

## **The salmonid and the subsurface: Hillslope storage capacity determines the quality and distribution of fish habitat**

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# The salmonid and the subsurface: Hillslope storage capacity determines the quality and distribution of fish habitat

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**Abstract.** Water in rivers is delivered via the critical zone that mantles landscapes. Consequently, the success of stream-rearing salmonids depends on the structure and resulting water storage and release processes of this zone. Physical processes below the land surface (the subsurface component of the critical zone) ultimately determine how landscapes ‘filter’ climate to manifest ecologically significant stream flow and temperature regimes. Subsurface water storage capacity of the critical zone has emerged as a key hydrologic variable that integrates many of these subsurface processes, helping to explain flow regimes and terrestrial plant community composition. Here, we investigate how subsurface storage controls flow, temperature and energetic regimes that matter for salmonids. We illustrate the explanatory power of broadly applicable, storage-based frameworks across a lithological gradient that spans the Eel River watershed of California. Study sites are climatically similar but differ in their geologies and consequent subsurface critical zone structure that dictates water storage dynamics, leading to dramatically different hydrographs, temperature, and riparian regimes – with consequences for every aspect of salmonid life history. Lithological controls on the development of key subsurface critical zone properties like storage capacity suggest a heretofore unexplored link between salmonids and geology, adding to a rich literature that highlights various fluvial and geomorphic influences on salmonid diversity and distribution. Rapidly advancing methods for estimating and observing subsurface water storage dynamics at large scales present new opportunities for more

42 clearly identifying landscape features that constrain the distributions and abundances  
 43 of organisms, including salmonids, at watershed scales.

## 44 1. Introduction

45 Riverine biota, including salmonids, depend on multiple facets of streamflow. Flow  
 46 regime (the timing and magnitude of streamflow) determines the accessibility and  
 47 hydraulic features of habitat, and influences the timing of key life history events,  
 48 such as migration and spawning [e.g. Beechie et al., 2006, Sykes et al., 2009]. Stream  
 49 temperature and riparian light environment impact habitat suitability, fish metabolism,  
 50 prey productivity, and salmonid growth potential [e.g. Atlas et al., 2021]. Human-caused  
 51 changes in land use and climate have impacted riverine flows, temperatures, and riparian  
 52 characteristics, altering aquatic ecosystems globally [Lehner et al., 2011]. Proper  
 53 attribution of drivers of change, as well as the development of successful mitigation and  
 54 restoration strategies for aquatic ecosystems, require that we understand the physical  
 55 controls of these elements at watershed scales [e.g. Sturrock et al., 2019, Quinn et al.,  
 56 1997].

57 Although climate strongly influences light environment, temperature, and water  
 58 quantity and quality [e.g. Berghuijs et al., 2020, Maurer et al., 2021], a full understanding  
 59 of watershed function requires a critical zone (CZ) perspective, which integrates above-  
 60 ground processes (e.g., atmospheric fluxes, vegetation patterns, land use changes) with  
 61 subsurface dynamics (infiltration, rooting zone processes, weathering) and water storage  
 62 [e.g. Anderson et al., 2007, 2008, Brantley et al., 2007, Riebe et al., 2016, Grant  
 63 and Dietrich, 2017]. In the upland freshwaters that host many rearing and spawning  
 64 salmonid populations, water storage in the subsurface CZ occurs in both the shallow soil  
 65 layer (commonly < 0.5 m thick) and deeper underlying layers of saprolite and weathered  
 66 bedrock [Rempe and Dietrich, 2018, McCormick et al., 2021, Wald et al., 2013, Dawson  
 67 et al., 2020].

68 The dynamic water storage capacity (storage capacity) of the subsurface CZ has  
 69 emerged as an integrative catchment characteristic because of its ability to explain  
 70 flowpaths, flow generation, and plant-community composition [Hahm et al., 2019b, Illien  
 71 et al., 2021, Sayama et al., 2011, McDonnell et al., 2018, Klos et al., 2018]. A watershed's  
 72 dynamic storage has been defined in various ways [e.g. Staudinger et al., 2017, Dralle  
 73 et al., 2018, Buttle, 2016], but we here focus on a simple definition of the term: Dynamic  
 74 storage ( $\Delta S$ ) is the change in volume of water stored in a catchment *relative* to some  
 75 reference storage state [Hahm et al., 2019b] as inferred through mass balance:

$$76 \quad \Delta S = P - Q - ET, \quad (1)$$

77 where P, Q, and ET are precipitation, stream discharge, and evapotranspiration,  
 78 respectively. In hilly and mountainous landscapes underlain by bedrock, water storage  
 79 capacity in the subsurface is set by the depth of chemical and physical weathering fronts

80 that alter fresh bedrock and generate porosity capable of retaining and releasing water  
81 [e.g. Klos et al., 2018, Pedrazas et al., 2021, Callahan et al., 2020, Holbrook et al., 2014].  
82 These weathering processes, and therefore water storage capacity, depend on tectonics,  
83 climate, biota, and, importantly, the underlying bedrock geology [e.g. Riebe et al.,  
84 2016]. Exactly determining the water storage capacity of a watershed is intractable,  
85 but it can be roughly estimated by simply calculating the maximum observed value  
86 of dynamic storage [Hahm et al., 2019b]. Storage capacity in the subsurface sets the  
87 maximum volume of water that can be stored for later use by vegetation, which itself  
88 interacts with climate and subsurface storage to determine the timing and magnitude of  
89 groundwater recharge, and thus runoff generation and river flow regime features [Klos  
90 et al., 2018, Dralle et al., 2018].

91 Here, we propose that subsurface water storage capacity can explain between-  
92 catchment differences in stream hydrologic and energetic features that matter for  
93 salmonid life history. Importantly, this is not a difference in larger-scale or regional  
94 aquifers tied to intrinsic properties of fresh bedrock. Rather, the variability occurs  
95 at the hillslope scale and is tied to the weathering-driven development of the critical  
96 zone, which depends on material properties of the bedrock. We use California’s Eel  
97 River—designated by the California Department of Fish and Wildlife and State Water  
98 Resources Control Board as a priority salmonid conservation watershed, and one of  
99 the few mostly undammed rivers on the US Pacific Coast—as a case study. First,  
100 we synthesize results from a decade-plus effort of subsurface monitoring enabled by  
101 deep drilling [Salve et al., 2012, Hahm et al., 2020, Schmidt and Rempe, 2020, Tune  
102 et al., 2020, Rempe and Dietrich, 2018] showing how subsurface structure and rock  
103 type explain variations in water storage capacity and plant community composition at  
104 two intensively monitored Eel River subcatchments. These sites are underlain by two  
105 different bedrock lithologies (the Coastal Belt turbidites and Central Belt melange of the  
106 Franciscan Complex) that have weathered differently (deep and shallow, respectively)  
107 owing to different rock properties [Hahm et al., 2019b]. We explore how lithologically-  
108 determined storage capacity drives differences in functional features of the flow regime  
109 [Yarnell et al., 2020] and energy, and the dynamics and climatic sensitivity of stream  
110 temperature between the two subcatchments. We then explore the consequences of  
111 these differences for stream rearing salmonids – at specific life stages, and holistically  
112 as life history syndromes. Further, we demonstrate in sparsely monitored catchments  
113 that estimates of storage capacity explain key metrics of flow behavior, and, across  
114 larger watershed scales, storage capacity reflects the contribution from subcatchments  
115 of mixed geological composition. Our findings indicate that lithologically controlled  
116 storage capacity has widespread impacts on the spatial distribution of habitat quantity  
117 and quality, factors that influence the diversity of salmonid life histories.

## 118 2. Geology and subsurface structure

119 The study area encompasses multiple subwatersheds (Table 3 and Figure 2) of the Eel  
120 River in the Northern California Coast Ranges. The regional climate is Mediterranean-  
121 type with warm, dry summers and cool, wet winters. The Eel River basin is underlain  
122 by the Franciscan Complex, a geological assemblage in Northern California consisting  
123 of three north-south running belts of different rock type that decrease in age and  
124 metamorphic grade from east to west [Blake Jr and Jones, 1974, McLaughlin et al.,  
125 1994].

126 Two intensively monitored watersheds within the Eel River basin, Elder Creek  
127 and Dry Creek, serve as representative end members of two (out of three) belts  
128 of the Franciscan—the Coastal Belt (Elder Creek) and Central Belt melange (Dry  
129 Creek). Figure 1 illustrates lithologically-determined contrasts in hillslope subsurface  
130 structure—and thus water storage capacity—in the two watersheds. We here provide a  
131 brief overview of the subsurface structures and water storage dynamics at the sites. For  
132 more details, we refer the reader to Hahm et al. [2019b].

