- 1 A direct evaluation of long-term global Leaf Area Index (LAI)
- 2 products using massive high-quality LAI validation samples

3 derived from Landsat archive

- 4 Junjun Zha^{1,2}, Muyi Li^{1,2}, Zaichun Zhu^{1,2,*}, Sen Cao^{1,2}, Yanan Zhang^{1,2},
- 5 Weiqing Zhao^{1,2}, Yue Chen^{1,2}
- ⁶ ¹School of Urban Planning and Design, Shenzhen Graduate School, Peking University,
- 7 Shenzhen 518055, China
- 8 ²Key Laboratory of Earth Surface System and Human—Earth Relations, Ministry of
- 9 Natural Resources of China, Shenzhen Graduate School, Peking University, Shenzhen
- 10 518055, China.
- 11 Correspondence: Zaichun Zhu (zhu.zaichun@pku.edu.cn)
- 12 Junjun Zha: 1901212937@pku.edu.cn
- 13 Muyi Li: limuyi@pku.edu.cn
- 14 Sen Cao: sencao@pku.edu.cn
- 15 Yanan Zhang: ynzhang@stu.pku.edu.cn
- 16 Weiqing Zhao: wqzhao@stu.pku.edu.cn
- 17 Yue Chen: yueccchen@pku.edu.cn
- 18
- 19 This paper is a non-peer reviewed preprint submitted to EarthArXiv. Updates would be
- 20 made in subsequent versions of the paper.

21	Abstract: The long-term global Leaf Area Index (LAI) products are critical supports
22	for characterizing the changes in land surface and its interactions with other
23	components of the Earth system under the dramatic global change. However,
24	intercomparisons between current available long-term global LAI products present
25	significant spatiotemporal inconsistencies which have been a persistent source of
26	uncertainties in global change ecology. Yet, a direct and systematic evaluation of current
27	long-term LAI products is still lacking due to the absence of appropriate LAI references,
28	especially before 2000. Here, we proposed a novel evaluation framework to directly
29	evaluate the mainstream long-term global LAI products (GIMMS LAI3g, GLASS LAI,
30	and GLOBMAP LAI) using massive high-quality LAI validation samples. The LAI
31	validation samples, derived from the Landsat archive using machine learning and
32	MODIS LAI, have a global distribution, a long temporal coverage (1982–2020), and a
33	large amount of 4.9 million. They substantially address the issue of insufficient LAI
34	reference data and can enable quantitative LAI assessments. The long-term global LAI
35	products showed reasonable quality in terms of absolute value, with GIMMS LAI3g
36	having better performance (R:0.96; MAE: 0.29 m^2m^{-2} ; RMSE: 0.49 m^2m^{-2}),
37	followed by GLASS LAI (R:0.96; MAE: 0.31 m^2m^{-2} ; RMSE: 0.51 m^2m^{-2}) and
38	GLOBMAP LAI (R:0.90; MAE: 0.52 m^2m^{-2} ; RMSE: 0.91 m^2m^{-2}). For all LAI
39	products, the data quality after 2000 was better than before 2000. Their annual
40	maximum LAI trends presented mediocre consistencies with the LAI validation
41	samples (R: 0.20-0.29) which showed a significantly larger area of greening. The
42	evaluation of ten state-of-the-art ecosystem models demonstrated varied capabilities in

	simulating global LAI trends, with the standard deviations ranging from ~ 0.01 to 0.04
44	$m^2m^{-2}a^{-1}$. Although the Multi-Model Ensemble Mean LAI agreed with satellite-
45	based LAI products, they differed with vegetation biomes especially for the tropics. The
46	Landsat LAI validation dataset produced in this study can facilitate the development of
47	long-term global LAI products and provide a quantitative reference for vegetation
48	dynamic studies.
49	KEYWORDS
50	Vegetation trend; Long-term global LAI products; LAI validation samples; Landsat
51	archive; TRENDY; Random Forests regressor
52	
53	1 INTRODUCTION
54	Leaf Area Index (LAI), defined as one-half the total green leaf area per unit
55	
22	horizontal ground surface, is a basic ecological variable to characterize the vegetation
55 56	horizontal ground surface, is a basic ecological variable to characterize the vegetation states and ecosystem functions (Myneni et al., 2002). Compared to other vegetation
55 56 57	horizontal ground surface, is a basic ecological variable to characterize the vegetation states and ecosystem functions (Myneni et al., 2002). Compared to other vegetation indices, e.g., Normalized Difference Vegetation Index (NDVI) and Enhanced
55 56 57 58	horizontal ground surface, is a basic ecological variable to characterize the vegetation states and ecosystem functions (Myneni et al., 2002). Compared to other vegetation indices, e.g., Normalized Difference Vegetation Index (NDVI) and Enhanced Vegetation Index (EVI) (Huete, 2012; Liu & Huete, 1995), LAI provides a more
55 56 57 58 59	horizontal ground surface, is a basic ecological variable to characterize the vegetation states and ecosystem functions (Myneni et al., 2002). Compared to other vegetation indices, e.g., Normalized Difference Vegetation Index (NDVI) and Enhanced Vegetation Index (EVI) (Huete, 2012; Liu & Huete, 1995), LAI provides a more specific description of the plant canopy structure and could better indicate the mass and
 55 56 57 58 59 60 	horizontal ground surface, is a basic ecological variable to characterize the vegetation states and ecosystem functions (Myneni et al., 2002). Compared to other vegetation indices, e.g., Normalized Difference Vegetation Index (NDVI) and Enhanced Vegetation Index (EVI) (Huete, 2012; Liu & Huete, 1995), LAI provides a more specific description of the plant canopy structure and could better indicate the mass and energy exchange processes between atmosphere, vegetation, and soil (Fang et al.,2019;
 55 56 57 58 59 60 61 	horizontal ground surface, is a basic ecological variable to characterize the vegetation states and ecosystem functions (Myneni et al., 2002). Compared to other vegetation indices, e.g., Normalized Difference Vegetation Index (NDVI) and Enhanced Vegetation Index (EVI) (Huete, 2012; Liu & Huete, 1995), LAI provides a more specific description of the plant canopy structure and could better indicate the mass and energy exchange processes between atmosphere, vegetation, and soil (Fang et al., 2019; Piao et al., 2013). The global climate observing system (GCOS) and the
 55 56 57 58 59 60 61 62 	horizontal ground surface, is a basic ecological variable to characterize the vegetation states and ecosystem functions (Myneni et al., 2002). Compared to other vegetation indices, e.g., Normalized Difference Vegetation Index (NDVI) and Enhanced Vegetation Index (EVI) (Huete, 2012; Liu & Huete, 1995), LAI provides a more specific description of the plant canopy structure and could better indicate the mass and energy exchange processes between atmosphere, vegetation, and soil (Fang et al., 2019; Piao et al., 2013). The global climate observing system (GCOS) and the Intergovernmental Panel on Climate Change (IPCC) thus used LAI as a critical climate

64 Long-term global LAI data has been essential to enhancing our understanding of

65	the response and feedback of vegetation under climate change and human disturbances
66	(Piao et al., 2020), from the perspective of greenness (Zhu et al., 2016), phenology (Shen
67	et al., 2022), and carbon (Forkel et al., 2016; Piao et al., 2018), water (Yuan et al., 2019),
68	and nutrition cycling (Liang et al., 2020). A major finding based on different long-term
69	global LAI products was a continuous greening trend of global vegetation since the
70	1980s. The main driver was CO ₂ fertilization globally but also varied with the region
71	(Piao et al., 2020; Zhu et al., 2016). The long-term global LAI products are also critical
72	inputs for Earth system models and other theoretical models. These models revealed
73	that terrestrial vegetation could significantly mitigate global warming through
74	biogeochemical (absorption of atmospheric CO ₂) and biogeophysical processes (e.g.,
75	transpiration cooling) (Zeng et al., 2017).

76 However, there are also widespread inconsistencies between the LAI products at both regional and global scales regarding the magnitude of vegetation trends and 77 78 interannual changes in anomalies, which have raised common concerns on the current 79 interpretation of terrestrial ecosystem changes (Jiang et al., 2017). As spatiotemporally 80 consistent LAI values can be only derived from remote sensing data, the primary 81 sources of the inconsistencies are the choice of remote sensing data and the LAI 82 inversion methods (Fang et al., 2019). Before the late 1990s, the Advanced Very High 83 Resolution Radiometer (AVHRR) onboard the National Oceanic and Atmospheric 84 Administration (NOAA) was the only data source to derive global LAI data, but it 85 underwent the effects of NOAA satellite orbital drift and AVHRR sensor degradation 86 (Mao et al., 2012, Jiang et al., 2017). After 2000, advanced satellite sensors became

87	increasingly available and LAI products such as the Moderate Resolution Imaging
88	Spectroradiometer (MODIS) LAI presented validated accuracies (Myneni et al., 2002;
89	Justice et al., 2002). Current long-term global LAI products utilized the overlapped
90	period between AVHRR and MODIS to establish LAI models and applied the model to
91	pre-2000 AVHRR data (Claverie et al., 2016; Pinzon and Tucher, 2014). They differ in
92	the AVHRR input (raw reflectance or NDVI), LAI reference (field measurements,
93	MODIS LAI or its variants), and LAI model (neural networks).
94	Recent advances in global change research appeal to addressing the inconsistencies
95	between long-term global LAI products. This can hardly be achieved by
96	intercomparison analysis between LAI products as it only provides relative differences
97	(Fang & Liang, 2005; Fang et al., 2013; Garrigues et al., 2008; Gessner et al., 2013;
98	Jiang et al., 2017; Xu et al., 2018); rather, a direct validation that quantifies the absolute
99	accuracies is preferred. The direct validation however requires high-quality LAI
100	references either from field measurements or satellite products (Baret et al., 2006;
101	Buermann et al., 2001). To date, the field LAI measurements are limited to small areas
102	or short periods. They also suffer from a spatial mismatch with satellite image pixels
103	(Fang et al., 2012). Satellite-derived LAI products of high reliability such as MODIS
104	LAI can provide globe-wide sample reference, but they became available only after the
105	year 2000 (Fan et al., 2014). A huge gap exists between the demand for direct LAI
106	validation and sufficient high-quality LAI sample reference.
107	In this context, this study aims to provide a systematic assessment of current long-

108 term global LAI products using a high-quality LAI validation dataset with massive

109	samples, a long-time span, and global coverage. The creation of the validation dataset
110	(1984-2020) takes advantage of the Landsat archive available since the 1970s and
111	employs the MODIS LAI product and a machine learning method. We evaluate the
112	quality of the validation dataset via field LAI measurements. Then, three mainstream
113	long-term global LAI products of old and new versions, namely, the third generation
114	Global Inventory Modeling and Mapping Studies LAI (GIMMS LAI3g) (Zhu et al.,
115	2013), the Global Land Surface Satellite (GLASS) LAI (Xiao et al., 2016), and the
116	Long-term Global Mapping (GLOBMAP) LAI (Liu et al., 2012) are compared to the
117	validation dataset. Absolute accuracies are presented. We also evaluate the annual
118	vegetation trends and anomalies in the LAI products and ecosystem models for different
119	vegetation biomes and periods.

