

1 **Redox-informed models of global biogeochemical cycles**

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22 *This article is a non-peer-reviewed preprint submitted to EarthArXiv.*

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24
25 **Abstract**

26 Microbial activity mediates the global fluxes of carbon, oxygen, nitrogen, and other elements,
27 which include gases that impact the climate. However, in the global models of the marine and
28 terrestrial biospheres used for climate change projections, typically only photosynthetic
29 microbial activity is resolved mechanistically. To move forward, models need a theoretically
30 grounded framework with which to constrain parameterizations of diverse microbial
31 metabolisms. Here, we explain how the key redox chemistry underlying metabolisms provides
32 this framework. Using this first-principles approach, the presence or absence of each metabolism
33 emerges dynamically from ecological interactions, expanding model applicability to unobserved
34 environments.

35
36 “Nothing is less real than realism. It is only by selection, by elimination, by emphasis, that we
37 get at the real meaning of things.” – Georgia O’Keefe

38

39 **Overview**

40 Microorganisms are the “engines that drive Earth’s biogeochemical cycles”¹ (Fig. 1).
41 Photoautotrophic microorganisms are responsible for about half of CO₂ fixation and O₂
42 production on earth, and heterotrophic microorganisms are responsible for much of the return
43 reaction: the oxidation of organic matter back into CO₂. The temporal and spatial separation of
44 photoautotrophy and heterotrophy in the global environment drives the biological sequestration
45 of carbon, the reduction of atmospheric CO₂, and the maintenance of elevated atmospheric and
46 oceanic O₂²⁻⁵. Chemoautotrophic microorganisms also fix CO₂ and, together with anaerobic
47 heterotrophic metabolisms, carry out diverse chemical transformations including the fluxes of
48 nitrogen to and from biologically available states and the formation of the potent greenhouse gas
49 nitrous oxide (N₂O)^{6,7}. Since these transformations respond to, and feedback on, changes in
50 climate (Fig. 1), estimating microbial activity accurately at global scales is important for climate
51 science.

52 However, understanding and projecting the impacts of microbial processes are limited in
53 part due to oversimplified representation in earth system models. For example, in marine
54 biogeochemical models, much attention is given to the complex impacts of phytoplankton – the
55 photoautotrophic microorganisms responsible for primary production -- and their small
56 zooplankton predators⁸⁻¹⁰. The bacterial and archaeal activities responsible for other critical
57 aspects of biogeochemical cycling in the land and ocean – remineralization, denitrification,
58 nitrogen fixation, methanogenesis, etc. – are often crudely parameterized¹¹. Such models have
59 limited prognostic capability. For example, models typically prescribe the ecological niche of a
60 given metabolism with imposed, empirically determined parameters that are site- or organism-
61 specific. These parameterizations may or may not apply to other environments, including past
62 and future ecosystems.

63 These simplistic approaches have been largely necessary due to the difficulties of
64 characterizing the taxonomy and metabolic capabilities of natural microbial communities.
65 However, the rapid expansion of “-omics” sequencing capability has enabled a clearer view of
66 microbial biogeography and activity in the environment. In consequence, computational
67 biogeochemistry is opening up the ‘black box’ of remineralization and other microbially
68 mediated processes in marine and terrestrial environments¹²⁻²¹.

69 As we expand models to include the full metabolic potential of microorganisms, how can
70 we organize and reduce the complexity of the descriptions of metabolic diversity? Non-
71 photosynthetic organisms oxidize chemical species for energy, and thus their respiration is
72 biogeochemically significant²². Here, we explain how the key reduction-oxidation (redox)
73 reactions that supply energy for metabolisms can provide an additional organizing principle for
74 explicit descriptions of microbial populations in ecosystem models. This redox basis can be
75 exploited to quantitatively resolve chemical transformations in terms of assimilatory and
76 respiratory fluxes. While not yet incorporated into earth system models, this view has been
77 advocated for such applications^{23,24}, and has been embraced and employed in the field of
78 environmental biotechnology, such as in the interpretation and modeling of wastewater
79 bioreactors²⁵. Just as models of ocean and atmospheric circulation are constrained by
80 conservation of energy and potential vorticity, complementing mass balance with powerful redox
81 and energetic constraints enables self-consistent descriptions of diverse microbial metabolisms.

82 This approach aims to advance ecological modeling beyond species-specific descriptions
83 to those that matter for biogeochemical function, in line with trait-based modeling approaches¹⁰.
84 In analogy to the use of redox chemistry, trait-based functional type models of phytoplankton
85 have used cell size as an organizing principle – a “master trait” – for understanding
86 phytoplankton biogeography, biodiversity, and impact on biogeochemistry^{10,26–30}. These types of
87 theoretical constraints allow for including more functional types without introducing as many
88 degrees of freedom as would be necessary if each were empirically described. The guiding
89 perspective is that organizing complex biological behavior by its underlying chemical and
90 physical constraints gives more universally applicable descriptions of large-scale
91 biogeochemistry.

92 When incorporating a redox-balanced approach into ecosystem models, microbial
93 function emerges from underlying chemistry as a consequence of interactions between
94 populations modeled as ‘metabolic functional types’ and their environment. Resulting
95 theoretically grounded ecosystem models independently simulate microbial growth, respiration,
96 and abundances in ways that we can compare with observations such as sequencing datasets.
97 Thus, sequencing datasets are used as critical tests for the models, as external constraints rather
98 than as input to the models, allowing for an iterative relationship between theory, observations,
99 and models.

100 In contrast with empirically informed models, this approach involves constructing a
101 model of microbial activity theoretically, and then comparing the results with the observations in
102 order to gain an understanding of the system. The goal is to understand why biology functions as
103 it does, in addition to anticipating global impacts. From a first-principles biogeochemical
104 perspective with respect to physical and chemical forcing, genes are an intermediate step
105 between forcing and function, with the detailed complexity of biological reality following the
106 underlying chemical and physical constraints (analogous to the “form follows function” principle
107 of architect Louis Sullivan). This does not equate to thinking that biology (or genetic
108 information) does not matter or can be replaced entirely by physics and chemistry. Rather, we
109 want to fundamentally understand biological activity as an integrated part of an ecosystem, and
110 physics and chemistry become tools for doing so.

