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Defining plastic pollution hotspots

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Abstract

Plastic pollution in the natural environment poses a growing threat to ecosystems and human health, prompting urgent needs for monitoring, prevention and clean-up measures, and new policies. To effectively prioritize resource allocation and mitigation strategies, it is key to identify and define plastic hotspots. UNEP's draft global agreement on plastic pollution mandates prioritizing hotspots, suggesting a potential need for a defined term. Yet, the delineation of hotspots varies considerably across plastic pollution studies, and a definition is often lacking or inconsistent without a clear purpose and boundaries of the term. In this paper, we applied four common hotspot definitions to plastic pollution datasets ranging from urban areas to a global scale. Our findings reveal that these hotspot definitions encompass between 0.8% to 93.3% of the total plastic pollution, covering <0.1% to 50.3% of the total locations. Given this wide range of results and the possibility of temporal inconsistency in hotspots, we emphasize the need for fit-for-purpose criteria and a unified approach to defining plastic hotspots. Therefore, we designed a step-wise framework to define hotspots by determining the purpose, units, spatial scale, temporal scale, and threshold values. Incorporating these steps in research and policymaking yields a harmonized definition of hotspots, facilitating the development of effective plastic pollution prevention and reduction measures.

1 Introduction

Plastic pollution in natural environments has received significant attention from academia and policymakers in recent years, because of the potential detrimental impacts on ecosystems, economies, and human health (van Emmerik and Schwarz, 2020). Plastic pollution can enter the terrestrial and aquatic ecosystems from diverse sources, such as wastewater treatment plants, landfills, agricultural activities, and mismanaged waste in urban areas (Lechthaler et al., 2020; Li et al., 2021). Recent estimates of global plastic leakage to aquatic systems are uncertain and range between 4.8 and 23 million metric tons annually (Boucher and Billard, 2019; Borrelle et al., 2020; Roebroek et al., 2022). Yet, only a small fraction of this mismanaged plastic waste is transported to oceans (Meijer et al., 2021; van Emmerik et al., 2022b). Most mismanaged plastics can accumulate and be retained for decades in aquatic and terrestrial compartments of rivers (van Emmerik et al., 2022b; Kaandorp et al., 2023).

38 Yet, the driving factors and dynamics of plastic transport and accumulation in these compartments
39 are largely unknown (Ford et al., 2022). Observations of plastic pollution are key to solving these
40 unknowns. As such, multiple monitoring efforts were made to understand the transport and fate of
41 plastics in urban areas (Tramoy et al., 2022; Tasserone et al., 2023a), river systems (González-Fernández
42 and Hanke, 2017; Schwarz et al., 2019; Liro et al., 2020; van Emmerik and Schwarz, 2020), beaches
43 (Morales-Caselles et al., 2021; Fruergaard et al., 2023), and oceans (Shim et al., 2022).

44 The methodologies and results from these studies are subsequently used to design effective poli-
45 cies and mitigation strategies to reduce and prevent plastic waste in the environment. Several of
46 these focus on targeting “hotspot” areas. For example, the “National Guidance for Plastic Pollution
47 Hotspotting and Shaping Action” was developed to present a structured framework for the identifi-
48 cation of plastic leakage hotspots, assessments of their impacts across the entire plastic value chain,
49 and subsequent prioritization of actions after hotspots are identified (Boucher and Initiative, 2020).
50 Whilst such frameworks should provide relevant insights for policymakers to tackle plastic pollution,
51 hotspots are not clearly defined, resulting in different definitions across plastic pollution studies. For
52 example, Franceschini et al. (2021) used an inverse distance weighting interpolation method to obtain
53 a map of marine zones where plastic litter tends to accumulate, and subsequently defined hotspots as
54 areas where the number of particles exceeded the 90th percentile of all values. Schuyler et al. (2016)
55 established a global risk map to highlight hotspot areas with a high probability of marine debris inges-
56 tion by sea turtles. In their study, hotspots are qualitatively defined as areas with high concentrations
57 of marine debris and high turtle species diversity. Hotspots are occasionally restrained to specific
58 spatial scales. Lessler et al. (2017) highlighted the importance of defining the spatial scale of hotspots
59 based on various practical considerations, such as data availability and resources, or a specific scale
60 at which the planned interventions are relevant to be implemented. Tasserone et al. (2020) proposed a
61 seven-step approach for plastic hotspot mapping specifically designed for urban water systems, where
62 hotspots were defined when an arbitrary threshold was exceeded. Another key aspect to consider is
63 that hotspot locations might change over time. Temporal variability in plastic abundance can influ-
64 ence whether a hotspot is permanent, or occurs only in extreme events such as peak river discharges
65 (van Emmerik et al., 2023). Lastly, plastic pollution can be reported in either item counts per unit
66 length, area, volume, or mass, which might yield different hotspots based on the composition of the
67 plastic litter. For example, de Lange et al. (2022) shows a strong divergence between the top ten most
68 abundant items when assessed by item counts compared to assessments based on the mass of items.

