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## PREPRINT

This is a non-peer reviewed preprint submitted to EarthArXiv

# Contextual learning improves forest above-ground biomass estimates from UAV-LiDAR: use of tree trait associations.

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### Funding information

Forest structure analyses and biomass prediction systems are key tools for advancing forest trait-based ecology and ecosystem stewardship. The combination of near-field remote sensing techniques—e.g., Unmanned Aerial Vehicles (UAV) and Light Detection and Ranging (LiDAR) systems—with machine-learning methods enhances the accuracy of forest structure analyses and above ground-biomass (AGB) estimates. In this study, we utilized a UAV-LiDAR system to map the 3D architecture of a monoculture Norway spruce forest in Davos, Switzerland, where a field-based inventory served as ground truth data. The objectives of this effort were (i) to gain insights into variation and gradients of structural traits (i.e., tree height) and (ii) to evaluate whether this knowledge of community structure may prove useful as contextual information to improve predictions of AGB at the individual tree level. To investigate the local association of structural traits, we segmented the point cloud data scene into individual trees and treated tree height as the morphological variable of interest. We then used local indicators of spatial association to determine the extent of significant local context, and defined tree neighborhoods

within the forest. For the task of AGB regression, we obtained results of several feature-based regression methods (i.e., AdaBoost, Lasso and Random Forest) and evaluated these based on nested cross-validation.

We applied this approach to two separate tree data sets within the same site, one being clustered and continuous, the other discontinuous and scattered in separate sampling plots. In both cases, we found evidence of enhanced AGB prediction performance in context-aware regressions, indicating that gradients in morphological tree traits across the ecosystem proxy for unveiled ecological information that influence tree growth, which can be leveraged to enhance predictions of AGB.

#### KEYWORDS

above-ground biomass, forest structure, functional trait mapping, machine learning, contextual learning, UAV-LiDAR, quantitative ecology

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This preprint manuscript is currently being considered for formal peer-reviewed publication. Please note that, this is the submitted version of the study, and has yet to be formally accepted. Subsequent versions of this manuscript may present slight differences in content. If accepted, the final version of this manuscript will be updated with the accepted manuscript. This preprint will also be linked to the formal publication via its Digital Object Identifier (DOI). Please, feel free to contact any of the authors; we welcome feedback.

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## 46 1 | INTRODUCTION

47 Above-ground biomass (AGB) is a critical component for determining global land carbon (C) budgets. Worldwide,  
48 forests are critical agents of the global C-cycle, as major sinks of atmospheric carbon dioxide [1, 2]. However, current  
49 estimates of C-cycling from land ecosystems have large margins of uncertainty, partly because of uncertainties in AGB  
50 estimates [3]. To date, the existence of seemingly equivalent but disparate AGB products hinders a more frequent use  
51 of such data products in conservation management [4] or current policy making [5]. Therefore, the growing number  
52 of AGB products need to be harmonized and techniques standardized.

53 Ongoing efforts within the remote sensing (RS) community aim at reducing the uncertainty of AGB predictions  
54 to allow reliable estimates across scales [6]. This is a considerable undertaking, since the technology, data sources  
55 and methods employed at different scales vary greatly, making it difficult to track propagated errors [6], or to de-  
56 termine how different end-products (i.e., AGB maps) perform comparatively [7]. This lack of standardization results  
57 in AGB and trait-mapping products with different degrees of agreement, making it particularly relevant to compare  
58 data-acquisition methods [8] and validation procedures [7, 9] of the AGB products [10, 11]. In this scenario, Un-  
59 manned Aerial Vehicle (UAV) & Light Detection and Ranging (LiDAR) monitoring systems are regarded as particularly  
60 versatile [12], accurate and cost-effective [13] tools to be bridged to regional scale maps seamlessly [6]

61 Current RS-driven biomass research focuses on algorithmic developments for the detection and segmentation of  
62 single trees, in order to enable more precise estimates of structural tree traits [14, 15, 16]. Also, recent reference work  
63 analyzing forest structure exploits the use of laser sensors to develop methods for volume reconstruction from point  
64 cloud data (PCD) [17, 18]. Furthermore, large-scale AGB mapping initiatives pursue characterizing scale-independent  
65 LiDAR-derived predictors to develop LiDAR-to-AGB models across scales [10, 19]. More specifically, in relation to  
66 recent advances in forest monitoring using close-range LiDAR technologies the development of versatile, practical and  
67 new cost-effective sensors [13] and platforms [20] has seen a rapid growth, widening the applicability of the emerging  
68 LiDAR systems [12, 21]. Their emergence has triggered discussions and investigations related to sensor accuracy,  
69 sensor types [22] and purpose-adapted surveying methods [23]. However, to date, efficient tree-level phenotyping  
70 has been challenged by several forest structural conditions, such as crown-shift [24], canopy closure [25] and tree  
71 clumping effects [26, 27].

72 Traditionally, assessments of tree structural traits from middle- and close-range RS data focused on individual tree  
73 attributes as predictors (e.g., tree height, tree canopy metrics) [28]. Over time, methods that consider plot-level metrics  
74 to improve the regression of individual tree traits emerged, e.g., non-linear mixed effects (NLME) methods [29, 30, 31],  
75 or competition-based methods [32, 33, 34]. In fact, plot-level information has long been reported as beneficial in  
76 diverse tree-level assessments, e.g., diameter at breast-height (DBH) [31], surface-based fuel potential [35] or tree  
77 height and crown structural metrics [31, 36, 37].

78 While all these theoretical and technological advances have accelerated the progress of forest biomass research  
79 in an unprecedented manner, there is still room for improvement as regards integrating ecological reasoning into  
80 biomass research. For instance, scholars argue that understanding local ecological processes requires monitoring  
81 biomass of individual trees [14, 33]. However, the opposite idea is seldom discussed: how and to what extent can  
82 community ecology processes be harnessed in tree-level AGB regression experiments [32, 34]? We consider this  
83 line of work within AGB research as yet relatively unexplored, with some exceptions. Earlier works have proposed to  
84 account for the effects of immediate competition pressure on tree growth with either distance-based [34] or distance-  
85 independent metrics [38], and judge such approaches beneficial. More recently, Sun et al. (2019) [32] evaluated the  
86 potential of distance- and ranking-based competition metrics for improving predictions of tree diameter growth, and  
87 found them outperforming competition-unaware prediction models. Similarly, Zhang et al. (2020) [33] ranked trees

88 by quantiles and competition levels to enhance predictions of the tree height-to-diameter ratio.

89 Despite the utility of current methods that leverage plot-level metrics, they remain unsatisfactory in some re-  
90 spects. Many of such methods are not directly transferable to a RS framework because they use understory metrics  
91 as predictors [32, 33]. More importantly, questions remain about the optimal scale at which such metrics should be  
92 retrieved. We noticed that, in the reviewed studies, the spatial scale at which ecological phenomena operate was not  
93 questioned. Instead, the focus is often on plot-level metrics, measured at an arbitrary distance that corresponds to the  
94 size of artificially-bounded forest inventory plots [29, 30, 31, 32, 33]. To the best of our knowledge, tree-level AGB  
95 and trait assessments considering context information are currently limited for one or more of the following reasons:  
96 (i) they characterize context with uniquely process-specific indices (e.g., competition pressure from immediate neigh-  
97 bors) [32, 33, 34]; (ii) calibrating models with neighbor-metrics retrieved from artificially-bounded inventory plots (e.g.,  
98 NLME methods) [29, 30, 31]; (iii) they do not sufficiently account for the spatial scale at which the ecological phenom-  
99 ena affect the trait under investigation. Moreover, when the relationship between the plot-level predictors used and  
100 the ecological phenomena is described, often ancillary data sources are incorporated (e.g., tree stand age) [33, 39] or  
101 non-standardized, forest management terms, e.g., "stand quality", "site index", "dominance index" [30, 33, 39]. These  
102 shortcomings currently hinder the transferability of such methods to other regions, larger scales or different data  
103 collection surveys.

104 Given the need for methods to be scalable and transferable, it certainly appears beneficial to characterize biotic  
105 interactions (e.g., tree competition) or environmental filtering (e.g., soil depth, nutrient availability) with metrics that  
106 can be remotely sensed such as tree height and crown dimensions, rather than understory predictors, as has been  
107 proposed [32].

108 A central question in community ecology is how functional trait diversity (e.g., the spatial distribution of tree  
109 heights) relates to ecosystem dynamics across environmental gradients [40, 41]. In this regard, current AGB research  
110 and mapping initiatives [10, 42] have not yet thoroughly investigated the opportunity to consider two-dimensional  
111 spatial patterns [43] of remotely-sensed predictors (e.g., tree height, crown dimensions) to enhance tree-level AGB  
112 estimations. These predictors, being subject to a concert of spatially continuous ecological factors—e.g., adaptation  
113 to different lighting conditions [44, 45] and soil depth variation [46], or the availability of nutrients and nonstructural  
114 carbohydrates [47]—exhibit, as a response, local spatial association (i.e., geographical clusters and gradients of similar  
115 tree heights) [48]. Such spatial associations of predictors may serve as proxy for the combined effect of the ecological  
116 phenomena being considered. Therefore, provided that spatially continuous ecological factors mediate individual tree  
117 growth [1, 49]—and these can be remotely sensed—, it seems plausible to use this information about the local context  
118 to improve tree-level AGB assessments. In addition, it appears relevant to examine the significance and spatial extent  
119 of the local context, as well as the relationship between context-based traits and individual tree traits.

120 In this framework, machine learning (ML) regression methods seem to be an interesting approach to incorporating  
121 a contextual analysis, given that they are commonly integrated into UAV-based forest mapping studies [50]. In such  
122 approaches it has been shown that the inclusion of information of local context (i.e., information about the surround-  
123 ings of the target object) improves their performance [51, 52]. This information can be included in a learning model by  
124 either enlarging the receptive field size (i.e., widening the field of view) [53, 54] or by incorporating context-aware fea-  
125 tures that encode neighboring information into the target object [55] (i.e., a subject tree in our case). In other research  
126 fields, such contextual analyses have been successfully incorporated into learning models to improve assessments in,  
127 e.g., land-use dynamics [56], Earth system modelling [57] or urban growth [58].

128 To date, the absence of standardized and scalable approaches to incorporate context information into AGB regres-  
129 sion experiments has hindered the potential to harness context to enhancing AGB mapping products. The potential of  
130 the spatial association patterns of individual tree traits to represent the effect of local ecological phenomena on tree

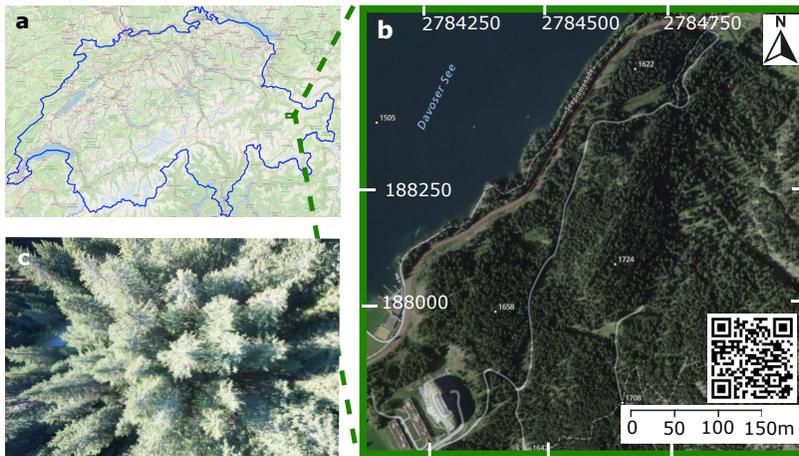
131 structure is an area that yet remains partially unexplored in AGB research. Specifically, the question that is still unan-  
132 swered is how context-awareness can be incorporated into a RS framework and leveraged to enhance AGB estimates  
133 at the individual tree level.

