

1 **Limitations in historical satellite archives bias SDG monitoring**

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11 **Abstract**

12 Satellite remote sensing is vital to monitoring, research, and policy addressing sustainability
13 challenges from climate and ecosystem changes to food and water security. Here, Landsat
14 satellite data play a crucial role, thanks to their unique global, long-term, and high-resolution
15 coverage. Yet, disregarded biases in the Landsat data archive threaten the validity of their
16 applications. Here, we demonstrate that global Landsat data are spatiotemporally highly
17 uneven, frequently interrupted, and have seasonally incomplete coverage and quality. We
18 show that these limitations are inherited in prominent global time-series products, leading to
19 biased perceptions of changes in forests, croplands, and water resources that impair reliable
20 assessments of related sustainability issues. Several data limitations and their biasing effects
21 disproportionately affect lower-income countries. We provide global data-quality information to
22 support their explicit consideration in future mapping efforts. Our results call for better data-
23 bias reporting and control in satellite-based sustainability monitoring and analyses.

24 **Main**

25 193 countries committed to 17 Sustainable Development Goals (SDGs)¹ to comprehensively
26 address the environmental and social impacts of economic development. Yet, nearing the
27 target year 2030, we are still far from meeting these goals². Widespread and rapid land
28 alterations³ cause biodiversity loss⁴, accelerate climate change⁵ and threaten regional food and
29 water security⁶.

30 The global 2030 Agenda for Sustainable Development foresees regular progress monitoring
31 and reporting towards these SDGs, as a basis for their periodic recalibration to regional
32 development differences⁷, and to identify national responsibilities for sustainability issues and
33 secure practical commitments⁸. Satellite remote sensing allows monitoring many SDG
34 indicators at multiple spatial and temporal scales⁹ and, thanks to open-data policies of key
35 satellite archives¹⁰, (geo)computational advances¹¹, and investments in technical capacity-

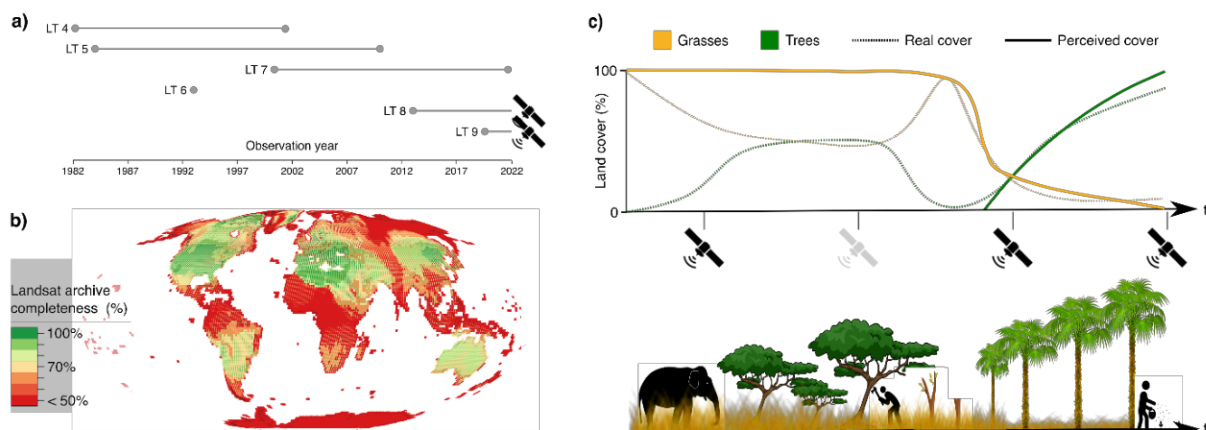
36 building¹², has become a primary tool for countries to meet their reporting obligations⁹. In
37 particular, the Landsat program¹³ fulfils a vital role for continuous land-surface monitoring, due
38 to its unique combination of long historical coverage (**Fig. 1a**) and relatively high spatial and
39 temporal image resolution (30-m since 1982, typically every 16-days). Thanks to the program's
40 longevity, Landsat data are key to evaluating long-term environmental changes against
41 historical baselines or in response to human interventions¹⁴. Moreover, Landsat data play an
42 important role in providing forward-looking policy support by enabling the construction of
43 credible future scenario projections that help anticipate emergent development issues and their
44 possible responses to policies¹⁵. Correspondingly, Landsat data have become an integral part
45 of a growing number of mapping applications¹⁶, including many designed specifically for
46 tracking progress towards SDGs¹⁷⁻¹⁹.

47 However, the archive of historical Landsat satellite images contains extensive data gaps (**Fig.**
48 **1b**), and existing images often have limited quality due to cloud cover or data degradation²⁰.
49 These gaps and quality limitations can lead to misclassifications of land-surface features in
50 derived time-series products, and thus to misinterpretations of land changes (**Fig. 1c**). Regional
51 and temporal variations in the magnitude and prevalence of these satellite data limitations can
52 hamper efforts to reliably map change patterns and to establish common baselines across
53 countries. For example, interruptions in the annual continuity of satellite observations may lead
54 us to misinterpret the timing and magnitude of long-term changes in forest cover²¹. Changes
55 in time intervals between usable images, in turn, affect our ability to detect periodic, abrupt
56 vegetation changes that often distinguish agricultural from other vegetated lands²². Similarly,
57 changes in within-year image availabilities affect our ability to capture dynamics of seasonally
58 occurring land resources²³.

59 Although many remote sensing experts are aware of such issues²⁰, their maps and time-series
60 products do not generally control for the different data limitations. Moreover, their products
61 ultimately reach a much broader, non-expert user community of resource managers, policy-
62 makers, and scientists from different disciplines. Additionally, open-access policies and cloud-
63 computing technologies have mainstreamed remote sensing and enabled non-experts to
64 generate their own products²⁴. While these developments helped accelerate scientific
65 progress²⁵, many dataset developers, as well as data users, are now largely unfamiliar with the
66 limitations of satellite archives²⁶. Moreover, as data choices become vast, more users rely on
67 the 'Landsat' reputation as an indicator of data quality²⁶. Most literature on satellite data
68 limitations is highly technical and directed at remote sensing experts²⁰, whereas literature
69 bridging remote sensing and other fields mostly focuses on promoting different remote-sensing
70 methods or applications²⁷ and disregards their limitations. Frank discussions of the magnitude

71 and implications of different types of satellite data limitations targeted at non-expert data
 72 developers and data users are largely lacking.

73 To tackle this gap, we expose the magnitude of spatial and temporal variations in different
 74 types of data limitations in the global, historical Landsat archive since the 1982 launch of the
 75 first 30-meter-resolution sensor. To this end, we map different dimensions of data limitations
 76 in the archive, quantifying the between- and within-year frequency, recurrence, and quality of
 77 daytime Landsat images. Additionally, we demonstrate how spatial and temporal fluctuations
 78 in image coverage and quality affect our perception of global changes in forest cover, seasonal
 79 water availability, and arable-land extents as mapped by state-of-the-art monitoring products.



80
 81 **Figure 1. Temporal gaps in satellite observations and their conceptual link to biased perceptions of land**
 82 **changes.** a) Five operational Landsat (LT) satellite missions have provided seemingly uninterrupted global
 83 coverage between 1982 and 2022 (grey lines; satellite icons indicate active missions as of 2022). b) However,
 84 available satellite images in the Landsat archive provide a much less continuous data coverage. c) These data gaps
 85 influence our perception of land-change trajectories. Consider a natural succession (bottom image) from an open
 86 grassland to savanna (grasses and trees), in which tree cover is subsequently removed by logging, before dense
 87 tree cover emerges through artificial plantations. These land-change processes correspond to time-series of grassy
 88 and tree cover (dashed lines in top image). Grassy/tree cover is perceived via repeated satellite observations
 89 (continuous line) only at distinct time steps (x-axis ticks with black satellite icons), and may thus miss important land-
 90 change processes when there are larger temporal gaps (e.g., due to sensor failure; grey satellite icon), such as
 91 here, the gains and losses of savanna.

92 Results and Discussion

93 *An uneven history of satellite data limitations*

94 Between 1982 and 2022, the Landsat archive contains 17,553,123 daytime images (per-pixel
 95 avg.: 1618, ± 1599 ; **Extended Data Fig. 1a**), meaning that 33.1% of images that would be
 96 expected under a 16-day revisit frequency are missing (see **Methods**). Moreover, the average
 97 pixel effectively lost an additional 44.8% ($\pm 17.6\%$) of available observations due to cloud cover
 98 or degradation of Landsat images (**Extended Data Fig. 1b/d**), leaving an average quality-
 99 weighted number of 1,429 (± 887.4) images per pixel across the 40 years (**Extended Data Fig.**
 100 **1c**). Data gaps are spatially and temporally highly uneven and particularly prevalent in the
 101 Tropics, Arctic, and Antarctic (**Fig. 1b**) and in earlier years (**Fig. 2a**). Notwithstanding the

102 regionally important role of cloud cover²⁸, extensive data gaps primarily relate to the historical
 103 development of the Landsat program²⁰.