133 The Elder Creek watershed is underlain by the fractured shale and sandstones  
134 of the Coastal Belt (Figure 1 right column). Deep weathering profiles in the Coastal  
135 Belt have resulted in large water storage capacity of the subsurface, most of which  
136 is unsaturated storage [Dralle et al., 2018] in a thick vadose zone that includes soil,  
137 saprolite, and weathered bedrock. This unsaturated reservoir can hold upwards of 300  
138 mm of seasonally dynamic water storage, equal to more than 1/3 of annual wet season  
139 precipitation during dry years [Rempe and Dietrich, 2018]. The large dynamic storage in  
140 the vadose zone is the primary water source for the productive, dense conifer-hardwood  
141 evergreen forests found in the Coastal Belt. Following the long dry season, tree-driven  
142 storage deficits (the amount of water input to the root zone required to replenish that  
143 which vegetation removed) in the unsaturated zone are typically replenished within the  
144 first few months of the wet season (Oct. to Dec.), whereupon the soil and weathered  
145 bedrock layers wet to a characteristic maximum storage. Analogously to ‘field capacity’  
146 in soils [Grindley, 1968], additional rainfall beyond this characteristic maximum value  
147 triggers gravitational drainage that recharges an underlying fractured-rock hillslope  
148 groundwater system, which flows laterally down to streams through a system of seeps  
149 and springs [Salve et al., 2012, Lovill et al., 2018a]. This deeper saturated reservoir  
150 can store upwards of 200 mm of groundwater that slowly drains to adjacent streams,  
151 supporting year-round cold baseflows.

152 Dry Creek is underlain by the Central Belt melange geology (Figure 1 left column),  
153 a chaotic mixture of bedrock of varying lithology and size suspended in a shale-derived,  
154 clay-rich matrix that is perennially saturated at depths typically not exceeding 2 to 3 m  
155 below the ground surface. The thin weathering zone of the Central Belt completely fills  
156 after approximately 200 mm of wet season rainfall, at which point the groundwater table  
157 rises to the surface, generating widespread saturation overland flow that is rapidly routed  
158 to dense drainage networks. Consequently, the Central Belt watersheds are unable to

159 store large volumes of wet season precipitation, resulting in fast draining hillslopes,  
160 and streams that cease flowing within the first couple months of the dry season [Lovill  
161 et al., 2018a]. Low storage also results in a more water-stress tolerant savanna plant  
162 community comprised primarily of Oregon white oak (*Quercus garryana*) and annual  
163 grasses [Hahm et al., 2018].

### 164 3. Water

165 Using hydrological and climatic data from the Dry and Elder Creek end members, we  
166 explore storage-capacity controls on flow regime features that matter for fish: timing of  
167 wet season flow activation (I), peakflow magnitude (II), flow recession rate (III), and low  
168 flow magnitude (IV) [Yarnell et al., 2020]. Figure 4 provides an overview of contrasting  
169 flow regime features in the end-member Coastal and Central Belt watersheds. Center  
170 panels plot discharge on linear (top-central panel) and log (middle-central panel) scales,  
171 as well as cumulative discharge and runoff, for the 2019 water year. The paneled subplots  
172 aim to highlight the major functional flow components I - IV. Importantly, the sites' 20  
173 km separation results in nearly identical rainstorm magnitudes through the winter.

174 Conceptual figures illustrate many of these outcomes. A four-quadrant hillslope  
175 diagram (Figures 1) depicts representative hydrology for the two end-member geologies  
176 in both the wet and dry seasons, and a four-quadrant stream diagram (Figures 3) depicts  
177 the typical trajectory of stream conditions from the spring/early summer flow recession  
178 to the late-summer low flow period.

#### 179 3.1. Wet season flow activation

180 Differences in vegetation cover between Elder and Dry (left versus right column in  
181 Figure 1) result in different magnitudes of plant water use, and thus differences in storage  
182 deficits in the root zone at the end of the dry season [Dralle et al., 2018, McCormick et al.,  
183 2021, Wang-Erlandsson et al., 2016]. Replenishment of these deficits via infiltration of  
184 rainfall and filling of the critical zone at the start of the wet season mediates drainage  
185 from the root zone, thereby determining the amount of rainfall (more at Elder) required  
186 to recharge the hillslope aquifers that drive streamflow production (either via subsurface  
187 flow or groundwater-driven saturation overland flow) at the hillslope-channel boundary  
188 [Müller et al., 2014, Dralle et al., 2018, Lapidés et al., 2021, Grindley, 1968].

189 To quantify storage controls on wet-season flow activation, we turn to a storage-  
190 activation approach introduced by Sayama et al. [2011], wherein early wet season  
191 discharge is plotted as a function of cumulative seasonal dynamic watershed storage  
192 to identify storage thresholds that lead to rapid increases in discharge. Figure 4I plots  
193 stream discharge against a catchment water storage approximation, calculated as the  
194 running sum of input (precipitation, P) minus output (discharge, Q) fluxes beginning on  
195 October 1, under the assumption that differences between these two dominant fluxes can  
196 be attributed to accumulation of storage in the watershed (approximation of Equation 1,

207  $\Delta S = P - Q - ET \approx P - Q$ ). Evapotranspiration is neglected because it is expected to  
208 be relatively small in October - January when the sum is calculated. Flow activation in  
209 Dry Creek begins around approximately 150 mm of cumulative rainfall at the start of the  
210 dry season, whereas Elder Creek flow does not activate until approximately 300 mm of  
211 rain has fallen. Moreover, Elder Creek discharge sensitivity to storage is much lower, as  
212 can be seen by the relatively muted increase in discharge with storage increases beyond  
213 300 mm. The storage-discharge relationship in Dry Creek is more nonlinear ('flashier'),  
214 with a very steep increase in flow rate beyond 200 mm of storage.

### 215 3.2. Peakflow magnitude

216 The relationship between storage and discharge also explains the difference in peak flow  
217 magnitudes in Figure 4II, where the smaller storage capacity at Dry Creek results in  
218 more extreme peak flows. Very small changes in storage generated by the addition  
219 of precipitation result in rapid, highly nonlinear increases in flow at Dry Creek, and  
220 correspondingly large peak flows, which contrast Elder Creek's muted peak flow response  
221 during rainfall events. These flow behaviors can be attributed to water storage dynamics  
222 in the hillslope, where complete filling of the critical zone in the Central Belt results  
223 in flashy streams fed predominantly by saturation overland flow, as compared to the  
224 muted groundwater-dominated signal in Elder Creek, as illustrated in the top row of  
225 Figure 1.

### 226 3.3. Rate of recession and low (base)flow magnitude

227 Whereas Figure 4I and II show the effect of storage capacity on rising limb and  
228 peakflow behaviors, Figure 4III and IV demonstrate that storage capacity is also a  
229 strong determinant of the drainage behavior of the study catchments. At Dry Creek,  
230 small storage capacity drives water to the surface, where shallow and overland flowpaths  
231 result in fast flow recessions and very little retention of drainable storage. In the Elder  
232 Creek watershed, the drainage of deeper fractured rock hillslope groundwater results in  
233 a much slower recession and high retention of drainable groundwater storage going into  
234 the dry season. The consequences of these recession dynamics are illustrated in the top  
235 row of Figure 3. The rapid drop in flows in Central Belt catchments results in relatively  
236 mild flow conditions during a short period in the spring (see much of April and May in  
237 Figure 4), which contrasts the persistently higher and more turbulent flows during the  
238 early dry season in Coastal Belt watersheds.

239 Over longer periods of drainage—California's protracted dry season can persist for  
240 more than six months of the year—storage capacity may dictate whether a stream has  
241 any water at all. In our study catchments, two distinct dry season flow regimes emerge:  
242 high storage capacity Elder Creek supports robust baseflows that persist through the  
243 dry season, whereas ephemeral flows in Dry Creek result in dry streambeds typically  
244 before July (Figure 3).

235 Interestingly, although there is lower *dry season* baseflow in the Dry Creek  
236 watershed, over the course of an entire year, Dry Creek typically generates more total  
237 runoff for a given precipitation event, as quantified by each watershed’s runoff ratio in  
238 4V. From water year 2016 through water year 2020, the average runoff ratio in Elder  
239 Creek is approximately 0.6, compared to an average runoff ratio of approximately 0.8 at  
240 Dry Creek. The small storage capacity at Dry Creek generates enough runoff during the  
241 wet season to overwhelm its small dry season runoff totals, thus producing an overall  
242 higher runoff ratio. We attribute this difference primarily to the significant amounts of  
243 water stored in the thick weathered bedrock vadose zone at Elder Creek, which results  
244 in more water being returned to the atmosphere via transpiration during the growing  
245 season.