134 In order to address this question, this study aims to evaluate whether AGB regression models can leverage context-  
135 awareness to improve AGB estimates at the individual tree level in a mono-culture forest, where the context is defined  
136 without using external ancillary data sources, or using neighborhood metrics of artificially-bounded inventory plots.  
137 The objectives for achieving this aim include: (i) collecting close-range point cloud data (PCD) via UAV-LiDAR survey-  
138 ing, (ii) retrieving contextual information based on the geographic spatial association of tree heights, (iii) developing  
139 methods that allowed the context to be defined and incorporated into regression experiments and (iv) evaluating the  
140 effect of introducing context-awareness in tree-level AGB estimates.

## 141 2 | MATERIALS AND METHODS

### 142 2.1 | Study Area

143 The Seehornwald Davos research site is located in a managed subalpine coniferous forest on the western flank of the  
144 Seehorn mountain, near Davos, in the Swiss Alps ( $46^{\circ} 48'55.2''$  N,  $9^{\circ} 51'21.3''$  E, 1640 m a.s.l.). The site is labeled  
145 as a class-1 forest ecosystem station of the Integrated Carbon Ecosystem Station (ICOS) network [59] where regular  
146 forest inventory measurements are collected following standardized protocols. The site is covered by spruce trees  
147 (*Picea abies* (L.) Karst.) with an average height and age of 18 m and 100 years, respectively, while some trees reach a  
148 height of 35m and an age of 300 years. Patchy vegetation (i.e., dwarf shrubs and mosses) covers around 30% of the  
149 forest floor. The research site is part of national (LWF[60], TreeNet[61], SwissFluxNet [62]) and international research  
150 networks (ICOS [63], ICP Forests [64], eLTER [65]).

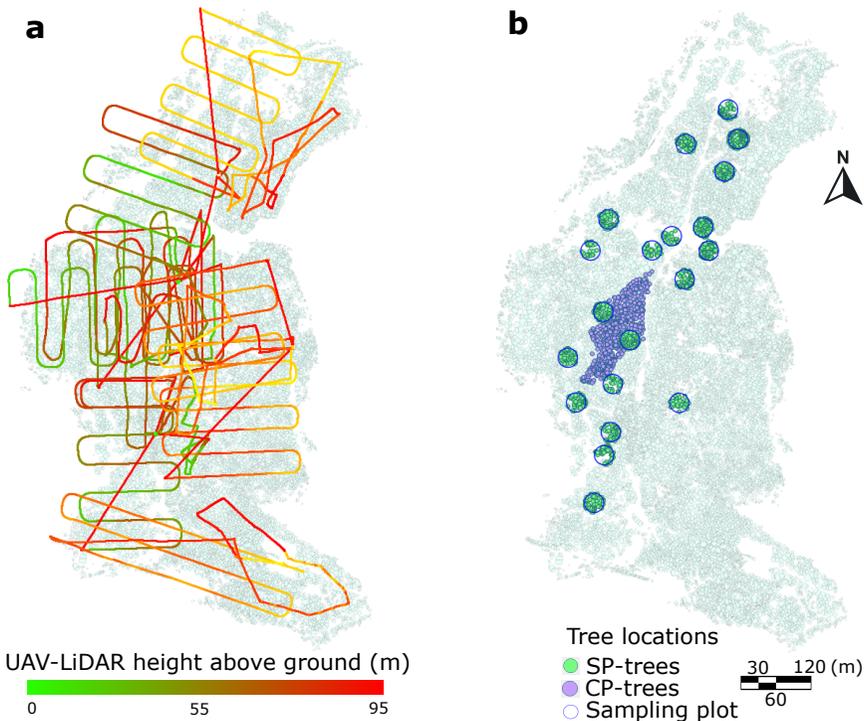


**FIGURE 1** a: location of the study site; the blue outline delineates the national territory of Switzerland (adapted from open.sourcemap.com). b: orthoimage of the study site (adapted from swisstopo.admin.ch); coordinate units are in m, with LV95 as a projected reference system; QR code links to additional information of the study site. c: RGB image of forest canopy from a nadir angle taken during the survey.

## 151 2.2 | UAV-LiDAR Survey and Field-Based Measurements

152 We used a UAV-borne LiDAR system mounted to a DJI Matrice 600 Pro payload at a 90° pitch angle, and same heading  
 153 and roll as the UAV platform. The system included a discrete infrared LiDAR scanner (M8 sensor, Quanenergy Systems,  
 154 Inc. Sunyvale, CA, USA) and the corresponding state-of-the-art inertial and navigation systems. In addition, we used a  
 155 ground based differential Global Positioning System (dGPS, Trimble R8) during the UAV-LiDAR survey, set up in post-  
 156 positioning kinematic (PPK) mode, which logged real-time satellite coverage (cf. Ravenga et al. 2022 [66] for details on  
 157 the airborne and ground system). The coupling of the satellite coverage data with the UAV-based laser and navigation  
 158 data produced, allowed the generation of georeferenced point clouds, following Davidson et al. (2019) [67].

159 Data were acquired with a terrain-adapted flight height (Figure 2, a) and 20% overlap between individual LiDAR  
 160 scans of ca. 50 m width and 250 points/ $m^2$  (cf. Revenga et al. 2022 [66] for additional details on applied flight  
 161 parameters). The surveys were conducted in October 2021, coinciding with the end of the growing season. Figure  
 162 2 (a) shows the trajectories of the individual UAV-LiDAR flights during the survey campaign. While the standard  
 163 survey coverage followed a regular auto-pilot flight grid, certain flight lines had to be manually piloted to adapt to the  
 164 topography and local forest structure.



**FIGURE 2** a: trajectories of individual flights during survey of the Unstaffed Aerial Vehicle (UAV) Light Detection and Ranging (LiDAR) sensor; color gradient indicates height above ground at take-off point. b: spatial distribution of field-based forest inventory. Dots represent the locations of the ground-truth labels. The sampling plot-trees (SP-trees,  $N = 1635$  trees) are shown in green; the control plot-trees (CP-trees,  $N = 845$  trees) are shown in purple. In both a and b, the underlying polygon dataset shows the individual tree canopies (ITC) after the canopy height model (CHM) segmentation.

165 The field-based measurements (shown in Figure 2, b) are taken on a yearly basis as part of a long-term ecosystem  
166 monitoring initiative—jointly organized by ICOS [63] and the Swiss Federal Institute for Forest, Snow and Landscape  
167 Research (WSL) [60]. Based on standardized methods (i.e., *Sanasilva Inventory* protocol [68]), expert field workers  
168 monitor tree crown status in terms of color and density, focusing on three groups of indicators: variations in size,  
169 density and color. The number of trees that have died since the previous survey, as well as the new ones that reached  
170 a minimum DBH of 5 cm are also recorded. As ancillary data, the following parameters are also evaluated: identified  
171 causes of defoliation, competition in the canopy, and the presence of epiphytes, mistletoe and climbing plants [69].

172 We treated two different datasets separately as ground-truth measurements within the same study area: con-  
173 trol plot trees (CP-trees, 4 adjacent monitoring units) and sampling plot trees (SP-trees, 20 scattered units of 15 m  
174 radius). Several factors led us to consider both datasets separately: (i) the CP dataset is clustered and spatially con-  
175 tinuous, while the SP dataset is spatially discontinuous and distributed along the valley. (ii) the two datasets present  
176 significant differences in morphological trait distribution (see Supporting Information, Annex IV). (iii) the variability in  
177 context metrics between the two datasets varied markedly. (iv) the field-based instrumentation and protocols used  
178 for monitoring presented minor differences between both datasets. Figure 2 (b) shows the spatial distribution of the  
179 field-based forest inventory. The CP tree position was recorded using a Leica System 1200 (GPS total station). The  
180 location and size of the sampling plots were defined according to ICOS protocols [70]. The center location of the  
181 SP plots was determined using a GPS Leica CS20 (antenna GS15) with a real-time kinematic (RTK) signal (accuracy  
182 measurements ranges from 0.03m to 0.7m). Next, the trees in the SP plots were positioned by measuring the azimuth  
183 with a field goniometer, while the horizontal distance and the inclination from the plot centers was determined using  
184 a Vertex Laser Geo meter. The accuracy of foot location of trees in the SP plots is within 0.5 m and 1.2 m. The  
185 field-based inventories used as ground-truth contain measurements taken between October 2019 and July 2021.

## 186 2.3 | Data Processing

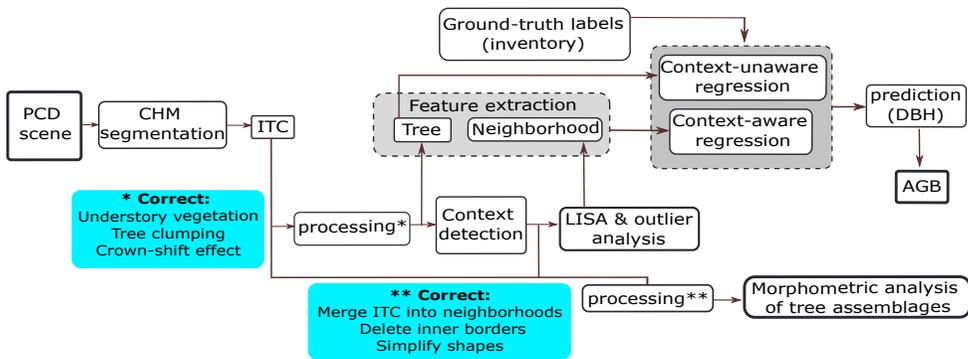
187 The workflow followed in this study is presented in Figure 3. Initially, the PCD generation followed the approach de-  
188 scribed in Revenga et al. (2022) [66]. The resulting PCD scene was normalized and rasterized to obtain a canopy height  
189 model (CHM), which in turn was subject to individual tree crown segmentation [71] producing a two-dimensional poly-  
190 gon dataset. For the CHM segmentation, we utilized a watershed algorithm that is specifically designed for coniferous  
191 forests [71] (implemented in the LiDAR360 software [72]). The match between field-based measurements and indi-  
192 vidual tree crown (ITC) polygons was conducted based on the closest distance between the field-based GPS point  
193 measurement and the ITC polygon centroid.

194 In order to ensure that only the LiDAR-detected trees would be accounted for in the regression experiment, a  
195 pre-processing manual task was undertaken (marked \* in Figure 3). First, understory trees that passed unnoticed  
196 to the UAV-LiDAR survey were removed. Second, we filtered clumped trees based on tree height by selecting the  
197 field-based measurement of the highest tree when two measurements were less than 1 m apart, while removing the  
198 measurement of the other tree. Third, we corrected for a crown shift effect, i.e., some high and skewed trees were  
199 affected by the presence of a smaller neighboring tree (affecting about 5% of trees) being closer to its corresponding  
200 ITC polygon centroid, thus introducing a wrong match between the field-based measurement and the LiDAR-derived  
201 metrics.

202 Afterwards, using the LiDAR-derived height as polygon attribute, we calculated the distance at which the spatial  
203 autocorrelation of tree height was most significant in order to define the optimal neighborhood size (as explained  
204 in Section 3.1). Once the optimal neighborhood size had been defined, we conducted the local indicators of spatial  
205 association (LISA) analysis [43, 48] and outlier analysis [73, 74] to retrieve neighborhood metrics. Finally, two separate

206 supervised regression experiments were performed, in order to predict DBH based on LiDAR-derived metrics: one  
 207 including the neighborhood metrics (context-aware regression), the other without taking those metrics into account  
 208 (context-unaware regression). Finally, AGB was estimated from the predicted DBH via an allometric function (as  
 209 defined in Eq. 5).

