

1 Integrating the water planetary boundary with water 2 management from local to global scales

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28 Abstract

29 The planetary boundaries framework defines the ‘safe operating space for humanity’ in terms of nine
30 important global processes influenced by humans which can destabilize the Earth System if perturbed.
31 The planetary boundary for freshwater use attempts to provide a global limit to anthropogenic water
32 cycle modifications, but it has been challenging to translate and apply it to the regional and local scales
33 at which societally relevant water problems and management typically occur. We develop an integrative
34 cross-scale approach considering how the water planetary boundary could help guide sustainable water
35 management and governance at different sub-global contexts defined by physical features (e.g.
36 watershed or aquifer), political borders (e.g. city, nation, or group of nations), or commercial entities
37 (e.g. a corporation, industry or trade group, financial institution). The integration of the water planetary
38 boundary at these sub-global contexts occurs via two approaches: (i) calculating *fair shares*, in which
39 local water cycle modifications are compared to that context’s allocation of the global safe operating
40 space, taking into account biophysical, socio-economic, and ethical considerations; and (ii) defining a
41 *local safe operating space*, in which interactions between water stores and Earth System components
42 are used to define local boundaries required for sustaining the local water system in stable conditions,
43 which we demonstrate with a case study of the Ciénaga Grande de Santa Marta wetland complex in
44 Colombia. By harmonizing these two approaches, the water planetary boundary can be used in multiple
45 contexts to ensure that water cycle modifications remain within both local and global boundaries. This
46 provides a broadly applicable method by which the planetary boundaries framework complements
47 existing water management and governance approaches.

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49 1. Local water resources and Earth System stability

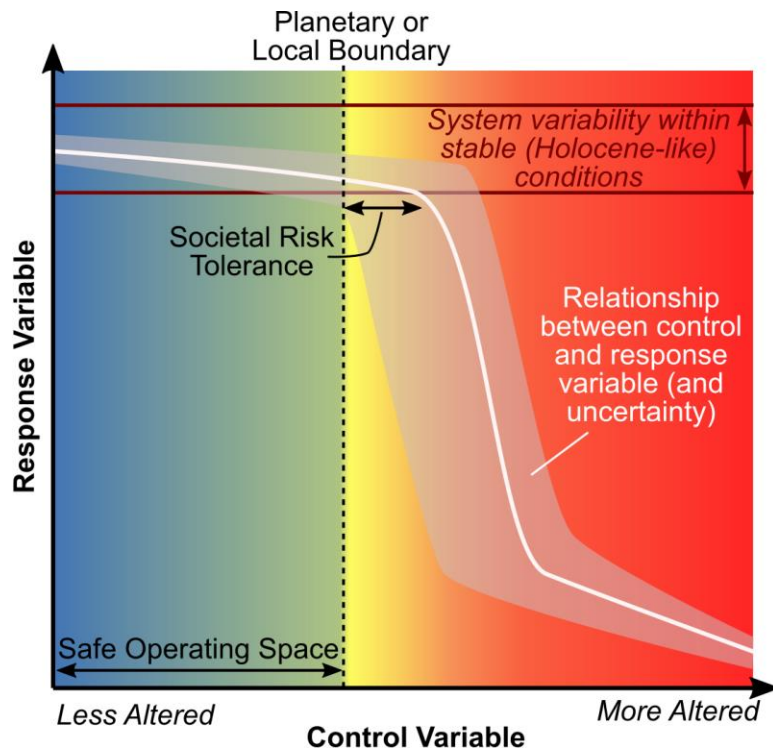
50 Water, the 'bloodstream of the biosphere' (Ripl, 2003), is fundamental to Earth System functioning and
51 human society. Due to the central role of water for maintaining global biosphere integrity, regulating
52 climate, and mediating carbon and nutrient cycling, changes to the water cycle can propagate through
53 the Earth System and disrupt processes interacting across numerous scales. For example, land use
54 change in one setting can alter evapotranspiration and lead to precipitation change downwind (Wang-
55 Erlandsson et al., 2018; Zipper et al., 2019a). Self-amplifying land-water interactions mean that
56 deforestation may lead to regional forest die-back in areas such as the Amazon (Zemp et al., 2017) with
57 potential cascading impacts on Earth System stability as a whole (Steffen et al., 2018). In addition to
58 physical processes, socio-economic factors external to a watershed can impact local hydrological
59 conditions: agriculture, by far the largest user of freshwater, is driven by global socio-economic
60 decisions as crops are shipped all over the world (Dalin et al., 2017; Hoekstra & Mekonnen, 2012;
61 Jaramillo & Destouni, 2015; Marston et al., 2015). Anthropogenic climate change, a global challenge,
62 also has diverse impacts on local water systems (Cook et al., 2018; Pfahl et al., 2017). In other words, the
63 local water cycle is shaped by global processes and modifications to local hydrological processes can
64 have global consequences.

65 This emerging, integrated understanding of interconnections between local and global water systems
66 requires integrated management and governance strategies across scales (Biermann et al., 2012; Keys et
67 al., 2017; Sivapalan et al., 2014). However, developing generalizable understanding of the
68 spatiotemporal scales spanned by the water cycle has been a longstanding challenge in hydrology, water
69 management, and at their intersection (Blöschl et al., 2019; Blöschl & Sivapalan, 1995; Daniell &
70 Barreteau, 2014; Klemeš, 1983). In particular, recent synthetic work has identified translating
71 understanding of coupled human and natural systems across scales as a key future research priority to
72 provide management-relevant science (Konar et al., 2016; Kramer et al., 2017). While socio-hydrology
73 has been suggested as a potential tool to bridge the gaps between watershed-scale and global-scale
74 water management (Di Baldassarre et al., 2019; Konar et al., 2016), specific approaches for integrating
75 global water sustainability targets with local water management remain lacking.

76 The planetary boundaries framework, introduced by Rockström et al. (2009) and further elaborated by
77 Steffen et al. (2015), offers one approach to bring a global perspective to local water management
78 (Konar et al., 2016). This framework identifies nine boundaries representing critical Earth System
79 processes that characterise deviation from the relatively stable environmental conditions of the

80 Holocene, the epoch during which the world's societies have developed. For most of these processes, a
81 quantifiable *control variable* has been suggested, which may cause some *response variable* to depart
82 from relatively stable Holocene-like conditions if sufficiently perturbed, either alone or through
83 interactions with other Earth System processes (Rockström et al., 2009; Steffen et al., 2015). For
84 effective boundary-setting, the control variable should be both quantifiable and subject to influence by
85 human actions, while the response variable should describe Earth's stable conditions and be influenced
86 by the control variable (Gleeson et al., 2019b). For safeguarding each planetary boundary the control
87 variable is set some distance upstream from departure of the response variable from stable conditions,
88 typically at the lower end of uncertainty due to systemic and/or scientific factors (Figure 1). The secure
89 area bounded by the nine planetary boundaries has been referred to as a "safe operating space for
90 humanity" (Rockström et al., 2009) in that it describes the Holocene-like Earth System conditions, which
91 so far are the only ones known in which human civilization has thrived and in which the risk of
92 destabilizing global dynamics is low.

93 Regarding water - our focus here - the planetary boundary for freshwater use has originally been based
94 on the amount of freshwater needed to maintain rivers' environmental flow requirements (Gerten et al.,
95 2013; Steffen et al., 2015). Recently, Gleeson et al. (2019a, 2019b) proposed a more holistic
96 representation of the various functions of water in maintaining Earth system stability, distinguishing six
97 sub-boundaries relating to the five major stores of freshwater (Table 1). Environmental flow
98 requirements are retained in this approach, together with new water planetary sub-boundaries for
99 frozen water, groundwater, soil moisture, surface water, and two sub-boundaries for different aspects
100 of atmospheric water. Gleeson et al. (2019a, 2019b) suggested potential control and response variables
101 for these new sub-boundaries (Table 1) meant to represent the primary function of water in the Earth
102 System, but a significant program of cross-disciplinary collaborative work remains to select and evaluate
103 appropriate variables and develop suitable methods to aggregate from distributed to global values. To
104 provide a sound societal relevance for these efforts, it is necessary to first determine whether the water
105 planetary boundary can be meaningfully integrated with existing management approaches.



106

107 Figure 1. A planetary or local boundary (dashed line) is set where the system shifts from stable to
 108 possibly destabilized conditions in response to change in the control variable. Based on (Gleeson et al.,
 109 2019b; Rockström et al., 2009; Steffen et al., 2015). A precautionary approach takes into account the
 110 system variability of the response variable (dark red horizontal lines), scientific and systemic uncertainty
 111 about the relationship between the control and response variables (grey zoning around the white line)
 112 and societal tolerance of risk. The relationship between the control and response variable shown here is
 113 just one possible relationship, and these relationships are not necessarily threshold-type or even
 114 monotonic.