#### 246 3.4. Spatial and temporal variability in water availability

247 In addition to impacting average flow regime features ‘at a station’, storage capacity can  
248 also mediate the spatial availability of water via the wetted channel network, and the  
249 sensitivity of runoff dynamics to year-to-year swings in rainfall totals. For example,  
250 Lovill et al. [2018a] demonstrate significant differences in wetted channel drainage  
251 density between the Elder Creek and Dry Creek watersheds. Figure 5 reproduces results  
252 from Lovill et al. [2018a], plotting a snapshot of wetted channel extent in late August  
253 2014. On this date, Elder Creek wetted channel drainage density is nearly 10 times  
254 greater than that observed in Dry Creek, despite nearly identical rainfall totals in the  
255 preceding wet season.

256 Regarding temporal variability in water availability, Hahm et al. [2019a] introduce  
257 the idea that *small* subsurface water storage capacity *relative* to annual average rainfall  
258 has the potential to decouple initial storage going into the summer (which sets plant  
259 water availability during the dry growing season) from total annual rainfall. This  
260 mechanism was dubbed ‘storage-capacity limitation’, wherein the limited subsurface  
261 water storage capacity fills completely in both wet and dry years, resulting in a  
262 hydrological mechanism of drought resilience. Although Hahm et al. [2019a] explore  
263 this concept with respect to vegetation response to annual rainfall variability, it extends  
264 to components of hydrograph variability as well. Specifically, since storage-controlled  
265 flow conditions going into the dry season control summer low flows [Dralle et al., 2016],  
266 if storage is decoupled from *total* rainfall, so too will be low flows. In both Elder Creek  
267 and Dry Creek, we find that dry season low flows are essentially decoupled from annual  
268 swings in total rainfall (both are storage-capacity limited). This is because storage  
269 capacity in both watersheds is significantly lower than average annual precipitation,  
270 and therefore annual variations in total rainfall do not lead to annual variations in  
271 dynamic storage going into the dry season, resulting in a decoupling of low flows from  
272 total rainfall. In the case of Dry Creek, this decoupling result is trivial; low flows are  
273 reliably zero in all years, and therefore total annual rainfall does not dictate low flow  
274 magnitudes. At Elder Creek, Figure 8a illustrates how dry season initial flow conditions



275 and dry season duration are likely more important drivers of low flows (not total annual  
 276 rainfall). Figure 8b quantifies the decoupling between low flow and annual rainfall,  
 277 demonstrating that low flow magnitude (over 21 years, 2001 to 2021) does not vary  
 278 strongly with total rainfall. Consistent with the storage-capacity limitation mechanism  
 279 hypothesized by Hahm et al. [2019a], a linear regression reveals no significant slope on  
 280 the low flow vs. water year precipitation relationship, with an  $R^2$  of 0.14, indicating  
 281 that water year precipitation explains little of the variation in summer low flows.

### 282 3.5. Hydrological scaling in mixed-lithology watersheds

283 Large watersheds can be conceived of as a collection of hillslopes connected through a  
 284 shared channel network [Harman et al., 2009]. Using this “unit hillslope” concept, and  
 285 under the assumption that a map of Coastal and Central Belt geologies may serve as  
 286 a proxy for hillslope storage capacity across the Eel River, we hypothesize that “mixed  
 287 lithology” watersheds will behave, hydrologically, like a superposition of the Elder  
 288 Creek and Dry Creek geological end members. To test this hypothesis, we identified 5  
 289 subwatersheds (Table 3) of the Eel River with contributing areas less than 1000 km<sup>2</sup>  
 290 that span a gradient in percent melange composition, where Dry Creek serves as the  
 291 100% melange watershed, and Elder Creek serves as the 0% melange watershed. We  
 292 explore scaling of flow recession and dry season water availability across this geological  
 293 spectrum of sites.

294 We perform two analyses to explore drivers of the scaling of dry season flow recession  
 295 and water availability in the watersheds listed in Table 3. First, we calculated a  
 296 dimensionless metric of summer water availability, the summer runoff fraction, which is  
 297 calculated as the long term average of the annual ratio of total flow from June through  
 298 September (summer dry season) divided by total flow from the previous October through  
 299 May (preceding wet season). Second, we calculated the average flow recession rate using  
 300 the widely used exponential flow recession model:

$$\frac{dQ}{dt} = -\frac{1}{\tau}Q \quad \rightarrow \quad Q = Q_0 e^{-t/\tau},$$

301 where  $Q_0$  is the initial flow value at the start of the recession. The recession timescale,  
 302  $\tau$  (with units of time), is a scale-independent measure of the rate of recession and is  
 303 therefore directly comparable between watersheds of different sizes and with different  
 304 average flow values.

305 Watersheds with greater percent Central Belt mélangé contributing area have less  
 306 summer runoff relative to annual total runoff (Figure 6c). At Elder Creek (0% mélangé),  
 307 we see an average 3.75% of annual runoff discharges during the summer months. There  
 308 is a monotonic decrease in summer runoff fraction, with effectively 0% of annual runoff  
 309 occurring during the summer months in Dry Creek (100% mélangé).

310 Figure 6a,b show why summer runoff fractions are so small in mélangé-dominated  
 311 watersheds: typical flow recession rates, as quantified by the linear recession timescale  
 312 ( $\tau$ ) decrease nearly ten-fold across the geological gradient, from approximately  $\tau = 10$

313 days at Elder Creek, to  $\tau = 1$  day at Dry Creek. Actual recession data in Figure 6a  
314 demonstrates how smaller recession  $\tau$  results in a more rapid drop in flows in watersheds  
315 with higher mélange fraction.

## 316 4. Energy

317 Different subsurface water storage capacities support distinct vegetation types: dense,  
318 shady forest at Elder Creek, and at Dry Creek a deciduous oak savanna admitting more  
319 solar radiation (Figure 3). Differences in runoff pathways and flow volumes between  
320 the Elder Creek and Dry Creek watersheds should also affect stream temperature and  
321 its sensitivity to changes in atmospheric conditions. Deeper groundwater flowpaths in  
322 Coastal Belt geologies (Figure 1) are more buffered from variations in air temperature.  
323 The thermal inertia of a stream can also be expected to vary with flow volumes; all  
324 else equal, stream temperature will be more responsive to changing air temperature at  
325 smaller flow volumes. For all these reasons, stream temperature in watersheds with  
326 lower storage capacity, like Dry Creek, should track changes in air temperature more  
327 closely.

### 328 4.1. Channel shading

329 Large differences in canopy cover between the two geologies can be seen in Figure 7a.  
330 These storage-driven differences in vegetation community and canopy cover affect the  
331 riparian light environment. To quantify this, we calculated a simplified light penetration  
332 index (LPI) metric [Bode et al., 2014] as the number of LiDAR point returns that  
333 strike either ground or water divided by the total number of LiDAR returns at a 1 m  
334 resolution across all available LiDAR datasets in the Eel River [Power, 2013, Dietrich,  
335 2014, Roering, 2006, Perkins, 2009, Dietrich, 2015]. The LPI estimates the fraction  
336 of shortwave radiation that penetrates the vegetation canopy to reach the ground  
337 or water surface. Using mapped LPI, we computed reach-averaged canopy closure  
338 (Figure 7b) throughout the Eel River basin, illustrating a significant shift in stream  
339 shading across the Coastal and Central Belt lithologic contact. We then explored how  
340 this shift in riparian cover impacts the growing season stream energy budget by first  
341 extracting total summer shortwave solar radiation delivery (June through August) the  
342 midpoint (latitude 39.64, longitude -123.53) of the centroids of the Elder and Dry Creek  
343 watersheds [Thornton et al., 2020], then multiplying this figure by reach-averaged LPI  
344 to arrive at an estimate of total shortwave solar energy delivery to streams over the  
345 peak growing season months.

346 Figure 7c plots the reach-scale, LPI-filtered solar shortwave radiation energy  
347 delivery across a range of watershed contributing areas for the two lithologies. At small  
348 contributing areas, solar energy delivery to the channel is nearly three times greater in  
349 Central Belt versus Coastal Belt watersheds. However, as contributing area increases,  
350 channel widths tend to increase, leading to an overall increase in solar radiation delivery

351 and a convergence between the two lithologies: streamside vegetation shading matters  
352 less in wide channels.

#### 353 4.2. Stream temperature

354 Figure 7d plots air temperature and stream temperature at the end-member sites, clearly  
355 demonstrating that Dry Creek water temperatures fluctuate more widely than those in  
356 Elder Creek despite nearly identical air temperatures between the two sites. Figure  
357 7e removes the time element to reveal the relationship between water temperature and  
358 air temperature. Best fit lines between Elder Creek and Dry Creek show that water  
359 temperature is generally buffered relative to air temperature as expected from its higher  
360 heat capacity (slope less than 1) but that Dry Creek's best fit line slope is nearly double  
361 that of Elder Creek, indicating that Dry Creek water temperature is much more sensitive  
362 to air temperature changes.

