

1 **Synchrony dynamics of dissolved organic carbon in high-mountain streams:**  
2 **insights into scale-dependent processes**

3 Lluís Gómez-Gener<sup>1,2\*</sup>, Nicola Deluigi<sup>1</sup>, Tom Battin<sup>1</sup>  
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5 <sup>1</sup> River Ecosystems Research Laboratory, Alpine and Polar Environmental Research Center (ALPOLE),  
6 Ecole Polytechnique Fédérale de Lausanne, EPFL, Switzerland

7 <sup>2</sup> Centre for Research on Ecology and Forestry Applications, Universitat Autònoma de Barcelona, Campus  
8 de Bellaterra, Edifici C, 08193 Cerdanyola del Vallès, Barcelona, Spain

9 \* E-mail: gomez.gener87@gmail.com (Corresponding author)

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11 Running head: High-mountain stream DOC synchrony dynamics  
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43

44 **Abstract**

45 In high-mountain landscapes, biological resources such as organic carbon (OC) are often limited and  
46 heterogeneously stored in poorly developed soils, snow, ground ice and glaciers. Climate change influences  
47 the dynamics of OC mobilization to- and processing into- the recipient streams. These patterns can vary from  
48 seasonal (e.g., snow melt in spring) to daily (e.g., ice melt in summer) depending on the location of the streams  
49 within the catchment. Capturing the temporal richness of stream biogeochemical signals is now a reality with  
50 the advent of high-resolution sensors. In this study, we used wavelet analysis on high-frequency discharge (Q)  
51 and dissolved organic carbon (DOC) measurements from nine streams in the Swiss Alps to investigate whether  
52 synchrony (S) of Q ( $S_Q$ ) and DOC ( $S_{DOC}$ ) persisted or collapsed among streams, and its response to spatial  
53 position, climate and land cover gradients across different time scales. Our findings reveal that short-term (0-  
54 10 days)  $S_Q$  and  $S_{DOC}$  were strongly influenced by distance between streams and network connectivity. In  
55 contrast, catchment-related properties (i.e., altitude or land cover) were more important in driving  $S_Q$  and  $S_{DOC}$   
56 dynamics at longer time scales (>50 days). However, the degree to which local catchment properties controlled  
57 S patterns at the longest timescales depended both on the response variable (i.e., Q vs. DOC) as well on the  
58 main land cover (i.e., vegetation vs. glacier). Elucidating most prominent timescales of  $S_{DOC}$  is relevant given  
59 the hydrological alterations projected for high-mountain regions. We show that glaciers impose a unique  
60 seasonal regime on DOC concentration, potentially overriding the effects of other local hydrological or  
61 biogeochemical processes during downstream transport. Consequently,  $S_{DOC}$  dynamics in high-mountain  
62 streams may change as glaciers shrink, thereby altering downstream opportunities for biogeochemical  
63 transformations.

64

## 65 **Introduction**

66 Predicting how spatial and temporal variability changes with scale of observation is one of the fundamental  
67 interests of both landscape ecology (Turner et al. 1989; Levin 1992) and catchment hydro-biogeochemistry  
68 (McGuire et al. 2014). Part of this impetus is the recognition that many environmental processes are regional,  
69 continental or even global in nature (Turner et al. 1989). However, our ability to scale up process dynamics  
70 from single sites to broader areas is still limited.

71 Ecosystem dynamics result from both extrinsic and intrinsic controls. In freshwater ecosystems, extrinsic  
72 controls are predominantly linked to climatic factors, such as air temperature, precipitation, solar radiation or  
73 wind, and they exert a rather uniform influence over large regions (Magnuson et al. 1990). Accordingly, to the  
74 extent that spatially separated ecosystems (or their properties) within a region behave similarly over time (i.e.,  
75 spatial synchrony or temporal coherence, *sensu* Magnuson et al., 1990), it becomes easier to make  
76 generalizations about how they will respond to climate variations. Therefore, freshwater ecosystems within a  
77 certain region are expected to respond similarly to climatic events, such as a multi-year drought, a series of  
78 cool, wet years, or systematic year-to-year variations (Kratz et al. 1998). However, a series of pioneering multi-  
79 lake studies (Magnuson et al. 1990; Kratz et al. 1998; Soranno 1999; Baines et al. 2000) have also shown that  
80 synchrony between lakes depends on both the nature of the variable under study (e.g., physical, chemical, and  
81 biological properties) and the sensitivity of such variables to intrinsic controls such as local characteristics or  
82 internal lake dynamics (e.g., lake shape or basin characteristics). For example, some physical variables, such  
83 as surface temperature and ice-out date, exhibit strong synchrony over distances greater than 1000 km (Wynne  
84 et al. 1996; Kratz et al. 1998). On the other hand, some biological and chemical variables (e.g., algal biomass  
85 or nutrients) are less synchronous than most physical variables and semi-conservative chemical species  
86 (Magnuson et al. 1990; Kratz et al. 1998). Consequently, by examining patterns of temporal coherence, we  
87 can identify the spatial extent to which dynamics within aquatic ecosystems change or collapse in response to  
88 climatic change versus local processes (i.e., the degree of regional generalizations for effects of climate  
89 change). In other words, temporal coherence indicates the extent to which observed responses may be  
90 extrapolated to the broader landscape (Magnuson et al., 1990).

91 In contrast to lake ecosystems, rivers are relatively open ecosystems that form networks spanning multiple  
92 catchments and are inherently connected to landscapes through matter exchange (Hynes 1975; Vannote et al.  
93 1980; Frissell et al. 1986; Fisher et al. 1998). As a consequence, the transferability of lake-based models might  
94 not be straightforward for understanding the synchrony dynamics between stream pairs within fluvial networks  
95 (e.g., Seybold et al., 2021). Instead, accounting for other filters such as landscape position, land-stream  
96 interaction or network connectivity might be necessary to better understand the synchronous versus  
97 asynchronous responses of streams to climate variability. To overcome this issue, we propose adapting the  
98 landscape position concept from lake districts (e.g., *sensu* Webster et al. 1996; Baines et al. 2000) to fluvial  
99 networks and exploring its applicability for the analysis of synchrony dynamics between contrasted variables  
100 in both connected or disconnected streams within the same region. Climate signals are filtered (i.e., attenuated  
101 and/or scattered) by both landscapes and riverscapes to reduce the synchrony between stream pairs or create

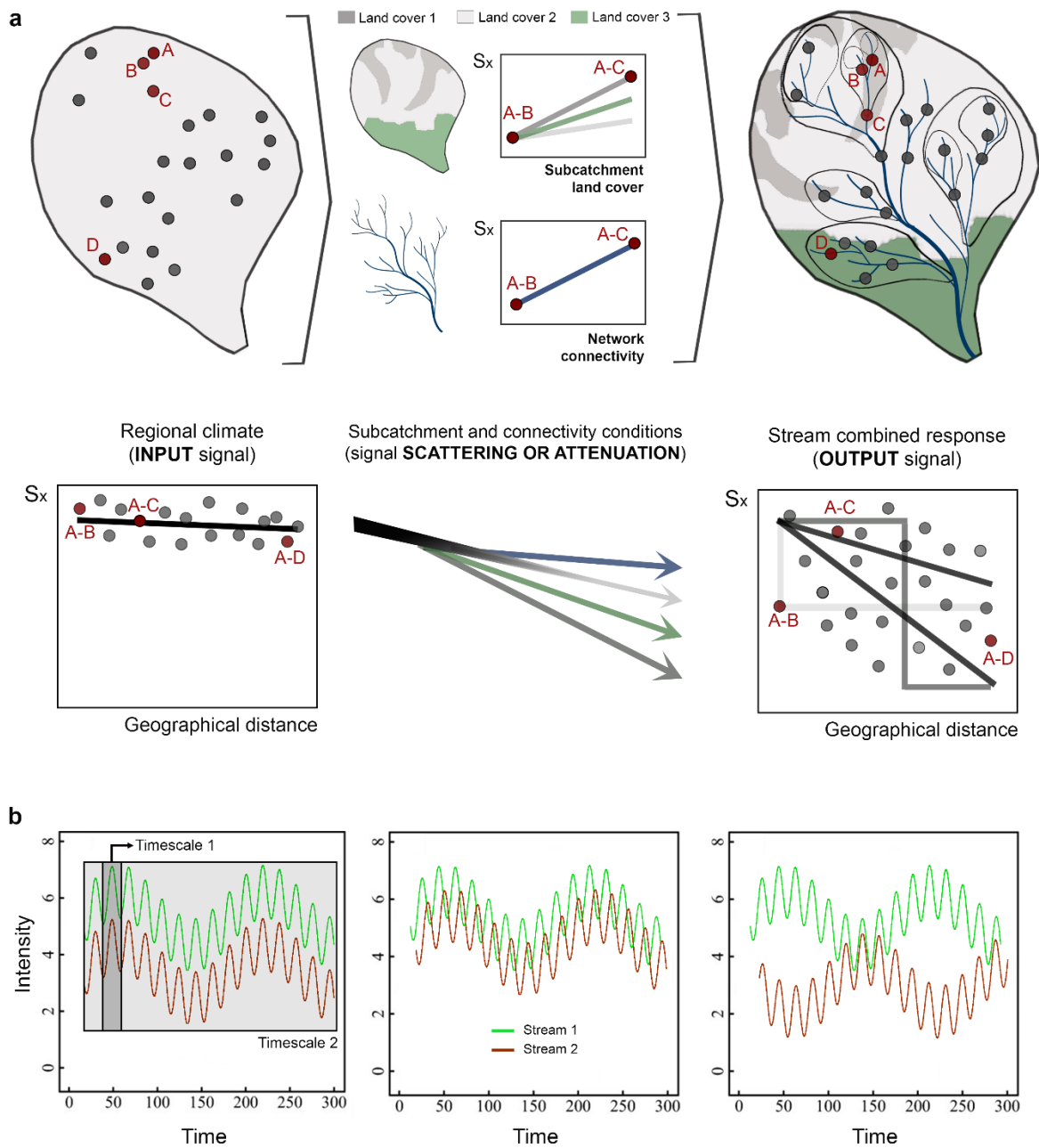
102 nonlinear responses of synchrony across distance gradients (Figure 1a). The dynamics of any dissolved  
103 constituent in a stream are largely determined by both land cover and its connectivity with upstream waters  
104 and lateral catchment sources (Hynes 1975). Based on this, a pair of stream segments draining catchments with  
105 different landscape compositions and/or not connected through the river network (A-B pair in Figure 1a) will  
106 be less synchronized for a specific dissolved constituent in stream water, regardless their lineal geographical  
107 distance, compared to a pair of streams with similar land cover proportions or sharing the same network branch  
108 (A-C in Figure 1a). As an example, forested and non-forested catchments not only differ in how dissolved  
109 constituents are spatially created and stored but also, because this heterogeneity includes differences in their  
110 vertical profiles, in how they are mobilized into streams in response to changes in flow (Tank et al. 2018;  
111 Gómez-Gener et al. 2021).

112 In addition to the spatial dimension, timescale-dependent controls on aquatic ecosystem processes have been  
113 identified in numerous long-term time series (Likens 1989; Hampton et al. 2019; Wilkinson et al. 2020). This  
114 includes the effect of processes operating within fractions of seconds (biochemical reactions), from seconds to  
115 days (non-biological processes such as rain events/hydrological pulses), and from days to weeks or even  
116 months (snowmelt events, solar cycles). Because the mechanism driving synchrony of a certain water  
117 constituent may operate at specific timescales (Defriez et al. 2016, 2017; Sheppard et al. 2016; Walter et al.  
118 2017; Anderson et al. 2019), synchrony will manifest itself differently across timescales (Figure 1b).  
119 Nevertheless, the complexity of high-frequency signals challenges our ability to identify synchronies (or  
120 asynchronies) between stream locations and/or variables of interest because (1) traditional monitoring,  
121 normally based on weekly or monthly grab measurements (Kirchner et al. 2004; see examples above), fails in  
122 capturing some of these responses (e.g., short-term events driven by rapid changes in water flow paths), and  
123 (2) standard approaches for assessing synchrony (i.e., correlation analysis) assume linearity and uniformity  
124 across timescales. With the advent of affordable and reliable sensor technology, high-resolution measurements  
125 across a broad range of solutes over prolonged periods are possible (Rode et al. 2016; Marcé et al. 2016). This  
126 enables an increase in the temporal richness of signals and, more importantly, the capture of processes  
127 dynamics occurring at high frequencies. Additionally, novel approaches have enabled the isolation of complex  
128 patterns into those timescales of variability with the most interest and/or significance (i.e., wavelet coherence;  
129 Torrence and Compo 1998; Grinsted et al. 2004; Cazelles et al. 2008).

130 Here we take advantage of these advances to improve the understanding of the patterns and processes  
131 regulating the transport and cycling of solutes in heterogeneous headwater landscapes spanning a high diversity  
132 of temporal signals. To do so, we applied synchrony (S) analysis to three years of high frequency measurements  
133 (every 10 minutes) of discharge and dissolved organic carbon (DOC) from nine high-mountain stream  
134 segments distributed over three catchments in the Western Swiss Alps. With this approach, S patterns can  
135 reveal whether similarities between stream pair dynamics are more persistent across temporal scales or, in  
136 contrast, frequency-dependent non-linearities arise in cases of lower pair similarity. Identifying groups of  
137 streams that respond similarly to hydro-climatic fluctuations will enhance our ability to monitor the effects of  
138 climate variability and land cover change in high-mountain regions, choose appropriate reference ecosystems,

139 and detect local deviations from the regional norm. This has become a remarkable task given the profound  
 140 climate-driven hydrological and physicochemical changes that these streams are facing.

