

Recent methane surges reveal heightened emissions from tropical inundated areas

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This is a non-peer reviewed preprint that has been submitted to *Nature Communications*.

30 **Abstract**

31 Record breaking atmospheric methane growth rates were observed in 2020 and 2021 (15.2 ± 0.4
32 and 17.6 ± 0.5 ppb yr⁻¹), reaching their highest level since the commencement of ground-based
33 observations in the early 1980s. Here we use an ensemble of atmospheric inversions informed
34 by surface or satellite methane concentration observations to infer emission changes during
35 these two years relative to 2019. We found a global increases of methane emissions of 20.3 ± 9.9
36 Tg CH₄ in 2020 and 24.8 ± 3.1 Tg CH₄ in 2021. The emission rise was dominated by tropical
37 and boreal regions with inundated areas, as a result of elevated groundwater table. Strong,
38 synchronous, and persistent emission increases occurred in regions such as the Niger River
39 basin, the Congo basin, the Sudd swamp, the Ganges floodplains and Southeast Asian deltas
40 and the Hudson Bay lowlands. These regions alone contributed about 70% and 60% of the net
41 global increases in 2020 and 2021, respectively. Comparing our top-down estimates with
42 simulation of wetland emissions by biogeochemical models, we find that the bottom-up models
43 significantly underestimate the intra- and inter-annual variability of methane sources from
44 tropical inundated areas. This discrepancy likely arises from the models' limitations in
45 accurately representing the dynamics of tropical wetland extents and the response of methane
46 emissions to environmental changes. Our findings demonstrate the critical role of tropical
47 inundated areas in the recent surge of methane emissions and highlight the value of integrating
48 multiple data streams and modeling tools to better constrain tropical wetland emissions.

49 Main

50 In the years 2020 and 2021, the methane growth rate (MGR) in the atmosphere reached
51 15.2 ± 0.4 and 17.6 ± 0.5 parts per billion per year (ppb yr^{-1}) respectively, hitting record high
52 since systematic measurements started in early 1980s by NOAA's Global Monitoring
53 Laboratory (GML) (Lan et al., 2023; https://gml.noaa.gov/ccgg/trends_ch4/). The
54 unprecedented methane growth during 2020 and 2021 coincided with the reduced human
55 activities and pollutant emissions during COVID-19 lockdowns and the gradual recovery
56 afterwards (Davis et al., 2022; Jackson et al., 2022; Laughner et al., 2021; Miyazaki et al.,
57 2021), together with the occurrence of a moderate and prolonged La Niña event (Li et al., 2022;
58 <https://psl.noaa.gov/enso/mei/>), which offers a unique opportunity to examine the drivers of
59 methane variabilities on a year-to-year basis.

60 Both process-based studies of sources ("bottom-up" estimates) and atmospheric-based inverse
61 analyses ("top-down" estimates) pointed to pronounced emission growth in 2020 compared to
62 2019, arising from tropical and northern sources, likely driven by enhanced wetland emissions
63 (Feng et al., 2023; Peng et al., 2022; Qu et al., 2022; Zhang et al., 2023), the main source
64 component of natural methane emissions. This is consistent with the overall earlier and larger
65 increase of MGR observed in the Tropics and Northern high-latitudes than Southern extra-
66 tropics from marine boundary layer sites of the surface network (Fig. 1b; Supplementary Fig.
67 1). The exceptionally high methane growth occurred again in 2021 over most latitude bands,
68 although followed by a drop in MGR towards the end of the year (Fig. 1a, b). Observations of
69 total column methane concentrations (XCH_4) by Greenhouse Gases Observing SATellite
70 (GOSAT), whether obtained from the National Institute for Environmental Studies (Japan) full
71 physics retrievals (hereafter "GSNIES"; Inoue et al., 2016) or from the University of Leicester
72 proxy retrievals (hereafter "GSUoL"; Parker et al., 2020), confirmed the unexpected methane
73 surge in 2020 and 2021 (Fig. 1c, d). Note that both GOSAT retrievals showed a larger increase
74 in MGR over Southern extra-tropics than that observed from surface network, with GSUoL
75 exhibiting higher MGR globally and in the Tropics as well (Fig. 1; Supplementary Fig. 1).

76 The zonally-averaged changes in MGR reveal the integrated variations of regional sources and
77 sinks, atmospheric transport, and removal by OH. To infer the spatiotemporal patterns in flux
78 changes from 2019 to 2021, we applied three-dimensional (3D) atmospheric inversions using
79 an inversion system PYVAR-LMDZ-SACS that assimilated either surface or satellite-based

80 CH₄ observations. An ensemble of six inversions was performed using the same inversion setup,
81 but different in assimilated observation datasets (surface network, GSNIES or GSUoL) and
82 transport model physical parameterizations (the “classic” and “standard” versions)
83 (Supplementary Table 1; see Methods). This allows us to test the consistency of flux change
84 patterns inferred from different types of measurements while accounting for some of the
85 uncertainties due to imperfect representation of atmospheric mixing. Surface networks offer a
86 good coverage of northern mid-to-high latitudes (especially over Europe and North America),
87 whereas satellite data have improved data densities from 60°S to 60°N, including the tropics
88 (Supplementary Figs. 2 & 3). We prescribed changes in OH concentration fields simulated
89 from a full chemistry transport model LMDZ-INCA (Hauglustaine et al., 2004, 2014), driven
90 by interannually varying meteorology from ECMWF ERA5 reanalysis (Hersbach et al., 2018)
91 with natural and anthropogenic emissions of NO_x, CO and hydrocarbons updated to 2021 that
92 account for reductions and rebound of NO_x and CO emissions by anthropogenic activities and
93 changes in wildfires (Carbon Monitor; Community Emissions Data System (CEDS); van der
94 Werf et al., 2017). Our chemistry transport model LMDZ-INCA simulated a reduction of
95 global tropospheric OH by 3% in 2021 relative to 2019, more than the 1.6–1.8% decrease in
96 2020 as reported by Peng et al. (2022) (Supplementary Fig. 4 & Table 2).

97 The ensemble of six inversions showed global increases in surface CH₄ emissions by an
98 average of 20.3±9.9 Tg CH₄ yr⁻¹ and 24.8±3.1 Tg CH₄ yr⁻¹ in 2020 and 2021 respectively,
99 compared to 2019. A large portion of this surge in global emissions was accounted for by the
100 northern tropics (0°–30°N), which contributed about 80% (16.2±8.3 Tg CH₄ yr⁻¹) and 95%
101 (23.2±4.0 Tg CH₄ yr⁻¹) of the global increases in 2020 and 2021, respectively (Fig. 2a, b;
102 Supplementary Fig. 5 & Table 3). Strong emission increases were consistently found over
103 tropical Africa and Southeast Asia in both years, according to all six inversions (Fig. 2a, b;
104 Supplementary Figs. 5–7). Overall, most of the regions with strong and consistent emission
105 changes overlap with major wetland complexes and inundated areas. These trends align well
106 with changes in Gravity Recovery and Climate Experiment Follow-On (GRACE-FO) liquid
107 water equivalent (LWE) height anomalies (Fig. 2, Supplementary Fig. 5), a proxy for water-
108 table depth and water stored in wetland systems (Bloom et al., 2010).

109 Focusing on 20 major wetland regions that represent about 60% of global wetland areas based
110 on the regularly flooded wetland map of Tootchi et al. (2019), a significant correlation was
111 noted between top-down estimates of emission anomalies and changes in LWE from GRACE-

112 FO (Fig. 2). Notably, the Niger River basin, the Congo basin, the Sudd swamp, the Ganges
113 floodplains and Southeast Asian deltas in the tropics, and the Hudson Bay lowlands in the
114 boreal region, exhibited persistent emission enhancements in response to increased LWE.
115 Emission increases over each of these regions were at least 1.5–2 times the interannual
116 variability ($1.5-2\sigma$) of methane emissions during 2010–2019, based on a former study that also
117 used the inversion system PYVAR-LMDZ-SACS and GSUoL as constraints for CH₄ (Zheng
118 et al., 2019). The six wetland regions together contributed around 70% (14.1 ± 4.2 Tg CH₄ yr⁻¹)
119 and 60% (14.9 ± 2.6 Tg CH₄ yr⁻¹) of the global emission increases for 2020 and 2021,
120 respectively (Fig. 2 & Fig. 3e, f; Supplementary Fig. 5 & Table 4). The strong and synchronous
121 emission increases over these inundated regions were also coincident with the occurrence of
122 La Niña (Li et al., 2022), which supports previous findings that showed enhanced wetland
123 emissions in the tropics and boreal North America during historical La Niña periods (Hodson
124 et al., 2011; Pandey et al., 2017; Zhu et al., 2017). Other regions, such as the western Siberian
125 lowlands in 2020 and the Amazon Basin and the Orinoco floodplain in 2021, also show
126 substantial emission increases, but reduced emissions were seen in the other year over these
127 regions, aligning with corresponding variations in LWE. Conversely, the Pantanal and the
128 Paraná floodplains in central and southeastern South America exhibited consistent reduction in
129 emissions in both 2020 and 2021, likely due to continued drier conditions and lower water
130 tables (Fig. 2; Supplementary Fig. 5 and Table 4).

131 Bottom-up inventories and process-based wetland emission models corroborate increased
132 emissions over the northern tropics during 2020–2021, albeit with much smaller magnitudes
133 than inversions (Fig. 2 & 3; Supplementary Fig. 8 and Table 3). Globally, the net emission
134 changes were -0.7 ± 4.0 Tg CH₄ yr⁻¹ and 6.9 ± 4.1 Tg CH₄ yr⁻¹ in 2020 and 2021 relative to 2019
135 based on bottom-up methodologies, with emission increases of 3.2 ± 1.3 Tg CH₄ yr⁻¹ and
136 4.1 ± 1.9 Tg CH₄ yr⁻¹ estimated for the northern tropics (Fig. 3a, b; Supplementary Fig. 8 and
137 Table 3). A breakdown into different processes showed that the anthropogenic CH₄ emissions
138 in 2020 and 2021 were higher than the 2019 level by 0.5 Tg CH₄ yr⁻¹ and 4.7 Tg CH₄ yr⁻¹,
139 respectively (Fig. 3a, b; Supplementary Fig. 8 and Table 3). The global fire CH₄ emissions
140 declined in 2020 by 6.5 Tg CH₄ yr⁻¹ relative to 2019, mainly contributed by reduced fire
141 emissions of 5.1 Tg CH₄ yr⁻¹ in the southern tropics (30°S–0°) (van der Werf et al., 2017). A
142 similar reduction of fire emissions occurred again in the southern tropics in 2021, but the
143 extreme fires in boreal North America and eastern Siberia during the hotter and drier

144 summertime (Zheng et al., 2023) led to an increase in emissions of 4.7 Tg CH₄ yr⁻¹ in the
145 northern extra-tropics and therefore a net global emission reduction of only 1.8 Tg CH₄ yr⁻¹
146 relative to 2019 (Fig. 3a, b; Supplementary Fig. 8 and Table 3). For wetlands, an ensemble of
147 process-based wetland model simulations from ORCHIDEE-MICT and LPJ-wsl driven by four
148 different climate forcings (see Methods) reported an increase in global wetland emissions by
149 5.3±4.0 Tg CH₄ yr⁻¹ in 2020 compared to 2019, which was dominated by enhanced wetland
150 emissions in the northern extra-tropics and tropics as a result of warmer and wetter climate
151 (Peng et al., 2022). The updated simulations showed slightly smaller wetland emission
152 increases by 4.0±4.1 Tg CH₄ yr⁻¹ in 2021 compared to 2019, with similar latitudinal patterns
153 (Fig. 3a, b; Supplementary Figs. 8 & 9 and Table 3).

154 Given the spread of simulated wetland emissions (Supplementary Fig. 9), the total emission
155 increases from bottom-up estimates were only 7.4 Tg CH₄ yr⁻¹ and 14.4 Tg CH₄ yr⁻¹ in 2020
156 and 2021 if the maximum increase of wetland emissions were considered, which are still 12.9
157 Tg CH₄ yr⁻¹ and 10.4 Tg CH₄ yr⁻¹ below the mean of the inversion ensemble estimates. While
158 there was a general agreement in the broad spatial patterns of emission changes between the
159 two methodologies, large discrepancies were found over the tropics. The strong and persistent
160 emission increases in tropical Africa and Asia, as inferred from atmospheric CH₄ observations,
161 were consistently underestimated by bottom-up estimates (Fig. 2 & 3; Supplementary Fig. 8).
162 Specifically, for the Niger River basin, the Congo basin, the Sudd swamp, the Ganges
163 floodplains and Southeast Asian deltas, the bottom-up estimates of emission increases were
164 only about 11% and 5% of the estimates from top-down inversions (Fig. 3e, f; Supplementary
165 Table 4).