69 While a common characteristic among these studies is the identification of areas in which plastics
70 are abundantly concentrated, specific criteria and methodologies differ greatly in both temporal and
71 spatial domains. Acknowledging these diverse perspectives and their influence on the reported hotspot
72 areas is crucial for advancing the development of effective, targeted mitigation strategies and their
73 cost-benefit analyses. The draft of UNEP’s legally binding instrument to target plastic pollution
74 worldwide states that government bodies and local parties should prioritize hotspots, and note a
75 definition of the term may be required (UNEP, 2023). For example, in Bangladesh, at national and
76 urban levels, an action plan for sustainable plastic management was developed, in which hotspots were
77 defined as “A place where plastics leak into the environment, including land, air, water, and marine
78 environment, where waste accumulates regularly and is not collected and transported to landfills
79 for proper disposal” (Yoshijima et al., 2021). While this definition points out where pollution is
80 happening, it resulted in over 1,200 hotspots in the city of Dhaka alone, which poses a challenge to
81 efficient mitigation strategies. Another example concerns Queensland’s Plastic Pollution Reduction
82 Plan, which states that state-wide plastic pollution hotspots should be identified and monitored,
83 without giving a hotspot definition (Queensland Government, 2019). It is evident that the concept of
84 hotspots varies significantly across plastic pollution studies. Different concepts may lead to different

85 amounts and localization of hotspots and can have strong impacts on the design and effectiveness of
86 local, national, or international waste management strategies and plastic pollution treaties. Therefore,
87 this study aims to compare four quantitative approaches to define hotspots using plastic pollution
88 datasets on four different spatial scales: global (Meijer et al., 2021), continental (González-Fernández
89 et al., 2021), national (van Emmerik et al., 2020a; Kiessling et al., 2021) and urban (van Emmerik
90 et al., 2020b; Tasseron et al., 2023a). By exploring the similarities, differences, and limitations of these
91 approaches, our study seeks to provide a comprehensive understanding of the definition of hotspots
92 and their impact on hotspot identification. We propose a framework toward a harmonized definition
93 of plastic hotspots that can be used to support local, national, and international ambitions to end
94 plastic pollution.

95 2 Methods

96 2.1 Datasets

97 In this study, seven plastic pollution datasets on four different spatial scales were used. An overview
98 of these datasets is shown in Table 1, and the distribution of values within each dataset in Figure 1.
99 Firstly, a global dataset presented by Meijer et al. (2021) modeled over 100,000 outlets of rivers and
100 streams, of which nearly 32,000 locations are reported to leak plastic litter into marine environments.
101 In this study, the amount of plastic was quantified in million tonnes per year [MT yr⁻¹]. A subset of
102 this dataset is taken that includes all plastic leaking locations in continental Europe, resulting in 2,626
103 locations (administrative areas obtained from <https://gadm.org/data.html/>). Next, a continental
104 dataset presented by González-Fernández et al. (2021) modeled the annual floating macrolitter load
105 (FML) from Europe into the ocean at over 32,500 locations in plastic items per year [items yr⁻¹].
106 Two datasets at the national scale are used. Firstly, the dataset presented by Kiessling et al. (2021)
107 contains riverine floating macrolitter observations for nearly 150 sites in Germany conducted in the
108 spring of 2017, reported in items per meter per sampling hour [items m⁻¹ h⁻¹]. Secondly, another
109 national dataset presented by van Emmerik et al. (2020a) includes over 500 sampling locations along
110 riverbanks in the Dutch Rhine-Meuse delta. In the spring of 2021, the riverbank macrolitter was
111 sampled and categorized using the River-OSPAR method [items m⁻¹ riverbank] (van Emmerik et al.,
112 2020a). Lastly, we used two datasets at the urban scale. One was a dataset of crowd-based observations
113 of floating plastic in the city of Leiden, the Netherlands (van Emmerik et al., 2020b). The abundance
114 of floating litter was registered for over 200 locations within the urban waterways [items]. The other
115 dataset at the urban scale was collected in Amsterdam (Tasseron et al., 2023b). The abundance of
116 litter was registered at 150 distinct locations in the canals of the historic city center [items]. Floating
117 items were observed, counted, and categorized according to the River-OSPAR protocol. A map
118 showing the locations and observed floating litter counts is found in Figure 6 in the Appendix.

Table 1: Datasets used in this study, including the reference, type of data, the spatial scale, type of water body, number of locations, and the unit of the measurements.

Reference	Type	Spatial scale	Water body	# locations	unit
Meijer et al. (2021)	Model output	Global	Outlets into ocean	31,820	MT yr ⁻¹
Meijer et al. (2021)	Model output	Continental (Europe)	Outlets into ocean	2,626	MT yr ⁻¹
González-Fernández et al. (2021)	Model output	Continental (Europe)	Outlets into ocean	32,652	items yr ⁻¹
Kiessling et al. (2021)	Observations	National (Germany)	Rivers and streams	141	items m ⁻¹ h ⁻¹
van Emmerik et al. (2020a)	Observations	National (Netherlands)	Riverbanks	512	items m ⁻¹
van Emmerik et al. (2020b)	Observations	Urban (Leiden)	Waterways	217	items
Tasseron et al. (2023b)	Observations	Urban (Amsterdam)	Waterways	150	items