120 2 MATERIALS AND METHODS

121 **2.1 Data**

122 **2.1.1 Landsat surface reflectance**

123 The Landsat surface reflectance was acquired from the Google Earth Engine (GEE) 124 (Kang et al., 2021). We employed six spectral bands (blue, green, red, Near Infrared 125 [NIR], Short-wave infrared 1 [SWIR 1], and SWIR 2) with a 30 m resolution in UTM projections from Landsat 8 Operational Land Imager (OLI), Landsat 7 Enhanced 126 127 Thematic Mapper Plus (ETM+), Landsat 4–5 Thematic Mapper (TM), and Landsat 1– 5 Multispectral Scanner (MSS) products. All product has been geometrically corrected 128 and radiometrically calibrated (Li et al., 2018). Each Landsat scene provides 129 130 information on its geographic spatial location (latitude and longitude) and solar zenith and azimuth at the time of acquisition. Most clouds and shadows have been labeled
using the Fmask algorithm (Zhu & Woodcock, 2012). We used the atmospheric opacity
(AOP) index to identify remaining thin clouds mainly in tropical regions and retained
scenes with AOP < 0.1 (clear sky).

135 **2.1.2 MODIS LAI**

136 The MODIS LAI product (MCD15A2H, Collection 6), acquired from Atmosphere Archive & Distribution System (LAADS) Distributed Active Archive Center (DAAC) 137 (https://ladsweb.modaps.eosdis.nasa.gov/search/), was generated every 8 days in 500 138 139 m spatial resolution (Huang et al., 2008). The MODIS LAI (C6) represented the new 140 version (nv) that spanned from 2000 to the present. The main algorithm of the MODIS 141 LAI applied biome-specific Look-up-Tables (LUTs) based on a three-dimension 142 radiative transfer model and the back-up algorithm that used empirical relationships between NDVI and LAI (Knyazikhin et al., 1998; Myneni et al., 2002). Compared to 143 144 the old version (ov) of MODIS LAI (C5), MCD15A2H incorporated data from Terra 145 and Aqua satellites and used the latest MODIS land cover product. It provided a quality 146 control (QC) layer and saturation information. We applied the Savitzky–Golay (SG) filter on the MODIS LAI time series (Yuan et al., 2011). 147

148 2.1.3 Long-term global LAI products of old and new versions

149The third generation Global Inventory Modeling and Mapping Studies LAI150(GIMMSLAI3g), acquiredfrom151https://drive.google.com/drive/folders/0BwL88nwumpqYaFJmR2poS0d1ZDQ?resour152cekey=0-9IRE9s-0tFGfwB5qTpLjZw&usp=sharing/, was generated every 15 days at

153	1/12° spatial resolution (Zhu et al., 2013). The algorithm of GIMMS LAI3g used the
154	feed-forward neural network model to relate GIMMS NDVI3g with MODIS LAI of
155	Beijing Normal University (BNU) version between 2001 and 2009. One neural network
156	model was generated from each month. The main difference between the new (v4) and
157	old (v2) versions of GIMMS LAI3g is that the new version employed the latest GIMMS
158	NDVI3g data.

159 Global Mapping (GLOBMAP) LAI, The Long-term acquired from https://zenodo.org/record/4700264/, provided consistent long-term global LAI values 160 (1981–2020) at 8 km resolution. The GLOBMAP LAI was a combination of AVHRR 161 162 LAI (1981-2000) (Tucker et al., 2005) and MODIS LAI (2001-2020). The MODIS LAI was derived from MODIS land surface reflectance data (MOD09A1). Pixel-wise 163 164 relationships were established between MODIS LAI and AVHRR NDVI in the overlapping periods (2000-2006) and were then applied back to AVHRR NDVI to 165 166 generate LAI between 1981 and 2020 (Deng et al., 2006; Liu et al., 2012). The improvement of the new version (v3) over the old (v2) version of GLOBMAP LAI 167 168 includes the use of updated MOD09A1 (C6), a new cloud detection algorithm for MOD09A1, and a new clumping index map for calculating MODIS LAI (Chen et al., 169 170 2020).

The Global Land Surface Satellite (GLASS) LAI, acquired from http://www.bnudatacenter.com/, was generated every 8 days in 1 km spatial resolution from 1981 to
2020. The product was based on the general regression artificial neural network, which
built relationships between MOD09A1 and LAI reference data. The LAI reference was

175	created by fusing Terra/MODIS LAI (MOD15) with clump-corrected CYCLOPES LAI
176	over Benchmark Land Multisite Analysis and Intercomparison of Products
177	(BELMANIP) sites (Xiao et al., 2014). All the global LAI products were resampled to
178	have a temporal resolution of half-month and a spatial resolution of 8 km. The main
179	difference between the new (v4) and old (v2) versions of GLASS LAI products is that
180	the new version used the latest version of AVHRR surface reflectance.

181 2.1.4 LAI from TRENDY Process-based Ecosystem Models

182 This study used ten sets of global monthly LAI data simulated by the TRENDY Process-based Ecosystem Models (https://globalcarbonbudgetdata.org/) at 0.5° spatial 183 184 resolution for 1984-2016 (Wong et al., 1979). These models take into account the 185 effects of temperature, soil moisture, atmospheric CO₂ concentration, climate change, 186 nitrogen deposition, and land cover changes. They have been widely used in the study of the carbon cycle process of the global terrestrial ecosystem (Sitch et al., 2003). The 187 188 TRENDY Process-based Ecosystem Models include simulations of multiple scenarios. 189 This study used the scenario when the models were driven by all factors of atmospheric 190 CO₂ concentration, climate, and land use.

191

2.1.5 Field LAI measurements

The field LAI measurements consist of the LAI datasets at BELMANIP network sites (Baret et al., 2006) and from the Oak Ridge National Laboratory (ORNL) (Breda et al., 2003), available at http://calvalportal.ceos.org/web/olive/site-description and http://www.ornl.gov, respectively. The BELMANIP network was a good representation of global land cover types (Baret et al., 2006). Its latest version completed the spatial

- distribution of sites according to the GLC2000 land cover classification and added 25
 sites in bare soil areas and tropical forests. The ground LAI measurements by ORNL
 covered a long period from 1932 to 2000. A total of 190 valid field LAI measurements
 from 1982–2020 were involved in this study.
- 201 2

2.1.6 MODIS Land Cover product

The MODIS Land Cover product (MCD12Q1, Collection 6), acquired from 202 203 https://lpdaac.usgs.gov/products/mcd12q1v006/, was generated based on the fusion of 204 Terra and Aqua observations from 2001 to 2019 with a spatial resolution of 500 m 205 (Friedl et al., 2010). The product includes five traditional classification systems. This 206 study selected the third classification scheme of MODIS-derived LAI which divides the global vegetation biome into eight types, including Grassland (GRA), shrubland (SHR), 207 208 Cropland (CRO), Savannas (SAV), Evergreen Broadleaf Forest (ENF), Deciduous Broadleaf Forest (DBF), Evergreen Needleleaf Forest (ENF), and Deciduous 209 210 Needleleaf Forest (DNF). This study further used GLO in data analysis to represent the 211 global vegetation biome (the ensemble of eight vegetation types).

212 **2.2 Generating global LAI validation dataset**

We selected massive training sample pairs from Landsat reflectance and MODIS LAI for different vegetation biome types. These sample pairs were then rigorously refined based on a series of criteria (Figure 1). The remaining ones were used to build biome-specific machine learning models that related Landsat surface reflectance to MODIS LAI (Zhou et al., 2018; Kang et al., 2021). The models were finally applied to the Landsat data to generate the LAI validation dataset so that long-term global LAI 219 products from 1982 to 2010s can be evaluated.

220 **2.2.1 Initializing training sample pairs**

221 Based on the LAI classification scheme in the MODIS Land Cover product, we 222 identified locations (in 500 m resolution) whose vegetation biome type remains unchanged for 19 consecutive years (2001-2019). A systematic random sampling 223 224 method was applied at the locations to select seventy thousand (70,000) samples for 225 DNF and one hundred thousand (100,000) samples for other vegetation biome types. 226 We used GEE to extract MODIS LAI (in 500 m resolution) and Landsat surface reflectance (20×20 pixels in 30 m resolution) at the sample locations, each creating 227 228 one sample pair. Based on quality information in MODIS LAI and Landsat surface 229 reflectance datasets, the sample pair was considered valid if (1) the MODIS LAI value 230 was derived from the main algorithm (rather than the back-up algorithm), (2) no sensor degradation and no clouds/cloud shadows were present in the MODIS pixel, and (3) 231 232 more 90% Landsat pixels (360) have a good quality with QC=0 and AOP smaller than 233 0.1.



235 FIGURE 1 Workflow of the methodology. (a) The generation of massive high-quality 236 LAI validation samples. (b) Accuracy validation for current LAI products of different versions. The sample pair screening process includes quality control, outlier removal, 237 and saturation misclassification removal. SR means surface reflectance. 238

239

2.2.2 Sample pair screening

240 For each sample pair, Landsat surface reflectance of six bands (blue, green, red, NIR, SWIR1, and SWIR2) were aggregated as mean (μ) and standard deviation (σ) for 241 242 good-quality pixels. The coefficient of variation (CV) was calculated as the ratio of μ 243 and σ . Sample pairs were considered homogeneous if their average CV across six bands was lower than 0.15 to ensure the purity of the samples (Kang et al., 2021). NDVI was 244 245 then calculated. The homogeneous sample pairs were screened to exclude those NDVI values less than 0 or greater than 1. We also removed the sample pairs whose NDVI fell 246 247 outside of the normal range for different MODIS LAI values (Kang et al., 2016). For this purpose, LAI values in all sample pairs were binned into $0.2 m^2 m^{-2}$ intervals. 248 Within each bin, NDVI fell outside of the 1.5 interquartile range (IQR) were identified 249

and the corresponding sample pairs were removed. Note that we used NDVI rather than
individual bands (e.g., red and NIR) because of its stronger relationship with LAI (Kang
et al., 2021). EVI1, EVI2, and Normalized Difference Water Index (NDWI) were also
calculated based on Landsat surface reflectance.

254 In the pre-experiments, we found that the saturation state (saturated or not) of the 255 MODIS LAI would significantly impact the model accuracy in LAI inversion. This 256 impact has seldom been reported in previous studies (Kang et al., 2021). To account for this impact, we introduced a saturation indicator as an explanatory variable in the LAI 257 258 inversion model (see the following section). The saturation indicator can be retrieved 259 from the MODIS LAI product. In the MODIS algorithm, an LAI pixel was classified as "saturated" if the surface reflectance fell within a predefined saturation threshold 260 261 (Knyazikhin et al., 1998). However, this threshold-based classification would fail as unsaturated pixels with lower LAI values frequently presented similar surface 262 263 reflectance as the saturated ones with higher LAI values. Misclassification of the 264 saturation state could lead to overestimation of unsaturated LAI and underestimation of 265 saturated LAI in LAI inversion.

This study removed the sample pairs whose LAI saturation states were possibly misclassified. First, we reclassified the saturation state of the MODIS LAI in sample pairs via the Random Forest classifier. The classification adopted a ten-fold crossvalidation strategy for each vegetation biome and Landsat sensor (TM, ETM+, and OLI), where nine splits were used for training to determine the saturation state of the remaining one split. The target variable was MODIS LAI, and the explanatory variables included the Landsat surface reflectance, vegetation indices (NDVI, EVI1, EVI2, and
NDWI), and solar illumination angles. Then, for each sample pair, the reclassified
saturation state was compared to that from MODIS LAI. The sample pair was removed
if the saturation states were conflicted.