210 In parallel, we conducted a second task to characterize the morphometry of tree assemblages (i.e., groups of  
 211 adjacent trees fulfilling a specific criterion of height similarity, as explained in Section 2.6) stemming from the ITC  
 212 polygon dataset. Prior to the morphometric analysis of tree assemblages, a second pre-processing task was conducted  
 213 (marked \*\* in Figure 3). First, ITC polygons were merged based on either local Moran's  $I_i$  [43] or  $SL_i$  [75] (see Section  
 214 2.4). These new larger polygons describe the two-dimensional projection of tree assemblages. Then, as our interest  
 215 focused on the extent and shape of the tree assemblages, the inner borders of the merged polygons were disregarded.  
 216 To reduce computation time, the polygon shapes were simplified by reducing the number of vertices and edges to 70 %  
 217 while keeping the polygon shape.



**FIGURE 3** Workflow followed in this study. PCD: point cloud data, CHM: canopy height model, ITC: individual tree crown, LISA: local indicators of spatial association, DBH: diameter at breast-height, AGB: above-ground biomass. The two colored boxes describe the subtasks constituting each of the processing steps, marked \* and \*\* in the diagram.

## 218 2.4 | Definition of Context Via Tree Neighborhood

219 We determined at what distance neighborhood metrics should be calculated (i.e., how many surrounding trees should  
 220 be accounted as neighbors) based on local similarity of tree height. Accordingly, the selection of an appropriate  
 221 neighborhood size around each individual tree (i.e., context detection) [76] was calculated through the analysis of  
 222 spatial autocorrelation as function of incremental distance. Based on the global peak in the significance of spatial  
 223 autocorrelation, we defined a characteristic distance within which all included trees should be considered as neighbors.  
 224 All so-defined neighbor trees were accounted for to compute context-aware metrics.

225 The local context information was encoded as metrics derived from the individual tree heights in each neighbor-  
 226 hood, calculated at each tree location. Specifically, the metrics computed to define tree neighborhoods were: local  
 227 Moran's  $I$  [43] clustering (i.e., an estimate of local significance of similarity with respect to global variance); and ( $SL_i$ )  
 228 of tree height (i.e., a weighted average of heights calculated entirely locally) [75].

## 2.5 | Neighborhood Information as Predictive Features

### 2.5.1 | Neighborhood Metrics

Local Moran's  $I_i$  is a well-established distance statistic in spatial data analysis [77], used for detecting local spatial autocorrelation and included within the family of LISA methods [48]. Like other methods [78], it relates attribute similarity with locational similarity, mapping autocorrelation across the geographic space. In the following definitions,  $\sigma$  is the global sample standard deviation of tree height;  $n$  and  $m$  represent the total number of instances (i.e., all trees in the forest) and the number of neighbors to each tree, respectively;  $y_i$  indicates the magnitude of interest at a particular point of interest (i.e., tree height) while the overline (i.e.,  $\bar{y}$ ) indicates global average;  $w_{i,j}$  indicates the distance weighting of each neighboring tree (here defined as inverse distance weighting); subindexes  $i$  and  $j$  indicate the tree of interest and a neighbor tree, respectively. Let  $y_1, \dots, y_n$  be the tree height values of all the  $n$  trees in the dataset. Then, the Local Moran  $I_i$  [43] is defined as

$$I_i = \frac{y_i - \bar{y}}{\sigma^2} \sum_{j \in \mathcal{N}_i, j \neq i} w_{i,j} (y_j - \bar{y}), \quad (1)$$

where  $\mathcal{N}_i \subset \{1, \dots, n\}$  is the set of indices corresponding to the nearest neighbors of tree  $i \in \{1, \dots, n\}$  in the overall set, with  $\sum_{j \in \mathcal{N}_i, j \neq i} w_{i,j} = 1$  and where

$$\bar{y} = \frac{1}{n} \sum_{i=1}^n y_i, \quad (2)$$

and

$$\sigma = \sqrt{\frac{\sum_{i=1}^n (y_i - \bar{y})^2}{n - 1}}, \quad (3)$$

are the global average height and the global sample standard deviation, respectively. It should be noted that insofar  $I_i$  includes global metrics (such as  $n$ ,  $\sigma$  and  $\bar{y}$ ), it is not entirely locally computed, but may present correlation with global features (i.e., characteristics derived from the entire dataset) [79]. The Spatial Lag ( $SL_i$ ) of tree height for a tree  $i$  is a spatial smoother [80] defined as:

$$SL_i = \sum_{j \in \mathcal{N}_i, j \neq i} w_{i,j} y_j \quad (4)$$

Therefore,  $SL_i$  can be seen as a weighted average of the attributes of neighboring trees [81]. The neighborhood metrics finally chosen as context-aware predictors are the following: local Moran's Index ( $I_i$ ), z-score of  $I_i$ , p-value of  $I_i$ , z-transformed value of  $I_i$  and  $SL_i$  - computed at 20 m, 30 m, 40 m and 50 m distance bands. Additionally, the mean heights of the k-nearest trees, with  $k \in (5 - 75)$ , were also included.

## 251 2.5.2 | Environmental Variables

252 We also included the topographic wetness index (TWI) [82] as environmental variable. TWI is a steady state wetness  
 253 index used to evaluate topography-dependent surface hydrology processes. According to [82], TWI is defined as  
 254  $\frac{a}{\tan(b)}$ , where  $a$  represents the upslope area draining through the point of interest, and  $b$  indicates the local slope. The  
 255 parameterization considered to calculate TWI followed the suggestions of Kopecký et al. (2021) [83] for estimating  
 256 soil moisture. In order to discern how much the contribution of TWI is influenced by granularity, we calculated it at  
 257 a 2  $m^2$  resolution, and resampled to 5 and 10  $m^2$ , via bilinear interpolation. Therefore, TWI was included at a spatial  
 258 resolution of 2, 5 and 10  $m^2$  as separate predictors.

## 259 2.6 | Tree Assemblages: Definition and Morphometry

260 In order to define tree assemblages, local Moran's  $I_i$  and  $SL_i$  were both computed at the optimal distance band to  
 261 obtain neighborhood metrics, i.e., based on the global peak in the significance of spatial autocorrelation as a function  
 262 of distance (using ArcGIS Pro software [84]). Tree assemblages were therefore defined as geographically continuous  
 263 groups of trees delineated according to either (i) variation of local Moran's  $I_i$  of tree height, or (ii) according to quantiles  
 264 of  $SL_i$  of tree height. The rationale for using two different statistics to calculate tree neighborhood metrics and thus  
 265 delineate different tree assemblages was that while  $SL_i$  is entirely locally calculated, local Moran's  $I_i$  includes global  
 266 features (and is therefore sensitive to the statistical characteristics of the dataset as a whole), as explained in Section  
 267 2.5.1. In order to discern which of the two approaches seemed most convenient in delineating tree assemblages (the  
 268 former *entirely* local; the latter only *partially* local), both were included.

269 Tree assemblages defined according to local Moran's  $I_i$  are geographically continuous groups of trees with signif-  
 270 icantly different heights than the global tree height average, and they also lie in a region with significantly different  
 271 neighbors. Local Moran's  $I_i$  identifies regions where the clustering of either high or short trees occurs. In the standard  
 272 notation [75] (i.e., *High-High* or *Low-Low*), the first term refers to the individual tree and the second to the neighborhood  
 273 (e.g., a tree belonging to a *High-High* assemblage is a "significantly high tree" in a "significantly high neighborhood").  
 274 The areas not showing statistical significance ( $p\text{-value} \geq 0.002$ ) were labeled as *Not-Significant*. The significance test  
 275 is based on random permutations ( $n = 499$ ) of neighboring tree-height values at each step in the computation. Then,  
 276 for every permutation, a local Moran's  $I_i$  value is calculated by randomly rearranging the tree heights of neighbor-  
 277 ing values. The result is a randomly generated reference distribution of expected local Moran's  $I_i$  that is compared  
 278 against the observed local Moran's  $I_i$  (Eq. 1) [48]. In this way, tree assemblages defined according to local Moran's  $I_i$   
 279 are classified as: *High-High*, *Low-Low*, or *Not-Significant*.

280 Likewise, tree assemblages defined according to  $SL_i$  of tree height are geographically continuous groups of trees  
 281 delimited according to the local weighted average of tree height [81], as defined above (Eq. 4). For the purpose of  
 282 this study, 5 subdivisions based on quantiles were deemed convenient, rendering a classification of tree assemblages  
 283 based on  $SL_i$  ranking as: *Highest*, *High*, *Mid*, *Low* and *Lowest*.

284 The morphometric analysis used as its objects of analysis the outer boundaries of tree assemblages, defined  
 285 either by local Moran's  $I_i$  or  $SL_i$  of tree height, as defined above. Twenty basic morphometric parameters as well  
 286 as 20 derived parameters were calculated for each type of tree assemblage. The 20 basic morphometric variables  
 287 are simple parameters obtained by fitting elemental geometric shapes to each tree assemblage polygon (e.g., area  
 288 of maximum inscribed circle), and basic positional parameters (e.g., XPOL, which is the X coordinate of the centroid  
 289 of the tree assemblage polygon). The 20 derived parameters are adimensional metrics (except for concavity [85],  
 290 measured in m) computed from the 20 basic morphometric variables, as explained in [86] (a full description of the 40

291 morphometric parameters is given in Annex I). The morphometric analysis of tree assemblages was conducted using  
 292 PolyMorph-2D algorithm [86], which is a toolbox for the morphometric analysis of vector-based polygon objects,  
 293 available as a plug-in for the open source JumpGIS software [87].

## 294 2.7 | Regression Models Selected

295 The regression experiments were designed to predict DBH, since AGB is a variable determined by the combination of  
 296 DBH, height and wood density. The AGB estimates were derived from the DBH prediction outputs by means of an  
 297 allometric fit (Eq. 5). Predicting DBH, instead of AGB directly was chosen as more suitable, as it avoids burdening the  
 298 learning models with the statistical error contained in the allometric fit. Several feature-based regression methods  
 299 were selected: namely AdaBoost [88, 89], Lasso [90] and Random Forest [91] regressors.

300 The AdaBoost regressor [92] is a gradient-boosting method based on stage-wise additive expansions; its effective-  
 301 ness rests on the combination of weak learners (i.e., decision trees) to produce a generalized prediction hypothesis.  
 302 Lasso is a linear model with  $L1$  prior penalty as a regularizer [93], while Random Forest is a tree-based ensemble re-  
 303 gression method. In our case, all three feature-based methods take as input the features derived from the ITC polygon  
 304 dataset resulting from the CHM segmentation.

305 Context-unaware regressions are defined as those in which a learning model performs DBH regression by taking  
 306 as predictors only individual tree attributes derived from the ITC polygon dataset (i.e., tree height, canopy area and  
 307 canopy perimeter), as it is a common approach [28]. We defined context-aware regressions as those regressions in  
 308 which context-aware features are additionally introduced as input. These come in the form of either neighborhood  
 309 metrics, e.g.,  $SL_i$  of tree height, or as environmental variables, i.e., TWI at different spatial resolutions. Both the  
 310 neighborhood metrics and environmental variables used are described in Section 2.5.1 and Section 2.5.2, respectively.  
 311 For every model predicting DBH from individual tree attributes (i.e., context-unaware conditions) we implemented a  
 312 context-aware counterpart.