115

116 Since there is no global governance nor common organizational institutions for planetary-scale water
 117 management and governance (Biermann et al., 2012), the water planetary boundary needs to be
 118 translated to the local and regional scales where water management and governance typically operate
 119 (CISL, 2019; Downing et al., In review; Konar et al., 2016). In this study, we address three questions
 120 necessary to integrate local water management and governance with global water sustainability:

121 (i) How can global-scale values be meaningfully disaggregated to the diverse spatial scales at which
 122 water management and governance occurs such as watersheds, nations, commercial entities?

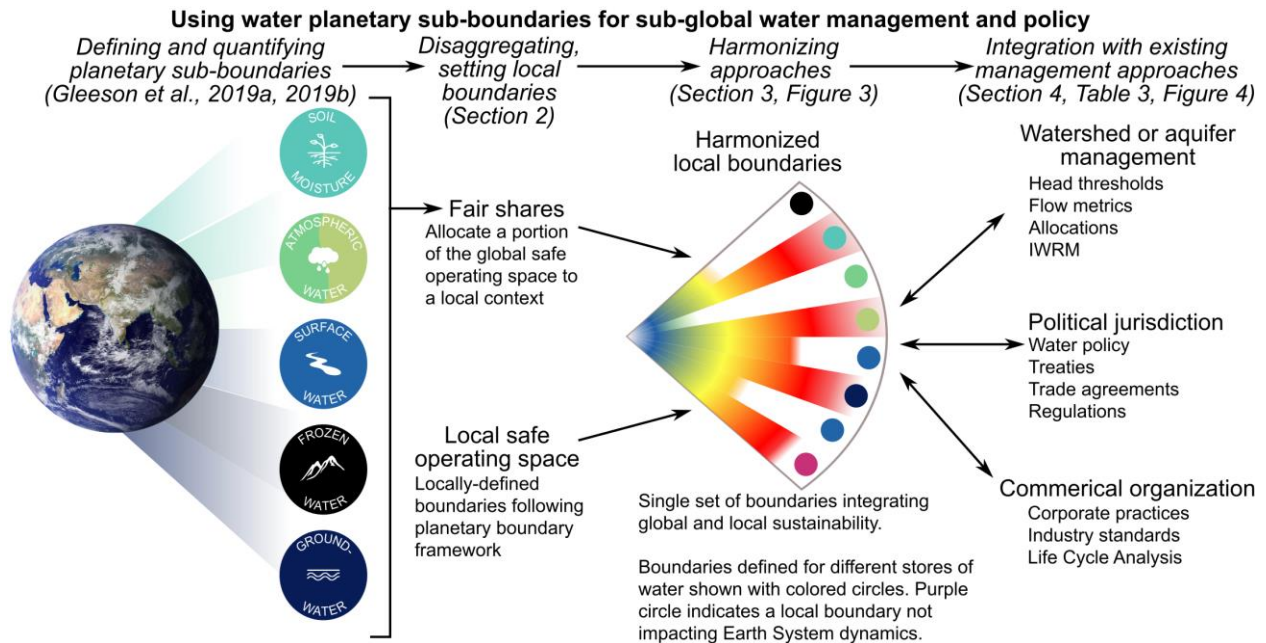
123 (ii) How does the planetary boundary framework complement existing water management
 124 approaches at each of these spatial scales?

125 (iii) What scientific questions need to be addressed to move the planetary boundary framework
126 forward as a potential water management approach?

127 We develop a flexible approach to applying the water planetary boundary across different scales and
128 jurisdictions, in order to complement existing management and governance approaches by accounting
129 for interactions across traditional water system borders and scales and for relationships among different
130 components of the Earth System. The water planetary boundary can be used in sub-global domains
131 defined using physical features (e.g. watershed or aquifer management), political borders (e.g. a city,
132 nation, or group of nations), or commercial entities (e.g. companies, industries, or trade groups
133 operating within or across national borders). As one example, we demonstrate how the water planetary
134 boundary can be used in the context of a degraded hydrological system, the Cienaga Grande de Santa
135 Marta wetland complex, in Colombia. Furthermore, because the water sub-boundaries highlight key
136 interactions between water cycle change, climate change, and land system change, we therefore
137 broaden the term “water management” to refer to not only management of liquid water in surface
138 water and aquifers, but any type of management of Earth System processes that have significant
139 interactions with the water cycle. For brevity, we will use the term ‘local contexts’ to refer
140 interchangeably to any of these sub-global applications.

141 2. Principles for using the planetary boundary in sub-global water management

142 Here, we synthesize previous literature on sub-global use of the planetary boundary framework (Table
143 2) to classify and explore two approaches, *fair shares* and *local safe operating space*, in order to identify
144 strengths, weaknesses, and principles for effective implementation of each approach. In Section 3, we
145 present a methodology to integrate these two approaches to calculate local harmonized boundaries
146 with strong connections to both global and local water sustainability.



147
 148 **Figure 2. Translating water’s planetary role in Earth System dynamics to local management and**
 149 **governance. In the harmonized approach, colored circles indicate a water sub-boundary corresponding**
 150 **to a specific store of water and the red-yellow-blue color gradient corresponds with the current position**
 151 **of the control variable with respect to the boundary, as in Figure 1. The fair shares approach will**
 152 **subdivide each water planetary sub-boundary to the local context. The local safe operating space**
 153 **approach may have local boundaries corresponding to all or some of the sub-boundaries, as well as**
 154 **additional locally-relevant boundaries which may not have an impact on Earth System function. Further**
 155 **details about each step are provided in the text sections and figures referenced in the top of the figure.**

156 **2.1 Fair shares approach**

157 The *fair shares* approach is a top-down approach which treats the planetary boundary value of the
 158 control variable as a global safe operating space “budget”, and then allocates a portion of that global
 159 safe operating space to a given local context. The fair shares approach has been used in diverse local
 160 contexts including nations, cities, companies and industries (Table 2). This approach has strong global
 161 relevance because it is directly connected to the Earth System functions that define the planetary
 162 boundaries, but this also means it may have limited local relevance because the globally defined control
 163 variable may not be the most effective descriptor or authoritative guidance for modifications of the
 164 water system in some local contexts. If global fair shares exceed socially and ecologically important
 165 thresholds in local hydrological processes (discussed in Section 2.2), local decision-makers would
 166 intervene in the water system well before the global boundary is reached and vice versa.

167 The fair shares approach requires three steps: (1) setting the planetary boundary value(s); (2) allocating
168 a fraction of the global safe operating space to a local context; and (3) comparing current performance
169 to the allocation (i.e. to the local fair share) for each control variable.

170 2.1.1 Setting the planetary boundary

171 The first step in operationalizing the fair shares approach is defining the safe operating space by
172 determining values for each of the water planetary sub-boundaries. While recent work has proposed
173 new planetary sub-boundaries for different stores of water in the Earth system (Gleeson et al., 2019a,
174 2019b), previous efforts to operationalize the planetary boundaries framework using the fair shares
175 approach have so far all used the control variable and estimated boundary values from Rockström *et al*
176 (2009) and Steffen *et al* (2015) (Table 2). In the case of water, this means that the planetary boundary is
177 typically taken as 4000 km³ yr⁻¹ of consumptive blue water use. To better account for the ecological
178 impacts of human water use, Gerten et al. (2013) suggested a spatially explicit quantification of
179 environmental flow requirements to explicitly focus on the impacts of freshwater use for local biosphere
180 integrity. Using this approach, they calculated monthly environmental flow requirements for all surface
181 water globally, which resulted in a boundary between 1100-4500 km³ yr⁻¹. Reframing the water
182 planetary boundary for different stores of water as proposed by Gleeson et al. (2019a, 2019b) improves
183 the ability of the fair shares approach to accurately represent water's role in Earth System dynamics,
184 and therefore we suggest that future work on the fair shares approach should shift to these new control
185 and response variables (Table 1).

186 2.1.2 Allocating the global safe operating space to local contexts

187 The fair shares approach requires allocating the global safe operating space (for any planetary
188 boundary) to sub-global scales and entities. For commercial organizations, economic indicators are
189 typically used such as a corporation's global market share (Ryberg et al., 2018; Sandin et al., 2015). For
190 political contexts (e.g. nations), allocation of the global safe operating space is most often done using a
191 per capita approach in which the global value of the control variable of the planetary boundary is
192 apportioned based on the number of people living in that nation relative to the global population (Dao
193 et al., 2015; Hoff et al., 2014; Nykvist et al., 2013; O'Neill et al., 2018). The major flaw of per capita
194 allocation is that it allocates a larger portion of the global safe operating space to more populous
195 nations, without taking into account local hydrological factors nor people's capacity to respond to
196 environmental challenges (Häyhä et al., 2018). Alternative allocation principles, for example based on
197 equality, rights, capacities and the responsibility of different groups of people could be used instead (see

198 Häyhä et al., 2016; Lucas & Wilting, 2018; or the extensive literature on allocating greenhouse gas
199 emissions). Different allocation approaches can result in substantial differences in what counts as a fair
200 share of the global safe operating space (Ryberg et al., 2018; Sandin et al., 2015).