276 **2.2.3 Enhancing the Random Forest model**

277 Given the differences in the radiative transfer process between biomes and the 278 discrepancies in spectral response between Landsat sensors, we built individual 279 Random Forest regression models for each vegetation biome and Landsat sensor (TM, ETM+, and OLI). The explanatory variables included Landsat surface reflectance (blue, 280 281 green, red, NIR, SWIR1, and SWIR2), NDVI, NDWI, EVI1, EVI2 (Gao, 1996), geographic coordinates (longitude and latitude) of the sample center, and solar zenith 282 283 and azimuth angles at the scene center. The vegetation indices were included because they could explain the variations of LAI from different aspects (You et al., 2017). 284 285 Geographic coordinates account for the spatial variation in LAI. The solar illumination 286 geometry can reduce the Bidirectional Reflectance Distribution Function (BRDF) effect 287 on LAI retrieval.

To avoid the issue of over-fitting, we determined the model hyperparameters (e.g., number of trees, minimum leaf population, and number of variables per split) by fivefold cross-validation. In the end, each regression model included 100 trees and 5 minimum leaves. The model performance was also evaluated by a five-fold crossvalidation strategy, with 80% data as training data and the remaining as test data. We used R-Square (R^2), Mean Absolute Error (MAE), Root Mean Squared Error (RMSE), 294 normalized RMSE (nRMSE), and bias as the error metrics. Normalized RMSE was
295 computed as the ratio of RMSE and the mean reference MODIS LAI. Bias was the
296 mean difference between prediction and reference LAI.

297 2.2.4 LAI prediction from Landsat data during 1982–2020

298 The established Random Forests regression models were applied to Landsat data 299 from 1982 to 2020 to generate the final LAI validation dataset. First, locations where 300 LAI would be predicted were determined. We produced a global land cover map with a spatial resolution of $1/12^\circ$, whose pixel type was set as the most frequent vegetation 301 302 biome based on MODIS Land Cover products between 2001 and 2019. A total of 40 303 thousand sample grids $(1/12^\circ)$ were randomly selected from the land cover map. Within 304 each sample grid, nine (3×3) locations were evenly placed. This step produced 40 305 thousand \times 9 sample locations.

306 Second, Landsat data at the sample locations were extracted and refined through a 307 series of criteria similar to those in Section 2.2.1 and 2.2.2. We extracted 20×20 308 Landsat pixels (30 m resolution) around each location from all available Landsat 309 records. The sample location with Landsat data of a certain date was considered valid if (1) more 90% of Landsat pixels (360) have a good quality with QC=0 and AOP 310 311 smaller than 0.1; (2) the average CV across six spectral bands was lower than 0.15; and 312 (3) the average NDVI was between 0 and 1. For each valid sample location, we 313 calculated its average Landsat surface reflectance, geographic coordinates, VIs, and 314 solar zenith and the azimuth of a specific date. All valid sample locations with Landsat 315 data formed the predicting samples.

316	Third, the LAI values of predicting samples were estimated. The estimation was
317	based on the established Random Forests regression model for each vegetation biome
318	and each Landsat sensor. A $1/12^{\circ}$ grid was considered valid if more than 5 of 9 sample
319	locations were valid. The predicted LAI values within each $1/12^{\circ}$ grid were averaged.
320	The final Landsat LAI validation dataset included all 1/12° sample grids with their LAI
321	values.
322	Accuracies of the Landsat LAI validation dataset were assessed by field LAI
323	measurements. We derived LAI values at the geographic locations of the BELMANIP
324	and ORNL sites based on Landsat data and established Random Forests models. The
325	derived LAI data were then compared to field-measured LAI.
326	2.3 Trend analysis for long-term global LAI products
327	2.3.1 Time-series analysis
327 328	2.3.1 Time-series analysis We used an Ensemble Empirical Mode Decomposition (EEMD) method along with
327328329	2.3.1 Time-series analysis We used an Ensemble Empirical Mode Decomposition (EEMD) method along with the classical linear model method to detect trends in the long-term global LAI products.
327328329330	 2.3.1 Time-series analysis We used an Ensemble Empirical Mode Decomposition (EEMD) method along with the classical linear model method to detect trends in the long-term global LAI products. The EEMD method decomposes the time series into a set of oscillatory components at
327328329330331	 2.3.1 Time-series analysis We used an Ensemble Empirical Mode Decomposition (EEMD) method along with the classical linear model method to detect trends in the long-term global LAI products. The EEMD method decomposes the time series into a set of oscillatory components at different frequency levels while overcomes the scale mixing problem (Huang et al.,
 327 328 329 330 331 332 	 2.3.1 Time-series analysis We used an Ensemble Empirical Mode Decomposition (EEMD) method along with the classical linear model method to detect trends in the long-term global LAI products. The EEMD method decomposes the time series into a set of oscillatory components at different frequency levels while overcomes the scale mixing problem (Huang et al., 1998). We used EEMD to decompose the long-term LAI products into four components,
 327 328 329 330 331 332 333 	 2.3.1 Time-series analysis We used an Ensemble Empirical Mode Decomposition (EEMD) method along with the classical linear model method to detect trends in the long-term global LAI products. The EEMD method decomposes the time series into a set of oscillatory components at different frequency levels while overcomes the scale mixing problem (Huang et al., 1998). We used EEMD to decompose the long-term LAI products into four components, and MODIS LAI products into three components depending on the length of the time
 327 328 329 330 331 332 333 334 	2.3.1 Time-series analysis We used an Ensemble Empirical Mode Decomposition (EEMD) method along with the classical linear model method to detect trends in the long-term global LAI products. The EEMD method decomposes the time series into a set of oscillatory components at different frequency levels while overcomes the scale mixing problem (Huang et al., 1998). We used EEMD to decompose the long-term LAI products into four components, and MODIS LAI products into three components depending on the length of the time series. The last two components were summed to generate adaptive trends. Annual
 327 328 329 330 331 332 333 334 335 	2.3.1 Time-series analysis We used an Ensemble Empirical Mode Decomposition (EEMD) method along with the classical linear model method to detect trends in the long-term global LAI products. The EEMD method decomposes the time series into a set of oscillatory components at different frequency levels while overcomes the scale mixing problem (Huang et al., 1998). We used EEMD to decompose the long-term LAI products into four components, and MODIS LAI products into three components depending on the length of the time series. The last two components were summed to generate adaptive trends. Annual anomalies were obtained by subtracting adaptive trends from the original time series.
 327 328 329 330 331 332 333 334 335 336 	2.3.1 Time-series analysis We used an Ensemble Empirical Mode Decomposition (EEMD) method along with the classical linear model method to detect trends in the long-term global LAI products. The EEMD method decomposes the time series into a set of oscillatory components at different frequency levels while overcomes the scale mixing problem (Huang et al., 1998). We used EEMD to decompose the long-term LAI products into four components, and MODIS LAI products into three components depending on the length of the time series. The last two components were summed to generate adaptive trends. Annual anomalies were obtained by subtracting adaptive trends from the original time series. We calculated the standard deviation of each detrended anomaly as a quantitative metric

338 2.3.2 Annual maximum LAI analysis

339 We developed annual estimates of maximum summer LAI from 1984 to 2016 (LAI_{max}) by 320,000 sampling sites in the global scope using Landsat surface 340 341 reflectance (Landsat Collection 1; 30 m resolution). The number of sampling sites for each vegetation type was proportional to the global area of the vegetation type. We first 342 343 buffered each site by 50 m (radius) and then used GEE to extract all Landsat 5, 7, and 344 8 surface reflectance acquired from June to August for the Northern Hemisphere, December to February for the Southern Hemisphere, and annually for the tropics, 345 during 1984-2016. 346

347 The annual LAI_{max} could be sensitive to multiple factors including the radiometric difference between Landsat sensors and the availability and timing of Landsat 348 349 observations. This study used the phenological curve reconstruction method to estimate annual LAImax from clear-sky Landsat images (Berner et al., 2020). The method 350 351 modeled seasonal land surface phenology at each site for every 17 years between 1984 352 to 2016 and then predicted annual LAImax using individual summer observations and 353 the phenology information during the corresponding period. As such, annual LAI_{max} can be reliably estimated even if few clear-sky summer measurements were available. 354

- **355 3 RESULTS**
- 356 **3.1 The Landsat LAI validation dataset**

For LAI model training, approximately 19.32 million sample pairs were acquired after the screening process. The SHR had the largest sample size (nearly 6,960,000) and ENF had the smallest sample size (654,800) (Table S1). These sample sizes were

360	considered sufficient for all biomes (Figure S1). Based on the training sample pairs,
361	biome-specific and Landsat sensor-specific Random Forest regression models were
362	built with R^2 of all models > 0.85 (Table S2). The LAI models were assessed by
363	generating Landsat LAI at the field site locations. We finally acquired 38 Landsat LAI
364	values that temporally coincided with the field measurements. The comparison
365	produced a R ² of 0.76 and the scatterplot was around the 1:1 line, which demonstrated
366	the effectiveness of the Landsat LAI inversion algorithm (Figure 2).



FIGURE 2 The comparison of field LAI measurements and Landsat estimated LAI. (a)
is the spatial distribution of temporally coincided field sites and (b) is the scatterplot
between field measurements and Landsat estimated LAI.

371 For Landsat LAI prediction, a total of 68,542,200 predicting samples were obtained

372	after the screening process with GRA having the largest sample size (27,370,300) and
373	DBF having the smallest (2,059,400) (Table S3). LAI of all predicting samples was
374	estimated from Landsat data (20×20 pixels in 30 m resolution) and then aggregated to
375	the spatial resolution of 1/12°. The final sample size of the Landsat LAI validation
376	dataset was about 4.9 million (1984-2020). We used a temporal subset (1984-2016)
377	for the validation work, with a sample size of 3.6 million (Figure 3). Details on the
378	training sample pairs, training model accuracies, and the predicting sample pairs can be
379	found in Supplementary materials.



FIGURE 3 Spatial distribution of the predicting sample size in 1/12°.

382 **3.2** Systematic evaluation of global LAI products in old and new versions

383 **3.2.1 Intercomparison of LAI trends among the products**

384 (a) Trends in terms of versions and periods

385 During 1982–2014, the average LAI in the new version of GLOBMAP LAI

386 increased steadily, reaching $1.55 \times 10^{-3} m^2 m^{-2} a^{-1}$, while that of the old version

- decreased by $2.69 \times 10^{-3} m^2 m^{-2}$ per year (Table 1; Figure 4). The average LAI in old
- and new versions of GIMMS LAI3g and GLASS LAI showed substantial increasing
- trends. The average LAI trend of GIMMS LAI3g presented a minor difference between

390	old and new versions $(4.61 \times 10^{-3} m^2 m^{-2} a^{-1} \text{ vs } 3.2 \times 10^{-3} m^2 m^{-2} a^{-1})$. The average
391	LAI trends of old $(11.42 \times 10^{-3} m^2 m^{-2} a^{-1})$ and new $(4.38 \times 10^{-3} m^2 m^{-2} a^{-1})$
392	versions of GLASS LAI were larger than others.

The LAI trends differed significantly between the periods of 1982–1999 (Phase I) 393 and 2000–2014 (Phase II) (Table 1; Figure 4). The growth rate in Phase I was generally 394 395 larger than Phase II, except for the new version of GLOBMAP LAI. Both versions of 396 GIMMS LAI3g showed rapidly increasing trends in Phase I (ov: $8.22 \times$ $10^{-3} m^2 m^{-2} a^{-1}$; nv: 8.24× $10^{-3} m^2 m^{-2} a^{-1}$), but slowed down after 2000. The new 397 version of GLASS LAI showed a continued increasing trend in Phase I 398 $(6.49 \times 10^{-3} m^2 m^{-2} a^{-1})$ and a decreasing trend in Phase II $(-1.59 \times 10^{-3} m^2 m^{-2} a^{-1})$; 399 400 while its old version was continuously rising, but with a significant decreased rate from 401 (Phase I) to (Phase II). From Phase I to Phase II, GLOBMAP LAI showed a rapid decline for the old version $(2.78 \times 10^{-3} m^2 m^{-2} a^{-1} \text{ to } -8.1 \times 10^{-3} m^2 m^{-2} a^{-1})$ and an 402 403 increase for the new version. The mean LAI of the old and new versions of MODIS LAI showed opposite trends, with a small decrease for the old version and a significant 404 405 increase for the new version. Since both GLASS LAI and GLOBMAP LAI were based on MODIS data in Phase II, the version update of MODIS data had profound but 406 407 different impacts on GLASS LAI and GLOBMAP LAI products due to their distinct 408 algorithms.