166 The lower bottom-up estimates of emission increases over tropical Africa and Asia inundated
167 areas suggest that biogeochemical models substantially underestimate emissions from these
168 water-logged ecosystems in response to changes in water-table depth and moisture conditions.
169 Indeed, a close look into a tropical wetland emission hotspot, the Sudd swamp in the eastern
170 Africa (3°N–17°N, 25°E–40°E; see also Lunt et al., 2019, 2021; Pandey et al., 2021) showed
171 that the annual variations in top-down CH₄ emissions and GRACE-FO LWE anomalies were
172 highly correlated (Supplementary Fig. 10; see also Lunt et al., 2019), implying a strong impact
173 of water-table depth on seasonal emissions. The annual peak of LWE occurs during
174 September–October, about one month preceding the annual peak of emissions averaged across
175 top-down inversions. In contrast, the seasonal maximum emissions during September–

176 November in Sudd (about 0.3 Tg CH₄ month⁻¹ relative to April–June) were coarsely
177 underestimated by most process-based wetland simulations (giving only a non-significant rise
178 of 0.02 Tg CH₄ month⁻¹ during September–November relative to April–June; Supplementary
179 Fig. 10). This could be partly explained by the simulation of too small wetland extents of the
180 Sudd swamp and weak intra- and inter-annual variations, as shown by the comparison with
181 those derived from the satellite-based CYGNSS inundated areas (Supplementary Fig. 11;
182 Gerlein-Safdi et al., 2021). The weaker seasonal changes in wetland extent and associated CH₄
183 emissions were also identified previously for process-based or data-driven biogeochemical
184 models over tropical regions (Pandey et al., 2021; Parker et al., 2018, 2022), resulting in smaller
185 estimates of year-to-year emission anomalies. On the other hand, large uncertainties remain in
186 the model representation of CH₄ production, oxidation and vegetation-mediated transport
187 processes for tropical wetlands (Pangala et al., 2017; Shaw et al., 2022), where direct flux
188 measurements are sparse (Delwiche et al., 2021; Helfter et al., 2022).

189 In summary, based on both ground-based and satellite-based atmospheric methane
190 observations, we infer that surface methane emissions increased by 20.3±9.9 Tg CH₄ yr⁻¹ and
191 24.8±3.1 Tg CH₄ yr⁻¹ in 2020 and 2021 respectively, compared to 2019. The emission increases
192 were primarily driven by a few major wetland complexes and inundated areas in the tropics
193 and boreal regions as a result of elevated groundwater table. Strong, persistent and synchronous
194 emission increases were found over the Niger River basin, the Congo basin, the Sudd swamp,
195 the Ganges floodplains and Southeast Asian deltas and Hudson Bay lowlands, contributing to
196 approximately 70% and 60% of emission increases in 2020 and 2021. Our findings underscore
197 that wetland emissions dominate the interannual variability of methane sources and played key
198 roles in the exceptionally high methane growth in 2020 and 2021. However, such strong
199 emission increases from tropical inundated areas were substantially underestimated by current
200 biogeochemical models, reflecting model deficiencies in representing year-to-year dynamics
201 of tropical wetland extents and related methane emissions. This finding highlights the need for
202 integrating multiple data streams and modeling tools to better constrain tropical wetland
203 emissions and to understand their environmental sensitivities.

204 **Methods**

205 We followed the methodologies described in Peng et al. (2022) to combine both top-down and
206 bottom-up approaches for a synthesis study of recent methane growth during 2020–2021. The

207 analyses of methane emission changes were extended to 2021 on the basis of Peng et al. (2022)
208 using similar data sources and modeling tools. We further included satellite-based CH₄
209 observations and flux inversions in addition to the analyses derived from surface CH₄ networks,
210 which allows for intercomparison of the emission change patterns that are informed by different
211 datasets of atmospheric CH₄ observations.

212 **Atmospheric observations**

213 For surface CH₄ observations, in-situ continuous and flask-air CH₄ measurements from a total
214 of 121 stations for the inversion analyses were included (Supplementary Fig. 2), most of which
215 are operated and maintained by the NOAA (Lan et al., 2022) and ICOS networks. Observations
216 from other networks were obtained from the World Data Centre for Greenhouse Gases
217 (<https://gaw.kishou.go.jp>) and the Global Environmental Database (<https://db.cger.nies.go.jp>).
218 All observations are reported on or linked to the WMOX2004 calibration scale.

219 For satellite CH₄ observations, we used two retrievals of GOSAT XCH₄ provided by National
220 Institute for Environmental Studies in Japan and the University of Leicester in the UK (denoted
221 as “GSNIES” and “GSUoL” respectively). Launched by the Japan Aerospace Exploration
222 Agency (JAXA) in early 2009, GOSAT achieves a global coverage every 3 days with a swath
223 of 750 km and a ground pixel with a diameter of approximately 10.5 km at nadir. The Thermal
224 And Near-infrared Sensor for Carbon Observation – Fourier Transform Spectrometer
225 (TANSO-FTS) onboard enables the measurements of column-averaged dry-air CO₂ and CH₄
226 mole fractions by solar backscatter in the shortwave infrared (SWIR) with near-unit sensitivity
227 across the air column down to the surface (Butz et al., 2011; Kuze et al., 2009). The two
228 GOSAT XCH₄ products used here differ in their algorithms to treat the scattering-induced
229 issues in the retrieval of total column concentrations from spectral data. The GSNIES XCH₄
230 retrieval was produced using a full-physics algorithm to infer CH₄ column together with
231 physical scattering properties of the atmosphere (Yoshida et al., 2011, 2013). Alternatively, the
232 GSUoL XCH₄ retrieval employed a proxy algorithm that simultaneously retrieves CH₄ and
233 CO₂ columns using the absorption features around the wavelength of 1.6 μm to minimize the
234 scattering effect on the retrieval (Parker et al., 2011, 2015). While the two conceptually
235 different approaches have their own advantages and disadvantages (Schepers et al., 2012), the
236 proxy retrieval is less sensitive to aerosol distribution and instrumental issues than the full-
237 physics retrieval, therefore has much higher data density over geographic regions with

238 substantial aerosol loading, such as in the tropics (Supplementary Fig. 3). In this study, we used
239 the GSNIES XCH₄ retrieval version 2.95/2.96 (Inoue et al., 2016) and the GSUoL XCH₄
240 retrieval version 9.0 (Parker et al., 2020), which are bias-corrected and in good agreement with
241 ground-based XCH₄ measurements from the Total Column Carbon Observing Network
242 (TCCON) and aircraft-based CH₄ profile measurements. These two retrievals have been widely
243 used in global or regional methane inverse modeling to study recent trends and interannual
244 variabilities (Feng et al., 2023; Qu et al., 2022; Wang et al., 2021; Wilson et al., 2021; Yin et
245 al., 2020). Note that only retrievals over land were assimilated in our inversions in order to
246 avoid potential retrieval biases between nadir and glint viewing modes.

247 To calculate the atmospheric CH₄ growth rate, for surface observations, we used zonally
248 averaged marine boundary layer (MBL) references for CH₄ constructed by NOAA's Global
249 Monitoring Laboratory (NOAA/GML) using measurements of weekly air samples from a
250 subset of sites in the NOAA Cooperative Global Air Sampling Network (Dlugokencky et al.,
251 2021). Only sites that measure background atmospheric compositions are considered, typically
252 at remote marine sea level locations with prevailing onshore winds. For GOSAT XCH₄
253 observations, we used daily means of all valid land retrievals per 10° latitude band for
254 subsequent growth rate calculation. The smoothed CH₄ growth rate for each latitude band
255 shown in Fig. 1 and Supplementary Fig. 1 was extracted from time series of the zonally-
256 averaged MBL CH₄ references or GOSAT XCH₄ observations over the period 2010–2021
257 following the curve fitting procedures of Thoning et al. (1989).

258 **Atmospheric 3D inversion**

259 We used a variational Bayesian inversion system PYVAR-LMDZ-SACS to optimize weekly
260 CH₄ surface fluxes at a spatial resolution of 1.9° in latitude by 3.75° in longitude over the
261 period 2019–2021. An ensemble of six inversions was performed (Supplementary Table 1),
262 each using a combination of three different observation datasets described above for constraints
263 and two physical parameterizations for the transport model Laboratoire de Météorologie
264 Dynamique with zooming capability (LMDZ), the atmospheric component of the coupled IPSL
265 climate model participating in IPCC Assessment Reports (AR). These setups allowed us to
266 explore consistency of the emission change patterns informed by different observation datasets
267 while accounting for some of the uncertainties in atmospheric transport. The two physical
268 parameterizations, denoted here as the “classic” and “standard” versions, represent two

269 development stages of LMDZ for IPCC AR3 and AR6 (Hourdin et al., 2006, 2013, 2020). The
270 “classic” AR3 version uses the vertical diffusion scheme of Louis (1979) to represent the
271 turbulent transport in the boundary layer and the scheme of Tiedtke (1989) to parameterize
272 deep convection (Hourdin et al., 2006). The “standard” AR6 version combines the vertical
273 diffusion scheme of Mellor & Yamada (1974) and the thermal plume model by Rio & Hourdin
274 (2008) to simulate the atmospheric mixing in the boundary layer, and the deep convection is
275 represented using the scheme of Emanuel (1991) coupled with the parameterization of cold
276 pools developed by Grandpeix et al. (2010) (Hourdin et al., 2013, 2020). While the “standard”
277 version showed overall improved representation of boundary layer mixing and large-scale
278 atmospheric transport (Locatelli et al., 2015a; Remaud et al., 2018, 2023), which would benefit
279 trace gas transport simulations and inversions despite its comparatively larger computational
280 costs, the “classic” version or its physical parameterization schemes are still widely used in the
281 scientific community for methane studies (see Supplementary Table S6 in Saunois et al., 2020).
282 A previous study by Locatelli et al. (2015b) showed that changing physical parameterizations
283 would have small impact on the inverted methane emissions at the global scale (around 1%),
284 but could lead to significant differences in the north–south gradient of emissions and the
285 emission partitioning between regions.

286 Depending on the observations assimilated and the physical parameterizations used in the
287 inversion system, discrepancies in the derived emission changes do exist among inversions at
288 global or regional scales. The emission growth inferred from surface observations was much
289 lower than those from GOSAT-based inversions for 2020 (Supplementary Fig. 5a), possibly
290 because surface networks have limited spatial coverage over certain key source regions (e.g.,
291 the tropics), thus being blind to the methane growth there (Supplementary Fig. 2). Among the
292 four GOSAT-based inversions, the ones constrained by GSUoL retrievals always gave 15–30%
293 higher global net emission increases than those constrained by GSNIES retrievals
294 (Supplementary Fig. 5a, b), consistent with the steeper rise of the methane growth rate seen
295 from GSUoL (Fig. 1; Supplementary Fig. 1). For several important emitting regions such as
296 eastern China, northern India and southern Africa, the directions of emission changes disagreed
297 among inversions assimilating different observations (Supplementary Figs. 6 & 7), reflecting
298 uncertainties in flux solution related to sparse data density or GOSAT XCH₄ retrieval
299 algorithms.

300 Other configurations of the inversions followed the descriptions in Peng et al., (2022). The
301 prior CH₄ fluxes were built on bottom-up inventories or process-based land surface models for
302 different categories (Supplementary Table 5). The OH and O(¹D) fields were prescribed from
303 the simulation of a chemistry-climate model LMDZ-INCA with a full tropospheric
304 photochemistry scheme (Hauglustaine et al., 2004, 2014). The model was run at the resolution
305 of 1.27° in latitude by 2.5° in longitude, driven by interannually varying horizontal winds from
306 ECMWF ERA5 reanalysis (Hersbach et al., 2018) and with natural and anthropogenic
307 emissions of NO_x, CO and hydrocarbons updated to 2021 (Carbon Monitor; Community
308 Emissions Data System (CEDS); van der Werf et al., 2017). The resulting oxidant fields were
309 not adjusted in the inversions in order to keep the simulated OH changes from LMDZ-INCA.
310 The preprocessing of surface CH₄ observations and the assignment of observation uncertainty
311 were based on the protocol described in Peng et al. (2022). For GOSAT XCH₄ observations,
312 the valid data were averaged into model grids for each time step (30 mins) to create “super-
313 observations”, with the observation errors defined as the retrieval errors reported by the data
314 product plus model errors whose standard deviations were empirically set as 1% (Cressot et al.,
315 2014; Yin et al., 2021).