141



142

143 **Figure 1.** Conceptual figure illustrating both the spatial (**a**) and temporal (**b**) structure of solute synchrony between  
 144 streams draining heterogeneous landscapes. (**a**) Climatic signals are filtered (e.g., scattered or attenuated) by the landscape  
 145 and/or stream networks, resulting in reduced synchrony between stream pairs or the emergence of nonlinear synchrony  
 146 responses across distance gradients. Red circles represent specific examples of streams exhibiting differences in pairwise  
 147 characteristics: 1) geographical distance or linear proximity (i.e., from A-B to A-D), 2) land cover (LC) of the sub-  
 148 catchment (e.g., A-C: LC1 vs. LC1; A-B: LC1 vs. LC2; A-D: LC1 vs. LC3), and 3) connectivity within the drainage  
 149 network (e.g., A-B: streams not connected; A-C: streams connected longitudinally through the main channel). (**b**)  
 150 Synchrony can exhibit different manifestations across timescales due to mechanisms operating at specific temporal scales.  
 151 Green and red solid lines represent simulated solute signal fluctuations for a specific parameter at two different stream

152 locations. In the first and third examples, the two streams exhibit synchrony for the specific parameter at the shorter period  
153 of the signal (Timescale 1). In contrast, in the second example, the two streams are synchronized at longer periods  
154 (Timescale 2) while showing asynchrony at the shorter timescale (Timescale 1). Figure 1(b) has been adapted from  
155 Defriez and Reuman (2017).

156

## 157 **Methods**

### 158 *Study sites*

159 We selected nine stream segments distributed across three catchments (Vallon de Nant, Valsorey and Val  
160 Ferret) in the Western region of the Swiss Alps (Figure S1). The selected streams and their corresponding  
161 catchments are located within a 40 km radius of each other (Figure S1a). While the proximity of these  
162 catchments leads to comparable meteorological patterns (e.g., temperature regimes and precipitation levels),  
163 they encompass a range of topo-climatic and environmental gradients characteristic of high-altitude systems  
164 (Figure S1b and Table S1). Among these streams, two (PEU and VEL) are not influenced by glaciers, while  
165 the remaining seven receive meltwaters from glaciers, which cover 3.4% to 33.5% of their catchments (Table  
166 S1). The drainage areas of these streams vary in size from 3.1 to 23.2 km<sup>2</sup>. The catchment altitudes range from  
167 1200 to 2161 m above sea level (a.s.l.) and stream channel slopes vary from 0.05 to 0.16 mm<sup>-1</sup>. Six of the  
168 stream segments (VAU, VAD, VEL, FEU, FED and PEU) are situated above the tree line, draining steep  
169 slopes with sparsely vegetated rocky terrain with grassland and shrubs. In contrast, the remaining three (RIC,  
170 AND, ANU) drain partially forested catchments mainly composed of coniferous tree species. The total  
171 vegetation cover across the catchments ranges from 21.1% to 70.2% (Boix Canadell et al., 2019). The  
172 geological composition of the catchments exhibits certain variability. The Vallon de Nant catchment is  
173 dominated by limestones, calcareous shales, and flysch. In the Val Ferret catchment, the geology is primarily  
174 composed of limestones and sandstones with pronounced schistosity. In contrast, the Valsorey catchment  
175 stands out with its metamorphic lithology, including gneisses, crystalline shales, and blue-grey schists. These  
176 geological differences also contribute to the diverse characteristics and dynamics observed in the streams.

177 To assess the similarity between stream pairs, we calculated a similarity index based on geographical and  
178 landscape characteristics (Table S2). We considered six main characteristics: linear distance (or horizontal  
179 proximity), sub-catchment land cover, topographic attributes, and connectivity within the drainage network.  
180 The stream connectivity index was assigned a value of 2 when two streams were connected longitudinally  
181 through the main channel, 1 when one stream was a tributary of the main channel, and 0 when the stream  
182 segments were not connected or belonged to different catchments (Figure S1b). This analysis aimed to  
183 hypothesize which stream pairs would exhibit greater temporal coherence based on their geographic and  
184 landscape attributes.

185

186 *Automated sensor measurements*

187 At each stream segment, we deployed high-resolution sensors to measure parameters over a three-year period  
 188 (from October 28, 2016, to October 20, 2019) as part of the ongoing METALP stream monitoring network  
 189 (<https://metalp.epfl.ch/>). Sensors measured every ten minutes water depth (WT-HR 1000, TruTrack),  
 190 chromophoric DOM (CDOM) sensor (Cyclops-7 CDOM/FDOM, PME Inc.), and turbidity (Cyclops-7  
 191 Turbidity, PME Inc.); all of them also measuring streamwater temperature. Before deployment, the optical  
 192 sensors were calibrated in the laboratory according to the manufacturers' specifications. Turbidity sensors were  
 193 calibrated using sensor-specific Turbidity NTU Calibration Standard-Formazin (Sigma-Aldrich®) up to 2000  
 194 NTU, and measurements were reported in nephelometric turbidity units (NTU). The CDOM sensor measured  
 195 fluorescence and was calibrated in parts per billion (ppb) of quinine sulfate equivalents (fluorescence of 1 ppb  
 196 quinine sulfate monohydrate (Sigma-Aldrich®) in 0.05 M H<sub>2</sub>SO<sub>4</sub>) using a two-point calibration curve (0 to  
 197 1,000 ppb quinine sulfate equivalents). Additionally, monthly spot measurements of the same parameters were  
 198 conducted during site visits to validate the recorded data, assess data quality and eventually post-correct time  
 199 series.

200 Discharge (m<sup>3</sup> s<sup>-1</sup>) was estimated using slug-injections of sodium chloride as a conservative tracer (Gordon et  
 201 al. 2004). Rating curves relating streamwater depth to discharge were established for each site using a power-  
 202 law model. The data was log-transformed, and a linear regression was performed to fit the rating curve  
 203 equation. This approach allowed for the estimation of discharge based on water depth measurements obtained  
 204 from the deployed sensors.

205

206 *Relationships between streamwater FDOM and DOC concentrations*

207 We compensated raw sensor FDOM data for streamwater temperature (FDOMtemp) and turbidity  
 208 (FDOMcorr) as follows (Downing et al. 2012; Boix Canadell et al. 2019):

209 
$$FDOM_{temp} = \frac{FDOM_{raw}}{1 - 0.016 (Temp_m - Temp_{ref})} \quad (1)$$

210 
$$FDOM_{corr} = \frac{FDOM_{temp}}{0.95 e^{(-0.002 Turb_m)} + 0.03} \quad (2)$$

211 where  $FDOM_{raw}$  refers to the uncorrected FDOM values,  $Temp_m$  is the streamwater temperature,  $Temp_{ref}$  is the  
 212 reference temperature at 10 °C, and  $Turb_m$  represents the streamwater turbidity (Watras et al. 2011; Lee et al.  
 213 2015). The coefficients were derived from laboratory experiments that assessed the sensor specific effect of  
 214 temperature and turbidity on the FDOM signal. These coefficients were comparable to those reported in  
 215 previously studies (Watras et al. 2011; Downing et al. 2012; Lee et al. 2015).

216 Additionally, we conducted regular site visits, with a minimum frequency of once a month for sensor  
 217 maintenance, data downloading, and grab sampling for DOC analysis. Streamwater samples for DOC analysis  
 218 were filtered using precombusted GF/F filters (Whatman) and collected in acid-washed, precombusted glass  
 219 vials; DOC concentrations were analyzed using a Sievers M5310c TOC Analyzer (GE Analytical Instruments).

220 The time series of streamwater DOC concentrations were calculated through least-squares linear regressions  
221 between  $FDOM_{\text{corr}}$  sensor data and DOC concentrations measured from grab samples (for more details on the  
222 methods, see Boix Canadell et al. 2019).

223

### 224 *Wavelet coherence*

225 We applied wavelet coherence (Torrence and Compo 1998; Grinsted et al. 2004; Schmidt et al. 2019) to  
226 identify synchronous fluctuations in Q and DOC concentration of stream pairs. To achieve this, we  
227 decomposed both time series using continuous wavelet transforms, which convert time series from the time  
228 domain to the time-frequency domain. The wavelet transform involves convolving a time series  $x(t)$  with a  
229 wavelet  $\psi(t)$ , which is a basis function localized in both time and frequency. We selected the Morlet wavelet  
230 as the basis function, due to its balance between time and frequency resolution and its widespread usage  
231 (Torrence and Compo 1998; Cazelles et al. 2008; Cauvy-Fraunié et al. 2013; Schmidt et al. 2019; Kneib et al.  
232 2020). The Morlet wavelet consists of a plane wave modulated by a Gaussian distribution and it defined as

$$233 \quad \psi_0(t) = \pi^{-\frac{1}{4}} e^{-i\omega_0 t} e^{-\frac{t^2}{2}} \quad (3)$$

234 where  $t$  represents time, and  $\omega_0$  is the central angular frequency. The continuous wavelet transform at scale  $f$   
235 and time  $\tau$  is given by:

$$236 \quad |W(f, \tau)| = \frac{1}{\sqrt{f}} \int_{-\infty}^{+\infty} x(t) \psi^* \left( \frac{t-\tau}{f} \right) dt = \int_{-\infty}^{+\infty} x(t) \psi_{f,\tau}^* \left( \frac{t-\tau}{f} \right) dt \quad (4)$$

237 where (\*) denotes the complex conjugate form (Torrence and Compo 1998, Cazelles et al. 2008). In Layman's  
238 terms, the wavelet function can be considered a portion of a sine function, which is then compared to different  
239 segments of the time series of observables. In the next step, the frequency of the wavelet function is modified  
240 by compressing or extending it along the time axis. As a result, this wavelet provides specific information  
241 about frequency and time localization. Wavelet coherence is determined by squaring of the product of the  
242 wavelet transform of the first time series,  $W_{f,\tau}^X$ , with the complex conjugation of the second time series,  $W_{f,\tau}^Y$ ,  
243 and normalizing it by the individual power spectra of each time series (see Schmidt et al. 2019).

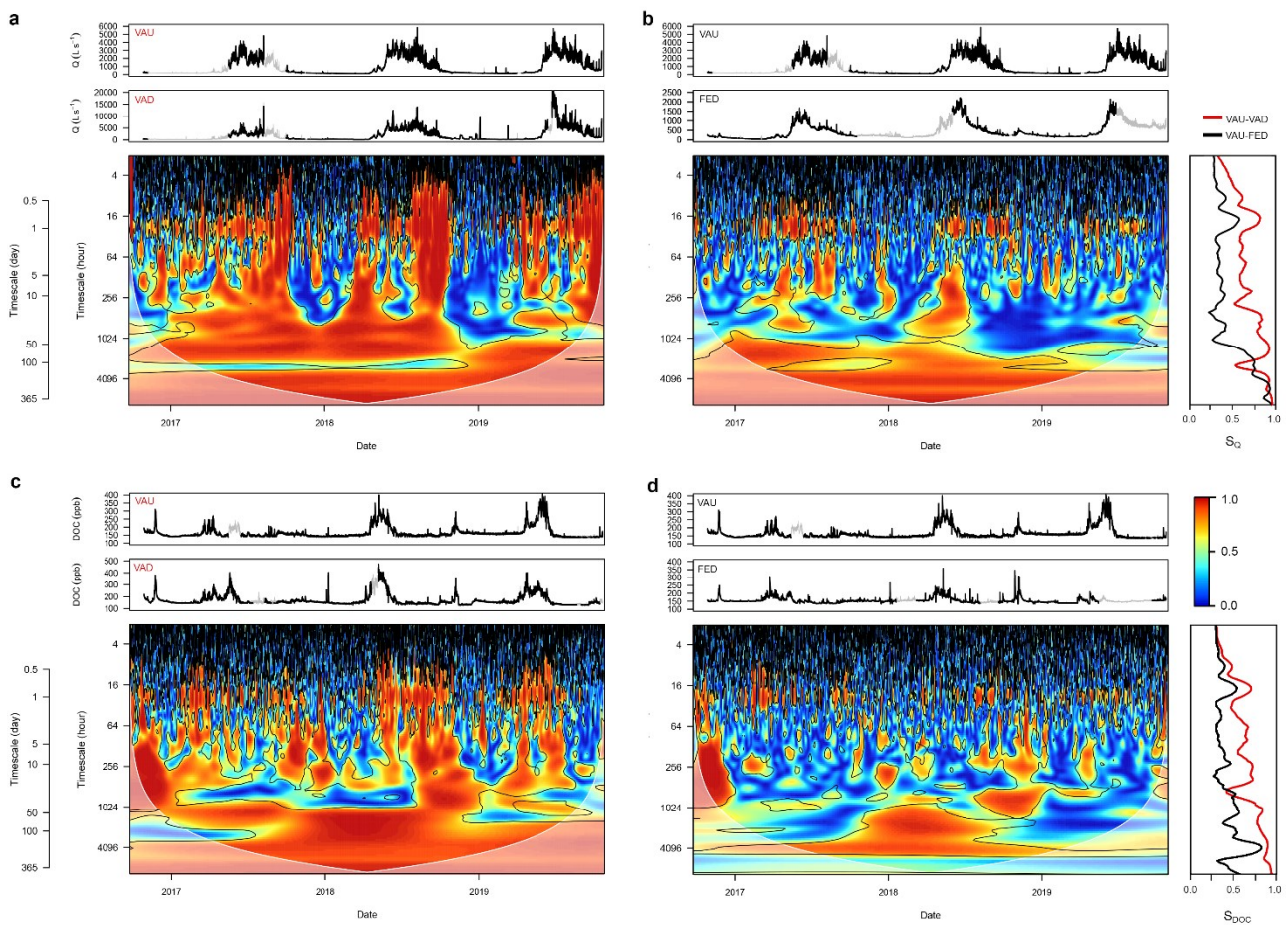
244 Wavelet coherence ranges between 0 and 1, and identifies phases of local cross-correlation between two time  
245 series as a function of frequency. High values of wavelet coherence indicate strong similarity between time  
246 series at a specific frequency and during a particular time window. All analyses were performed using R (R  
247 Core Team 2021). Wavelet coherence relationships were calculated using the R biwavelet package (Gouhier  
248 et al. 2018), which is based on MATLAB's WTC package (Grinsted et al. 2004). Starting from version 0.14,  
249 biwavelet also includes the plotting of bias-corrected wavelet and cross-wavelet power spectra, following the  
250 methods described by Liu et al. (2007) and Veleda et al. (2012). This correction is necessary because the  
251 traditional approach for computing the power spectrum (e.g., Torrence and Compo 1998) leads to an artificial  
252 and systematic reduction in power at lower periods. The significance of wavelet coherence was determined by  
253 comparing it to red noise, generated by 100 Monte Carlo randomizations of a first-order autoregressive process



254 (AR1) with the same autocorrelation coefficients as the respective input time series. Significance was tested  
 255 at a significance level of 0.95.