### 363 5. Discussion

#### 364 5.1. Salmonids and the Subsurface - Life history framework and hypotheses

365 Properties of the subsurface critical zone have consequences for salmonids across every  
366 stage of their life history. Environmental regimes determined by a watershed's storage  
367 capacity in turn constrain (Table 2) opportunities for salmon, influencing access for  
368 migrating and spawning adults, the survival of eggs, and the rearing habitats and  
369 movements of juveniles during their freshwater residence. Below we delve deeper into  
370 these connections and consider how watershed storage capacity influences core events  
371 in the life history of salmonids, using Elder Creek and Dry Creek as representative  
372 ecological end-members. In particular, we emphasize that contrasting conditions that  
373 favor different life history traits. Further, we suggest that the watershed behaviors of  
374 Elder and Dry Creek, driven by their distinct subsurface storage capacities, could select  
375 for life history syndromes, that is, suites of correlated traits (e.g. spawn timing, growth  
376 and foraging, and movement decisions), associated with different degrees of dependence  
377 on rearing in natal habitat (i.e., habitat in which they emerged) versus rearing outside  
378 their natal habitat (i.e., non-natal rearing). Life history variation among populations in  
379 contrasting CZ environments could reflect natural selection (i.e., life history adaptation)  
380 or plasticity. Regardless, we posit the habitat conditions in different CZ environments  
381 tend to favor expression of different (and specific) life histories.

382 5.1.1. Adult migration and spawning Migration and spawning require suitable depth of  
383 flow for passage between the ocean and riverine spawning sites and suitable hydraulics  
384 for spawning once fish arrive at their spawning destination. The site-specific lag time  
385 between a rainfall event and the hydraulic response in streams determines how quickly  
386 migrating and spawning salmon can respond to precipitation. Subsurface storage deficits  
387 at the end of the dry season dictate how much precipitation is needed to elevate winter

388 stream flows in the early wet season [Rempe and Dietrich, 2018, Dralle et al., 2018]. For  
389 example, when root-zone storage is fully depleted, flow activation requires twice as much  
390 precipitation in Elder Creek as in Dry Creek. Thus, in low-storage capacity watersheds  
391 like Dry Creek, less precipitation is needed to provide suitable flows for migration and  
392 spawning, potentially allowing spawning to occur earlier in these systems. In higher  
393 storage capacity systems like Elder Creek, a later wet season flow activation, but more  
394 prolonged flow recession, suggests adults may arrive and spawn later – but enjoy a longer  
395 duration of suitable spawning conditions. The difference in flow activation between Elder  
396 and Dry Creek may vary from hours to weeks depending on their respective storage  
397 capacity deficits and initial wet season rainfall patterns. Variation in subsurface storage  
398 capacity, therefore, is likely to generate differences in access and spawn timing during  
399 many years, similar to the influence of stream temperature in influencing diversity in  
400 this trait among populations [e.g. Lisi et al., 2013].

401 *5.1.2. Egg incubation* Female salmon build their redds (nests) in the gravel of  
402 streambeds. The shape of the redd, the location, and the size of the gravels all differ  
403 among species. The developmental rate of eggs also differs among species and is strongly  
404 influenced by water temperature [From and Rasmussen, 1991]. Temperature variation  
405 among streams driven by subsurface properties (see Figure 7), and contributions to  
406 discharge from ground water, likely have consequences for egg development rates and fry  
407 emergence timing. Riparian vegetation composition and flow volume (thermal mass)–  
408 both controlled by features of the Critical Zone–affect the sensitivity of stream water  
409 temperature to air temperature and solar radiation (Figure 7). Ephemeral Dry Creek,  
410 with its minimal channel shading, warms (and cools) more rapidly than heavily shaded  
411 Coastal Belt streams with sustained contributions from groundwater, like Elder Creek.  
412 Eggs incubating in redds in the warmer waters of intermittent streams are likely to  
413 develop more rapidly, leading to earlier alevin emergence.

414 In addition to the effect of temperature on egg emergence, salmon redds are at risk  
415 of scour and dessication depending, in part, on the flashiness of stream hydrology [Orth  
416 et al., 2005]. Rapid declines in streamflow can potentially cause redds to dry before  
417 alevin are able to emerge. Conversely higher peak flows lead to greater risk of redd  
418 scour. In short, in flashy melange streams like Dry Creek, redds are likely also more at  
419 risk from scour and desiccation than in more stable streams with lower peak flows like  
420 Elder Creek (Figure 3).

421 *5.1.3. Juvenile growth and summer survival* Once salmon fry have emerged from the  
422 gravels and begin exogenous feeding, differences in CZ structure have implications for  
423 the prey production and growth of fish during their early life stages. During the spring  
424 months (March-May), streamflow recession coincides with increasing photo-period and  
425 primary and secondary productivity in salmon-bearing food webs of coastal streams  
426 like Elder and Dry Creek [Rossi et al., 2022]. However, the relative timing and rate  
427 of streamflow recession, water temperature warming, and food web phenology all vary

428 between stream types (Figure 6), driving different seasonal patterns in growth potential  
429 for rearing salmonids [Rossi et al., 2022]. All else being equal, streams fed by Critical  
430 Zones with low storage capacity will support an earlier spring increase in juvenile  
431 salmonid growth potential, and an earlier decrease in summer growth potential; whereas  
432 in perennial streams fed by Critical Zones with high storage, growth potential for juvenile  
433 salmonids will peak later and be sustained longer [Rossi et al., 2022].

434 While growth potential may peak earlier in intermittent streams with low storage  
435 potential, these systems also experience an earlier onset of inhospitable conditions  
436 for summer rearing salmonids. With warming and hypoxia, and eventual drying and  
437 disconnection of the wetted channel network, fish that remain in the stream (as opposed  
438 to outmigrating, see next section) can perish [Rossi et al., 2022, Labbe and Fausch, 2000].  
439 Importantly, the magnitude of summer mortality varies considerably among years due  
440 to interannual variation in rainfall patterns [Hwan et al., 2018, Obedzinski et al., 2018].  
441 Recent work by Vander Vorste et al. [2020], however, highlights that some intermittent  
442 systems provide more reliable habitats for juvenile coho rearing. For example, across  
443 the geologically complex Russian River watershed in Sonoma County, California, inter-  
444 annual variation in summer survival was high at some sites, but much more stable  
445 at other sites, hinting at the importance of CZ properties in influencing sensitivity  
446 of different systems to inter-annual variation in rainfall and consequences for salmon  
447 survival (e.g. Figure 6; see also Moidu et al. [2021]).

448 *5.1.4. Life history syndromes* A particularly diverse component of the many life  
449 histories of anadromous salmonids [Shapovalov and Taft, 1954, Hodge et al., 2016] is  
450 residence time of juveniles in their natal streams and their corresponding dependence,  
451 or lack thereof, on rearing outside their natal habitats. Here we suggest that watersheds  
452 with different and distinct subsurface storage capacities will favor the emergence of life  
453 history syndromes, that is, suites of correlated traits associated with different degrees  
454 of dependence on natal and non-natal rearing.

455 In non-perennial systems such as Dry Creek, faster recession that lowers summer  
456 base flow should favor a “grow fast and outmigrate early in life” strategy [e.g. Erman and  
457 Leidy, 1975], while earlier flow activation will allow adults to access and spawn sooner  
458 than in perennial streams (Figure 3). Warmer water would also directly accelerate egg  
459 development in non-perennial streams (Figure 7), so both earlier spawn timing and  
460 faster incubation should result in earlier emergence of juveniles. Earlier increases in  
461 food availability in sunlit channels support an earlier spring peak in growth potential in  
462 intermittent versus perennial streams [Rossi et al., 2022, Ebersole et al., 2006]. When  
463 the wetted channel dries completely or conditions become lethal, however, outmigration  
464 before the stream dries is the only option for survival. Erman and Leidy [1975] reported  
465 that large numbers of *O. mykiss* fry outmigrated from an intermittent stream prior to  
466 stream drying, suggesting that these systems contribute to diversity of outmigration  
467 timing. In the second year of their study, with more precipitation and perennial  
468 summer flows, many juveniles over-summered in the tributary, highlighting the influence

469 of interannual variation in precipitation on life history expression. Over-summering  
470 salmonids have also been regularly observed in intermittent streams with remnant  
471 pools with adequate water quality that persists through the summer [e.g. Obedzinski  
472 et al., 2018, Hwan et al., 2018, Woelfle-Erskine et al., 2017, Grantham et al., 2012,  
473 Vander Vorste et al., 2020]. However, wetted habitat area in such channels can be  
474 extremely limited, particularly habitats where older (i.e., age 1+ and 2+) fish can  
475 rear. In short, these streams can produce large numbers of outmigrating juveniles, but  
476 the success of this strategy relies on non-natal growth opportunities elsewhere in the  
477 watershed.