316 **Bottom-up estimates of methane emissions**

317 Anthropogenic methane emissions were compiled from a combination of existing inventories.
318 For the 42 Annex-I countries that report their national greenhouse gas inventories (NGHGs)
319 to UNFCCC each year, we used the reported anthropogenic methane emissions updated to 2021
320 from coal mining, oil and gas production, agriculture and waste sectors, respectively
321 (<https://unfccc.int/ghg-inventories-annex-i-parties/2023#fn2>). For China, anthropogenic
322 methane emissions were computed and updated to 2021 based on the activity data collected
323 from national statistic books (National Bureau of Statistics of China, 2022) and specific
324 emission factors at provincial levels (Liu et al., 2021; Peng et al., 2016). For other countries,
325 emissions from coal mining, oil and gas production, agriculture and waste were obtained from
326 the Emissions Database for Global Atmospheric Research version 7.0 (EDGAR v7.0; Crippa
327 et al., 2021, 2022), with coal production and livestock data corrected by the activity data from
328 International Energy Agency (IEA) and Food and Agriculture Organization of the United
329 Nations (FAO). The national total emissions from the 42 Annex-I countries, China and other
330 countries were distributed on 0.1° × 0.1° grid cells based on the spatial patterns of EDGAR
331 v7.0. Note that the change in global anthropogenic methane emissions between 2020 and 2019

332 slightly differs from that reported by Peng et al. (2022), as different versions of EDGAR and
333 IEA data were used in this study to estimate emissions.

334 The global fire methane emissions were obtained from the Global Fire Emissions Database
335 version 4.1 including small fire burned area (Randerson et al., 2012; van der Werf et al., 2017).
336 The data set produces monthly gridded burned area and fire emissions at a spatial resolution of
337 $0.25^\circ \times 0.25^\circ$, based on satellite information on fire activity and vegetation productivity (van
338 der Werf et al., 2017).

339 For wetland emissions, we used two process-based wetland emission models (WEMs),
340 ORCHIDEE-MICT (Guimberteau et al., 2018) and LPJ-wsl (Zhang et al., 2017), to simulate
341 the global wetland CH₄ emissions. Based on the simulation protocol in Peng et al. (2022),
342 wetland methane emissions were updated to 2021 using these two WEMs with four climate
343 forcing datasets. The spatiotemporal dynamics of wetland areas were simulated by a
344 TOPMODEL-based diagnostic model and applied to ORCHIDEE-MICT (Xi et al., 2021, 2022)
345 and LPJ-wsl (Zhang et al., 2016), respectively. For ORCHIDEE-MICT in particular, we
346 utilized two wetland maps to calibrate the parameters in simulating the wetland area dynamics
347 (Xi et al., 2022). The static map of Regularly Flooded Wetlands (RFW; Tootchi et al., 2019)
348 was applied for the grid-based calibration of the long-term maximum wetland extent, whereas
349 the Global Inundation Estimate from Multiple Satellites version 2 (GIEMS-2; Prigent et al.,
350 2020b) was applied to calibrate the yearly maximum wetland extent for each grid. Together,
351 eight ORCHIDEE-MICT simulations and four LPJ-wsl simulations of wetland CH₄ emissions
352 were included in our analyses.

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629

630 **Data availability**

631 The datasets that support the findings of this study are publicly available as follows. The global
632 atmospheric methane growth rates and marine boundary layer references are obtained from
633 https://gml.noaa.gov/ccgg/trends_ch4 and <https://gml.noaa.gov/ccgg/mbl/> respectively. The
634 assimilated surface CH₄ observations from NOAA and ICOS networks are available at
635 <https://doi.org/10.15138/VNCZ-M766> and <https://doi.org/10.18160/KCYX-HA35>; surface
636 observations from other networks are available from World Data Centre for Greenhouse Gases
637 (<https://gaw.kishou.go.jp/>) and Global Environmental Database (<https://db.cger.nies.go.jp/>).
638 The GOSAT NIES full physics XCH₄ retrievals are available at <https://data2.gosat.nies.go.jp/>
639 through registration; the GOSAT University of Leicester proxy XCH₄ retrievals are available
640 at <https://catalogue.ceda.ac.uk/uuid/18ef8247f52a4cb6a14013f8235cc1eb>. The EDGAR v7.0
641 time series of country-level emissions and sector-specific gridmaps are downloaded from
642 https://edgar.jrc.ec.europa.eu/dataset_ghg70. The hourly ERA5 reanalysis data are obtained
643 from <https://www.ecmwf.int/en/forecasts/dataset/ecmwf-reanalysis-v5>. The monthly dataset
644 of temperature and precipitation from CRU TS v4.06 are obtained from
645 https://crudata.uea.ac.uk/cru/data/hrg/cru_ts_4.06/. The monthly precipitation data from
646 MERRA2 are obtained from <https://gmao.gsfc.nasa.gov/reanalysis/MERRA-2/>. The monthly
647 precipitation from MSWEP v2.8 are obtained from <http://www.gloh2o.org/mswep/>. The
648 dataset of monthly global water storage/height anomalies from GRACE-FO is available at
649 <https://doi.org/10.5067/TEMSC-3JC63>. The Regular Flooded Wetlands maps are available at
650 <https://doi.pangaea.de/10.1594/PANGAEA.892657>. The monthly fire emissions from Global
651 Fire Emissions Database version 4.1, which includes small fire burned area, are obtained from
652 <https://www.geo.vu.nl/~gwerf/GFED/GFED4/>. The anthropogenic emissions from the CEDS
653 emission inventory up to 2019 are available at <https://data.pnnl.gov/dataset/CEDS-4-21-21>.
654 The gridded near-real time fossil fuel combustion data that include confinement-induced
655 reductions in 2020 and rebound in 2021 are obtained from <https://carbonmonitor.org/>.

656 **Code availability**

657 The codes and documentation for the process-based wetland models ORCHIDEE-MICT
658 (v8.4.4) and LPJ-wsl are publicly available at <http://forge.ipsl.jussieu.fr/orchidee/> and
659 https://github.com/benpoulter/LPJ-wsl_v2.0.git, respectively. The global chemistry transport
660 model LMDZ-INCA is part of the coupled IPSL climate model, with its codes and
661 documentation available at <https://cmc.ipsl.fr/ipsl-climate-models/ipsl-cm6/>.

662 **Acknowledgements**

663 The study was funded by the ESA CCI RECCAP2 project (ESRIN/4000123002/18/I-NB). P.C.
664 also acknowledges the support from the ANR CLAND convergence institute. R.J.P. and H.B.
665 were funded via the UK National Centre for Earth Observation (grant nos. NE/R016518/1 and
666 NE/N018079/1), as well as from the ESA GHG-CCI and Copernicus C3S projects (grant no.
667 C3S2_312a_Lot2). We are very grateful to the scientists and technicians who are running
668 surface networks of CH₄ observations and make these datasets available. We thank the
669 Japanese Aerospace Exploration Agency, National Institute for Environmental Studies and the
670 Ministry of Environment for the GOSAT data and their continuous support as part of the Joint
671 Research Agreement. This research is supported through HPC resources of TGCC under the
672 allocation A0110102201 made by GENCI for LMDZ-INCA simulations and computing mass
673 fluxes, the Peking University supercomputing facility for ORCHIDEE-MICT simulations, and
674 the ALICE high-performance computing facility at the University of Leicester for the GOSAT
675 retrievals and analysis. We also acknowledge the technical support from the IT team of LSCE.

676 **Author contributions**

677 X. Lin, S.P. and P.C. designed the study. X. Lin performed atmospheric 3D inversions, with
678 precomputed mass fluxes prepared by M. Remaud. G.L. and S.P. built the bottom-up
679 anthropogenic emissions inventory. S.P. and Y.X. performed ORCHIDEE simulations; Z.Z.
680 and B.P. performed LPJ-wsl simulations. D.H. and B.Z. performed LMDZ-INCA simulations.
681 M. Ramonet and X. Lan provided surface CH₄ observations; Y. Yoshida provided GOSAT
682 NIES full physics XCH₄ retrievals; R.J.P. and H.B. provided GOSAT University of Leicester
683 proxy XCH₄ retrievals. C.G.-S. and T.P. provided CYGNSS inundation maps. X. Lin
684 coordinated the research and conducted the analyses; S.H. and D.B. help preprocess the data.
685 X. Lin drafted the first manuscript; S.P., P.C., X. Lan, Y. Yin, Z.Z., H.B., P.B., F.C., C.G.-S.,
686 R.J.P, B.P., M. Remaud, A.R., M.S., R.L.T., Y. Yoshida and B.Z. contributed to writing and
687 commenting on the draft manuscript.

