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#### Using Detrending to Assess SARS-CoV-1

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#### Sewersheds in North Carolina 6

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# 32 Abstract

33 Wastewater surveillance emerged during the COVID-19 pandemic as a novel strategy for tracking the burden of illness in communities. Previous work has shown that trends in 34 35 wastewater SARS-CoV-2 viral loads correlate well with reported COVID-19 case trends over 36 longer time periods (i.e., months). We used detrending time series to reveal shorter sub-trend 37 patterns (i.e., weeks) to identify leads or lags in the temporal alignment of the wastewater/case 38 relationship. Daily incident COVID-19 cases and twice-weekly wastewater SARS-CoV-2 viral 39 loads measured at 20 North Carolina sewersheds in 2021 were detrended using smoothing 40 ranges of ∞, 16, 8, 4 and 2 weeks, to produce detrended cases and wastewater viral loads at progressively finer time scales. For each sewershed and smoothing range, we calculated the 41 42 Spearman correlation between the cases and the wastewater viral loads with offsets of -7 to +7 43 days. We identified a conclusive lead/lag relationship at 15 of 20 sewersheds, with detrended 44 wastewater loads temporally leading detrended COVID-19 cases at 11 of these sites. For the 11 45 leading sites, the correlation between wastewater loads and cases was greatest for wastewater 46 loads sampled at a median lead time of 6 days before the cases were reported. Distinct lead/lag 47 relationships were the most pronounced after detrending with smoothing ranges of 4–8 weeks. 48 suggesting that SARS-CoV-2 wastewater viral loads can track fluctuations in COVID-19 case 49 incidence rates at fine time scales and may serve as a leading indicator in many settings. These 50 results could help public health officials identify, and deploy timely responses in, areas where 51 cases are increasing faster than the overall pandemic trend.

# 52 Introduction

53 The first lab-confirmed COVID-19 case was reported in North Carolina (NC) on March 3. 54 2020, and over the next two and a half years, the number of reported positive cases statewide 55 increased to more than three million [1,2]. However, the true burden of disease far exceeded 56 this number due to underreporting, access to testing, unreported at-home tests, asymptomatic 57 illness and other factors [3-5]. Testing was not uniformly distributed among populations due to 58 unequal availability and pervasive mistrust of public health recommendations by historically 59 marginalized persons [6-8]. As a result, there is need for non-clinical means of tracking COVID-60 19 trends to augment case-based reporting. 61 One promising approach is wastewater-based epidemiology (WBE), which measures 62 substances shed in human feces and derived from a condition of interest, such as pathogen 63 nucleic acids or pharmaceutical metabolites, by sampling sewage containing human fecal waste 64 and byproducts of water usage [9]. WBE has been increasingly utilized to track COVID-19 infection trends at the community level by guantifying SARS-CoV-2 RNA in sewage. Twice-65 66 weekly testing of SARS-CoV-2 loads in wastewater can provide information on changes in COVID-19 burden in the sewershed population and can be used as a method to detect periods 67 68 of increasing COVID-19 cases from far fewer samples than required for clinical case reporting 69 since wastewater samples represent pooled samples of multiple individuals [10]. Unlike case-70 based surveillance, wastewater surveillance does not rely on individual healthcare-seeking 71 behavior or access to testing, which are strongly impacted by well-documented societal 72 inequities [11]. Additionally, SARS-CoV-2 is shed in the feces of both symptomatic and 73 asymptomatic individuals, allowing the capture of data on a range of infected individuals [12–14] 74 at varying stages of infection. Numerous studies have shown that when clinical testing coverage 75 is high, wastewater SARS-CoV-2 loads and documented COVID-19 cases follow similar trends 76 and are highly correlated [15–18]. Therefore, given the cost and human resource savings, WBE

may provide an effective complement to case-based surveillance that addresses some of thelimitations of traditional clinical surveillance approaches.

79 However, the values typically measured in wastewater, such as viral genome copies per 80 liter, are not directly interpretable in terms of familiar population health metrics, like the 81 prevalence or incidence rate of infection in a defined population. To effectively inform public 82 health response and mitigation strategies using WBE, it is necessary to relate wastewater-83 based measurements to interpretable population-level metrics. One critical aspect is the 84 temporal relationship between SARS-CoV-2 wastewater loads measured at a wastewater 85 treatment plant (WWTP) and reported COVID-19 cases in the corresponding sewershed served 86 by the plant [5,19]. Past work has demonstrated that increases in SARS-CoV-2 wastewater 87 loads may occur prior to a rise in lab-confirmed sewershed COVID-19 cases in a sewershed, 88 allowing for WBE to be used as an early warning system [4,20–22]. Such leading signals in 89 wastewater were reported during the earlier phases of the pandemic in some North Carolina 90 sewersheds [10,23] as well as during more recent pandemic phases [24].

