument specifications during ICP-MS analysis.

Supplementary Table 3. Instrument specifications of Wavelength Dispersive XRF.

Supplementary Table 4. Alpha and Beta diversity indices of soil microbiome in three phase soil samples.

Detailed process of metagenomics study-

• DNA extraction from soil and metagenomic library preparation

A small amount of soil (0.25 gm) was used for the DNA isolation using QiagenDNeasy Powersoil Kit (Cat#12888-50). Soil samples were vortexed for 2 mins within a Powerbead tube in which 60µl of solution C1 was added and vortexed again. Rest of the process was followed according to the kit manufacturer guidelines. 50µl of nuclease free water (Ambion, Cat#AM9938) was used for the DNA elution. Nanodrop 2000 and agarose gel electrophoresis were used for qualitative and quantitative analysis. Region-specific targeting proprietary primers at Genotypic Technology Pvt. Ltd., Bangalore, India, were used for all the metagenomics analysis and V3-V4 metagenome library preparation. All the detailed processes have been mentioned in the supplementary file. Genomic DNA was amplified for 26 cycles (Round 1) using KAPA HiFi HotStart PCR Kit (KAPA Biosystems Inc., Boston, MA USA). Both forward and reverse primer concentrations were 5µM on 1.2% agarose gel bed. PCR amplicons from round 1 (1µl diluted volume) were used for the round 2 PCR indexing, amplified for 10 PCR cycles. Then the loadings were added to the Illumina sequencing barcoded adaptors (Nextera XT v2 Index Kit, Illumina, U.S.A.); another gel electrophoresis was done to normalize and pool the sequenced libraries.

Illumina Adapter Sequences (Majumdar et al., 2020b):

5' -AATGATACGGCGACCACCGAGATCTACAC [i5] TCGTCGGCAGCGTC

5' -CAAGCAGAAGACGGCATACGAGAT [i7] GTCTCGTGGGCTCGG

[i5, i7] –Unique dual index sequence to identify sample-specific sequencing data.

Samples were then loaded into an Illumina MiSeq v3 600 cycles cartridge (Illumina, CA, USA) and the run was performed according to standard Illumina protocol.

• Data output analysis for metagenomics reads

V3-V4 primer sequence from the Illumina paired-end raw reads were selected from the high quality bases. Further, the reads were joined using Fastq-join. These joined reads were used for the microbiome search using QIIME functions. In further step, query sequences were clustered

using UCLUST process against a trimmed chimera free 16SrRNA database (Greengenes v 13.8). RDP6 classifier was used to assign taxonomies with clustered at $\geq 97\%$ sequence similarities while comparing to the Greengene database which created a biome file. This biome file was further used for the advanced analysis and visualization. In the biome file, information was stored about the reads number assigned to specific taxa. Microbiome identification and the number of operational taxonomic units (OTUs) for each sample were analyzed using giime scripts. Calculation of relative abundance from phyla to species from read counts were assigned to OTUs and divided by the total considered reads for the microbiome search. Alpha and beta diversity for microbial species richness were calculated for each sample. Krona charts were made that are interactive html files consisting the phylogenetic information at each taxonomic level. Venn diagram and multivariate analysis were made to identify the difference of microbial diversity in different soil samples.







Supplementary Figure 1. Soil microbial phylum diversity under three field conditions.

Combined soil metagenomics results of soils from three plant growth phases have been represented here in form of special pie chart showing the majority of the microbial phyla encountered with high read numbers in the main pie chart and phyla with low read numbers are casted out as a secondary pie. Observed numbers of read are mentioned alongside of each phylum. Field soils have been cross-checked at two different seasons of pre-monsoon (**a**, **b**, **c**) and monsoon (**a1**, **b1**, **c1**) to compare any altered microbial community. A higher read was observed during the harvesting phase of rice cultivation.

