

1 Single-blind test of airplane-based hyperspectral methane detection via 2 controlled releases

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8 **Abstract**

9 Methane leakage from point sources in the oil and gas industry is a major contributor to global greenhouse
10 gas emissions. The majority of such emissions come from a small fraction of “super-emitting” sources.
11 We evaluate the emission detection and quantification capabilities of Kairos Aerospace’s airplane-based
12 hyperspectral imaging methane emission detection system for methane fluxes of 18 to 1,025 kilograms
13 per hour of methane ($\text{kg}(\text{CH}_4)$). In blinded controlled releases of methane conducted over four days in
14 San Joaquin County, California, USA, Kairos detected 182 of 200 valid nonzero releases, including all
15 173 over 15 $\text{kg}(\text{CH}_4)$ per meter per second (mps) of wind and none of the 12 nonzero releases below
16 8.3 $\text{kg}(\text{CH}_4)$ /mps. 9 of the 26 releases in the partial detection range of 5 to 15 $\text{kg}(\text{CH}_4)$ /mps were
17 detected. There were no false positives: Kairos did not detect methane during any of the 21 negative
18 controls. Plume quantification accuracy depends on the wind measurement technique, with a parity slope
19 of 1.15 ($\sigma=0.037$, $R^2=0.84$, $N=185$) using a cup-based wind meter and 1.45 ($\sigma=0.059$, $R^2=0.80$, $N=157$)
20 using an ultrasonic anemometer. Performance is comparable even with only modeled wind data. For
21 emissions above 15 $\text{kg}(\text{CH}_4)$ /mps, quantification error scales as roughly 30-40% of emission size, even when
22 using wind reanalysis data instead of ground-based measurements. This reflects both uncertainty in wind
23 measurements and in Kairos’ estimates. These findings suggest that at 2 mps winds under favorable
24 environmental conditions in the US, Kairos could detect and quantify over 50% of total emissions by
25 identifying super-emitting sources.

26 **Introduction**

27 US natural gas (NG) production reached 40.1 trillion cubic feet in 2019, a 56% increase since 2009 EIA
28 (2020). The shift from coal toward less carbon-intensive NG and renewables has reduced the carbon
29 intensity of the US power sector Schivley et al. (2018). However, the climate benefits of NG cannot be
30 fully realized if methane leaks into the atmosphere at significant rates, as methane has a global warming
31 potential that is 28-36 times that of carbon dioxide over a 100-year period EPA (2017).

32 The US Environmental Protection Agency (EPA) greenhouse gas inventory states that NG and
33 petroleum systems accounted for 32% of total US methane emissions and about 4% of total US green-
34 house gas emissions in 2017 EPA (2019). Field surveys in gas-producing regions suggest that the EPA
35 inventory underestimates NG methane emissions, likely because EPA’s process-based approach does not
36 sufficiently account for emissions from extremely large sources Brandt et al. (2014); Lyon et al. (2016);
37 Zavala-Araiza et al. (2015). Emission sizes in the North American NG supply chain are found to fol-
38 low a heavy-tailed distribution, where the top 5% of point sources, so-called “super-emitters,” contribute

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39 over 50% of total emissions Brandt et al. (2014). A recent study indicates that 10% of the methane point
40 sources in California, including oil and gas facilities, landfills, wastewater treatment plants, and dairy
41 manure management sites, are responsible for 60% of the detected point-source emissions Duren et al.
42 (2019). Therefore, leak detection and repair (LDAR) programs could reduce the cost of detection and
43 mitigation by allowing mitigation efforts to focus on the largest sources. Given the limited resources and
44 manpower available for detection and repair, technologies for rapidly and accurately identifying super-
45 emitters are essential for guiding mitigation efforts.

46 Close-range approaches, such as optical gas imaging, are widely employed in ground-based LDAR
47 programs in the oil and gas industry. These methods are effective for source identification Ravikumar
48 et al. (2018), but can be slow and labor-intensive. Mobile systems with sensors placed on trucks, drones,
49 or aircraft have the potential advantage of speeding up detection by avoiding the need for manual detec-
50 tion via in-person site visits Ravikumar et al. (2019). In particular, mobile remote sensing via airplanes
51 or satellites can be used to target super-emitters, providing benefits of “low per-site cost, high spatial
52 coverage, and frequent sampling” Fox et al. (2019).

53 We examine a system developed by Kairos Aerospace (henceforth “Kairos”). Kairos’ LeakSurveyor
54 is a hyperspectral methane imaging system that is mounted on a light aircraft flown at general aviation
55 altitudes of approximately 900 m (3,000 feet) above ground level. The system uses an infrared imaging
56 spectrometer to detect methane with 3 m resolution, an optical camera to create an optical surface map
57 of the surveyed region, and GPS and inertial measurement units to record the position and orientation
58 of the sensor Berman et al. (2021). This system is capable of surveying roughly 400 square kilometers
59 (150 square miles) of oil and gas infrastructure in a single day Berman et al. (2021). See Supplementary
60 Information (SI) section S2 for further detail.