91 As the pandemic becomes endemic, trends lasting several months have been widely 92 reported to anticipate trends in COVID-19 infections, as later indicated by population 93 surveillance metrics [21,22,25,26]. However, the time alignment between trends in wastewater 94 load and trends in cases can be difficult to determine since its small temporal lead or lag may be 95 eclipsed by the longer time scale of trends. In this situation, kernel detrending can be used to 96 remove these longer pandemic trends and reveal shorter-term fluctuations that may help identify 97 leads or lags in the temporal alignment of the detrended wastewater and detrended case 98 relationship [27–31]. While the correlation between wastewater-based measurements of 99 pathogens of concern and clinical cases over longer time periods (i.e., months) is useful for 100 informing longer-term public health response, much less is known about short-term sub-trends 101 (i.e., weekly or even daily), which may be more relevant for ongoing, day-to-day public health 102 decision making. Therefore, there is a need for research to better understand and anticipate

103 changes in COVID-19 incidence on shorter time scales to inform timely, targeted, and cost-104 effective public health action, particularly at the local level. Detrending the wastewater and case 105 data is done by modeling these longer-term trends and removing them to obtain detrended 106 wastewater loads and detrended cases, also referred to as wastewater load residuals and case 107 residuals, respectively. If wastewater load residuals can predict a fine-scale fluctuation in case 108 residuals, then public health measures can be taken *locally* and for *short* durations in 109 sewersheds where cases are anticipated to rise at levels greater than that of the baseline trend. 110 This methodology may also be applicable for other pathogens beyond SARS-CoV-2 as 111 wastewater surveillance expands to new targets in the future. 112 Our work aims to contribute to previous studies by refining the time scale at which 113 correlations between wastewater and cases are assessed. Accordingly, we investigate the 114 temporal relationship (i.e., lead or lag) that maximizes correlation between detrended 115 wastewater SARS-CoV-2 viral loads and detrended COVID-19 clinical cases at the finest time-116 scale possible for 20 sewersheds across North Carolina in 2021. Furthermore, to operationalize 117 this approach, we propose and validate a set of reproducible criteria that can be easily deployed 118 by public health agencies to support the application of WBE approaches beyond North Carolina. 119

120 Materials and Methods

### 121 Ongoing Wastewater-Based Epidemiology in North Carolina

In collaboration with University of North Carolina (UNC) system researchers, the North
Carolina Department of Health and Human Services (NCDHHS) was one of eight state health
departments initially funded by the Centers for Disease Control and Prevention (CDC) to
participate in the National Wastewater Surveillance System (NWSS). The NCDHHS NC
Wastewater Monitoring Network is a multi-disciplinary collaboration between epidemiologists,

127 laboratory scientists, water reclamation managers, environmental engineers, and public health 128 officials with promising applications for genomic, large-scale pathogen monitoring, as well as 129 COVID-19. The development of this state surveillance network benefited from a collaboration 130 funded by the North Carolina State Legislature among North Carolina universities at the start of 131 the pandemic in 2020. This group of experts created the NC Wastewater Pathogen Research 132 Network to develop sampling techniques, laboratory capabilities, and analysis of SARS-CoV-2 133 in wastewater [32]. The NC Wastewater Pathogen Research Network, in collaboration with 134 NCDHHS, established a strong foundation for WBE, and founding contributors continue to be 135 essential partners in the NC Wastewater Monitoring Network using a framework of innovative 136 research to inform public health surveillance and action in North Carolina.

137 As part of the NC Wastewater Monitoring Network data collection in 2021, wastewater 138 samples were collected twice per week by WWTP staff and shipped to the UNC-Chapel Hill 139 Institute of Marine Sciences (IMS, Morehead City, NC) for laboratory analysis. Samples were 140 analyzed for SARS-CoV-2 by reverse-transcription droplet digital polymerase chain reaction 141 (RT-ddPCR) following a standardized protocol [33], for which additional details are provided in 142 the Supplementary Material [34]. Sewer network spatial data (e.g., gravity mains, force mains, 143 manholes, pump stations) obtained from North Carolina wastewater utilities and local 144 geographic information systems departments were used to delineate a sewershed polygon 145 using ArcGIS Pro 2.8 (ESRI, Redlands, CA). COVID-19 clinical cases reported to NCDHHS 146 were geocoded in ArcMap 10.7.1 (ESRI) and matched to the sewershed within which they 147 resided using a custom composite geocoder built from state and county address data. Lastly, 148 wastewater sample results and recorded clinical cases in the sewershed were submitted to 149 NCDHHS and uploaded weekly the CDC NWSS analytics platform for epidemiologic trend 150 analysis. COVID-19 cases were given a date based on the following hierarchy: date of symptom onset, date of specimen collection, and date of result. Daily incidence rates per 100,000 151 152 estimated sewershed population were calculated. Wastewater sample results were normalized

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to flow within each municipal utility to represent a 24-hour viral load. These analyzed data are
posted publicly on the CDC COVID-19 Data Tracker and the NCDHHS COVID Dashboard
(https://covid19.ncdhhs.gov/dashboard/wastewater-monitoring).