Develop Milking Harvest	row min		row max					
	Alphaproteobacteria		[Fimbriimonadia]	• • •	0319-6E2	•	• •	SC72
	Acidobacteria-6		Clostridia	•••	Elusimicrobia	•	• •	Verruco-5
	Gammaproteobacteria		Sphingobacterija		SJA-176	•	•••	[Methylacidiphilae]
	Planctomycetia		PAUC37f		TM7-3	•	• •	SC3
	Deltaproteobacteria		Thaumarchaeota	• • •	C6	•	• •	SBRH58
	Anaerolineae		Ellin6529		WCHB1-64	•	• •	Mollicutes
	[Chloracidobacteria]		TK17		ML635J-21	•	•••	Spirochaetes
	Nitrospira		iii1-8		EC1113	•	• •	[Brachyspirae]
	Retanroteobacteria		TM7-1		AT-s54	•	• •	TA18
	[Sanrosnirae]		Synechococconhycideae		SHA-109		• •	Epsilonproteobacteria
	Phycisphaerae		Chthonomonadetes		SJA-4		•	OP8_1
	Solibacteres		Gemm-5		[Leptospirae]	•	• •	ABY1
	Cytophagia		OM190		Pla3	•	•	12-24
	[Spartobactoria]		SM2E11		PBS-25		• •	[Lentisphaeria]
	Pacilli		DDC102		Fibrobacteria		• •	GN15
	S005		Gitt GS 126		SIA-20		• •	BD1-5
	Nactacophysidaga		Gill-03-130	•••	DDD 11		• •	BB34
	Actinobactoria	•••	Gerillatarianhyaidaga	• • •	TM4		• •	3BR-5F
	Acidimicrobiio		Asidebasterija	• • •			• •	Fusobacteriia
	Comm 1		Acidobacterila	•••	VHS-85-50		• •	Endomicrobia
	Gemmeleenhilie		Rubrobacteria	•••	VadinHA49		• •	[Thermobacula]
	[Dedeephoeree]	••	4000-2	• • •	028H05-P-BN-P5		•	Dehalococcoidetes
	[Pedospriaerae]	•••	5035	•••	GKS2-1/4			BSV26
	PRR-12		SVa0/25	• • •	TG3		• •	NPL-UPA2
	Acidobacteria-5	•••	RB25	• • •	SAR202			Bacteroidia
•••	Pla4	••	Verrucomicrobiae	• •	Ignavibacteria			Armatimonadia
	Chlorotlexi	•••	DA052	• • •	Chlamydiia			OPB41
	MB-A2-108	• • •	BD7-11		[Rhodothermi]			Holophagae
	Gemmatimonadetes	• • •	ZB2		[Parvarchaea]			SHA-114
	Flavobacterila	••	Thermomicrobia		Thermoplasmata		• •	[Micrarchaea]
	Opitutae		P2-11E		Methanobacteria			Methanomicrobia
	TK10	• • •	OPB56		MCG			Unclassified

Supplementary Figure 2. Microbial class distribution in three soils with availability and dominance.



Supplementary Figure 3. Principal component analysis (PCA) (**a**) and principal coordinate analysis (PCoA) (**b**) in three different phase soil microbiomes.

Fraction	Solution	Applied Conditions
F1: Exchangeable	8 ml of MgCl ₂ (1M, pH=7.0)	1 hr. at Room temperature, continuous agitation
F2: Carbonates	8ml of NaOAc (1M, pH=5.0) w/ CH ₃ COOH	5 hrs. at Room temperature, continuous agitation
F3: Iron and Manganese Oxides	20ml of NH ₂ OH*HCl in 25% (v/v) HOAc	6 hrs. at 96°C, intermittent agitation
F4: Organic Matter	3ml of HNO ₃ (0.02M) 5ml of 30%H ₂ O ₂ (pH=2) w/HNO ₃	2 hrs. at 85°C, intermittent agitation
	3ml of 30%H ₂ O ₂ (pH=2) w/HNO ₃	3hrs. at 85°C, intermittent agitation
	5ml of NH4OAc(3.2M) in 20% (v/v)HNO3 diluted to 20mL with	30 mins. at Room temperature, continuous
	Milli-Q water	agnation
F5: Residual	$HF-HClO_4(5:1)$	-

Supplementary Table 1. Details of sequential extraction analysis method with reactant application dosages.

PARAMETER	ICPMS (Parkin-Elmer)
RF Power	4.2 Kw
Carrier Gas Flow Rate	20 ml/min
Argon Plasma Flow Rate	1.01 L/min
(Nebulizer cold Plasma) Sampling Depth	150 to 180
Collision Cell Gas	He (93%) + H (7%)
Collision Cell Gas Flow Rate Nebulizer Pump Rate	2-10 ml /min 0.5 rps (30 RPM)
Uptake Time	30 s
Wash Time	60- 70 s

Supplementary Table 2. Instrument specifications during ICP-MS analysis.

Parameters	WD XRF (Bruker S8 Tiger) Beryllium to Uranium		
Analysis range			
Sample form	Powder solid		
Concentration range	Concentrations from sub ppm to 100%		
Sample size	Less than 200mm mess size		
Excitation	End window 4kW Rh anode tube (60Kv, 170mA)		
Detector gas	P10 gas (10 % methane, 90 % argon)		
X-ray Detectors	Gas flow proportional counters and Scintillation counters		

Supplementary Table 3. Instrument specifications of Wavelength Dispersive XRF.

Field samples	Alpha diversity indices				
Pre-monsoon	Shannon	Simpson	Chao1		
Developing	9.39	0.946	2418.05		
Milking	9.18	0.906	2014.34		
Harvest	9.69	0.954	2862.10		
Monsoon	Shannon	Simpson	Chao1		
Developing	8.91	0.822	2314.24		
Milking	8.67	0.856	1936.78		
Harvest	8.95	0.871	2639.22		
	Beta diversity indices				
Pre-monsoon	Whittaker	Cody	Williams		
Developing:Milking	0.272	102.5	0.115		
Milking:Harvest	0.335	144.5	0.174		
Harvest:Developing	0.318	117	0.126		
Monsoon	Shannon	Simpson	Chao1		
Developing:Milking	0.214	94.25	0.093		
Milking:Harvest	0.299	108.61	0.095		
Harvest:Developing	0.296	102.97	0.098		

Supplementary Table 4. Average Alpha and Beta diversity indices of soil microbiome in three phase samples.