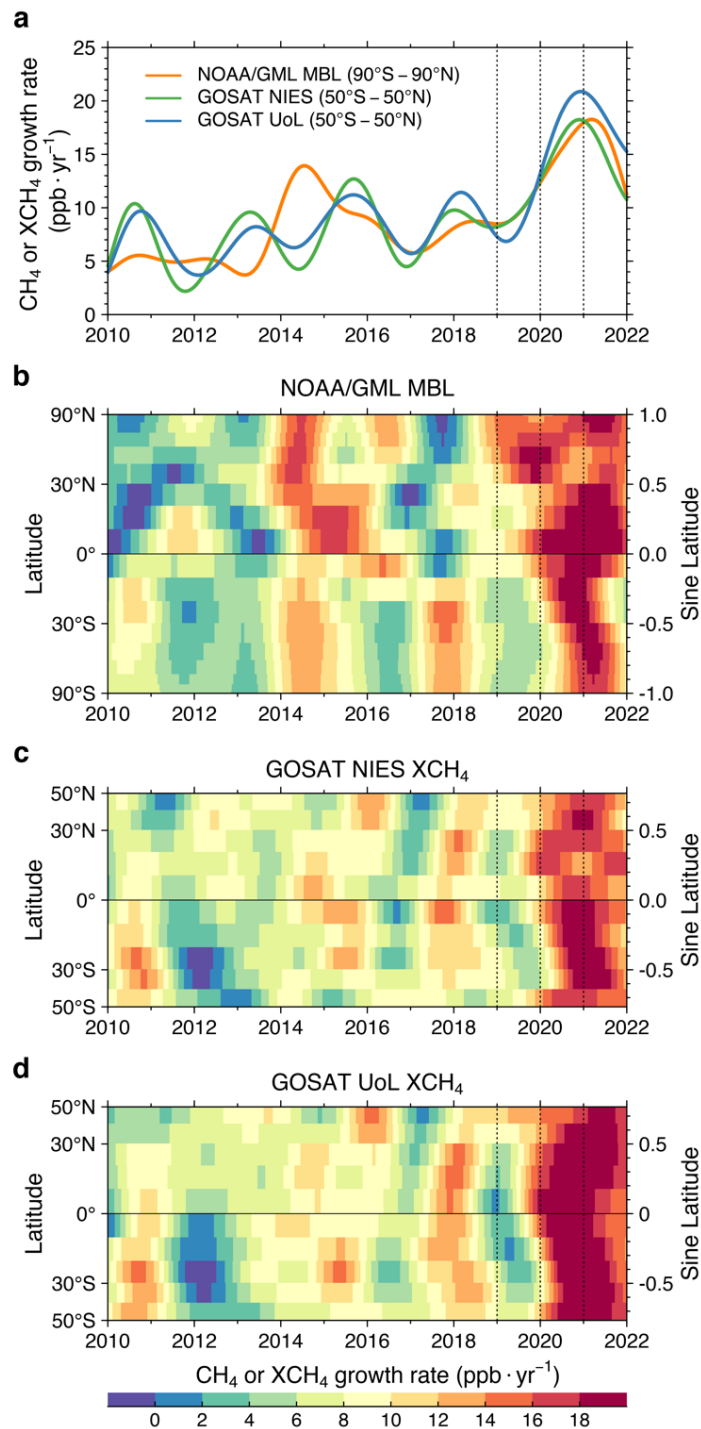
688 **Competing interests**

689 The authors declare no competing interests.

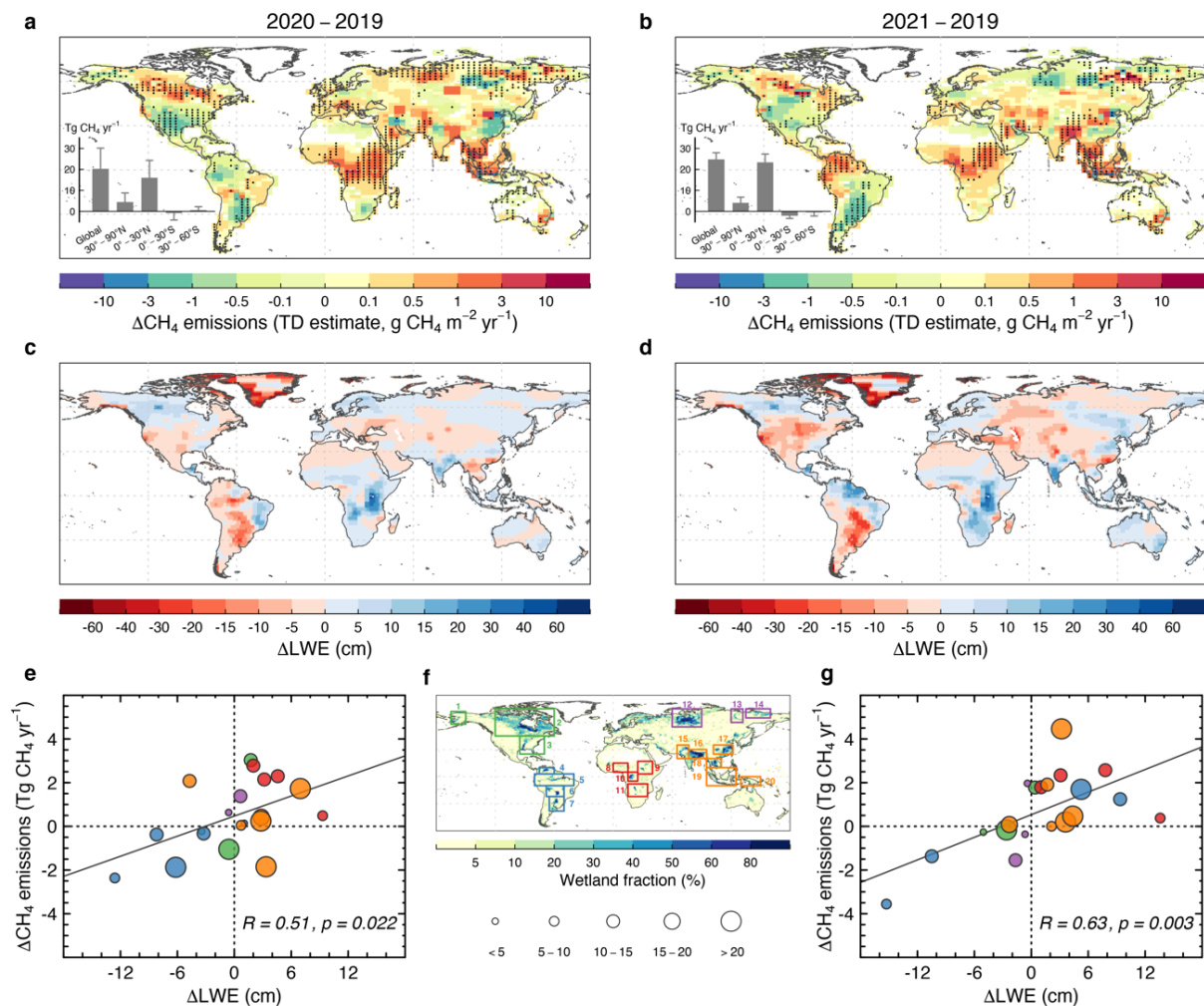
690 **Additional information**

691 Correspondence should be addressed to Xin Lin (xin.lin@lsce.ipsl.fr).

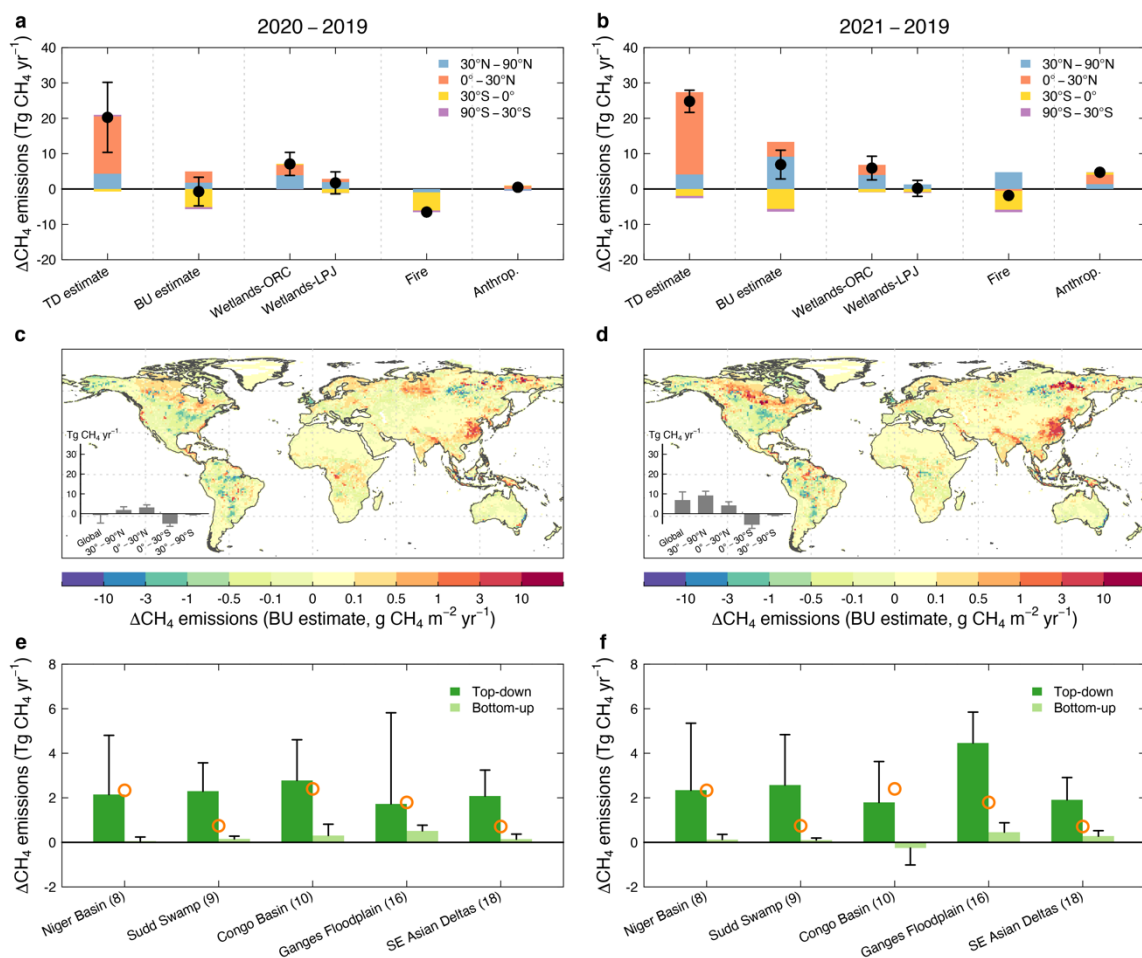
692 **Figure 1. Variations of atmospheric methane growth rates between 2010 and 2021. a–d,**
 693 Methane growth rates were derived from zonally averaged observations of NOAA Global
 694 Monitor Laboratory (NOAA/GML) marine boundary layer (MBL) sites (Lan et al., 2023),
 695 GOSAT NIES retrievals (Inoue et al., 2016) or GOSAT UoL retrievals (Parker et al., 2020),
 696 following the curve-fitting routines of Thoning et al. (1989). For **c, d**, results are not shown
 697 north of 50°N or south of 50°S due to data gaps of GOSAT retrievals over these regions during
 698 winter.



700 **Figure 2. Methane emission anomalies in 2020 and 2021 relative to 2019 inferred from six**
 701 **inversions in relation to changes in satellite observed liquid water equivalent (LWE). a,**
 702 **b,** Spatial patterns of the mean CH₄ emission anomalies averaged over the six inversions of the
 703 ensemble. The shaded areas indicate that posterior fluxes from all six inversions have the same
 704 changing direction. The inset bar plots summarize the net emission changes at the global scale
 705 and for four latitude bands. Error bars represent one standard deviation of the six inversions. **c,**
 706 **d,** Spatial patterns of changes in LWE from the GRACE-FO satellites. **e, g,** Scatterplot of CH₄
 707 emission anomalies versus changes in GRACE-FO LWE across 20 major wetland regions
 708 defined based on the regularly flooded wetland map (Tootchi et al., 2019) in **f**. Each circle
 709 represents a major wetland region, with the size scaling with the magnitude of the region's
 710 posterior methane emission in 2019 (unit: Tg CH₄ yr⁻¹). The color of the circle corresponds to
 711 the color code for each wetland region in **f**.



713 **Figure 3. Top-down versus bottom-up estimates of methane emission anomalies in 2020**
714 **and 2021 relative to 2019. a, b,** Mean methane emission anomalies of four latitude bands
715 derived from the ensemble of six top-down inversions and bottom-up estimates. The black dots
716 represent the net global emission changes relative to 2019. **c, d,** Spatial patterns of bottom-up
717 CH₄ emission anomalies summed up from process-based wetland models, inventories of
718 anthropogenic and fire emissions. The color scale is the same as for the top-down CH₄ emission
719 anomalies in Fig. 2a, b. The inset bar plots summarize the net emission changes at the global
720 scale and for four latitude bands. **e, f,** Top-down versus bottom-up estimates of methane
721 emission anomalies for five tropical inundated areas. The delineation of each inundated area is
722 shown in Fig. 2f and Supplementary Fig. 5g. The open circle indicates two times the
723 interannual variability (2σ) of methane emissions during 2010–2019 derived from a previous
724 study using the inversion system PYVAR-LMDZ-SACS and GSUoL as constraints for CH₄
725 (Zheng et al., 2019). Error bars in all panels denote one standard deviation of the methane
726 emission anomalies from the ensemble of top-down or bottom-up estimates.



Supplementary Information for

Recent methane surges reveal heightened emissions from tropical inundated areas

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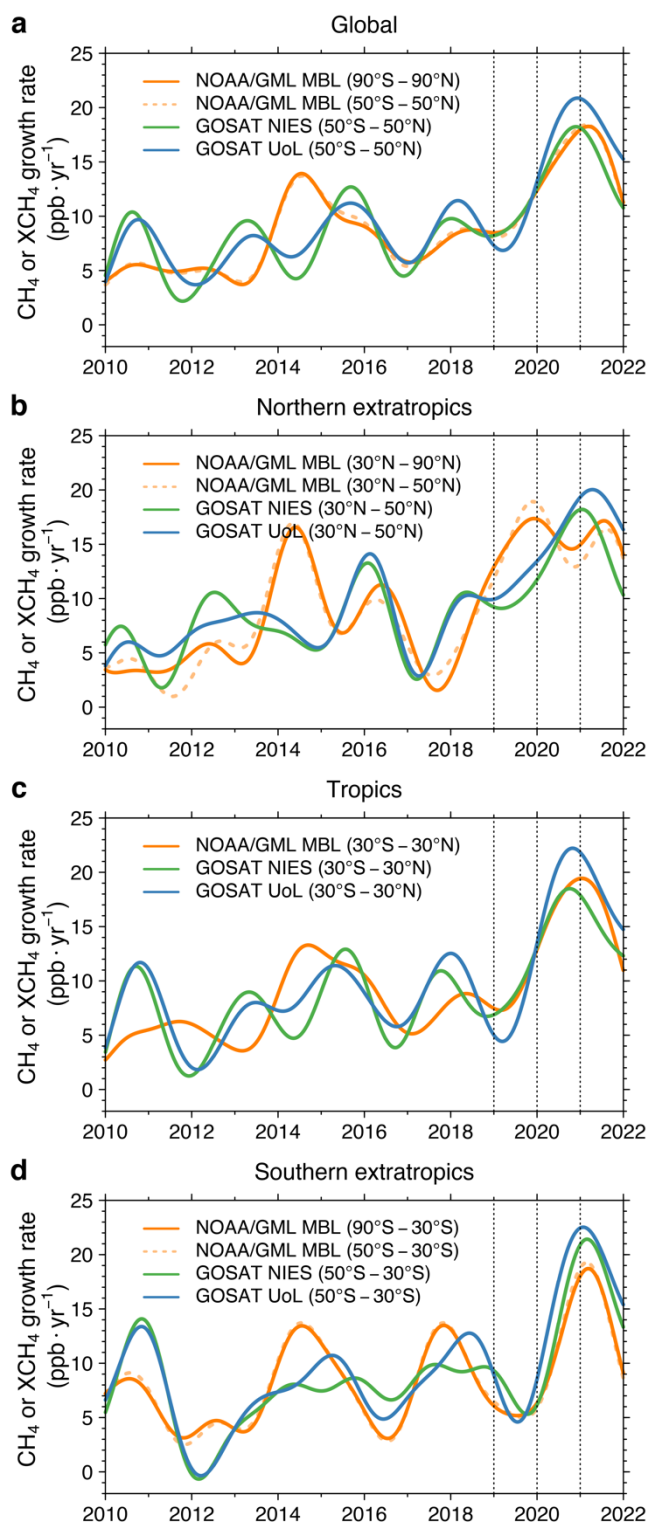
*Correspondence to Xin Lin (xin.lin@lsce.ipsl.fr)