409 **TABLE 1** The trend of growing season mean LAI $(10^{-3} m^2 m^{-2} a^{-1})$ in periods of 410 1982–1999 (Phase I) and 2000–2014 (Phase II).

1982-2014	2000-2014	1982-1999

Thi	is is c	a non-peer	reviewed	EarthAr	Xiv pre	eprint (Septembe	er 11,	2023)
-----	---------	------------	----------	---------	---------	----------	----------	--------	-------

GIM (ov)	4.61**	6.57**	8.24
GIM (nv)	3.2**	2.61*	8.22**
GLA (ov)	11.42**	1.93	11.19*
GLA (nv)	4.38**	-1.59	6.49*
GLO (ov)	-2.69**	-8.1**	2.78**
GLO (nv)	1.55*	5.76**	2.93*
MOD (ov)	١	-0.85	\
MOD (nv)	١	13.96**	\

411 ov: old version. nv: new version. GIM represents GIMMS. GLA represents GLASS. GLO





414 **FIGURE 4** Annual variations of the growing season mean LAI for global LAI products.



416 Figure 5a shows the LAI adaptive trends within each decade. During 1982–1991,

- 417 the new version of GLOBMAP LAI and old versions of GIMMS LAI3g and GLASS
- 418 LAI showed decreasing and then increasing trends; the old version of GLOBMAP LAI
- 419 and new version of GIMMS LAI3g showed significant increasing trends; and the new

420	version of GLASS LAI showed a steady trend. During 1992-2001, all products showed
421	increasing and decreasing trends except for the old version of GLOBMAP LAI which
422	showed a continuous decrease. During 2002–2014, the new version of GLOBMAP LAI
423	and both versions of GIMMS LAI3g had a decreased and then increased trend; the old
424	version of GLASS LAI showed an increased and decreased trend; and the new version
425	of GLASS LAI and the old version of GLOBMAP LAI showed a continuous decreasing
426	trend. All products showed large anomalies in the pre-MODIS period with different
427	magnitudes (Figure 5b). GLOBMAP LAI and GLASS LAI inherited the anomalies
428	from MODIS LAI in the post-MODIS period.



430 **FIGURE 5** Global mean LAI, adaptive trends detected by the EEMD method, and 431 detrended anomalies $(m^2m^{-2}a^{-1})$ for old and new versions of LAI products. 432 Anomalies (solid curves in b) were subtracted from adaptive trends (dashed curves in 433 a) from global mean LAI values (solid curves in a).

During 1982–2014, new versions of long-term LAI products presented similar interannual variability $(2.08-2.46 \times 10^{-2} m^2 m^{-2})$ (Table 2). The interannual variability of old versions followed the descending order of GLASS LAI, GIMMS LAI3g, and GLOBMAP LAI. The interannual variability in the period 1982–1999 or from the old version was larger than that in the period 2000–2014 or from the new

439	version. The largest interannual variability was from the old version of GLASS LAI
440	during 1982–1999 (8.37 × $10^{-2} m^2 m^{-2}$) and the smallest was from MODIS LAI
441	during 2000–2014 (0.99–1.06× $10^{-2} m^2 m^{-2}$)

	GIM	GIM	GLO	GLO	GLA	GLA	MOD	MOD
	(ov)	(nv)	(0V)	(nv)	(ov)	(nv)	(ov)	(nv)
1982-2014	4.89	2.46	1.68	2.08	6.63	2.41	\	\
1982–1999	6.18	3.15	1.81	2.43	8.37	2.72	\	\
2000-2014	2.75	1.36	1.55	1.57	3.87	2.02	1.06	0.99

442 **TABLE 2** Interannual variability of four global LAI products $(10^{-2}m^2m^{-2}a^{-1})$.

443 (c) Spatial patterns

444 For old versions, the greening area dominated the landscape during 1982–2014 for 445 GLASS LAI (69%) and GIMMS LAI3g (52%) (Figure 6). GLOBMAP LAI had the 446 highest browning area of about 41%, mainly in Australia, India, and the eastern coastal 447 region of China. Before 2000, all of GLOBMAP LAI, GLASS LAI, and GIMMS LAI3g showed significant greening areas. The proportions of the greening area in 448 449 GLASS LAI (44%) and GIMMS LAI3g (43%) were close, both larger than that of 450 GLOBMAP LAI. After 2000, GLOBMAP LAI showed a vast area of significant 451 browning (36%), except for Australia, India, and the eastern coastal region of China. In 452 GLASS LAI, 14% area appeared browning and 15% of the terrestrial area in the 453 northern high latitudes, southern South America, South Africa, and western Australia were significantly greening. The GIMMS LAI3g showed continuous significant global 454 455 greening in about 31% of the global area, mainly in the eastern Amazon, Congo Basin, 456 and Eurasia. The browning area was in north-central Russia, the middle eastern Amazon, 457 and the Congo region.

458 For new versions, all long-term LAI products showed significant global greening over a majority of the vegetated area and the browning area took only 7-15% during 459 460 1982–2014 (Figure 7). GLOBMAP LAI had the largest browning area, mainly in Asia, 461 Europe, north-central North America, and southern South America. During 1982–1999, 462 only 1%-3% of the area was significantly browning for all long-term LAI products and 463 GIMMS LAI3g had the largest area of significant greening (59%). After 2000, the browning area of the long-term LAI products increased, with the largest area in GLASS 464 LAI (21%). The significant greening area in MODIS LAI was about 58%, mainly in 465 466 the Amazon, Congo Basin, and eastern part of Eurasia.

The trends were compared between versions (Figure 8). During 1982–2014, the 467 468 new version of GLOBMAP LAI had a higher greening trend than the old version for 88% of the area concentrated in the tropics. For GIMMS LAI3g, the old version had a 469 470 higher trend for 65% of the area mainly in Asia, Europe, and South America but the mean differences between versions were small ($\pm 0.005 \ m^2 m^{-2} a^{-1}$). In GLASS LAI, 471 472 the new version had a higher LAI trend in about 62% of the area, mainly in the central and eastern regions of Asia and Europe, the tropics, and the central-eastern part of North 473 474 America. During 1982–1999, the differences in the trend were more obvious for 475 GLASS LAI, especially in the high northern latitudes and tropics. The spatial pattern of trend differences between versions for 2000-2014 was significantly different from 476 1982–1999. For GIMMS LAI3g and GLASS LAI, their new versions had smaller LAI 477 478 growth rates at 68% and 57% of the global area, respectively. The new version of

- 479 GLOBMAP LAI had a larger growth rate for about 91% of the global area, which were
- 480 similar to MODIS LAI (97%).



- 482 **FIGURE 6** The spatial pattern of LAI trends for the old version of global LAI products
- 483 in the growing season. The black point represents the p < 0.05.



- 485 FIGURE 7 The spatial pattern of global LAI trends for the new version of global LAI
- 486 products in the growing season. The black point represents the p < 0.05.



488 FIGURE 8 Spatial patterns of trend differences between the old and new versions of

LAI products. Numbers in square brackets mean the proportional area that the LAI
product in the new version showed a faster (first number) or slower (second number)
growth rate than the old version.

- 492 **3.2.2** Direct evaluation using the Landsat LAI validation dataset
- 493 (a) Overall accuracies

494 The comparison results between the Landsat LAI validation dataset and the new version of long-term global LAI products showed a higher correlation for GIMMS 495 LAI3g (R=0.96-0.97) and GLASS LAI (R=0.95-0.96) than GLOBMAP LAI 496 (R=0.88-0.90) (Figure 9). The MAE and RMSE of GIMMS LAI3g (MAE=0.27-0.29 497 498 m^2m^{-2} , RMSE=0.47-0.49 m^2m^{-2}) were also slightly lower than those of GLASS LAI 499 $(MAE=0.31-0.32 \ m^2m^{-2})$, RMSE=0.51-0.55 m^2m^{-2}) and GLOBMAP LAI (MAE=0.52-0.54 m^2m^{-2} , RMSE=0.91-0.98 m^2m^{-2}). For the old versions, GLASS 500 LAI (R=0.95-0.97) had a higher correlation than GIMMS LAI3g (R=0.95) and 501 GLOBMAP LAI (R=0.89-0.90). The MAE and RMSE of the old version of GIMMS 502 LAI3g (MAE= $0.33-0.35 m^2m^{-2}$, RMSE= $0.58-0.59 m^2m^{-2}$) were significantly 503 504 larger than others. The deviation of GLOBMAP LAI before 2000 was larger in the new version. In summary, the data quality from high to low followed the order of GIMMS 505 506 LAI3g, GLASS LAI, and GLOBMAP LAI.

507 (b) In terms of vegetation biome type

From the perspective of vegetation biome type, the data quality of SHR was higherthan other vegetation types. GLOBMAP LAI improved the data quality of SHR in the

510 new version, with R of above 0.68 and MAE and RMSE of less than 0.27 m^2m^{-2} . For

511	GRA with the largest amount of validation samples, GIMMS LAI3g and GLASS LAI
512	showed better quality (R=0.79-0.86; MAE=0.13-0.20 m^2m^{-2} ; RMSE=0.19-0.32
513	m^2m^{-2}) than GLOBMAP LAI whose MAE and RMSE were twice as high as others.
514	EBF presented the lowest LAI quality for all products mainly due to its distribution in
515	the tropics where remote sensing data suffered from frequent cloudiness. As for the
516	quality of EBF, the new version of GLASS LAI (R=0.37-0.48) had a higher correlation
517	with the LAI validation dataset than GIMMS LAI3g (R=0.25-0.27) and GLOBMAP
518	LAI (R=0.23-0.25); yet the new version of GIMMS LAI3g had lower MAE (0.58-0.60
519	m^2m^{-2}) and RMSE (0.76–0.80 m^2m^{-2}).

520 (c) In terms of periods

To explore the data quality differences between 1984–1999 (p2) and 2000–2014 521 (p3) for the GIMMS LAI3g, GLOBMAP LAI, and GLASS LAI products, we used the 522 correlation analysis method to quantify the consistency based on the validation 523 accuracies during 1984-1999 and that during 2000-2014. The mean correlation 524 525 coefficient for the consistency of the old and new versions of GIMMS LAI3g, 526 GLOBMAP LAI, and GLASS LAI were 0.99, 0.96, and 0.99, respectively. The consistency of updated GLASS LAI was slightly improved, and other global LAI 527 528 products remained steady. The results showed that the data quality consistency of 529 GIMMS LAI3g and GLASS LAI was better than that of GLOBMAP LAI. In terms of 530 different vegetation assessment accuracy, the correlation value in p3 phase was higher 531 than in p2 phase. GIMMSLAI 3g for the global area in p3 phase was slightly lower, but 532 it showed that the correlation value for all vegetation types in p3 phase were higher than in p2 phase, and the data quality of the period 2000–2014 was better than that of theperiod 1984–1999.