#### 157 Relating Wastewater Loads and COVID-19 Incidence

158 During a ten-month study period from January 2021 through October 2021, we 159 compared SARS-CoV-2 viral loads in influent wastewater collected at the 20 WWTPs in the NC 160 Wastewater Monitoring Network with COVID-19 incidence in the corresponding sewersheds. 161 Nine sites were sampled for the entire duration of the study period, two sites were sampled 162 beginning in January and ending before October 2021, and nine sites were added in the 163 summer and sampled from June 2021 through October 2021 (Table 1). We retrieved calculated 164 wastewater viral loads and clinical COVID-19 incidence rates in the sewershed for each of the 165 20 sites from the CDC NWSS analytics platform. Twice-weekly wastewater loads were provided 166 as the sample-specific geometric mean of measured N1 and N2 target copy numbers per liter 167 (L) of wastewater [35], normalized by multiplying by the average daily flow and dividing by the 168 estimated sewershed population. Half the target-specific limit of detection (LOD) was substituted 169 for the concentration when a target was not detected in the sample (see Supplemental 170 Material). The resulting population-normalized viral loads, with units of SARS-CoV-2 N gene 171 copies (GC) per person per day (pppd), were  $\log_{10}$ -transformed for all analyses, which were 172 conducted in R version 4.1.2 [36].

WWTP Name	Population (2019)	Area (km²)	Capacity (ML/day)	First Sample	Last Sample	Number of Samples
Newport	3,731	6.1	5	1/12/2021	8/18/2021	64
Pittsboro	3,799	10.3	3	1/5/2021	5/25/2021	39
Beaufort	3,992	7.4	7	1/5/2021	10/20/2021	83
Marion	7,793	22.9	14	6/17/2021	10/14/2021	35
Laurinburg	15,407	37.4	18	6/17/2021	10/19/2021	36
Roanoke Rapids	19,335	43.9	38	6/19/2021	10/20/2021	33
Wilson	51,285	164.4	64	6/19/2021	10/19/2021	33
New Hanover Co.	51,401	81.4	48	1/22/2021	10/20/2021	78
Wilmington	65,081	62.5	25	1/5/2021	10/20/2021	79
Charlotte 1	77,278	126	55	1/5/2021	10/19/2021	76
Chapel Hill	84,729	89.8	66	1/6/2021	10/20/2021	81
Greenville	94,194	95.2	80	1/5/2021	10/20/2021	81
South Durham	98,068	100.7	91	1/6/2021	10/20/2021	81
Charlotte 3	122,063	122.2	55	6/3/2021	10/19/2021	38
Greensboro	144,539	143.6	82	6/18/2021	10/20/2021	36
Charlotte 2	154,519	105.3	127	1/4/2021	10/19/2021	80
Fayetteville	159,000	250.8	95	6/19/2021	10/20/2021	36
Winston Salem	177,520	319.6	70	6/19/2021	10/20/2021	33
MSD of Buncombe Co.	188,927	534.4	182	6/19/2021	10/20/2021	33
Raleigh	551,534	536.7	341	1/6/2021	10/20/2021	74

173	Table 1.	Characteristics	of NC	Wastewater	Monitoring	Network Sites

174

175 Exponential kernel smoothing is a technique used in space/time geostatistics to estimate 176 spatial and temporal trends of environmental and health processes at a variety of spatial and 177 temporal scales [27-31]. Here, we used exponential kernel smoothing to estimate trends in 178 wastewater viral loads and COVID-19 incidence rates at different temporal scales. For each 179 observed response, a smoothed estimate was obtained as the average of all observations 180 weighted by an exponentially decaying function of the temporal distance from the estimation 181 time point. The rate of exponential decay was determined by a smoothing range parameter, 182 corresponding to the temporal duration below which variations in the response are smoothed 183 out of the mean trend to retain only those variations of greater duration than the smoothing

184 range. For a response y(t) observed at time t, the smoothed estimate was obtained as the

185 mean trend  $m_v(t;T)$  with smoothing range of duration T:

$$m_y(t;T) = \sum_{j=1}^{N} k_j y(t_j) \tag{1}$$

186 where  $y(t_j), j=1,...N$ , are the observations at observation times  $t_j$  and the exponential kernel 187 smoothing weights  $k_i$  are given by

.

Sincolining weights 
$$k_j$$
 are given by

$$k_{j} = \frac{exp\left(\frac{-3|t_{j} - t|}{T}\right)}{\sum_{j=1}^{N} exp\left(\frac{-3|t_{j} - t|}{T}\right)}$$
(2)

188 Scaling the exponential decay function by -3 ensured that the influence of observations with 189 temporal distance equal to the smoothing range T was diminished by ~95%, with the estimation 190 point itself receiving the highest weight. As T increased, observations further away in time were 191 allowed greater influence on the mean trend, increasing the extent of smoothing until 192 converging to a constant value at the arithmetic mean of all the data for T of infinite duration. 193 As the mean trend  $m_v(t;T)$  only retained variations in the response of greater duration 194 than the smoothing range T, we detrended the observed responses by subtracting the mean 195 trend estimated at time t to obtain the residual response:

$$\tilde{y}(t;T) = y(t) - m_y(t;T) \tag{3}$$

196 which captured the fluctuations around the trend at temporal scales shorter than the smoothing 197 range T (including any measurement error). In short, we decomposed the signal  $y_i(t)$  into a time 198 trend  $m_{y_i}(t;T)$  that captured variation of time scales greater than T and a detrended signal  $\tilde{y}_i$ 199 (t;T) that captured fluctuations of time scale shorter than T, corresponding to the shorter-term 200 variations around pandemic trends that are of particular relevance to timely public health action. 201 To examine the time scales at which wastewater signals may lead (i.e., precede) or lag 202 (i.e., follow) clinical cases at North Carolina Wastewater Monitoring Network sites, we evaluated 203 the cross-correlation between *detrended* wastewater viral loads, denoted  $\tilde{w}(t;T)$ , and *detrended*  This manuscript is a preprint and has not been peer reviewed. The copyright holder has made the manuscript available under a Creative Commons Attribution 4.0 International (CC BY) license and consented to have it forwarded to EarthArXiv for public posting.