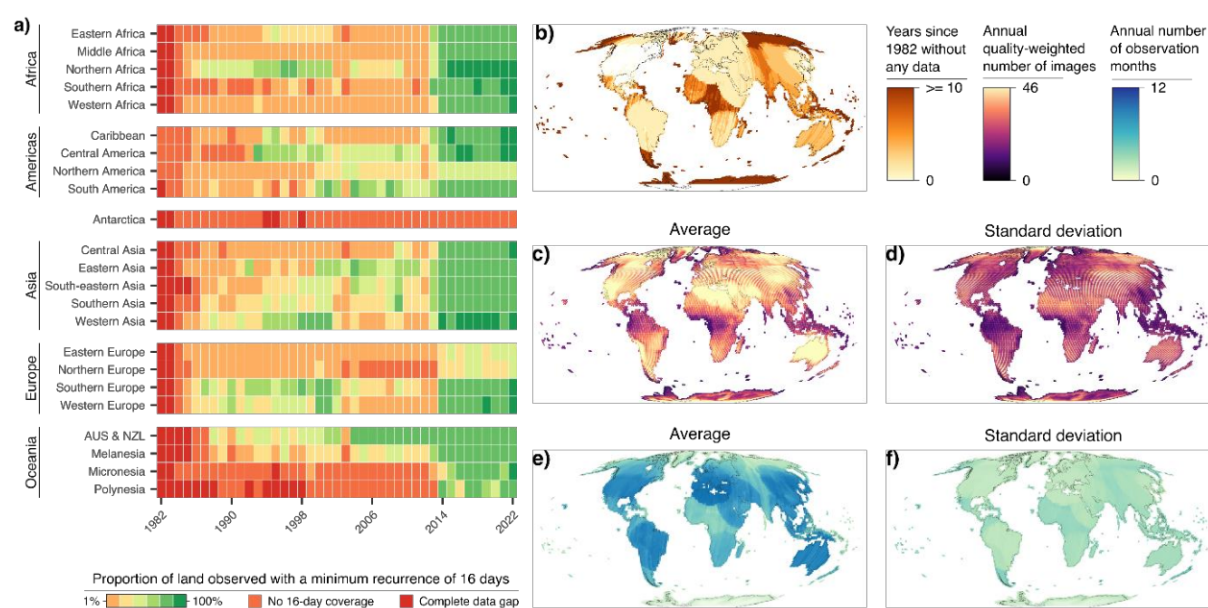
104 Global coverage of Landsat data evolved only gradually (**Fig. 2a, Extended Data Fig. 3**) with
 105 the support of a global network of receiving stations established in all continents except
 106 Antarctica (**Extended Data Fig. 2a**). Whereas the archive grew by an average of 118,362
 107 ($\pm 67,104$) images per year during the 1980s, this rate increased nearly five-fold by the 2010s
 108 (to 575,169, $\pm 159,924$; **Extended Data Fig. 3**). However, improvements were not uniform.
 109 Many countries lacked (or still lack) the infrastructure and know-how to continuously collect
 110 and preserve data from overpassing satellites²⁰ (**Extended Data Fig. 2a-b**). As a result, many
 111 world regions, particularly at low and very high latitudes, have lagged behind in developing a
 112 dense Landsat data record (**Fig. 2a**). Improvements happened particularly late over islands,
 113 with, for example, 65.4% of Oceanian Island areas never observed until the 1990s.

114 Beyond these gradual increases, changes in Landsat satellite technologies caused several
 115 abrupt global changes in data coverage. Since the launch of Landsat 7 in 1999, Landsat
 116 satellites include on-board data storage that reduced the reliance on global networks of
 117 receiving stations for assuring data collection and archiving²⁰. Combined with improving data
 118 transmission and warehousing, this helped the Landsat archive expand rapidly during the 21st
 119 century. Since 2003, however, mechanical issues in Landsat 7 degraded as much as 25% of
 120 pixels per image²⁹. With the end of Landsat 5 in 2010, the quality of available images decreased
 121 until the launch of Landsat 8 in 2013, where data coverage improved dramatically (**Fig. 2a**).
 122 The average proportion of countries' lands with a full-year coverage increased from 17.5%
 123 ($\pm 30.8\%$) for the years before 2013 to 89.4% ($\pm 26.4\%$) thereafter. Additionally, interoperability
 124 between Landsat 8 and earlier missions is hindered by differences in the sensed spectral
 125 information³⁰, a challenge not addressed explicitly in state-of-the-art mapping applications.

126 With the exception of much of North America, nearly all land areas have continuous gaps
 127 lasting ≥ 1 year in their post-1982 Landsat data record (**Fig. 2b**). 36.2% of those lands have ≥ 1 -
 128 year interruptions after their first data coverage, including most islands, most of Africa,
 129 Mesoamerica and north-eastern South America, northern Beringia, Patagonia, and Antarctica
 130 (**Extended Data Fig. 4a**). Oftentimes, these interruptions persisted over extended periods
 131 (averaging 5.4 years, ± 3.7), with several Central African regions not having a single usable
 132 Landsat image during >10 consecutive years (**Extended Data Fig. 4b**). These archive
 133 interruptions often reflect short-lived or inconsistent receiving and storage capacities
 134 (**Extended Data Fig. 2b**), and are particularly severe during the mid-1990s²⁰.

135 When data are available, their frequency and quality vary (**Fig. 2c-d**). Whereas much of the
 136 world has a large number of *any* Landsat images (**Extended Data Fig. 1a**), consistent high

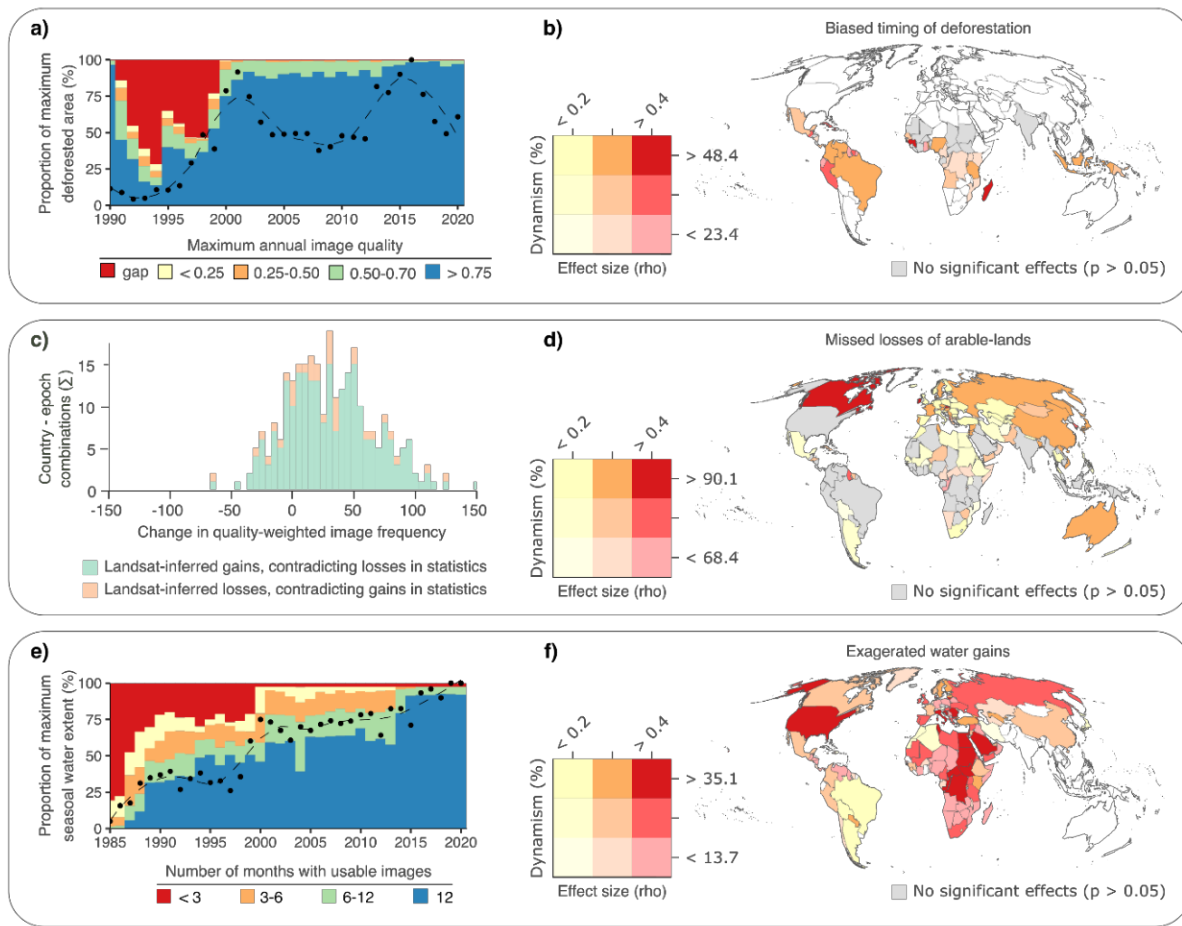
137 coverage with *high-quality* images only exists for dryland regions of the Middle East, North
 138 Africa, Australia, and North America (**Extended Data Fig. 1c**). By contrast, for most equatorial
 139 forest regions of the world, less than half of existing images are usable, due to persistent cloud
 140 cover²⁸ (**Extended Data Fig. 1b**). In some areas such as the Sahel belt and much of Central
 141 Asia, annual fluctuations in quality-weighted image numbers exceed annual averages (**Fig. 2c-**
 142 **d**). In similar regions, seasonal data coverage is highly incomplete, often with less than three
 143 months covered with data in a given year, and fluctuations of a similar magnitude between
 144 years (**Fig. 2e-f**).