201 Choosing among allocation approaches requires ethical and political decisions that account for local
202 differences in the contribution to global environmental challenges, the ability to respond to them, and
203 differing definitions of ‘fair’ among stakeholders (Biermann, 2012; Häyhä et al., 2016). For instance, the
204 suggested control variable for the frozen water sub-boundary is the volume of global ice melt (Table 1),
205 which is strongly driven by polar air temperature increases due to anthropogenic climate change;
206 however, the countries which contribute most to climate change are not the same countries that feel
207 the strongest (downstream) impacts (Althor et al., 2016; Biemans et al., 2019), posing an ethical
208 challenge. Climate change has been found to increase inequality of income distribution between
209 polluting countries and those countries with the least share of pollution responsibility (Differbaugh &
210 Burke, 2019). Addressing ethical issues is particularly challenging because different definitions of equity
211 lead to different outcomes (Häyhä et al., 2016; Ryberg et al., 2018). Also, regardless of the allocation
212 approach used, any fair shares approach will be sensitive to the estimated position of the planetary
213 boundary and control variable, both of which may include substantial uncertainty which must be
214 accounted for and communicated as part of the fair shares downscaling.

215 2.1.3 Comparing current performance to the allocation for each control variable

216 Previous planetary boundaries applications have used both production-based and consumption-based
217 approaches to calculate a local context’s performance relative to the allocation of a local fair share, both
218 of which are originally derived from climate accounting (Table 2). A production-based approach
219 considers impacts of the production of goods and services on the water cycle only within the local
220 context, such as within national territory. However, many environmental impacts (particularly from
221 wealthy nations) are partially externalized via trade (Marston et al., 2015; Wiedmann & Lenzen, 2018)
222 and/or felt in locations distant from where the water use occurs (e.g., transboundary effects; Munia et
223 al., 2016; Veldkamp et al., 2017). A consumption-based approach therefore considers the global impacts
224 on the water cycle associated with the water used to supply all the goods and services consumed within
225 the local context. This is known as embedded or virtual water, and there is growing consensus that it
226 should be accounted for in sustainable resource-use decision making. For example, more than 40% of
227 the water use associated with the supply of products and services to Europe occurs elsewhere (Häyhä et
228 al., 2018), and 11% of global nonrenewable groundwater use is attributable to international trade (Dalin

229 et al., 2017). Geographically explicit approaches such as water footprints can be used for consumption-
230 based quantification (Hoekstra & Mekonnen, 2012; Mekonnen & Hoekstra, 2011; Vanham et al., 2019).
231 Companies and industries, which operate across borders, frequently use consumption-based life cycle
232 analyses, which account for the various materials and impacts of processes required in complex global
233 supply chains, and have begun to integrate the planetary boundaries framework as a lens to interpret
234 the results of life cycle analyses (Brejnrod et al., 2017; Ryberg et al., 2018; Wolff et al., 2017). Life cycle
235 analyses can also use the water footprint approach to quantify the performance of a product, company,
236 or supply chain for water impacts all over the world (Chapagain et al., 2005; Kounina et al., 2013; Pfister
237 et al., 2009). The water footprint can be adjusted based on water scarcity at the location of production,
238 i.e. where the consumptive use takes place, to more directly reflect environmental impacts (Ridoutt &
239 Pfister, 2010), though this is controversial because it introduces additional value choices and non-
240 physical dimensions to the footprint metric (Hoekstra, 2016; Pfister et al., 2017).

241 Gleeson et al.'s (2019a, 2019b) new control variables will require the development of new approaches
242 for consumption-based quantification. As an example, they suggest that the control variable for
243 atmospheric water's role in hydroclimatic regulation may be the degree of human-caused change in
244 evapotranspiration, for instance due to land use change and water use. The water footprint concept
245 described above offers a way to link different changes in evapotranspiration to specific local actions
246 such as land use change or irrigation (Schyns et al., 2019). Land use footprints and indirect land use
247 change metrics (Searchinger et al., 2008; Weinzettel et al., 2013) can likely be adapted to meet the
248 needs of the fair shares approach for assessing sub-global responsibility. However, these sorts of
249 attribution studies are disputed due to the difficulties involved in tracing how national policies
250 propagate through the global economic system to influence land use (Mathews & Tan, 2009; Zilberman,
251 2017). Ongoing research aims to make environmental footprints, including land use, more spatially
252 explicit.

253 *2.2 Local safe operating space approach*

254 The *local safe operating space* approach is a bottom-up approach that uses the principles of the
255 planetary boundary framework to generate locally meaningful control variables, response variables, and
256 boundary values defining the local stable conditions of the water system (Figure 1). Local variables may
257 or may not be the same as the planetary boundary control and response variables, because the drivers
258 of hydrological stability in local water systems can differ from drivers of stability at the Earth System

259 scale. This approach allows stakeholders and water managers working on a specific region to define safe
260 operating spaces that have a strong local relevance and can inform efficient water management
261 interventions locally. However, local safe operating spaces potentially have weak relevance at the global
262 scale. Furthermore, while the planetary boundaries were designed as global measures of biophysical
263 Earth System conditions (Rockström et al., 2009; Steffen et al., 2015), local safe operating spaces are
264 more likely to integrate social-environmental considerations regarding the aspects of the water system
265 most important to society (Dearing et al., 2014; Teah et al., 2016).

266 The local safe operating space approach also typically contains three steps: (1) defining locally
267 meaningful control and response variables, which may differ from the variables used for the planetary
268 boundaries; (2) setting boundary values which defined the local safe operating space; and (3)
269 quantifying the current state of each control variable. These can be connected to the fair shares
270 approach using the harmonization approach presented in Section 3.

271 2.2.1 Defining the control and response variables

272 The local safe operating space approach complements the fair shares approach by focusing on the local
273 water system, and there is no explicit relationship to Earth System stability (though local effects may
274 scale up to impact Earth System stability). The definition of locally relevant control and response
275 variables should be based on the biophysical and socio-economic properties of the local water system,
276 which may already be identified in thresholds or allocations from existing water management
277 agreements. Like the planetary boundary control and response variables (Section 1), the control variable
278 should be quantifiable and can be influenced by human actions, while the response variable should
279 describe stable conditions for the local water system and be influenced by the control variable. There
280 may be cases where the control variable is the same in the local safe operating space and fair shares
281 approaches (Section 3).

282 Past attempts to define local safe operating spaces have typically used consumptive freshwater use, the
283 same control variable as the water planetary boundary (Table 2; Cole et al., 2014; Fanning & O'Neill,
284 2016; Teah et al., 2016). However, the local safe operating space approach may require alternative or
285 additional control and response variables based on the unique conditions of the local water system
286 context. This flexibility is well-aligned with the multiple water sub-boundaries corresponding to different
287 water stores proposed by Gleeson et al. (2019a, 2019b) in Table 1. For example, in an analysis of
288 regional safe operating spaces for two regions in China, Dearing et al. (2014) did not include a
289 freshwater use boundary because the primary regional water challenges were related to water quality

290 and sedimentation, rather than water quantity. Similarly, environmental flow requirements could be
291 used as a local process of interest. In this way, the local safe operating space can complement existing
292 management approaches.

293 2.2.2 Setting the local boundary

294 For operational purposes, setting a boundary value requires a clearly defined relationship between a
295 control variable and the response variable(s), though this was not explicitly defined in the original
296 freshwater use planetary boundary (Rockström et al., 2009; Steffen et al., 2015). In local contexts, the
297 stable conditions of the response variable may be defined using the observed range during the Holocene
298 (as in the planetary boundary for water) or based on some other locally-relevant environmental
299 thresholds. For instance, Dearing et al. (2014) use historical measurements of hydrological and
300 ecological variables in anthropogenically perturbed regions to identify boundary values. Identification of
301 water regime shifts (reviewed in Falkenmark et al., 2019) can be used to define locally-relevant ranges
302 of the response variable and corresponding control variables.

303 Using the planetary boundary framework to define a local safe operating space has primarily been an
304 academic exercise to date (Table 2), but for practical water management and governance, socio-
305 economic and equity concerns will come into play (Figure 1). During the setting of boundary values, both
306 the characterisation of stable hydrological conditions and assessments of acceptable levels of risk are
307 likely to vary among stakeholders within a community as well as across local contexts. For example,
308 poorer stakeholders may be less resilient to short-term hydrological variability (i.e., define stable
309 conditions as a narrower range in Figure 1) and have fewer options to reduce exposure to risk (i.e., set
310 the boundary further back from estimated thresholds in Figure 1). Community-level involvement has not
311 been prioritized in past efforts to apply the planetary boundaries framework, but from a sustainable
312 development perspective, local boundary setting can be rooted in environmental justice to define a
313 “safe and just operating space” (Dearing et al., 2014; Leach et al., 2013; Raworth, 2012). This would
314 require that all communities within the local context - not just the historically advantaged groups in a
315 position of power - are meaningfully involved in defining the local safe operating space and
316 implementing environmental policies or regulations to maintain the local safe operating space in a
317 manner which is responsive to changes in local conditions driven by external factors such as global
318 climate change (Martín-López et al., 2019).