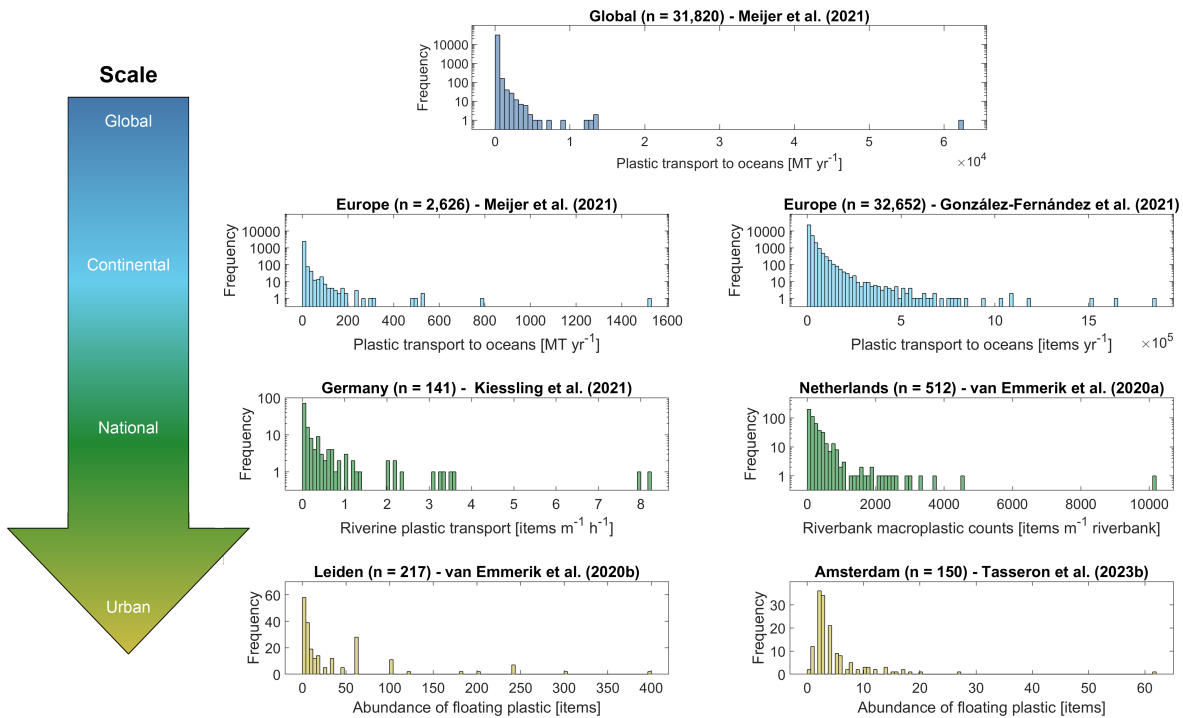


Figure 1: Distribution of plastic pollution values for each dataset used in this study. Note the difference in scale on both the y and x axes.

2.2 Hotspot definitions and statistical analyses

Four quantitative definitions were used based on the distribution of the values in the plastic pollution datasets. Values above these thresholds are defined as hotspots:

1. **Above average values.** With this approach, any value above the average value of the dataset is considered a hotspot. This definition is based on Fok and Cheung (2015), who identified microplastic pollution in Hong Kong to be higher than international averages, highlighting Hong Kong as a hotspot of marine plastic pollution. Another example of the above-average definition is used by Fruergaard et al. (2023), who determined that the Nha Trang coast (Vietnam) is a global hotspot for plastic pollution because the mean abundance of plastic litter items found was higher than other beaches worldwide. The “above average” definition is also used in other disciplines, for example, in criminology, hotspots can be classified as areas with an above-average number of criminal indices or areas where the risk of becoming a crime victim is higher than average (Eck et al., 2005).
2. **Values in the highest interval.** Datasets can be divided into several (constant) intervals, in which the highest interval represents hotspots. For example, Tasseron et al. (2020) use five intervals for two urban plastic pollution datasets, in which the highest interval is depicted as a hotspot. These intervals can be chosen arbitrarily or with equal steps in between the intervals. Here, five equal intervals are used for each dataset, in which the highest interval represents hotspots.
3. **Outliers** Outlier values can be defined as hotspots. Extreme events such as high river discharge and (flash) floods can cause high rates of plastic transport (van Emmerik et al., 2022a). Aquatic systems with favorable hydrodynamic conditions for the accumulation of plastic litter can, in turn, be characterized by extreme levels of pollution, which are likely to be outliers in the dataset (Jayasiri et al., 2013). In our study, we use the common definition of outliers to describe hotspots, which are any values above the 3rd quantile plus 1.5 times the interquartile range ($Q3 + 1.5 * IQR$).
4. **Values in the top percentile.** Hotspots can be defined based on values within the upper

146 n-th percentile range (e.g. 90th, 95th, or 99th, depending on the desired number of hotspots).
 147 For example, Franceschini et al. (2021) identified microplastic hotspots as areas in which the
 148 number of particles exceeded the 90th percentile of all values. Here, we use the 90th percentile
 149 value to classify hotspots. The percentile threshold can be adjusted based on specific research
 150 objectives and context.

151 2.3 Temporal variability and consistency

152 Assessing the consistency and changes of hotspots over time can be crucial to effectively allocate
 153 resources for cleanup practices. Here, six monitoring rounds of plastic pollution along Dutch riverbanks
 154 and beaches are used to identify this consistency. Between the fall of 2020 and the spring of 2023, 101
 155 locations were monitored twice a year using the River-OSPAR protocol (van Emmerik et al., 2020a).
 156 In the resulting six datasets, hotspots are defined as values above the 90th percentile, resulting in 10
 157 hotspots for each dataset. The resulting hotspots are analyzed in terms of consistency, whether they
 158 are 1) consistently defined as hotspots for multiple monitoring rounds, 2) hotspots on just a single
 159 occurrence, or 3) not a hotspot.