535 FIGURE 9 The data quality of long-term global LAI products (GIMMS LAI3g,

GLOBMAP LAI, and GLASS LAI) assessed by the Landsat LAI validation dataset.
p1-p3 represent the period of 1984–2016, 1984–1999, and 2000–2014, respectively.
The quality was assessed by indicators of R (a and b), MAE (c and d), and RMSE (e
and f). a, c, and e were for LAI products in the new versions. b, d and f were for LAI

- 540 products in the old versions.
- 541 3.3 Annual maximum LAI trends
- 542 **3.3.1 Global LAI products**
- 543 During 1984–2016, global LAI products had mediocre consistencies with Landsat
- 544 LAI samples in LAI_{max} trend, following descending order of GLOBMAP LAI, (R=0.29),
- 545 GIMMS LAI3g (R=0.22), and GLASS LAI (R=0.20) (Table 3). In GLOBMAP LAI,
- 546 LAI_{max} trends were negatively correlated with Landsat LAI_{max} for SHR (R=-0.05) and
- 547 DBF (R=-0.09) and relatively well correlated with GRA (R=0.44) and CRO (R=0.34).
- 548 GIMMS LAI3g also presented a higher correlation of LAI_{max} trends for CRO (R=0.52)
- 549 and GRA (R=0.33). In GLASS LAI, the high consistencies with Landsat LAI_{max} trend
- 550 appeared for CRO (R=0.55), ENF (R=0.49), and GRA (R=0.40).
- 551 **TABLE 3** The correlations between annual maximum LAI (LAI_{max}) of the Landsat
- validation dataset and long-term global LAI products for different vegetation biomes.

	GRA	SHR	CRO	SAV	EBF	DBF	ENF	GLOBAL
Landsat-GIMMS	0.33	0.18	0.52	0.17	0.09	0.18	0.10	0.22
Landsat-GLOBMAP	0.44	-0.05	0.34	0.21	0.01	-0.09	0.04	0.29
Landsat-GLASS	0.40	0.31	0.55	0.08	0.03	0.49	0.20	0.20
samples size	16020	6614	1021	14010	2701	40	92	40498

553	In terms of the spatial pattern, the Landsat LAI_{max} showed a large-scale increasing
554	trend globally, especially in the Asian and European continental regions (Figure 10).
555	Three global LAI products had a similar spatial pattern of LAI_{max} trend with Landsat
556	LAI in most vegetated areas. In the northern region of Canada, however, the LAI_{max} of
557	Landsat LAI, GIMMS LAI3g, and GLASS LAI showed an increasing trend while

- 558 GLOBMAP LAI showed a decreasing trend. In the eastern part of Asia and Europe, the
- increasing trend of Landsat LAI_{max} and GLOBMAP LAI_{max} exceeded 0.03 $m^2m^{-2}a^{-1}$,
- 560 greater than that of GIMMS LAI3g and GLASS LAI.



561 FIGURE 10 The spatial pattern of trends in annual maximum LAI (LAI_{max}; 562 $m^2m^{-2}a^{-1}$) for the Landsat estimated LAI validation dataset and the long-term global 563 LAI products during 1984–2016.

564 **3.3.2 Ecosystem models**

565 We compared the LAI_{max} trends from ten Dynamic Global Vegetation Models to Landsat LAI at the global scale during 1984-2016 for different types of vegetation 566 567 (Figure 11). The quality of IBIS data was highest in the TRENDY model (STD<0.01, RMS<0.02), while the uncertainty was larger in the LPX-Bern model data (STD=0.04) 568 569 and CLM5.0 model data (STD=0.028). The LAI datasets simulated by TRENDY 570 models differed from each other both in values and uncertainties. We considered the 571 mean value of the model simulated LAI data (Multi-Model Ensemble Mean LAI or MMEM LAI) a higher representation. The uncertainty differed significantly in 572

573	vegetation types. Compared to the Landsat LAI_{max} trend, the MMEM LAI_{max} trend had
574	the highest similarity (R>0.5) and less uncertainty (STD<0.01, RMS<0.01) for SHR.
575	The LAI_{max} trend correlation between Landsat LAI and satellite-based LAI was higher
576	than the LAI dataset simulated by ten models whose dispersion was larger.
577	We characterized the uncertainty among ten global ecosystem models using the
578	standard deviation of TRENDY LAI at pixels (Figure 12). To avoid over-fitting, we
579	analyzed the spatial pattern of the MMEM LAI and found that the pattern agreed with
580	averaged satellite-based LAI products. The MMEM LAI_{max} showed a decreasing trend
581	in southern Australia, central Russia, and western North America where Landsat LAI_{max}
582	had an increasing trend. MMEM LAI_{max} showed an increasing trend in the tropics, and
583	the growth rate exceeds 0.03 $m^2 m^{-2} a^{-1}$. In terms of the spatial pattern of uncertainties,
584	LAI_{max} based on TRENDY simulation had larger uncertainties in the tropics
585	(>0.05 $m^2m^{-2}a^{-1}$) and smaller uncertainties (<0.01 $m^2m^{-2}a^{-1}$) in regions of
586	southern Australia, South America and high northern latitudes.



FIGURE 11 Comparison of ten Dynamic Global Vegetation Models and long-term global LAI products with the Landsat estimated LAI from 1984–2016 at the global scale for different types of vegetation using Taylor diagrams. The standard deviation represents the interannual variability of the Landsat LAI_{max} trend and LAI_{max} trend derived from the model or satellite. The red line showed a centered root mean square error (RMS) between the Landsat LAI_{max} trend and LAI_{max} trend derived from the model or satellite.



594 FIGURE 12 The trends (a) and uncertainties (b) of mean LAI simulated by TRENDY
595 Process-based Ecosystem Models.

596 4 DISCUSSION

597 4.1 Inconsistencies between current long-term global LAI products

598 Interinconsistencies were found between the long-term global LAI products in 599 trend, interannual variability, and spatial pattern for different product versions and 600 vegetation biome types. Old and new versions of GIMMS LAI3g presented temporally 601 consistent increasing trends in annual average LAI especially around 2000, primarily due to the constant use of AVHRR data across periods despite the sensor turnover from 602 603 AVHRR-2 to AVHRR-3. In contrast, GLASS LAI and GLOBMAP LAI changed the 604 data source from NOAA/AVHRR to Terra/MODIS in 2001 and exhibited significant discrepancies in linear trends between periods (pre-2000 and post-2000) (Figure 4), i.e., 605 606 their post-2000 linear trends were subject to that of MODIS LAI. MODIS C5 suffered

607	from the effect of sensor degradation, leading to questionable LAI trends. The sensor
608	degradation was resolved in MODIS C6 and the LAI trend was corrected in the new
609	versions of GLOBMAP LAI and GLASS LAI. The effects of sensor change were also
610	manifested in the adaptive detrends of GLOBMAP LAI and GLASS LAI, where
611	remarkably different annual anomaly oscillations existed before and after 2001 (Figure
612	5b). The annual anomaly oscillations shall not be explained by environmental factors
613	such as solar radiation, temperature, precipitation, and the CO ₂ fertilization effect
614	(Keenan et al., 2016; Sanchez-Lorenzo et al., 2015; Yan et al., 2013), but rather by
615	changes in satellite platforms and sensors (Jiang et al., 2017).
616	The effect of NOAA satellite orbital drift and AVHRR sensor degradation led to
617	interannual variability in all long-term LAI products (Vermote et al., 2009). The effect
618	could explain the greater interannual variability in the period 1982-1999 over
619	2000–2014 and in GIMMS LAI3g over GLOBMAP LAI and GLASS LAI. This study
620	confirmed a better intraconsistency in GLOBMAP LAI, which has been attributed to
621	its LAI retrieval algorithm (Jiang et al., 2017). Aerosol and cloudiness were other
622	potential factors driving the interannual variability, especially for tropical evergreen
623	forests which contribute most to global LAI year-to-year variations (Samanta et al.,
624	2010).

625 **4.2 Findings from the direct evaluation using Landsat LAI samples**

626 The consistent radiometric performance and high resolution (30 m) make Landsat 627 data a potentially solid LAI reference; and the long archive since the 1970s and a global 628 coverage of observation make Landsat data the only and best choice to evaluate long-

35

629	term global LAI products when other LAI reference was absent before the year 2000
630	(Wulder et al., 2019; Hermosilla et al., 2019). The massive high-quality Landsat LAI
631	validation samples generated in this study enabled a direct evaluation of current long-
632	term global LAI products namely, GIMMS LAI3g, GLASS LAI, and GLOBMAP LAI.
633	A large amount of pre-2000 Landsat validation samples (1,453,228) was created in this
634	study. To guarantee the quality of Landsat based LAI samples, individual random forest
635	models were built according to vegetation biomes and Landsat sensors (TM, ETM+,
636	OLI) so that the different radiative transfer mechanisms in vegetation biomes and
637	distinct spectral characteristics in Landsat sensors could be accounted for. With
638	sufficient LAI samples produced for all vegetation biome types, we were able to not
639	only identify the significant variations between LAI products at the regional scale
640	(Wang et al., 2022; Jiang et al., 2017) but also detect LAI data quality for different
641	biomes. Annual trends of LAI could also be directly validated using Landsat LAI
642	validation samples by calculating LAImax. As such, from different perspectives we
643	could determine the best LAI products rather than merely the relative differences
644	between them.

Based on the Landsat LAI samples, this study found the best data accuracy from GIMMS LAI3g, followed by GLASS LAI and GLOBMAP LAI. The quality of updated GIMMS LAI after 2000 and GLOBMAP LAI before 2000 was relatively low. In the EBF of Africa, for instance, the GIMMS LAI3g exhibited a decreasing trend from the year 2000 while the MODIS LAI showed an increasing trend (Wang et al., 2022). We developed annual estimates of maximum summer LAI from 1984 to 2016 to detect

whether the vegetation was greening or browning. A significant finding was that all 651 652 current long-term global products potentially underestimate the greening area of the 653 Earth to different extents. This finding prompted a more solid evaluation of vegetation 654 responses and feedback under current environmental changes. The consistency with Landsat LAImax trends followed a descending order of GLOBMAP LAI, GIMMS 655 656 LAI3g, and GLASS LAI. This can be explained by GLOBMAP LAI better reflected 657 trends in SHR and GRA which dominate the global landscape and had higher data qualities than other vegetation types. The low LAI quality of EBF was mainly due to 658 its distribution in the tropics where remote sensing data suffered from frequent 659 660 cloudiness. Vegetation in northern high latitudes with the polar night phenomenon and 661 low solar altitude angle also presented higher LAI uncertainties.

662 **4.3 Potential uncertainties**

Despite our efforts, uncertainties existed in the Landsat LAI validation dataset. 663 First, the data quality of training and predicting sample pairs could be lowered by the 664 geometric errors between Landsat and MODIS data and the heterogeneous nature of 665 land cover (Yan et al., 2016). This type of uncertainty was also presented in other studies 666 that employed multiple remote sensing data and can hardly be eliminated. 667 668 Misclassification in the MODIS Land Cover product was another source of 669 uncertainties that affected the sample quality (Fang et al., 2013; Fang et al., 2019). 670 Second, the size of Landsat LAI samples was limited in certain regions, e.g., the 671 northern high latitudes and tropical areas. Future work could involve other high-672 resolution satellite images, e.g., Sentinel-2, to improve the availability of global cloud673 free observation. Last, spatiotemporally continuous ground LAI measurements were 674 desired to optimize our Random Forest regression models. Deep learning methods of 675 higher complexity and stronger prediction power were also welcomed to improve the 676 accuracy of LAI estimation.