319 2.2.3 Quantifying the current value of each control variable

320 Quantifying the current value of the control variable and comparing it to the local boundary value
321 informs whether the local context is within its local safe operating space. Where data are available to
322 define the relationship between the control and response variables, quantifying the current value of the
323 control variable can be done in a fairly straightforward manner (Cole et al., 2014; Dearing et al., 2014).
324 In many cases, however, data and deep understanding of the system needed to accurately estimate
325 control variable values are lacking for some or all local boundaries. This may provide an opportunity to
326 further integrate local communities in the definition of the local safe operating space, as demonstrated
327 by Teah et al. (2016). In addition to quantifying biophysical limits for a portion of the Heihe River in
328 China, Teah et al. (2016) surveyed local residents on their perceptions on the current status of the
329 control variables and the potential impacts of regional boundary transgression, finding that resident
330 perceptions of the values of the control variables relative to their regional boundaries were mostly
331 consistent with the quantified values. While the social survey was not used to guide the setting of
332 boundaries nor identification of the ecological thresholds, it does indicate that local stakeholder
333 involvement has the potential to both identify relevant control variables, estimate the present value for
334 control variables where monitoring data do not exist, and evaluate the potential impacts of
335 transgression. Where uncertainty in the estimation of the control variable is large, stakeholder
336 involvement may be particularly important to constrain acceptable control and response variable values.

337 3. Harmonizing approaches to integrate local to global water sustainability targets

338 *3.1 Harmonizing fair shares and local safe operating space approaches*

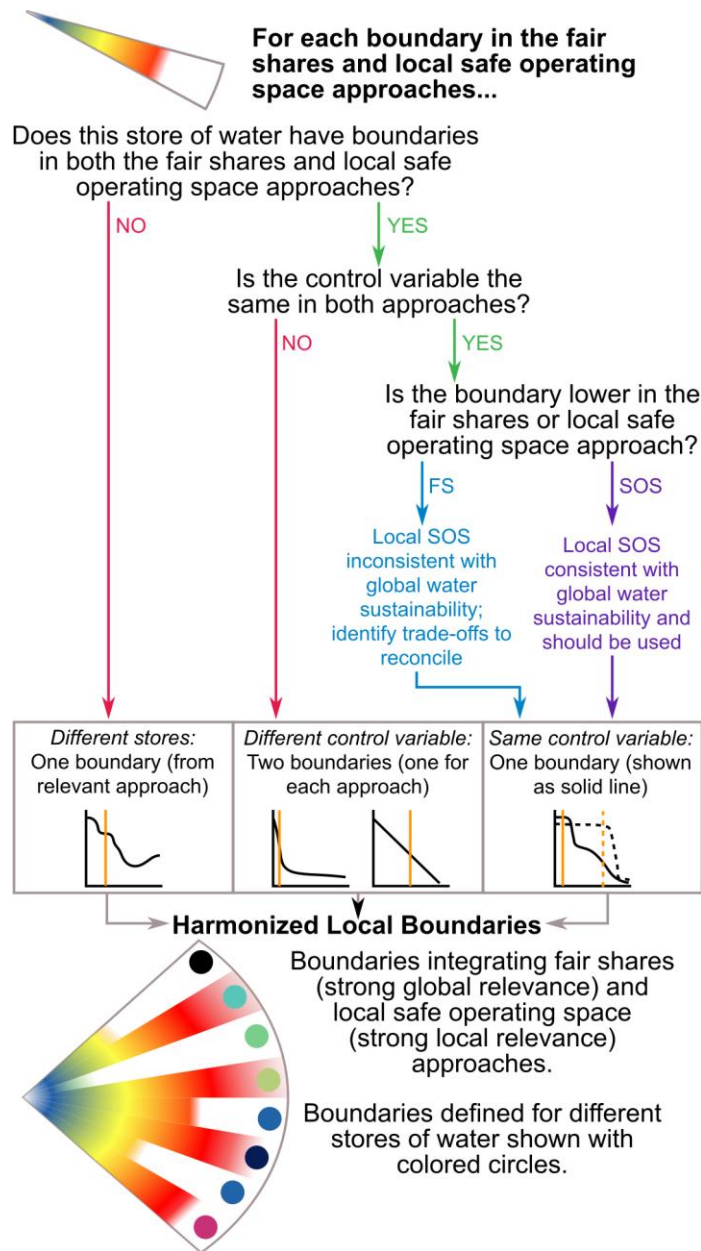
339 The fair shares and local safe operating space approaches each have benefits and drawbacks. The fair
340 shares approach provides a framework for assessing global responsibility and allocating responsibility
341 for local contribution to global processes; however, it does not provide any guidance for whether the
342 water cycle remains within locally important limits which are the primary concerns of water managers
343 and policymakers. In contrast, the local safe operating space approach addresses local limits to water
344 system modifications, but does not provide any information about potential impacts external to the
345 local context which are essential for global citizenship.

346 Here, we propose a methodology to harmonize the fair shares and local safe operating space
347 approaches to develop a set of local boundaries which are consistent with both local and global water

348 sustainability. For a given water-related boundary (here defined for each store of water), there are three
349 potential relationships between the fair shares and local safe operating space approaches (Figure 3):

- 350 1) *Different stores*: For stores which are relevant in only one of the two approaches, there is no
351 harmonization needed since there will only be a single boundary value. A hypothetical example
352 of this may be that frozen water is unimportant in many tropical catchments so would be
353 ignored in the local safe operating space approach, but is still considered in the fair shares
354 approach due to its impact on global processes. A second hypothetical example may be a locally
355 important store of water - for example, the water level in a lake - which is considered in the
356 local safe operating space approach but makes no significant impact globally so is ignored in the
357 fair shares approach. In this case, the local control variable could be water withdrawals directly
358 from the lake or surface water or groundwater in the lake's contributing area, the response
359 variable could be the aquatic biosphere integrity of the lake, and the local boundary defined
360 based on the change in lake levels which would lead to a collapse of the aquatic food web
361 (AghaKouchak et al., 2015; Kraft et al., 2012). In both these examples, the control variable,
362 response variable, and boundary value from whichever of the two approaches is relevant can be
363 used.
- 364 2) *Different control variables*: For stores where there are different core water functions at the local
365 and global scales, the control variables may differ between the fair shares approach (which
366 focuses on global processes) and the local safe operating space approach (which focuses on
367 local processes). Again, a hypothetical example illustrates this. For the groundwater sub-
368 boundary, the primary function of groundwater globally is providing baseflow to rivers during
369 dry periods to maintain environmental flow requirements (Table 1). For the fair shares
370 approach, this suggests a potential control variable of stream-aquifer flux, a response variable of
371 aquatic biosphere integrity, and a boundary value based on global environmental flow
372 requirements (Gerten et al., 2013; Gleeson et al., 2019a, 2019b; Wada et al., 2013). However, in
373 some local contexts there are many groundwater-dependent terrestrial ecosystems, suggesting
374 a potential control variable of groundwater depth below the land surface, a response variable of
375 terrestrial biosphere integrity, and a boundary value when groundwater drops below the
376 rooting depth (Eamus et al., 2015; Qiu et al., 2019; Rohde et al., 2017). For this type of
377 relationship, a harmonized approach would require a unique set of sub-boundaries for this
378 water store, with separate control and response variables for each of the two approaches (i.e.,
379 fair share and local safe operating space).

380 3) *Same control variable*: For stores where the same control variable is used for the local safe
381 operating space and fair shares approaches, the relationship between the control and response
382 variables may not be the same in the two approaches. Modifying our hypothetical example for
383 the groundwater sub-boundary from the previous type, if the control variable is stream-aquifer
384 flux and the response variable is aquatic biodiversity for both the local safe operating space and
385 fair shares approaches, the boundary value may be different in the two approaches if small
386 changes in aquatic biodiversity would push the local socio-environmental system beyond the
387 local safe operating space without having a negative global impact. Where the local safe
388 operating space boundary is lower (more environmentally conservative) than the fair shares
389 boundary, this indicates that prioritizing local management will be consistent with global Earth
390 System stability and therefore this lower value can be used. However, in cases where a fair
391 shares boundary is lower than the local safe operating space boundary indicates that locally
392 acceptable water management practices are problematic when upscaled to the planetary level.
393 In this case, ethical and socio-economic considerations are necessary to reconcile this mismatch
394 between scales (Häyhä et al., 2016), for example by determining whether excessive local
395 impacts can be compensated for by conservation elsewhere at the global scale via trade-off
396 analysis (Section 3.2). In either case, the end result would be a single harmonized boundary.
397 Because the control-response relationship and therefore boundary value may change through
398 time due to changing biophysical conditions and interconnectedness with other non-water-
399 related processes, the relationship between the local safe operating space and fair shares
400 boundary values may change through time.



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Figure 3. Decision tree for harmonization process between fair shares and local safe operating space approaches. Each sub-boundary in the fair shares and local safe operating space approaches should be evaluated to determine whether a given store of water should have one or two boundaries in the harmonized set. The hypothetical plots corresponding to each type of boundary show the relationship between a control variable (x-axis) and response variable (y-axis) as shown in Figure 1, and the orange line shows the boundary value. Each of the colored circles indicates a water sub-boundary corresponding to a specific store of water from either the fair shares or the local safe operating space approach, as in Figure 2. The red-yellow-blue color gradient corresponds with the current position of the control variable with respect to the boundary, as in Figure 1.