## 313 2.8 | Training, Validation and Test

314 A hard validation of AGB is not possible without harvesting trees destructively, which raises obvious ethical, legal and  
 315 economic issues. Thus, non-invasive methods that use RS data and allometric functions are the standard procedure  
 316 for estimating AGB [18]. Here, we chose two variables to validate our predictions: (i) DBH, a key morphological trait  
 317 contained in the field-based forest inventory; and (ii) tree-level AGB estimates derived via species-specific allometric  
 318 and wood density functions. Specifically, the allometric model used was the one proposed by Dalponte and Coomes  
 319 (2016) [94]:

$$AGB_{tree} = \alpha \cdot WD_{spruce}^{\beta} \cdot (DBH - d_0)^{\gamma} \cdot H^{\delta}, \quad (5)$$

320 where the wood density value ( $WD_{spruce}$ ) was taken from Alpine spruce dendrometric models [95]; diameter  
 321 at breast-height ( $DBH$ ) and height ( $H$ ) are allometric measurements, while  $\alpha$ ,  $\beta$ ,  $\gamma$ ,  $\delta$  and  $d_0$  are species-specific fitted  
 322 parameters [96]. The AGB assessment was derived from the predictions of DBH (and LiDAR-derived height) in ei-  
 323 ther aware or unaware conditions. Therefore, the predicted value of DBH was input into Eq. 5, in order to obtain  
 324 predictions of AGB. This allowed to compare AGB predictions to the ground-truth values of AGB, which were simi-  
 325 larly obtained via the field-based measurements (provided by the regular tree-monitoring campaigns of ICOS [63] and

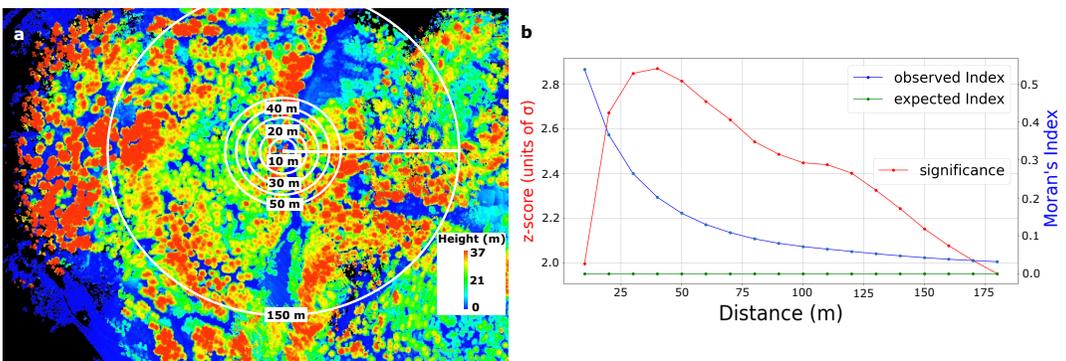
326 WSL [60]) and Eq. 5.

327 For training and validating the regression models, the instances with empty ground-truth labels were initially re-  
 328 moved (i.e., trees with no DBH or tree height recorded). Afterwards, data stratification was done via five commonly  
 329 used percentiles (i.e., 0-10, 10-25, 25-50, 50-75, 75-90, 90-100) to ensure that input data is independently drawn  
 330 from an identical sample distribution (i.i.d. assumption) [97]. This assured us that most parts of the target distribution  
 331 are represented, in particular the tail ends. Then, the technique used to estimate model prediction error consisted of  
 332 a nested cross-validation (NCV) [98]. Following the NCV scheme, we divided the input dataset (either CP, or SP, corre-  
 333 spondingly) into 10 inner and 10 outer folds. The inner cross-validation was used for hyperparameter optimization and  
 334 feature selection, while the outer cross-validation was used to evaluate model performance (the method description  
 335 is extended in Section 4.4 and further details are given in Annex III). The significance of the enhancement in context-  
 336 aware predictions and effect size was assessed using Wilcoxon signed-rank test [99] and Cliff's Delta analysis [100],  
 337 respectively.

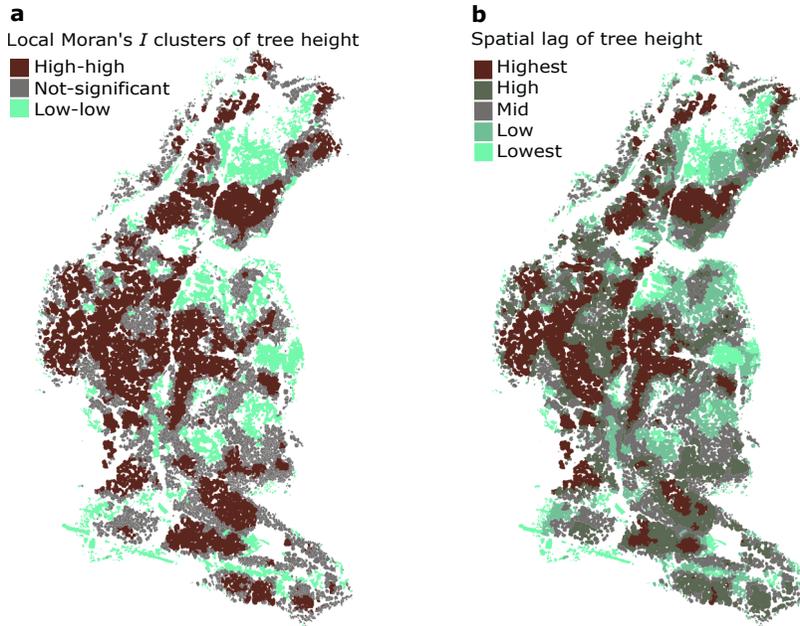
### 338 3 | RESULTS

#### 339 3.1 | Context Detection and Tree Assemblage

340 The selection of the specific distance for computing tree neighborhood metrics was calculated based on the degree  
 341 of spatial autocorrelation of tree height by incremental distance, as in previous studies [101]. This resulted in a global  
 342 maximum at a distance of 40 m. Figure 4 (a) shows the calculation of local Moran's index ( $I_i$ ) of tree height at different  
 343 distance bands. Figure 4 (b) shows the z-score of  $I_i$  obtained at each distance band, resulting from comparing the  
 344 observed  $I_i$  and the expected  $I_i$  under the tree height randomness assumption (details included in the Annex II). As  
 345 a sanity check, we ran context-aware regression experiments including context features retrieved at shorter (i.e., 20  
 346 m, 30 m) and larger (i.e., 50 m) distances than the optimal range (i.e., 40 m). The context features retrieved at these  
 347 distances and contributing to improved predictions of DBH were also included in the final regression models.



**FIGURE 4** Context detection. **a**: normalized point cloud data (PCD) scene colored by tree height overlaid with a selection of the appropriate radii for defining the neighboring context. **b**: Autocorrelation of tree height as function of distance. The red line shows the number of standard deviations ( $\sigma$ ) that an observation is away from the expected value (under the assumption of heights being randomly distributed). The blue and green lines show the actually observed local Moran's Index and the expected value under randomness assumption, respectively.

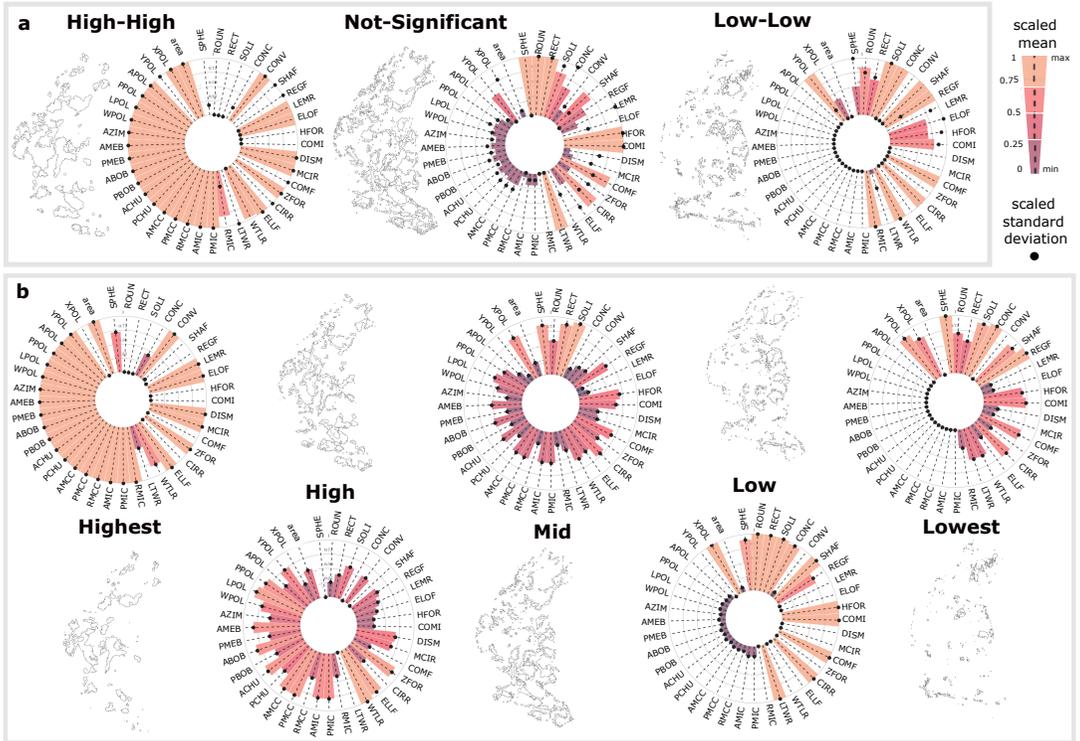


**FIGURE 5** Tree assemblages defined by local similarity of tree height. **a:** delineated according to local Moran's  $I_i$  of tree height; **b:** delineated according to spatial lag of tree height.

348 Figure 5, a and b, show the spatial distribution of different tree assemblages defined by local Moran's  $I_i$  and by  
 349  $SL_i$  of tree height respectively. While both types of assemblages show similarities as regards extent and location,  
 350  $SL_i$  captures more local variability. This is not only due to a higher discretization (5 groups in  $SL_i$ , vs. 3 groups in  
 351 local Moran's  $I_i$ ), but also to the fact that  $SL_i$  is insensitive to the variance in the dataset beyond the range of its  
 352 neighborhood, as explained in Section 2.5.1 (in Figure 5, both assemblage types shown in Figure 5 were derived from  
 353 these two metrics, calculated at 40 m range).

354 Figure 6, panels a and b, show the results of the morphometry analysis of tree assemblages defined by local  
 355 Moran's  $I_i$  and by  $SL_i$  respectively. The results are based on the shape of the outer contours of the resulting tree  
 356 assemblages. The circular barplots show the average magnitude as bar lengths, and the standard deviation as dots.  
 357 Both mean and standard deviation values are shown as min-max scaled (across assemblage types) to present all vari-  
 358 ables on the same radial axis and to ease visual comparison, i.e., for every morphometric variable, the highest value is  
 359 replaced by 1, the minimum is replaced by 0, and the intermediate values are linearly interpolated between 0-1.

360 While not for all variables a systematic trend was found, for several basic morphometric variables a clear positive  
 361 correlation between them and  $SL_i$  was observed. This is the case for polygon area, perimeter of polygon (PPOL) and  
 362 radius of the minimum circumscribed circle (RMCC). Additionally, a positive correlation was found for some derived  
 363 morphometric variables, namely: length-to-width ratio (LTWR) [102], circularity ratio (CIRR) [105], compactness factor  
 364 (COMF) [86], dispersion measure (DISM) [105], complexity index (COMI) [86], lemniscate ratio (LEMR) [109], regularity  
 365 factor (REGF) [104], and concavity (CONC) [85]. Conversely, other morphometric variables showed a decreasing trend  
 366 with increasing  $SL_i$ . A negative correlation between  $SL_i$  and the following derived morphometric variables was found:  
 367 Miller's circularity ratio (MCIR) [107], Horton's form factor (HFOR) [102], elongation factor (ELOF) [108], shape factor  
 368 (SHAF) [104], convexity [110], solidity [111], rectangularity (RECT) [112] and roundness (ROUN) [110].