111 Here, we outline the basis for using redox chemistry as an organizing principle and its
112 translation into quantitative descriptions of microbial activity that are simple enough for global
113 earth system models. We then discuss the benefits of this approach in the context of their
114 implications for improved understanding and projections of global change impacts. Finally, we
115 discuss limitations and possible future developments.

116

117 **Predicting microbial activity**

118 From one perspective, microbial communities are characterized by interactions at the
119 micro-scale: gene expression, enzymatic capabilities, metabolites, species-specific
120 interdependencies, etc., as well as the physical and chemical environment surrounding small
121 cells^{31–35}. The information from sequencing in particular has allowed for a huge expansion of
122 insight into the detailed *in situ* activity of uncultivated species. When investigating global-scale
123 impacts, how do we decide which of these details may be bypassed for simplicity? Or, if this
124 simplification is impossible, must we incrementally construct a microbial ecosystem model that
125 incorporates all known micro-scale detail?

126 Another way forward arises from a “macro-scale” perspective, which examines how
127 ecosystem function relates to the chemical potential utilized by organisms for energy^{23,36,37}. For
128 example, it is well known that microbial communities in sediments and anoxic zones organizes
129 according to the ‘redox tower’— the ranking of half-reactions by electrochemical potential^{19,38,39}.
130 Furthermore, respiration by living organisms increases the entropy of the environment by

131 dissipating concentrated sources of chemical energy in accordance with the Second Law of
132 Thermodynamics^{37,40,41}.

133 This perspective suggests that chemical potential can be used to predict the activity of
134 microbial communities and their biogeochemical impact. However, given the notorious
135 complexity of microbial cells and systems⁴², which is many steps away from governing chemical
136 or physical equations, how can we be sure that this activity is indeed predictable? Frenzt et al.
137 2015 demonstrated that external conditions cause the seemingly random fluctuations observed in
138 microbial growth, rather than stochastic variation in gene expression⁴³. This provides direct
139 evidence of deterministic behavior, and so the authors conclude that microbial systems can in
140 principle be determined by macroscopic laws.

141 How is this determinism manifested? If microbial communities can respond relatively
142 quickly to changes to their local environment, they may predictably optimize the exploitation of
143 locally available resources. In the ocean, dispersal in microbes is thought to be a highly efficient
144 process such that microbial communities can in effect draw from an extensive seed bank^{44,45}, as
145 captured in the phrase “everything is everywhere, the environment selects”⁴⁶. Furthermore,
146 recent evidence also shows that gene acquisitions and deletions happen quickly enough to allow
147 for horizontal gene transfer to dominate bacterial adaptation⁴⁷⁻⁵¹, implying that evolution can
148 occur within few generations and thus on timescales similar to ecological interactions. Perhaps
149 consequentially, similar geochemical environments have been demonstrated to have high
150 microbial functional redundancy despite different taxonomic compositions^{16,52,53}. This may be
151 interpreted with the hypothesis that physics and chemistry selects for metabolic traits, and that
152 these traits can be housed in different organisms with taxonomic composition shaped by micro-
153 scale or biotic interactions^{16,52,54}.

154 The prediction of microbial activity from environmental chemical potential has a long
155 history in microbiology^{23,55-65}, and is conceptually similar to other redox-balanced approaches to
156 understanding microbial activity in sediments, soils, subsurfaces, and aquatic systems<sup>14,24,38,57,66-
157 70</sup>. Illustrating the power of these approaches, anticipating metabolism from chemical potential
158 resulted in a prediction that anaerobic ammonium oxidation (anammox) should exist decades
159 before it was observed^{71,72}. Quantitatively understanding microbially mediated rates of
160 conversion of substrates has practical implications for wastewater treatment, and thus the field of
161 biotechnology has established methodologies for an approach in textbook form²⁵. Flux balance

162 analysis (FBA) models can be considered as much more highly detailed analogues of this
163 approach that resolve the electron flow among a multitude of chemical reactions within a
164 cell^{73,74}.

165

166 **Redox-balanced metabolic functional types**

167 We can resolve microbial activity in global ecosystem models using the underlying redox
168 chemistry of diverse metabolisms as a constraint. One specific way forward is to model distinct
169 metabolisms as populations of metabolic functional types. This systematically quantifies relative
170 rates of substrate consumption, biomass synthesis, and excretions of transformed products
171 associated with each metabolism. Coupled with parameterizations of substrate uptake, this
172 replaces parameterizations of processes such as organic matter consumption, oxygen depletion,
173 and denitrification with electron-balanced respiratory fluxes of dynamic microbial populations.
174 Box 1 provides a detailed description of this methodology for multi-dimensional models.

175 A particular set of redox reactions may distinguish a functional type, such as the
176 oxidation of organic matter using oxygen (aerobic heterotrophy), or the oxidation of ammonium
177 or nitrite using oxygen (chemoautotrophic nitrification) as exemplified in Table 1. For each
178 metabolism, an electron-balanced description consists of multiple half-reactions: biomass
179 synthesis, oxidation of an electron donor, and reduction of an electron acceptor^{25,55,60}. The ratio
180 of anabolism and catabolism can then be represented by the fraction f of electrons fueling cell
181 synthesis vs. respiration for energy, following Rittman and McCarty 2001²⁵. This provides a
182 yield y (moles biomass synthesized per mole substrate utilized) of each required substrate that
183 reflects two inputs: electron fraction f and the coefficients of the half-reactions (Fig. 2). The
184 interlinked yields reflect the energy supplied by the redox reaction, the energy required for
185 synthesis and other cellular demands, and the inefficiencies of energy conversion. Either f or y
186 for any one of the substrates may be estimated theoretically with Gibbs free energies of
187 reactions²⁵ or with a combination of theoretical and empirical strategies⁷⁵.