256 We generated a total of thirty-six wavelet coherence and cross-wavelet power spectrums for both discharge  
 257 and dissolved organic carbon using the methods described above. Due to the large number of graphs generated,  
 258 we have included examples of the wavelet coherence spectrum for two stream pairs with contrasting synchrony  
 259 patterns for discharge (Figure 2a and 2b) and for dissolved organic carbon concentrations (Figure 2c and 2d).  
 260 The left panels of Figure 2 depict streams (VAU-VAD) with the highest pair similarity ranks for all geographic  
 261 distance, sub-catchment land cover, and network connectivity (representing case A-C in Figure 1). The right  
 262 panels of Figure 2 depict streams (VAU-FED) with relatively low pair similarity ranks for all geographic  
 263 distance, sub-catchment land cover, and network connectivity (representing case A-D in Figure 1).

264



265

266 **Figure 2.** Time series and wavelet coherence spectrum of two contrasting stream pairs for discharge (a, b) and dissolved  
 267 organic carbon concentration (c, d). The left panels compare streams VAU and VAD, which exhibit the highest pair  
 268 similarity ranks across all geographic distances, sub-catchment land cover, and network connectivity (corresponding to  
 269 case A-C in Figure 1). The right panels compare streams VAU and FED, which have relatively low pair similarity ranks  
 270 for geographic distance, sub-catchment land cover, and network connectivity (corresponding to case A-D in Figure 1). In  
 271 the time series plots, grey regions indicate data gaps that were imputed. The black contours represent regions of significant  
 272 coherence against red noise, determined using Monte Carlo AR1 time series with a significance level of 0.95. The lighter  
 273 shade denotes the cone of influence, where edge effects may distort coherence patterns. The y-axis on the coherence

274 spectrum indicates timescale length, expressed in both hours and days. The panels on the right margin display the global  
275 wavelet coherence or synchrony for discharge ( $S_Q$ , upper panel) and dissolved organic carbon ( $S_{DOC}$ , lower panel),  
276 calculated as the arithmetic mean over time.

277

## 278 **Results and discussion**

### 279 *Timescale-aggregated synchrony*

280 Timescale-aggregated synchrony, measured as the median synchrony across all timescales, ranged from 0.34  
281 to 0.68 (median: 0.43,  $n=36$ ) for discharge ( $S_Q$ ) and from 0.25 to 0.62 (median: 0.36,  $n=36$ ) for DOC ( $S_{DOC}$ )  
282 among stream pairs (Figure S2). Out of thirty-six stream pairs, thirty had higher aggregated  $S_Q$  than  $S_{DOC}$   
283 (Figure S2). However, timescale aggregated  $S_Q$  than  $S_{DOC}$  was not significantly related ( $r^2 = 0.10$ ;  $p > 0.05$ ,  $n$   
284  $= 36$ ; Figure S2). This low correlation is consistent with other studies showing that discharge fluctuations  
285 between streams are often more synchronous over spatial scales compared to soil-derived dissolved  
286 constituents (Basu et al. 2010; Thompson et al. 2011). Similar to water levels in lakes (Magnuson et al. 1990;  
287 Kratz et al. 1998), stream discharge dynamics are likely to be more directly influenced (i.e., less filtered) by  
288 climatic factors, whereas DOC concentrations are either indirectly influenced by climatic factors or  
289 interactively influenced (or filtered) by other localized factors such as soil production, atmospheric deposition,  
290 terrestrial-aquatic transfer, and internal processing.

291

### 292 *Timescale-decomposed synchrony*

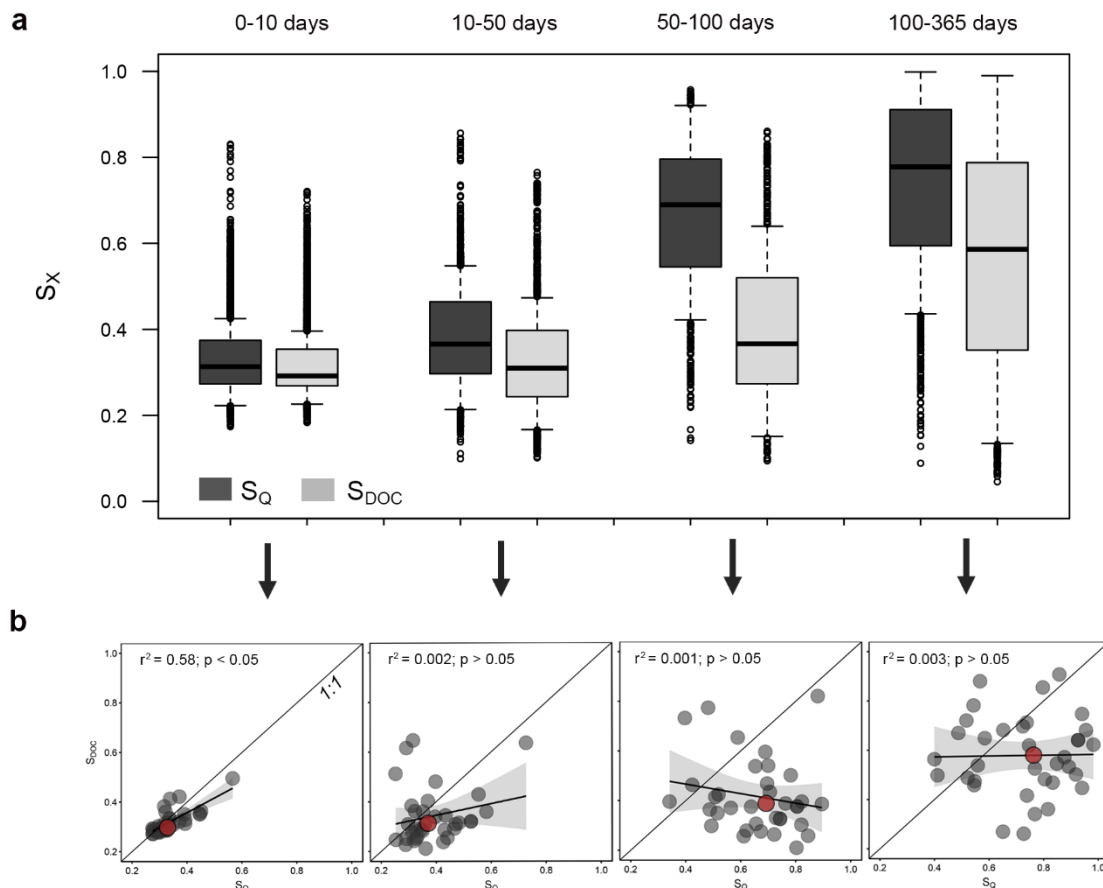
293 Decomposition of timescale-aggregated synchrony revealed that both  $S_Q$  and  $S_{DOC}$  were strongly influenced  
294 by the timescale of fluctuations (Figure S3). As timescales increased from shorter to longer periods, both  
295 individual (Figure S3a and S3b) and averaged (Figure S3c) stream pair synchrony for Q and DOC showed a  
296 continuous increase (averaging from 0.24 to 0.94 for  $S_Q$  and from 0.27 to 0.82 for  $S_{DOC}$ ; Figure S3c). The only  
297 exception to this pattern was a pronounced increase in both  $S_Q$  and  $S_{DOC}$  at the daily timescale, reflecting the  
298 synchronized response of these variables to daily climatic fluctuations (e.g., air temperature, solar radiation,  
299 soil evapotranspiration). At certain specific timescales,  $S_Q$  and  $S_{DOC}$  exhibited different responses (Figure 3a),  
300 with  $S_Q$  increasing more rapidly than  $S_{DOC}$  for timescales of approximately 50 days and beyond (Figure 3a).  
301 This difference becomes more pronounced when comparing short-term (0-50 days) and long-term (50-365  
302 days) timescale bands, as  $S_Q$  increased from 0.32 to 0.68 while  $S_{DOC}$  remained relatively stable, at 0.31 to 0.33.  
303 This pattern suggests that regional landscape drivers have a relatively greater influence (or less filtration) on  
304 discharge fluctuations at longer timescales (e.g., beyond 50 days) compared to DOC fluctuations. In contrast,  
305 shorter-term fluctuations in both discharge and DOC (0-50 days) shared similar drivers.

306 In accordance with the previous pattern, the strength of the relationship between  $S_Q$  and  $S_{DOC}$  was also  
307 dependent on the timescale of fluctuations (Figure 3b). At shortest timescales (0-10 days),  $S_Q$  and  $S_{DOC}$  were  
308 significantly related ( $r^2 = 0.58$ ;  $p < 0.05$ ,  $n = 36$ ). However, this covariation between variables did not persist  
309 at timescales ranging from 10 to 50 days ( $r^2 = 0.002$ ;  $p > 0.05$ ,  $n = 36$ ), 50 to 100 days ( $r^2 = 0.001$ ;  $p > 0.05$ ,  $n$

310 = 36), or 100 to 365 days ( $r^2 = 0.003$ ;  $p > 0.05$ ,  $n = 36$ ). Consistent with the aggregated data, the changing  
 311 dynamics of  $S_Q$  and  $S_{DOC}$  from shorter to longer timescales suggest that mechanisms operating at shorter  
 312 timescales (directly linked to climatic fluctuations) affect both variables similarly. However, the mismatch  
 313 between  $S_Q$  and  $S_{DOC}$  observed at longer timescales suggests that DOC dynamics, which are likely more  
 314 influenced by local catchment drivers, tend to be further removed from direct climate forcing or regional  
 315 climatic drivers. Although climate also influences the processes driving DOC patterns due to its intrinsic  
 316 hydrological connection, many other local-scale processes, including soil DOC production, terrestrial-aquatic  
 317 DOC transfer, and internal DOC processing, are likely filtering such signals (Li et al. 2018; Boix Canadell et  
 318 al. 2019; Rosset et al. 2020).

319 The variability of  $S_Q$  and  $S_{DOC}$ , measured as the standard deviation (SD) of  $S_Q$  and  $S_{DOC}$  for the 36 stream pairs  
 320 at each timescale, also increased significantly with the timescale of fluctuations. For  $S_Q$ , the SD ranged from  
 321 0.030 to 0.242, while for  $S_{DOC}$ , it ranged from 0.007 to 0.280. Notably, the threshold at which SD increased  
 322 more markedly was approximately 10 days (Figure S3d). This consistent increase in dispersion suggests that  
 323 controls on the variables are less constrained and more diversified (Abbott et al. 2018).

324



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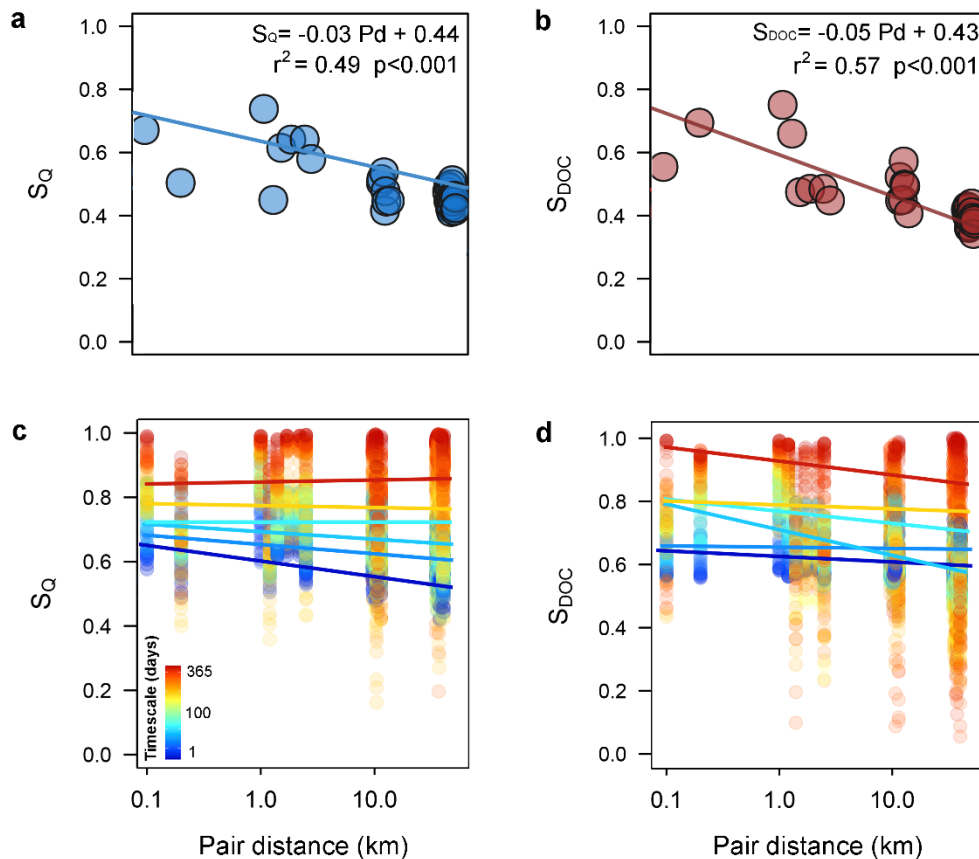
326 **Figure 3. (a) Comparison and (b) relationships between aggregated  $S_Q$  and  $S_{DOC}$  at four specific timescale ranges (see**  
 327 **headings). In panel (a) box plots display the 25<sup>th</sup>, 50<sup>th</sup>, and 75<sup>th</sup> percentiles; whiskers display 10<sup>th</sup> and 90<sup>th</sup> percentiles.**  
 328 **In panel (b) red circles represent medians of the time-scale aggregated data, while lines represent least-squares linear**

329 regression models between aggregated  $S_Q$  and  $S_{DOC}$  at four specific timescale ranges. The corresponding text provides  
 330 statistical outputs ( $r^2$  and p-values) for each regression model.