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 146 **Figure 2. Spatiotemporal variation in Landsat satellite data coverage and quality.** a) Year-to-year and regional
 147 variation in proportions of land areas observed during every month of the year, highlighting both interruptions and
 148 gradual improvements and a sharp increase in coverage in 2013, after the launch of Landsat 8. b) Accumulated ≥ 1 -
 149 year data gaps. c) Global variation in average annual quality-weighted image numbers and d) their multi-annual
 150 fluctuation. e) Global variation in average annual numbers and f) multi-annual fluctuation of observation months.

151 ***Data limitations bias perceived changes in SDG indicators***

152 Unless carefully accounted for, the described spatial and temporal differences in data coverage
 153 and quality can introduce biases into the derived time-series products used for monitoring
 154 SDGs. For example, discontinuous historical satellite-data coverage impairs our perception of
 155 the timing of change events, while varying frequencies of quality images affect our ability to
 156 perceive time-sensitive changes, and fluctuations in seasonal data completeness limit our
 157 ability to distinguish multi-year changes from seasonal fluctuations. These biases in monitoring
 158 can lead to poor policy-making. For example, biased information on trends in forest and
 159 wetland areas may lead to ill-conceived protection and restoration goals as Nationally
 160 Determined Contributions under the Paris Agreement. In the following paragraphs, we will
 161 highlight how different types of Landsat data limitations indeed bias perceptions of changes in
 162 different SDG indicators derived from state-of-the-art monitoring products.



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Figure 3. Effects of satellite data limitations on perceived land changes. **A)** Points: annual lost tropical moist forest areas normalized by maximum extent (1990–2020), revealing positive/negative outliers relative to smoothed trend (LOWESS with 0.4 span; dashed line). Colors: percentages of maximum forest extent observed with different maximum annual data qualities (red: complete data gap). **b)** Country-level differences in strengths of causal effects (ρ ; x-axis) of anomalies in maximum within-year Landsat image quality on anomalies in perceived deforestation rates, intersected with differences in forest-area dynamism (percentages of maximum forest extent experiencing change; colored by tertiles; y-axis). **c)** Frequency of disagreements in the direction of Landsat-inferred arable-land area losses/gains between mapped 4-year epochs and changes derived from official statistics, distributed along a gradient of changes in quality-weighted Landsat image frequencies between the epochs. **d)** As in b), but for effects of improvements in quality Landsat data on Landsat-inferred arable-land gains contradicting statistics-inferred losses. **e)** Analogous to a), points indicate annual global area that is seasonally covered by surface water, normalized to maximum extent, and colors indicate percentages of the maximum extent observed with varying seasonal Landsat data completeness (number of months covered with ≥ 1 usable image). **f)** As in b), but for effects of anomalies in seasonal Landsat data completeness on anomalies in seasonal water areas.

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Gaps in quality data bias perceived timings of deforestation events. Changes in forest

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areas affect multiple SDGs and are the focus of SDG indicator 15.1.1. Tropical moist forests

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accounted for $>90\%$ of global deforestation since 2000³¹, and are particularly important for

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capturing and storing carbon (SDG 13), preserving and restoring biodiversity (SDG 15), and

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providing billions of people with income, food, and/or medicine from forest products (SDGs 1,

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2, and 3). Accordingly, the recently published Landsat-based Tropical Moist Forest product

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(TMF)¹⁹, which maps the onset of deforestation since 1982, is poised to play prominent role in

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global monitoring under diverse policy frameworks, including the Paris Agreement, the Post-

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2020 Global Biodiversity Framework, or EU regulations on deforestation-free supply chains.

187 The TMF addresses temporal gaps in the Landsat archive by preserving the last-recorded
188 classes into the gap periods and only mapping any class changes once new satellite images
189 confirm those¹⁹. The resulting annual time-series thus hide uncertainties regarding the true
190 timing of forest-change events. This means that the inferred deforestation years and perceived
191 change trajectories may commonly be biased by unaccounted gaps and quality differences in
192 the Landsat archive.

193 We examined the TMF and, indeed, found strong indications of such biases. Globally,
194 perceived annual deforestation affected disproportionately large portions of the tropical moist
195 forest biome during two periods since 1990, both of which mark periods of particularly rapid
196 improvements in Landsat satellite-data coverage and quality (**Fig 3a**). For example, from 1999
197 to 2000, right after the launch of Landsat 7, the World seemingly experienced an increase in
198 deforested areas of 65.2%, more than twice the maximum year-over-year increase (31.9%)
199 registered anywhere between 1990 and 1999. Similarly, the 2012-2013 deforestation increase
200 of 60.5% coincides with increased image frequencies following the launch of Landsat 8 and is
201 nearly twice the recorded maximum over the 2001-2012 period (30.8%), in which Landsat 7
202 was the sole data source. Regions experiencing ≥ 1 -year periods with either no data or
203 potentially unusable, low-quality data show disproportionately higher deforestation rates during
204 years immediately following those periods, compared to their smoothed trend line (**Fig. 3a**; see
205 **Methods**). This results in 67,329,203 ha of globally deforested areas that are potentially
206 allocated to the wrong year (**Extended Data Fig. 5a**; see **Methods**), corresponding to 58.4%
207 of total gross deforestation mapped since 1990.

208 We wanted to know whether gaps in Landsat data actually bias the perceived timing of
209 deforestation. To this end, we used a formal causal analysis technique developed for detecting
210 causal relationships between two time-series called Convergent Cross Mapping^{32,33} (CCM; see
211 **Methods**). Based on the results of these analyses, we attribute deforestation anomalies to
212 anomalies in maximum annual image quality in preceding periods in 68.8% of tropical-moist-
213 forest countries (**Fig. 3b**; see **Methods**).

214 Resulting biases in perceived deforestation years may bias any timing-sensitive applications
215 related to achieving SDGs, including modelling of carbon emissions⁵, restoration prioritization
216 to mitigate extinction debts³⁴, or attributions of forest changes to changing socio-political
217 conditions³⁵. For example, deforestation inside the Luo Scientific Reserve (Democratic
218 Republic of Congo) that reportedly happened during the first Congo war (1996-1997) due to
219 human displacement³⁵ would be falsely attributed to processes in the immediate post-war
220 period, for which Landsat data are again available (**Extended Data Fig. 5b**). These biases
221 may also cast unfair perceptions of national progress in curbing deforestation. Twelve countries

222 indicated by the TMF as having increasing deforestation rates around the Landsat 8 launch –
 223 when data improvements were particularly strong (**Fig. 2a**) – in fact reported decreases in the
 224 Forest Resource Assessments (relative to the previous reporting period)³⁶, including countries
 225 with successful restoration and conservation programs over that period (e.g., Cuba, India,
 226 Vietnam, Thailand)³¹.

227 **Increasing frequencies of quality data miss regional arable-land losses.** Accurately
 228 capturing dynamics in arable-land extents is a critical component of measuring SDG indicator
 229 2.4.1 on agricultural lands under sustainable use, and is also closely linked to indicators aimed
 230 at avoiding deforestation (15.2.1) and loss of water-related ecosystems (6.6.1)⁶.

231 Mapping arable-land requires temporally dense satellite observations to capture phenological
 232 land-surface changes driven by crop planting and harvesting, and as such is highly sensitive
 233 to cloud-related gaps in Landsat data³⁷. To tackle this, a recently developed global product
 234 (GLAD)¹⁸ maps arable-land in four-year epochs, exploiting the highest-quality images of an
 235 entire epoch (aggregated into an annualized 16-day time-series) for more accurate detections
 236 of cropping-related phenological patterns¹⁸. Yet, this approach is not immune to increasing
 237 densities of high-quality images in the Landsat archive over time (**Extended Data Fig. 3**), which
 238 may lead to overestimations of arable-land gains by reducing the likelihood of missing existing
 239 arable-lands, compared to earlier time periods. Simultaneously, changes in newer Landsat
 240 sensors relative to earlier missions (e.g., different band spectral ranges)³⁰ are likely to misinform
 241 classification algorithms fed mainly with data from earlier periods.

242 In fact, we identified 123 countries where the GLAD mapped gains despite reported losses in
 243 national statistics (**Extended Data Fig. 6**), casting doubts on 74,975,550 ha of arable-land
 244 expansion, an area larger than the total of all arable-lands across all Amazonian countries in
 245 2021³⁸ (**Extended Data Fig. 5c**). Most (80.0%) of these positive disagreements relative to
 246 statistics are associated with improvements in the frequency of quality Landsat images (**Fig.**
 247 **3c**). These disagreements peak between the 2008-2011 and 2012-2015 epochs (36.0% of
 248 cases), coinciding with the 2013 launch of Landsat 8 which massively increased quality-image
 249 numbers (**Fig. 2a**). Doubtful arable-land gains concentrate in Southern Asia (19.5% of doubtful
 250 gains), South America (18.8%), and Western Africa (18.5%, **Extended Data Fig. 5c**), with
 251 countries such as Ghana in Nepal consistently experiencing positive disagreements between
 252 all epochs. Our causal analysis using CCM attribute the former to the latter in 48.4% of
 253 countries (**Fig. 3d**; see **Methods**).

254 These biases in perceived arable-land changes can severely bias perceptions of global food
 255 security issues. The 54 countries with moderate to high bias-causing effects include top food-
 256 producing countries (e.g., China, Russia, France) and together accounted for 38.4% of global

257 cereal production in 2021³⁹. However, they also include many food-insecure countries (e.g.,
258 Central African Republic, Niger, Somalia, South Sudan, Yemen, Zimbabwe), where
259 misinterpreting losses of arable-lands for gains bears risks that policy-makers might fail to
260 recognize emerging crises.