61 Several other airborne methane detection technologies exist, including other forms of airborne re-
62 mote sensing infrared imaging spectrometry, both thermal and shortwave Tratt et al. (2014); Thorpe et al.
63 (2016). The AVIRIS-NG airplane-based instrument developed at Jet Propulsion Laboratory (JPL) is one
64 such infrared spectrometer Thorpe et al. (2016); Duren et al. (2019), as are technologies produced by
65 Advisian, Baker Hughes (GE), and Seek Ops Inc., which employ laser absorption spectrometry on a he-
66 licopter, a drone, and a drone, respectively Ravikumar et al. (2019). Another airborne methane sensing
67 approach, employed by Scientific Aviation, requires the aircraft to circle a suspected source multiple
68 times at different altitudes while taking in situ methane concentration measurements to estimate the total
69 flux within the encircled area Schwietzke et al. (2019); Conley et al. (2017). Picarro employs a similar
70 cavity ringdown spectrometer in a joint drone-vehicle system that relies on transects of the plume Raviku-
71 mar et al. (2019). Ball Aerospace detects methane using airplane-based differential LIDAR Ravikumar
72 et al. (2019). Satellites such as GHGSat have begun to produce estimates of local methane emissions,
73 often using infrared spectrometry approaches Varon et al. (2020).

74 Many of these technologies, although none of the satellites, have been evaluated using controlled re-
75 lease experiments. Only some of these trials demonstrate a clear blinded experimental design Ravikumar
76 et al. (2019). Some have sample sizes below 10, too small to draw meaningful statistical conclusions
77 Schwietzke et al. (2019); Conley et al. (2017); Thorpe et al. (2016). Some, including Tratt et al. (2014)
78 and the Kairos evaluation in Schwietzke et al. (2019), focus only on detection, without evaluating methane
79 emission quantification performance. None test genuine controlled, metered emissions above 100 kgh,
80 far below emissions Kairos reports quantifying in the field. See the SI, Section S1 for further detail.

81 This study performs large-volume single-blind controlled releases, motivated in part by the Mobile
82 Monitoring Challenge (MMC), organized by the Stanford Natural Gas Initiative and the Environmental
83 Defense Fund (EDF). The 2018 MMC tested ten methane detection technologies through single-blind
84 controlled releases, with 6 out of the 10 participating technologies “correctly detecting over 90% of test
85 scenarios (true positive plus true negative rates)” Ravikumar et al. (2019). A similar set of single-blind

86 tests through the Methane Observation Networks with Innovative Technology (MONITOR) program
87 compares twelve handheld, mobile, and continuous monitoring approaches to methane detection at mod-
88 est emission rates Bell et al. (2020). The MONITOR findings demonstrate higher accuracy for handheld
89 and mobile methods over continuous monitoring techniques, which similarly highlights the importance
90 of high-precision follow-up detection for methane remote sensing systems.

91 We focus on characterizing quantification accuracy of the super-emitting methane point sources that
92 Kairos' technology was designed to quickly identify through aerial surveys. As a result, our emission
93 rates are two to three orders of magnitude larger than those in the MMC, reaching over 1,000 kilograms
94 of methane per hour ($\text{kg}(\text{CH}_4)$), as opposed to 0.29 $\text{kg}(\text{CH}_4)$ for most near-ground technologies in the
95 MMC and 29 $\text{kg}(\text{CH}_4)$ for two airplane and truck-based technologies Ravikumar et al. (2019).

96 **Materials and Methods**

97 *Airplane-based methane sensing technology*

98 Kairos' methane detection technology uses hyperspectral imaging from the wing of a small aircraft to
99 construct a two-dimensional image of excess methane concentrations integrated along the path between
100 the airplane and the ground. Each image is generated through a single pass over an area. Kairos' au-
101 tomated processing identifies methane plumes and calculates a wind-adjusted methane emission rate
102 in kilograms of methane per hour per meter per second of wind ($\text{kg}(\text{CH}_4)/\text{mps}$), henceforth denoted
103 kg/mps).

104 As described in Schwietzke et al. (2019) and Berman et al. (2021) the Kairos system combines signal
105 processing of optical and hyperspectral infrared images to produce estimates of the probability that a
106 given 3 m pixel displays excess methane above the background. Kairos then identifies connected clusters
107 of pixels identified as having high probability of excess methane and applies a simple physics-based
108 algorithm to estimate the associated emission rate (assuming a point source) Branson et al. (2021). As
109 described in the patent for this system Jones and Dieker (2019), the spectral resolution for systems of
110 this sort are "typically around 0.5 nm or better/finer," suggesting a spectral resolution in this range for
111 the Kairos system. Since data collection for Schwietzke et al. (2019) Kairos' technology has improved
112 in several ways, including a more sensitive infrared camera and a completely new atmospheric retrieval
113 algorithm. Most importantly, Kairos now produces methane quantification estimates, which it did not
114 during the trials in Schwietzke et al. (2019), which focused solely on detection.