331 *Spatial variability of synchrony*

332 In our study region, streams that were closer to each other exhibited higher  $S_Q$  and  $S_{DOC}$  values compared to  
 333 streams that are more spatially separated (Figure 4a and 4b). Therefore, consistent with Kling et al., 2000,  
 334 timescale-aggregated  $S$  shows relatively low geographical persistence across the distance gradient (ranging  
 335 from 0.1 to 40 km) for both  $Q$  and  $DOC$  (Figure 4a and 4b, respectively). As depicted in the conceptual model  
 336 (Figure 1a, left panels), large-scale synchrony persistence can be attributed to the presence of strongly  
 337 “regionalized” environmental variables, such as shared exposure to similar climatic shifts or homogeneous  
 338 landscape properties. However, in our case, the observed trend of decreasing synchrony with spatial position  
 339 suggests that localized controls are overriding or strongly influencing the effects of higher-scale drivers  
 340 operating at the entire region and acting as synchrony stabilizers. These findings align with previous studies  
 341 conducted in lake districts, where a negative correlation between lake pairs  $S$  and lake similarity in exposure  
 342 to climatic factors (measured as pair distance and area/depth ratio) was moderate to high in most cases  
 343 (Magnuson et al. 1990; Soranno 1999; Kling et al. 2000).

344



345

346 **Figure 4.** Relationships between timescale aggregated (a)  $S_Q$  and (b)  $S_{DOC}$  (measured as the median coherence between  
 347 stream pairs for all timescales) and timescale segregated (c)  $S_Q$  and (d)  $S_{DOC}$  (measured as the median coherence of six  
 348 specific timescale bands) and stream pair distance. Lines in panels (a), (b), (c) and (d) represent to least-squares linear

349 regression models. Refer to Table S4 for the statistical outputs of the least-squares linear regression models associated  
350 with panels (c) and (d).

351

352 In our study, the loss of synchrony with linear distance could be attributed to the local landscape heterogeneity  
353 characteristic of high-mountain regions. This heterogeneity includes differences in the size and arrangement  
354 of landscape patches that act as sources and sinks of water and solutes, such as DOC (e.g., soils, groundwater,  
355 snow, ground ice, and glaciers), which ultimately control the routing of water and the delivery of solutes to  
356 streams as water flows superficially and sub-superficially (Tank et al. 2018) through the landscape (Malard et  
357 al. 1999; Brown et al. 2003). Additionally,  $S_{\text{DOC}}$  may be uniquely influenced by the local effects of consistent  
358 processing, such as ecosystem respiration, as water flows within river ecosystems. If these interpretations are  
359 correct, we would expect the reduction of S with distance to be more pronounced or correlated for DOC.  
360 However, despite the slope of the relationship between timescale-aggregated  $S_{\text{DOC}}$  and pair distance being  
361 slightly higher than that for  $S_{\text{Q}}$  (Table S3), a regression analysis indicated that only  $S_{\text{DOC}}$  (not  $S_{\text{Q}}$ ) was  
362 additionally influenced by local topographical and land cover properties, such as altitude and catchment  
363 vegetation coverage (Table S3).

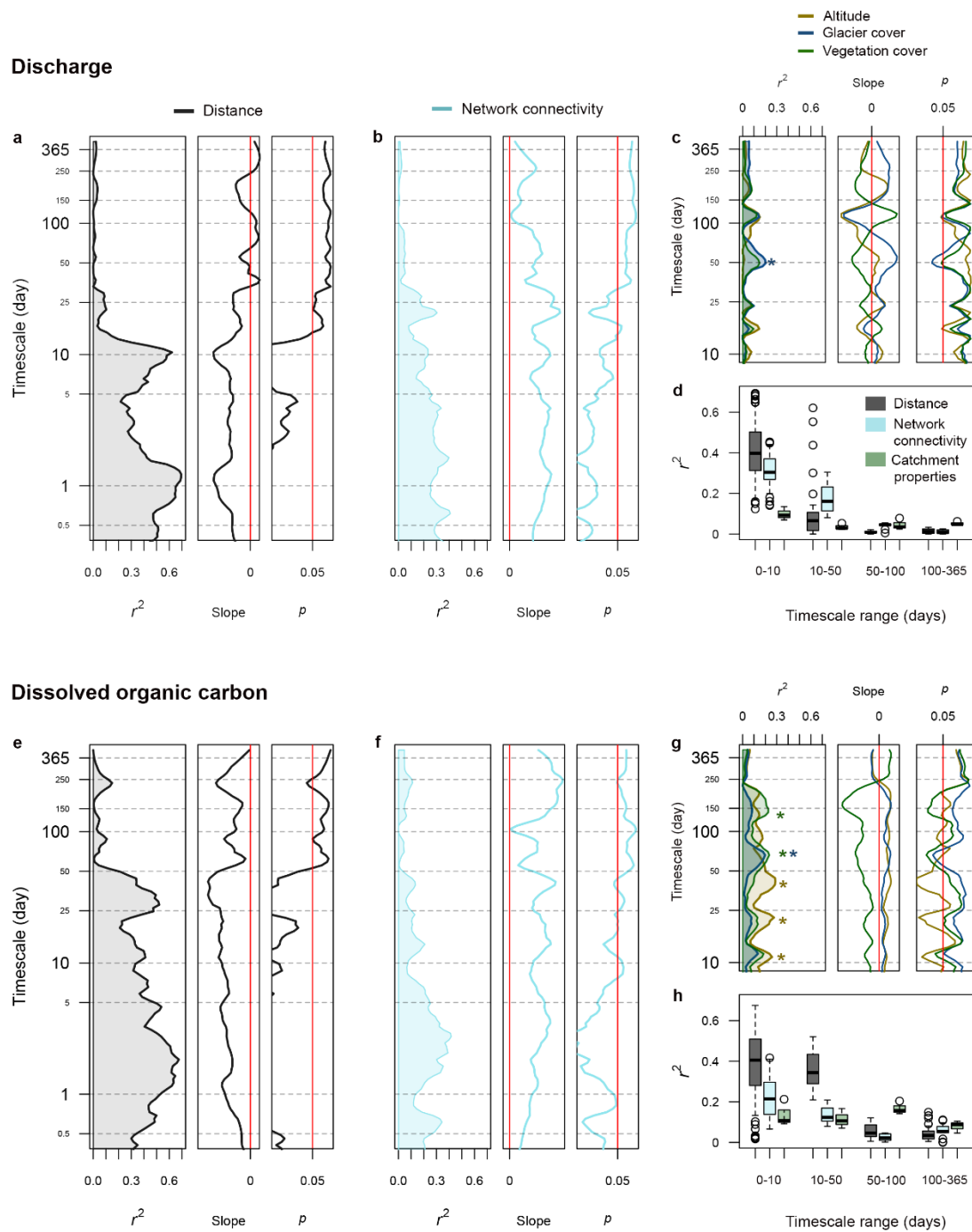
364 In accordance with the timescale-aggregated data, negative correlations between  $S_{\text{DOC}}$  and stream pair distance  
365 persisted even after segregating the data by timescale (Figure 4c and 4d; Table S4). This pattern was more  
366 consistent at longer timescales of fluctuations and contrasted with the observations for  $S_{\text{Q}}$ , which exhibit higher  
367 spatial persistence at longer and intermediate timescales but showed higher correlation at shorter timescales  
368 (as described in the drivers below). Taken together, these results suggest that  $S_{\text{DOC}}$  dynamics are more sensitive  
369 to locally unique conditions compared to  $S_{\text{Q}}$ , and that these local conditions exert distinct effects on the  
370 variables across the range of explored timescales.

371

### 372 *Timescale-dependent drivers of synchrony*

373 Given that the magnitude and dynamics of  $S_{\text{Q}}$  and  $S_{\text{DOC}}$  depended on the timescale of observation, we  
374 hypothesized that the mechanisms driving synchrony would also vary across timescales. To investigate this,  
375 we conducted linear regression analysis to examine whether potential drivers of  $S_{\text{Q}}$  and  $S_{\text{DOC}}$  (Figure 1a and  
376 Table S1) operated differently at each specific timescale considered ( $n=148$ , ranging from 1 hour to 365 days).  
377 The statistical outputs from these models provided evidence that the mechanisms driving  $S_{\text{Q}}$  and  $S_{\text{DOC}}$  dynamics  
378 indeed differed across timescales (Figure 5). Considered drivers were stream pair distance, network  
379 connectivity, and altitude, glacier extent and vegetation cover (hereafter catchment properties). At short  
380 timescales (ranging from 0 to 10 days),  $S_{\text{Q}}$  and  $S_{\text{DOC}}$  were strongly correlated with stream pair distance (median  
381  $r^2_{\text{Q}}=0.40$ ,  $r^2_{\text{DOC}}=0.41$ ) and network connectivity (median  $r^2_{\text{Q}}=0.30$ ,  $r^2_{\text{DOC}}=0.21$ ), while weaker relationships  
382 were observed with catchment-related properties within the same timescale range (median  $r^2_{\text{Q}}=0.08$ ,  
383  $r^2_{\text{DOC}}=0.10$ ). Similarly, at timescales ranging from 10 to 50 days, stream pair distance (median  $r^2_{\text{Q}}=0.11$ ,  
384  $r^2_{\text{DOC}}=0.37$ ) and network connectivity (median  $r^2_{\text{Q}}=0.20$ ,  $r^2_{\text{DOC}}=0.14$ ) remained the most important factors

385 driving  $S_Q$  and  $S_{DOC}$  (Figure 5). These results indicate that linear proximity and connectivity between stream  
 386 pairs, either through the landscape or the stream network, controlled short-term S dynamics regardless the  
 387 differences in local landscape characteristics observed between the catchments they drain (Table S1 and S2).



388  
 389 **Figure 5.** Temporal (or timescale) structure of the spatial drivers of (a) discharge synchrony (SQ) and (b) dissolved  
 390 organic carbon synchrony (SDOC). Linear regression models were used to assess relationship between SQ and SDOC  
 391 and each explanatory variable (distance, network connectivity, catchment altitude, catchment glacier extent and catchment  
 392 vegetation cover) for each timescale or period evaluated. A total of 148 models were analyzed, corresponding to 148  
 393 periods ranging from 0.5 to 365 days. Statistical parameters including  $r^2$ , slope and p-values are provided for each model.  
 394 The horizontal dashed lines represent reference timescales, while vertical red lines indicate the reference slope (at slopes  
 395 of 0) and significance level (p-value < 0.05). Asterisks in the top right panels indicate statistically significant variables



396 for specific time periods. Boxplots summarize the proportion of variance ( $r^2$ ) on the dependent variable ( $S_Q$  and  $S_{DOC}$ )  
397 explained by each group of independent variables. The averages are presented for specific categories described in Figure  
398 1a, namely distance, network connectivity, and catchment properties, within four specific timescale windows (i.e., 1-10,  
399 10-50, 50-100, and 100-365 days).

400

401 This shared control of  $S_Q$  and  $S_{DOC}$  suggests a coupled responsiveness of Q and DOC dynamics to different  
402 processes associated with regional physical weather dynamics, including short-term fluctuations in air  
403 temperature, solar radiation, rain events, as well as intermediate-term fluctuations associated with cold fronts,  
404 ice or snow melting episodes. Interestingly, the influence of pair distance on S dynamics extended up to  
405 timescales of approximately 50 days for DOC, while it appeared to diminish after around 10 days for Q. In  
406 most of the studied headwaters, DOC export is primarily limited by hydrological transport during mobilization  
407 events meaning that additional sources of DOC are mobilized as discharge increases (Zarnetske et al. 2018;  
408 Boix Canadell et al. 2019). This finding supports the observed decoupling between  $S_Q$  and  $S_{DOC}$  synchrony  
409 dynamics within considered timescales.

410 For timescales longer than 50 days, the coefficient of determination ( $r^2$ ) derived from the regression models  
411 including local catchment properties as explanatory variables, was higher than those with stream pair distance  
412 and network connectivity (Figures 5d and 5h). The extent to which local catchment properties controlled  
413 synchrony patterns at the longest timescales depended on both the type of response variable ( $S_Q$  versus  $S_{DOC}$ )  
414 and the specific nature of the catchment property (altitude versus vegetation cover versus glacier extent).  
415 Specifically, the effect of catchment properties on driving the S patterns at the longest timescales was  
416 significantly stronger for  $S_{DOC}$  (median  $r^2_{DOC}=0.17$ ) than for  $S_Q$  (median  $r^2_Q=0.04$ ). In terms of intra-catchment  
417 variations, relative glacier extent emerged as the most important variable driving higher  $S_Q$  at the longest  
418 timescales (Figure S4c and S4d), while vegetation cover, together with altitude, explained most of the variation  
419 in  $S_{DOC}$  for the same timescale bands (Figure S4g and S4h). Fluctuations at timescales longer than 50 days can  
420 be attributed to two major regional signals. First, spring snowmelt drives high spring discharge and associated  
421 DOC transport. Snowmelt is recognized as one of the major drivers determining inter-annual and seasonal  
422 dynamics in lake districts (e.g., Soranno 1999; Baines et al. 2000). Second, glacier ablation results in elevated  
423 glacier runoff and likely increases in DOC yield (Milner et al. 2017). In fact, upon melting, ice locked DOC  
424 can contribute to the downstream DOC pool (Singer et al. 2012; Hood et al. 2015; Li et al. 2018). Additionally,  
425 temperature seasonality across the studied region can lead to large seasonal variations in evapotranspiration,  
426 which in turn influences observed S patterns in Q and DOC by regulating the composition and fluxes of DOC.  
427 Finally, other landscape-dependent regional drivers of S have been identified in other studies, such as droughts,  
428 El Nino events, atmospheric deposition of DOC, although their role in determining synchrony is still being  
429 explored (Dillon and Molot 1997; Futter et al. 2007).