412 Based on these three types of relationships, the locally harmonized water boundary will always have at
413 least as many sub-boundaries as the water planetary boundary, which may be supplemented by
414 additional sub-boundaries derived from the local safe operating space approach. In this manner, the
415 harmonized local boundaries will always be consistent with both global and local water sustainability
416 goals, and provide a framework for determining whether local water management is consistent with
417 both local socio-environmental processes and global Earth System function. An example of these types
418 of relationships for the Ciénaga Grande de Santa Marta wetland complex (Colombia) is explored in Box
419 1.

420 *3.2 Recognizing and respecting real-world complexity*

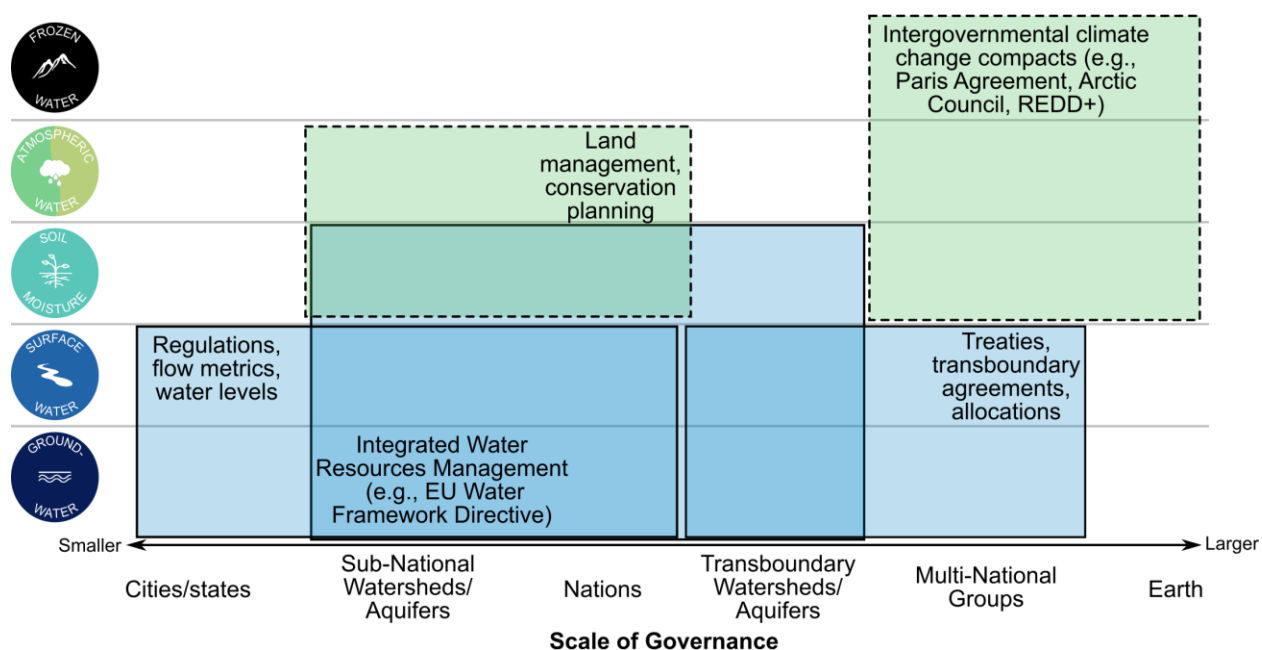
421 Comparison between the local safe operating space and fair shares approaches can provide valuable
422 insight into the cross-scale relationships between sub-global-scale water systems and Earth System
423 function. For instance, Gleeson et al. (2019) suggest that there may be ‘keystone regions’ where water
424 cycle modifications have a disproportionate impact on Earth System function which manifest as changes
425 in the core planetary boundaries of climate change and biosphere integrity. In keystone regions, we
426 hypothesize that the local safe operating space and fair shares approaches would have similar control
427 and response variables and boundary values, since the degradation of the local water system would lead
428 to outsize impacts at the global scale. The keystone region concept should not be taken to imply that
429 local water cycle modifications with little detectable influence on Earth System stability are unimportant
430 to the water planetary boundary. All local changes contribute to the cumulative effect on the global
431 water cycle, and through the to the pressures on the water and other planetary boundaries, especially
432 the core boundaries of climate change and biosphere integrity. To avoid this unintended consequence of
433 the planetary boundary framework, harmonization with the fair shares approach as described above
434 and existing management and governance approaches is essential. Additionally, countries and
435 organizations in non-keystone regions might also yield a high indirect influence over keystone regions,
436 for example through trade and geopolitics (Bierkens et al., 2019; Dalin et al., 2017; D’Odorico et al.,
437 2019), which can be accounted for in the fair shares approach. Thus, while this paper primarily focuses
438 on disaggregation from global to local scales, comparison between the local safe operating space and
439 fair shares approaches may also be useful for determining appropriate techniques to aggregate from
440 local to global scales (CISL, 2019).

441 While the typologies presented in Figure 3 are comprehensive of all potential local-global interactions,
442 the complexity of the real world will introduce trade-offs among water sub-boundaries and other
443 planetary boundaries across spatial scales, time scales, degrees of reversibility, stakeholders, and types
444 of environmental impacts (Booth et al., 2016; Heck et al., 2018b; Qiu et al., 2018; Rodríguez et al., 2006).
445 In addition, boundaries in the real-world may not be expressed as two-dimensional plots as shown in
446 Figure 1, but rather as multidimensional parameter spaces representing multiple interconnected
447 ecohydrological processes. As one example, groundwater withdrawals for irrigation may enhance local
448 net primary productivity and food production, but alter regional-scale hydroclimate (spatial scale trade-
449 off; DeAngelis et al., 2010; Harding & Snyder, 2012a, 2012b; Keune et al., 2018), impair groundwater-
450 dependent ecosystems (stakeholder trade-off; Barlow & Leake, 2012; Gleeson & Richter, 2017; Zipper et
451 al., 2018, 2019), and make groundwater resources unavailable for future generations (reversibility and
452 time scale trade-off; Butler et al., 2018; Gleeson et al., 2012; Wada & Bierkens, 2014; Whittemore et al.,
453 2016) and increase cropland at the expense of other ecosystem services or planetary boundaries, such
454 as biodiversity and biochemical flows (environmental impact trade-off; Anache et al., 2019; Foley et al.,
455 2005, 2011; Hanaček & Rodríguez-Labajos, 2018; VanLoocke et al., 2017).

456 There are numerous existing frameworks, models, and tools for understanding interactions and
457 managing trade-offs with diverse approaches including cluster analysis, integrated (nexus) modelling,
458 multi-criteria analyses, and trade-off curves (Cavender-Bares et al., 2015; Deng et al., 2016; Heck et al.,
459 2018b; Hurford et al., 2014). While it is beyond the scope of this paper to address how the water
460 planetary sub-boundaries can be integrated with existing trade-off analysis tools, we note that
461 managing trade-offs will require a detailed and process-based understanding of interactions among the
462 water sub-boundaries and the rest of the planetary boundaries at local and global scales, and motivates
463 further research on understanding cross-scale feedbacks between local water systems, social-ecological
464 conditions, and the Earth System. To make decisions among local and global trade-offs will inherently be
465 a normative and participatory exercise based on clear visualisation and communication of trade-offs to
466 local stakeholders and decision-makers (see Section 5), and may be a tool to mobilize non-local financial
467 resources as incentive or compensation for foregone local benefits when contributing to global
468 sustainability targets and thus help a region stay within both local and global boundaries. This can follow
469 examples from the climate finance domain such as the Amazon Fund (Forstater et al., 2013; Nakhoda
470 et al., 2013).

471 **4. Opportunities for complementing existing water management approaches**

472 Numerous approaches to water management exist, most of which focus on either surface water or
 473 groundwater (Figure 4). In this section, we discuss how, and why, local applications of the water
 474 planetary boundary can complement these existing approaches (Table 3; CISL, 2019). Most importantly,
 475 the Earth Systems focus of the planetary boundaries demonstrates the necessity of expanding ‘water
 476 management’ beyond the traditional focus on surface water and groundwater to include aspects of land
 477 management (related to atmospheric water, precipitation and soil moisture) and climate change
 478 governance (related to changes in frozen water and in water availability).



479
 480 **Figure 4.** Examples of water management and governance approaches at different spatial scales
 481 targeting each store of water proposed for the water planetary boundary. Blue boxes are water
 482 resource management approaches and green dashed boxes are management approaches which are not
 483 designed specifically for water but likely to have a strong effect on that specific water store.