478 In contrast, in perennial streams with year-round flow, more critical zone storage  
479 of precipitation delays runoff pulses that allow adults to access spawning locations,  
480 delaying spawn timing (Figure 4), while cooling groundwater inputs during the spring  
481 as eggs incubate (Figure 7) slow development of incubating eggs, delaying emergence  
482 relative to the timing of these events in non-perennial streams. However, the slower rate  
483 of flow recession in spring and higher mean summer base flows can support fish that  
484 over-summer in the stream and rear for at least one year before outmigration [Kelson  
485 and Carlson, 2019]. Secondary production increases later than in intermittent streams,  
486 which along with sustained recessions and greater channel shading, leads to a later  
487 peak in growth potential [Rossi, 2020]. Perennial flow creates a greater extent of wetted  
488 channel and sustained summer rearing environment (both space and water quality), and  
489 reduces summer mortality relative to intermittent streams.

## 490 5.2. Storage capacity's influence on stream energetics

491 We identified three mechanisms by which hillslope storage dynamics could impact  
492 stream temperature and light environment. First, storage-controlled plant community  
493 composition has consequences for stream shading. Second, flow volumes impact the  
494 thermal inertia of water in the channel; all else being equal, lower flows in low storage  
495 Dry Creek result in higher in-channel sensitivity of water temperature to radiation fluxes  
496 and air temperature [Webb et al., 2008]. Finally, storage dictates flowpaths to streams,  
497 and because near-surface versus deep flowpaths will have different sensitivities to air  
498 temperature, this ultimately will impact the temperature of groundwater and water  
499 delivery to channels [Kurylyk et al., 2015]. We found water temperature dynamics were  
500 consistent with all three of these mechanisms; specifically, the low storage Dry Creek  
501 catchment has flow temperatures that are both hotter and more sensitive to climate than  
502 the high storage capacity Elder Creek catchment during the warm summer months.  
503 Although we did not determine the relative strengths of these three mechanisms, we  
504 did demonstrate potential scale-dependence in their impacts. At large scales, channels  
505 are wide and therefore hillslope plant communities have less impact on shading. At  
506 small scales (headwaters), water has not resided in channel for long and in-stream  
507 temperatures may be more representative of water temperatures being delivered to  
508 the channel by the hillslope [Dugdale et al., 2017]. Although there have been exciting

509 advances toward incorporating the impacts of flowpaths and hillslope processes in stream  
510 temperature prediction [Leach and Moore, 2015, Hrachowitz et al., 2010], most efforts  
511 focus on climate factors or heat exchange at the stream surface or with channel substrate  
512 [Brown, 1969]. Increased focus on hillslope processes will be especially important for  
513 understanding the fate of headwater refugia during low flow periods [Isaak et al., 2016,  
514 Leach and Moore, 2017], where prediction of water temperature sensitivities to climate  
515 are highly dependent hillslope processes [e.g. groundwater flow Leach and Moore, 2019]  
516 and properties [e.g. depth to bedrock Briggs et al., 2018].

### 517 5.3. Measuring and modeling storage capacity

518 Storage capacity has predictive potential for ecology, yet is difficult to measure or  
519 estimate; the *in situ* methods deployed at Dry Creek and Elder Creek cannot be  
520 realistically deployed at larger scales relevant to managers. Geological maps can be  
521 used to extrapolate behaviors over larger scales (under the assumption that rock-type is  
522 the primary driver of hillslope weathering profiles), but inferences need to be grounded  
523 with data from intensively monitored sites. Alternative approaches to analyzing storage  
524 in environments where data are sparse have emerged in recent years . Where discharge  
525 data are available, storage-discharge methods and models, or dynamic storage tracking,  
526 can provide important insights into subsurface storage processes and their controls on  
527 hydrology [Sayama et al., 2011] and salmonid-relevant flow metrics [Soulsby et al.,  
528 2016]. Satellite remote sensing methods have emerged as a scaleable approach for  
529 monitoring plant-driven storage deficits [Wang-Erlandsson et al., 2016], which control  
530 flow activation. Maximum observed storage deficits have been correlated with storage  
531 capacity as well [McCormick et al., 2021, Stocker et al., 2021]. Geomorphological,  
532 ecological, and hydrological model inversion and inferential methods may also provide  
533 some insights into the thickness of weathering profiles and water storage capacity  
534 [Pelletier et al., 2016, Ichii et al., 2009, Kleidon, 2004, Schenk, 2008]. Finally, geophysical  
535 methods, such as seismic refraction, have shown promise for understanding ecologically  
536 important hillslope-scale storage dynamics with significantly less effort than invasive  
537 methods [Holbrook et al., 2014, Briggs et al., 2018].

### 538 5.4. The evolution of hillslopes and salmon

539 Geologic history [e.g. Waples et al., 2008], landscape evolution [e.g. Montgomery, 2000]  
540 and channel network dynamics [e.g. Stokes and Perron, 2020, Val et al., 2022], and,  
541 on the shorter time scales, erosional and flood dynamics [Waples et al., 2008], all  
542 influence salmon diversity and resilience in the Anthropocene. Here we add another  
543 geomorphic component: the critical zone. The evolution of *subsurface* hillslope critical  
544 zone properties—including subsurface water storage capacity—depend on complex  
545 interactions between hydrology, weathering, erosion and tectonics [Riebe et al., 2016].  
546 We highlighted the dependence of storage capacity on rock properties, but other work  
547 demonstrates how subsurface architecture may also be influenced by frost cracking,

548 regional tectonic stresses, and groundwater geochemistry [Riebe et al., 2016, Rempe  
549 and Dietrich, 2014, St. Clair et al., 2015, Brantley and Lebedeva, 2011, Anderson et al.,  
550 2013, Harman and Kim, 2019]. Watershed hydrologic behavior arises from the collected  
551 dynamics of individual hillslopes, whose subsurface capacity to transmit flow downslope  
552 dictates the spatial extent of runoff generation. Over longer timescales, runoff drives  
553 the erosion of channels and the evolution of river networks [Beven and Kirkby, 1979,  
554 Litwin et al., 2020, Deal et al., 2018], the properties of which have direct consequences  
555 for the extent of wetted channel [Prancevic and Kirchner, 2019, Moidu et al., 2021] and  
556 aquatic habitat [Hwan and Carlson, 2016, Sabo et al., 2010]. Additional lithologically  
557 influenced hillslope and channel processes that impact habitat include base-level change  
558 induced propagation of knickpoints up channel networks as well as the introduction of  
559 static knickpoints in the form of megaboulders [Roering et al., 2015, Bennett et al.,  
560 2016], both of which occur in the Eel River watershed and can act as barriers to fish  
561 passage. Longer-term geologic and tectonic processes have been used to explain aspects  
562 of salmonid evolution, spatial distribution, and life-history strategies [Montgomery,  
563 2000, Montgomery et al., 2003, Hassan et al., 2008, Waples et al., 2008], but the potential  
564 indirect effects of these processes on salmonids via the formation of hillslopes, patterns  
565 of subsurface storage, and the genesis river flow-temperature regimes have not been  
566 previously identified.

567 We build on earlier research to emphasize that subsurface critical zone diversity  
568 likely favors expression of distinct salmonid life histories, and may lead to the emergence  
569 of life history syndromes. These include fry dispersing from ephemeral streams early in  
570 life and rearing downstream in non-natal habitats prior to ocean entry [Everest et al.,  
571 1973], fish over-summering in intermittent streams with refuge pools and out-migrating  
572 the following year [e.g. Hwan et al., 2018], and perennial streams supporting juveniles for  
573 one to two years prior to out-migration and, in the case of *O. mykiss*, trout completing  
574 the entire life history in the stream (as resident rainbow trout) [e.g. Kelson et al.,  
575 2020]. Thus, different critical zones within a watershed create a mosaic of habitats  
576 with different seasonalities and channel characteristics, which likely favor and support  
577 distinct life histories.

578 The success of different life histories will also vary across years due to variation in  
579 flow activation and access to tributary breeding habitats, potential for redd scour, spring  
580 flow recession and channel warming dynamics, connectivity to downstream non-natal  
581 rearing habitat, and disconnectivity of habitats and exposure to lethal temperatures – all  
582 of which are consequences of how climate is filtered through the critical zone. Across the  
583 watershed, maintaining a suite of salmonid populations that differ in their life histories  
584 may generate a portfolio effect, wherein the complex of populations is more stable than  
585 the individual populations [Schindler et al., 2010]. This suggests that the geography  
586 of critical zone structure may be an important factor contributing to the stability of  
587 salmonid population complexes, and that mapping the diversity of critical zones across  
588 the watershed may be essential to developing successful strategies for sustaining salmon  
589 in an era of change.