430 Our findings demonstrated that the difference in vegetation cover between stream pairs, but not the glacier  
431 extent, was negatively correlated with  $S_{DOC}$  at the longest timescales (Figure S4g, S4h). This pattern is  
432 consistent with a diversification of vegetation cover classes in the lower altitude catchments below the tree

433 line. On the other hand, headwater catchments covered with similar vegetation classes in the highest parts of  
434 the studied area (i.e., more homogeneous areas covered by moors and mineral surfaces) exhibited more  
435 synchronized inter-annual DOC fluctuations. Unexpectedly, the seasonal fluctuations associated with glacier  
436 ablation drove  $S_Q$  patterns but not  $S_{DOC}$ . This result suggests that  $S_Q$  dynamics are retained regardless the  
437 variability in glacier extent, whether the streams are subject to more less or non-glacier ablation activity.  
438 However, this does not hold true for DOC. One possible explanation is that the amount of DOC transported  
439 by glacier-fed streams is not sufficient or has a source limitation, leading to a lack of homogenization in the  
440  $S_{DOC}$  responses. In other words, the temporal pattern of DOC transported by non-glacier-fed streams, which  
441 drain forested catchments with more groundwater and rain contribution during periods of maximum glacier  
442 ablation (Malard et al., 1999; Tockner et al., 2002), is markedly different and does not overlap seasonally with  
443 that from glacier-fed streams. Ultimately, using a simple space-for-time substitution approach, we predict that  
444 at seasonal and inter-annual scales,  $S$  might decrease for  $Q$  but increase for DOC as glaciers shrink and high-  
445 mountain trees migrate to higher altitudes.

446

#### 447 *Time-series of synchrony from two neighboring streams*

448 To exemplify how processes operating at multiple spatial scales drive the temporal structure (or timescale  
449 structure) of stream  $S_Q$  and  $S_{DOC}$  patterns in high mountains (as discussed in previous section), we analyzed  
450 the time series of synchrony for two of our streams (VAU and VEL, Figure S1) across five specific timescale  
451 ranges (1, 1-10, 10-50, 50-100, and 100-365 days; Figures 6 and S5 $S_{DOC}$  and  $S_Q$ , respectively). These two  
452 streams represent an extreme case of similarity in their exposure to climatic factors (measured as horizontal  
453 proximity; Table S1 and S2), but they differ significantly in their sub-catchment land cover and hydro-  
454 geomorphological attributes (e.g., the relative influence of glacier to vegetated drainage area in the catchment;  
455 Table S1 and S2). As a reference for the maximum expected pair synchrony, we also conducted the same  
456 analysis for two streams (VAU and VAD; Figure S1, Table S1 and S2) that showed the highest similarity  
457 indices for both exposure to climatic factors and sub-catchment attributes. This comparison aims to address  
458 how  $S_{DOC}$  and  $S_Q$  change in high-mountain streams throughout a typical hydrological cycle (including  
459 snowmelt, glacier ablation, and base flow dynamics), but varies across the five specific timescale ranges.

460 At the 1-day and 1- to 10-day timescale bands, the median  $S_{DOC}$  between VAU (glacier-fed stream;  $G$ ) and  
461 VEL (non-glacier-fed stream;  $NG$ ) was 0.63 and 0.49 respectively. These values were comparable to the  $S_{DOC}$   
462 between VAU ( $G$ ) and VAD ( $G$ ) (0.68 and 0.64, respectively; Figure 6). This relatively high agreement  
463 between stream pairs subjected to different glacier influences ( $G-G$  vs.  $G-NG$ ) confirms the role of landscape  
464 position (or proximity) in determining short-term  $S_{DOC}$  dynamics. However, each of the two stream pairs  
465 exhibited a unique seasonal pattern in  $S_{DOC}$ . While both stream pairs showed the highest  $S_{DOC}$  during the initial  
466 part of the snowmelt period (as shown in the left inset), differences in  $S_{DOC}$  were observed between the two  
467 pairs for the rest of the studied periods. For example, the largest differences in  $S_{DOC}$  between the two pairs ( $G-$   
468  $G$  vs.  $G-NG$ ) were observed during the glacial ablation period when the  $S_{DOC}$  for the  $G-NG$  pair was the lowest,  
469 while for the  $G-G$  pair, it peaked most of the time (as shown in the right inset).



470 Consistent with our previous results, the difference in  $S_{\text{DOC}}$  between the two stream pairs (*G-G* vs. *G-NG*)  
471 increased significantly at timescales longer than 10 days. Specifically, the median  $S_{\text{DOC}}$  between VAU (*G*) and  
472 VEL (*NG*) was 0.47, 0.2, and 0.14, while between VAU (*G*) and VAD (*G*) it was 0.6, 0.85, and 0.95 for the  
473 timescale ranges of 10-50, 50-100, and 100-365 days, respectively. This more detailed exploration confirms  
474 that larger differences in local landscape attributes reduce  $S_{\text{DOC}}$  at the longest timescales. Importantly, the low  
475 synchrony between *G-NG* at the timescale associated with glacial ablation activity (while high  $S$  for the *G-G*  
476 pair) further indicates that glaciers play a significant role in desynchronizing DOC dynamics in high-mountain  
477 regions. Our findings show that glaciers impose a unique seasonal regime on DOC, which likely overrides the  
478 effects of other local hydrological or biogeochemical processes occurring during downstream transport within  
479 the main channel or between the main channel and the floodplain (Singer et al. 2012; Hood et al. 2015; Li et  
480 al. 2018; Boix Canadell et al. 2019). In contrast, groundwater-fed streams in high-mountain landscapes (such  
481 as VEL, used for comparison), are typically characterized by a relatively constant DOC concentration regime  
482 (Malard et al. 1999; Tockner et al. 2002), particularly during the glacial ablation period when they show little  
483 seasonal fluctuation compared to glacier-fed streams. Instead, they are mainly driven by short-scale  
484 fluctuations coinciding with rain events (Figure 6). Unlike the pattern of decreasing median  $S_{\text{DOC}}$  with  
485 timescale observed between VAU and VEL, the median  $S_{\text{Q}}$  remained relatively high across all the explored  
486 temporal ranges, resulting in a reduced difference with the reference pair (Figure S5). This observation, using  
487 two contrasting neighboring streams, is consistent with the previous results obtained from all the stream pairs.  
488 Additionally, the  $S_{\text{Q}}$  time series for the two studied stream pairs differed from that of  $S_{\text{DOC}}$ , especially at the  
489 shorter timescales (1 and 1-10 timescale ranges). In this case, apart from some specific events (as shown in the  
490 right inset), the overall coherence between the contrasting and reference stream pairs remained high and  
491 relatively consistent throughout most of the year.

492

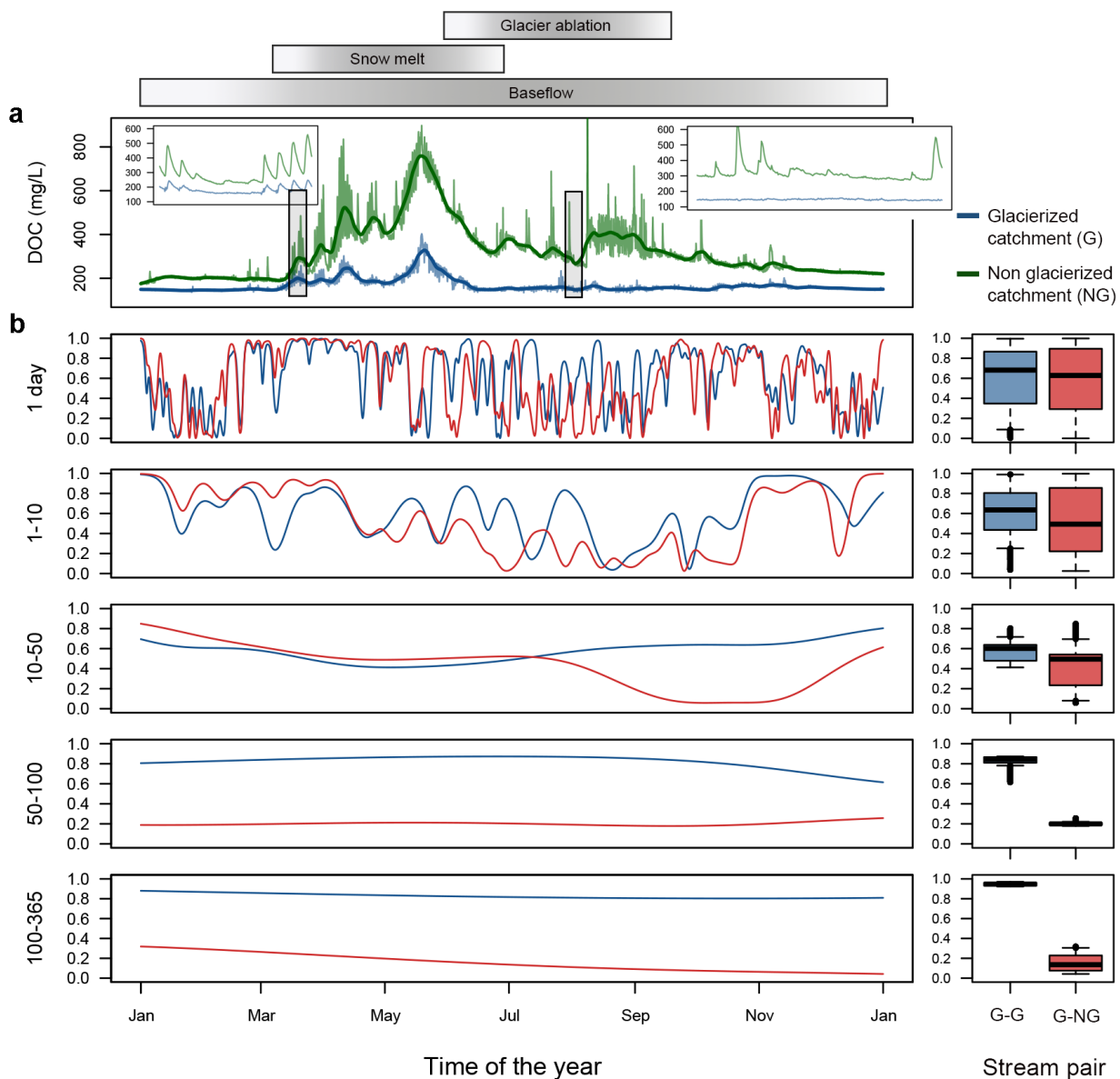
## 493 **Conclusions**

494 Our study provides valuable insights into the temporal dynamics and synchrony patterns of discharge ( $S_{\text{Q}}$ ) and  
495 dissolved organic carbon ( $S_{\text{DOC}}$ ) in high-mountain streams. Through a comprehensive analysis of multiple  
496 timescales, a deeper understanding of the underlying processes driving the dynamics of  $S_{\text{Q}}$  and  $S_{\text{DOC}}$  has been  
497 achieved. We identified distinct mechanisms driving  $S_{\text{Q}}$  and  $S_{\text{DOC}}$  dynamics at different timescales. At shorter  
498 timescales, stream pair distance and network connectivity have emerged as important factors influencing  $S_{\text{Q}}$   
499 and  $S_{\text{DOC}}$ , highlighting the role of linear proximity and connectivity. At longer timescales, local catchment  
500 properties (e.g., altitude, vegetation cover, and glacier extent) exert stronger control on synchrony patterns.  
501 The differential responses of  $S_{\text{Q}}$  and  $S_{\text{DOC}}$  to these catchment properties underscore the unique dynamics of  
502 these variables and their sensitivity to local conditions.

503 The study also contributes to our understanding of the interplay between glaciers and stream dynamics. Glacier  
504 extent induces a significant desynchronizing effect on  $S_{\text{DOC}}$  dynamics, resulting in reduced synchrony between  
505 glacier-fed and non-glacier-fed streams at longer timescales. This highlights the dominant influence of glaciers  
506 on the seasonal regime and variability of stream DOC concentration. Moreover, our study reveals the

507 limitations of the spatial scalability of  $S_{DOC}$  synchrony, suggesting that the concept of sentinel ecosystems  
 508 cannot be directly applied to headwater streams in high-mountain regions due to the low spatial persistence of  
 509  $S_{DOC}$  patterns. It is important to note that our study focuses on a portion of the Swiss Alps, and the extent and  
 510 arrangement of landscape features, such as glacierized and post-glacierized areas, may vary in other high-  
 511 mountain regions. This suggests that landscape synchrony dynamics within and beyond the study area may  
 512 exhibit even greater heterogeneity. Nevertheless, the findings contribute to the fundamental understanding of  
 513 the processes shaping high-mountain stream dynamics, providing insights into the intricate interplay between  
 514 climate, hydrology, and landscape characteristics.

515



516

517 **Figure 6.** High-frequency time series of DOC (a) concentrations and (b) synchrony ( $S_{DOC}$ ) between two selected stream  
 518 pairs across five specific timescale ranges (1, 1-10, 10-50, 50-100 and 100-365 days). The first pair (VAU and VEL)  
 519 consists of two neighboring streams (Figure S1) with contrasting sub-catchment land cover and hydro-geomorphological  
 520 attributes (Table S1 and S2). VAU is influenced by glacier ice melt dynamics (G), while VEL is a groundwater-fed stream

521 draining a non-glacierized catchment (NG). In contrast, the two streams of the second pair (VAU and VAD), both  
522 influenced by glacier dynamics (G), exhibit the highest similarity indices for exposure to climatic factors and sub-  
523 catchment attributes (Table S1 and S2) and are used as a reference for maximum expected pair S. The right panels feature  
524 boxplots comparing differences in S between stream pairs (G-G vs. G-NG) at the five specific timescale ranges. The  
525 boxplots display the 25<sup>th</sup>, 50<sup>th</sup>, and 75<sup>th</sup> percentiles; whiskers represent the 10<sup>th</sup> and 90<sup>th</sup> percentiles.