677 **5 CONCLUSION**

678 In this study, we generated an LAI validation dataset of massive samples and used 679 the validation dataset to provide a direct evaluation of current long-term global LAI products. The LAI validation dataset, with 4.9 million high-quality samples from 1984 680 681 to 2020, was derived from rigorously selected and refined Landsat samples with the 682 Random Forests regressor and MODIS LAI. It addressed the lack of long-term globewide LAI reference, especially before 2000. We used an ensemble empirical mode 683 684 decomposition method along with the classical linear model to detect the LAI trend of long-term global LAI products (GIMMS LAI3g, GLOBMAP LAI, GLASS LAI, and 685 686 MODIS LAI) in various versions. The temporal and spatial inconsistency of the LAI 687 products of different versions were explored. We also constructed a phenological curve to develop annual estimates of the maximum summer LAI (LAI_{max}) dataset to assess 688 the consistency of trends and interannual variability of the long-term global LAI 689 690 products and the LAI simulated by TRENDY ecosystem process models. The results 691 showed the best data quality of GIMMS LAI3g, followed by GLASS LAI, and 692 GLOBMAP LAI. The data quality in the EBF was generally poor. The LAImax trend of GLOBMAP LAI best matched the Landsat LAI_{max} trend, followed by GIMMS LAI3g 693 694 and GLASS LAI. The Landsat LAI validation dataset produced in this study can

- facilitate the development of long-term global LAI products. The evaluation results of
 current global LAI products can provide a quantitative reference for the rational
 application of LAI for global vegetation dynamic monitoring in the context of climate
 change.
- 699 ACKNOWLEDGMENTS

This work was supported by the National Natural Science Foundation of China (42271104, 41901122), the Shenzhen Fundamental Research Program (GXWD20201231165807007–20200814213435001), and the Shenzhen Science and Technology Program (JCYJ20220531093201004).

- 704 **REFERENCES**
- 705 Baret, F., Morissette, J. T., Fernandes, R. A., Champeaux, J. L., Myneni, R. B., Chen,
- 706 J., Plummer, S., Weiss, M., Bacour, C., Garrigues, S., and Nickeso, J. E. (2006).
- 707 Evaluation of the representativeness of networks of sites for the global
- 708 validation and intercomparison of land biophysical products: proposition of the
- 709 CEOS-BELMANIP. IEEE Transactions on Geoscience and Remote Sensing,
- 710 44(7): 1794–1803. https://doi.org/10.1109/TGRS.2006.876030
- 711 Berner, L. T., Massey, R., Jantz, P., Forbes, B. C., Macias-Fauria, M., and Myers-Smith,
- I. (2020). Summer warming explains widespread but not uniform greening in
 the Arctic tundra biome. Nature Communications, 11: 4621.
- 714 https://doi.org/10.1038/s41467-020-18479-5
- Breda, N. J. J. (2003). Ground-based measurements of leaf area index: a review of
 methods, instruments and current controversies. Journal of Experimental
- 717 Botany, 54(392): 2403–2417. https://doi.org/10.1093/jxb/erg263

- 718 Breiman, L. (2001). Random Forests. Machine Learning, 45(1): 5–32.
 719 https://doi.org/10.1023/A:1010933404324
- 720 Buermann, W., Wang, Y. J., Dong, J. R., Zhou, L. M., Zeng, X. B., Dickinson, R. E.,
- Potter, C. S., and Myneni, R. B. (2002). Analysis of a multiyear global
 vegetation leaf area index data set. Journal of Geophysical Research, 107(D22):
 4646. https://doi.org/10.1029/2001JD000975
- Chen, Y., Chen, L., Cheng, Y., Ju, W., and Ruan, H. (2020). Afforestation promotes
 the enhancement of forest LAI and NPP in China. Forest Ecology and
 Management, 462: 117990. https://doi.org/10.1016/j.foreco.2020.117990
- 727 Claverie, M., Matthews, J., Vermote, E., and Justice, C. (2016). A 30+ year AVHRR
- LAI and FAPAR climate data record: Algorithm description and validation.
 Remote Sensing, 8(3): 263. https://doi.org/10.3390/rs8030263
- Deng, F., Chen, J. M., Plummer, S., Chen, M., and Pisek, J. (2006). Algorithm for
 global leaf area index retrieval using satellite imagery. IEEE Transactions on
 Geoscience and Remote Sensing, 44(8): 2219–2229.
 https://doi.org/10.1109/TGRS.2006.872100

734 Eyring, V., Gillett, N. P., Achutarao, K., Barimalala, R., Barreiro Parrillo, M., Bellouin,

- 735 N., Cassou, C., Durack, P., Kosaka, Y., McGregor, S., Min, S.-K., Morgenstern,
- 736 O., and Sun, Y.: Human Influence on the Climate System, in: Climate Change
- 737 (2021). The Physical Science Basis. Contribution of Working Group I to the
- 738 Sixth Assessment Report of the Intergovernmental Panel on Climate Change
- 739 [Masson-Delmotte, V., Zhai, P., Pirani, A., Connors, S.L., Péan, C., Berger, 570

740	S., Caud, N., Chen, Y., Goldfarb, L., Gomis, M.I., Huang, M., Leitzell, K.,
741	Lonnoy, E., Matthews, J.B.R., Maycock, T.K., Waterfield, T., Yelekçi, O., Yu,
742	R., and Zhou, B. (eds.)], Cambridge University Press, Cambridge, United
743	Kingdom and New York, NY, USA, 423–552,
744	https://doi.org/10.1017/9781009157896.005
745	Fan, L. Y., Franz, H., Berger, and Huizhi, et al. (2014). Validating modis land surface
746	reflectance products using ground-measured reflectance spectra – a case study
747	in semi-arid grassland in inner mongolia, china. International Journal of Remote
748	Sensing, 35(5): 1715-1728. https://doi.org/10.1016/j.rse.2006.09.031
749	Fan, L., Berger, F. H., Liu, H., and Bernhofer, C. (2014). Validating MODIS land
750	surface reflectance products using ground-measured reflectance spectra – a case
751	study in semi-arid grassland in Inner Mongolia, China. International Journal of
752	Remote Sensing, 35(5): 1715–1728.
752 753	Remote Sensing, 35(5): 1715–1728. https://doi.org/10.1080/01431161.2014.882031 1715–1728. 1715–1728.
752 753 754	Remote Sensing, 35(5): 1715–1728. https://doi.org/10.1080/01431161.2014.882031 Fang, H., Baret, F., Plummer, S., and Schaepman-Strub, G. (2019). An overview of Fang, H., Baret, F., Plummer, S., and Schaepman-Strub, G. (2019). Fang, H., Baret, F., Plummer, S., and Schaepman-Strub, G. (2019). Schaepman-Strub, G. (2019). Fang, H., Baret, F., Plummer, S., and Schaepman-Strub, G. (2019). Schaepma
752 753 754 755	RemoteSensing,35(5):1715–1728.https://doi.org/10.1080/01431161.2014.882031Fang, H., Baret, F., Plummer, S., and Schaepman-Strub, G. (2019). An overview of global leaf area index (LAI): Methods, products, validation, and applications.
 752 753 754 755 756 	RemoteSensing,35(5):1715–1728.https://doi.org/10.1080/01431161.2014.882031Fang, H., Baret, F., Plummer, S., and Schaepman-Strub, G. (2019). An overview of global leaf area index (LAI): Methods, products, validation, and applications. ReviewsofGeophysics,57(3):739–799.
 752 753 754 755 756 757 	RemoteSensing,35(5):1715–1728.https://doi.org/10.1080/01431161.2014.882031Fang, H., Baret, F., Plummer, S., and Schaepman-Strub, G. (2019). An overview of global leaf area index (LAI): Methods, products, validation, and applications. ReviewsReviewsofGeophysics,57(3):739–799.https://doi.org/10.1029/2018RG000608
 752 753 754 755 756 757 758 	RemoteSensing,35(5):1715–1728.https://doi.org/10.1080/01431161.2014.882031Fang, H., Baret, F., Plummer, S., and Schaepman-Strub, G. (2019). An overview of global leaf area index (LAI): Methods, products, validation, and applications. ReviewsReviewsofGeophysics,57(3):739–799.https://doi.org/10.1029/2018RG000608Fang, H., Li, W., and Myneni, R. B. (2013). The impact of potential land cover
 752 753 754 755 756 757 758 759 	RemoteSensing,35(5):1715–1728.https://doi.org/10.1080/01431161.2014.882031Fang, H., Baret, F., Plummer, S., and Schaepman-Strub, G. (2019). An overview of global leaf area index (LAI): Methods, products, validation, and applications. ReviewsReviewsofGeophysics,57(3):739–799.https://doi.org/10.1029/2018RG000608Fang, H., Li, W., and Myneni, R. B. (2013). The impact of potential land cover misclassification on modis leaf area index (LAI) estimation: a statistical
 752 753 754 755 756 757 758 759 760 	RemoteSensing,35(5):1715–1728.https://doi.org/10.1080/01431161.2014.882031Fang, H., Baret, F., Plummer, S., and Schaepman-Strub, G. (2019). An overview of global leaf area index (LAI): Methods, products, validation, and applications. Reviews of Geophysics, 57(3): 739–799. https://doi.org/10.1029/2018RG000608Fang, H., Li, W., and Myneni, R. B. (2013). The impact of potential land cover misclassification on modis leaf area index (LAI) estimation: a statistical perspective.RewiewsRemoteSensing,5(2):830–844.

- Fang, H. L. and Liang, S. L. (2005). A hybrid inversion method for mapping leaf area
 index from MODIS data: experiments and application to broadleaf and
 needleleaf canopies. Remote Sensing of Environment, 94(3): 405–424.
 https://doi.org/10.1016/j.rse.2004.11.001
- Fang, H., Wei, S., Jiang, C., and Scipal, K. (2012). Theoretical uncertainty analysis of
- 767 global MODIS, CYCLOPES, and GLOBCARBON LAI products using a triple
- collocation method. Remote Sensing of Environment, 124: 610–621.
 https://doi.org/10.1016/j. rse.2012.06.013
- 770 Forkel, M., Carvalhais, N., Roedenbeck, C., Keeling, R., Heimann, M., and Thonicke,
- K. (2016). Enhanced seasonal CO₂ exchange caused by amplified plant
 productivity in northern ecosystems. Science, 351(6274): 696–699.
 https://doi.org/10.1126/science.aac4971
- Friedl, M. A., Sulla-Menashe, D., Tan, B., Schneider, A., Ramankutty, N., Sibley, A.,
 and Huang, X. (2010). MODIS Collection 5 global land cover: Algorithm
 refinements and characterization of new datasets. Remote Sensing of
- 777 Environment, 114(1): 168–182. https://doi.org/10.1016/j.rse.2009.08.016
- Gao, B. (1996). NDWI—A normalized difference water index for remote sensing of
 vegetation liquid water from space. Remote Sensing of Environment, 58(3):
 257–266. https://doi.org/10.1016/S0034-4257(96)00067-3
- Garrigues, S., Lacaze, R., Baret, F., Morisette, J. T., Weiss, M., Nickeson, J. E.,
 Fernandes, R., Plummer, S., Shabanov, N. V., Myneni, R. B., Knyazikhin, Y.,
- and Yang, W. (2008). Validation and intercomparison of global Leaf Area Index