204 COVID-19 incidence rates  $\tilde{y}(t;T)$  across various detrending kernel smoothing ranges for 205 observations from January – October, 2021. The cross-correlation between two time series was 206 determined as the set of correlations between pairs of observations for different temporal offsets

207  $\tau$ , given by

$$r(\tau;T) = corr(\tilde{w}(t+\tau;T),\tilde{y}(t;T))$$
(4)

for which  $\tau < 0$  indicated the detrended wastewater load signal leads the detrended signal obtained from COVID-19 incidence rates; conversely,  $\tau > 0$  indicated the signal from detrended wastewater loads lags that of detrended COVID-19 incidence.

211 We examined detrended wastewater loads and detrended COVID-19 incidence rates 212 with detrending smoothing ranges of  $T = \infty$ , 16, 8, 4 and 2 weeks separately for each site. 213 Because subtracting a constant does not affect correlation estimates, using the  $T = \infty$ 214 detrended residuals was equivalent to performing the analysis without detrending. As we 215 anticipated nonlinear associations, we estimated Spearman rank correlations to assess the 216 monotonic relationships between the two surveillance systems for temporal offsets ranging from 217  $\tau = -7$  to  $\tau = +7$  days. The optimal combination of detrending smoothing range and temporal 218 offset to characterize the lead/lag relationship between wastewater and incidence over relevant 219 time scales was identified for each site by applying a reproducible set of criteria. For each 220 detrending smoothing range T, starting from  $T = \infty$  down to T = 2, we: 221 1. Identified the span of consecutive lead/lag values  $\tau$  for which  $r(\tau;T)$  was a statistically 222 significant positive correlation. 223 2. Accepted  $\tau$  if (a) it was less than 7 days (identifiable), (b) it lasted at least 2 days 224 (persistent), and (c) it contained the maximum  $r(\tau;T)$  value (predictive). Otherwise, it 225 was rejected and deemed inconclusive.

Finally, the optimal smoothing range was obtained by choosing the shortest detrending
 smoothing range *T* that successfully identified a conclusive lead or lag. Detecting fluctuations

- 228 over a shorter duration is ideal because these can be addressed with more timely public health
- measures. We selected criteria that favor identifiability, persistence, and predictivity; however,
- this framework may easily be extended to additional or alternative criteria as required by the
- 231 specific application.
- 232 This analysis did not involve human subjects in its research.
- 233

## 234 **Results**

#### 235 Charlotte 1 Sewershed Case Study

236 In this case study, we demonstrate the use of kernel detrending in the cross-correlation 237 analysis of SARS-CoV-2 wastewater loads and COVID-19 incidence in the Charlotte 1 238 sewershed. One of three WWTPs in the Charlotte metropolitan area monitored by the NC 239 Wastewater Monitoring Network during the study period, the Charlotte 1 sewershed covers 126 240 km<sup>2</sup> in the northeast of the city and serves approximately 80,000 people. From January to 241 October 2021, 76 wastewater samples were collected at Charlotte 1 with a SARS-CoV-2 RNA 242 detection frequency of 98% and a mean daily load of 9.2 x 10<sup>6</sup> GC pppd. The maximum load 243 was an order of magnitude higher at 4.7 x  $10^7$  GC pppd and the minimum load was 7.9 x  $10^4$  GC 244 pppd. A total of 6,039 COVID-19 cases were reported in the Charlotte 1 sewershed over the 10-245 month study period, with a daily incidence rate of 30 cases/100,000 people on average and a 246 maximum of 132 cases/100,000 people. There was only one day with zero COVID-19 cases 247 reported (0.3%, n = 293 days).

Visual inspection of trends in the Charlotte 1 sewershed indicated the wastewater loads generally mirrored the COVID-19 incidence rates, with a peak in January, a gradual decline through July followed by a sharper increase in August and second peak around September (Figure 1a and 1b). The mean trend was estimated at each time point for smoothing ranges of

252  $T = \infty$ , 16, 8, 4 and 2 weeks. Using  $T = \infty$  resulted in a flat (i.e. constant) trend line. Then, as 253 the kernel smoothing range became finer (i.e. T = 16, 8, 4 and 2 weeks), the trend line captured 254 more of the inflections in the wastewater and case trends.