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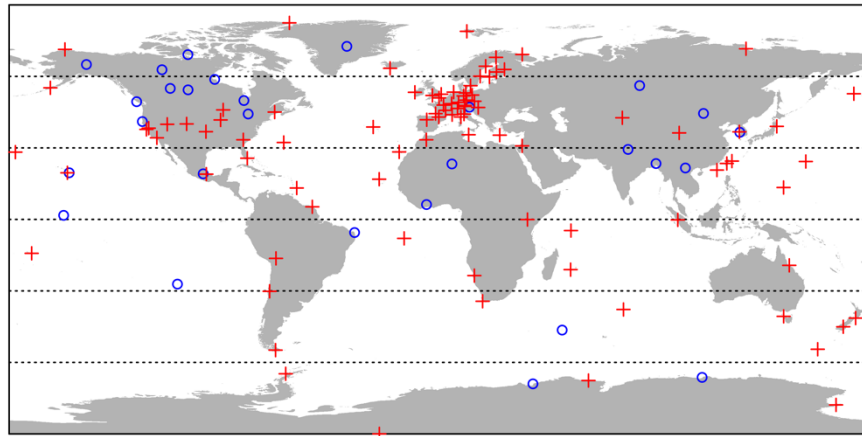
Supplementary Figures 1 to 11

Supplementary Tables 1 to 5

Supplementary Figure 1 | Variations of atmospheric methane growth rates between 2010 and 2021 at the global scale and for three latitude bands. **a–d**, Methane growth rates were derived from zonally averaged observations of NOAA Global Monitor Laboratory (NOAA/GML) marine boundary layer (MBL) sites (Lan et al., 2023), GOSAT NIES retrievals (Inoue et al., 2016) or GOSAT UoL retrievals (Parker et al., 2020), following the curve-fitting routines of Thoning et al. (1989).

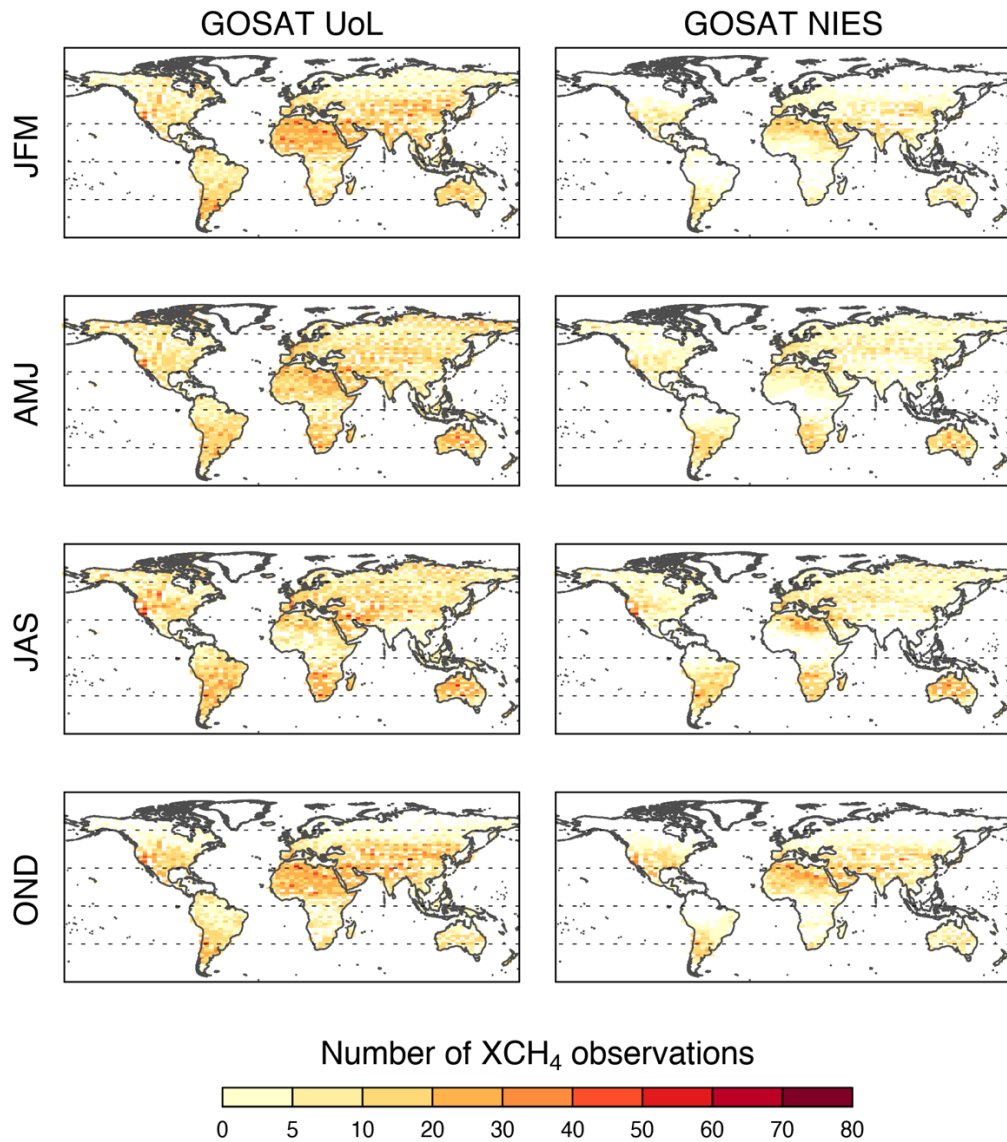


Supplementary Figure 2 | Map of surface stations for in-situ and flask-air CH₄ samplings used in the inversions. Blue circles represent stations with observations before 12/2021, while red crosses represent stations with observations extending to 12/2021 and afterwards.

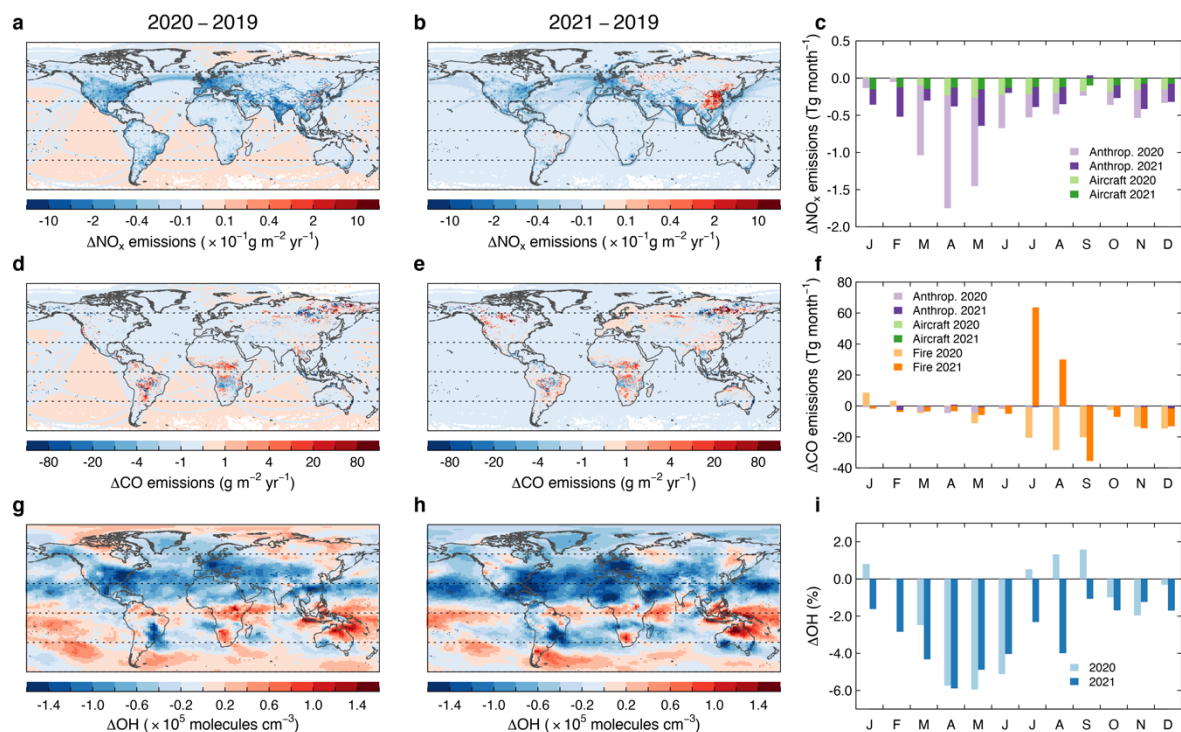


- + Obs. extended to 12/2021 and afterwards (93)
- o Obs. before 12/2021 (28)

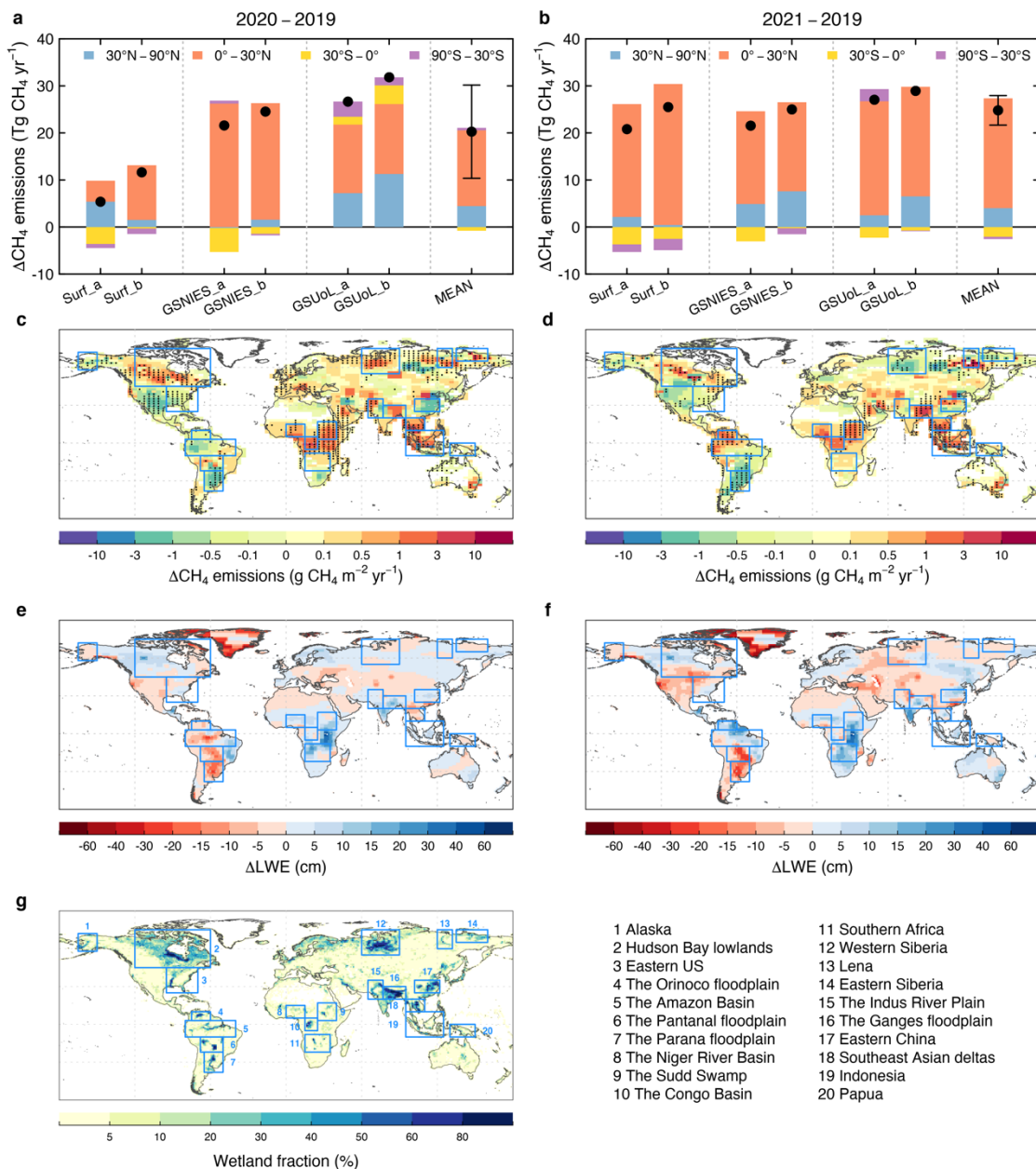
Supplementary Figure 3 | The number of XCH₄ observations for GOSAT University of Leicester (UoL) proxy retrievals and GOSAT NIES full physics retrievals. Each panel maps the number of XCH₄ “super-observations” (see Methods) on the model grids (1.9° in latitude by 3.75° in longitude) for a 3-month period of 2020.