784	products derived from remote sensing data. Journal of Geophysical Research,
785	113(G2): G02028. https://doi.org/10.1029/2007JG000635
786	GCOS: Systematic observation requirements for satellite-based products for climate,
787	2011 update, 2011.
788	Gessner, U., Niklaus, M., Kuenzer, C., and Dech, S. (2013). Intercomparison of leaf
789	area index products for a gradient of sub-humid to arid environments in West
790	Africa. Remote Sensing, 5(3): 1235 - 1257. https://doi.org/10.3390/rs5031235
791	Hermosilla, T., Wulder, M. A., White, J. C., Coops, N. C., Pickell, P. D., and Bolton,
792	D. K. (2019). Impact of time on interpretations of forest fragmentation: Three-
793	decades of fragmentation dynamics over Canada. Remote Sensing of
794	Environment, 222: 65-77. https://doi.org/10.1016/j.rse.2018.12.027
795	Huang, D., Knyazikhin, Y., Wang, W., Deering, D., Stenberg, P., Shabanov, N., Tan,
796	B., and Myneni, R. (2008). Stochastic transport theory for investigating the
797	three-dimensional canopy structure from space measurements. Remote Sensing
798	of Environment, 112(1): 35-50. https://doi.org/10.1016/j.rse.2006.05.026
799	Huang, N. E., Shen, Z., Long, S. R., Wu, M. C., Shih, H. H., Zheng, Q., Yen, NC.,
800	Tung, C. C., and Liu, H. H. (1998). The empirical mode decomposition and the
801	Hilbert spectrum for nonlinear and non-stationary time series analysis.
802	Proceedings Mathematical Physical & Engineering Sciences, 454(1971): 903-
803	995. https://doi.org/10.1098/rspa.1998.0193

804	Huete, A. R. (2012). Vegetation Indices, Remote Sensing and Forest Monitoring.
805	Geography Compass, 6(9): 513-532. https://doi.org/10.1111/j.1749-
806	8198.2012.00507.x
807	Jiang, C., Ryu, Y., Fang, H., Myneni, R., Claverie, M., and Zhu, Z. (2017).
808	Inconsistencies of interannual variability and trends in long-term satellite leaf
809	area index products. Global Change Biology, 23(10): 4133-4146.
810	https://doi.org/10.1111/gcb.13787
811	Justice, C. O., Townshend, J. R. G., Vermote, E. F., Masuoka, E., Wolfe, R. E., Saleous,
812	N., Roy, D. P., and Morisette, J. T. (2002). An overview of MODIS Land data
813	processing and product status. Remote Sensing of Environment, 83(1-2): 3-15.
814	https://doi.org/10.1016/S0034-4257(02)00084-6
815	Kang, Y., Özdoğan, M., Zipper, S., Román, M., Walker, J., Hong, S., Marshall, M.,
816	Magliulo, V., Moreno, J., Alonso, L., Miyata, A., Kimball, B., and Loheide, S.
817	(2016). How universal is the relationship between remotely sensed vegetation
818	indices and crop leaf area index? A global assessment. Remote Sensing, 8(7):
819	597. https://doi.org/10.3390/rs8070597
820	Kang, Y., Ozdogan, M., Gao, F., Anderson, M. C., White, W. A., Yang, Y., Yang, Y.,
821	and Erickson, T. A. (2021). A data-driven approach to estimate leaf area index
822	for Landsat images over the contiguous US. Remote Sensing of Environment,

- 823 258: 112383. https://doi.org/10.1016/j.rse.2021.112383
- Keenan, T., Prentice, I. C., Canadell, J., Williams, C., Wang, H., Raupach, M., and
 Collatz, J. (2016). Recent pause in the growth rate of atmospheric CO₂ due to

- 826 enhanced terrestrial carbon uptake. Nature Communications, 7(1): 1–9.
 827 https://doi.org/10.1038/ncomms13428
- 828 Knyazikhin, Y., Martonchik, J. V., Diner, D. J., Myneni, R. B., Verstraete, M., Pinty,
- B., and Gobron, N. (1998). Estimation of vegetation canopy leaf area index and
- fraction of absorbed photosynthetically active. Journal of Geophysical Research,

831 103(D24): 32239–32256. https://doi.org/ 10.1029/98JD02461.

- Li, S., Wang, W., Ganguly, S., and Nemani, R. R. (2018). Radiometric Characteristics
- of the Landsat Collection 1 Dataset. Advances in Remote Sensing, 7(3): 203–
- 834 217. https://doi.org/10.4236/ars.2018.73014
- 835 Liang, X., Zhang, T., Lu, X., Ellsworth, D. S., Bassirirad, H., and You, C. (2020).

836 Global response patterns of plant photosynthesis to nitrogen addition: a meta -

- 837 analysis. Global Change Biology, 26(6): 3585–3600.
 838 https://doi.org/10.1111/gcb.15071
- 839 Liu, H. Q. and Huete, A. (1995). A feedback based modification of the NDVI to
- 840 minimize canopy background and atmospheric noise. IEEE Transactions on
- 841 Geoscience and Remote Sensing, 33(2): 457–465.
 842 https://doi.org/10.1109/36.377946
- Liu, Y., Liu, R. G., and Chen, J. M. (2012). Retrospective retrieval of long-term
 consistent global leaf area index (1981–2011) from combined AVHRR and
 MODIS data. Journal of Geophysical Research-Biogeosciences, 117(G4): 14.
- 846 https://doi.org/10.1029/2012JG002084

847	Mao, D., Wang, Z., Ling, L., and Ren, C. (2012). Integrating AVHRR and MODIS data							
848	to monitor NDVI changes and their relationships with climatic parameters in							
849	Northeast China. International Journal of Applied Earth Observation and							
850	Geoinformation, 18(1): 528-536. https://doi.org/10.1016/j.jag.2011.10.007							
851	Myneni, R. B., Hoffman, S., Knyazikhin, Y., Privette, J. L., Glassy, J., Tian, Y., Wang,							
852	Y., Song, X., Zhang, Y., Smith, G. R., Lotsch, A., Friedl, M., Morisette, J.T.,							
853	Votava, P., Nemani, R. R., and Running, S. W. (2002). Global products of							
854	vegetation leaf area and fraction absorbed PAR from year one of MODIS data.							
855	Remote Sensing of Environment, 83(1–2): 214–231.							
856	https://doi.org/10.1016/S00344257(02)00074 - 3							
857	Piao, S., Yin, G., Tan, J., Cheng, L., Huang, M., Li, Y., Liu, R., Mao, J., Myneni, R. B.,							
858	Peng, S., Poulter, B., Shi, X., Xiao, Z., Zeng, N., Zeng, Z., and Wang, Y. (2015).							
859	Detection and attribution of vegetation greening trend in China over the last 30							
860	years. Global Change Biology, 21(4): 1601–1609.							
861	https://doi.org/10.1111/gcb.12795							
862	Piao, S., Wang, X., Park, T., Chen, C., Lian, X., He, Y., Bjerke, J. W., Chen, A., Ciais,							
863	P., Tømmervik, H., Nemani, R. R., and Myneni, R. B. (2020). Characteristics,							
864	drivers and feedbacks of global greening. Nature Reviews Earth & Environment							
865	1(1): 14-27. https://doi.org/10.1038/s43017-019-0001-x							
866	Piao, S., Zhuo, L., Wang, Y., Ciais, P., Yao, Y., and Peng, S. (2018). On the causes							
867	of trends in the seasonal amplitude of atmospheric co2. Global Change Biology,							
868	24(2): 608-616. https://doi.org/10.1111/gcb.13909							

- Piao, S., Sitch, S., Ciais, P., Friedlingstein, P., Peylin, P., and Wang, X. (2013). 869 870 Evaluation of terrestrial carbon cycle models for their response to climate variability and to CO₂ trends. Global Change Biology, 19(7): 2117–2132. 871 https://doi.org/10.1111/gcb.12187 872 Pinzon, J. E. and Tucker, C. J. (2014). A non-stationary 1981–2012 AVHRR NDVI3g 873 874 time series. Remote Sensing, 6(8): 6929-6960. 875 https://doi.org/10.3390/rs6086929 Samanta, A., Ganguly, S., Hashimoto, H., Devadiga, S., Vermote, E., Knyazikhin, Y., 876 Nemani, R. R., and Myneni, R. B. (2010). Amazon forests did not green-up 877 878 during the 2005 drought. Geophysical Research Letters, 37(5), L0540. https://doi.org/10.1029/2009GL042154 879 880 Sanchez-Lorenzo, A., Wild, M., Brunetti, M., Guijarro, J. A., Hakuba, M. Z., Calbo, J., and Bartok, B. (2015). Reassessment and update of long-term trends in 881 downward surface shortwave radiation over Europe (1939-2012). Journal of 882 120(18), 883 Geophysical Research: Atmospheres, 9555-9569. 884 https://doi.org/10.1002/2015JD023321
- 885 Shen, M., Wang, S., Jiang, N., Sun, J., Cao, R., Ling, X., Fang, B., Zhang, Lei, Zhang,
- 886 Lihao, Xu, X., Lv, W., Li, B., Sun, Q., Meng, F., Jiang, Y., Dorji, T., Fu, Y.,
- 887 Iler, A., Vitasse, Y., Steltzer, H., Ji, Z., Zhao, W., Piao, S., and Fu, B. (2022).
- 888 Plant phenology changes and drivers on the Qinghai–Tibetan Plateau. Nature
- 889 Reviews Earth & Environment, 3(10): 633–651.
- 890 https://doi.org/10.1038/s43017-022-00317-5

- Sitch, S., Smith, B., Prentice, C. I. (2003). Evaluation of ecosystem dynamics, plant
 geography and terrestrial carbon cycling in the LPJ dynamic global vegetation
 model. Global Change Biology, 9(2): 161–185. https://doi.org/10.1046/j.13652486.2003.00569.x
- 895 Tucker, C. J., Pinzon, J. E., Brown, M. E., Slayback, D. A., Pak, E. W., Mahoney, R.,
- 896 Vermote, E. F., and El Saleous, N. (2005). An extended AVHRR 8-km NDVI
 897 dataset compatible with MODIS and SPOT vegetation NDVI data. International
- 898 Journal of Remote Sensing, 26(20): 4485–4498.
 899 https://doi.org/10.1080/01431160500168686
- Vermote, E., Justice, C., and Breon, F. (2009). Towards a generalized approach for
 correction of the BRDF effect in MODIS directional reflectances. IEEE
 Transactions on Geoscience and Remote Sensing, 47(3): 898–908.
 https://doi.org/10.1109/TGRS.2008.2005977
- 904 Wang, Z., Wang, H., Wang, T., Wang, L., Liu, X., Zheng, K., and Huang, X. (2022).
- Large discrepancies of global greening: Indication of multi-source remote
 sensing data. Global Ecology and Conservation, 34: e02016.
 https://doi.org/10.1016/j.gecco.2022.e02016
- Wong, S., Cowan, I., and Farquhar, G. (1979). Stomatal conductance correlates with
 photosynthetic capacity. Nature, 282(5737): 424–426.
 https://doi.org/10.1038/282424a0
- Wulder, M. A., Loveland, T. R., Roy, D. P., Crawford, C. J., Masek, J. G., and
 Woodcock, C. E. (2019). Current status of Landsat program, science, and