255 Subtracting the various mean trends from the wastewater and case observations vielded 256 residuals retaining the variation in the observations at time scales shorter than the 257 corresponding smoothing range T. With an 8-week range, the detrended wastewater loads and 258 detrended cases demonstrated lower temporal variability compared to the variability seen 259 without detrending (Figures 1c and 1d). Scatterplots comparing the detrended wastewater loads 260 and detrended cases on the same day (i.e., temporal offset  $\tau = 0$ ) are presented in Figures 1e 261 and 1f for detrending smoothing ranges  $T = \infty$  weeks and T = 8 weeks, respectively. As 262 anticipated, we observed that the pairwise correspondence between detrended wastewater 263 loads and detrended cases declined with decreasing detrending smoothing range (i.e., as T = $\infty$ , 16, 8, 4 and 2 weeks) because more of the pandemic-scale trend was removed and only 264 265 shorter-term fluctuations remained. However, detrended residuals were significantly positively 266 correlated for all detrending smoothing ranges other than T = 2 (the shortest range considered, 267 Spearman's  $\rho = 0.19$ , p = 0.11).

268 We then calculated, for each detrending smoothing range T, not only the correlation for 269 detrended wastewater observations on the same day as each case date ( $\tau = 0$ ), but also for 270 wastewater observations up to 7 days before ( $\tau = -7$ ) and 7 days after ( $\tau = +7$ ) each case date (Figure 1g). Based on our proposed criteria, we determined the shortest smoothing range T to 271 272 conclusively identify a time offset  $\tau$  for predicting detrended cases from detrended wastewater 273 loads in the Charlotte 1 sewershed was T = 8 weeks, which revealed positive correlations for wastewater measured 0 to 3 days before cases were reported. This set of contiguous positive 274 275 correlations spanned more than 2 and fewer than 7 contiguous days and included the maximum 276 correlation value, satisfying our proposed criteria for identifiable and predictive lead/lag

277 relationships. Longer detrending smoothing ranges ( $T = \infty$  and T = 16 weeks) demonstrated 278 significant positive correlations at all temporal offsets, suggesting that the lead/lag relationships 279 were not identifiable because they were dominated by overall pandemic trends that obscured 280 the short-term fluctuations relevant to timely public health action. Conversely, the shorter 4- and 281 2-week detrending smoothing ranges removed so much of the trend that the residuals were not 282 predictive at any contiguous sets of temporal offsets, rendering the lead/lag relationships 283 inconclusive. We therefore concluded that the finest detrending time-scale at which wastewater 284 loads predicted COVID-19 cases in the Charlotte 1 sewershed during our study period—based 285 on our reproducible criteria for identifiability, persistency and predictivity-was 8-weeks, and 286 that the correlation between detrended wastewater loads and detrended cases was greatest for 287 wastewater loads sampled with a lead time of 0 to 3 days before the cases were reported.

288 Figure 1. Kernel smoothing of the (A) SARS-CoV-2 wastewater loads (log GC pppd) and (B) 289 COVID-19 incidence (cases/100k) observed at Charlotte 1 sewershed from January to October 290 2021, using various range parameters indicated by the colored lines in the legend. The 291 smoothed estimates were subtracted from the observations to yield the (C) detrended 292 wastewater loads and (D) detrended cases, shown here for a detrending smoothing range of 8-293 weeks. The pairwise correspondence of the detrended wastewater and case residuals on the 294 same day (i.e. temporal offset of zero) were compared in scatterplots with added spearman 295 correlation lines prior to evaluating any temporal offsets for detrending smoothing ranges of (E) 296  $T = \infty$  weeks and (F) T = 8 weeks. A cross-correlation plot (G) between the detrended 297 wastewater and case residuals was created for each detrending smoothing range and temporal 298 offset to be used with the criteria to assess the lead/lag relationship. Note: The temporal offset 299 values on the x-axis are in relation to the case date, such that negative values indicate the 300 correlation was performed when the wastewater preceded the cases and positive values 301 indicate the correlation was performed when the wastewater lagged the cases. Statistically 302 significant correlations are indicated with a filled-in circle and the intersecting line represents the 303 95% confidence interval.

#### 305 Wastewater Loads and COVID-19 Incidence Across All Sites

306 The observed COVID-19 incidence rates and SARS-CoV-2 wastewater loads varied 307 across the 20 North Carolina sewersheds participating in this study (Figure 2). The sites were 308 distributed across North Carolina, covering approximately 20% of the population and about 2% 309 of the land area. There was a wide range in sewershed size, with the largest sewershed, 310 Raleigh, serving 551,534 people at a capacity of 341 ML/day and the smallest sewershed, 311 Newport, serving 3,731 people at a capacity of 5 ML/day. During the study period, the number 312 of samples collected per site ranged from 33 (Wilson, Buncombe, Roanoke Rapids, and 313 Winston-Salem) to 83 (Beaufort). SARS-CoV-2 RNA was detectable in 74% of the 1,129 314 wastewater samples across all 20 sites. Sewersheds with larger populations tended to have 315 higher detection frequencies, with 50% of all the non-detects occurring at the three smallest 316 sites with populations under 5,000 people. The lowest mean daily load was 5.0 x 10<sup>5</sup> GC pppd, 317 observed at Newport, while the highest mean daily load of  $2.3 \times 10^7$  GC pppd was observed at 318 Fayetteville. The median mean daily load was 7.8 x 10<sup>6</sup> GC pppd, observed at Buncombe 319 County. There was a total of 122,444 COVID-19 cases reported across all 20 sites during the 320 study period, with the average daily incidence rate ranging from 1 case/100.000 people 321 (Pittsboro) to 148 cases/100,000 people (Raleigh). The median daily incidence rate was 16 322 cases/100,000 people, observed at South Durham. Comparable to the wastewater loads, the 323 three smallest sewersheds accounted for almost 75% of the observed days with zero reported 324 COVID-19 cases.