484 **4.1 Watershed or aquifer management (single jurisdiction or transboundary)**

485 At the watershed or aquifer scale, existing management approaches often identify critical threshold(s) of
 486 streamflow, aquifer or reservoir water levels, or some other metric describing stores or fluxes of water.
 487 When watersheds or aquifers cross administrative boundaries, relevant metrics include water
 488 allocations to different stakeholders. These metrics often balance socio-economic and biophysical
 489 considerations - for instance, how much water is necessary to support irrigated agriculture in the

490 watershed while preserving sufficient in-stream flows for aquatic ecosystems. Integrated water
491 resources management (IWRM) is a watershed-focused water management framework, which is
492 typically applied at the watershed scale. Several core principles underlying IWRM and the planetary
493 boundaries overlap: both frameworks treat different stores of water as inherently interconnected, and
494 explicitly include land surface processes in the scope of water management (Badham et al., 2019). Since
495 IWRM is designed to be applied at the watershed scale, the water planetary boundary provides a
496 complementary framework for considering water flows that are not “seen” in IWRM - atmospheric,
497 subsurface, and biosphere flows which cross watershed boundaries (Srinivasan et al., 2017). There is
498 growing acknowledgment that human activities can directly or indirectly influence the stocks and flows
499 of water beyond the watershed or nation in which they occur, through pathways including: (i) use in the
500 production of goods and services, which enter or exit the area, i.e. through virtual water trade (Dalin *et*
501 *al* 2017, Marston *et al* 2015) or foreign direct investment; (ii) flow through the atmosphere, i.e. moisture
502 recycling (Wang-Erlandsson *et al* 2018, Keys *et al* 2012, 2014, 2016), or subsurface, i.e. regional
503 groundwater flow (Ameli et al., 2018; Krakauer et al., 2014; Maxwell & Condon, 2016; McKenzie & Voss,
504 2013); (iii) through physical infrastructure, i.e. interbasin water transfers (Garrick et al., 2019; McDonald
505 et al., 2014), within the realm of traditional water management; and (iv) altered land-atmosphere-ocean
506 interactions, i.e. local groundwater depletion leading to global sea level rise (Döll et al., 2014; Wada et
507 al., 2010). In particular, the importance of atmospheric water flows highlighted by the planetary
508 boundaries framework indicates a need for integrating existing (blue) water management with domains
509 typically thought of as land or climate governance, for example via land use planning and thus managing
510 green water fluxes (Figure 4; Keys et al., 2017).

511 One example of how existing watershed-scale water management approaches may be complemented
512 by water planetary boundary framework is the European Water Framework Directive (WFD), a water
513 management approach at the watershed scale, which aims to reach and maintain “good” ecological and
514 chemical status, defined relative to natural conditions (Kallis & Butler, 2001). However, the WFD does
515 not aim to establish or define any boundaries, nor does it situate this status relative to potential broader
516 scale feedbacks with Earth System processes (Bishop et al., 2009). Thus, the water planetary boundary
517 framework complements the WFD in two primary ways. First, the fair shares approach to downscaling
518 the water planetary sub-boundaries provides a tool for assessing whether management actions in
519 Europe are sufficient to maintain water stocks and flows and good ecological conditions from an Earth
520 System perspective, especially through quantifying externalized environmental impacts via
521 consumption-based methods. Second, the local safe operating space approach provides a tool that can

522 be used to set locally relevant boundaries within the context of the WFD, which is particularly needed
523 for heavily modified or artificial water features, with WFD status that contain significant subjectivity, and
524 prioritize outcomes based on the local social-ecological system. As of 2015, 47% of the European waters
525 had not reached good ecological status, so it can also be argued that the Directive has fallen short in
526 delivering coherent and sustainable water management in Europe (Voulvoulis et al., 2017). In sum, the
527 water planetary boundary framework provides opportunities to better contextualize the WFD at both
528 local and Earth System scales.

529 *4.2 Political jurisdictions (state/province, national, multi-national)*

530 Political jurisdictions use a number of policies, regulations, and incentives to govern water. These
531 instruments are created and applied at numerous and often overlapping spatial scales (Wardropper et
532 al., 2015). States/provinces, cities, or other local jurisdictions may supplement national water-related
533 policies at sub-national scales. Multiple nations can be bound together through political and trade
534 agreements, including political units such as the European Union and African Union; trade agreements
535 such as North American Agreement on Environmental Cooperation and the ASEAN Free Trade Area; and
536 intergovernmental organizations such as the United Nations.

537 Chapron et al. (2017) argue that these “legal boundaries” need to be designed and enforced in ways that
538 translate the biophysical planetary boundaries into effective delimiters of human activities, either to
539 prevent transgression of planetary boundaries or to scale down the human pressures on boundaries
540 that are already exceeded. At the scale of an individual political unit such as a nation, the planetary
541 boundaries framework provides a basis for considering the effects on stocks, flows and processes of
542 water external to the unit’s borders (whether geographic or institutional), which activities within the
543 unit may impact. Transboundary water management frameworks have been developed for some
544 watersheds and aquifers that cross political boundaries (Earle, 2013; Eckstein, 2011; Puri & Aureli,
545 2005). However, these are unequally distributed globally and focus primarily on surface water
546 watersheds and to a lesser degree aquifers (McCracken & Meyer, 2018). To our knowledge, no existing
547 water management agreements address atmospheric water flows across watershed boundaries, i.e.
548 continental moisture recycling (Keys et al., 2016, 2017; Zemp et al., 2014), through which land use
549 change or human water use can alter precipitation remotely (Keune et al., 2018; Wang-Erlandsson et al.,
550 2018; Zemp et al., 2017; Zipper et al., 2019a). These transboundary atmospheric water flows indicate a

551 need to expand the scope of water management beyond watershed and national scale and beyond blue
552 water to include green and frozen water (see Section 4.1).

553 Political jurisdictions are inherently limited in their ability to regulate outside their physical boundaries.
554 Calculating fair shares of planetary boundaries provides important information and incentives for setting
555 aspirational and also operational goals for water managers and other actors to motivate action that
556 addresses sustainability targets beyond their local geographic context, such as downwind and
557 transboundary effects on water systems (Keys et al., 2018; Munia et al., 2016). The planetary boundary
558 framework provides a tool, which can consistently be applied in multiple locations to estimate the
559 relative contribution to a defined planetary boundary aligned with local boundaries.

560 One example of how the water planetary boundary may contribute to expand water management
561 beyond watershed and national scale is provided by Häyhä et al. (2018). Taking a fair shares approach to
562 downscaling, Häyhä et al. find that a consumption-based approach to water use (e.g., water
563 footprinting) is essential to accurately calculate the European Union's total impact on water systems,
564 because >40% of water use caused by EU consumption of goods and services takes place outside its
565 borders, mainly through agricultural imports. By systematically applying the same method across all
566 countries in the EU, the authors are able to provide a consistent framework for inter-regional
567 comparisons. Additionally, Häyhä et al. conclude that the primary benefit of the planetary boundary
568 framework relative to existing management approaches is the focus on interconnections between the
569 water planetary boundary and other Earth System processes such as land system change and
570 biogeochemical flows. While Häyhä et al. evaluated countries relative to the previously defined
571 consumptive blue water use boundary (Rockström et al., 2009; Steffen et al., 2015), we expect that
572 applying their same methods to the other suggested planetary boundaries (Table 1) would provide
573 additional useful management context to compare among nations and regions and guide national-level
574 water and land sustainability policy.

575 *4.3 Commercial organizations (corporations, industries, financial institutions)*

576 The water planetary boundary can also be used to guide decision-making of commercial organizations. It
577 may benefit a private sector stakeholder by providing a basis for traceable metrics for demonstrating a
578 commitment to global sustainability, and also scoping for shifts in risk exposure along the value chain.
579 Since the UN 2030 Agenda and Paris Agreement explicitly include the private sector as actors, there is a
580 growing demand for the development of science-based sustainability targets for commercial

581 organizations, which consider negative impacts on environmental and hydrological resources (and
582 societal spillover effects) across scales, from local to global. Clift et al. (2017) discuss the challenges and
583 opportunities for businesses to use the planetary boundaries in their decision-making strategies. In
584 particular, they note that the planetary boundary framework may provide a consistent approach to
585 compare performance among different regions or companies using the fair shares approach outlined
586 above (Table 3).