**FIGURE 6** Morphometric analysis of tree assemblages grouped by (a) local Moran's  $I_i$ , and (b) by spatial lag of tree height. Bar length and color gradient represent the mean value, while black dots represent the standard deviation (SD) over all tree assemblages. Both mean and SD are scaled (min-max) to allow comparison of all metrics across assemblage types on the same axis (i.e., for every morphometric variable, the highest value of a certain assemblage type is replaced by 1, the minimum value is replaced by 0, and the intermediate values are linearly interpolated in between the range (0-1)). YPOL: northing of centroid of the tree assemblage; XPOL: easting of centroid of the assemblage; APOL: area of polygon (P); N-S: defined as  $|\sin(\text{azimuth})|$ , shows the alignment of the main axis of P with the North-South direction; PPOL: perimeter of P; LPOL: major axis length (L) of P; WPOL: minor axis length (W) of P; ABOB: area of the bounding box fully containing P; PBOB: perimeter the bounding box fully containing P; AMEB: area of the minimum enclosing box fully containing P; PMEB: perimeter of the minimum enclosing box fully containing P; ACHU: area of containing hull ; PCHU: perimeter of convex hull fully containing P; AMCC: area of the minimum circumscribed circle (MCC); PMCC: perimeter of MCC; RMCC: radius of MCC; AMIC: area of maximum inscribed circle (MIC); PMIC: perimeter of MIC; perimeter of MCC; RMIC: radius of MCC; LTRW: length-to-width ratio [102]; WTLR: width-to-length ratio [103]; ELLF: ellipticity factor [104]; CIRR: circularity ratio [105]; ZFOR: Zavoianu's form factor [106]; COMF: compactness factor [86]; MCIR: Miller's circularity ratio [107]; DISM: dispersion measure [105]; COMI: complexity index [86]; HFOR: Horton's form factor [102]; ELOF: elongation ratio [108]; LEMR: lemniscate ratio [109]; REGF: regularity factor [104]; SHAF: shape factor [104]; CONV: convexity [110]; CONC: concavity [85]; SOLI: solidity [111]; RECT: rectangularity [112]; ROUN: roundness [110]; SPHE: sphericity [113].

369 The correlations between local Moran's  $I_i$  and morphometric variables followed the same trends as for  $SL_i$ . An  
 370 observed difference between  $SL_i$  and local Moran's  $I_i$  was found in the heteroscedasticity of the morphometric  
 371 variables calculated. In the former case, we observed that the variance of all metrics scaled with magnitude (i.e.,

372 constantly increasing variance), while in the latter an irregular trend was found (i.e., an uneven trend in the variance).  
373 We visualized these observations in the distribution of scaled mean values and scaled standard deviations in Figure 6,  
374 a and b.

### 375 3.2 | AGB Predictions: Aware vs. Unaware of Local Context

376 Regression experiments including context-aware features improved predictions of DBH consistently (see Tables 1  
377 and 2), resulting in spatially resolved enhanced tree-level AGB predictions via allometry (Eq. 5). Although consistent,  
378 the degree of prediction enhancement differed between both datasets considered. Predictions in the CP-dataset  
379 observed a lower enhancement in comparison to predictions in the SP-dataset. For instance, RMSE was reduced  
380 by 9.1% (SP-dataset) vs. 4.0% (CP-dataset), and R2 increased by 3.5% (SP-dataset) vs. 3.2% (CP-dataset). This was  
381 expected, due to less variability in context in the CP-dataset.

382 In Figure 7, the left panel (a) shows the ground-truth labels (i.e., field based estimates of AGB), which were derived  
383 from the field measurements and a species-specific allometric fit (i.e., Eq. 5). The central panel (b) shows the spatial  
384 distribution of residuals (i.e.,  $\epsilon = AGB_{ground-truth} - AGB_{prediction}$ ) of the AdaBoost context-aware regression results.  
385 The mean values converge towards zero (i.e.,  $\bar{\epsilon}_{SP} = 3.8$  kg,  $\bar{\epsilon}_{CP} = -3.2$  kg), while the spread of the error distribution  
386 varies between SP and CP datasets (i.e.,  $\sigma(\epsilon_{SP}) = 123$  kg,  $\sigma(\epsilon_{CP}) = 140$  kg).

387 The lack of high spatial autocorrelation of errors (i.e., low clustering of errors) indicates that predictions are not  
388 geographically biased. The upper-right panel (c) displays the error distributions in both datasets. SP-errors show a uni-  
389 modal distribution with a slight overestimation of DBH of -28 mm (i.e., overestimation). CP-errors present a similar  
390 overestimation bias (-25 mm) with a bimodal distribution (the second mode is located at 25 mm of underestimation).  
391 The two bottom-right panels show the error distribution of DBH predictions along the ground-truth measurements  
392 of DBH and tree height, respectively. It can be observed that, generally, smaller and thinner trees tend to be slightly  
393 overestimated (i.e., in the first two quantiles), while the largest trees (i.e., quantile 5 and highest trees) tend to under-  
394 estimation.

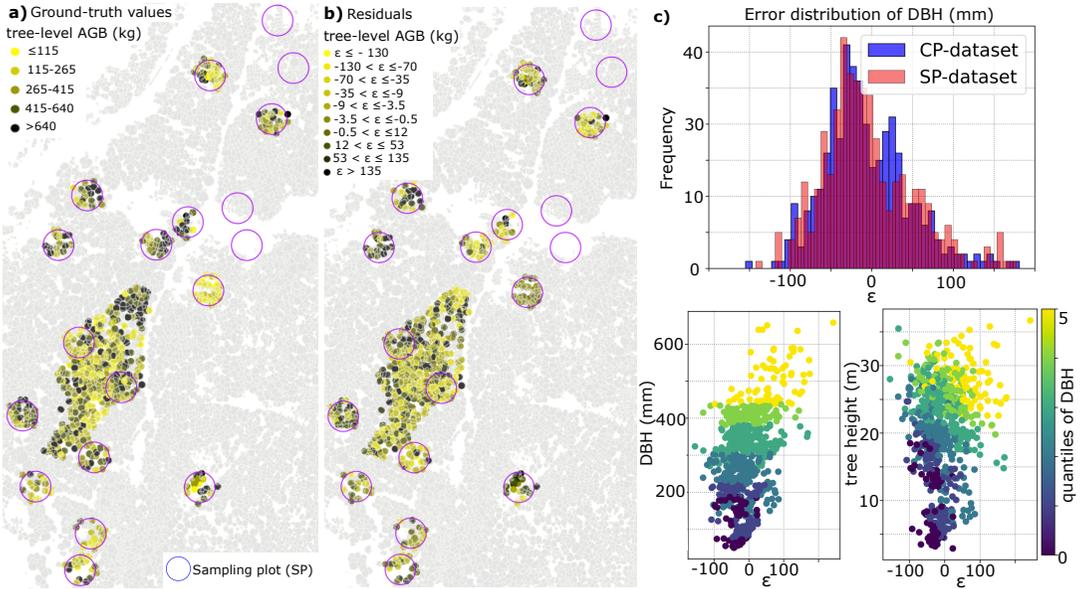
395 Figure 8 presents a detailed analysis of the relative importance of all predictors considered in the context-aware  
396 DBH regression with the AdaBoost regression model. We used the permutation importance inspection technique as  
397 proposed by Altmann et al. (2010) [114]. The analysis reveals that in both SP and CP datasets, the most important  
398 context-aware predictors are the average heights of the 5, 10, and 15 nearest neighboring trees, outperforming some  
399 individual-tree metrics, such as the crown metrics.

**TABLE 1** Results (on test set) of the sampling plot (SP) dataset. Predictor variables are LiDAR-derived features; target variable is diameter at breast-height (DBH, in mm). The values are presented as mean  $\pm$  standard deviation of the 10 outer CV folds. One asterisk (\*) marks results where the enhancement introduced by context-awareness is statistically significant with "small" size effect, while \*\* and \*\*\* mark "medium" and "large" size effect, respectively. The best results are shown in bold.

Regression model	$R^2$	RMSE (mm)	MAE (mm)	MAPE (%)
AdaBoost (unaware)	0.830 $\pm$ 0.05	58.0 $\pm$ 9.0	43.3 $\pm$ 4.4	19.1 $\pm$ 1.9
AdaBoost (aware)	<b>0.860 <math>\pm</math> 0.03 ***</b>	<b>52.7 <math>\pm</math> 5.3 ***</b>	<b>41.0 <math>\pm</math> 3.1 **</b>	<b>19.5 <math>\pm</math> 1.7</b>
Random Forest (unaware)	0.818 $\pm$ 0.04	60.2 $\pm$ 7.3	46.8 $\pm$ 4.5	22.8 $\pm$ 5.8
Random Forest (aware)	0.838 $\pm$ 0.05 *	56.5 $\pm$ 9.2 *	41.6 $\pm$ 5.4 ***	22.4 $\pm$ 5.1
Lasso (unaware)	0.851 $\pm$ 0.02	54.6 $\pm$ 4.9	4.20 $\pm$ 3.3	19.1 $\pm$ 1.4
Lasso (aware)	0.852 $\pm$ 0.02	54.4 $\pm$ 4.9	4.17 $\pm$ 3.5	18.8 $\pm$ 1.7

**TABLE 2** Results (on test set) of the control plot (CP) dataset. The predictive variables are LiDAR-derived features; the target variable is diameter at breast-height (DBH, in mm). The values are presented as mean  $\pm$  standard deviation of the 10 outer CV folds. One asterisk (\*) marks results where the enhancement introduced by context-awareness is statistically significant with "small" size effect. The best results are shown in bold.

Regression model	$R^2$	RMSE (mm)	MAE (mm)	MAPE (%)
AdaBoost (unaware)	0.713 $\pm$ 0.07	54.7 $\pm$ 5.98	43.0 $\pm$ 5.26	15.5 $\pm$ 2.4
AdaBoost (aware)	0.737 $\pm$ 0.05 *	52.9 $\pm$ 5.28 *	42.2 $\pm$ 4.43 *	15.7 $\pm$ 3.1
Random Forest (unaware)	0.688 $\pm$ 0.07	57.0 $\pm$ 5.9	43.8 $\pm$ 5.1	15.7 $\pm$ 3.1
Random Forest (aware)	0.705 $\pm$ 0.04	55.6 $\pm$ 5.3	41.3 $\pm$ 5.5 *	15.9 $\pm$ 4.3
Lasso (unaware)	0.741 $\pm$ 0.09	51.3 $\pm$ 6.6	39.1 $\pm$ 5.2	13.6 $\pm$ 1.6
Lasso (aware)	<b>0.750 <math>\pm</math> 0.08</b>	<b>50.4 <math>\pm</math> 5.9</b>	<b>38.6 <math>\pm</math> 4.1</b>	<b>13.6 <math>\pm</math> 1.1</b>



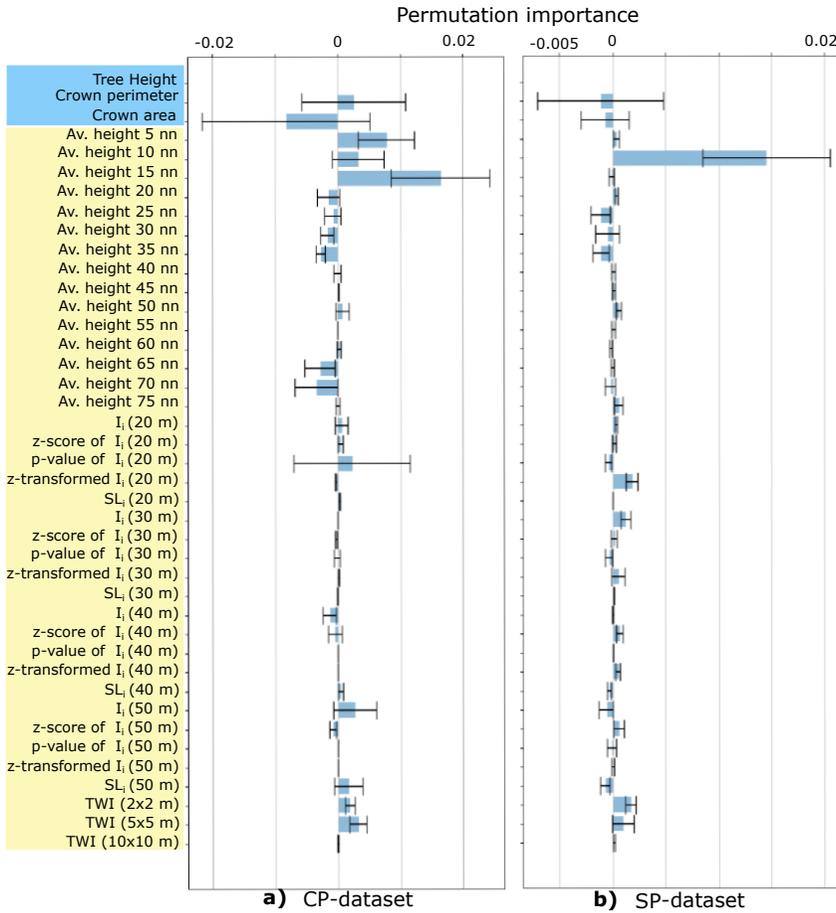
**FIGURE 7** a: spatial distribution of tree-level above-ground biomass (AGB) according to ground-truth measurements (provided by the tree-monitoring campaigns of ICOS [63] and WSL [60]) and Eq. 5, grouped by quantiles. b: spatial distribution of residuals ( $\epsilon = AGB_{ground-truth} - AGB_{prediction}$ ) of AGB predictions with AdaBoost context-aware regression, grouped by quantiles. Negative values indicate overestimation. The empty SP-plots correspond to areas where the quality of the UAV-LiDAR data collection was compromised. c: error distributions of diameter at breast-height (DBH) in sampling plot (SP) and control plot (CP) datasets. The two bottom-right panels show the error distribution of DBH (in x-axis) vs. field-measurements of DBH and tree height.