261 Overestimated arable-land gains can also lead to unfair evaluations of progress towards SDG
262 target 2.4 (sustainable food production) that exaggerate conflicts of food security with
263 ecosystem protection and climate-change mitigation. For example, two recent studies^{40,41} using
264 GLAD data reported extensive cropland expansion into global protected areas, with massively
265 accelerating expansion rates between the mid-2000s and mid-2010s. The above-described
266 data biases associated with the 2013 launch of Landsat 8 (**Fig. 2a, Fig. 3c**), however, may
267 render these assessments unreliable. This is illustrated in India, where sudden changes in
268 Landsat data led the GLAD mapping algorithm to falsely re-classify an entire protected Ramsar
269 wetland of >3,000 ha into arable-land (**Extended Data Fig. 5d**).

270 **Improving seasonal data completeness exaggerates water gains.** SDG Indicator 6.6.1
271 tracks changes in surface water bodies, such as lakes, rivers, and reservoirs, and is informed
272 by the Landsat-based Global Surface Water product (GSW)¹⁷. Particularly in many dryland
273 regions of the world, seasonal water bodies that only exist for a few months per year play a
274 crucial role for water security, both as seasonal sources of drinking water and water for
275 livestock and cropping⁴², as well as for filling aquifers that sustain water supplies during dry
276 seasons. Even outside drylands, seasonal flooding of river plains affects both natural nutrient
277 inputs in, and leaching from, major agricultural production regions⁴³. The GSW maps seasonal
278 (as well as permanent) surface water extents annually based on monthly classifications of
279 water occurrences.

280 These data show a nearly 5-fold increase in global seasonal water areas between 1984 and
281 2020, with increasing trends over 90.1% of the maximum seasonal-water extent. However,
282 because the GSW maps water if as little as 43.5% of expected images per year are available,
283 seasonally biased distributions of those images could either entirely miss seasonal water
284 occurrences or misclassify seasonal for permanent water (if only covering the dry or wet
285 season, respectively). Therefore, long-term increases in seasonal completeness could be
286 falsely mapped as increasing seasonal-water extents⁴⁴.

287 We found that, indeed, global seasonal surface water gains correlate with improvement in
288 seasonal data completeness (number of months with usable data; $r^2=0.80$; **Fig. 3e, Extended**
289 **Data Fig. 5e**), which are largely unsupported by local discharge measurements (64.5% of
290 gauge stations show disagreements, **Extended Data Fig. 7**; see *methods*). Our causal
291 analysis using CCM found moderate to high bias-causing effects in 144 countries (**Fig. 3f**),

292 including several with severe water stress (e.g. Yemen, Sudan)³⁹, mischaracterizing persistent
293 and expanding water scarcity issues driven by increasing drought frequencies⁴⁵ (e.g., in
294 Somalia⁶; **Extended Data Fig. 5f**).

295 ***Biases disproportionately affect lower-income countries***

296 We found that Landsat data limitations, as well as the resulting biases in perceptions of land
297 changes, occur disproportionately often in countries with lower financial capacity to sustain
298 remote sensing monitoring programs. Specifically, biasing effects on perceived arable-land and
299 seasonal water trends were significantly more frequent in lower-income than in higher-income
300 countries (McNemar's tests, arable-land: 51.9% of lower-income vs. 46.1% of higher-income,
301 p -value=0.00; water: 89.7% vs. 73.1%, p =0.00; note there was a near-significant difference in
302 deforestation bias in the opposite direction among the respective income groupings of tropical-
303 moist-forest countries; 51.9% vs. 46.1%, p =0.07; *details in **Methods***). Unless ensuing biases
304 in SDG indicators are accounted for, misperceptions of progress in food- and water-security
305 goals in those countries may hamper adequate international support and timely policy
306 interventions.

307 Similarly, we found higher average frequencies of years without any usable data in lower-
308 income countries (Wilcoxon test, p =0.0, avg. of 4.9% [\pm 2.9] vs. 3.7% [\pm 4.0] for higher-income
309 countries), affecting 43.3% of their combined area, compared to only 17.2% of the combined
310 area of higher income countries (mainly high-latitude and offshore territories). Similarly, we
311 found that pixels in lower-income countries were more frequently affected by fluctuations in
312 usable-data frequencies exceeding the expected frequencies under a 16-day recurrence
313 (60.2%, vs. 40.6% for upper-middle-/high-income countries; p =0.00), and also by fluctuations
314 in usable-data months exceeding a typical climate-season length (88.2% for lower-income vs.
315 70.73% for higher-income countries; p =0.0).

316 ***Future needs: bias corrections, fair product validations, and support to users***

317 While this paper focuses on Landsat data as the most important resource for long-term, global
318 land-change monitoring, all satellite data archives are affected by uneven data coverage and
319 quality^{46,47}. Given the importance of satellite-based land-change observations for sustainability
320 policy, monitoring, and related scientific fields, addressing the highlighted biases caused by
321 limitations in global satellite archives becomes imperative. This will require more rigorous bias-
322 control and more honest validations by data developers, as well as better support for (and
323 commitment by) data users for detecting and addressing remaining uncertainties.

324 Firstly, expert communities developing remote-sensing-based time-series products should
325 raise standards for correcting for satellite data limitations before applying classification
326 algorithms. An increasing array of sophisticated approaches can fill gaps in satellite

327 archives^{48,49}, for example, by fusing sparse Landsat with coarser-resolution but less incomplete
328 data from the MODIS and AVHRR satellite missions to generate global, seamless data cubes⁵⁰,
329 but such approaches remain rarely applied in operational land-surface monitoring. To further
330 improve their performance, information on data coverage and quality, as provided here for
331 Landsat (see **Data availability** and **Code availability** sections), could be made available for
332 all sensor systems, enabling its explicit use by gap-filling models for correcting satellite data to
333 desired, high-quality levels.

334 Secondly, we need higher standards for assessing uncertainties in the derived land-change
335 products. All three products scrutinized here were, in fact, extensively validated by their
336 developers. Yet, validation samples were mostly generated by visually interpreting Landsat
337 images^{17–19} – as is true for nearly all global time-series, especially for pre-2000 periods, where
338 few alternative sources of validation data exist⁵⁰. For accuracy assessments to be meaningful,
339 however, accuracies must be comparable between validated and non-validated pixels and
340 years⁵¹. In reality, the selection of validation samples and their correct visual interpretations are
341 both biased away from the most data-limited regions and periods, which is also where the
342 classification algorithms are most likely to fail. This likely results in exaggerated accuracy
343 scores and hence unwarranted trust in Landsat-based monitoring products. To be honest
344 towards data users, accuracy tests should directly incorporate information on limitations in both
345 satellite and validation data. Again, models could be used to generate seamless predictions of
346 class-confusion probabilities in between existing samples that are representative of all pixels
347 and years, including those with limited data. Much more than allowing 'corrections' of all pixel
348 values⁵¹, this should allow mapping remaining uncertainties in ways that enable their due
349 propagation into change assessments and indicators⁵², for example, in form of probability-mass
350 functions of alternative class sequences for each pixel.

351 Such higher standards would imply more time needed for the development and quality-
352 assurance of time-series products, and thus fewer, more transparent products that pass peer-
353 review and enter the market every year. This would be desirable from the perspective of data
354 users, who are already overwhelmed by too many products to choose from with little guidance
355 on which products they should trust²⁶. Many data users may be similarly overwhelmed by fewer
356 but more voluminous products with rich, pixel-level uncertainty information, as they lack the
357 technical capacity to effectively use them. Thus, we additionally need easy-to-use tools helping
358 with their use, as well as with selecting the most fit-for-purpose product for a given desired
359 application. For example, software packages could support easier incorporation of data
360 uncertainties into, and propagation between, different types of applications (mapping, change
361 assessment, causal analysis, etc.). Similarly, cloud-based tools could automatically test where

362 within a user-specified region and period a given product could plausibly support the desired
363 application, given the product's uncertainties and/or underlying satellite data limitations.

364 At the same time, data users should acknowledge that remote-sensing "data" on land changes
365 are not facts, but model-based interpretations of (satellite) data that often inherit large
366 uncertainties. Ultimately, data users carry the burden of validating their original results on land
367 changes, even when they rely on existing products. Easy-to-use and freely available webtools
368 for exploring historical time-series of high-resolution images (e.g., Google Earth Pro) empower
369 them to do so.

370 By highlighting data limitations and biases in perceived land changes, and offering data-quality
371 layers and suggestions for addressing these, we provide an essential first step. We hope that
372 this may serve as a starting point for the needed collaborative actions, to ensure that satellite-
373 based data can reliably guide progress towards a sustainable future.

374 **Methods**

375 ***Quantification of satellite data limitations***

376 ***Landsat quality metrics.*** We developed a suite of quality metrics that characterise the quality
377 and coverage of Landsat satellite observations. The data-coverage aspects considered are
378 annual and year-to-year frequency and recurrence of images, weighted by the quality of
379 individual observations. The quality of individual images reflects the reported geometric and
380 spectral image quality⁵³, as well as limitations in image usability caused by cloud cover²⁸. We
381 calculated these quality metrics for each year and each descending tile drawn in the World
382 Reference System 2 (WRS-2), which is used to partition Landsat data into publicly available
383 chunks⁵³. We then combined the tile-specific metrics by averaging them into global grids with
384 a 1-km resolution.