115 The quantification algorithm described in Branson et al. (2021) uses a simple cross-sectional integra-
116 tion of excess methane concentrations within a detected plume. First, Kairos' proprietary algorithm uses
117 data from the spectrometer, optical camera, and GPS receiver, described in Jones and Dieker (2019), to
118 compute pixel-level estimates of excess methane column density between the airplane and the ground.
119 Next, they designate a spatially contiguous region as a plume if each pixel in the region has a methane
120 level that is statistically distinguishable from the background concentration (based on a proprietary met-
121 ric). Kairos estimates the wind direction based on the orientation of the vector between the point of
122 highest excess methane concentration in the plume and the furthest point within the plume from this
123 maximum concentration. Kairos then selects a core segment of the plume along this direction, passing
124 through the maximum and extending about 50% of the way to each end of the plume. Kairos then es-
125 timates total excess methane levels in the plume by summing the excess methane levels in each pixel
126 within the plume along that core segment, as in Equation 1.

$$127 \text{Methane}(\text{kg}) = \sum_i \text{ExcessColumnDensity}_i(\text{kg}/\text{m}^2) * \text{Area}_i(\text{m}^2) \quad (1)$$

128 Converting this methane mass to an emission rate requires assumptions of a constant emission rate,
constant wind speed, and slow methane diffusion compared to wind speed. Under these conditions,

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<http://dx.doi.org/10.1525/elementa.2021.00063>.

$$Rate(kg/s) = \frac{Methane(kg)}{Length(m)} * WindSpeed(m/s) \quad (2)$$

129 Where Length is the length of the core segment above used to estimate the excess methane mass
130 within the plume. Thus, Kairos' algorithm produces an estimate of the wind speed-normalized methane
131 emission rate, which must then be multiplied by an estimate or measurement of wind speed at the height
132 of the plume (Branson et al. (2021)).

133 Note that the spectrometer detects only methane and not other constituent components of natural gas,
134 such as ethane. See the SI, Section S2 and Kairos' patent for its system, Jones and Dieker (2019), for
135 further technical detail.

136 *Test location and set-up*

137 The Stanford team performed four days of single-blind controlled releases in San Joaquin County, Cal-
138 ifornia, on October 8th, 10th, 11th, and 15th, 2019. Kairos personnel were in the aircraft but were not
139 present at the ground release site. Stanford personnel designed the methane release schedule and con-
140 trolled the release rates with assistance from a natural gas release operator, Rawhide Leasing.

141 We measured methane flow rates through Sierra Instruments QuadraTherm 740i thermal mass flow
142 meters Sierra (2019). We measured wind speed and direction using both a Vantage Vue Sensor Suite with
143 a cup-based wind meter and a Gill Instruments WindSonic 60 two-dimensional ultrasonic anemometer
144 (not present on the first day of data collection) Davis (2018); Gill (2019). See the SI, Section S3 for
145 further detail.

146 *Single-blind experimental design*

147 The aerial test used a two-person airplane occupied by one pilot and one Kairos engineer, with Kairos'
148 LeakSurveyor instrument fastened to one wing strut. The Kairos engineer oversaw operations and radio
149 communication with ground crews from Stanford and Rawhide. As the aircraft passed over the test site,
150 the Kairos instrument attempted to detect any methane below. The aircraft flew repeated North-South
151 round-trip passes on a fixed route, passing overhead roughly every four minutes, varying from three to
152 five minutes depending on wind and other environmental conditions.

153 Kairos did not have access to data collected on the ground until they reported final results to Stanford
154 on October 24th. Kairos then received actual release rates and ground-based wind measurements on
155 October 29th. See the SI, Section S4 for further detail.

156 *Performance metrics*

157 We test Kairos' technology for detection accuracy, minimum detection threshold, and quantification accu-
158 racy. Here, detection accuracy is defined as the sum of true positive and true negative rates. The minimum
159 detection threshold analysis characterizes both the minimum release rate that the technology can detect
160 with some nonzero probability and the rate above which all releases are detected. Quantification accuracy
161 compares the estimated methane release rates to the true release rate. We compute quantification accu-
162 racy using a linear fit of released v. detected methane to assess the accuracy of the detection method. For
163 simplicity and intercomparability with other controlled release tests of methane detection technologies,
164 we use an ordinary least squares linear regression in the main analysis, although we discuss the potential
165 implications of weighted least squares approaches that account for variation in uncertainty across points
166 in the SI, Section S5.

167 **Results**

168 *Data summary*

169 A total of 230 data points were collected during the four-day single-blind tests, among which 21 (~9%)
170 were negative controls during which no methane was released. 40 releases (~17%) were dedicated to
171 characterizing the detection threshold by releasing at a rate between 0-50 kgh. The remaining large re-
172 leases (~74% of releases) were focused on testing the ability of the system to quantify high release
173 volumes. Among these, 110 releases were within the range of 50 and 500 kgh; and 59 were over 500 kgh.
174 Note that reported methane flow rates are 93.5% of metered natural gas flow rates PG&E (2019). The
175 volume rates are converted to mass rates based on the molar mass of methane and the molar volume at
176 the standard condition of 1 atm and 15 °C GPSA (2011).