401 **4 | DISCUSSION**

402 **4.1 | Enhancement of Tree-Level AGB Prediction**

403 This study presents a method of enhancing tree-level AGB estimates in forests using UAV-LiDAR surveying and  
 404 context-aware ML regression methods. The results consistently showed that context-aware regressions outperformed  
 405 context-unaware regressions across models. This finding indicates that gradients in morphological tree traits across  
 406 the ecosystem may be a proxy for unveiled environmental and biotic factors (e.g., windstorm disturbance, nutrient and  
 407 soil moisture abundance, light harvesting competition [44, 45]) that influence tree growth, which can be leveraged to  
 408 enhance predictions of AGB.

409 The accuracy enhancement gained from including context-aware features in the regression experiments varied  
 410 between the two datasets considered (i.e., SP-trees and CP-trees). Context-aware regressions of DBH in SP-trees  
 411 experienced greater enhancement than CP-trees. This is consistent with the fact that the CP-dataset contains less  
 412 variability of context, since it is a clustered and more homogeneous dataset, while the SP-dataset includes more  
 413 variability in context-aware features. The investigated mono-specific forest presents a heterogeneous landscape,  
 414 where the distribution of tree heights varies in space. Hence, the UAV-LiDAR survey gives rise to a non-stationary  
 415 tree dataset [76], showing both smooth gradients and sharp changes in height values, a non-trivial question in tree-  
 416 phenotyping and species mapping [50]. As SP-trees are grouped in scattered plots across the forest, their spatial



**FIGURE 8** Inspection of predictors' importance via the permutation method [114] in AdaBoost regression experiment in context-aware conditions. The left panel (a) shows results in the control plot (CP)-dataset, and the right panel (b) shows results in the sampling plot (SP)-dataset. Bar length and error bar show the mean and standard deviation of a predictor's importance, respectively. A negative mean value indicates that a predictor is less useful than when being randomly shuffled, so it lowers the model's predicting performance. Predictors highlighted in light blue are individual tree traits; predictors highlighted in light yellow are context-based (i.e., either neighborhood metrics or environmental variables). In both datasets, it can be noted how the average heights of the nearest neighbors (nn) stand out as the strongest context-based predictors. In both plots (a and b), individual tree height has been removed to ease visual comparison of the remaining predictors.

417 distribution spans hundreds of meters, making them subject to a more diverse context than the very local CP-dataset.

## 418 4.2 | The Role of Neighboring Context in AGB Prediction Performance

419 All regression models achieved enhanced predictions when contextual information was considered. Thereby, the  
 420 degree of local similarity of tree height (i.e.,  $SL_i$ , local Moran's  $I_i$ ) was most important and, to a lesser extent, environ-  
 421 mental variables (i.e., TWI).

422 Conversely, including features informing about neighbor dissimilarity, such as local outliers detected using Lo-  
423 cal Outlier Factor [73] and Isolation Forest [74] algorithms did not result in enhanced predictions. We hypothesize  
424 that metrics containing information about the degree of local similarity may reveal the combined effect of ecological  
425 processes that are specific to the immediate neighboring context. Conversely, metrics containing information of the  
426 dissimilarities of the individual trees do not help to uncover such processes, but remain useful in detecting outstanding  
427 trees (i.e., local outliers).

428 Context-based features at closer distances generally showed larger predictive power but also larger variance (as  
429 less neighboring trees are computed), therefore producing a strong and fluctuating signal, that in some cases was  
430 challenging for the ML model to incorporate in the learning process. For instance, the p-value of Local Moran's  $I_i$   
431 at a 20 m range in the CP-dataset has an average positive effect but is not a stable predictor (Figure 8, a). This can  
432 be observed in the general trend of larger standard deviations in the permutation importance of predictors retrieved  
433 at short ranges than at greater distances (Figure 8). After the peak in the spatial autocorrelation of tree height (i.e.,  
434 at larger distance bands than 40 m), the significance of clustering of tree height values declined, presenting another  
435 shoulder at a distance of 110 m (Figure 4, b). As the neighborhood size increased beyond the 40-meter distance  
436 range, the predictive power of the metrics derived from the neighboring trees (i.e., the influence of local context)  
437 progressively smoothed down [80].

438 In accordance with competition-based studies [32, 33, 34], we observe that the strongest context-based predic-  
439 tors are those retrieved from the immediate neighboring trees in both datasets, i.e., the average height of 5, 10 and  
440 15 nearest neighbors (Figure 8). This observation indicates that individual tree structural traits are primarily medi-  
441 ated by competition mechanisms. However, our method additionally allows to compare the relative importance of  
442 competition-derived metrics and other context-based metrics operating at larger scales. For instance, in Figure 8 (a)  
443 it is shown that local Moran's  $I_i$  retrieved at a 50 m range is comparable in importance to the average height of the  
444 closest 10 neighboring trees.

445 A general difference observed between the CP and the SP dataset is that the predictors' importances in the CP-  
446 dataset fluctuate more (i.e., larger standard deviations). Further, in the SP-dataset, predictors rarely become negative  
447 and if they do, it is to a lesser extent. Given its broader spatial distribution and greater contextual variability, we  
448 contend that the SP-dataset can be regarded as a more representative sample of the entire forest population compared  
449 to the clustered CP-dataset. Consequently, the finding that context-based features demonstrate greater stability  
450 within the SP-dataset is noteworthy.

451 In relation to the environmental metrics used, TWI exhibited a greater impact on improved predictive performance  
452 at finer spatial resolutions in both datasets (Figure 8), whereas its contribution decreased at coarser resolutions (e.g., it  
453 did not significantly contribute as a predictor at 10 m resolution). This observation indicates that the spatial resolution  
454 at which TWI is most informative of individual tree traits, is similar to the usual tree crown size (i.e., 2-5 m resolution),  
455 while at larger scales its contribution as predictor becomes negligible.

456 The morphometric analysis (Figure 6, a and b) provided 40 additional features that were evaluated as potential  
457 predictors of DBH. However, including morphometric variables calculated from the tree assemblages in the regression  
458 experiments—either defined by  $SL_i$  or by local Moran's  $I_i$ —did not result in improved predictions of DBH. As shown  
459 in Figure 5, the shape of tree assemblages shows sensitivity to the method used. The shapes of tree assemblages  
460 indicated a trend of convergence assembly patterns at the group level [115], as discussed in Section 4.3. Nevertheless,  
461 the group morphometry did not prove useful to improve predictions of DBH.

462 Including context to enhance estimates of structural traits at the individual tree level has previously been pro-  
463 posed in seminal works [36] and been adopted subsequently for various applications in forest research [35, 37]. Lo  
464 and Lin (2012) [34] proposed a competition-specific index to capture the effect of the competing pressure of imme-

465 diate neighbors. More recent research conducted in this area [32, 33] has motivated the further development of  
466 competition-aware approaches to improve the prediction accuracy of individual tree traits, using overstory tree traits  
467 as predictors, such as tree height and crown metrics, which enables the potential transferability of these methods to  
468 a RS framework.

469 In forest biomass research, a commonly recognized approach is calibrating regression models with plot-level met-  
470 rics for predicting tree-level structural traits (e.g., parameters accounting for plot-level random effects in NLME meth-  
471 ods). However, such approaches do not question the influence of the artificially-delineated plot size on prediction  
472 enhancement, even if it is observed that accuracy increases with plot width and number of tree neighbors [29, 31].  
473 Furthermore, how diverse context-based attributes retrieved at different distance ranges affect tree-level predictions  
474 had not been investigated before. In this regard, our results show that the variability and extent of context determines  
475 its beneficial leverage for prediction of tree-level structural traits.

476 Our study continues this line of work and sheds light on how the local spatial context can be defined and leveraged  
477 in tree-level structural trait predictions (i.e., DBH), making a case for AGB estimates. The analysis shows that there  
478 is an optimal range to computing neighborhood metrics. In the case of the monoculture forest studied here, this  
479 corresponded to a 40 m range distance, based on the spatial autocorrelation of tree heights. Further, we found  
480 that the predictive power of context-based metrics is sensitive to context extent (i.e., the distance at which such  
481 metrics are calculated). This observation indicates that considering context based on plot-level metrics retrieved from  
482 artificially bounded units (plot-level metrics, as in [29, 30, 31]) may be seen as a suboptimal approach [116]. Likewise,  
483 in the light of this observation, and in line with recent studies [117], determining the significant contextual extent  
484 of individual functional traits based on fixed pixel-size [118] appears to be a subpar technique. Therefore, future  
485 forest research would probably benefit from including context-awareness determined by spatial association of tree  
486 traits, bearing in mind that context-detection is trait-dependent and may vary depending on dataset source (e.g.,  
487 spatial autocorrelation as a function of distance is sensitive to CHM segmentation quality) and method applied (e.g.,  
488 delineation of tree assemblages varied slightly between local Moran's  $I_i$ , and  $SL_i$ , as we show in Figure 5, a and b).  
489 The motivation for our study has been to introduce more quantifiable terms to ecological reasoning and to propose a  
490 standardized method of incorporating context-awareness into AGB research. The method proposed is conceived for  
491 a RS framework. Since we do not make use of external data sources but, on the contrary, every predictor is native to  
492 the UAV-LiDAR dataset, it is readily transferable.

493 Lastly, we note that RS studies usually define the optimal scale of analysis as a trade-off between the observational  
494 extent (i.e., area surveyed) and the unit resolution (i.e., pixel size) [117]. Also, in ecological research, it is common  
495 to subsample datasets using natural subregions based on ancillary ecological criteria (i.e., ecoregions, conservation  
496 status) [4]. Conversely, here we defined the range of influence of context-based metrics (i.e., the boundaries of tree  
497 neighborhoods) using a dataset-native approach, based entirely on the spatial association of individual tree traits.  
498 This permitted us to determine the context of influence unhampered by the RS technique and not using external data  
499 sources. In computer vision studies that investigate contextual learning, image analyses typically do not assume a  
500 specific optimal scale [119, 120], such as in geographic analysis [121]. In this study, local context was defined based  
501 on the spatial association of a real physical attribute of the target objects (i.e., tree height), and not defined by an  
502 artificially bounded unit (e.g., pixel size [118] or plot size) so that the resulting distance (i.e., 40 m) could be considered  
503 informative of the forest ecosystem.