160 3 Results

161 Applying four hotspot definitions on seven plastic pollution datasets resulted in widely different rep-
 162 resentations of hotspots (Figure 2). A large range was observed in the percentage of the total plastic
 163 pollution contained in hotspots (0.8-93.3%). Similarly, a large range in the percentage of the total
 164 locations defined as hotspots is present (<0.1-50.3%). Here, the findings for each definition and its
 165 implications are presented.

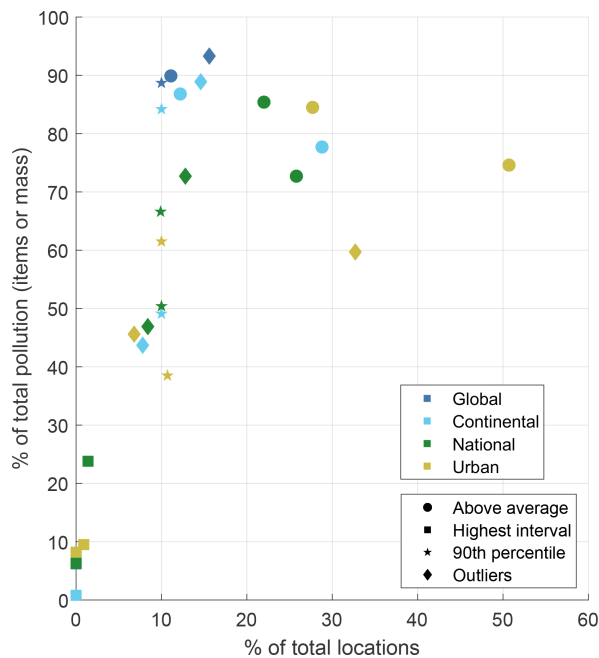


Figure 2: The four different hotspot definitions, and their effect on the percentage of total locations and total items/mass these definitions cover. The colors indicate spatial scale, where dark blue is global, light blue is continental, green is national, and yellow is urban. Different symbols indicate the four hotspot definitions, in which the circle is “Above average”, the square is “Highest interval”, the star is “90th percentile”, and the diamond is “Outliers”.

166 Defining hotspots as any value above the average value of the pollution dataset results in hotspots
 167 containing between 72.7% and 89.9% of the total plastic pollution across the four spatial scales (Figure
 168 2). Out of all the locations in the datasets, between 11.1% (Global) and 50.7% (Urban, Leiden)
 169 are identified as hotspots. This suggests the above-average definition might not effectively pinpoint
 170 extremely polluted areas as hotspots, since certain regions with slightly elevated pollution levels could

171 be overemphasized. When aiming to identify hotspots using this method, efforts and resources can
172 be diverted away from areas with more pressing pollution concerns.

173 Dividing the data into five equal intervals and using the highest interval as the definition for
174 hotspots results in a contrasting situation. Here, hotspots contain between 0.6% and 23.8% of the
175 total pollution, distributed over <0.1%-1.4% of the total locations. In absolute numbers, this method
176 identifies only 1-3 locations at each of the four spatial scales as hotspots. Since the intervals are
177 strongly determined by the highest value in the dataset, an extreme outlier results in many data
178 points in the lowest interval(s) and only a few in the highest. Therefore, given its highly selective
179 nature, this method might not capture areas with moderate to high levels of pollution and may
180 therefore not be suitable for offering comprehensive insights into the spatial distribution of hotspots.

181 Next, using outliers results in 43.3%-93.3% of the total plastic pollution to be included in hotspots.
182 Between 6.8% and 15.6% of the total locations are classified as hotspots (Figure 2). Compared to
183 “Above Average” hotspots and “highest interval” hotspots, the outlier-based method seems to capture
184 a more balanced portion of the pollution, providing a broader spatial perspective. However, it is
185 essential to consider that statistical outliers may not always be present in every dataset. In these rare
186 instances, the outlier-based method is ineffective in pinpointing hotspots.

187 Lastly, using the 90th percentile as the threshold for hotspots, between 38.5% and 88.7% of the
188 total plastic pollution is included for the different spatial scales. Since this statistical threshold is
189 based on a certain percentage of the data, in all cases 10.0% of the total locations are defined as
190 hotspots (Figure 2). This standardized proportion of hotspot locations is a valuable feature of this
191 method, providing a consistent basis for hotspot identification irrespective of the dataset size. When
192 the original number of locations included in monitoring or model outputs is not too large, this method
193 effectively captures plastic hotspots while maintaining a reasonable number of locations for practical
194 management purposes.

195 **3.1 Arbitrary thresholds and spatial variability**

196 In the absence of well-established guidelines for classifying plastic pollution hotspots, researchers and
197 policymakers may resort to using arbitrary thresholds as a pragmatic approach to identify hotspots.
198 In such cases, the threshold is often determined based on a combination of scientific judgment, data
199 availability, and policy objectives (Bank et al., 2021; Lu et al., 2019). For example, an objective of
200 cleanup strategies and actions can be to target the top ten most polluted locations in a city, or the
201 top 50 most polluted global rivers. These thresholds can be based on several factors such as item
202 count, concentration, plastic weight, or a specific number of locations. Contrary to using statistical
203 thresholds, the use of arbitrary thresholds across different scales or studies can hinder meaningful
204 comparisons between hotspots.