188 The result is a stoichiometric budget of the metabolism of the whole organism (Table 1).
189 These descriptions quantify the elemental ratios of utilized substrates, biomass, and the excretion
190 of waste products. For example, the descriptions account for the CO₂ produced by heterotrophic
191 metabolisms as well as the CO₂ fixed by chemoautotrophic metabolisms (Table 1, Fig. 2, Fig.
192 S1), linking microbial activity directly to global carbon cycling.

193 To estimate the growth rate of each functional type, the yields from the metabolic
194 budgets are combined with the uptake rates of the required substrates (Box 1). Limiting uptake
195 rates can be estimated theoretically from diffusive supply, cell size, membrane physiology, and
196 other physical constraints⁷⁶⁻⁷⁸. Together, physical constraints on substrate acquisition and redox
197 chemical constraints on energy acquisition can provide an entirely theoretical estimate of the
198 growth of each metabolic functional type.

199 One strategy is to represent the populations carrying out each of these discrete
200 metabolisms as one functional type population, which aggregates the diverse community of
201 many species that are fueled by the same (or a similar) redox reaction (Fig. 3). Such aggregation
202 has been deemed a useful strategy for representing the biogeochemical impacts of microbial
203 communities for certain research questions^{79,80}. However, for other questions this wipes out
204 critical diversity among the aggregated populations. For example, diverse aerobic heterotrophic
205 populations consume organic matter over a wide range of rates, and these rates dictate the
206 amount of biologically sequestered carbon in the ocean. Redox chemistry and physical
207 limitations alone may not inform the heterogeneity among similar metabolisms. One additional
208 constraint is the limited capacity of the cell and thus its allocation of proteome towards different
209 functions⁸¹. While the electrons supplied to the cell must be conserved following the redox
210 balance, the electrons may be partitioned differently into machinery for substrate uptake vs.
211 biomass synthesis, for instance, for different phenotypes. This partitioning can be quantitatively
212 related to ecological fitness and biogeochemical impact via uptake kinetics, effective yields, and
213 other traits¹⁰ (Supplementary Note 1 and Supplementary Fig. 1).

214

215 **Benefits and implications for anticipating global change**

216 Redox chemistry aids in reducing the number of degrees of freedom in descriptions of
217 diverse microbial metabolisms. We next discuss the benefits of this electron-balanced approach,
218 each contextualized by specific projected impacts of global change due to microbial activity and
219 broad challenges in the fields of microbial ecology and biogeochemistry.

220

221 **1. Flexible and broadly applicable metabolic thresholds**

222 A key question for microbial biogeochemical studies, for which biogeochemical models
223 are primed to answer, is how the biogeographies of diverse, active metabolisms vary with

224 changes in the physical and chemical environment. What threshold determines the viability of a
225 given metabolism?

226 Redox-balanced metabolic budgets obviate the need to impose critical concentrations or
227 other thresholds that determine the presence of any given metabolism. Rather than being
228 imposed following empirical relationships, metabolic biogeography emerges dynamically from
229 ecological interactions and reflects environmental chemical potential. This flexibility aids in
230 understanding metabolic thresholds more fundamentally, and it expands model applicability to
231 diverse and unobserved environments. This is of particular importance for understanding global
232 change, since past and future worlds may include very different ecosystems that do not reflect
233 current empirical trends.

234 For example, the oceans are currently losing oxygen due to global warming^{82,83}. If anoxic
235 zones continue to expand, this will increase the habitat of anaerobic microorganisms, whose
236 respiration results in emissions of N₂ and N₂O to the atmosphere⁸⁴. Many biogeochemical
237 models prescribe O₂ concentrations that inhibit anaerobic activity in accordance with
238 observations of specific organisms or communities in experimental conditions. This assumes that
239 the same O₂ concentrations limit metabolism similarly in all environments, and often trades
240 mechanistic understanding of oxygen limitation for empirical correlations that may reflect a
241 variety of natural and introduced biases, such as microscale heterogeneity, physical mixing in the
242 ocean, and experimental bottle effects.

243 In contrast, a metabolic functional type model does not require imposed oxygen threshold
244 concentrations (Supplementary Fig. 3). When oxygen supply is abundant, anaerobic types are
245 competitively excluded because growth using alternative electron acceptors is lower than with
246 oxygen. When oxygen supply is low, aerobic populations may persist and continue to deplete
247 any available oxygen even as their growth is limited by oxygen, allowing for a steady state stable
248 coexistence of aerobic and anaerobic metabolisms, which is consistent with a variety of
249 observations⁸⁵.

250 Descriptions of microbial growth that reflect underlying chemical potential can enable
251 predictions of many other metabolic transitions, such as nitrogen fixation, nitrification, and the
252 transition to sulfur oxidation and reduction^{14,17,21,86,87}. As another example, this approach predicts
253 the restriction of nitrification from the sunlit surface ocean as a consequence of competitive
254 exclusion by phytoplankton in many environments (Fig. 4, Supplementary Fig. 2), as well as

255 active nitrification in some surface locations where phytoplankton are limited by another factor
256 not affecting the chemoautotrophs, such as at high latitudes where phytoplankton are limited by
257 light¹⁷. The emergent exclusion from most of the surface ocean anticipates that many clades of
258 nitrifying microorganisms have adapted to long-term exclusion from the surface and
259 consequentially lost (or did not develop) photoprotective cellular machinery.

260

261 **2. Replacing empirical descriptions of organic matter remineralization**

262 The fate of organic matter dictates the amount of carbon sequestered in the marine and
263 terrestrial biospheres. Microbial consumption mediates the carbon stored in soils, the carbon
264 stored in the ocean as dissolved organic matter (DOM), and the sinking flux of organic carbon
265 that constitutes the marine “biological carbon pump⁴,” without which atmospheric CO₂ would be
266 100-200 ppm higher than current levels. We want to understand how these carbon reservoirs
267 respond to changes in climate, such as increased temperatures and changes in precipitation
268 patterns. However, in biogeochemical models, simple rate constants often dictate the
269 remineralization of elements from organic back into inorganic constituents. In many global
270 ocean models, the biological pump has been modeled following empirical relationships with a
271 power law (the “Martin curve”⁸⁸).