177 Of the 230 data points collected, we exclude 9 from the baseline analysis due to technical issues such
178 as an incomplete plume image or controlled release practices that deviated from protocol. 4 additional
179 overflights did not result in valid data collection due to an incorrect flight altitude (see the SI, section
180 S3 for detail). When using wind speed from the cup meter, we exclude an additional 8 data points from
181 the 230 data points with measured 1-minute gust wind speed lower than 0.9 mps (2 miles per hour), the
182 rated uncertainty. We also exclude data points from the quantification analysis if there is not sufficient
183 time after a change in release level for full plume development. See the SI, Section S6 for further detail.

184 Figure 1 shows false color images of methane plumes detected by the Kairos instrument during the
185 trial, with blue and white representing low and high concentrations, respectively. All connected pixels,
186 with high enough confidence in detected excess methane, are considered to be within a single plume. If
187 there are multiple disconnected plumes, we consider the closest plume to the release point, consistent
188 with Kairos' internal practices. Figure 1a shows a Kairos image while no methane is being released.
189 Figure 1b shows a small plume at a release rate of 36 kgh, approaching the minimum detection threshold
190 of the instrument. The plume in Figure 1c is clearly visible, with a wind-adjusted release rate of 87 kgh.
191 Figures 1d-f show larger plumes with a wider field of view. See the SI, Section S7 for plume images in
192 terms of raw pixel-level excess methane column concentration.

193 *Detection probability and false positive rate*

194 Kairos previously published work reporting a 50% probability of detection at 9.2 kgh/mps Berman et al.
195 (2021). Considering the limited resources available for this study and the interest in testing quantification
196 accuracy at large emission rates, only 17% of data points have nonzero release rates below 50 kgh,
197 generally corresponding to at most rates of 25 kgh/mps, a level at which previous internal Kairos tests
198 suggest the instrument reliably detects emissions with close to 100% probability. After accounting for
199 data exclusion criteria and the wind speed conditions at the time of the release, 36 valid data points fall
200 in the range of 0-25 kgh/mps. We present this subset of the full dataset in Figure 2.

201 Note that we present results in these wind-normalized units for two reasons. First, Kairos' instrument
202 outputs readings in wind-normalized terms, so this presentation of results disentangles instrument capa-
203 bilities from the wind profile of the region in question. Second, these releases were on the low end of what
204 our release apparatus could accurately meter. As a result, for many of these smaller releases we left the
205 release level constant and allowed the wind to provide the variability. Thus, converting wind-normalized
206 releases to absolute methane fluxes would remove this variability. See the SI, Section S6.4 for minimum
207 detection results presented without wind normalization.

208 Figure 2 shows the fraction of emissions detected by Kairos as a function of the wind-speed-
209 normalized methane release rate for the 35 points below 25 kgh/mps, using 1-minute gust measurements
210 from the cup wind meter. Small circles on the top and bottom of the histogram represent each emission
211 and whether it was detected. Only 1 of the 14 data points in the 5-10 kgh/mps range was detected, with