587 Effective use of the planetary boundaries can inform business decision-making by improving the
588 understanding of the interdependence between economic activity and global sustainability. As long as
589 full lifecycle impacts on water resources are accounted for, as described in Section 2, a corporation's fair
590 share of the water planetary boundary could provide a globally consistent way to assess the water
591 sustainability of a product or company, similar to a 'fair trade' or 'ocean-wise' product branding.
592 Additionally, corporations can use the water planetary boundary framework to identify risks to water
593 along their global supply chain (CISL, 2019). Recent work has shown that a relatively small number of
594 transnational corporations, deemed 'keystone actors' as an analogy to the ecological concept of
595 keystone species, have a disproportionate influence over Earth System function including marine
596 ecosystems (Österblom et al., 2015), deforestation in boreal forests and the Amazon (Galaz et al., 2018)
597 and more (Folke et al., 2019). Identifying and working with keystone actors in different local contexts
598 could provide one mechanism to effectively manage water resources; such science-business
599 partnerships are currently emerging for global fisheries management (Österblom et al., 2017). A
600 reframing of management education may also be needed to better situate Earth System science within
601 business education programs (Edwards et al., 2018)

602 Several companies are exploring ways to implement the planetary boundary framework. Butz et al.
603 (2018) described an approach for translating planetary boundaries into 'economic intensities' to inform
604 investment decisions. To account for effective water availability, they included toxic emissions to water
605 and groundwater into their calculation. L'Oréal, the multinational beauty company also include
606 measures of freshwater ecotoxicity as well as water resource depletion (Vargas-Gonzalez et al., 2019).
607 Houdini, the outdoor clothing company, now includes a planetary boundaries assessment in its
608 sustainability reporting (Haeggman et al., 2018). Alpro, the plant-based foods company, has piloted the
609 use of the planetary boundaries framework as a basis for setting science-based targets for nature
610 (Gladek et al., 2019), focused on translating corporate activities into environmental impacts. They
611 propose accounting for both blue and green water use at the basin scale, similar to the context-based

612 approach to target setting for water (Gillespy et al., 2017). The Kering fashion company has partnered
613 with the University of Cambridge Institute for Sustainability Leadership to identify how businesses can
614 best use the planetary boundary framework for assessing corporate sustainability (CISL, 2019). This
615 report highlights the differences between ‘downscaling’ (i.e., fair shares) and ‘upscaling’ (i.e., local safe
616 operating space approaches). CISL suggests that businesses should explore the opportunities of using
617 the local safe operating space approach to guide corporate practices. Rather than trying to assess what
618 is left to exploit, as might happen by comparing conditions to the global safe operating space under the
619 fair shares approach, the focus should shift to actions needed to restore local environmental functioning
620 in affected areas which would be identified using the local safe operating space approach. This is
621 particularly important for globally heterogeneous boundaries like water because it considers
622 hydrological impacts along the whole value chain, beyond the immediate local context.

623 **Box 1. Local-global connections in the Ciénaga Grande de Santa Marta Wetland** 624 **Complex, Colombia**

625 To demonstrate how the water planetary boundary framework may complement existing water
626 management approaches, we present a case study of the Ciénaga Grande de Santa Marta wetland
627 complex in Colombia. The Ciénaga wetlands are among the most productive coastal wetlands in the
628 world (Cloern & Jassby, 2008) and a listed Ramsar Site (Ramsar.org, 2019). The mangrove-dominated
629 system is highly susceptible to changes in salinity (Figure 5; Cardona & Botero, 1998; Elster et al., 1999),
630 which is driven by the balance of freshwater inputs from precipitation and rivers, saltwater inputs from
631 the ocean, and concentration of salinity via evapotranspiration within the wetland. The natural state of
632 exchange between the rivers, ocean, and lagoon has been disrupted by human activity. Road
633 construction in the 1950s cut off most surface water and groundwater exchange between the ocean and
634 the wetland. In the 1970s and 1980s road and berm construction along the Magdalena River decreased
635 freshwater inflows – these led to an increase in water salinity. Concurrently, irrigation and changes in
636 land cover upstream led to a decrease in freshwater inputs and an increase in sediment loading to the
637 wetland (Jaramillo et al., 2018a, 2018b; Perdomo et al., 1998; Rodríguez-Rodríguez, 2015). Beginning in
638 the 1990s, the Colombian government developed a long-term environmental management plan for the
639 Ciénaga wetlands (Botero & Salzwedel, 1999; Vilarly et al., 2011), focused on restoring hydrological and
640 ecological conditions by mangrove reforestation and dredging to increase freshwater inflows (Figure 5).

641 Mangrove cover has increased since the mid-1990s (Jaramillo et al., 2018a), but recovery is slower than
642 expected (Röderstein et al., 2014).

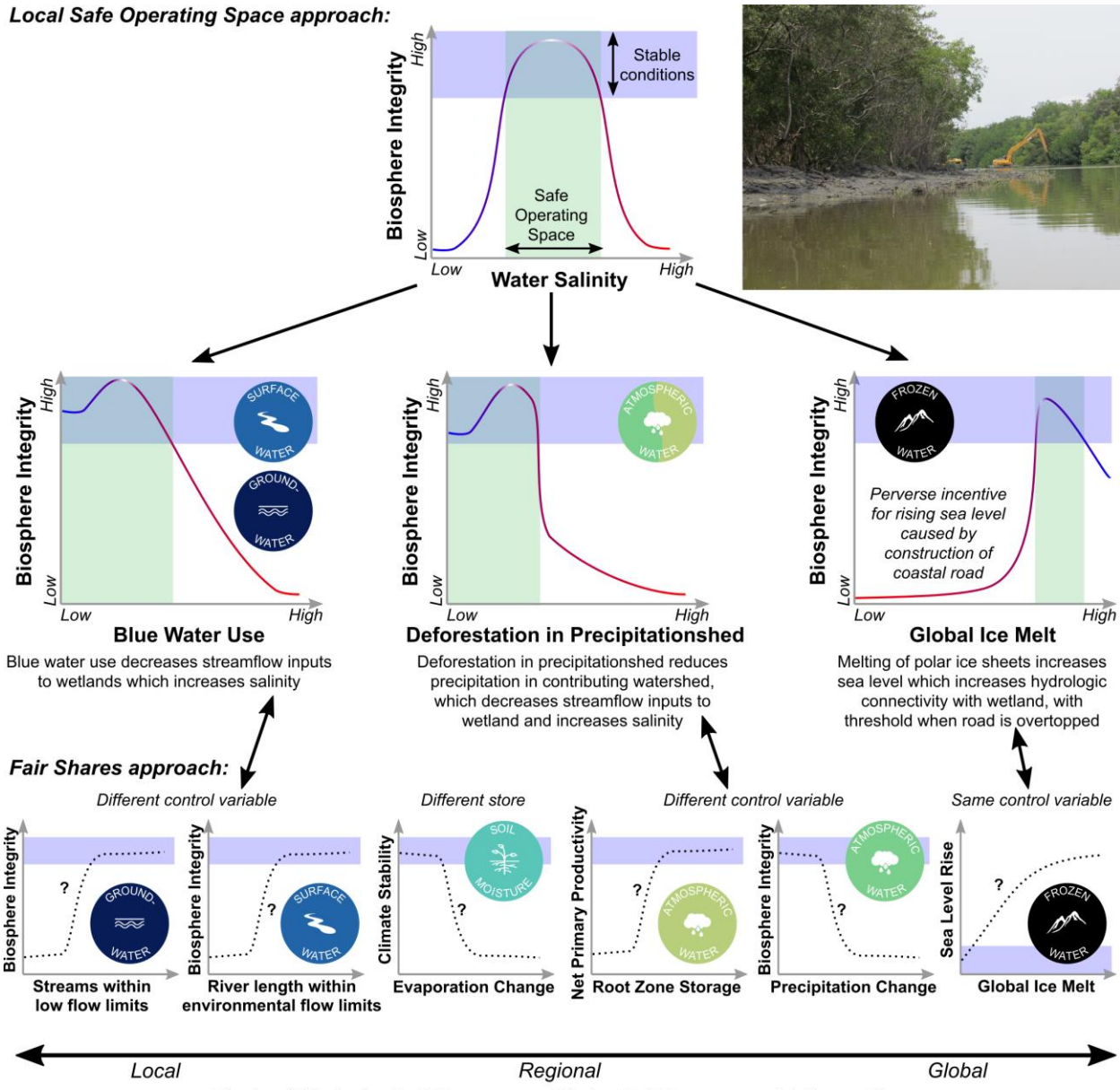
643 The planetary boundaries framework helps identify multiple feedback mechanisms occurring at nested
644 spatial scales which define the local safe operating space of the Cienaga wetlands (Figure 5). The
645 primary goal of local management in the Cienaga wetlands is protecting the mangrove ecosystem. For
646 the local safe operating space approach, *biosphere integrity* (response variable) depends on keeping
647 *water salinity* (control variable) within a narrow optimal range (Figure 5, top row). Three hydrological
648 mechanisms occurring at three different spatial scales influence salinity and thereby bound the local
649 safe operating space: At the local scale, freshwater inflows to the wetlands are influenced by upstream
650 blue water withdrawals for intensive agriculture, primarily banana plantations (Botero & Salzwedel,
651 1999; Vilardy et al., 2011). At the regional scale, the amount of incoming precipitation to the Cienaga
652 wetland is influenced by ocean-atmosphere oscillatory cycles such as ENSO (Blanco et al., 2006; Hoyos
653 et al., 2019; Jaramillo et al., 2018a; Restrepo et al., 2014) and also by terrestrial moisture recycling via
654 evapotranspiration from the wetland's precipitationshed which has areas of high deforestation,
655 including the Magdalena River Basin (Keys et al., 2016; Wang-Erlandsson et al., 2018; Zemp et al., 2014).
656 At the global scale, sea level rise linked to ice-sheet melt and climate change increases ocean-wetland
657 exchange. However, attention to the global scale alone could lead to a perverse incentive to allow
658 increases in sea levels to improve wetland biosphere integrity, since sea salinity (~35 ppm) is lower than
659 the current hypersaline conditions causing mangrove mortality (>100 ppm). These hydrological
660 processes and likely boundary values are additionally modified and affected by ongoing environmental
661 change in the Cienega wetlands, for instance upstream erosion and sedimentation associated with land
662 use change reducing the hydrological connectivity between the river and the wetland and would
663 therefore exacerbate the negative impacts of reduced flows caused by upstream blue water use (Botero
664 & Salzwedel, 1999; Jaramillo et al., 2018b).