272 Replacing simplistic parameterizations with dynamic metabolic functional types means
273 that electron-balanced descriptions of growth and respiration instead drive the fate of organic
274 matter in earth system models (Fig. 3). In addition to a more sophisticated and responsive
275 description of carbon sequestration, non-living organic matter is fully integrated into ecosystem
276 frameworks, enabling theoretical studies of phytoplankton-bacteria interactions to complement
277 observational and experimental approaches.

278

279 **3. Relationships between abundances, rates, nutrient concentrations, and elemental ratios**

280 An overarching puzzle challenging microbial ecology is to understand how chemical
281 transformations in the environment are set by the ecological interactions at the organism level,
282 among individual microscopic cells. It is clear that abundances of populations are not simply and
283 directly correlated with biogeochemical impact (i.e., higher abundance does not necessarily
284 imply an associated higher rate of chemical transformation). Untangling the relationship between

285 abundances and biogeochemical function is also necessary for interpretation of genetic evidence
286 that provide insight into this complex ecosystem structure.

287 Redox-balanced metabolic functional type modeling links rates of biomass synthesis
288 associated with a particular metabolism to its rate of respiration as well as the standing stock of
289 limiting nutrients. Because functional type modeling is coupled with estimates of population loss
290 rates due to grazing, viral lysis, or other mortality, simulations also resolve the standing stocks of
291 functional biomass. This quantifies the relationship between biomass concentrations and
292 volumetric rates of chemical transformations, emphasizing how relatively low biomass may be
293 associated with relatively high bulk rates⁸⁵.

294 For example, the approach has revealed a clear example of the signature of chemical
295 potential in the ecology of marine nitrification¹⁷ (Fig. 4). In this model, the two steps of
296 nitrification are represented by two functional type populations. This predicts about a three-fold
297 difference in the abundances of the organisms responsible for each of the two steps of
298 nitrification, despite the fact that the two populations carry out the same rate of subsurface N-
299 cycling at steady state^{6,17}. A three-fold or greater difference in abundance and associated
300 ammonia (NH_4^+) and nitrite (NO_2^-) concentrations is consistent with observed differences^{17,89},
301 and it reflects that the oxidation of one mole of NH_4^+ generates three times more electrons than
302 the oxidation of one mole of NO_2^- , with differences in cell size further contributing to differences
303 in abundances (Fig. 4, Supplementary Fig. 2). Recent observations confirm the redox-based
304 difference in NH_4^+ and NO_2^- biomass yield^{90,91}, although measured rates from a nonsteady
305 environment suggest that NO_2^- -oxidizing bacteria can partition electrons more efficiently than
306 NH_4^+ -oxidizing archaea⁹⁰ (i.e. higher fraction f despite lower yield y ; see Supplementary Note 2).

307 Because redox-based descriptions resolve the stoichiometry of whole organism
308 metabolism, they also link together elemental cycles. Explicit description of relative elemental
309 flow through the ecosystem, and specifically their variation from average values, is critical for
310 understanding climate-biogeochemical feedbacks⁹²⁻⁹⁵. For example, the nitrification model also
311 estimates the CO_2 fixation rates associated with nitrification rates (Supplementary Fig. 2),
312 enabling global-scale, electron-balanced projections of the amount of carbon converted to
313 organic form by chemoautotrophic nitrifying microorganisms.

314

315 **4. Connections with sequencing datasets**

316 How do we relate metabolic functional type models to sequencing datasets measuring
317 genetic, transcriptomic, and proteomic diversity? Connecting biogeochemical models with
318 sequencing data is critical because this data provides an enormous amount of information about
319 ecosystem structure and function. Genes (or transcripts) themselves are not necessarily the most
320 concise or useful currency given functional redundancies, unattributed function, and variation in
321 gene dosage from horizontal gene transfer as well as growth rate⁴⁸. Recent “gene-centric” models
322 aim to resolve the abundances of key genes as proxies for a predetermined set of metabolic
323 pathways^{14,16,19}. However, the parameters used to describe metabolic pathways in these models
324 are estimated similarly to the redox-balanced yields and efficiencies described here.

325 The innovation of gene-centric models is the sophisticated conversion of estimates of
326 biogeochemical activity and biomass to genes. For example, the model of Coles et al. 2017
327 resolves biomass and nutrient concentrations prognostically, and then uses a three-part formula –
328 representing constitutive, regulated, and steady state transcription -- to diagnostically calculate
329 transcription rates from modeled biomass and growth rates¹⁶. Thus, the two types of modeling
330 are complimentary, with redox chemistry providing estimates of metabolic activity from
331 fundamental principles, and the careful calibrations between activity and sequencing providing a
332 comparative metric.

333 The examples here externalize the conversion between modeled activity and sequencing
334 information as a transparent process. In Fig. 4, the predicted functional biomass of ammonia
335 oxidizing population is related to archaeal Marine Group I (MGI) and *Nitrospina*-like 16S rRNA
336 genes with two conversion factors: the cell elemental quota (fmol N cell⁻¹) and the number of
337 cellular gene copies. Conversion error arises since cell mass and size vary with growth rate^{96,97}.
338 Maintaining transparency of the conversion from predicted microbial activity to genes and
339 transcripts allows interdisciplinary audiences to understand and critique the models.

340

341 **Limitations and possible extensions**

342 Using chemical potential as a theoretically grounding organizing principle for the
343 resolution of diverse metabolisms can greatly improve microbial descriptions in global
344 biogeochemical models. However, the approach does have its limitations, which generally
345 increase in significance with increased temporal or spatial resolution.

346 The proposed modeling approach relies on estimates of the limiting uptake rates of
347 required substrates. Uptake kinetics are complex, but for the limiting resource, encounter
348 effectively controls the uptake, and the physics of encounter has been relatively well described.
349 For example, uptake rates estimated from diffusive supply of substrate, cellular geometry, and
350 membrane physiology^{76–78} have been empirically supported⁹⁸.