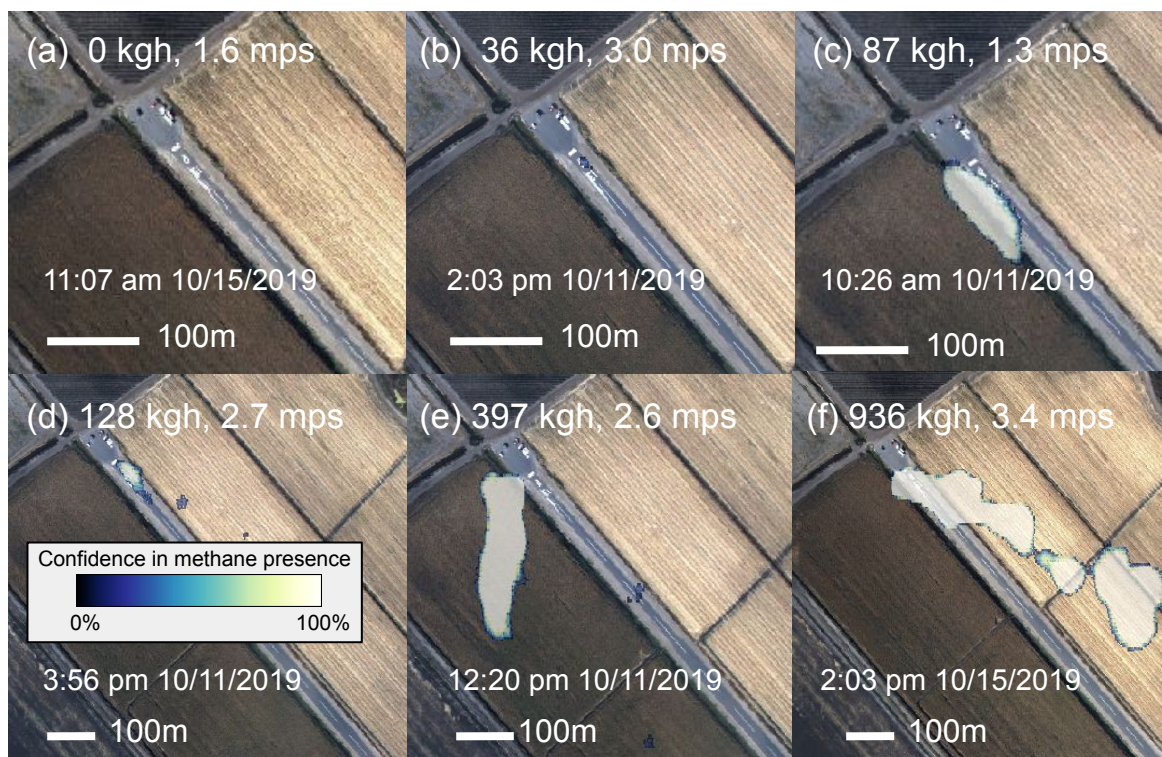


Figure 1. Examples of detected plumes associated with different methane release rates. Colorized plume images are based on post-processed spectrometer images, with blue and white representing low and high confidence of detected excess methane, respectively. Optical images were taken from the airplane as it passed overhead. Each image includes the measured methane release rate, in kgh, and wind speed from the ultrasonic anemometer. Note that the scale changes in the bottom row, d-f. (a) No release. (b) Small release, close to detection threshold. (c) Medium-sized release, low wind. (d) Medium-sized release, moderate wind. (e) Large release, moderate wind. (f) Approaching maximum release rate, moderate wind. Note that the plume images are based not on direct methane concentration measurements but on assessed confidence in the presence of excess methane, based on both the background concentration and the local variance in the strength of methane-indicating spectra.

212 a true emission rate of 8.3 kgh/mps. The detection rate rises to 67% for release rates of 10-15 kgh/mps.
213 Above 15 kgh/mps, 100% of emissions were detected, both in this subsample and in the data set as a
214 whole. Thus, the 50% probability of detection threshold likely occurs between 8.3 and 15 kgh/mps, con-
215 sistent with Kairos' internal trials. Error bars represent twice the standard error assuming a binomial
216 distribution, with no error bars shown for cases with 100% or 0% detection rates. This suggests that the
217 instrument can detect all emissions above about 15 kgh/mps with high probability.

218 Note that due to sensitive manual flow controls and high relative meter error and flow variability at
219 these low flow rates, for this section of the analysis we opted to hold the overall methane release rate
220 relatively constant for extended periods of time, allowing changes in wind speed to provide variability
221 in the wind-normalized release rates that Kairos' method produces. As a result, we characterize the min-
222 imum detection threshold in terms of wind-normalized methane release rates but do not have sufficient
223 variability in the overall flow rate to quantify the minimum detection threshold in terms of methane flow
224 rate.

225 To test for false positives, we devote ~9% of releases (21 releases) to negative controls with release

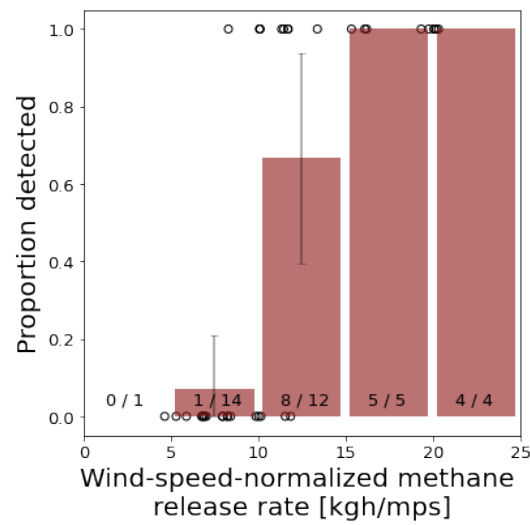


Figure 2. Binary detection results and the proportion of releases detected when the true release rates fall in the range of 0-25 kgh/mps. Each bin has a width of 5 kgh/mps. Kairos detected 100% of emissions above 15 kgh/mps. The smallest release detected was 8.3 kgh/mps. Error bars show twice the standard error assuming a binomial distribution. The fraction at the bottom of each bin denotes the number of true positives divided by the total number of releases in this range. Small circles on the top and bottom of the histogram represent each emission and whether it was detected.

226 rates of 0 kgh. Kairos reported no detections during these periods, leading to a false positive rate of
227 0%. This is in part because such remote sensing techniques are less sensitive than many other methane
228 detectors, missing small emissions but rarely triggering false positives. Thus, Kairos detected all 173
229 releases over 15 kgh/mps and none of the 12 nonzero releases below 8.3 kgh/mps.

230 In all, Kairos detected 182 of 200 valid nonzero releases and had no false positives in the 21 negative
231 controls, resulting in an overall accuracy of 91.9%, with 100% accuracy for releases above 15 kgh/mps.

232 100% detection above 15 kgh/mps represents an apparent improvement in the technology's minimum
233 detection capabilities compared to a controlled release field trial by Schwietzke et al. (2019) of an earlier
234 version of the technology. Schwietzke et al. (2019) found that Kairos detected emissions of 68.6 kgh
235 and 91.5 kgh with probabilities of 50% and 67%, respectively, including only passes with favorable
236 environmental conditions, or 17% and 29%, respectively using all 6 or 7 passes, respectively, regardless
237 of environmental conditions. That said, these results are not directly comparable because wind speed is
238 not reported in Schwietzke et al. (2019).