The maximum daily population normalized loads (henceforth referred to simply as loads) for each site ranged from  $4.7 \times 10^6$  GC pppd to  $4.3 \times 10^8$  GC pppd, with most of these values occurring in January or late August/early September, during which peaks in COVID-19 cases were also observed with daily incidence rates as high as 235 cases/100,000 people. For the 10 sites that were sampled for the entire 10-month period, there was also a noticeable lull during

- the period of May to July 2021 for both the wastewater loads and cases. All but one sewershed
- had significant positive correlations between the wastewater loads and cases observed on the
- same day, with the significant Spearman's coefficients ranging from  $\rho = 0.38$  to  $\rho = 0.85$ , with a
- median of  $\rho = 0.72$ . The smallest sewershed (Newport) had a non-significant correlation with a
- 334 coefficient of  $\rho = 0.21$  and p-value of 0.09.

Figure 2. Time series of SARS-CoV-2 wastewater loads (log GC pppd) and COVID-19 incidence (cases/100k) for each of the 20 sites from January through October 2021. Note: COVID-19 incidence is shown as a 7-day rolling average with the blue line. SARS-CoV-2 wastewater loads are depicted with the orange dots and a LOESS curve was fitted to these values, depicted by the orange line (span=0.3).

340

#### 341 Detrending Reveals Short-Term Associations

342	Applying each detrending smoothing range ( $T = \infty$ , 16, 8, 4 and 2 weeks) across
343	temporal offsets ( $\tau = -7$ to $\tau = +7$ days) allowed us to evaluate the lead/lag relationship
344	between the detrended wastewater and case residuals at progressively finer time scales.
345	Correlation plots similar to Figure 1e were generated for all 20 sewersheds, and the proposed
346	criteria were used to identify the optimal detrending smoothing range for each site, which was
347	defined as the shortest kernel smoothing range that revealed an identifiable lead or lag (Figure
348	3). Eighteen of the 20 sewersheds exhibited statistically significant correlation coefficients at all
349	temporal offsets when $T = \infty$ weeks, indicating that additional detrending was needed to reveal
350	the fine time scale fluctuations required for a lead/lag analysis. Beaufort and Pittsboro were the
351	only sewersheds for which the $T = \infty$ weeks range was optimal for identifying the lead/lag
352	relationship; the detrended wastewater and case residuals were no longer significantly
353	correlated over any 2-day span of temporal offsets using shorter detrending smoothing ranges.
354	Of the remaining sewersheds, one site had an optimal detrending smoothing range of $T$
355	= 16 weeks, eight sites had an optimal detrending smoothing range of $T$ = 8 weeks, and four
356	sites had an optimal detrending smoothing range of $T = 4$ weeks (Figure 3). As a general

357 pattern, the detrending smoothing ranges greater than the identified optimal T either had 358 significant positive correlations at all temporal offsets, such that no lead/lag pattern was 359 identifiable, or additional detrending allowed us to detect fluctuations over a shorter duration 360 while still meeting all the proposed criteria. Conversely, too much of the trend was removed 361 when using values for T smaller than the optimal detrending smoothing range, such that the 362 detrended residuals were no longer significantly correlated for any span of contiguous temporal 363 offsets. Five of the 20 sewersheds were deemed inconclusive as none of the detrending 364 smoothing ranges revealed an identifiable lead or lag between the detrended wastewater loads 365 and cases, according to the proposed criteria. We identified two reasons for this: 1) the span of 366 consecutive lead/lag values was longer than 7 days for larger T values (not identifiable) and 367 shorter than 2 days at smaller T values (not persistent), or 2) the longest range of consecutive 368 lead/lag values did not include the maximum correlation coefficient (not predictive). The 369 inconclusive nature of the lead/lag relationship in these sewersheds may be linked to the short 370 sampling duration or the small size of the sewershed as all five sites had data for only half of the 371 study duration and all but Buncombe County were among the smallest sewersheds.

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- 372 Figure 3. Plots of Spearman correlation coefficient versus temporal offset at the optimal
- 373 detrending smoothing range for each of the 15 conclusive sites, ordered from longest lead to
- 374 longest lag. The highest correlation value is colored in red, the identified lead or lag span is
- 375 represented with brackets, and the optimal smoothing range is listed in the bottom right corner
- of each plot. Note: The lead/lag relationship was inconclusive for Wilson, Laurinburg, Marion,
- 377 MSD of Buncombe County, and Roanoke Rapids, and these plots are therefore not presented.