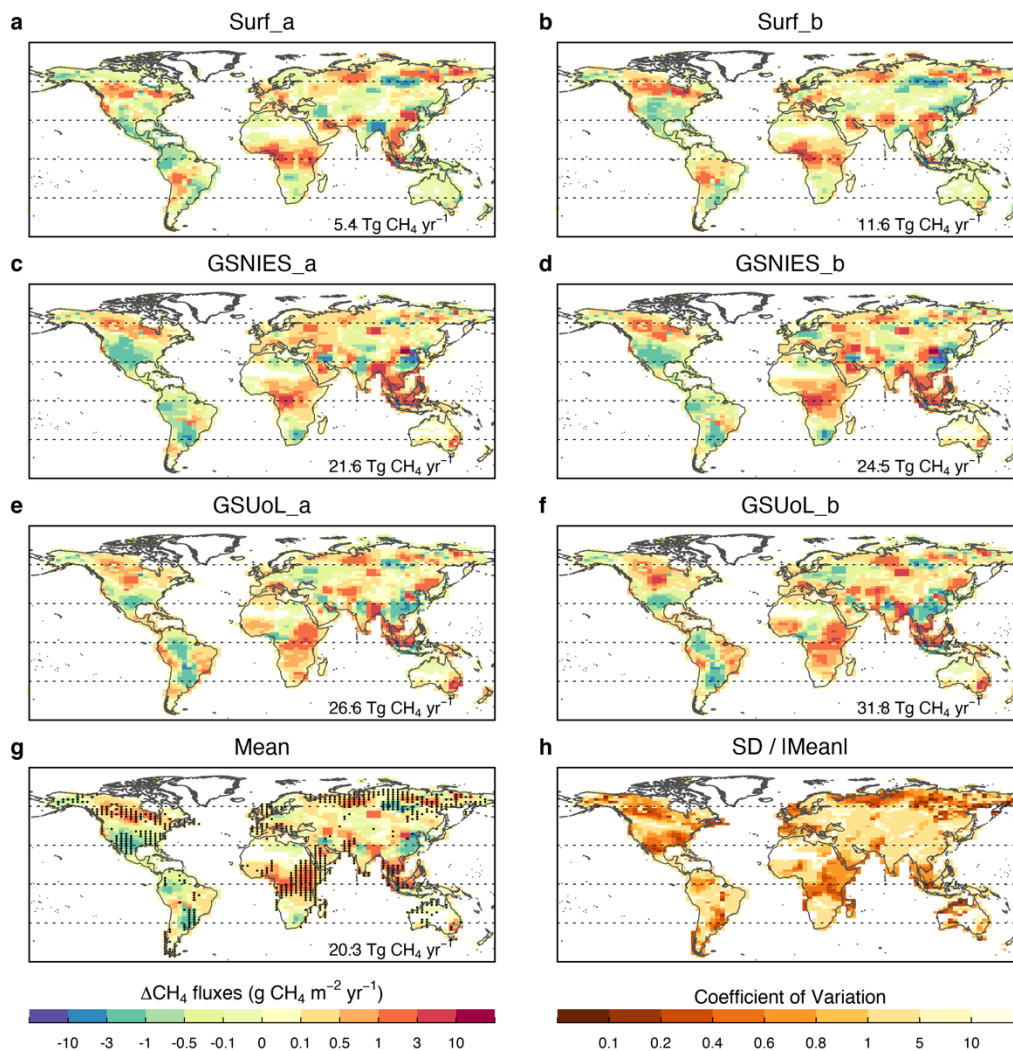
Supplementary Figure 4 | Anomalies of NO_x and CO emissions and tropospheric hydroxyl radical (OH) concentrations in 2020–2021 relative to 2019. **a-b, d-e, g-h**, Spatial patterns of emission anomalies of NO_x (ΔNO_x emissions) and CO (ΔCO emissions), and anomalies of tropospheric OH (ΔOH) in 2020–2021 relative to 2019. **c, f, i**, Changes in monthly global emissions of NO_x and CO and monthly tropospheric OH in 2020–2021 relative to 2019. The anthropogenic NO_x and CO emission data were obtained from the Community Emissions Data System (Community Emissions Data System (CEDS)) and Carbon Monitor (Carbon Monitor). The CO emissions from biomass burning were obtained from GFEDv4.1s (van der Werf et al., 2017). The OH concentration fields were simulated from a full chemistry transport model LMDZ-INCA (see the main text and Methods for details).



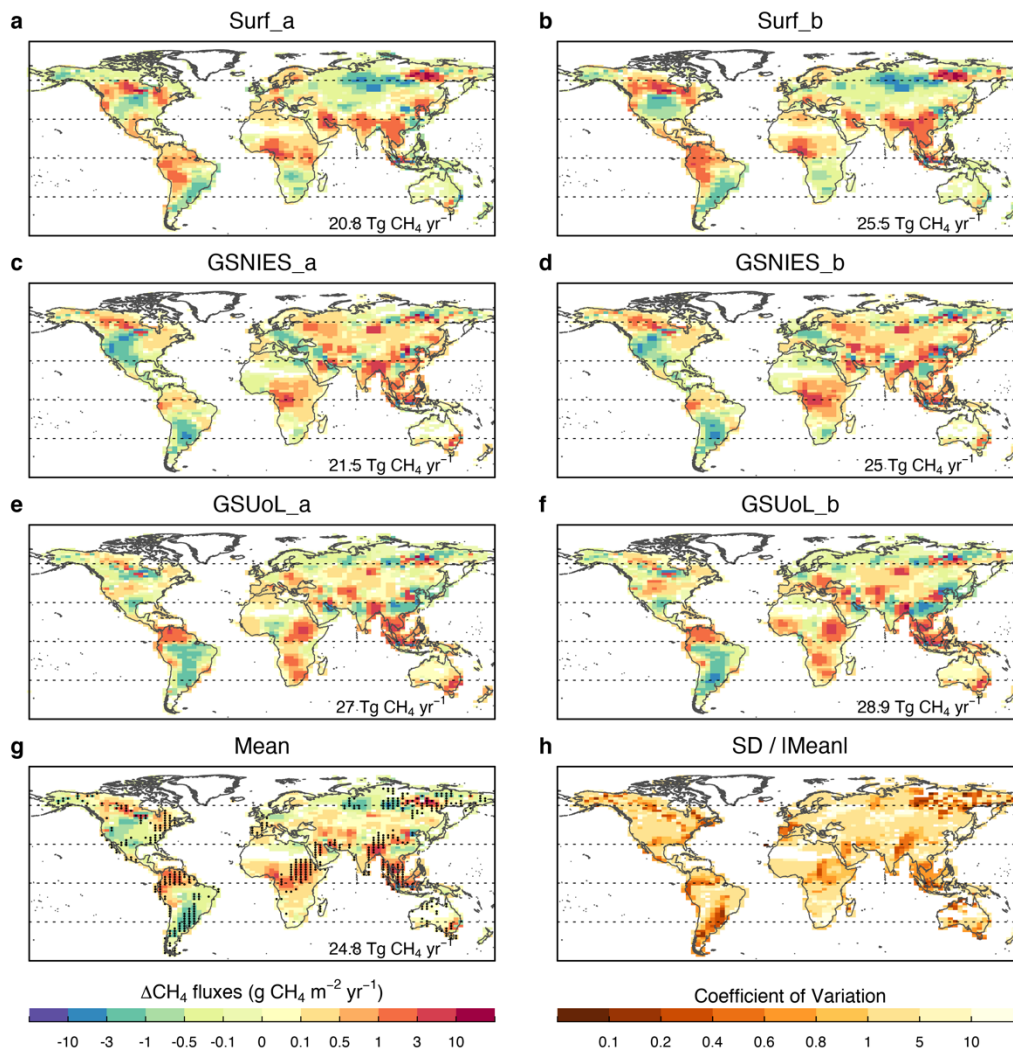
Supplementary Figure 5 | Methane emission anomalies in 2020 and 2021 relative to 2019 inferred from six inversions in relation to changes in GRACE-FO LWE. **a, b**, Methane emission anomalies of four latitude bands from six inversions. The black dots indicate the global net emission changes. The error bars represent one standard deviation. Refer to Supplementary Table 1 and Methods for descriptions and configurations of the inversion ensemble. **c, d**, Spatial patterns of CH₄ emission anomalies averaged over the six inversions. The shaded areas indicate that posterior fluxes from all six inversions have the same changing direction. **e, f**, Spatial patterns of changes in GRACE-FO LWE in 2020–2021 relative to 2019. **g**, Map of wetland fractions based on regularly flooded wetlands (Tootchi et al., 2019). Twenty major wetland regions were selected and marked in **c–f** to demonstrate emission anomalies over these regions in relation to changes in GRACE-FO LWE.