239 *Quantification accuracy*

240 Figure 3(a) shows 185 valid data points associated with nonzero release rates, comparing the metered
241 release rates (x-axis) to the estimated rates generated by Kairos. For consistency with Kairos' internal
242 testing procedures, we use 1-minute gust wind speed from the cup wind meter to convert the Kairos-
243 reported emission estimate in kgh/mps to kgh, using wind speed measured at the time of each pass.
244 Kairos reports point estimates in kgh/mps with no estimate of the uncertainty in these results. There is,
245 of course, uncertainty surrounding these point estimates and the data collected in this field trial allow
246 us to estimate it empirically. Uncertainties in wind measurements are $\pm \sim 0.9$ mps (± 2 mile per hour)
247 for the cup wind meter. This introduces errors in Kairos' estimates of release rates due to uncertainty in
248 wind speed, shown in the Y error bars in Figure 3(a). The length of the error bars is thus dependent on

249 the magnitude of the Kairos-reported number in kgh/mps. The y-axis of Figure 3(b) shows the Kairos-
250 reported number multiplied by 1-min gust wind speed measured with the ultrasonic anemometer, which
251 has a rated accuracy of roughly $\pm 2\%$, with some variation depending on wind speed Gill (2019). In this
252 case, the length of the error bars depends both on the magnitude of the measured wind speed and the
253 Kairos-reported quantification in kgh/mps. Because these error bars do not include uncertainty in Kairos'
254 quantification estimates, they necessarily underestimate uncertainty. Although the ultrasonic anemometer
255 has a much smaller measurement uncertainty, it was not present for the first day of data collection.
256 Therefore, we use results from the cup wind meter, shown in Figure 3(a) as a baseline. See the SI,
257 Section S8 for further detail. See the SI, Section S9 for further detail on uncertainty and variability in the
258 measured natural gas flow rate.

259 Using winds from the cup wind meter, the linear fit is relatively close to parity, with an R^2 of 0.84 and
260 a slope of 1.15 ($\sigma = 0.037$), shown in in Figure 3(a). The slope is statistically distinguishable from zero
261 at the $p=0.05$ level. This finding is robust to several techniques that correct for heteroskedasticity in the
262 data, shown in the SI, Section S5. Note that the confidence intervals in Figure 3 assume homoskedasticity,
263 which residual plots in Figure S7 suggest does not hold. Heteroskedastic confidence intervals would
264 widen further at higher release rates. Using ultrasonic anemometer wind data, R^2 drops to 0.80 and the
265 best fit line exhibits a larger slope of 1.45 ($\sigma = 0.059$), indicating somewhat more bias.

266 In the field, Kairos may not have access to on-the-ground wind measurements. In these circumstances,
267 one would likely use third-party data products to approximate local wind speed and direction. Figure
268 3(c) uses 1-minute gust wind reanalysis data from Dark Sky, a private company that estimates minute-
269 resolution wind speed at high spatial resolution across the United States based largely on publicly avail-
270 able data sources and atmospheric modeling Apple (2016). Dark Sky reports wind speed values at 10 m,
271 which we convert to 2.5 m values using a factor of $(2.5/10)^{0.15}$ for grassland terrain, based on Banuelos-
272 Ruedas et al. (2011). See the SI, Section S6.3.2 for further detail.

273 Figure 3(d) shows results using hourly surface gust data from the High-Resolution Rapid Refresh
274 (HRRR) wind reanalysis database, produced by the United States National Oceanic and Atmospheric
275 Administration NOAA (2020) averaging wind speed estimates over the nearest 9 km x 9 km area for the
276 three hours before, during, and after the Kairos measurement, based on Duren et al. (2019). For further
277 discussion of HRRR data, see the SI, Section S6.3.2. Note that (c) and (d) only exclude 6 and 10 data
278 points, respectively due to insufficient time for plume formation, while (a) excludes 15 of 200 nonzero
279 valid data points with either incomplete time for plume formation or wind speed measurements whose
280 uncertainty range contains zero. See the SI, Section S6.1 for further detail.

281 For both forms of wind reanalysis data, overall quantification performance is similar to the results
282 with ground-based wind data, with a slightly less precise linear fit. With Dark Sky in Figure 3(c), the R^2
283 falls slightly to 0.77 with a parity slope of 1.19, between the cup wind meter and ultrasonic anemometer
284 slopes. The R^2 for HRRR falls to 0.67 with a slope of 0.88, indicating average underestimates rather than
285 overestimates of total methane emissions.

286 Thus, Dark Sky data appears to provide a more precise estimate of overall emissions when ground-
287 based wind data are not available. However, because this is a proprietary product, the underlying algo-
288 rithms may change without notice. In addition, the data will likely not be publicly available after the
289 end of 2021 Grossman (2020). Although HRRR data have a lower spatial and temporal resolution, the
290 underlying process behind their production is more transparent. In addition, 15-minute HRRR data are
291 available for download within 48 hours of a given date, so future Kairos flights could likely acquire
292 publicly available HRRR data with a higher temporal resolution, potentially improving performance.