378 Detrended wastewater loads were temporally leading detrended COVID-19 cases in 11 379 of the 15 sewersheds where we were able to identify optimal detrending smoothing ranges 380 (Figure 4). For these sites, the highest correlation was observed for wastewater loads sampled 381 at a median lead time of 6 days before the cases were reported, with a contiguous span of elevated correlations lasting a median of 3 days. At four sewersheds, the correlation between 382 383 detrended wastewater loads and detrended cases was greatest when the detrended wastewater 384 loads were lagging, with the highest correlation identified at a median of 3.5 days after the 385 cases were reported and a median contiguous span of elevated correlations of 2 days. Although 386 the smaller sewersheds were more likely to be inconclusive, size did not seem to influence the 387 lead/lag relationship at the 15 conclusive sites, with about the same proportion of leading vs 388 lagging between groups of the smallest and largest sewersheds. However, the optimal 389 detrending smoothing range seemed to be related to the lead/lag relationship, as 64% (7/11) of 390 the leading sewersheds had an optimal detrending smoothing range of T = 8 weeks and 75% 391 (3/4) of the lagging sewersheds had an optimal detrending smoothing range of T = 4 weeks, 392 suggesting that it may be easier to identify detrended wastewater loads lagging detrended 393 COVID-19 cases at shorter detrending time scales. A summary of the optimal smoothing range, 394 relationship, span, and temporal offset with the highest correlation identified for each sewershed 395 is included in the Supplemental Material (Table S2). 396 Figure 4. Locations of NC Wastewater Monitoring Network sewersheds participating between 397 January and October 2021. In 11 sewersheds, detrended wastewater leads cases (lead), in 4 398 sewersheds detrended wastewater lags cases (lag) and in 5 sewersheds results were 399 inconclusive.

400

## 401 Discussion

Wastewater surveillance emerged during the pandemic as a potential leading indicator
of COVID-19 infection trends in the community. Although previous research analyzed the overall

404 correlation between SARS-CoV-2 wastewater loads and clinical cases, this analysis used kernel 405 detrending to characterize short-term relationships and identify sub-trends. By detrending 406 wastewater viral loads and cases in the sewershed using various kernel smoothing ranges, we 407 were able to characterize lead/lag relationships at 15 of the 20 North Carolina sewersheds 408 assessed using a set of reproducible criteria, reducing the proportion of inconclusive results 409 from 90% without detrending to 25% using the optimal detrending smoothing range. 410 Furthermore, we found that detrended wastewater loads temporally led detrended cases at 411 almost three times as many sewersheds (N=11) as sewersheds where detrended wastewater 412 loads lagged detrended cases (N=4), further highlighting the utility of wastewater as a leading 413 indicator of COVID-19 cases in North Carolina. The optimal detrending kernel smoothing range 414 that removed long-scale pandemic trends while retaining sufficient temporal correlation to 415 identify lead/lag relationships was in the range of 4 to 8 weeks at 12 of the 15 sites with 416 conclusive relationships. Because detrending with a given smoothing range retains only the 417 variation in the observations at time scales shorter than the corresponding timeframe, this 418 finding suggests that this approach is ideal for identifying the leading or lagging nature of 419 wastewater and case trends in most sewersheds experiencing a sustained period of increasing 420 SARS-CoV-2 infection rates lasting at least 4 to 8 weeks. A sustained 4 to 8 weeks increase in 421 COVID-19 incidence corresponding to the emergence of the Delta variant (B.1.617.2) in late 422 July 2021 was observed in wastewater loads at 19 of the 20 study sites, further supporting the 423 wider relevance of this range during the study period. However, due to onboarding schedules, 424 some sewersheds were only sampled for half of the study period, and the shorter sampling 425 history appeared related to inconclusive results at these sites.

A strength of our study is that we performed a lead/lag analysis across a wide-range of WWTP systems, including both rural and urban municipal systems serving sewershed populations ranging from under 4,000 to 550,000 people [16,24,37–39]. Although we identified a leading relationship in the majority of North Carolina sewersheds, those within the same county

430 or in adjacent counties did not always exhibit the same lead/lag relationship nor have the same 431 optimal detrending smoothing range (Figure 4). For example, we found that detrended 432 wastewater loads led detrended cases at Charlotte 1 and Charlotte 3 but lagged detrended 433 cases at Charlotte 2 (Figure 4, Table S2). Wastewater led cases in both the Wilmington 434 sewershed and the sewershed encompassing surrounding areas of New Hanover County, but 435 the optimal detrending smoothing range was 8 weeks for the city and 16 weeks in the county. 436 which covers a larger land area but serves fewer people (Table 1). Differences in the temporal 437 relationship or optimal smoothing range at each sewershed could be due to conditions at a 438 given site: virus loads measured in wastewater can be impacted by sewer network infrastructure 439 age, sewer residence time, or weather [38,40,41], and clinical surveillance is subject to 440 underreporting due to testing access, home test usage, or fluctuations in populations from 441 tourists and commuters [42]. To minimize the potential impact of testing behavior on the 442 evaluation of relationships between SARS-CoV-2 loads and COVID-19 cases presented in this 443 work, we chose to perform the analysis for a period ending prior to November 2021, when 444 clinical testing penetration was still relatively high and home testing was not yet widely used in 445 North Carolina communities.