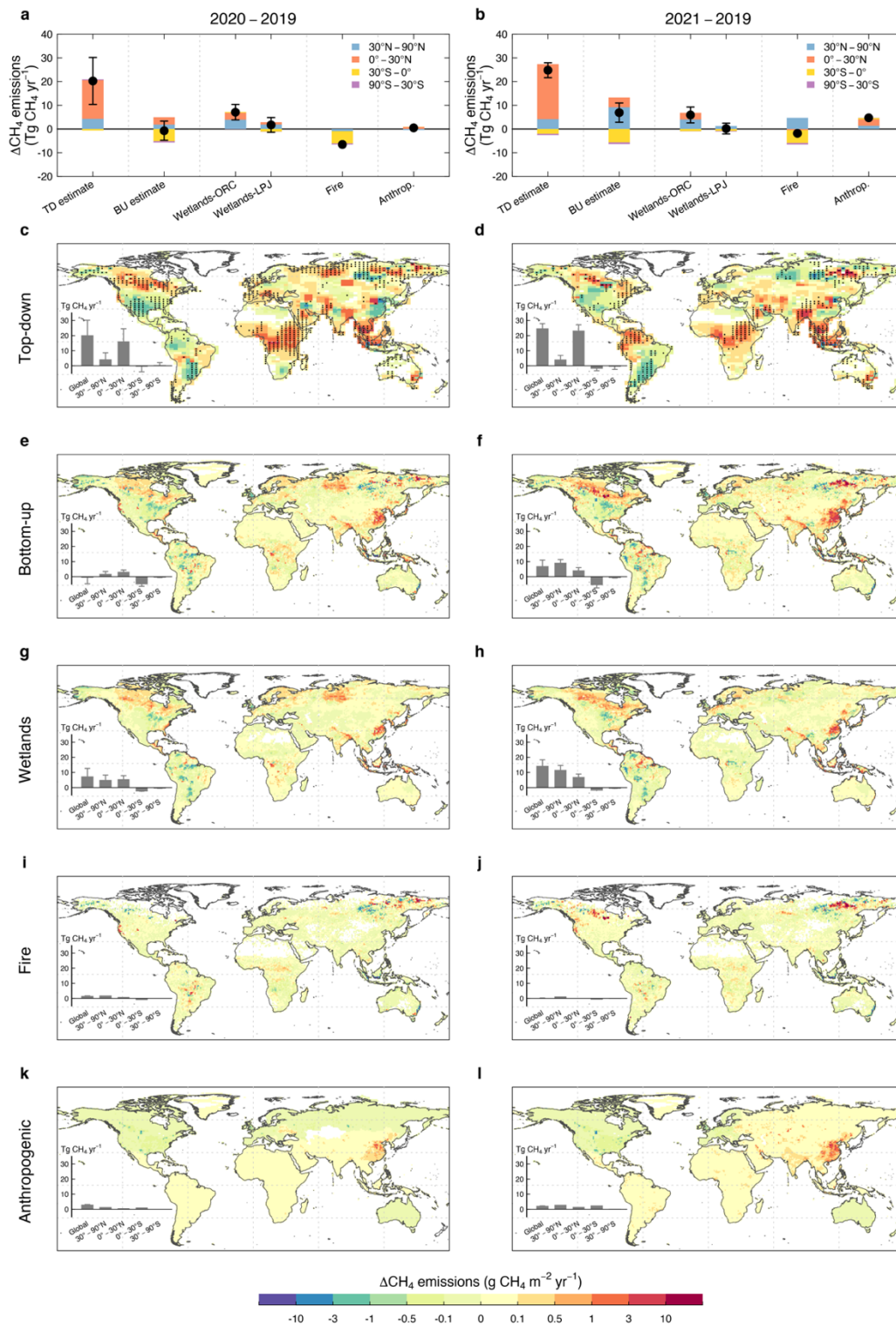
Supplementary Figure 6 | Anomalies of CH₄ emissions in 2020 relative to 2019 derived from the ensemble of six inversions. **a-f**, Spatial patterns of CH₄ emission anomalies derived from each member of the ensemble. The global net emission change (in unit Tg CH₄ yr⁻¹) is given for each panel on the bottom right. **g**, Spatial pattern of CH₄ emission anomalies averaged over the ensemble of six inversions. The shaded areas indicate that posterior fluxes from all six inversions have the same changing direction. **h**, Coefficient of variation in the CH₄ emission anomalies from the ensemble of six inversions. For each model grid, the coefficient of variation is defined as the standard deviation (SD) of emission anomalies from six inversions divided by the absolute value of their mean. Darker colors indicate better agreement among inversions. Refer to Supplementary Table 1 and Methods for detailed description and configurations of the inversion ensemble.



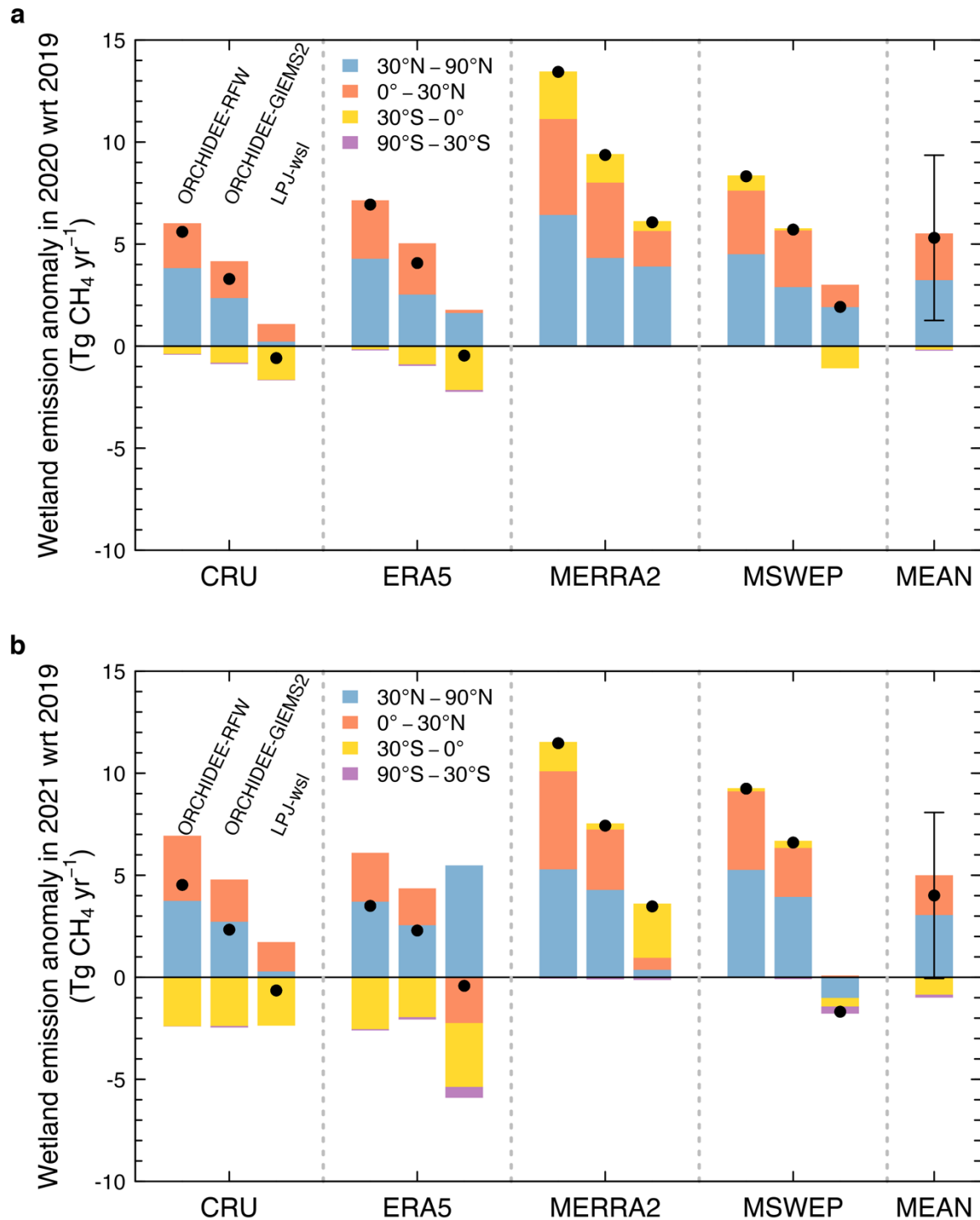
Supplementary Figure 7 | The same as Supplementary Figure 6, but for 2021.



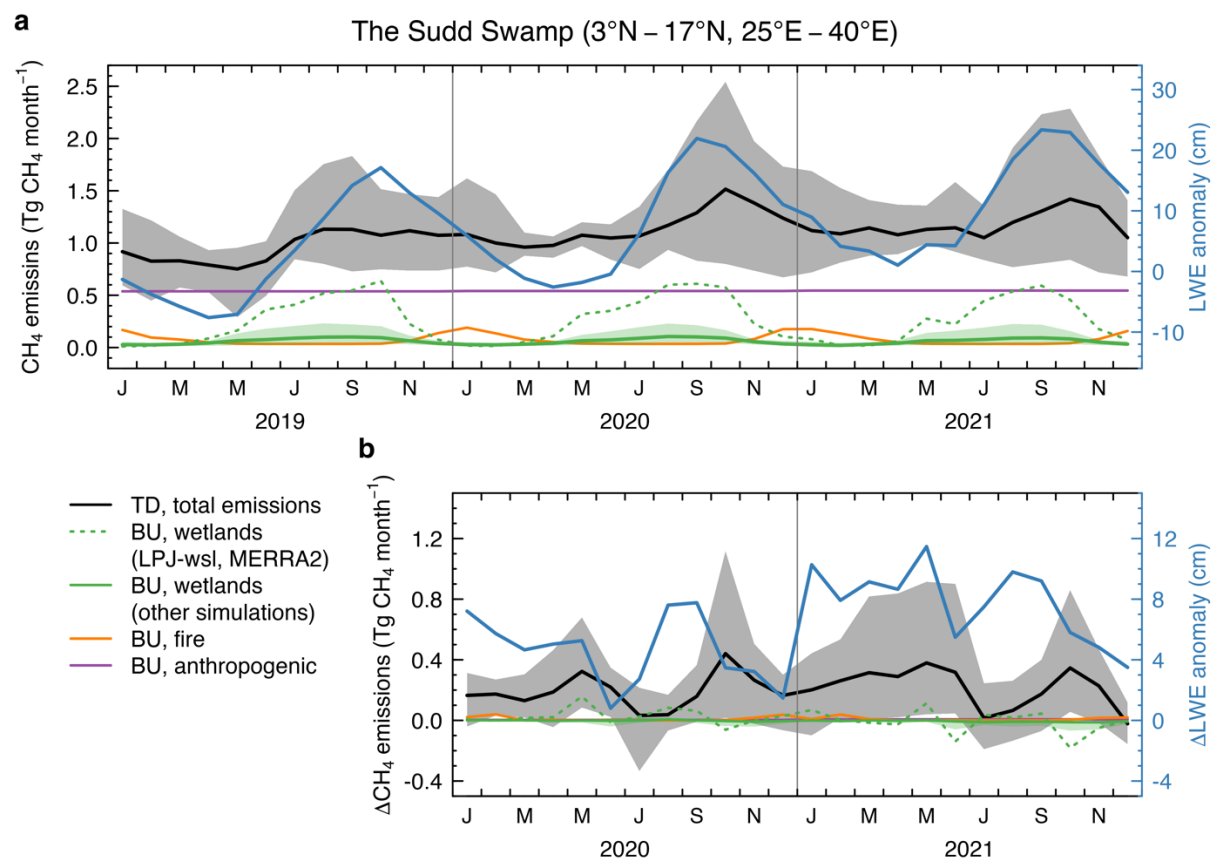
Supplementary Figure 8 | Top-down versus bottom-up estimates of methane emission anomalies in 2020 and 2021 relative to 2019. **a, b**, Mean methane emission anomalies of four latitude bands derived from the ensemble of six top-down inversions and bottom-up estimates. The black dots represent the net global emission changes relative to 2019. **c–l**, Spatial patterns of top-down or bottom-up CH₄ emission anomalies. The inset bar plots summarize the net emission changes at the global scale and for four latitude bands.



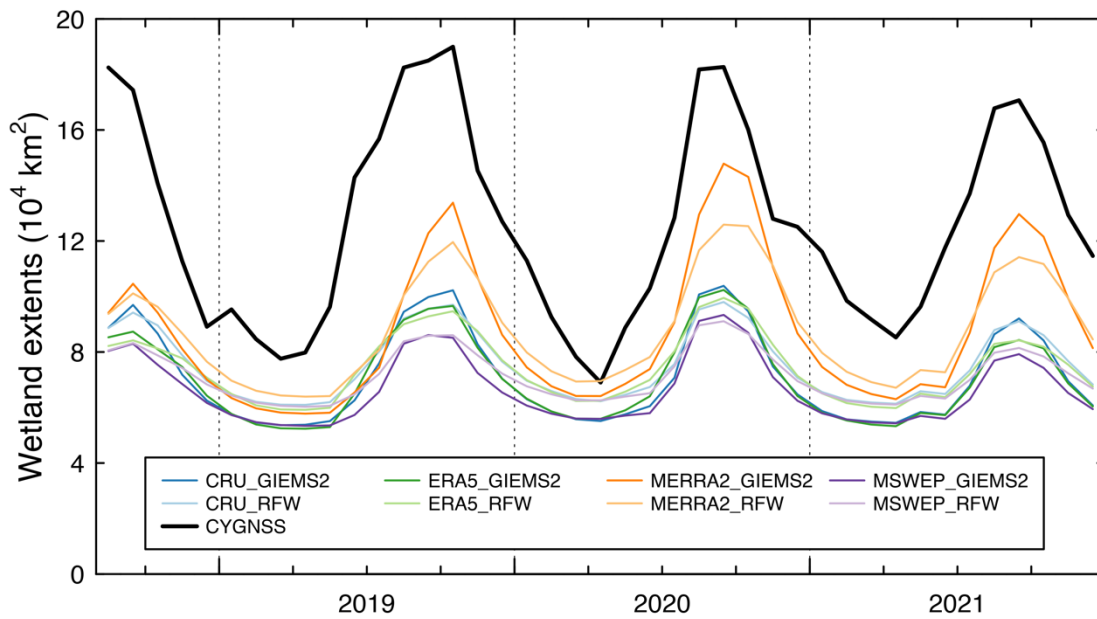
Supplementary Figure 9 | Wetland methane emission anomalies in 2020 and 2021 relative to 2019 simulated from two process-based wetland models ORCHIDEE-MICT and LPJ-wsl. **a**, **b**, Changes in wetland methane emissions were averaged for four latitude bands, based on simulations with four different climate forcing data (CRU, ERA5, MERRA2 and MSWEP). For ORCHIDEE-MICT, wetland area dynamics were calibrated by RFW and GIEMS2, respectively (noted as “ORCHIDEE-RFW” and “ORCHIDEE-GIEMS2”). The black dots represent the net changes of global wetland methane emissions.