385 We calculated the quality metrics base on the available metadata for all images acquired
386 between 1982 and 2022⁵⁴ with Landsat 4, 5, 7, 8, and 9. We thus disregarded images obtained
387 with Landsat's Multispectral Scanner System (MSS) that, while available since 1972, provide
388 data with a lower spatial resolution (60 to 90-m) compared to more recent sensors (30-m), the
389 latter resolution providing more adequate spatial detail for historical environmental
390 monitoring⁵⁵. MSS data also have a coarser spectral resolution, making them unsuitable for
391 many remote sensing applications^{56,57}.

392 ***Image quality.*** We calculated the quality of each Landsat image j (QI), for a given year y and
393 tile t , as $Q_{y,t,j}/9 \times (100 - CCover_{y,t,j})$. This results in a normalized metric between 0 (worst) and
394 1 (best) based on cloud-cover percentage (CCover) and the spectral and geometric 'image
395 quality' (Q). The latter is as qualitative metric between 0 and 9 that grades the number of bad

396 scans in an image⁵³, which we normalized by the maximum possible grade. As cloud cover
397 compromises the reliability of individual pixels⁵⁸, QI is 1 (best) if Q is 1 and CCover is 0, and
398 decreases as CCover increases. Similarly, QI is lower when Q is lower, such as for images
399 acquired by Landsat 7 post 2003 due to the degradation of the sensor.

400 ***Within-year variations in data quality.*** Using QI, we measured the within-year frequency and
401 completeness of the Landsat archive, and identified between-year interruptions in data
402 availability. Here, we only considered images with QI > 0, because images with a QI of zero
403 imply the absence of any data usable for downstream land-surface monitoring applications,
404 such as those discussed in this paper.

405 First, we counted the number of years without data since 1982 (i.e., the first year when Landsat
406 4 was operational), capturing inconsistencies in data collection efforts. Second, we summed
407 the QI of all images collected in each year, which weighs down the annual image frequency by
408 the proportion of usable pixels in each image (referred to as 'quality-weighted frequency' in this
409 paper), as not all images provide usable data due to clouds or otherwise obstructed visibility or
410 data degradation. Increases in quality-weighted data frequencies over time increase the
411 chance of misinterpreting the long-term changes, such as those related to land use (see
412 *'Increasing frequencies of quality data miss regional arable-land losses'*). Third, we counted
413 the number of months with usable images, which informs on our ability to perceive seasonal
414 change dynamics, such as those driven by regional climate forcing (see *'Improving seasonal
415 data completeness exaggerates water gains'*). Fourth, we estimated the maximum within-year
416 QI across all image acquisitions, which depicts the likelihood of detecting persistent changes
417 between in comparison to previous years (used in *'Gaps in quality data bias perceived timings
418 of deforestation events'*).

419 ***Estimating the completeness of the Landsat archive.*** We quantified the 'completeness' of
420 the Landsat archive as the proportion of usable images relative to the number of expectable
421 images when assuming one image every 16 days, or 23 per year, per active sensor. The
422 number of usable images corresponds to the sum of annual, quality-weighted image
423 frequencies.

424 ***Analyses of effects of satellite data limitations on perceived land changes***

425 ***Causal analysis approach.*** We tested for biasing effects of data-quality on different Landsat-
426 inferred land changes for three case studies, corresponding to three different types of data
427 limitations and three different types of land changes that are expected to be sensitive to those
428 (see *below*). We tested for causal effects using Convergent Cross Mapping (CCM)³², a
429 technique that can identify and quantify causal links between two variables (time-series), even

430 if the variables are not separable and if links are very weak or non-linear. To achieve this, CCM
431 first constructs co-called “shadow manifolds” of the two variables, which summarize their past
432 behavior over time. CCM then establishes whether there is a causal relationship between the
433 variables by finding corresponding points in the shadow manifolds of these variables (via
434 “cross-mapping”), and testing whether the shadow manifolds “converge”, i.e., whether
435 information from the causal variable has been embedded in the effect variable.

436 Given that the Landsat products analyzed in this study have varying, and sometimes very short
437 time-series lengths (the GLAD has a length of 5, whereas the TMF and the GSW have lengths
438 of ≥ 30 years), we used a further development of the original CCM method³³, that uses dewdrop
439 regression to combine information from multiple short time-series from similar systems,
440 leveraging time-series information from multiple pixels to identify causal effects. To this end,
441 we ran the CCM analyses for each case study (i.e., the different data limitation biasing
442 perceptions in different land changes) on a country-by-country basis. We could country-level
443 aggregations as this is the typical scale of SDG progress reporting⁵⁹.

444 **Deforestation year analyses.** Our case study on annual deforestation patterns is based on
445 per-pixel deforestation and degradation years mapped by the Tropical Moist Forest dataset¹⁹.
446 According to the developers, these data inform on the first change year between 1982 and
447 2021. To assure comparability with our quality metrics, we aggregated these data from their
448 native 30-m to 1-km resolution, by summing corresponding pixel area.

449 We then evaluated the ability of the TMF dataset to map deforestation trends as reported by
450 national statistics³¹. Because statistics are reported in 5 to 10-year intervals, we matched the
451 TMF to the same temporal resolution. Then, we quantified changes between consecutive
452 years, and then compared subsequent change magnitudes. Here, if the TMF accurately depicts
453 the evolution of forest change, we would expect that decreases in change magnitudes reported
454 by statistics are followed by similar decreases in the TMF. Using these data, we identified
455 disagreements in the direction of change.

456 Additionally, we used formal causal analyses (using CCM, *see previous section*) to analyse
457 effects of variation in the maximum annual Landsat data quality on perceived deforestation
458 timings. Our hypothesis was that abrupt improvements in data quality after extended periods
459 without quality data cause disproportionately high perceived deforestation rates in following
460 years, due to the backlog of deforestation events that accumulated over the data-gap year(s)
461 that suddenly became perceivable. We thus tested for effects of positive/negative outliers in
462 data quality on positive/ negative forest-change outliers, using the data of annual forest-change
463 areas and annual maximum image qualities. For this, we calculated magnitude of differences
464 between annual values and multi-annual trendlines, as derived by Locally Weighted Scatterplot

465 Smoothing (LOWESS)⁶⁰ using a span of 0.4. We focused these analyses on pixel time-series
466 that experienced ≥ 1 instance of forest-cover changes between 1990 and 2020, as well as one
467 or more years without usable data.

468 **Arable-land change analyses.** Our case study on arable-land changes is based on a global
469 dataset on the Global Cropland Extent dataset (GLAD)¹⁸, which maps arable-land extents in
470 subsequent four-year epochs between 2000 and 2019. To match these data to our quality
471 metrics, we aggregated the 30-m GLAD data to a 1-km resolution time-series of per-pixel areas
472 of arable-land, and recalculated the quality-weighted image frequencies for the same four-year
473 epochs.

474 We evaluated the ability of the GLAD product to detect arable-land gains and losses as
475 reported in FAO national statistics on the area of arable-land⁶¹. Specifically, we adjusted the
476 FAO-reported areas of arable-land by subtracting the reported areas of temporary pastures
477 and meadows, following recent recommendations⁶² by FAO data experts that this adjusted
478 metric most closely corresponds to the “cropland” definition adopted by the GLAD product. To
479 compare the FAO statistics to the GLAD data, we first aggregated the annual FAO values to
480 per-epoch maximum areas for each country. We chose maxima because the GLAD product
481 maps any arable-land occurrences perceivable at any points within a given epoch, making
482 GLAD data sensitive to maximum extents. We averaged the quality-weighted frequencies over
483 each epoch, and over each country to match the national scale of the FAO data, focusing only
484 on pixels where the GLAD product maps some arable-land at any point during the full
485 observation period.

486 Using these data, we tested our hypothesis that differences in quality-weighted Landsat data
487 frequencies cause bias in perceived changes in arable-land extents in the Landsat-based
488 GLAD product, causing disagreements with changes inferred from FAO data. Specifically, we
489 used CCM to test whether national-scale qualitative disagreements between GLAD and FAO
490 data in the directions of arable-land changes between subsequent epochs are causally linked
491 to changes in the quality-weighted image frequencies. To this end, we focused on pixels in
492 each epoch and country where the GLAD product mapped arable-land gains – the dominant
493 source of disagreement between the GLAD product and national statistics (**Fig. 3c**) – and
494 where we recorded increases in quality-weighted image frequencies.

495 **Surface-water change analyses.** We based our case study on seasonal surface-water
496 dynamics on the Global Surface Water dataset (GSW)¹⁷, which classifies annual occurrences
497 of permanent and seasonal water. These data are derived from monthly maps on water
498 occurrences, and seasonal water is distinguished from permanent whenever the presence of
499 water is intermittent. Given the dependency of these data on monthly observations, we

500 compared long-term changes in surface water to annual variations in the number of months
501 per year with ≥ 1 usable Landsat observation (**Fig. 3e**).

502 Additionally, we evaluated the plausibility long-term trends in surface-water changes mapped
503 by the GSW dataset relative to those depicted by data on water discharge recorded by gauge
504 stations⁶³. Because upstream surface water limits discharge⁶⁴, long-term changes in discharge
505 are likely to be correlated with true surface-water changes in upstream areas. To allow for
506 comparisons with annual surface water extents, we derived the annual maximum values of
507 discharge for each data year in each gauge station. The maximum annual value conceptually
508 fits the GSW data, which also maps maximum surface water extents.