293 Although absolute residual plots in Figure S7(a-b) exhibit heteroskedasticity, percent residuals in Fig-
294 ure S7(c-d) appear relatively stationary in release size. Analysis of the smallest and largest 50% of the
295 data (above the 100% detection threshold) demonstrates that the mean and variance are not statistically

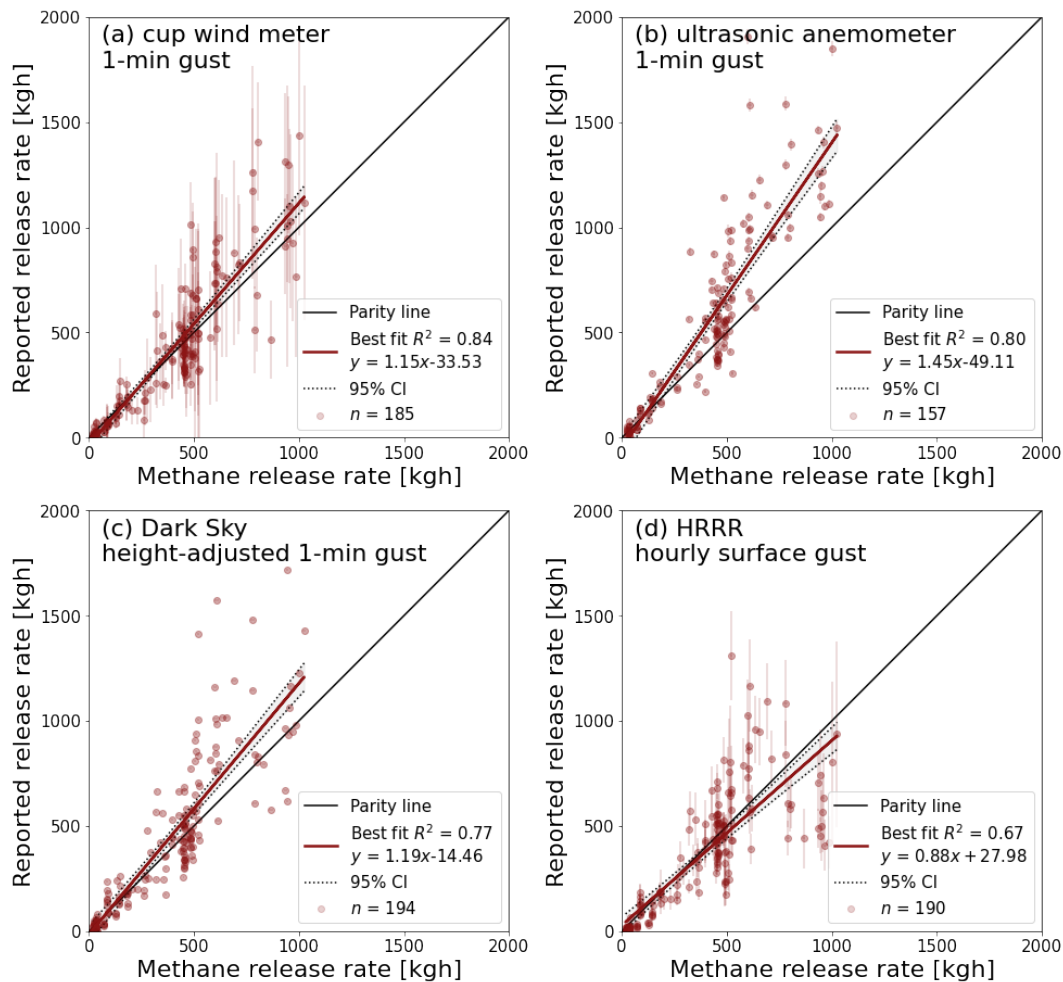


Figure 3. Parity chart of known nonzero methane release rates and the corresponding Kairos-reported estimate in kgh/mps multiplied by 1-min gust wind speed measured by (a) the cup wind meter, (b) the ultrasonic anemometer or reported by (c) height-adjusted values from the Dark Sky commercial wind reanalysis database, and (d) surface gusts from the High-Resolution Rapid Refresh (HRRR) database. The type of wind used in (a-c) is 1-minute gust wind speed. The X=Y parity line indicates perfect quantification. All four cases show a relatively close linear fit. (a-c) show mild to moderate bias toward overestimation based on minutely gust and (d) shows a mild underestimation based on hourly gust. The Dark Sky wind used in (c) is converted to 2.5-meter wind from 10-meter wind by applying a height adjustment factor. The HRRR wind used in (d) uses the method from Duren et al. (2019), averaging hourly surface gusts over three hours in the nearest 3x3 measurement locations (a box of 9 km by 9 km). See the SI, Section S6.3.2 for further detail on HRRR winds. 95% confidence intervals of the regression fits are shown. n = number of data points shown in each graph, which depends on data exclusion criteria described in the SI, Section S6.1. Y error bars are based on wind uncertainties, described in the SI, Section S8. Note that wind measurement uncertainty in the ultrasonic anemometer is smaller than point size, while Dark Sky does not report uncertainty. X error bars, not visible, are based on observed flow variability and flow meter error, described in the SI, Section S9.