351 Modeling metabolic diversity with functional type populations requires choosing how
352 metabolisms are distributed among the populations. This has consequences when interpreting
353 time-varying states: model solutions become dependent on the partitioning of metabolism among
354 the functional types as the timescales of physical change approach the timescales of microbial
355 growth (see Supplementary Note 3, Supplementary Fig. 3, and Supplementary Fig. 4 for a
356 detailed example). Other species-specific time-varying phenomena such as the lag response of
357 organisms to substrate availability also becomes relevant⁹⁹. On one hand, this is beneficial for
358 resolution of microbial processes in fine-grained ocean circulation models where flow can vary
359 on the order of days. However, incorporating another constraint, such as proteome allocation⁸¹, is
360 necessary to inform these choices. For example, considering enzymatic allocation in combination
361 with energetics allowed for the prediction of both the division of nitrification into a two-step
362 process in mixed-environments and the combined, complete pathway in one organism
363 (“comammox”) in biofilms, which preceded observations of the latter^{100–102}.

364 Uncertainty in distributions of metabolism lies not only in the length of a metabolic
365 pathway, but also in the degree of metabolic versatility (metabolic “mixotrophy”). Such
366 versatility characterizes key players in large-scale biogeochemistry, such as nitrite-oxidizing
367 bacteria and photoheterotrophs^{103–105}. Mixotrophic lifestyles can increase the fitness of
368 populations in their environments, impacting overall ecosystem function¹⁰⁶. In one sense, the
369 approach here provides a prediction of where we might expect such mixotrophy by resolving
370 stable co-existences of diverse metabolisms. In Fig. 3, for example, syntrophic co-existence
371 occurs at depth among heterotrophs, ammonia-oxidizers, and nitrite-oxidizers, and future work
372 could investigate what determines which combinations of these co-existences remain as ‘passive’
373 interactions, which develop into mutualistic dependencies as ‘active’ interactions¹⁰⁷, and which
374 evolve into mixotrophic phenotypes or endosymbionts. Additionally, by considering the potential
375 to carry out a metabolism as a trait, we can use the current framework along with an additional
376 constraint to investigate implications of metabolic mixotrophy. For example, Coles et al. 2017

377 impose a tradeoff between the degree of metabolic diversity of a single functional type and
378 growth rate, enabling the exploration the consequences of distribution of metabolism on the
379 biogeochemical state¹⁶.

380 Also, the metabolic functional type approach resolves only active functional biomass,
381 while evidence suggests that less than 10% to more than 75% of the microbial community may
382 be inactive¹⁰⁸. Some seemingly inactive populations may slowly metabolize over long
383 timescales, requiring longer model integration times and careful attention to their loss rates for
384 resolution, while some populations are periodically active as revealed by high-resolution
385 observations in time⁴⁵.

386 In Fig. 3, the electron-balanced description consists of an average stoichiometry and
387 electron fraction for one sinking pool of organic matter in the ocean, which, as mentioned above,
388 is not sufficient to accurately resolve the carbon storage that is shaped by a distribution of rates.
389 An energetics-based perspective can serve as a tool for further deciphering organic matter
390 complexity. This could be used to replace the phenomenological description of organic matter in
391 models, as “labile” vs. “non-labile,” for example, with a more mechanistic underpinning. For
392 example, organic matter may be organized by the nominal oxidation state of its carbon atoms,
393 which relates to a measure of free energy and accessibility^{62,109,110}. Descriptions also require
394 attention to the physical dynamics of sinking particles, soil chemistry and physics, and the
395 complex interactions between organic substrates and their bacterial consumers and enzymatic
396 activity^{12,111–113}.

397 Descriptions of phytoplankton are currently much more sophisticated than of bacteria and
398 archaea in models, reflecting a longer history of comprehensive sets of observations. However,
399 further work could develop simple descriptions of photoautotrophic metabolisms from
400 underlying energetics by connecting the supply of photons to available energy for biosynthesis
401 within the cell. Many biogeochemical models account for an inefficiency of phytoplankton
402 metabolism with a parameter that dictates their excretions of dissolved organic matter¹¹⁴.
403 Incorporating this excretion into an energetic framework would enhance studies of
404 phytoplankton ecology, such as studies of photoautotrophic-heterotrophic interactions in the
405 ocean surface or photoautotrophic-chemoautotrophic interactions at the base of the euphotic zone
406 where some phytoplankton excrete nitrite due to incomplete reduction of nitrate^{115,116}.

407 As a more radical extension, can we progress past population modeling and model
408 microbial consortia as one aggregate community biomass^{41,117-120}? This may improve resolution
409 of time-varying metabolic versatility. However, if both steps of nitrification were a part of such a
410 consortium, would the characteristic accumulation of nitrite be predicted (Supplementary Fig.
411 2)? We leave these questions for future research and conclude that the best choice for the
412 “resolution” of metabolism will depend on the specific research question and the available
413 observations.

414 We have described a useful approach for understanding and anticipating microbial
415 control of biogeochemical cycling that is suitable for global applications. The approach aims to
416 represent microbial growth and respiration explicitly and consistently from knowledge of
417 chemical gradients in the environment, towards a goal of building an independently constructed
418 theoretical ecosystem model that can then be compared to observations. Describing microbial
419 communities with underlying energetic constraints connects metabolisms dynamically with
420 global geochemical distributions, such as carbon dioxide, oxygen, and biologically available
421 nitrogen. This deepens our understanding of microbial ecosystems and enables the incorporation
422 of the feedbacks of microbial activity to changes in global biogeochemistry and the climate
423 system.

424

425 **Acknowledgements**

426 We thank Victoria Coles for constructive comments and suggestions during the review process,
427 Alyson Santoro for providing data, and Deepa Rao for providing the introductory quote. We are
428 grateful for helpful feedback from Andreas Richter, Victoria Orphan, and Naomi Levine. We
429 appreciate discussions over the past years with Amala Mahadevan, Dan Repeta, Penny
430 Chisholm, Joe Vallino, and Terry Hwa that have influenced this perspective. EJZ was supported
431 by the Simons Foundation (Postdoctoral Fellowship in Marine Microbial Ecology). MFP was
432 supported by a grant from the Simons Foundation (LIFE ID 572792), MJF was supported by the
433 Simons Foundation: the Simons Collaboration on Ocean Processes and Ecology (SCOPE
434 #329108) and the Simons Collaboration on Computational Biogeochemical Modeling of Marine
435 Ecosystems (CBIOMES #549931).