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532

533 **Author Contribution Statement**

534 L.G.-G. and T.B. conceived the study and wrote the paper with input from N.D. L.G.-G. and N.D. compiled,  
535 processed and analyzed the data. All authors contributed with data and commented on the earlier versions of  
536 this manuscript.

537

538 **Data Availability Statement**

539 The data that support the findings of this study are available through the METALP data (<https://metalp.epfl.ch/>;  
540 <https://doi.org/10.6084/m9.figshare.23522856.v3>).

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542 **References**

- 543 Abbott, B. W., G. Gruau, J. P. Zarnetske, and others. 2018. Unexpected spatial stability of water chemistry in  
544 headwater stream networks J. Grover [ed.]. *Ecol Lett* **21**: 296–308. doi:10.1111/ele.12897
- 545 Anderson, T. L., L. W. Sheppard, J. A. Walter, S. P. Hendricks, T. D. Levine, D. S. White, and D. C. Reuman.  
546 2019. The dependence of synchrony on timescale and geography in freshwater plankton. *Limnol Oceanogr* **64**:  
547 483–502. doi:10.1002/lno.11054
- 548 Baines, S. B., K. E. Webster, T. K. Kratz, S. R. Carpenter, and J. J. Magnuson. 2000. SYNCHRONOUS  
549 BEHAVIOR OF TEMPERATURE, CALCIUM, AND CHLOROPHYLL IN LAKES OF NORTHERN  
550 WISCONSIN. *Ecology* **81**: 815–825. doi:10.1890/0012-9658(2000)081[0815:SBOTCA]2.0.CO;2
- 551 Basu, N. B., G. Destouni, J. W. Jawitz, and others. 2010. Nutrient loads exported from managed catchments  
552 reveal emergent biogeochemical stationarity: BIOGEOCHEMICAL STATIONARITY. *Geophys. Res. Lett.*  
553 **37**: n/a-n/a. doi:10.1029/2010GL045168
- 554 Boix Canadell, M., N. Escoffier, A. J. Ulseth, S. N. Lane, and T. J. Battin. 2019. Alpine Glacier Shrinkage  
555 Drives Shift in Dissolved Organic Carbon Export From Quasi-Chemostasis to Transport Limitation. *Geophys.*  
556 *Res. Lett.* **46**: 8872–8881. doi:10.1029/2019GL083424

- 557 Brown, L. E., D. M. Hannah, and A. M. Milner. 2003. Alpine Stream Habitat Classification: An Alternative  
558 Approach Incorporating the Role of Dynamic Water Source Contributions. *Arctic, Antarctic, and Alpine*  
559 *Research* **35**: 313–322. doi:10.1657/1523-0430(2003)035[0313:ASHCAA]2.0.CO;2
- 560 Cauvy-Fraunié, S., T. Condom, A. Rabatel, M. Villacis, D. Jacobsen, and O. Dangles. 2013. Technical Note:  
561 Glacial influence in tropical mountain hydrosystems evidenced by the diurnal cycle in water levels. *Hydrol.*  
562 *Earth Syst. Sci.* **17**: 4803–4816. doi:10.5194/hess-17-4803-2013
- 563 Cazelles, B., M. Chavez, D. Berteaux, F. Ménard, J. O. Vik, S. Jenouvrier, and N. C. Stenseth. 2008. Wavelet  
564 analysis of ecological time series. *Oecologia* **156**: 287–304. doi:10.1007/s00442-008-0993-2
- 565 Defriez, E. J., D. C. Reuman, and D. Tittensor. 2017. A global geography of synchrony for marine  
566 phytoplankton. *Global Ecol Biogeogr* **26**: 867–877. doi:10.1111/geb.12594
- 567 Defriez, E. J., L. W. Sheppard, P. C. Reid, and D. C. Reuman. 2016. Climate change-related regime shifts have  
568 altered spatial synchrony of plankton dynamics in the North Sea. *Glob Change Biol* **22**: 2069–2080.  
569 doi:10.1111/gcb.13229
- 570 Dillon, P. J., and L. A. Molot. 1997. Dissolved Organic and Inorganic Carbon Mass Balances in Central  
571 Ontario Lakes. *Biogeochemistry* **36**: 29–42.
- 572 Downing, B. D., B. A. Pellerin, B. A. Bergamaschi, J. F. Saraceno, and T. E. C. Kraus. 2012. Seeing the light:  
573 The effects of particles, dissolved materials, and temperature on in situ measurements of DOM fluorescence  
574 in rivers and streams: Effects and compensation for in situ DOM fluorescence. *Limnol. Oceanogr. Methods*  
575 **10**: 767–775. doi:10.4319/lom.2012.10.767
- 576 Fisher, S. G., N. B. Grimm, E. Martí, R. M. Holmes, and J. B. J. Jr. 1998. Material Spiraling in Stream  
577 Corridors: A Telescoping Ecosystem Model. *Ecosystems* **1**: 19–34. doi:10.1007/s100219900003
- 578 Frissell, C. A., W. J. Liss, C. E. Warren, and M. D. Hurley. 1986. A hierarchical framework for stream habitat  
579 classification: Viewing streams in a watershed context. *Environmental Management* **10**: 199–214.  
580 doi:10.1007/BF01867358
- 581 Futter, M. N., D. Butterfield, B. J. Cosby, P. J. Dillon, A. J. Wade, and P. G. Whitehead. 2007. Modeling the  
582 mechanisms that control in-stream dissolved organic carbon dynamics in upland and forested catchments:  
583 MODELING SURFACE WATER DOC. *Water Resour. Res.* **43**. doi:10.1029/2006WR004960
- 584 Gene E. Likens. 1989. *Long-Term Studies in Ecology*, Springer New York.
- 585 Gómez-Gener, L., E. R. Hotchkiss, H. Laudon, and R. A. Sponseller. 2021. Integrating Discharge-  
586 Concentration Dynamics Across Carbon Forms in a Boreal Landscape. *Water Res* **57**.  
587 doi:10.1029/2020WR028806
- 588 Gordon, N. D., T. F. Allnutt, B. L. Finlayson, C. J. Gippel, and R. J. Nathan. 2004. *Stream hydrology: An*  
589 *introduction for ecologists*, John Wiley & Sons Inc.

- 590 Gouhier, T. C., A. Grinsted, and V. Simko. 2018. R package biwavelet: Conduct univariate and bivariate  
591 wavelet analyses (Version 0.20.17). Available from <https://github.com/tgouhier/biwavelet>.
- 592 Grinsted, A., J. C. Moore, and S. Jevrejeva. 2004. Application of the cross wavelet transform and wavelet  
593 coherence to geophysical time series. *Nonlin. Processes Geophys.* **11**: 561–566. doi:10.5194/npg-11-561-2004
- 594 Hampton, S. E., M. D. Scheuerell, M. J. Church, and J. M. Melack. 2019. Long-term perspectives in aquatic  
595 research. *Limnol Oceanogr* **64**. doi:10.1002/lno.11092
- 596 Hood, E., T. J. Battin, J. Fellman, S. O’Neel, and R. G. M. Spencer. 2015. Storage and release of organic  
597 carbon from glaciers and ice sheets. *Nature Geosci* **8**: 91–96. doi:10.1038/ngeo2331
- 598 Hynes, H. B. N. 1975. The stream and its valley., *Verhandlungen der Internationalen Vereinigung für*  
599 *theoretische und angewandte Limnologie*.
- 600 Kirchner, J. W., X. Feng, C. Neal, and A. J. Robson. 2004. The fine structure of water-quality dynamics:  
601 the (high-frequency) wave of the future. *Hydrol. Process.* **18**: 1353–1359. doi:10.1002/hyp.5537
- 602 Kling, G. W., G. W. Kipphut, M. M. Miller, and W. John O’Brien. 2000. Integration of lakes and streams in  
603 a landscape perspective: the importance of material processing on spatial patterns and temporal coherence.  
604 *Freshwater Biol* **43**: 477–497. doi:10.1046/j.1365-2427.2000.00515.x
- 605 Kneib, M., S. Cauvy-Fraunié, N. Escoffier, M. Boix Canadell, Å. Horgby, and T. J. Battin. 2020. Glacier  
606 retreat changes diurnal variation intensity and frequency of hydrologic variables in Alpine and Andean streams.  
607 *Journal of Hydrology* **583**: 124578. doi:10.1016/j.jhydrol.2020.124578
- 608 Kratz, T. K., P. A. Soranno, B. J. Benson, and R. C. Lathrop. 1998. Interannual synchronous dynamics in north  
609 temperate lakes in Wisconsin, USA. *Management of lakes and reservoirs during global climate change* **42**:  
610 273–287.
- 611 Lee, E.-J., G.-Y. Yoo, Y. Jeong, K.-U. Kim, J.-H. Park, and N.-H. Oh. 2015. Comparison of UV–VIS and  
612 FDOM sensors for in situ monitoring of stream DOC concentrations. *Biogeosciences* **12**: 3109–3118.  
613 doi:10.5194/bg-12-3109-2015
- 614 Levin, S. A. 1992. The Problem of Pattern and Scale in Ecology: The Robert H. MacArthur Award Lecture.  
615 *Ecology* **73**: 1943–1967. doi:10.2307/1941447
- 616 Li, X., Y. Ding, J. Xu, and others. 2018. Importance of Mountain Glaciers as a Source of Dissolved Organic  
617 Carbon. *J. Geophys. Res. Earth Surf.* **123**: 2123–2134. doi:10.1029/2017JF004333
- 618 Magnuson, J. J., B. J. Benson, and T. K. Kratz. 1990. Temporal coherence in the limnology of a suite of lakes  
619 in Wisconsin, U.S.A. *Freshwater Biol* **23**: 145–159. doi:10.1111/j.1365-2427.1990.tb00259.x
- 620 Malard, F., K. Tockner, and J. V. Ward. 1999. Shifting Dominance of Subcatchment Water Sources and Flow  
621 Paths in a Glacial Floodplain, Val Roseg, Switzerland. *Arctic, Antarctic, and Alpine Research* **31**: 135–150.  
622 doi:10.1080/15230430.1999.12003291

- 623 Marcé, R., G. George, P. Buscarinu, and others. 2016. Automatic High Frequency Monitoring for Improved  
624 Lake and Reservoir Management. *Environ. Sci. Technol.* **50**: 10780–10794. doi:10.1021/acs.est.6b01604
- 625 McGuire, K. J., C. E. Torgersen, G. E. Likens, D. C. Buso, W. H. Lowe, and S. W. Bailey. 2014. Network  
626 analysis reveals multiscale controls on streamwater chemistry. *Proc. Natl. Acad. Sci. U.S.A.* **111**: 7030–7035.  
627 doi:10.1073/pnas.1404820111
- 628 Milner, A. M., K. Khamis, T. J. Battin, and others. 2017. Glacier shrinkage driving global changes in  
629 downstream systems. *Proc. Natl. Acad. Sci. U.S.A.* **114**: 9770–9778. doi:10.1073/pnas.1619807114
- 630 Rode, M., A. J. Wade, M. J. Cohen, and others. 2016. Sensors in the Stream: The High-Frequency Wave of  
631 the Present. *Environ. Sci. Technol.* **50**: 10297–10307. doi:10.1021/acs.est.6b02155
- 632 Rosset, T., S. Binet, J.-M. Antoine, E. Lerigoleur, F. Rigal, and L. Gandois. 2020. Drivers of seasonal- and  
633 event-scale DOC dynamics at the outlet of mountainous peatlands revealed by high-frequency monitoring.  
634 *Biogeosciences* **17**: 3705–3722. doi:10.5194/bg-17-3705-2020
- 635 Schmidt, S. R., G. Lischeid, T. Hintze, and R. Adrian. 2019. Disentangling limnological processes in the time-  
636 frequency domain. *Limnol. Oceanogr.* **64**: 423–440. doi:10.1002/lno.11049
- 637 Sheppard, L. W., J. R. Bell, R. Harrington, and D. C. Reuman. 2016. Changes in large-scale climate alter  
638 spatial synchrony of aphid pests. *Nature Clim Change* **6**: 610–613. doi:10.1038/nclimate2881
- 639 Singer, G. A., C. Fasching, L. Wilhelm, J. Niggemann, P. Steier, T. Dittmar, and T. J. Battin. 2012.  
640 Biogeochemically diverse organic matter in Alpine glaciers and its downstream fate. *Nature Geosci* **5**: 710–  
641 714. doi:10.1038/ngeo1581
- 642 Soranno, P. A. 1999. Spatial Variation among Lakes within Landscapes: Ecological Organization along Lake  
643 Chains. *Ecosystems* **2**: 395–410. doi:10.1007/s100219900089
- 644 Tank, S. E., J. B. Fellman, E. Hood, and E. S. Kritzberg. 2018. Beyond respiration: Controls on lateral carbon  
645 fluxes across the terrestrial-aquatic interface: Controls on lateral carbon fluxes. *Limnol. Oceanogr.* **3**: 76–88.  
646 doi:10.1002/lol2.10065
- 647 Thompson, S. E., N. B. Basu, J. Lascurain, A. Aubeneau, and P. S. C. Rao. 2011. Relative dominance of  
648 hydrologic versus biogeochemical factors on solute export across impact gradients: HYDROLOGY  
649 CONTROLS SOLUTE EXPORT. *Water Resour. Res.* **47**. doi:10.1029/2010WR009605
- 650 Tockner, K., F. Malard, U. Uehlinger, and J. V. Ward. 2002. Nutrients and organic matter in a glacial river-  
651 floodplain system (Val Roseg, Switzerland). *Limnol. Oceanogr.* **47**: 266–277. doi:10.4319/lo.2002.47.1.0266
- 652 Torrence, C., and G. P. Compo. 1998. A Practical Guide to Wavelet Analysis. *Bulletin of the American*  
653 *Meteorological Society* **79**: 18.
- 654 Turner, M. G., V. H. Dale, and R. H. Gardner. 1989. Predicting across scales: Theory development and testing.  
655 *Landscape Ecol* **3**: 245–252. doi:10.1007/BF00131542