## 590 6. Conclusion

591 A lithological gradient across California’s Eel River illustrates the power of broadly  
592 applicable, storage-based frameworks to explain energetic and flow features of the  
593 stream environment that directly affect behavior and growth of riverine biota, such  
594 as salmonids. Different critical zones create a mosaic of habitats that likely favor and  
595 support different salmonid life histories, and may contribute to a stabilizing portfolio  
596 effect. Looking beyond the Eel River, our work motivates deeper study of geological  
597 and landscape controls on subsurface water storage capacity. At present, subsurface  
598 water supply is poorly mapped beyond shallow soils, despite increasing recognition that  
599 storage in deeper layers of weathered bedrock plays an essential role in determining  
600 moisture availability and runoff production in water-limited environments. Rapidly  
601 advancing methods for estimating and observing subsurface water storage dynamics at  
602 large scales present new opportunities for more clearly identifying landscape factors  
603 that influence aquatic biota. The linkages between water storage capacity, flow regime,  
604 stream energetics, and their consequences for salmonid life history expression highlight  
605 the need for a subsurface perspective on how landscapes and their evolution influence  
606 salmonid fishes. Better understanding the consequences of different critical zones for  
607 salmon life history diversity would help managers support resilient salmon populations.

## 608 Data and code availability

609 All plotting and analysis codes, and data used to generate plots, are available through  
610 Google Colaboratory Python notebooks posted on GitHub at [https://github.com/  
611 daviddralle/salmonid\\_and\\_subsurface](https://github.com/daviddralle/salmonid_and_subsurface).

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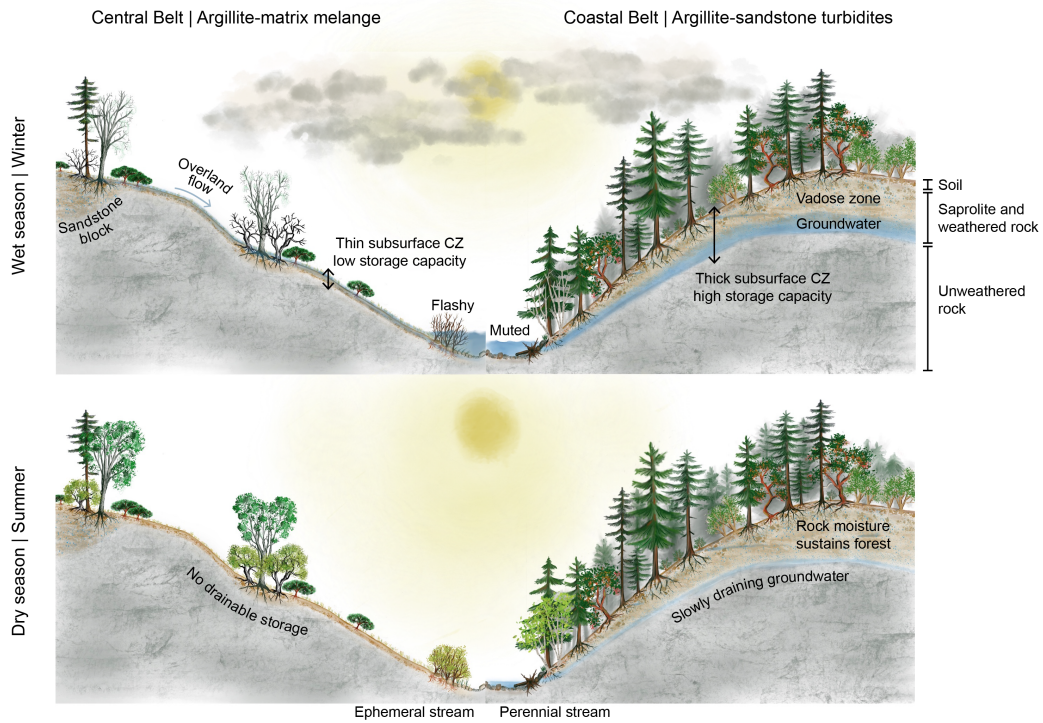


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Hillslope structure, subsurface water storage, and seasonal hydrological dynamics



**Figure 1.** Seasonal hydrological dynamics between hillslopes representing two dominant geologies in the Eel River watershed — Central Belt *mélange* (left) and Coastal Belt turbidites (right) — leading to contrasting CZ architectures and water storage capacities. A typical wet season (winter) snapshot is depicted in the top row, while the bottom row illustrates conditions later in the dry season (summer).