436 **Author Contributions**

437 E. J. Z. wrote the manuscript. All authors contributed to revising and editing the manuscript.

438 **Competing Interests**

439 The authors declare no competing interests in relation to this work.

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Box 1: Incorporating metabolic functional types into ecosystem models

A metabolic functional type can be represented as a population with a growth rate that is limited or co-limited by multiple required substrates. If Liebig's Law of the Minimum is employed, the limiting growth rate μ is described as

$$\mu = \min(V_i y_i) \quad (1)$$

where V_i is the specific uptake rate of substrate i , and yield y_i is the biomass yield with respect to that substrate. Yields for the different substrates and elements are interlinked in the metabolic budget derived from the underlying redox chemistry. Yields reflect Gibbs free energies of reaction among other factors. In the simplest model, non-limiting substrates are consumed in proportion to the limiting resource according to the metabolic budget, although in reality they may accumulate in the form of storage molecules.

Each metabolic functional type population can be incorporated into a multi-dimensional environmental model (e.g. an ocean simulation) with physical transport as

$$\frac{dB}{dt} = \mu B - L(B)B - \underbrace{\nabla \cdot (\mathbf{u}B)}_{\text{advection}} + \underbrace{\nabla \cdot (\boldsymbol{\kappa} \nabla B)}_{\text{diffusion}} \quad (2)$$

for biomass concentration B , loss rate L , velocity \mathbf{u} , and diffusion coefficient $\boldsymbol{\kappa}$. The loss rate function represents a combination of processes including predation, viral lysis, maintenance, and senescence, which remain largely unconstrained, although efforts have been made to relate losses to ecological dynamics^{121–125}.

The yield partitions the amount of substrate taken up by the population into that used for growth, $V_i y_i$, versus that exiting the cell in modified form as a waste product, $V_i(1 - y_i)$ (Fig. 2; Table 1). Eqn. 1 suggests a correlation between μ and y , but yields may be further modified by other factors. For example, accounting for maintenance energy decreases the ratio of growth to respiration, contributing to a decoupling between growth rate and yield particularly at low growth rates¹²⁶. Furthermore, a trade-off between uptake rate and yield at the cellular level reflects the allocation of enzyme towards machinery for substrate uptake vs. biomass synthesis, among other factors. Considering a proteome constraint can incorporate this trade-off (Supplementary Note 1).