913	applications. Remote Sensing of Environment, 225: 127-147.
914	https://doi.org/10.1016/j.rse.2019.02.015
915	Xiao, Z., Liang, S., Wang, J., Xiang, Y., Zhao, X., and Song, J. (2016). Long-time-
916	series global land surface satellite leaf area index product derived from MODIS
917	and AVHRR surface reflectance. IEEE Transactions on Geoscience and Remote
918	Sensing, 54(9): 5301-5318. https://doi. org/10.1109/TGRS.2016.2560522
919	Xiao, Z., Liang, S., Wang, J., Chen, P., Yin, X., and Zhang, L. (2014). Use of general
920	regression neural networks for generating the GLASS leaf area index product
921	from time-series MODIS surface reflectance. IEEE Transactions on Geoscience
922	and Remote Sensing, 52(1): 209–223. https://doi.
923	org/10.1109/TGRS.2013.2237780
924	Xu, B., Li, J., Park, T., Liu, Q., Zeng, Y., Yin, G., Zhao, J., Fan, W., Yang, L.,
925	Knyazikhin, Y., and Myneni, R. B. (2018). An integrated method for validating
926	long-term leaf area index products using global networks of site-based
927	measurements. Remote Sensing of Environment, 209: 134-151.
928	https://doi.org/10.1016/j.rse.2018.02.049
929	Yan, K., Park, T., Yan, G., Chen, C., Yang, B., Liu, Z., Nemani, R. R., Knyazikhin, Y.,
930	and Myneni, R. B. (2016). Evaluation of MODIS LAI/FPAR product collection
931	6. Part 1: consistency and improvements. Remote Sensing, 8(5): 359.
932	https://doi.org/10.3390/ rs8050359

- 933 Yan, H., Yu, Q., Zhu, Z. -C., Myneni, R. B., Yan, H. -M., Wang, S.-Q., and Shugart,
- 934 H. H. (2013). Diagnostic analysis of interannual variation of global land

935	evapotranspiration over 1982-2011: Assessing the impact of ENSO. Journal of
936	Geophysical Research: Atmospheres, 118(16): 8969–8983.
937	https://doi.org/10.1002/jgrd.50693
938	You, J., Li, X., Low, M., Lobell, D., and Ermon, S. (2017). Deep Gaussian process for
939	crop yield prediction based on remote sensing data. In: 2017 Association for the
940	Advancement of Articial Intelligence, pp. 4559–4565.
941	https://doi.org/10.5555/3298023.3298229
942	Yuan, H., Dai, Y., Xiao, Z., Ji, D., and Shangguan, W. (2011). Reprocessing the
943	MODIS leaf area index products for land surface and climate modelling.
944	Remote Sensing of Environment, 115(5): 1171–1187.
945	https://doi.org/10.1016/j.rse.2011.01.001
946	Yuan, W., Zheng, Y., Piao, S., Ciais, P., Lombardozzi, D., Wang, Y., Ryu, Y., Chen,
947	G., Dong, W., Hu, Z., Jain, A. K., Jiang, C., Kato, E., Li, S., Lienert, S., Liu, S.,
948	Nabel, J. E. M. S., Qin, Z., Quine, T., Sitch, S., Smith, W. K., Wang, F., Wu,
949	C., Xiao, Z., and Yang, S. (2019). Increased atmospheric vapor pressure deficit
950	reduces global vegetation growth. Science Advances, 5(8): eaax1396.
951	https://doi.org/10.1126/sciadv.aax1396
952	Zeng, Z., Piao, S., Li, L., Zhou, L., Ciais, P., and Wang, T. (2017). Climate mitigation

- 953 from vegetation biophysical feedbacks during the past three decades. Nature
- 954 Climate Change, 7(6): 432–436. https://doi.org/10.1038/nclimate3299

50

955	Zhou, J., Zhang, S., Yang, H., Xiao, Z., and Gao, F. (2018). The retrieval of 30-m
956	resolution LAI from Landsat data by combining MODIS products. Remote
957	Sensing, 10(8): 1187. https://doi.org/10.3390/rs10081187
958	Zhu, Z., Bi, J., Pan, Y., Ganguly, S., Anav, A., Xu, L., Samanta, A., Piao, S., Nemani,
959	R., and Myneni, R. (2013). Global Data Sets of Vegetation Leaf Area Index
960	(LAI)3g and Fraction of Photosynthetically Active Radiation (FPAR)3g
961	Derived from Global Inventory Modeling and Mapping Studies (GIMMS)
962	Normalized Difference Vegetation Index (NDVI3g) for the Period 1981 to 2011.
963	Remote Sensing, 5(2): 927–948. https://doi.org/10.3390/rs5020927
964	Zhu, Z., Piao, S., Myneni, R. B., Huang, M., Zeng, Z., Canadell, J. G., Ciais, P., Sitch,
965	S., Friedlingstein, P., Arneth, A., Cao, C., Cheng, L., Kato, E., Koven, C., Li,
966	Y., Lian, X., Liu, Y., Liu, R., Mao, J., Pan, Y., Peng, S., Peñuelas, J., Poulter,
967	B., Pugh, T. A. M., Stocker, B. D., Viovy, N., Wang, X., Wang, Y., Xiao, Z.,
968	Yang, H., Zaehle, S., and Zeng, N. (2016). Greening of the Earth and its drivers.
969	Nature Climate Change, 6(8), 791-795. https://doi.org/10.1038/nclimate3004
970	Zhu, Z., Woodcock, C. E., Holden, C., and Yang, Z. (2015). Generating synthetic
971	Landsat images based on all available Landsat data: Predicting Landsat surface
972	reflectance at any given time. Remote Sensing of Environment, 162: 67-83.
973	https://doi.org/10.1016/j.rse.2015.02.009
974	

1	Supplementary materials for "A direct evaluation of long-term
2	global Leaf Area Index (LAI) products using massive high-
3	quality LAI validation samples derived from Landsat archive"
4	Junjun Zha ^{1,2} , Muyi Li ^{1,2} , Zaichun Zhu ^{1,2,*} , Sen Cao ^{1,2} , Yanan Zhang ^{1,2} ,
5	Weiqing Zhao ^{1,2} , Yue Chen ^{1,2}
6	¹ School of Urban Planning and Design, Shenzhen Graduate School, Peking University,
7	Shenzhen 518055, China
8	² Key Laboratory of Earth Surface System and Human-Earth Relations, Ministry of
9	Natural Resources of China, Shenzhen Graduate School, Peking University, Shenzhen
10	518055, China.
11	Correspondence: Zaichun Zhu (zhu.zaichun@pku.edu.cn)
12	
13	1.1 Training sample pairs
14	The size of initial training sample pairs was over 20 million for GRA, SHR, and CRO,
15	over 10 million for SAV, DBF, ENF, DNF, and over 6 million for ENF (Table S1). After
16	outlier removal, saturation screening, quality controlling, and AOP index filtering,
17	approximately 16 % of all sample pairs were retained, which can be translated into a size
18	of 19.32 million (Table S1). GRA and SHR samples had relatively high retention rates (20%
19	and 32%). The retention rate for DBF and ENF was only 11% and 6%, respectively, but
20	their sample size was more than 929,800 and 654,800, respectively. The retention rates of
21	other types of vegetation ranged from 10% to 16%, with sample sizes all exceeding 1
22	million. The vegetation biome with the smallest sample size was EBF before screening

23	which suffered from extensive clouds in the tropics; and was ENF after screening primarily
24	due to a large number of samples of misclassified saturation state. For all vegetation biome
25	types, their final sample sizes were believed large enough for establishing robust Random
26	Forest regressors.

- 27 **TABLE S1** The size of training sample pairs (in 10 thousand) and the retention rate (%)
- after the screening. The values were summarized by vegetation biome type.

	Size before screening	Size after screening	Ratio
GRA	2090.87	431.06	20.62
SHR	2149.39	695.95	32.38
CRO	2339.88	251.35	10.74
SVA	1098.17	128.68	11.72
EBF	601.67	92.98	15.45
DBF	1467.04	161.96	11.04
ENF	1115.69	65.48	5.87
DNF	1042.88	104.9	10.06
ALL	11905.6	1932.36	16.23

29 **1.2 Random Forest Regressors**

Based on the refined training sample pairs, we built biome-specific and Landsat sensor-specific Random Forest regression models (Table S2). The R² of all models were considered high (> 0.85). SHR had the highest model accuracy with R² of 0.87–0.95, RMSE of 0.05–0.1, and MAE of 0.08–0.17. The RMSE and MAE values of forests, ranging from 0.22 to 0.43 m^2m^{-2} , were higher than other types of vegetation; meanwhile, their nRMSE values were smaller. The nRMSE of ENF was the lowest, with an average value of 0.06 m^2m^{-2} . This can be explained by the fact that evergreen forests were mainly

57	distributed in areas with high croadiness where good quanty data were searce. Also,
38	evergreen forests always had a low seasonal variability in spectral characteristics and LAI,
39	making the prediction models more sensitive to potential uncertainties in model inputs.
40	TABLE S2 The prediction accuracies of Random Forest regression models for each

41 vegetation biome type.

		MAE	RMSE			Bias
		(m^2m^{-2})	(m^2m^{-2})	R ²	nRMSE	(m^2m^{-2})
	Landsat5	0.13	0.07	0.91	0.26	-0.02
GRA	Landsat7	0.12	0.07	0.91	0.25	0.00
	Landsat8	0.12	0.06	0.92	0.25	0.02
	Landsat5	0.07	0.04	0.94	0.21	0.01
SHR	Landsat7	0.08	0.05	0.95	0.22	0.00
	Landsat8	0.08	0.05	0.95	0.21	0.00
	Landsat5	0.17	0.10	0.87	0.31	0.04
CRO	Landsat7	0.16	0.09	0.87	0.29	-0.01
	Landsat8	0.16	0.09	0.89	0.31	-0.01
	Landsat5	0.29	0.19	0.87	0.22	0.19
SVA	Landsat7	0.27	0.17	0.88	0.21	0.12
	Landsat8	0.31	0.20	0.88	0.21	0.04
	Landsat5	0.33	0.22	0.85	0.06	0.15
EBF	Landsat7	0.34	0.23	0.87	0.06	0.20
	Landsat8	0.35	0.24	0.88	0.06	-0.09
DDE	Landsat5	0.40	0.26	0.96	0.15	-0.11
DRL	Landsat7	0.38	0.24	0.96	0.15	-0.02

	Landsat8	0.45	0.31	0.95	0.13	0.21
	Landsat5	0.54	0.39	0.86	0.20	0.14
ENF	Landsat7	0.53	0.38	0.86	0.20	-0.01
	Landsat8	0.57	0.43	0.85	0.19	-0.09
	Landsat5	0.47	0.33	0.93	0.19	0.32
DNF	Landsat7	0.46	0.32	0.93	0.19	0.24
	Landsat8	0.51	0.37	0.92	0.17	0.17



FIGURE S1 Spatial distribution of the number of training samples for different vegetation
types. The numbers in square brackets represent the total number of training samples (in
500 m resolution) of this vegetation type.

46 **1.3 The final LAI validation dataset**

47 Table S3 shows the size of predicting samples before and after screening (section 2.2.4). 48 Before the screening, predicting sample size was more than 300 million. Grass has the 49 largest predicting sample size of more than 80 million, followed by ENF (63 million), ENF 50 (42 million), DNF (41 million), and SHR (32 million). The DBF had the fewest predicting 51 samples (13 million). After the screening, GRA still had the largest predicting sample size. 52 SHR with the highest retention rate of 49% had the second largest sample size (16 million). The sample size of all forest types greatly decreased (2 million–8 million) with a very low 53 54 retention rate (7% to 15%). The overall retention rate was 22%, eventually producing 68 55 million predicting samples.

56 **TABLE S3** The size of predicting samples (in 10 thousand) and the retention rate (%)

57 after screening. The values were summarized by vegetation biome type.

	Size before screening	Size after screening	Ratio
GRA	8006.08	2737.03	34
SHR	3240.08	1590.62	49
CAO	1740.83	345.90	20
SAV	1786.96	316.17	18
EBF	6347.21	800.59	13
DBF	1392.36	205.94	15
ENF	4374.05	301.04	7
DNF	4102.09	556.95	14

ALL	30989.66	6854.22	22

58