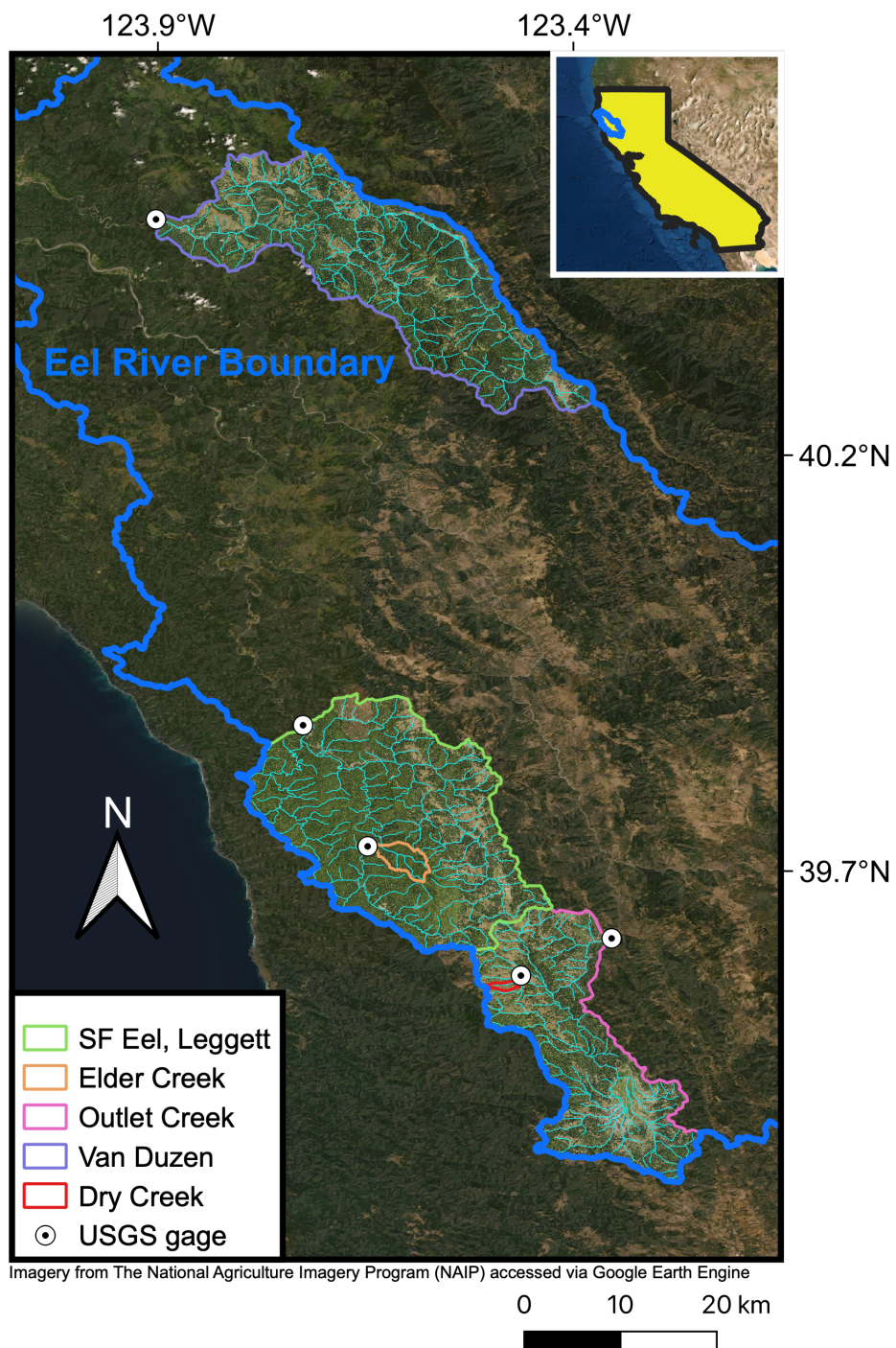
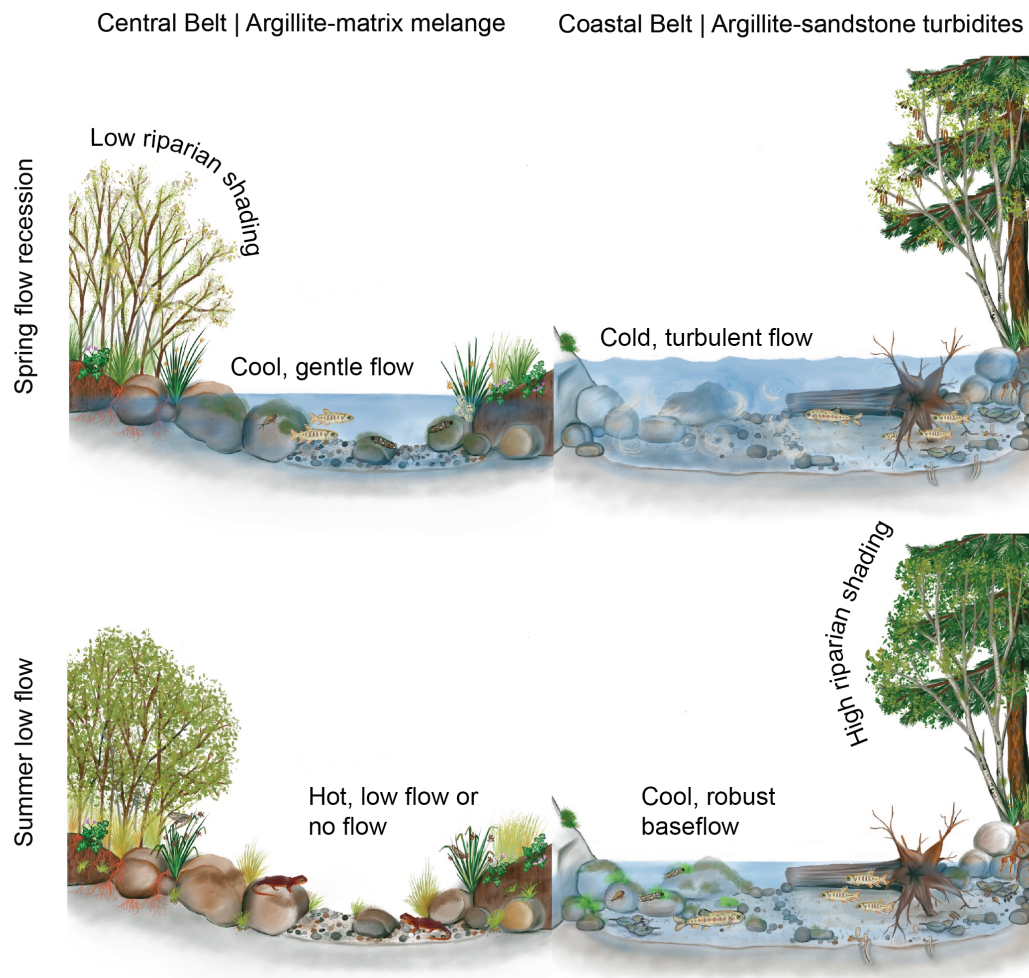
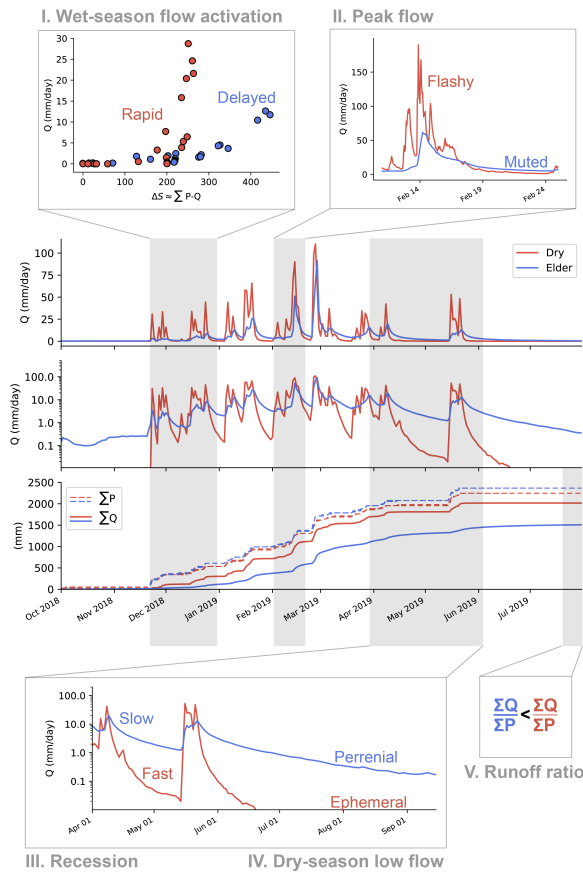


Figure 2. Overview map of the study watersheds.

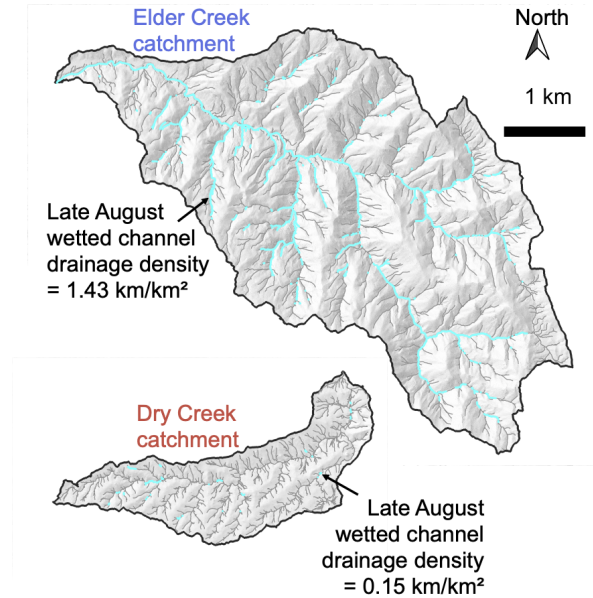


**Figure 3.** Typical progression of stream conditions between the Central Belt mélangé (left) and Coastal Belt turbidites (right) following the last significant rainfall event of the wet season. The top row illustrates conditions in the spring/early summer when air temperatures have begun to increase and stream flow is beginning its long seasonal recession. The bottom row depicts late summer low flow conditions when air temperatures are high and water availability in the stream is approaching its annual minimum.

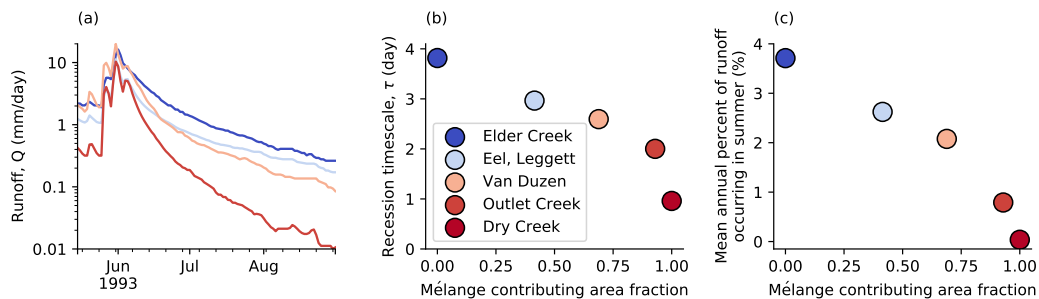


**Figure 4.** Storage capacity impacts important flow regime characteristics. Roman numerals correspond to entries in Table 2, while blue and red colors correspond to the Elder Creek and Dry Creek watersheds, which are representative end members of the Coastal Belt (relatively high storage capacity) and *mélange* (low storage capacity) geologies, respectively. The top two subplots of the center panel show 2019 water year hydrographs (on linear and log scales), while the bottom subplot shows cumulative precipitation ( $\sum P$ ) and cumulative discharge ( $\sum Q$ ). Focus panel (I) plots initial wet-season discharge as a function of the approximate dynamic storage  $\sum P - Q$ . (II) shows an expanded view of peak flows during a typical wet season storm sequence after initial wet-up. (III) illustrates differences in recession rates, while (IV) demonstrates how recession rate determines whether or not streams continue to flow through the entire dry season. (V) shows that a greater fraction of precipitation is converted to runoff in the Dry Creek watershed.