298 size. This error range represents an estimate of the uncertainty associated with methane emissions quan-
299 tification from the Kairos system together with one of several potential sources of wind measurement
300 (including model-based estimates from HRRR and Dark Sky). Even with perfect ground-based wind
301 speed measurements, wind speeds may also vary at different heights in the atmosphere as the plume
302 rises. As a result, it is not possible to fully disentangle uncertainty in Kairos' measurements from wind
303 uncertainty. Note that this computation does not include points below the 15 kgh/mps 100% detection
304 threshold because false negatives introduce an additional form of error that is not representative of the
305 error profile of larger emissions. See the SI, Section S5 for further detail.

306 These results demonstrate a high level of detector performance, even without on-the-ground wind
307 measurements, in terms of high R^2 and low bias compared to past controlled releases for mobile methane
308 detectors in Ravikumar et al. (2019); Duren et al. (2019); Schwietzke et al. (2019); Conley et al. (2017)
309 and Foster-Wittig et al. (2015). That said, most controlled release studies operate at one to two orders of
310 magnitude lower release volumes with smaller sample sizes predominantly clustered near the minimum
311 detection threshold and most do not appear to employ a blinded experimental design. See the SI, Section
312 S1 for further detail.

313 *Estimate of field efficacy*

314 Using a bottom-up inventory of 1,009 methane emission sites from the US oil and gas system from
315 Omara et al. (2018), a compilation of data from nine separate studies and eight oil and gas-producing
316 basins, we estimate that given 2 mps winds and emission detection fractions based on the probabilities
317 from Figure 2, adoption of this technology would detect 53% of total emissions, with 49% coming from
318 24 sites above the 100% detection threshold of 15 kgh/mps. At 1 mps winds, this rises to 63% of total
319 emissions. At 4 mps winds, this falls to 41% of total emissions. At 7 mps, the maximum wind speed at
320 which it is safe for these airplane-based surveys, Kairos would still detect 32% of total emissions. Note
321 that this inventory combines emissions from multiple basins. In practice, detector efficacy would likely
322 vary across basins due to different emission profiles. In addition, we do not perform a full stochastic
323 techno-economic analysis, such as that in the Fugitive Emissions Abatement Simulation Toolkit, which
324 would be necessary to determine the cost-effective mitigation potential of airplane-based methane sensing
325 technology Kemp et al. (2016).

326 **Discussion**

327 These results suggest that in suitable contexts, aerial surveys at modest wind speeds could detect 50%
328 or more of total methane emissions even without ground-based wind measurements. This process can
329 screen assets much more rapidly than traditional leak detection and repair methods, with few if any
330 resource-diverting false positives. Thus, this technology could provide rapid detection of super-emitting
331 methane leaks in upstream and midstream oil and gas, likely as a supplement to more precise but more
332 labor-intensive leak detection and repair programs. More sensitive instruments would likely be required
333 for most distribution system applications.

334 The overall cost of this field trial was roughly \$50,000 including materials, natural gas release equip-
335 ment rental, gas, and personnel, flight time, space rental, and miscellaneous expenses (not including
336 Stanford researchers' time, which is difficult to quantify but could approximately double the cost). Com-
337 panies will often participate in trials of their technology free of charge. Note that controlled release testing
338 for more sensitive sensors aimed at lower emission volumes is substantially less expensive. Basin-wide or
339 state-wide aerial methane emissions survey campaigns can cost \$1 million or more. Testing new instru-
340 ments with blinded controlled releases at a range of methane emission levels approaching those expected
341 in the field, with a statistically meaningful sample size of at least a few dozen, would increase confidence

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342 in the capabilities of these methods, thus adding substantial value to the data from such field campaigns.

343 **Contributions**

344 Substantial contributions to conception and design: EDS, YC, APR, ARB

345 Acquisition of data: EDS, YC

346 Analysis and interpretation of data: EDS, YC, ARB

347 Drafting the article or revising it critically for important intellectual content: EDS, YC, APR, ARB

348 Final approval of the version to be published: EDS, YC, APR, ARB

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358 **Competing interests**

359 The authors have no competing interests to declare

360 **Supplemental material**

361 The supplemental materials for this article can be found as follows:

- 362 • **Text S1.** Supplementary Information for Single-blind test of airplane-based hyperspectral methane
363 detection via controlled releases. <https://osf.io/vqnpb/>
 - 364 – **S1.** Comparison with other controlled release studies
 - 365 – **S2.** Kairos Aerospace technology
 - 366 – **S3.** Controlled release set-up
 - 367 – **S4.** Single-blind experimental design
 - 368 – **S5.** Addressing heteroskedasticity
 - 369 – **S6.** Data exclusion, sensitivity analyses
 - 370 – **S7.** Excess methane concentration plume images
 - 371 – **S8.** Wind variability and uncertainty
 - 372 – **S9.** Flow variability and uncertainty

373 **Data accessibility statement**

374 All data and code required to reproduce the results of this article are available on GitHub at:
375 [https://github.com/yuliachen/Single-blind-test-of-airplane-based-hyperspectral-methane-detection-via-](https://github.com/yuliachen/Single-blind-test-of-airplane-based-hyperspectral-methane-detection-via-controlled-releases)
376 [controlled-releases](https://github.com/yuliachen/Single-blind-test-of-airplane-based-hyperspectral-methane-detection-via-controlled-releases)

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