### 4.3 | Tree Assemblages

The quantitative comparison of morphometric variables between tree assemblages (Figure 6) permitted to examine whether trees—grouped by local association of an individual trait—persistently show different shapes at the group level, shedding light on the relationship between context-based traits and individual tree traits. Remarkably, it was observed that tree assemblages delineated according to the weighted average of individual tree heights (i.e.,  $SL_i$ ) presented positive correlations with two-dimensional morphometric features at the group level.

For instance, assemblages with higher trees (i.e., labeled as *Highest* according to  $SL_i$ , or *High-High* according to local Moran's  $I_i$ ) are consistently rounder, larger and more regular in shape. As visualized in Figure 6,  $SL_i$  correlates positively with shape regularity [104], two-dimensional concavity [85], length-to-width ratio [102] and size, indicating a consistent trait-convergence assembly pattern [115]. Higher trees seem to converge in most sheltered areas (i.e., thalwegs and local sub-basins) so that tree assemblages with highest  $SL_i$  tend to adopt the morphological features of the drainage network's shape (see Figure 9, in Annex I). Interpretation of this observation would go beyond the scope of this study. However, it may indicate that both the shape and location of tree assemblages of different heights are conditioned by underlying environmental and biotic driving mechanisms.

In the monoculture forest studied here, tree height clustering occurs (Figure 5, a), while spatial gradients of averaged tree height present preferential shapes and directions (Figure 5, b). These observations indicate that there is tree-height convergence and a tendency toward optimal phenotype expression (i.e., maximum growth performance) around the runoff drainage network (Figure 9, c, in Annex I). Higher trees are found in sheltered regions and concave channels—which generally benefit from more frequent runoff events and deeper soils [46, 47]. This may indicate that favorable environmental conditions (e.g., deeper soil, lower soil moisture-recession rates, greater availability of soil nutrients due to leaching) allow individuals to reach their optimal phenotype. Conversely, a lower  $SL_i$  of tree height in more exposed terrain (e.g., ridges, hilltops) indicates that environmental filtering (e.g., windstorm disturbance) or a reduced competition in light harvesting could play a significant role in determining the location of low  $SL_i$  tree assemblages (Figure 9, a, in Annex I). Thus, the relatively reduced tree height in exposed areas could indicate a passive response of tree height to harsher environmental conditions [49], an active response to higher light availability [44] or a limitation to tree growth caused by other local factors, such as lower soil depth or nutrients availability [1, 47]. Nevertheless, this study cannot provide an interpretation of such observations, as shifts in the variance of functional traits across environmental gradients (i.e., spatial patterns of trait similarity) do not bring strong evidence of either biotic or environmental filtering on their own [122].

### 4.4 | Methods Applied

The regression methods used (i.e., AdaBoost, Lasso and Random Forest regressors) are well-known methods that take as input features extracted from the polygon dataset obtained after CHM segmentation, abstracted from their spatial location (see Figure 3).

The NCV technique [98], used for model optimization and evaluation, follows the updated, most established recommendations to achieve an unbiased estimate of the generalization error, while making optimal use of the limited available data. The results in the inner folds report on training performance, as they are used for model optimization, while the mean performance on the outer folds is the one used for model evaluation. As a modification developed from standard cross-validation [123], NCV improves estimates of prediction accuracy and confidence intervals by accounting for the correlation between error estimates in different folds, an inconvenient phenomenon affecting standard cross-validation that may render error estimates overly optimistic (further details of how the NCV algorithm

544 is implemented are given in Annex III).

545 The inspection technique used to evaluate predictors' influence on the DBH regression results was the permu-  
546 tation importance method [114]. The feature-elimination procedure consisted of eliminating progressively those pre-  
547 dictors that presented a negative mean importance, as they were considered harmful to the model's performance.

548 In order to evaluate the statistical significance of the enhancement introduced by context-awareness, we used  
549 the Wilcoxon signed-rank test [99], while for the assessment of effect size we used the Cliff's Delta analysis [100].  
550 These two tests were conducted in the same 10 outer folds of the NCV routine (i.e., test data) in aware and unaware  
551 conditions, so that results were compared using the exact same test data folds.

## 552 5 | CONCLUSIONS

553 The model performance consistently showed improvements to AGB prediction when context-aware features were  
554 included as predictors. This phenomenon was observed across regression models. Features that provide information  
555 about the tree neighborhood (e.g.,  $SL_i$  of tree height, average height of k-nearest trees) contain useful information  
556 to improve predictions of different individual tree traits (e.g., DBH, AGB). This finding suggests that the information  
557 retrieved from the local context serves as a proxy for underlying mechanisms that exert influence on the variable of  
558 interest, i.e., tree heights adapt locally as a result of environmental and biotic processes [1, 46, 47].

559 Utilizing the spatial association of structural tree traits, e.g., tree height, to define the local context range is a  
560 more effective approach compared to methods that rely solely on plot-level data from artificially delineated units,  
561 such as the monitoring plot size [29, 30, 33]. This is because contextual features may contribute to enhanced AGB  
562 predictions at larger scales beyond the plot level. Moreover, as the method proposed uses metrics entirely native to  
563 the UAV-LiDAR dataset, it does not rely on tailored process-specific indices (e.g., competition metrics) [32, 33, 34] or  
564 ancillary data sources (e.g., biomes, conservation status, ecoregions) [4], making this approach more transferable to  
565 other regions or scales.

566 A promising continuation within the scope of this research is to investigate the relative importance of different  
567 context-based metrics in enhancing tree-level AGB predictions. This pathway may yield valuable insights into the  
568 predictive power of various biotic and abiotic environmental factors as explanatory variables. Furthermore, since indi-  
569 vidual adaptive responses can vary among tree populations [124], evaluating how diverging tree-height adaptations  
570 to the local conditions can be linked to tree populations and genotypes with UAV-based methods seems a valuable  
571 endeavor to pursue. In this line, recent work demonstrated that linking tree phenotyping to inheritable traits using  
572 UAV-based methods is possible with relative accuracy [125, 126, 127].

573 Regarding UAV-LiDAR data acquisition, we recommend establishing protocolized procedures for assessing PCD  
574 quality, in line with recent suggestions [18]. Also, standardized methods for structural tree-trait data acquisition have  
575 been proposed [19, 128]. However, as data collection surveys are commonly challenged by environmental conditions  
576 and conducted by different field experts and protocols, the need to deal with noisy and disparate datasets is likely  
577 to persist. Therefore, in order to successfully integrate ML models into real analysis pipelines in bio-geography, it  
578 will be necessary to devise methods that are able to perform in the presence of label noise [129] and dataset shift  
579 effects [97], as these, unlike benchmark datasets, are ubiquitous in real-world AGB applications.

580 Finally, we recommend adopting a context-aware approach in the growing number of forest AGB mapping projects  
581 [9, 11, 130]. Likewise, we recommend using metrics entirely locally computed (e.g.,  $SL_i$ ) to detect local patterns and  
582 leverage their use, as suggested by Westerholt et al. (2018) [79]. In this way, the metric is sensitive to neighboring  
583 differences while remaining totally independent from spatial structures beyond the border of the neighborhood (i.e.,

584 the dataset as a whole).

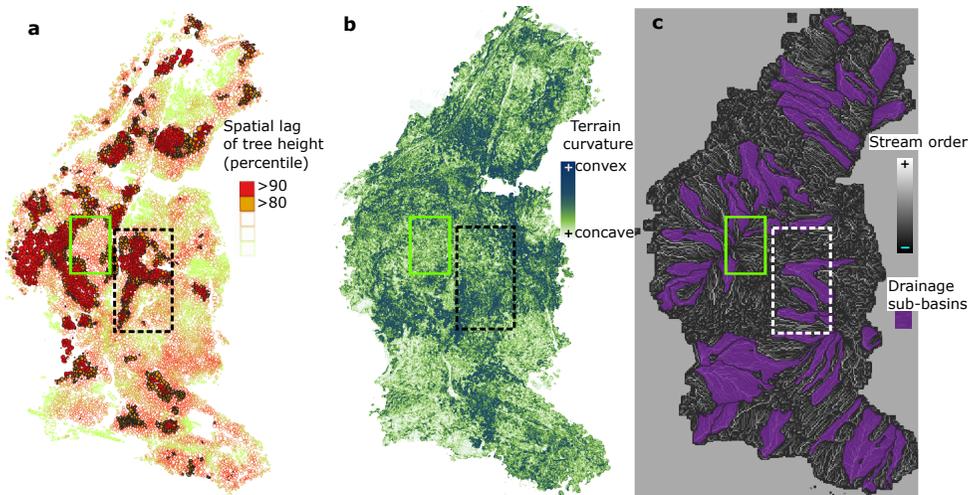
## 585 Acknowledgements

586 Helpful discussions with Thomas Friberg, Daniel Kükenbrink and Moritz Bruggisser are gratefully acknowledged. Like-  
 587 wise, we acknowledge the contribution of the field workers, who are responsible for collecting the forest inventory  
 588 data on a regular basis, used here as ground-truth. This project received funding support from the Talent Program  
 589 Horizon 2020/Marie Skłodowska-Curie Actions and the DeepCrop project (UCPH Strategic plan 2023 Data + Pool).  
 590 MG also acknowledges funding by Swiss National Science Foundation project ICOS-CH Phase 3 (20F120\_1982287).

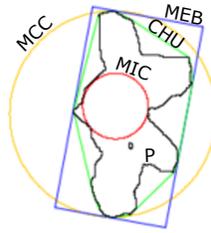
## 591 Supporting Information

### 592 | Annex I: Location and Morphometry of Tree Assemblages

593 The spatial distribution of  $SL_i$  presents directional anisotropy, occupying preferential areas which seem to match  
 594 sheltered sectors of the forest, such as concave thalwegs. Figure 9 highlights two neighboring areas with contrasting  
 595 values of  $SL_i$ , indicating that surface hydrology processes and terrain exposure (i.e., terrain convexity) condition tree  
 596 growth at the group level.



**FIGURE 9** a: Spatial lag of tree height derived from the individual tree crown (ITC) polygon dataset. b: map of terrain curvature derived from point cloud data (PCD) ground-returns. c: Hydrological network (Strahler's stream order [131, 132]). In all three panels, the dashed box indicates an area favored by surface hydrological conditions, hosting an assemblage of trees in the >90 % percentile of spatial lag of tree height. The solid green box indicates an area at a hilltop, unfavored by surface hydrological processes, more exposed to windstorm disturbance, and hosting an assemblage of trees in the < 60% percentile of spatial lag of tree height.



**FIGURE 10** Calculation of elementary geometries fitted to an exemplary tree assemblage. P: polygon of tree assemblage (black line). MCC: minimum circumscribed circle (in green). MIC: maximum inscribed circle (in red). CHU: convex hull (in yellow). MEB: minimum enclosing box containing P (in blue).

597 The morphometric analysis was conducted by taking into account the outer borders of tree assemblages defined  
598 either by  $SL_i$ , or by local Moran's  $I_i$  (delineated as explained in Section 2.6; results shown in Figure 5). The 20 basic  
599 morphometric variables (Table 3) result from fitting elementary geometries to the tree assemblage polygon. The 20  
600 derived variables are adimensional parameters (except for concavity, measured in m) obtained by combining the basic  
601 parameters.

**TABLE 3** Twenty basic morphometric variables derived from the tree assemblage polygon dataset (as described in [86]). P: tree assemblage polygon.