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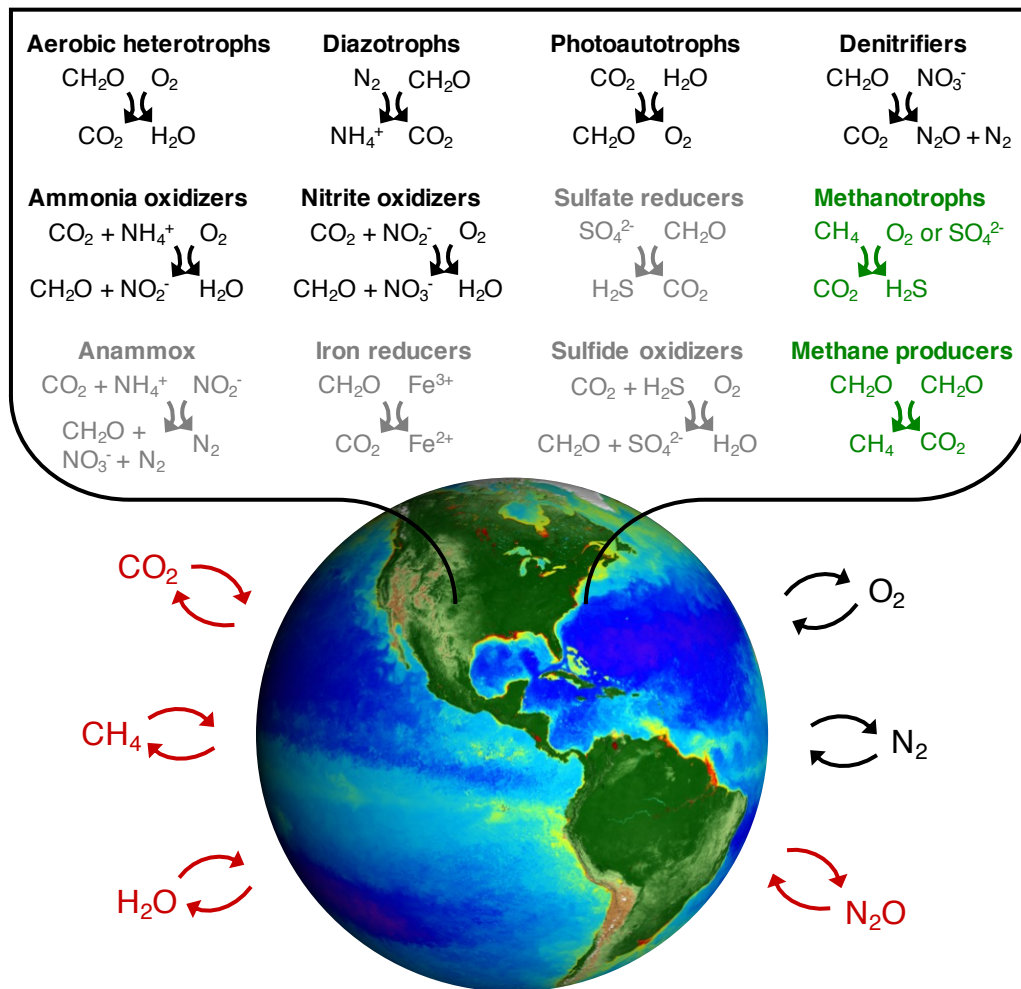
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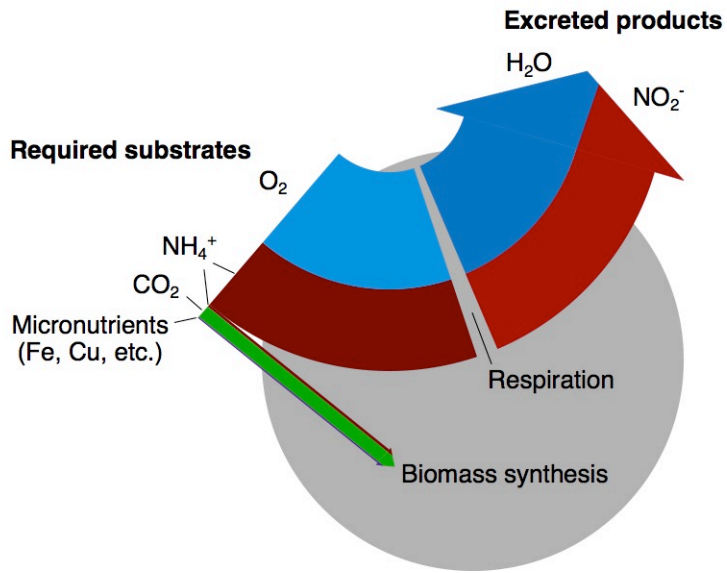
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 756 **Figure 1:** Key microbially driven redox transformations that mediate the atmospheric fluxes of
 757 climatically relevant gases. Radiatively active gases are notated with red type. The processes in
 758 black type are represented in some way (though not necessarily with electron balancing) in both
 759 the marine and terrestrial biosphere in earth system models within the Coupled Model
 760 Intercomparison Project (land: NCAR Community Earth System Model; ocean: GFDL
 761 COBALTv2¹²⁷) which are used for projections of climate change in reports by the
 762 Intergovernmental Panel on Climate Change. Processes in green type are represented in only the
 763 terrestrial model. Current models do not yet include other relevant reactions, some of which are
 764 represented in grey type, such as anaerobic ammonia oxidation (anammox), the marine
 765 production and consumption of methane, the redox cycling of iron, manganese, and other metals,
 766 and the methane-relevant redox chemistry of phosphorus¹²⁸. COBALTv2 does account for
 767 sulfate reduction in marine sediments, but sulfate is not represented. Image courtesy of NASA.
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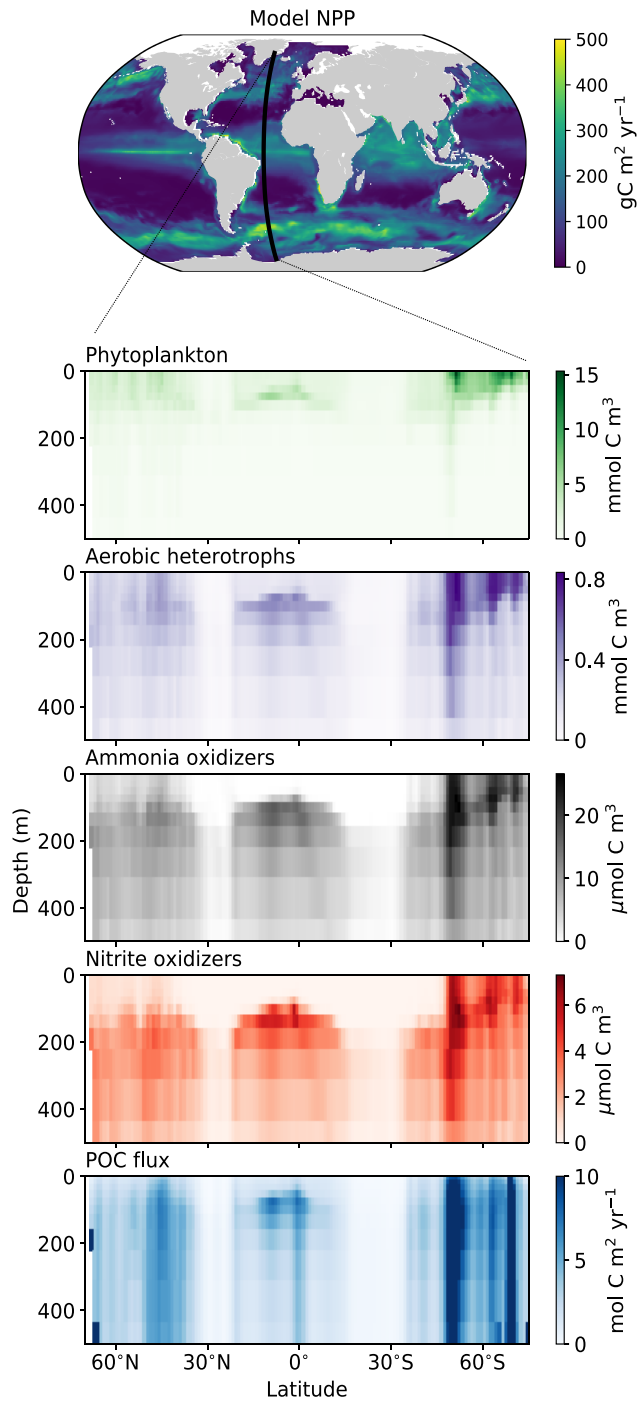