205 In Figure 3a, the cumulative distribution of pollution over the share of locations is shown. The
206 curves of the different datasets and spatial scales do not overlap, implying a threshold set on a
207 specific dataset - for example, the 15.0% most polluted locations are hotspots - covers an entirely
208 different share of the total pollution for the “Amsterdam - Urban” dataset (56.0%), compared to the
209 “Meijer et al. (2021) - Global” dataset (7.0%). This is a direct result of the diversity in data collection
210 protocols, units, and various patterns at different spatial scales. For example, when comparing the two
211 “Continental” datasets, it seems the plastic pollution in González-Fernández et al. (2021) is much more
212 diffuse compared to Meijer et al. (2021). Even though the distribution is different at various spatial
213 scales, at least 57.0% of the total pollution is concentrated in 25.0% of the total locations regardless
214 of scale, protocol, and the number of locations. The share of pollution is even higher for 25.0% of
215 the total locations when using a similar protocol, as depicted in Figure 3b. Here, cumulative plots of

216 the pollution distribution across the share of locations are shown using subsets at different scales for
 217 the Meijer et al. (2021) dataset. In this case, at least 79.0% of the total pollution is concentrated in
 218 25.0% of the locations, regardless of scale. These insights could be relevant for designing strategies
 219 tackling plastic pollution on multiple spatial scales simultaneously.

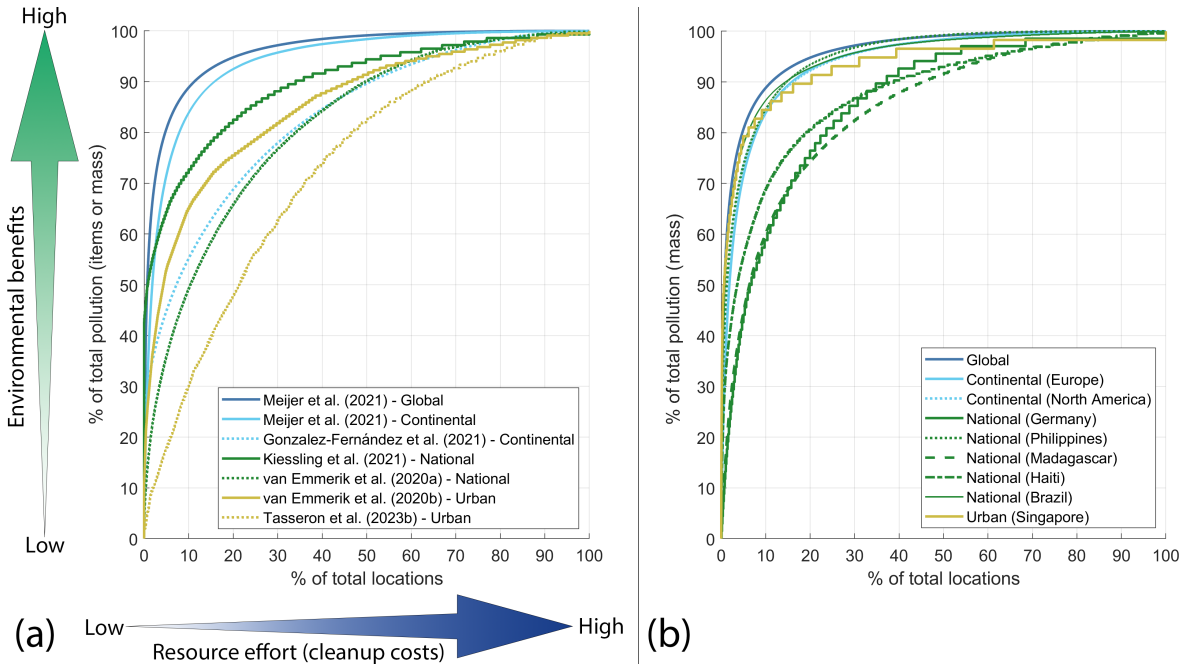


Figure 3: (a) Cumulative plots, expressing the percentage of total locations (x-axis) and percentage of total pollution (y-axis) using the seven different plastic datasets. For management purposes, a higher number of hotspot locations requires more resources, whereas targeting higher shares of plastic pollution for cleanups results in higher environmental benefits. (b) Cumulative plots using different continental, national, and urban subsets of the Meijer et al. (2021) dataset.

220 3.2 Temporal variability and consistency

221 Seasonal variability in hydrological and meteorological patterns can strongly impact the consistency
 222 and composition of plastic hotspots monitored in different periods. In Figure 4, an overview of the
 223 temporal consistency of hotspots along the Dutch riverbanks monitored between the fall of 2020
 224 and the spring of 2023 is depicted. Figure 4a shows the consistency of hotspots over time, using
 225 the “Values in the top percentile” definition, resulting in ten hotspots per monitoring round. In
 226 total, 33 unique locations emerge as hotspots at least once in any of the monitoring rounds. Only
 227 one location is always classified as a hotspot, and just six locations are a hotspot in three or more
 228 monitoring rounds. In Figure 4b, an arbitrary threshold was chosen to define hotspots for the same
 229 six monitoring rounds. Defining hotspots as locations in which >500 plastic litter items per 100-meter
 230 riverbank are monitored results in a different number of hotspots for each season, ranging from one
 231 (Fall 2021) to fifteen (Spring 2021) hotspots. As hotspots change over time, tracking their dynamics is
 232 crucial for the effective allocation of resources for targeted cleanup practices. Continuous monitoring
 233 of longer timescales will yield further insights into hotspots driven by short-term fluctuations (such
 234 as floods (van Emmerik et al., 2023) or shipping cargo spills (Saliba et al., 2022)), and consistent
 235 hotspots characterized by persistent accumulation of plastic litter.