Basic parameters	Description	units
XPOL	Easting of P centroid	m
YPOL	Northing of P centroid	m
APOL	Area of P	m <sup>2</sup>
PPOL	Perimeter of P	m
LPOL	Major axis' length of P	m
WPOL	Minor axis' length of P	m
N-S	North-South alignment of P, defined as $ \sin(\text{azimuth}) $ of major axis	∅
ABOB	Area of the bounding box fully containing P	m <sup>2</sup>
PBOB	Perimeter of the bounding box fully containing P	m
AMEB	Area of minimum enclosing box	m <sup>2</sup>
PMEB	Perimeter of minimum enclosing box	m
ACHU	Area of the convex hull fully containing P	m <sup>2</sup>
PCHU	Perimeter of the convex hull fully containing P	m
AMCC	Area of the minimum circumscribed circle enclosing P	m <sup>2</sup>
PMCC	Perimeter of the minimum circumscribed circle enclosing P	m
RMCC	Radius of the minimum circumscribed circle enclosing P	m
AMIC	Area of the maximum inscribed circle enclosing P	m <sup>2</sup>
PMIC	Perimeter of the maximum inscribed circle enclosing P	m
RMIC	Radius of the maximum inscribed circle enclosing P	m

**TABLE 4** 20 derived morphometric variables derived from the tree assemblage polygon dataset (as described in [86]). P: tree assemblage polygon. A: area of P. L: length of major axis of P. W: width of minor axis of P (i.e., width). ACHU: area of convex hull fully containing P. RMCC: radius of minimum circumscribed circle. PCHU: perimeter of convex hull fully containing P. AMEB: area of minimum enclosing box.

Derived parameters	Name	Definition	Source
LTWR	Length-to-width ratio	$L/W$	[102]
WTLR	Width-to-Length ratio	$W/L$	[103]
ELLF	Ellipticity Factor	$ L - W /(L + W)$	[104]
CIRR	Circularity Ratio	$P^2/A$	[105]
ZFOR	Zăvoianu's Form Factor	$(16A)/P^2$	[106]
COMF	Compactness Factor	$P/(4\pi A)^{0.5}$	[86]
MCIR	Miller's Circularity Ratio	$(4\pi A)/P^2$	[107]
DISM	Dispersion Measure	$1 - [(4\pi A)^{0.5}/P]$	[105]
COMI	Complexity Index	$1 - [(4\pi A)/P^2]$	[86]
HFOR	Horton's Form Factor	$A/L^2$	[102]
ELOF	Elongation Factor	$(4A/\pi)^{0.5}/L$	[108]
LEMR	Lemniscate Ratio	$(\pi L^2)/4A$	[109]
REGF	Regularity Factor	$(\pi LW)/4A$	[104]
SHAF	Shape Factor	$[(4\pi A)/P^2] \times (L/W)$	[104]
CONV	Convexity	$PCHU/P$	[110]
CONC	Concavity	$ACHU - A$	[85]
SOLI	Solidity	$A/ACHU$	[111]
RECT	Rectangularity	$A/AMEB$	[112]
ROUN	Roundness	$(4\pi A)/(PCHU)^2$	[110]
SPHE	Sphericity	$(4A/\pi)^{0.5}/(2 \times RMCC)$	[113]

## 603 | Annex II: Context Detection

604 The distance range selected around each tree to compute neighborhood metrics (i.e., context detection), was con-  
 605 ducted based on the peak of significance (determined using the standard z-score) of local spatial autocorrelation  
 606 (using Local Moran's  $I_i$ ) as function of increasing distance, in steps of 10 m (as explained in Section 3.1).

607 Local Moran's  $I_i$  is a spatial statistic that relates attribute similarity to locational similarity, mapping the autocorre-  
 608 lation of individual tree heights across the geographical space, as defined above (Eq. 1, Section 2.5.1). The expression  
 609 below (Eq. 6) defines the z-score, which is used to measure the significance of tree-height clustering. Z-scores shows  
 610 the significance of the clustering by subtracting the observed  $I_i$  values from the expectation (i.e.,  $E[I_i]$ ), and normal-  
 611 izing over the standard deviation of  $I_i$ . This produces a distance metric in units of standard deviations.  $E[I_i]$  is the  
 612 expected value of local Moran's  $I_i$  under the null hypothesis of no spatial autocorrelation.

$$z_{score} = \frac{I_i - E[I_i]}{\sqrt{V[I_i]}}, \quad (6)$$

613 Neighborhood size was determined according to the significance of spatial autocorrelation (defined as local Moran's  
 614  $I_i$ ) as function of distance, via the standard z-score. Z-score measures the distance of a measured value from the ex-  
 615 pectation in units of standard deviation, under the assumption of randomly distributed values.

616 and the expected value of Moran's  $I$  under the null hypothesis of no spatial autocorrelation is:

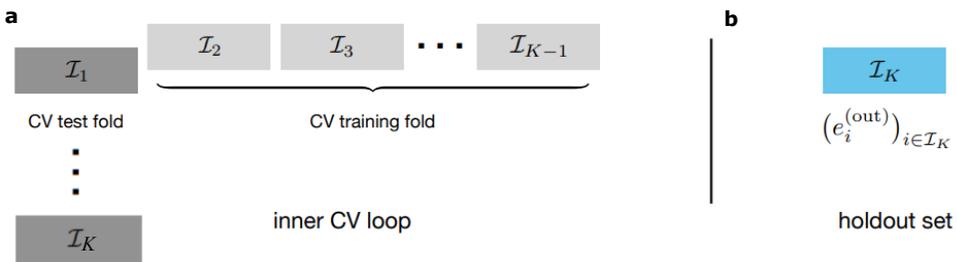
$$E[I_i] = -\frac{\sum_{j=1}^m w_{i,j}}{m-1} = -\frac{1}{m-1}, \quad (7)$$

617 where  $m$  equals the total number of trees in the neighborhood. At large sample sizes (i.e., for increasing values of  
 618  $m$ ), the expected value approaches zero. The spatial weights allocated to each neighboring tree  $j$  are standardized [81],  
 619 such that for each tree  $i$ ,  $\sum_j w_{i,j} = 1$ . The variance of local Moran's  $I_i$  is defined as the expectation of the square of  $I_i$ ,  
 620 minus the square of the expectations of  $I_i$ :

$$V[I_i] = E[I_i^2] - E[I_i]^2, \quad (8)$$

### 621 | Annex III: Training, Validation and Test

622 Nested cross-validation (NCV) [98] is an evaluation method for determining the accuracy of point estimates and  
 623 confidence intervals for prediction errors. How NCV is implemented is shown in Figure 11. The entire algorithmic  
 624 routine of NCV is presented immediately below, using pseudocode. The input data (i.e.,  $X, Y$ ) corresponds to the set  
 625 of predictors (i.e.,  $X$ ), and the target variable DBH (i.e.,  $Y$ ), respectively.



**FIGURE 11** Visualization of 10-fold nested cross-validation (CV). **a:** at each of the  $K$  steps ( $K = 10$ ), we perform standard cross-validation for model training (light grey folds), holding one of the folds out of the inner CV loop (dark grey fold). **b:** the fresh holdout folds (in blue) are never used for hyperparameter optimization or feature selection (figure adapted from Bates et al., 2021 [98]).

**Algorithm 1:** Nested cross-validation

---

**Input:** data  $(X, Y)$ , fitting algorithm  $A$ , loss function  $l$ , number of folds  $K$ , number of repetitions  $R$

```

procedure Nested cross-validation  $(X, Y)$                                 // ▷ primary algorithm;
   $es \leftarrow []$                                                     // ▷ initialize empty vectors;
   $a\_list \leftarrow []$                                               // ▷ (a) terms;
   $b\_list \leftarrow []$                                               // ▷ (b) terms;
  for  $r \in \{1, \dots, R\}$  do
    Randomly assign points to folds  $I_1, \dots, I_K$ ;
    for  $k \in \{1, \dots, K\}$  do
      // ▷ outer CV loop;
       $e^{(in)} \leftarrow$  inner cross-validation( $X, Y, \{I_1, \dots, I_K\} \setminus I_k$ ) // ▷ inner CV loop;
       $\hat{\theta} \leftarrow A((X_i, Y_i)_{i \in I \setminus I_k})$ ;
       $e^{(out)} \leftarrow (l(\hat{f}(X_i, \hat{\theta}), Y_i))_{i \in I_k}$ ;
       $b\_list \leftarrow$  append( $a\_list, (mean(e^{(in)}) - mean(e^{(out)}))^2$ );
       $b\_list \leftarrow$  append( $b\_list, var(e^{(out)})/|I_k|$ );
       $es \leftarrow$  append( $es, e^{(in)}$ )
     $\widehat{MSE} \leftarrow mean(a\_list) - mean(b\_list)$ ;
     $\widehat{Err}^{(NCV)} \leftarrow mean(es)$ ;
    return: ( $\widehat{Err}^{(NCV)}, \widehat{MSE}$ ) // ▷ prediction error estimate and MSE estimate;

procedure Inner cross-validation  $(X, Y, \{I_1, \dots, I_{K-1}\})$  // ▷ inner cross-validation subroutine;
   $e^{(in)} \leftarrow []$ ;
  for  $k \in \{1, \dots, K-1\}$  do
     $\hat{\theta} \leftarrow A((X_i, Y_i)_{i \in I_1 \cup \dots \cup I_{K-1} \setminus k})$ ;
     $e^{(temp)} \leftarrow (l(\hat{f}(X_i, \hat{\theta}), Y_i))_{i \in I_k}$ ;
     $e^{(in)} \leftarrow$  append( $e^{(in)}, e^{(temp)}$ )
  return:  $e^{(in)}$ ;

```

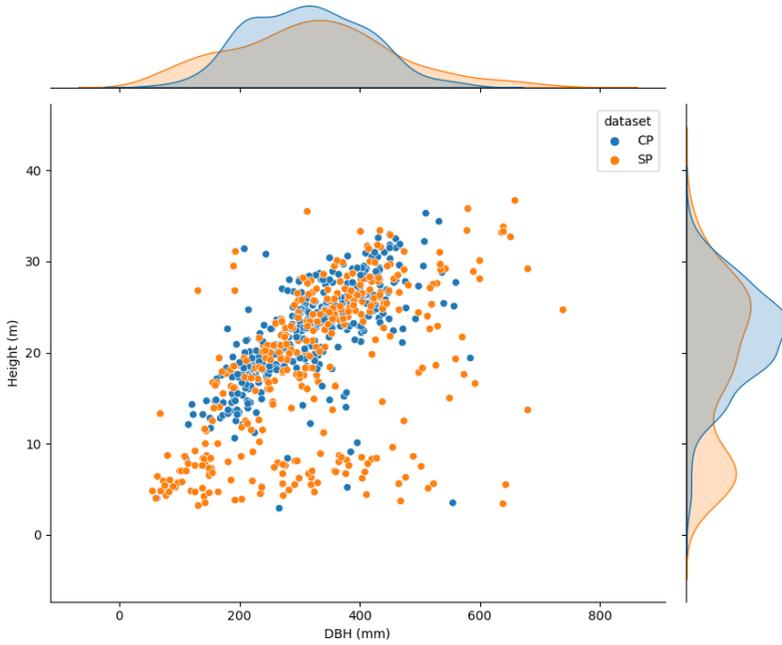
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**Output:** Nested cross-validation  $(X, Y)$

---

626 | **Annex IV: Distribution Shift Between CP-trees and SP-trees**

627 The joint distributions of (DBH, height) in both CP and SP datasets show a shift between the two [97]. For instance,  
 628 the kernel probability distribution of heights shows that the SP-dataset contains a higher amount of short trees (i.e.,  
 629 heights  $\in (3, \dots, 8)$  m), that cover a wide range of DBH values. Also, the range of DBH is broader in the SP-dataset  
 630 compared to the CP-dataset, and the instances do not exhibit an accumulation in the center as evident as the one  
 631 observed in the CP-dataset.



**FIGURE 12** Joint distributions of diameter at breast-height (DBH) and tree height from field-based inventory data. It should be noted that the two datasets are differently distributed—i.e., there is a dataset shift [97] between sampling plots (SP) and control plots (CP) datasets.

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