- 656 Vannote, R. L., G. W. Minshall, K. W. Cummins, J. R. Sedell, and C. E. Cushing. 1980. The River Continuum  
657 Concept. *Can. J. Fish. Aquat. Sci.* **37**: 130–137. doi:10.1139/f80-017
- 658 Walter, J. A., L. W. Sheppard, T. L. Anderson, J. H. Kastens, O. N. Bjørnstad, A. M. Liebhold, and D. C.  
659 Reuman. 2017. The geography of spatial synchrony B. Blasius [ed.]. *Ecol Lett* **20**: 801–814.  
660 doi:10.1111/ele.12782
- 661 Watras, C. J., P. C. Hanson, T. L. Stacy, K. M. Morrison, J. Mather, Y.-H. Hu, and P. Milewski. 2011. A  
662 temperature compensation method for CDOM fluorescence sensors in freshwater: CDOM temperature  
663 compensation. *Limnol. Oceanogr. Methods* **9**: 296–301. doi:10.4319/lom.2011.9.296
- 664 Webster, K. E., T. K. Kratz, C. J. Bowser, J. J. Magnuson, and W. J. Rose. 1996. The influence of landscape  
665 position on lake chemical responses to drought in northern Wisconsin. *Limnol. Oceanogr.* **41**: 977–984.  
666 doi:10.4319/lo.1996.41.5.0977
- 667 Wilkinson, G. M., J. Walter, R. Fleck, and M. L. Pace. 2020. Beyond the trends: The need to understand  
668 multiannual dynamics in aquatic ecosystems. *Limnol Oceanogr Letters* **5**: 281–286. doi:10.1002/lo12.10153
- 669 Wynne, R. H., J. J. Magnuson, M. K. Clayton, T. M. Lillesand, and D. C. Rodman. 1996. Determinants of  
670 temporal coherence in the satellite-derived 1987-1994 ice breakup dates of lakes on the Laurentian Shield.  
671 *Limnol. Oceanogr.* **41**: 832–838. doi:10.4319/lo.1996.41.5.0832
- 672 Zarnetske, J. P., M. Bouda, B. W. Abbott, J. Saiers, and P. A. Raymond. 2018. Generality of Hydrologic  
673 Transport Limitation of Watershed Organic Carbon Flux Across Ecoregions of the United States. *Geophys.*  
674 *Res. Lett.* **45**. doi:10.1029/2018GL080005
- 675



676 Supplementary information for:

677 **Synchrony dynamics of dissolved organic carbon in high-mountain streams: insights into**  
678 **scale-dependent processes**

679 Lluís Gómez-Gener<sup>1,2\*</sup>, Nicola Deluigi<sup>1</sup>, Tom Battin<sup>1</sup>

680 <sup>1</sup> River Ecosystems Research Laboratory, Alpine and Polar Environmental Research Center (ALPOLE),  
681 Ecole Polytechnique Fédérale de Lausanne, EPFL, Switzerland

682 <sup>2</sup> Centre for Research on Ecology and Forestry Applications, Universitat Autònoma de Barcelona, Campus  
683 de Bellaterra, Edifici C, 08193 Cerdanyola del Vallès, Barcelona, Spain

684 \* E-mail: gomez.gener87@gmail.com (Corresponding author)

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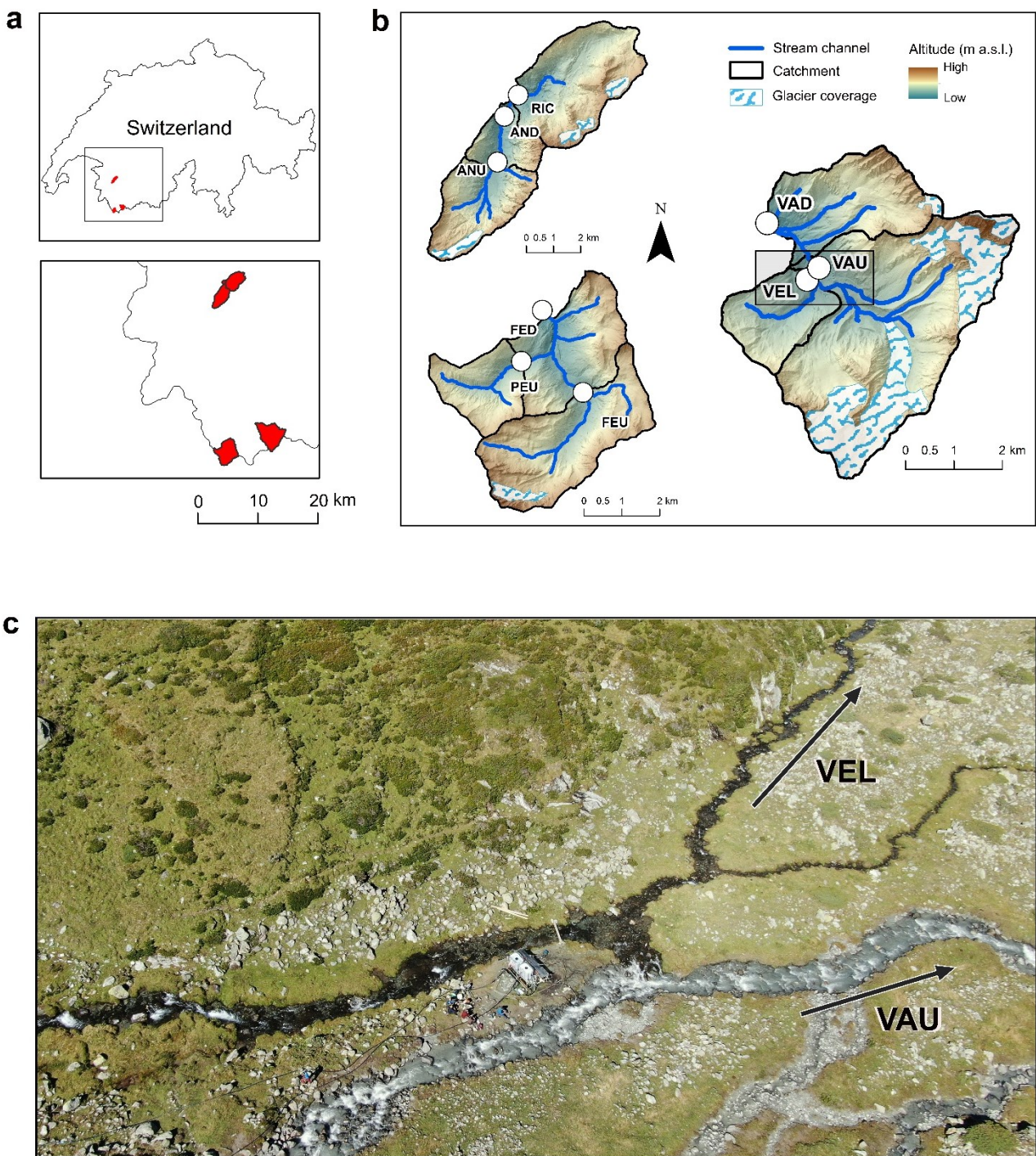
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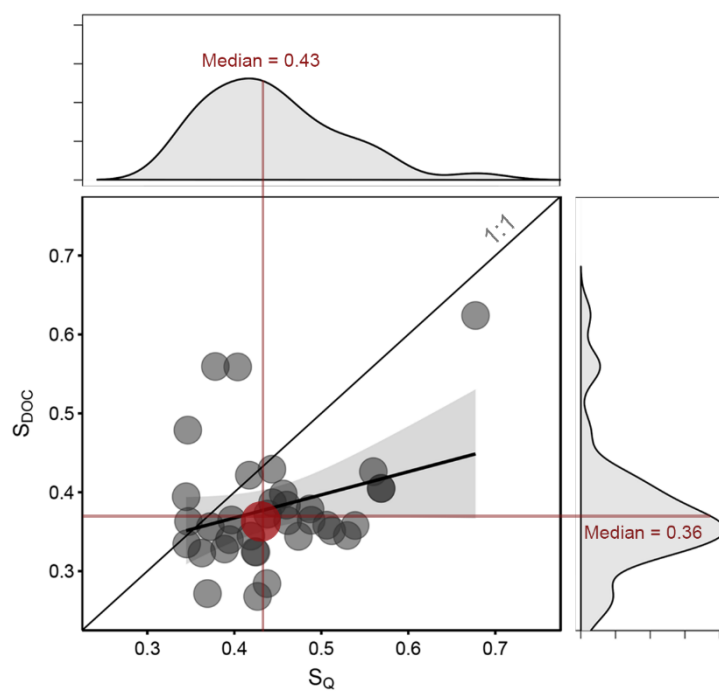
706 **Supplementary Figures**



707

708 **Figure S1.** (a) Study region with the catchments and (b) stream sampling locations in the Swiss Alps. Panel (c) shows a  
 709 visual example of two streams (VEL and VAU) representing an extreme case of similarity in their exposure to climatic  
 710 factors (measured as horizontal proximity; see Table S1 and S2) but dissimilarity in their sub-catchment land cover and  
 711 hydro-geomorphological attributes. VEL represents a groundwater-fed stream typical of high-mountain landscapes, while  
 712 VAU represents a glacier-fed stream typical of high-mountain landscapes.

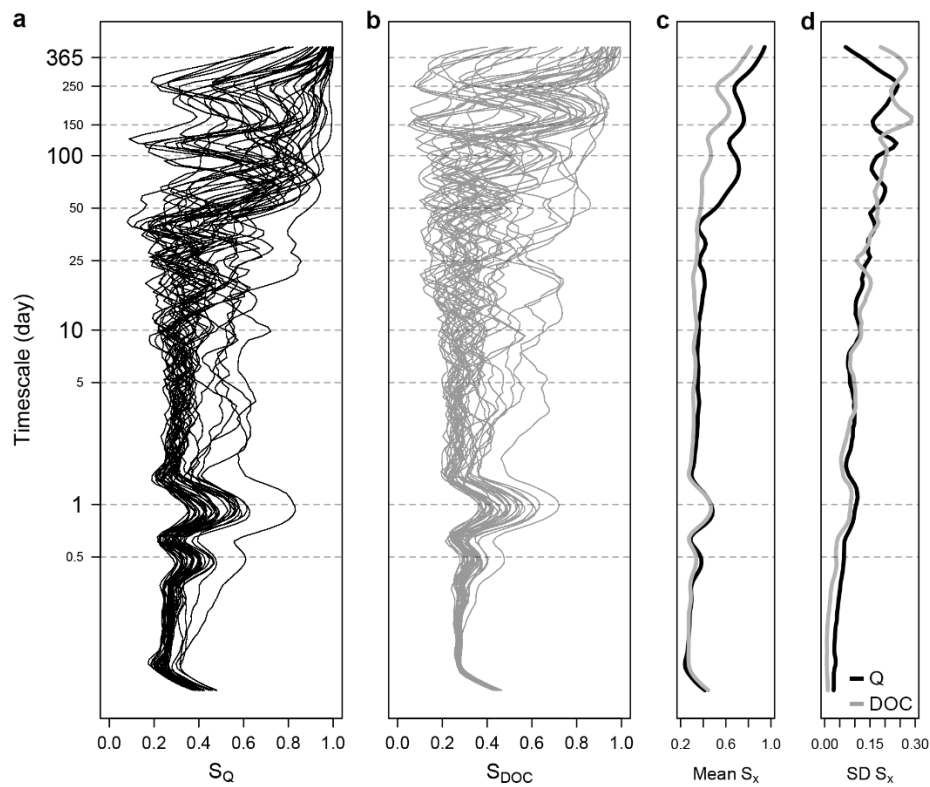
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715 [Figure S2. Relationship between timescale-aggregated  \$S\_Q\$  and  \$S\_{DOC}\$  for all stream pairs. The red circles indicate the](#)  
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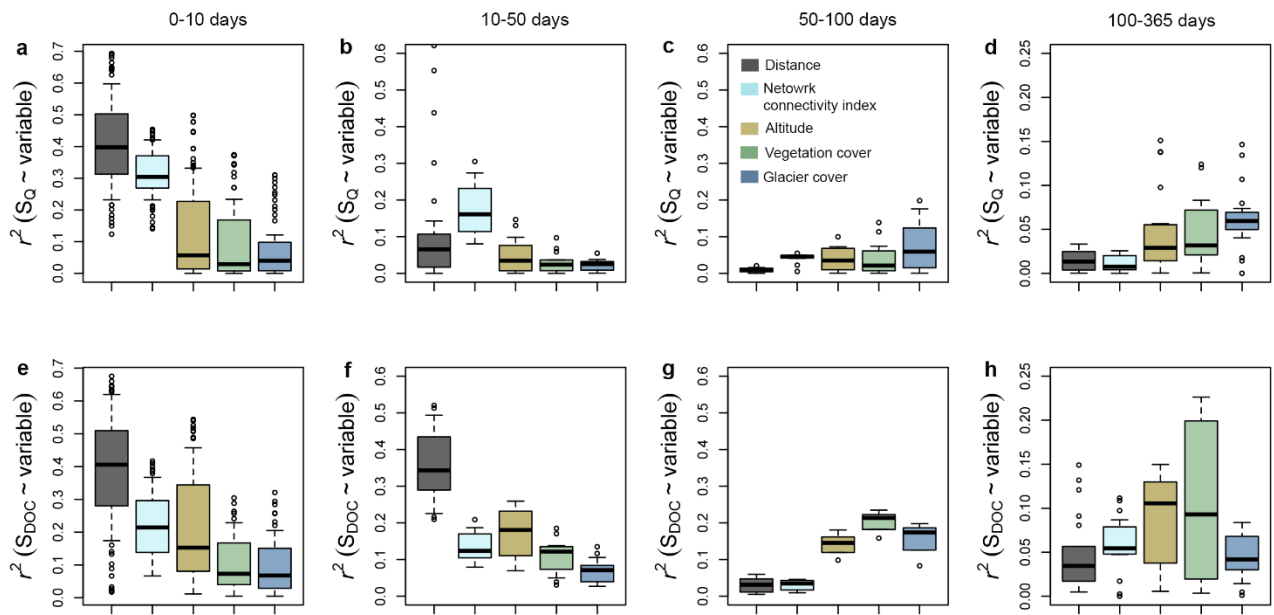


718

719 **Figure S3.** Temporal structure of (a) discharge synchrony ( $S_Q$ ) and (b) dissolved organic carbon synchrony ( $S_{DOC}$ ) for  
720 the 36 stream pairs assessed in the study. The temporal structure of the (c) mean and (d) standard deviation of  $S_Q$  (black  
721 line) and  $S_{DOC}$  (grey line) is depicted. The y-axis represents the log10-transformed period length or timescale of oscillation  
722 (in days). Horizontal dashed lines represent reference timescales.