509 We quantified surface water extents in each year with available discharge measurements over
510 the hydrological basins upstream of the respective gauge stations, using the spatial information
511 on (sub-)basin extents and their water-flow connections from the HydroSHEDS dataset⁶⁵.
512 Specifically, we quantified the proportions of upstream hydrological (sub-)basin extents that,
513 according to the GSW, were covered with seasonal surface water in each year. Here, we
514 excluded gauge stations placed downstream of dams⁶⁶, which may disrupt the natural links
515 between upstream surface-water extents and downstream river flows. For each of the
516 remaining 2,561 stations (**Extended Data Fig. 7a**), we then smoothed the discharge and
517 upstream surface water time-series using LOWESS. We then applied Mann-Kendall tests to
518 determine if both surface water and discharge experienced significant change trends ($p < 0.05$).
519 These tests are sensitive to abrupt changes in variables that are related to data gaps.
520 Therefore, when extracting the surface-water proportions for each (sub-)basin and year with
521 gauge-station data, we excluded any pixels without any usable observations in the respective
522 year.

523 We used CCM to test out hypothesis that changes in seasonal Landsat data completeness
524 bias perceptions of changes in seasonal surface-water extents, as observing the land surface
525 during more months of a given year increases the likelihood of detecting short-lived seasonal
526 water occurrences and of accurately distinguishing longer-lived seasonal from permanent
527 water bodies. Specifically, we tested for effects of positive/negative outliers in annual numbers
528 of months with usable Landsat data on positive/negative outliers in seasonal surface areas.
529 We calculated the magnitude of differences between annual values and trendline derived with
530 LOWESS. When applying this test for each country, we focused on pixels with seasonal
531 surface water observed at any point during the observation period.

532 **Comparisons between lower- and higher-income countries.** We evaluated whether there
533 are asymmetries in both data-quality limitations and in the related biases in perceived land
534 changes between higher- and lower-income regions of the world, as identified in World Bank

535 data⁶⁷. For simplicity, we defined only two income groups of countries, aggregating countries
536 classified as either 'low-income' or 'lower-middle-income' into a combined 'lower-income'
537 group, and those classified as either 'upper-middle-income' or 'high-income' into a combined
538 'higher-income' group. In all income-groups comparisons, we used a threshold of $p < 0.05$ to
539 identify significant differences.

540 For the income-group comparisons of different quality metrics, we compared average per-pixel
541 values across each group's entire land area (i.e., not distinguishing countries within the
542 respective groups). We report area-weighted mean values for compared quality metrics. Here,
543 we used the Wilcoxon-Signed-Rank⁶⁸ test to evaluate differences in means among each
544 income group for each quality issue. To account for spatial autocorrelation, we matched pixels
545 from the lower-income regions to the spatially closest pixels of the higher-income regions (with
546 replacement) using a nearest neighbour approach⁶⁹, and excluded all non-matched pixels from
547 these tests. Given this matching step, these comparisons consistently involved 167,494,453
548 and 53,134,895 numbers of pixels in the lower- and high-income groups, respectively.

549 For the income-group comparisons of bias-causing effects of data limitations on perceived land
550 changes (based on the country-level results of the CCM analyses). Specifically, we used
551 McNemar's tests⁷⁰ (a paired, non-parametric Chi-square test for categorical samples) to
552 evaluate whether the prevalence of Landsat-driven land-change biases identified at country-
553 level differed between lower-income and higher-income groups of countries. To this end, we
554 first classified country-level p -values of the corresponding CCM analyses as 'significant'
555 ($p < 0.05$) or 'non-significant' ($p > 0.05$). As with the previous tests, we used matching to account
556 for spatial autocorrelation, but here matching each higher-income country (with a population
557 size of 133) to its spatially closest neighbour lower-income country (with a population size of
558 83 countries) based on country centroids. Given this matching step, these income-group
559 comparisons were consistently performed with a sample size of 133, except for our example
560 on forests. This is because the TMF dataset is restricted to the moist tropics. Here, higher-
561 income countries had a population size of 36, compared to 41 for lower-income countries.

562 **Data availability**

563 All data underlying the findings of this manuscript are available through the associated Figshare
564 repository, with the exception of the original Landsat quality metrics, which are provided
565 through a dedicated repository (<https://doi.org/10.5281/zenodo.7901148>).

566 **Code availability**

567 The code used in the causal analysis of each example dataset is provided within the
568 supplementary material. In turn, the code used to calculate Landsat quality metrics is
569 available through a dedicated GitHub repository (<https://github.com/RRemelgado/ltqa>).

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577 **Author contributions**

578 RR and CM designed the study. RR developed the analysis with support from CM and CC. RR
579 developed the data underlying the study and ran the analysis. RR and CM designed the figures.
580 RR, CM, and CC interpreted the result and wrote the paper.