Supplementary Figure 10 | a, The seasonal variations in CH₄ emissions over the Sudd Swamp in the eastern Africa (3°N–17°N, 25°E–40°E) in relation to the seasonal changes in GRACE-FO LWE anomalies. **b**, Changes in monthly CH₄ emissions and GRACE-FO LWE anomalies between 2020–2021 and 2019. The black curves indicate the total CH₄ emissions (or emission changes relative to 2019) averaged across the ensemble of six inversions, with the grey shaded areas representing the ranges of estimates. Note that the total CH₄ emissions plotted here were the posterior CH₄ emissions with soil sinks from Murguia-Flores et al. (2018, 2021) excluded. The solid green curves indicate the mean wetland CH₄ emissions (or emission changes relative to 2019) averaged across the ensemble of wetland model simulations except LPJ-wsl driven by MERRA2 climate forcing, with the shaded areas representing the ranges of estimates. The dotted green curves show results from LPJ-wsl with MERRA2, which simulated larger seasonal variations of emissions, but failed to capture the emission enhancements during 2020–2021 as other simulations did. Refer to Supplementary Fig. 5g for the delineation of the region.



Supplementary Figure 11 | The monthly variations of simulated versus observed wetland extents for the Sudd Swamp in the eastern Africa (3°N–17°N, 25°E–40°E). The black line represents monthly observed wetland extents obtained from the inundation maps based on the Cyclone Global Navigation Satellite System (CYGNSS; Gerlein-Safdi et al., 2021; Gerlein-Safdi & Ruf, 2019). The colored lines represent simulated wetland dynamics by ORCHIDEE-MICT with four different climate forcing data (CRU, ERA5, MERRA2, MSWEP) and two wetland maps (GIEMS-2, RFW) for calibration of parameters (Xi et al., 2022).



Supplementary Table 1 | Configurations of the ensemble of six inversions performed in this study (see Methods for more details).

| Inversions | Observations assimilated | Physical parameterizations for transport |
|------------|----------------------------------|--|
| Surf_a | Surface networks | LMDZ “classic” AR3 version |
| Surf_b | | LMDZ “standard” AR6 version |
| GSNIES_a | GOSAT XCH ₄ | LMDZ “classic” AR3 version |
| GSNIES_b | NIES full physics retrievals | LMDZ “standard” AR6 version |
| GSUoL_a | GOSAT XCH ₄ | LMDZ “classic” AR3 version |
| GSUoL_b | Univ. Leicester proxy retrievals | LMDZ “standard” AR6 version |

Supplementary Table 2 | Changes in NO_x and CO emissions and tropospheric hydroxyl radical concentrations ([OH]_{trop}) in 2020 and 2021 relative to 2019. The numbers in the parentheses represent the percentages of changes in emissions with respect to the 2019 levels.

| | 2020 – 2019 | 2021 – 2019 |
|--------------------------------------|---|---|
| NO_x | –7.5 Tg yr^{–1} (–6.2%) | –4.2 Tg yr^{–1} (–3.5%) |
| NO _x – Surface emissions | –5.7 Tg yr ^{–1} (–4.8%) | –2.8 Tg yr ^{–1} (–2.4%) |
| NO _x – Aviation emissions | –1.9 Tg yr ^{–1} (–47.9%) | –1.4 Tg yr ^{–1} (–36.0%) |
| CO | –110.8 Tg yr^{–1} (–12.1%) | +1.3 Tg yr^{–1} (+0.15%) |
| CO – Surface emissions | –19.0 Tg yr ^{–1} (–3.6%) | –4.1 Tg yr ^{–1} (–0.78%) |
| CO – Aviation emissions | –0.35 Tg yr ^{–1} (–47.9%) | –0.26 Tg yr ^{–1} (–36.0%) |
| CO – Fire emissions | –91.4 Tg yr ^{–1} (–23.7%) | +5.7 Tg yr ^{–1} (+1.5%) |
| [OH]_{trop} | –1.5% | –3.0% |

Supplementary Table 3 | The bottom-up estimates versus top-down estimates of methane emission changes in 2020 (a) and 2021 (b) compared to 2019 at the global scale and over four latitudinal bands (unit: Tg CH₄ yr⁻¹). For comparison, the interannual variability (IAV) of emissions during 2010–2019 is derived from a previous study using the inversion system PYVAR-LMDZ-SACS and GSUoL as constraints for CH₄ (Zheng et al., 2019), calculated as one standard deviation of detrended emission anomalies.

(a)

| | Global | 30°N–90°N | 0°–30°N | 30°S–0° | 90°S–30°S |
|---------------------------------|----------|-----------|----------|----------|-----------|
| BU | -0.7±4.0 | 1.8±1.6 | 3.2±1.3 | -5.1±1.3 | -0.6±0.0 |
| Anthropogenic | 0.5 | -0.5 | 0.8 | 0.1 | 0.0 |
| Fire | -6.5 | -0.9 | 0.0 | -5.1 | -0.5 |
| Wetland | 5.3±4.0 | 3.2±1.6 | 2.3±1.3 | -0.2±1.3 | 0.0±0.0 |
| TD | 20.3±9.9 | 4.3±4.3 | 16.2±8.3 | -0.7±3.3 | 0.4±1.7 |
| IAV of emissions (2010–2019) | 6.7 | 4.6 | 3.6 | 5.3 | 0.4 |

(b)

| | Global | 30°N–90°N | 0°–30°N | 30°S–0° | 90°S–30°S |
|---------------------------------|----------|-----------|----------|----------|-----------|
| BU | 6.9±4.1 | 9.2±2.1 | 4.1±1.9 | -5.6±1.9 | -0.8±0.2 |
| Anthropogenic | 4.7 | 1.4 | 2.7 | 0.6 | 0.0 |
| Fire | -1.8 | 4.7 | -0.5 | -5.4 | -0.7 |
| Wetland | 4.0±4.1 | 3.1±2.1 | 1.9±1.9 | -0.9±1.9 | -0.1±0.2 |
| TD | 24.8±3.1 | 4.1±2.7 | 23.2±4.0 | -2.0±1.3 | -0.6±1.8 |
| IAV of emissions (2010–2019) | 6.7 | 4.6 | 3.6 | 5.3 | 0.4 |

Supplementary Table 4 | The bottom-up estimates versus top-down estimates of methane emission changes in 2020 and 2021 compared to 2019 over 20 major wetland regions (unit: Tg CH₄ yr⁻¹). For comparison, the interannual variability (IAV) of emissions during 2010–2019 is derived for each region from a previous study using the inversion system PYVAR-LMDZ-SACS and GSUoL as constraints for CH₄ (Zheng et al., 2019), calculated as one standard deviation of detrended emission anomalies. The delineation of the 20 major wetland regions is presented in Supplementary Fig. 5g. The lines in bold correspond to the five wetland regions plotted in Fig. 3e, f.

| No. | Wetland regions | 2020–2019 | | 2021–2019 | | IAV of emissions (2010–2019) |
|-----------|-------------------------------|----------------|----------------|-----------------|----------------|------------------------------|
| | | BU | TD | BU | TD | |
| 1 | Alaska | -0.4±0.2 | -0.2±0.2 | -0.3±0.1 | -0.3±0.2 | 0.1 |
| 2 | Hudson Bay lowlands | 0.1±0.4 | 3.0±1.0 | 3.5±1.0 | 1.8±2.8 | 1.2 |
| 3 | Eastern US | -0.6±0.3 | -1.1±0.3 | -0.8±0.4 | -0.2±0.2 | 0.1 |
| 4 | The Orinoco floodplain | -0.1±0.7 | -0.3±0.3 | 0.0±1.2 | 1.2±1.3 | 0.4 |
| 5 | The Amazon Basin | -1.5±0.6 | -1.9±1.6 | -0.6±0.8 | 1.7±2.7 | 1.6 |
| 6 | The Pantanal floodplain | -0.1±0.4 | -0.4±1.3 | -0.8±0.7 | -1.4±1.7 | 0.9 |
| 7 | The Parana floodplain | -0.5±0.5 | -2.4±1.3 | -0.6±0.4 | -3.6±1.4 | 0.5 |
| 8 | The Niger River Basin | 0.1±0.2 | 2.1±2.7 | 0.1±0.2 | 2.3±3.0 | 1.2 |
| 9 | The Sudd Swamp | 0.2±0.1 | 2.3±1.3 | 0.1±0.1 | 2.6±2.3 | 0.4 |
| 10 | The Congo Basin | 0.3±0.5 | 2.8±1.8 | -0.3±0.8 | 1.8±1.8 | 1.2 |
| 11 | Southern Africa | -0.1±0.2 | 0.5±1.3 | 0.0±0.1 | 0.4±1.7 | 0.3 |
| 12 | Western Siberia | 1.0±0.5 | 1.4±0.6 | 0.0±0.2 | -1.6±2.7 | 0.8 |
| 13 | Lena | 0.0±0.0 | 0.1±0.3 | 1.7±0.0 | 2.0±2.0 | 0.1 |
| 14 | Eastern Siberia | 0.6±0.1 | 0.6±0.4 | -0.1±0.1 | -0.4±0.2 | 0.1 |
| 15 | The Indus River Plain | 0.2±0.2 | 0.4±0.9 | 0.4±0.3 | 0.1±0.8 | 0.2 |
| 16 | The Ganges floodplain | 0.5±0.3 | 1.7±4.1 | 0.5±0.4 | 4.5±1.4 | 0.9 |
| 17 | Eastern China | 2.0±0.6 | -1.8±3.5 | 3.4±0.9 | 0.2±1.4 | 0.7 |
| 18 | Southeast Asian deltas | 0.1±0.2 | 2.1±1.2 | 0.3±0.2 | 1.9±1.0 | 0.4 |
| 19 | Indonesia | -3.4±0.3 | 0.2±2.1 | -3.4±0.5 | 0.5±2.1 | 2.3 |
| 20 | Papua | 0.5±0.4 | 0.0±0.4 | 0.6±0.6 | 0.0±0.4 | 0.2 |

Supplementary Table 5 | The prior CH₄ fluxes used in the inversions. All emission maps were regridded to 1.9° in latitude by 3.75° in longitude, the spatial resolution of the inversions.

| Categories | Data sources | Temporal resolution | Original spatial resolution | Total flux for 2020 (Tg CH ₄ yr ⁻¹) |
|-----------------|---|----------------------------|-----------------------------|--|
| Anthropogenic | EDGARv6.0 (Crippa et al., 2020). Data for the year 2019 were either scaled with FAO statistics (for the agricultural sector), or scaled with BP statistics (for coal, oil and gas production), or linearly propagated. Data for the year 2020 and 2021 were set equal to 2019. Emissions from agricultural waste burnings in EDGARv6.0 were excluded to avoid double counting with GFEDv4.1s. | Monthly, with IAV | 0.1° | 377.8 |
| Biomass burning | GEFDv4.1s (van der Werf et al., 2017). Data from 2017 onward are beta data. | Monthly, with IAV | 0.25° | 12.9 |
| Wetlands | Monthly climatological emissions averaged over 2008–2017 from 11 land surface models contributing to wetland emissions published in Saunois et al. (2020). TEM-MDM and ELM were excluded due to high negative fluxes. | Fixed, with seasonality | 1° | 151.1 |
| Termites | Climatological emissions without seasonality from S. Castaldi's simulation (Saunois et al., 2020). | Fixed, without seasonality | 1° | 9.9 |
| Ocean | Climatological emissions without seasonality from Weber et al. (2019). | Fixed, without seasonality | 1° | 11.5 |
| Geological | Climatological emissions without seasonality from Etiope et al. (2019), rescaled to a global total of 23 Tg (best value given by IPCC AR6 WG1 report). Only onshore emissions were included to avoid double counting of offshore emissions with ocean emission sources. | Fixed, without seasonality | 1° | 21.2 |
| Soil sink | Output from the MeMo model based on Murguia-Flores et al. (2018, 2021). Data for the years 2019–2021 were set equal to 2017. | Monthly, with IAV | 1° | -35.5 |
| Total | | | | 548.9 |

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