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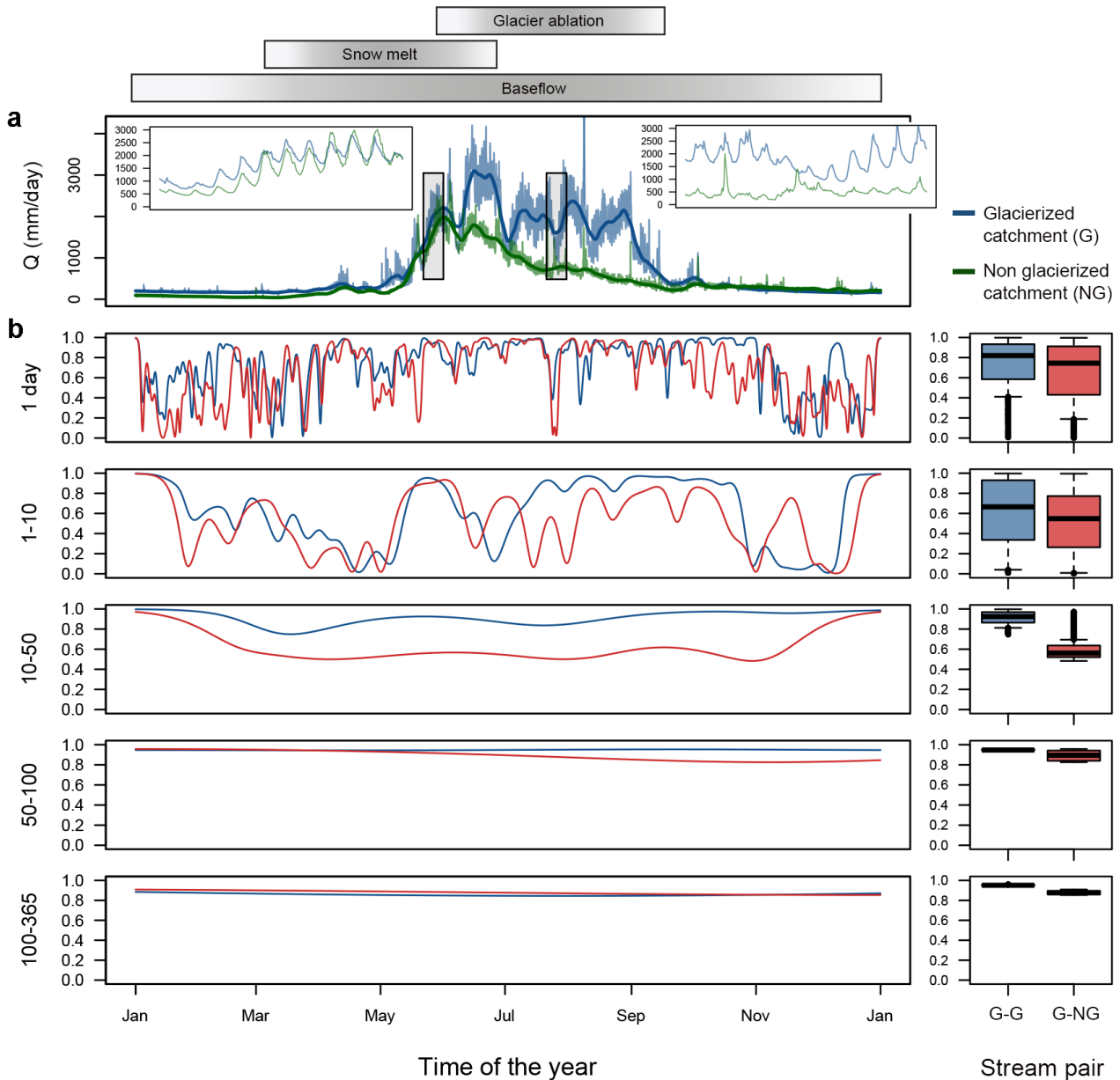
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[Figure S4](#). Boxplots summarizing the proportion of variance explained by each individual explanatory variable on  $S_Q$  and  $S_{DOC}$  at four specific timescale ranges (i.e., 1-10, 10-50, 50-100, and 100-365 days).



729

730 **Figure S5.** High-frequency time series of (a) discharge (Q) and (b) runoff synchrony ( $S_Q$ ) between two selected stream  
 731 pairs across five specific timescale ranges (1, 1-10, 10-50, 50-100, and 100-365 days). The first pair (VAU and VEL)  
 732 consists of two neighboring streams (Figure S1) that differ in their sub-catchment land cover and hydro-geomorphological  
 733 attributes (Table S1 and S2). VAU remains under the influence of glacier ice melt dynamics (G), while VEL is a  
 734 groundwater-fed stream draining a non-glacierized catchment (NG). On the other hand, the two streams in the second pair  
 735 (VAU and VAD), both influenced by glacier dynamics (G), show high similarity indices in terms of exposure to climatic  
 736 factors and sub-catchment attributes (Table S1 and S2) and are used as reference for the maximum expected pair S. The  
 737 right panels display boxplots comparing differences in S between stream pairs (G-G vs. G-NG) at the five specific  
 738 timescale ranges. Box plots show the 25<sup>th</sup>, 50<sup>th</sup>, and 75<sup>th</sup> percentiles while whiskers represent the 10<sup>th</sup> and 90<sup>th</sup>  
 739 percentiles. To normalize and compare discharge from the two streams with different catchment areas, we report specific discharge  
 740 (mm day<sup>-1</sup>) in panel (a).

741



742 **Supplementary Tables**

743 **Table S1.** Geographical and landscape characteristics of the nine streams in the Swiss Alps.

Stream code	Stream location		Catchment properties				
	Latitude (°)	Longitud (°)	Altitude (m a.s.l.)	Surface area (km <sup>2</sup> )	Vegetation (%) <sup>a</sup>	Glacier and perpetual snow (%) <sup>a</sup>	Mineral surfaces (%) <sup>a</sup>
VAU	45.9295	7.2446	2148	18.1	21.1	33.5	45.4
VAD	45.9350	7.2269	1937	23.2	24.2	27.4	48.4
VEL	45.5568	7.1475	2161	3.1	56.7	0.0	43.3
FEU	45.8831	7.1309	1995	9.3	46.3	7.4	46.3
FED	45.9051	7.1156	1774	20.2	62.4	3.4	34.2
PEU	45.8937	7.1080	2027	4.0	70.2	0.0	29.8
ANU	46.2316	7.1020	1465	9.0	54.0	6.8	39.2
AND	46.2534	7.1096	1201	13.4	63.9	4.6	31.5
RIC	46.2535	7.1101	1200	14.3	64.2	6.4	29.4

744

745 [<sup>a</sup> Vegetation, glaciers, and perpetual snow, and mineral surfaces cover percentages based on the CORINE Land Cover Inventory 2012.](#)  
 746 [The percent vegetation cover includes mixed and coniferous forests, moors, heathlands, pastures, sparsely vegetated areas, and natural](#)  
 747 [grassland land cover layers.](#)

748

749

**Table S2.** Similarity index between stream pairs for the geographical and landscape characteristics of the nine streams in the Swiss Alps.

750

**Linear distance (km)**

	VAU	VAD	VEL	FEU	FED	PEU	ANU	AND	RIC
VAU									
VAD	1.03								
VEL	0.22	1.22							
FEU	10.20	9.66	10.21						
FED	10.46	9.64	10.36	2.20					
PEU	11.37	11.37	10.69	1.70	1.42				
ANU	35.25	34.57	35.53	38.32	36.29	37.56			
AND	37.13	36.64	37.37	40.63	38.62	39.91	2.46		
RIC	37.23	36.74	37.47	40.73	38.72	40.01	2.53	0.10	

**Stream connectivity index**

	VAU	VAD	VEL	FEU	FED	PEU	ANU	AND	RIC
VAU									
VAD	2								
VEL	0	1							
FEU	0	0	0						
FED	0	0	0	2					
PEU	0	0	0	0	1				
ANU	0	0	0	0	0	0			
AND	0	0	0	0	0	0	2		
RIC	0	0	0	0	0	0	0	0	

**Mean catchment altitude (m a.s.l.)**

	VAU	VAD	VEL	FEU	FED	PEU	ANU	AND	RIC
VAU									
VAD	2042.5								
VEL	2154.5	2049.0							
FEU	2071.5	1966.0	2078.0						
FED	1961.0	1855.5	1967.5	1884.5					
PEU	2087.5	1982.0	2094.0	2011.0	1900.5				
ANU	1806.5	1701.0	1813.0	1730.0	1619.5	1746.0			
AND	1674.5	1569.0	1681.0	1598.0	1487.5	1614.0	1333.0		
RIC	1674.0	1568.5	1680.5	1597.5	1487.0	1613.5	1332.5	1200.5	

**Mean vegetation cover (%)**

	VAU	VAD	VEL	FEU	FED	PEU	ANU	AND	RIC
VAU									
VAD	22.6								
VEL	38.9	40.5							
FEU	33.7	35.2	51.5						
FED	41.7	43.3	59.6	54.3					
PEU	45.7	47.2	63.5	58.2	66.3				
ANU	37.5	39.1	55.4	50.1	58.2	62.1			
AND	42.5	44.1	60.3	55.1	63.2	67.1	59.0		
RIC	42.7	44.2	60.5	55.2	63.3	67.2	59.1	64.1	

**Mean glacier cover (%)**

	VAU	VAD	VEL	FEU	FED	PEU	ANU	AND	RIC
VAU									
VAD	30.4								
VEL	16.8	13.7							
FEU	20.5	17.4	3.7						
FED	18.5	15.4	1.7	5.4					
PEU	16.8	13.7	0.0	3.7	1.7				
ANU	20.2	17.1	3.4	7.1	5.1	3.4			
AND	19.0	16.0	2.3	6.0	4.0	2.3	5.7		
RIC	19.9	16.9	3.2	6.9	4.9	3.2	6.6	5.5	

**Mean mineral surface cover (%)**

	VAU	VAD	VEL	FEU	FED	PEU	ANU	AND	RIC
VAU									
VAD	48.9								
VEL	41.1	42.4							
FEU	52.1	53.3	45.6						
FED	44.4	45.6	37.9	48.8					
PEU	43.2	44.5	36.7	47.7	40.0				
ANU	49.2	50.4	42.7	53.6	45.9	44.8			
AND	46.5	47.8	40.1	51.0	43.3	42.1	48.1		
RIC	45.4	46.6	38.9	49.8	42.1	41.0	46.9	44.3	

751

752



753 **Table S3.** Statistical summary of regressions between timescale aggregated  $S_Q$  (above) and  $S_{DOC}$  (below) and the set of  
 754 potential explanatory variables.

$S_Q \sim$ variable	$r^2$	slope	$p$ -value
<b>Distance</b>	<b>0.49</b>	<b>-0.03</b>	<b>&lt; 0.001</b>
<b>Network connectivity</b>	<b>0.22</b>	<b>0.060</b>	<b>&lt; 0.001</b>
Altitude	0.01	0.001	0.65
Glacier cover	0.06	0.002	0.15
Vegetation cover	0.05	-0.002	0.20
<b><math>S_{DOC} \sim</math> variable</b>			
<b>Distance</b>	<b>0.57</b>	<b>-0.05</b>	<b>&lt; 0.001</b>
<b>Network connectivity</b>	<b>0.20</b>	<b>0.058</b>	<b>&lt; 0.001</b>
<b>Altitude</b>	<b>0.21</b>	<b>0.001</b>	<b>&lt; 0.001</b>
Glacier cover	0.06	0.003	0.15
<b>Vegetation cover</b>	<b>0.13</b>	<b>-0.003</b>	<b>0.03</b>

755

756

757 **Table S4.** Statistical summary of regressions between  $S_Q$  and  $S_{DOC}$  and distance between stream locations for both  
 758 timescale aggregated (measured as the median coherence for all the timescales) and timescale segregated at six specific  
 759 timescales or timescale bands (i.e., 1, 1-10, 10-50, 50-100, 100-365, and 365 days).

Timescale band (days)	$S_Q \sim \text{distance}$			$S_{DOC} \sim \text{distance}$		
	$r^2$	slope	p-value	$r^2$	slope	p-value
Aggregated	<b>0.49</b>	<b>-0.03</b>	<b>&lt; 0.001</b>	<b>0.57</b>	<b>-0.05</b>	<b>&lt; 0.001</b>
1	<b>0.52</b>	<b>-0.05</b>	<b>&lt; 0.001</b>	<b>0.39</b>	<b>-0.03</b>	<b>&lt; 0.001</b>
10-ene	<b>0.53</b>	<b>-0.03</b>	<b>&lt; 0.001</b>	<b>0.58</b>	<b>-0.02</b>	<b>&lt; 0.001</b>
10-50	0.18	-0.03	0.06	<b>0.46</b>	<b>-0.06</b>	<b>&lt; 0.001</b>
50-100	-0.02	0.001	0.96	<b>0.12</b>	<b>-0.04</b>	<b>0.02</b>
100-365	-0.02	-0.01	0.55	0.03	-0.02	0.14
365	0.01	0.01	0.73	<b>0.14</b>	<b>-0.05</b>	<b>0.01</b>

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