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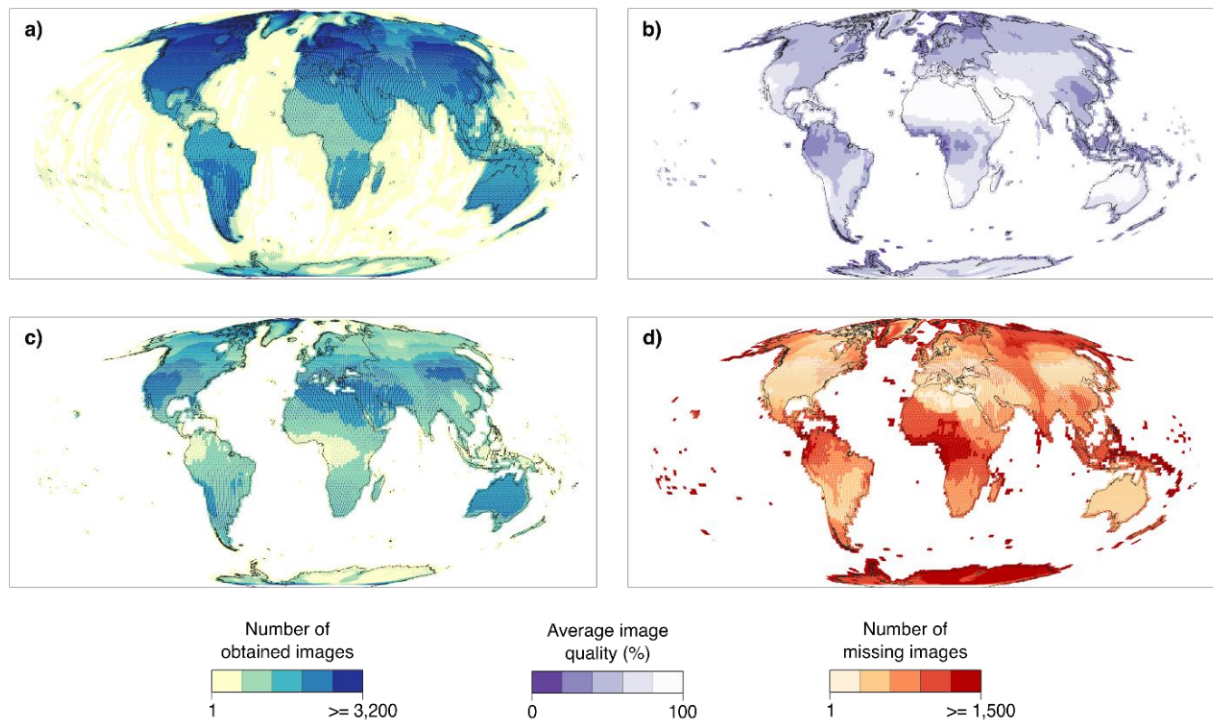
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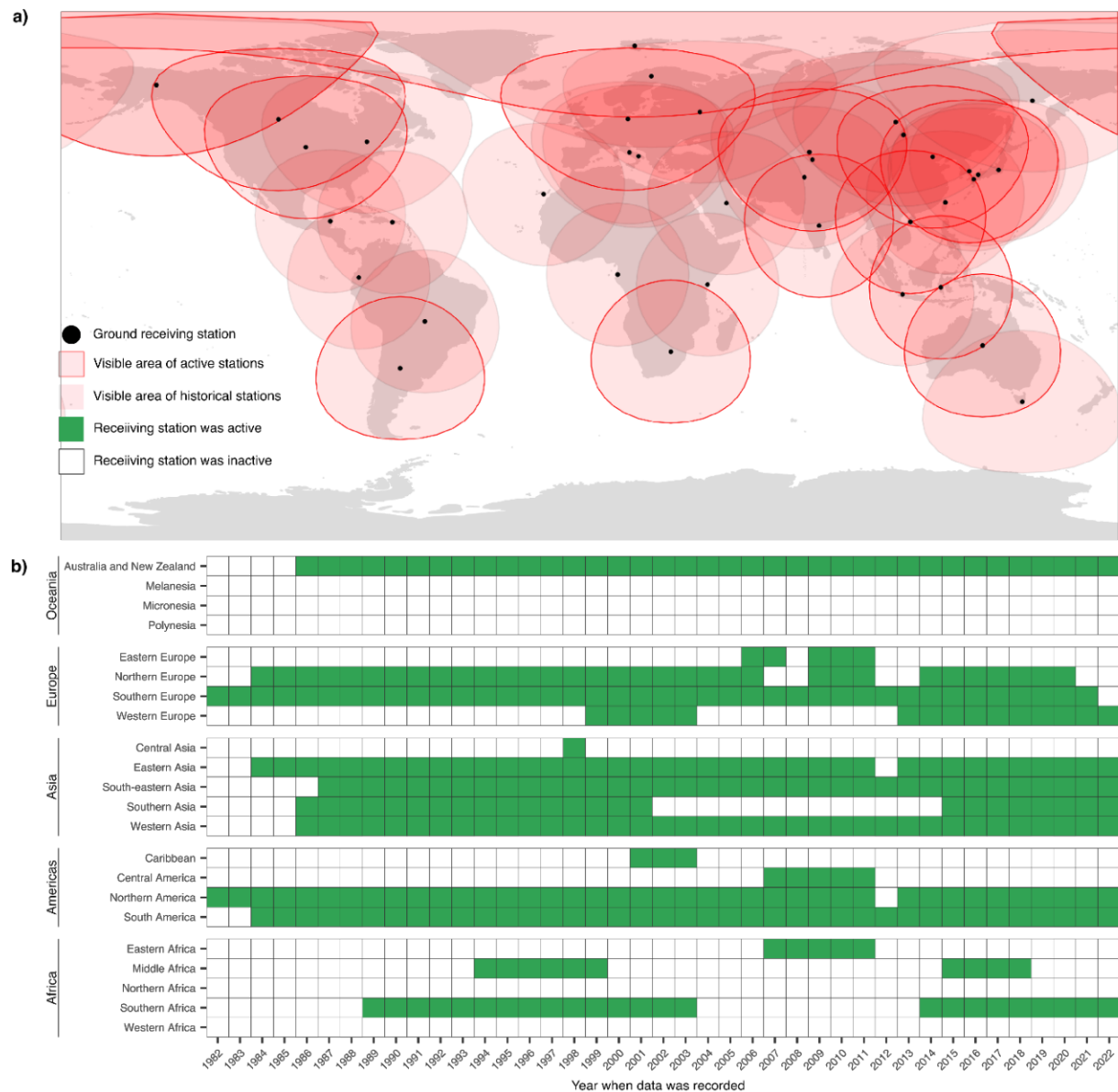
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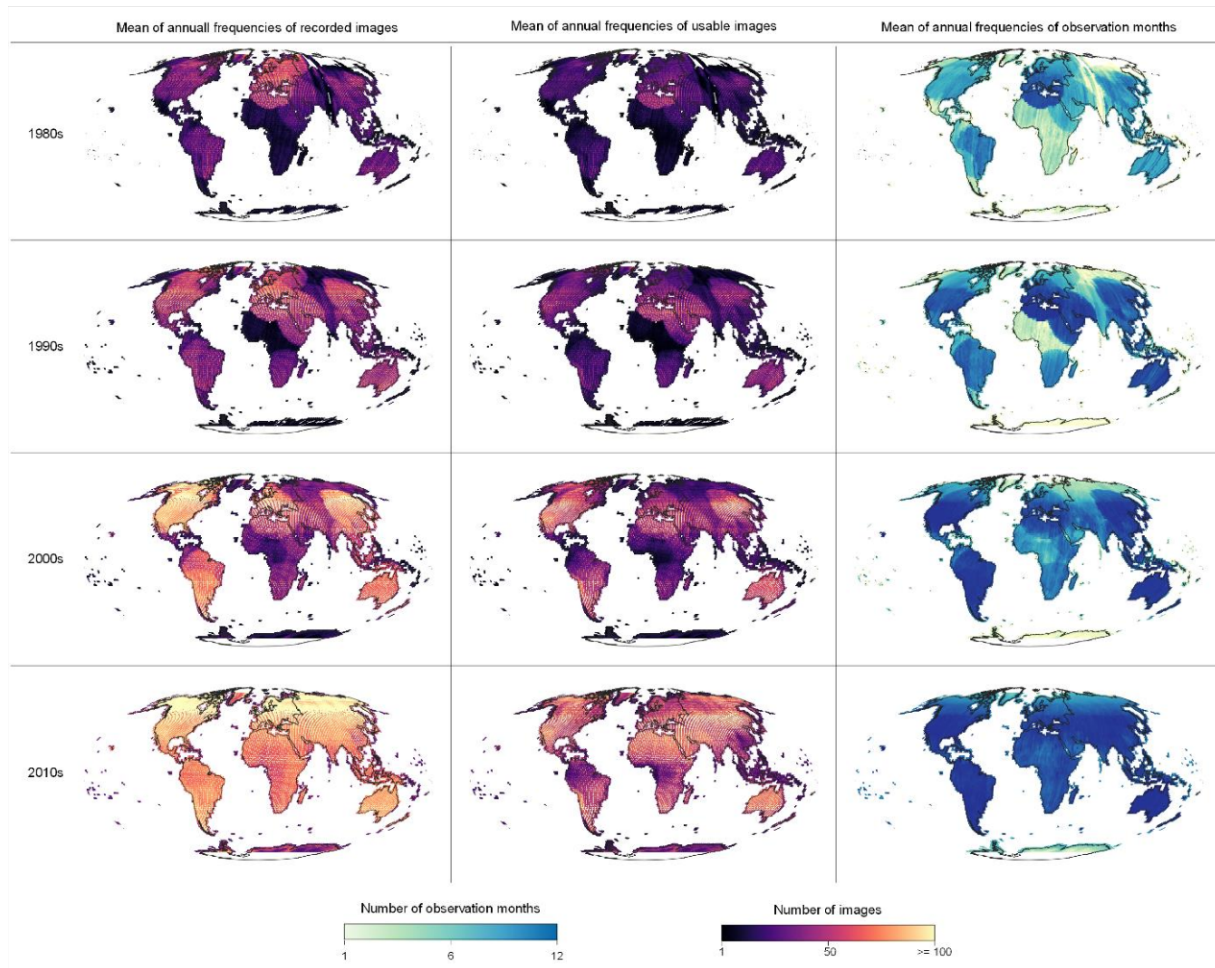
729 **Supplementary material**





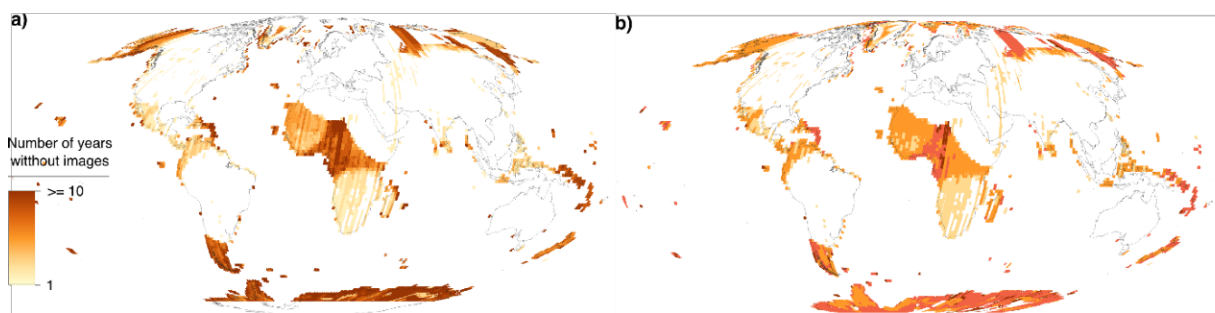
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739 **Extended data figure 2. Spatial and temporal coverage of Landsat data receiving stations.** a) Each of the 39
 740 original Landsat receiving stations (black markers) had a visibility circle of ~2,700 km (shown in pink)
 741 within which data was received from overpassing satellites and stored. Only 14 stations are still active as of
 742 2022 (pink with red outlines), covering most of North America, continental Europe, and Oceania, and portions of South America,
 743 South Africa, and Russia. Meanwhile, Central, Western, and Eastern Africa, as well as Central America and the
 744 northern edges of South America, are now no longer covered by any receiving stations, thus relying fully on the
 745 satellites' on-board storage. b) Temporal variation in data receiving capabilities within different geographical regions.
 746 In Asia and North America, most regions have a near complete coverage between 1982 and 2022, the period during
 747 which Landsat TM, ETM, and OLI missions were active. Still, some sub-regions are covered poorly (between 1-5
 748 years), such as Central Asia, the Caribbean, and Central America. In Oceania, small island territories in Melanesia,
 749 Micronesia and Polynesia lack receiving capabilities. In the same continent, while Australia and New Zealand
 750 compose one region, only Australia actively collected data.