236 4 Discussion and outlook

237 4.1 Temporal variation and consistency of hotspots

238 The seven datasets used in this study to examine the influence of different hotspot definitions were
 239 static, containing plastic pollution values of either model outputs, single observations, or observations

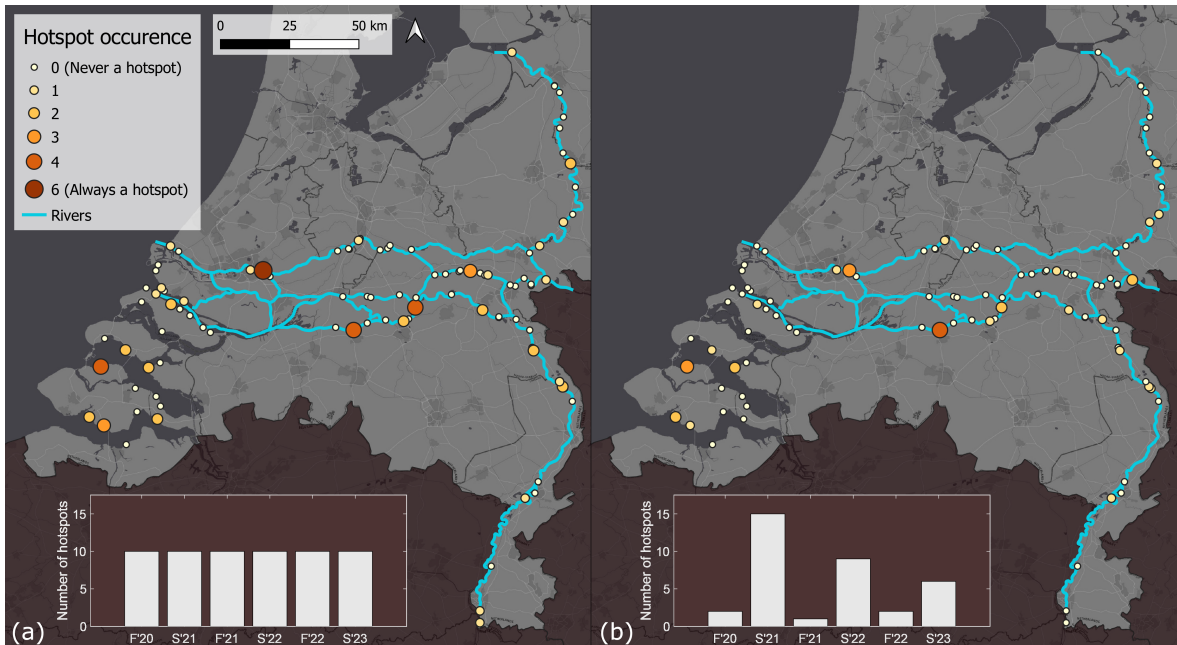


Figure 4: Temporal differences and consistency of hotspot occurrence of plastic pollution along Dutch riverbanks between Fall (F) 2020 and Spring (S) 2023. (a) Shows the hotspots defined according to values above the 90th percentile, resulting in ten hotspots for each monitoring round. (b) Shows the hotspots using an arbitrary threshold, in which hotspots are locations with >500 monitored plastic items per 100-meter riverbank, resulting in a different number of hotspots for each monitoring round. Basemap: Esri, HERE, DeLorme, MapmyIndia.

240 averaged over time. By using time-series data, for instance, twenty years of beach plastic litter
 241 data (Grundlehner et al., 2023), emerging hotspots can be identified and the evolution of previously
 242 identified ones can be tracked (Piacenza et al., 2015). Monitoring pollution hotspots over temporal
 243 timescales helps to distinguish between short-term fluctuations such as floods (van Emmerik et al.,
 244 2023) or areas in which pollution is recurring and consistent, such as accumulation zones (Schwarz
 245 et al., 2019; van Emmerik et al., 2022b). The latter may contribute to the efficient allocation of
 246 resources and cleanup practices. For instance, areas that are characterized by consistent hotspots may
 247 require continuous monitoring and a sustained allocation of resources, whereas regions experiencing
 248 occasional hotspots might demand more targeted interventions during peak pollution periods. Hotspot
 249 monitoring over time might benefit from utilizing a fixed threshold as opposed to statistical thresholds.
 250 Statistical thresholds such as “above average” and “values in the top percentile” will always lead
 251 to hotspots, even when litter abundance can reduce over time due to targeted cleanups. In these
 252 situations, using a definition that can result in the absence of hotspots might better describe the
 253 limited level of pollution. Lastly, monitoring the change of hotspots over time allows the determination
 254 of how effective specific mitigation measures are once implemented.