446 Given that site-specific conditions can influence wastewater results, public health 447 agencies leading wastewater surveillance programs in their jurisdictions may want to validate 448 their wastewater data against other foundational COVID-19 metrics to determine how 449 wastewater surveillance fits into their larger surveillance strategies. For states or jurisdictions 450 less familiar with wastewater data, a lead/lag analysis between wastewater loads and reported 451 cases would be a useful method to help understand the temporal relationship between 452 wastewater-based pathogen and other decision-making metrics. Our method can be employed 453 by public health agencies participating in CDC NWSS across the United States by using an R 454 Markdown document that applies set criteria to identify the leading or lagging relationships 455 between wastewater and reported cases [43]. As counts of reported cases become less reliable

over time due to an increase in non-reportable results from at-home-testing kits, as well as an
overall reduction in PCR-based, reportable, COVID-19 clinical testing, this method can be
adapted to utilize surveillance metrics besides cases, including hospitalizations, emergency
department visits (syndromic surveillance data), or mortality [17].

460 Results from our analysis characterizing the shortest time ranges at which wastewater 461 loads are associated with cases have been formative in elevating wastewater as a reliable 462 metric for tracking trends in North Carolina, not only to anticipate the start of long-term cycles 463 (such as the start of elevated rates in winter), but also for short duration fluctuations within any 464 given long-term cycle. The leading nature of wastewater-based COVID-19 findings at most sites 465 provides the foundation and rationale for including wastewater loads as an early warning metric 466 alongside reported cases, emergency department visits, and hospitalizations, which are 467 highlighted on statewide data surveillance dashboards such as the NCDHHS COVID-19

468 dashboard (<u>https://covid19.ncdhhs.gov/dashboard/wastewater-monitoring</u>).

469 In under two years, COVID-19 wastewater surveillance in the United States expanded 470 from 8 pilot state health agencies participating in the CDC National Wastewater Surveillance 471 System in 2020 to 46 states, 5 cities, 3 territories, and 7 tribes participating in 2022 [44]. 472 Similarly, the global portal expanded to coverage of 70 countries, reporting for 3,807 sites, 473 indicating widespread use of wastewater surveillance data [45]. With the explosive growth in 474 both the academic literature on, and implementation of, wastewater surveillance programs 475 globally, public health professionals developed a wide range of approaches to utilizing 476 wastewater data for decision making. Our method shows how detrended wastewater loads can 477 predict fine scale fluctuations in detrended cases, which can allow public health officials to 478 respond more *locally* and *timely* when COVID-19 burden, or other disease burden as 479 wastewater surveillance expands to new targets, is increasing at levels greater than the 480 baseline trend. Examples of mitigation strategies that can be deployed at local levels and for 481 short durations, while being complementary to long lasting statewide measures, may include the

482 following: (a) officials could quickly alert local hospitals about a potential increase in cases 483 above the statewide trend and provide recommendations to community leaders to implement 484 short-duration restrictions, such as limiting indoor gatherings and reducing business capacity 485 [46]; (b) jurisdictions could mobilize pop-up testing and take steps to increase vaccination in the 486 community [47]; (c) increasing public health communications regarding masking, handwashing, 487 vaccination, and social distancing to help contain the spread of the virus; and d) interacting with 488 local public health officials and hospital administrators to indicate periods of higher ICU bed, 489 PPE, and medical staffing needs. This has already been observed during a large sport fishing 490 tournament that took place in a small coastal North Carolina sewershed where NCDHHS 491 notified local health department and city officials of an increase in wastewater viral load. In 492 response to this increase, local health department and city officials reinforced recommended 493 mitigation strategies outlined in the Governor's Executive order to the event leadership, like 494 additional hand-washing stations and frequent disinfection of high touch surfaces (Nina Oliver. 495 Carteret County Health Director, personal communication, June 21-22 2021 & February 6, 496 2023). Local notices were also used to encourage the surrounding community to take 497 precautions through vaccinations, masking, social distancing, and frequent handwashing [48]. 498 Immediately following the event, county and city officials met routinely to review wastewater, as 499 well as other COVID-19 metrics, and to ensure levels were decreasing (Nina Oliver, personal 500 communication, February 6, 2023).

As public health officials and the scientific community continue to rely on wastewater surveillance both for large-scale pandemic decision-making and localized action as described here, there is a growing need for increasing equitable access to wastewater services, particularly in cases of municipal underbounding, and for investing in substantial infrastructure improvements. This is especially important in jurisdictions like North Carolina, where half of households rely on private septic and package treatment plants [49]. In some cases, racial disparities in access to and disproportionate exclusion from municipal water and sewer service

have been documented [49–51]. In other areas, distance, lack of gradient, and groundwater
height play a role in decisions to use centralized versus decentralized waste treatment systems.
For wastewater to continue to be useful for disease tracking and public health decision-making
beyond COVID-19, additional resources are needed to achieve equitable access to centralized
wastewater treatment where it is desired and environmentally relevant. In other rural areas
where this is not the case, we need to improve our technical capabilities to characterize
decentralized waste systems.

515

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Fig1



# Fig2



Fig3