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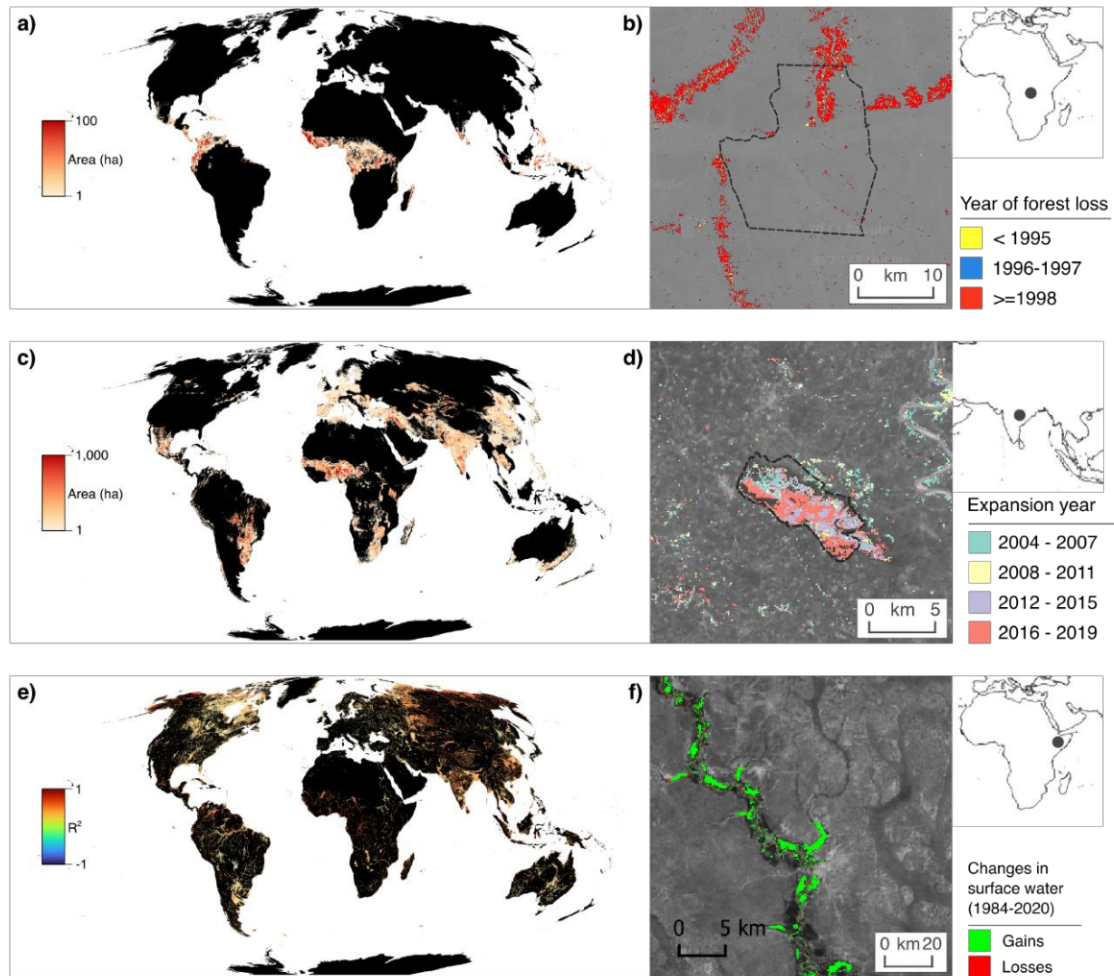
752 **Extended Data Figure 3. Decadal changes in Landsat data coverage and quality.** Historical variation in the
 753 within-year quality of daytime Landsat images for three quality metrics across four decades. **Left column:**
 754 average number of images available for downstream applications. **Middle column:** the same quantity,
 755 weighted by the quality of individual images. **Right column:** number of months with usable images.



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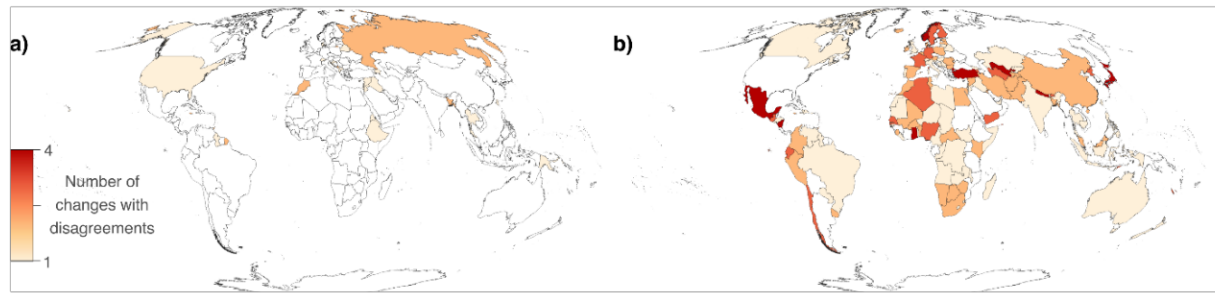
757 **Extended data figure 4. Extended interruptions in temporal Landsat data coverage.** a) Total number of years
 758 lacking any usable Landsat images, showing interruptions in year-to-year continuity of the Landsat archive. Only
 759 data gaps registered after the first imaging year are considered. b) Largest number of consecutive years without
 760 any data (after first imaging year).

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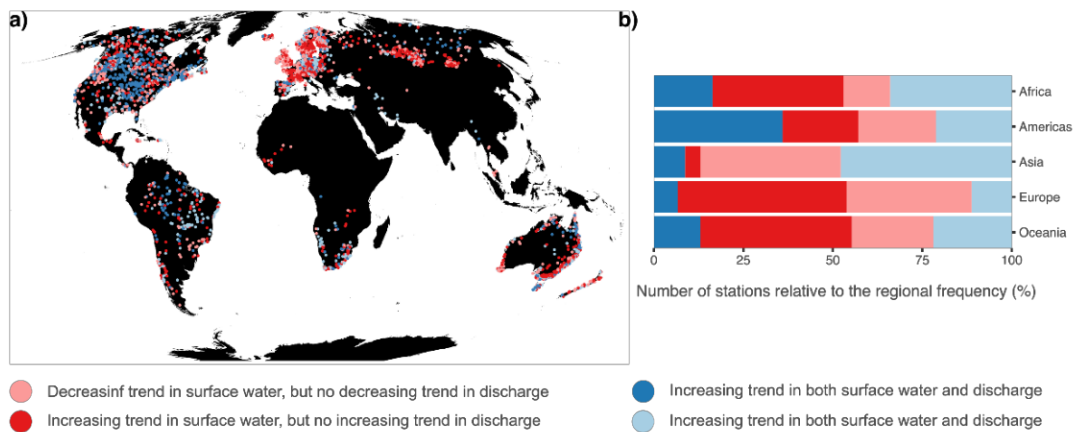


762
 763 **Extended Data Figure 5. Land change areas in doubt, due to associated changes in Landsat data coverage**
 764 **and quality.** Panels **a)**, **c)**, and **e)** map areas of doubtful changes (compare Fig. 3) at a 10-km resolution. Panels
 765 **b)**, **d)**, and **f)** showcase local examples where Landsat data limitations affect perceived land changes, as inferred
 766 from Landsat-based time-series products. **a)** Doubtful areas of deforestation mapped into years directly following
 767 ≥ 1 -year periods without data, casting doubt on true timing of deforestation events. **b)** Forest losses mapped by the
 768 TMF inside/surrounding the Luo Scientific Reserve, DRC (delineated by dashed black line) during the 1990s,
 769 showing that deforestation is mostly mapped into period after the end of the first Congo war (1996-1997), when
 770 Landsat data were again available, even though in-situ evidence largely attributes this deforestation to human
 771 displacement during the war. **c)** Doubtful arable-land gains mapped by the GLAD product in countries/periods for
 772 which FAO statistics report losses, which simultaneously, the quality-weighted frequency of Landsat images
 773 improved, thus likely representing overestimations by the Landsat-based GLAD product. **d)** Arable-land gains
 774 mapped by the GLAD product inside the Bakhira Wildlife Sanctuary (a protected Ramsar wetland and Important
 775 Bird Area in Northern India). Nearly the entire wetland area is re-classified into arable-land by the GLAD mapping
 776 algorithm over the two mapping epochs following the 2013 launch of Landsat 8, which massively increased coverage
 777 and quality of images, which moreover have impaired interoperability with images from earlier Landsat missions. By
 778 inspecting historical time-series of high-resolution imagery (using Google Earth Pro), we found these changes to be
 779 entirely artefactual. **e)** Doubtful increases in global, seasonal surface-water extents that may reflect increasing
 780 Landsat data coverage and quality. Colors indicate directions and strengths of (nearly exclusively positive) statistical
 781 associations of multi-annual trends between *i)* proportions of 10-km pixels that are perceived by the GSW product
 782 as being covered by seasonal surface water, and *ii)* annual numbers of months with usable Landsat observations.

783 f) Seasonal surface-water gains mapped by the GSW regionally misrepresent true surface-water losses, such as
 784 those along the drying Jubba river (Somalia), caused by droughts within the Horn of Africa.



785
 786 **Extended Data Figure 6. Disagreeing reports of arable-land extents.** Both maps show frequencies of changes
 787 in the GLAD product that disagree in change direction with changes reported in national FAO statistics on arable-
 788 land areas. **a)** Instances where GLAD maps losses whereas statistics indicates gains. **b)** Instances where GLAD
 789 maps gains whereas statistics map losses. Comparing **a)** and **b)** shows that whereas cases where GLAD potentially
 790 mischaracterizes true arable-land gains as losses are rare, those cases where GLAD potentially mischaracterizes
 791 true losses as gains are common and occur globally.



792
 793 **Extended Data Figure 7 – Comparison of trends in seasonal surface water between GSW and gauge station**
 794 **data.** **a)** Locations of gauge-station records of water-discharge changes that were compared against changes in the
 795 percentages of upstream hydrological (sub-)basins that are covered by seasonal surface water according to the
 796 Landsat-based GSW product. Colours indicate qualitative agreement/disagreement on significant directional trends
 797 ($p < 0.05$, based on Mann-Kendall tests) between discharge measurements (gauge stations) and seasonal surface-
 798 water data (GSW). Bright red: perceived increasing trends in surface-water cover are *not* supported by increasing
 799 discharge trends; pale red: perceived decreasing trends in surface-water cover *not* supported by decreasing
 800 discharge trends; bright blue: perceived decreasing trends in surface-water cover *supported* by decreasing
 801 discharge trends; pale blue: perceived increasing trends in surface-water cover *supported* by increasing discharge
 802 trends. **b)** Percentages of hydrological (sub-)basins per continent showing each type of agreement/disagreement
 803 between recorded water-discharge (gauge stations) and perceived seasonal surface-water trends.