## VI. Dry season wetted channel extent



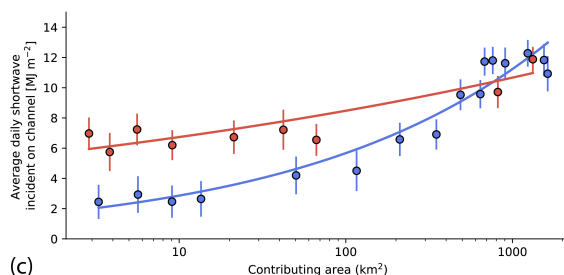
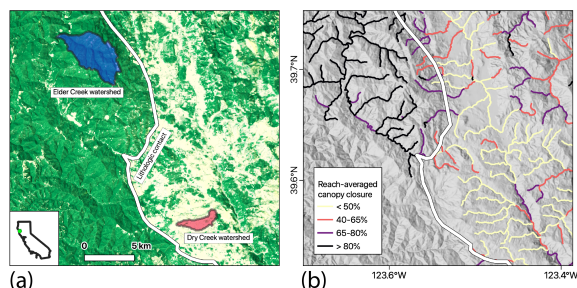
**Figure 5.** Dry season wetted channel extent is approximately tenfold higher in the representative Coastal belt watershed (Elder Creek) than the representative Central belt watershed (Dry Creek). Cyan lines denote liquid water at surface in channels (including all stagnant pools and flowing reaches, whether disconnected or continuous). Light grey lines denote approximate geomorphic channel extent. Each catchment is shown to scale but their relative locations have been modified for display purposes. Wetted channel data from Lovill et al. [2018b]; Elder Creek surveyed in 2014, Dry Creek in 2015. LiDAR-derived hillshade underlays from data collected by NCALM [Dietrich, 2014].



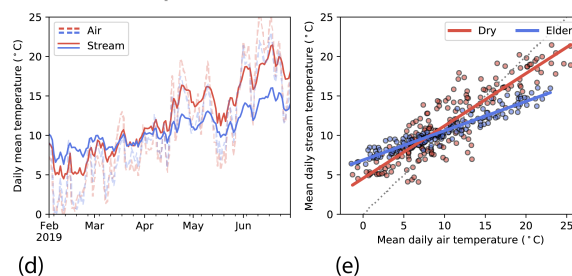
**Figure 6.** Watersheds across a gradient in fraction of mélange contributing area illustrate a range of flow recession behaviors (left subplot). The five colored points refer to the watersheds described in Table 3. Recession analysis (middle subplot) shows that larger fractional mélange contributing area results in faster recessions, as quantified by a simple exponential recession model:  $Q(t) = Q_0 e^{-t/\tau}$ . Smaller values of  $\tau$  in mélange-dominated watersheds correspond to faster recessions (i.e. rapid timescales of drainage). One consequence of the fast mélange recession is decreased water availability during the dry season, as demonstrated by the decrease in summer runoff fraction with increasing mélange fraction (right subplot). Conversely, Coastal Belt watersheds drain much more slowly, resulting in perennial flows and robust dry season discharge.



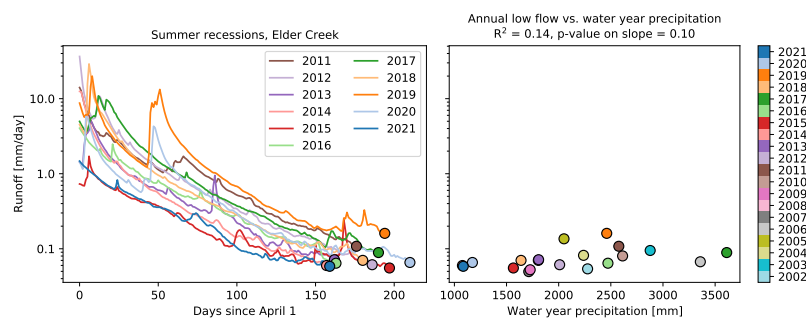
## VII. Channel shading



## VIII. Stream temperature



**Figure 7.** Differences in storage capacity across the geologic contact lead to stark differences in vegetation cover (a). Representative end-member catchments are outlined. Differences in canopy cover result in smaller delivery of shortwave radiation to headwater channel surfaces during summer months (June, July, August) in the Coastal belt (west of contact) versus Central belt (east of contact) (b). With increasing contributing area, channels widen, resulting in a convergence of channel-incident shortwave radiation between the two geologies (c). Red and blue points are binned averages  $\pm$  one standard deviation, from all Central belt and Coastal belt channels, respectively, in study area with available LiDAR data. Bin spacing varies to ensure a sufficient number of samples in each bin according to the procedure described in Kirchner [2009]. Contrasting stream temperature dynamics (d) due to differences in flow pathways, flow volumes, and riparian light environment. Subplot (e) demonstrates significantly higher sensitivity to changes in air temperature in the Dry Creek watershed.



**Figure 8.** Storage capacity decouples annual low flows from *total* water year precipitation at Elder Creek. (a) plots summer recessions as a function of days after April 1 from 2011 water year through 2021 water year, stopping on the day with the lowest observed flow for that calendar year (through December). The second subplot (b) shows that annual total precipitation is not a strong predictor of dry season low flows due to the mechanism of storage-capacity limitation. End-of-dry-season low flow conditions are more strongly controlled by rainfall conditions during the transition between wet and dry seasons. Annual rainfall data is derived from the PRISM Climate Group [PRISM Climate Group, 2004] daily rainfall product found on the Google Earth Engine Data Catalog.

GLOSSARY		
Term/quantity	Definition	Dimensions
Dynamic storage ( $\Delta S$ )	The volume of water stored in a catchment relative to some reference storage state, commonly taken to be zero at the driest time of year.	[L] or [L <sup>3</sup> ]
Storage capacity	The maximum observed value of dynamic storage.	[L] or [L <sup>3</sup> ]
Runoff (Q)	Stream discharge. Expressed in volumetric units (e.g. cubic meters per second), but also commonly reported in area-normalized units (e.g. mm/day) to facilitate runoff production intercomparisons between watersheds with different areas.	[L/T] or [L <sup>3</sup> /T]
Evapotranspiration (ET)	The sum of water use by vegetation (transpiration) and water returned to the atmosphere via evaporative losses from the ground surface or water bodies.	[L/T]
Recession timescale ( $\tau$ )	Determines the flow recession rate under the assumption that Q decline is linearly proportional to Q (i.e. $dQ/dt = -\frac{1}{\tau}Q$ ), leading to an exponential functional form for the streamflow recession.	[T]
Light penetration index (LPI)	The number of LiDAR returns from the ground or water surface divided by the total number of LiDAR returns.	Unitless
Runoff ratio	The ratio of total stream discharge to total precipitation over some time interval. Typically expressed over annual or longer timescales.	Unitless
Drainage density	The length of stream channel per area of watershed.	[L <sup>-1</sup> ]
Saturation overland flow	Overland flow that occurs when groundwater tables rise from below and intersect the ground surface, leading to runoff production via direct precipitation on saturated areas or water exfiltrating from the groundwater (return flow).	[L/T] or [L <sup>3</sup> /T]
Contributing area	Defined at a point, the total upstream watershed area draining all streams and hillslopes to that point.	[L <sup>2</sup> ]

**Table 1.** Table of terminology

Observed impacts of hillslope subsurface water storage capacity on streamflow characteristics in rain-dominated, Mediterranean climates

Category	Metric	Relative impact	Hypothesized reason/mechanism
<i>Water</i>	I. Wet season flow activation	Later with bigger storage capacity	More rain required to replenish bigger dry season hillslope water storage deficits
	II. Peakflow magnitude	Higher with smaller storage capacity	Small storage capacity more likely to fill, prompting greater and activation of faster (shallow near surface or overland) runoff pathways
	III. Rate of recession	Higher with smaller storage capacity	Deep slow flowpaths versus shallow fast flowpaths
	IV. Mean low (base)flow magnitude	Higher with bigger storage capacity	Greater reservoir to sustain dry season flow
	V. Annual runoff ratio	Lower with bigger storage capacity	More rainfall is partitioned to evapotranspiration where storage capacity is greater, enabling storage of wet-season rainfall for dry-season use by plants
	VI. Dry season wetted channel extent	Lower with smaller storage capacity	Lower supply of flow from hillslopes to channels
<i>Energy</i>	VII. Stream temperature	Colder in winter and warmer in summer with smaller storage capacity	Small storage capacity promotes shallower hillslope runoff pathways through regions similar to ambient air temperature; big storage capacity promotes deep hillslope runoff pathways through regions with mean annual air temperature
	VIII. Channel shading	Lower with smaller storage capacity in headwaters	Small storage capacity limit growth of plants, decreasing shade adjacent to channel; at large areas channel is sufficiently wide that riparian vegetation shading becomes less important

**Table 2.** Observed impacts of hillslope subsurface water storage capacity on streamflow characteristics in rain-dominated, Mediterranean climates

Watershed	USGS gage ID	Area (km <sup>2</sup> )	Mapped % mélange geology*
Dry Creek	N/A	3.5	100%
Outlet Creek	11472200	418	92%
Van Duzen River	11478500	574	68%
Eel River, Leggett	11475800	642	41%
Elder Creek, Branscomb	11475560	16.8	0%

**Table 3.** \* Non-mélange geology for these watersheds is predominantly Coastal Belt, with the exception of the Van Duzen River which also includes portions of the Eastern Belt.