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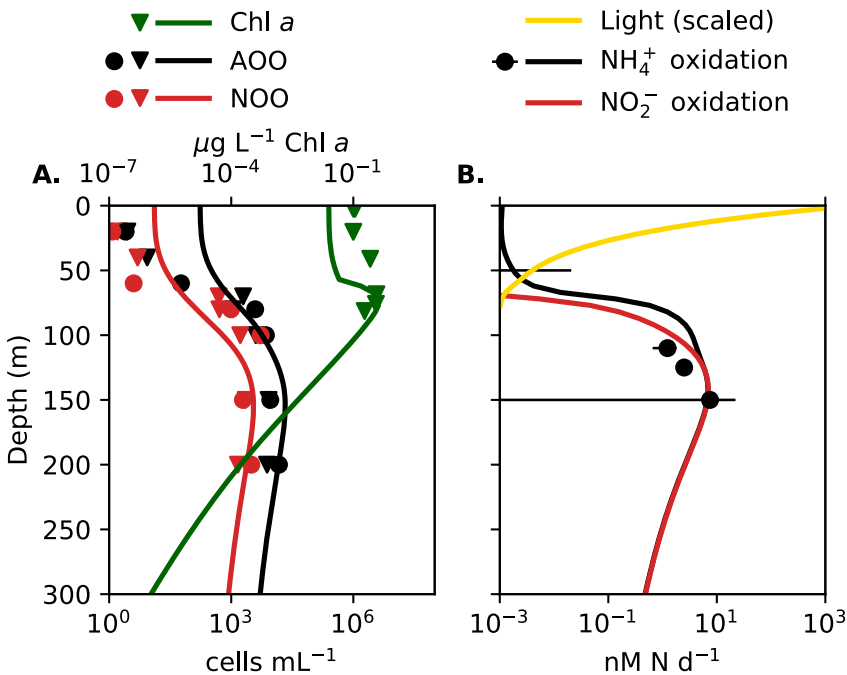
772 **Figure 2:** Schematic of a single cell represented as a metabolic functional type carrying out the
773 aerobic oxidation of ammonia (NH_4^+ ; Table 1). The redox balance informs the elemental ratios
774 of substrates utilized, biomass synthesized, and waste products excreted.

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777 **Figure 3.** Simulation of net primary productivity (NPP), the biomasses of metabolic functional
 778 types, and the sinking particulate organic carbon (POC) flux along a transect of a global microbial
 779 ecosystem model coupled with an estimate of the ocean circulation (Darwin-MITgcm¹⁷).



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781 **Figure 4: Model simulation and observations of the marine nitrification system.**

782 Biogeochemistry is driven by microbial metabolic functional types in a vertical water column
 783 model¹⁷. Lines are model solutions, and marked points are observations from two stations in the
 784 Pacific Ocean^{89,129} (see Fig. S1 for more detail) **A.** Chlorophyll *a* concentrations and abundances
 785 of ammonia-oxidizing organisms (AOO) and nitrite-oxidizing organisms (NOO). Observed
 786 abundances are of the 16S rRNA abundances of archaeal Marine Group I and *Nitrospina*-like
 787 bacteria^{89,129}. Model abundances are converted from biomass with 0.1 fmol N cell⁻¹ for AOO, 0.2
 788 fmol N cell⁻¹ for NOO, and one gene copy per cell. **B.** Solar irradiance (“light”) and bulk
 789 nitrification rates.

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794 **Table 1: Simplified equations describing two exemplary metabolic functional types.** For
 795 each, half-reactions combine to form the catabolic and anabolic full reactions²⁵: the oxidation of
 796 an electron donor (R_D ; here either organic matter or ammonium), the reduction of an electron
 797 acceptor (R_E), and biomass synthesis (R_S). The total reaction (R_T) sums each of these three
 798 multiplied by a factor of f , the fraction of electrons partitioned into the synthesis reaction vs.
 799 respiration. Denominator d represents the number of electron equivalents that correspond to the
 800 oxidation states of the inorganic constituents of that synthesis: with a microbial biomass
 801 composition of $C_5H_7O_2N$, $d_B = 4(5) + 1(7) - 2(2) - 3(1) = 20$. Organic matter oxidation and
 802 synthesis equations are written without H_2O on the left- and right-hand side, respectively, for
 803 conciseness. Example whole organism metabolic budgets are calculated using the listed example
 804 efficiencies for marine organisms, an average marine organic substrate composition¹³⁰ of
 805 $C_{6.6}H_{10.9}O_{2.6}N$, and the above biomass composition.

	Aerobic heterotroph	Ammonia-oxidizing chemoautotroph
R_D (1)	$\frac{1}{d_D} C_{c_D} H_{h_D} O_{o_D} N_{n_D} \rightarrow \frac{n_D}{d_D} NH_4 + \frac{c_D}{d_D} CO_2 + H^+ + e^-$	$\frac{1}{6} NH_4^+ + \frac{1}{3} H_2O \rightarrow \frac{1}{6} NO_2^- + \frac{4}{3} H^+ + e^-$
R_E (1-f)	$\frac{1}{4} O_2 + H^+ + e^- \rightarrow \frac{1}{2} H_2O$	$\frac{1}{4} O_2 + H^+ + e^- \rightarrow \frac{1}{2} H_2O$
R_S (f)	$\frac{n_B}{d_B} NH_4 + \frac{c_B}{d_B} CO_2 + H^+ + e^- \rightarrow \frac{1}{d_B} C_{c_B} H_{h_B} O_{o_B} N_{n_B}$	$\frac{n_B}{d_B} NH_4 + \frac{c_B}{d_B} CO_2 + H^+ + e^- \rightarrow \frac{1}{d_B} C_{c_B} H_{h_B} O_{o_B} N_{n_B}$
R_T	$\frac{1}{d_D} C_{c_D} H_{h_D} O_{o_D} N_{n_D} + \frac{(1-f)}{4} O_2$ $\rightarrow \frac{f}{d_B} B + \left(\frac{n_D}{d_D} - \frac{n_B f}{d_B} \right) NH_4 + \left(\frac{c_D}{d_D} - \frac{c_B f}{d_B} \right) CO_2$	$\left(\frac{1}{6} + \frac{f}{d_B} \right) NH_4 + \frac{c_B f}{d_B} CO_2 + \frac{(1-f)}{4} O_2$ $\rightarrow \frac{f}{d_B} B + \frac{1}{6} NO_2$
e^- donor yield	$y_D = f \frac{d_D}{d_B} \approx f$	$y_{NH_4^+} = \left(1 + \frac{d_B}{6f} \right)^{-1} \approx \frac{6f}{d_B}$
Example efficiency	$f = 0.1 - 0.2$; Marine bacteria (Robinson 2008)	$f = 0.02 - 0.04$; Marine archaea (Zakem et al. 2018)
Example budget	$7.1 C_{6.6} H_{10.9} O_{2.6} N + 47 O_2 \rightarrow B + 6.1 NH_4 + 42 CO_2$	$112 NH_4 + 5 CO_2 + 162 O_2 \rightarrow B + 111 NO_2$

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