255 4.2 Qualitative elements and non-measurable components

256 Within plastic pollution research, the identification of hotspots has initially been associated with
 257 quantitative measurements, reporting either an abundance, density, or concentration of pollution in
 258 aquatic compartments. Yet, as introduced by Boucher and Initiative (2020) hotspots can encompass
 259 qualitative and non-measurable components that contribute significantly to their characterization
 260 and complexity. While quantitative data remains fundamental in hotspot identification, qualitative
 261 aspects play a vital role in understanding the broader context of sources, sinks, and dynamics and
 262 their impact on the environment and communities. For example, the quality of waste management
 263 infrastructure, recycling capacities, or the frequency of (informal) cleanups can have a critical influence
 264 on plastic leakage and abundance (Boucher and Initiative, 2020; Mihai et al., 2021). The latter can
 265 also influence the definition of hotspots, for instance when a criterion is required that hotspot locations

266 have good access to infrastructure or are easily accessible for waste management personnel. Other non-
 267 measurable components that potentially influence hotspots include societal attitudes, local policies,
 268 and cultural practices. Therefore, Boucher and Initiative (2020) recommends formulating “actionable”
 269 hotspots to provide a comprehensive view of hotspots across the plastic value chain. These actionable
 270 hotspots should concisely specify the type of plastic involved, identify the expected source of leakage,
 271 and pinpoint potential key drivers for leakage within the waste management system. By combining
 272 measurable data with qualitative insights and non-measurable components, an integrated perspective
 273 emerges that fosters more efficient decision-making and facilitates targeted interventions.

274 4.3 Recommendations for future efforts

275 We recommend using the term “hotspot” in an explicit and meaningful way that is tailored to specific
 276 research objectives or demands for management strategies. While some studies report a detailed
 277 definition of hotspots, it is frequently employed in an imprecise manner as an evocative term to draw
 278 attention to a study (Lessler et al., 2017). Therefore, we propose a step-wise framework, which includes
 279 five main steps to use hotspots in plastic pollution research. These steps are 1) define the purpose,
 280 2) determine units of interest, 3) determine the spatial scale, 4) determine the temporal scale, and 5)
 281 determine and report threshold values (Figure 5).

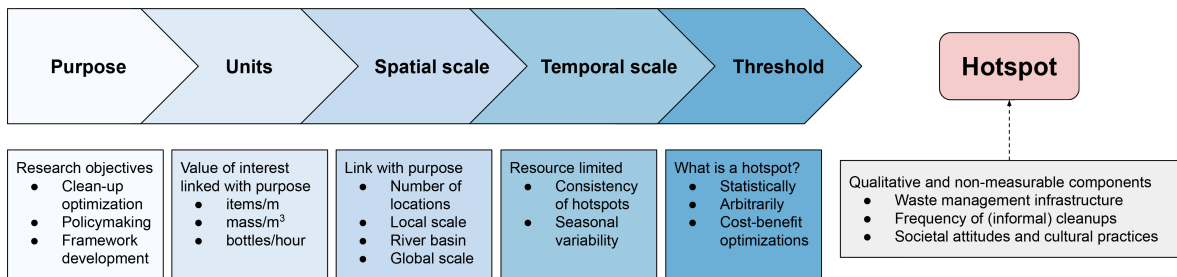


Figure 5: Framework with the recommended steps to be taken when using the term “hotspot” in plastic pollution research. The terms in bullet points are possible definitions for each step and are not limited to these examples.

- 282 1. **Define the purpose of hotspot mapping.** Before proceeding with hotspot identification,
 283 the research objectives and intended use of the hotspot data should be clearly defined. For
 284 example, the purpose could be to prioritize resource allocation for clean-up efforts (Prata et al.,
 285 2019). Another purpose could be to guide (international) policymaking and unified frameworks
 286 to target the most effective mitigative actions (Boucher and Initiative, 2020). Defining the
 287 purpose allows subsequent hotspot definitions to align with desired outcomes. While a unified
 288 hotspot definition is useful in terms of comparability, the purpose of the research might require
 289 specific definitions that differ.
- 290 2. **Determine units of interest** Next, appropriate units that align with the purpose of hotspots
 291 should be defined. For example, when a clean-up effort is designed to reduce the abundance of
 292 plastic mass on riverbanks, an appropriate unit would be [kg/m² riverbank]. Another purpose
 293 could focus very specifically on identifying the transport of PET bottles, in which the appropriate
 294 unit could be [PET bottles/hour] flowing past a measurement location. By selecting the correct
 295 units, goals linked to the purpose of hotspots will be measurable and quantifiable.
- 296 3. **Determine the spatial scale.** As evident from the results presented here, a specific hotspot
 297 definition might contain an entirely different percentage of the total pollution for datasets with
 298 other spatial extents. When assessing hotspots on a local scale, the number of included locations
 299 can be relatively low (Vriend et al., 2020), whereas achieving a more comprehensive understand-

300 ing of hotspots on (multi) river basin scales or even global scale, the number of required sampling
301 sites increases.

302 **4. Determine the temporal scale.** Assessing whether hotspots are persistent or transient might
303 be relevant in the formulation of effective management strategies. Seasonal variability, floods,
304 and various hydrodynamic conditions might all influence the abundance and accumulation of
305 plastic in aquatic environments (Cheung et al., 2016; van Emmerik et al., 2023; Roebroek et al.,
306 2021). Therefore, reporting the temporal scale (period, frequency, structure, and duration) is key
307 to assessing whether hotspots are consistent through time, or whether they are just temporary as
308 a result of specific or unique events. This assessment could be resource-limited, especially when
309 (pilot) monitoring efforts have limited funds available to facilitate high-frequency monitoring.

310 **5. Selecting and reporting the threshold values.** Lastly, researchers and policymakers need
311 to establish and report threshold values in a transparent way that is based on sound scientific
312 reasoning. They can be chosen arbitrarily, statistically, or result from complex cost-benefit
313 optimizations. For example, Christensen et al. (2021) introduced and illustrated a spatial cost-
314 benefit optimization framework allowing the prioritization of limited cleanup efforts of plastic
315 pollution, maximizing the environmental benefits. Another example includes a threshold that
316 could be directed by the development of a global legally binding treaty by the UN to end plastic
317 pollution (March et al., 2022). Such treaties would incentivize policymakers to first target the
318 most heavily polluted hotspots on multiple spatial scales with limited resources, influencing the
319 threshold of when a location is classified as a hotspot.

320 Conclusion

321 Hotspots in plastic pollution research and policymaking are often used to highlight areas in which
322 plastics are concentrated and should be prioritized in monitoring, prevention, and reduction actions.
323 Yet, the definition and characterization of key aspects related to plastic hotspots is often lacking or
324 inconsistent, without a clear purpose and boundaries of the term. Here, we compared different quanti-
325 tative ways of how hotspots can be defined, and shown they vary significantly on four different spatial
326 scales ranging from urban to global datasets. All hotspot definitions combined encompass between
327 0.8-93.3% of the total pollution, distributed across <0.1-50.3% of the total locations. Furthermore, we
328 highlighted hotspots can be dynamic over time, in which the temporal consistency varies greatly per
329 monitored location. Classifying hotspots appropriately is particularly relevant for resource allocation
330 and management strategies to target pollution hotspots. Therefore, we designed a five-step frame-
331 work for defining hotspots in plastic pollution research. By determining the purpose, units, spatial
332 scale, temporal scale, and threshold values, hotspots are defined in an explicit and meaningful way
333 that is suitable for specific research objectives or management strategies. Ultimately, the ability to
334 define, target, and address plastic hotspots effectively is necessary to safeguard our environments and
335 ecosystems and achieve the ambitions to end plastic pollution.

336 Author Contributions

337 Conceptualization, PT, TvE, PV, RH, YM; Data curation, PT; Formal analysis, PT; Funding acquisi-
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339 TvE, MvdP; Visualization, PT; Writing – original draft, PT; Writing – review & editing, All authors.

340 Declaration of Interest Statement

341 The authors declare that the research was conducted in the absence of any commercial or financial
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485 Appendix

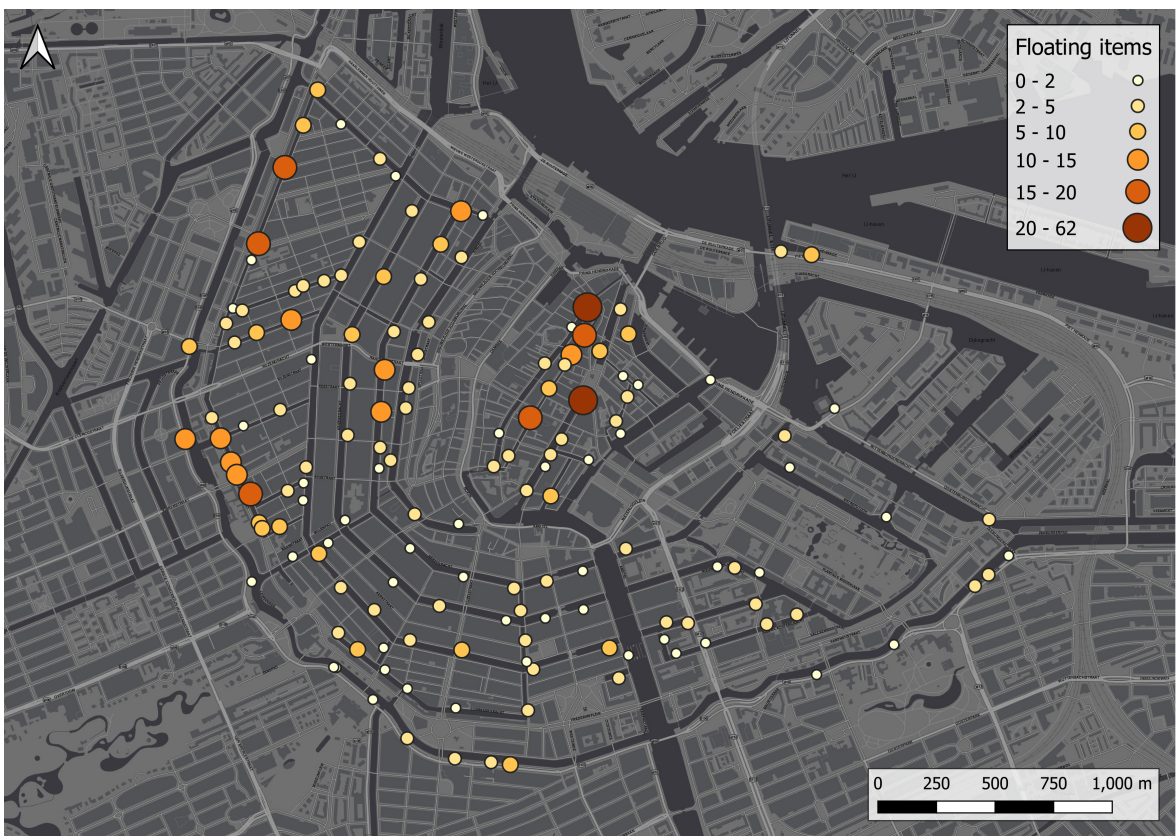


Figure 6: 150 locations within the historic city center of Amsterdam in which the abundance of plastic pollution was monitored using the River-OSPAR item categorization. The data sheet is available online (<https://doi.org/10.4121/6ee9946f-9ff5-4019-9d91-03e1c5283210>) (Tasseron et al., 2023b). Basemap: Esri, HERE, DeLorme, MapmyIndia.