665 Applying the fair shares approach would apportion the water planetary sub-boundaries to the Cienaga
666 wetlands. Using the sub-boundaries tentatively proposed by Gleeson et al. (2019a, 2019b) we apply the
667 harmonization approach proposed in Section 3. We see a mixture of relationships between the local safe
668 operating space and fair shares approaches. The surface water and groundwater sub-boundaries have
669 different control variables. For the fair shares approach, the control variables primarily focus on in-
670 stream conditions (environmental flows and low flow thresholds), while in the local safe operating space
671 approach the control variable is concerned with net inflows into the wetland lagoon which are a

672 function of blue water use. The soil moisture sub-boundary is only relevant in the fair shares approach
673 because local net primary productivity is not strongly dependent on soil moisture since the mangrove
674 wetlands are frequently inundated. The atmospheric water sub-boundaries also have different control
675 variables in the two approaches: the fair shares approach uses changes in evaporation and precipitation
676 change as control variables, and the local safe operating space approach uses deforestation in the
677 precipitationshed of the contributing watershed. Finally, the frozen water sub-boundary has the same
678 control variable (global ice volume) for both the fair shares and local safe operating space approaches,
679 but the local safe operating space creates a perverse incentive for rising sea levels due to the historic
680 construction of a coastal road. This mismatch indicates that the local safe operating space approach for
681 frozen water is inconsistent with global water sustainability (Figure 3), and thus only the fair shares
682 approach would be used for defining the safe operating space for frozen water in the Cienaga.

683 Combined, this analysis reveals several insights unaccounted for in current management efforts. First,
684 the local safe operating space approach in particular can be integrated with as a tool for setting both
685 limits to modification and targets for restoration of the local water system. Second, all aspects of the
686 water cycle are important to the restoration of the Cienaga wetlands, though management (in the
687 Cienaga and elsewhere; Figure 4) focuses primarily on surface water and, to a lesser degree,
688 groundwater (Vilardy et al., 2011). Third, accounting for atmospheric water and frozen water requires a
689 broad and cross-scale perspective which addresses for drivers and risks at the local, regional, and global
690 scales (Keys et al., 2019). In summary, the water planetary boundary framework discussed here provides
691 additional useful insight into the Earth System processes sustaining local ecosystems, and identifies
692 actions at multiple spatial scales from local to global which can help sustain the Cienaga wetlands. These
693 nested spatial scales of Earth System processes and feedbacks among management and other factors
694 indicates that collaborative, multi-scale governance approaches will be necessary to halt or reverse the
695 degradation of the Ciénaga wetlands (Biermann et al., 2012; Keys et al., 2017; Konar et al., 2016).

Local Safe Operating Space approach:



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Figure 5. Example definition of local safe operating space for Cienaga wetlands including underlying Earth System processes at local, regional, and global scales identified using the planetary boundaries framework (top section) and the suggested control and response variables for the fair shares approach (bottom section) from Gleeson et al. (2019). The lines on the local safe operating space plots show hypothesized relationships based on feedbacks described below plots, and the lines on the fair shares plots are placeholders since global relationships necessarily for downscaling are not yet known. Inset photo by INVEMAR shows dredging of channels to increase freshwater inflows.

705 5. Further development of the water planetary boundary forward to complement 706 current water management and governance approaches

707 While the planetary boundaries framework is recognised to have value in framing policy integration and
708 supporting policy coherence - both across scales and across sectors, through different tiers of
709 government - in practice, the studies we cite have generally remained as points for science-policy
710 discussion rather than actual policy shifts. Here, we highlight three important aspects and research
711 priorities that will enable better integration of planetary boundaries framework with existing water
712 management and governance approaches.

713 First, the planetary boundaries framework can be regarded as a simple snapshot of a complex adaptive
714 system. To integrate the planetary boundary framework with water management and governance, it is
715 essential to consider not just the current value of the water sub-boundaries, but also their temporal
716 dynamics and the current values, trajectories and potential systemic changes of the other boundaries,
717 over the course of a policy-relevant timeframe. In addition, changes in planetary boundaries are likely to
718 have knock-on effects on socio-economic dynamics, with potential feedbacks influencing the water sub-
719 boundaries (Crépin et al., 2017). Each of the planetary boundaries is defined in terms of a notional
720 Holocene-like state, which means that the control variables reflect the predominant interconnections
721 and feedbacks of the Holocene Earth system - but these properties are changing in the Anthropocene
722 (Waters et al., 2018), in particular because socio-economic feedbacks play an increasing role in the
723 trajectory of the Earth system. For example, the transgression of the land system change and climate
724 change planetary boundaries (Rockström et al., 2009; Steffen et al., 2015) are having consequences for
725 the water cycle and its role in biophysical feedbacks, and actions to remain within those boundaries may
726 help or hinder our ability to remain within the water planetary sub-boundaries (Heck et al., 2018b,
727 2018a; Stenzel et al., 2019). Policies based on water planetary boundary quantification should
728 investigate not just the current status of a control variable, but also its trajectory and relationship to the
729 temporal dynamics of other ecological and socio-economic components (Dearing et al., 2014). Since
730 different hydrological processes act on dramatically different timescales (residence times for
731 atmospheric water and groundwater differ by orders of magnitude; Oki & Kanae, 2006), adaptive water
732 management approaches are needed that consider the time taken for a store of water to change
733 relative to its current value and trajectory. Such adaptive approaches are key to sustainable water
734 management (Gleeson et al., 2012b; Pahl-Wostl, 2007, 2008; Rohde et al., 2017).

735 Second, accounting for scientific uncertainty is a longstanding challenge for water management
736 (Badham et al., 2019; Merz et al., 2015; Poff et al., 2016; Varela-Ortega et al., 2011; Wei et al., 2011).
737 Given the complexities inherent in Earth System dynamics, there are numerous aleatoric and epistemic
738 uncertainties embedded in the planetary boundaries framework which will need to be estimated and
739 communicated to policymakers and others (Westerberg et al., 2017): uncertainty regarding the complex
740 and nonlinear feedbacks in the Earth System (Steffen et al., 2018); uncertainty regarding the
741 relationship between the control and response variables at both local and global scales; uncertainty
742 inherent in both observational data and models used to quantify the current value of the control
743 variable (Long et al., 2014; Sperna Weiland et al., 2015); uncertainty in the approaches for
744 aggregating/disaggregating control variables and allocating the safe operating space globally (Mace et
745 al., 2014); uncertainty in the harmonization process of the two approaches here discussed; etc. These
746 uncertainties will most directly manifest in the definition of stable and unstable conditions for the
747 response variable and integrating uncertainty into water management and governance, particularly
748 aleatoric uncertainty which is inherently unpredictable (Poff et al., 2016). The precautionary principle
749 suggests that, as uncertainty increases, so should the setback from the estimated threshold value,
750 particularly given the large impacts of crossing these thresholds (Crépin & Folke, 2015; Margolis &
751 Nævdal, 2008) and the presence of substantial time-lags (Crépin & Nævdal, 2019). Integrating the
752 planetary boundaries into management frameworks will drive the creation and adoption of new data
753 analytics and modelling tools that can better handle the complex behaviour of water in the Earth
754 system, where it plays multiple physical, ecological and biogeochemical roles (Chang et al., 2018; Fan et
755 al., 2019; Milly et al., 2014; Sippel et al., 2015).

756 Third, the scientific community needs to closely work with policy and decision makers in order to
757 transparently identify relevant options that can meet existing local water management constraints while
758 satisfying the Earth System sustainability defined by the water planetary boundary, and indicate how
759 performance on a harmonized approach will be evaluated. This will require international commitment to
760 transdisciplinary fundamental and solutions-oriented research, for example through existing science-to-
761 policy approaches and boundary organizations such as Future Earth (Sun et al., 2016). As mentioned
762 above, the water planetary boundary is just one of a set of nine boundaries with strong
763 interrelationships. As scientific understanding increases about these interactions as well as about other
764 changes in interrelated aspects of the Earth System, care needs to be taken to ensure that this shifting
765 information baseline informs adaptive policy making rather than confounding it (Galaz et al., 2012a). If
766 we adopt the mindset that the planetary boundaries are guardrails defining a 'corridor' which can be

767 navigated through multiple socio-political pathways (Biermann, 2012; Qiu et al., 2018a), changes to
768 other planetary boundaries (e.g., land system change) may alter the effective size and shape of the
769 hydrological corridor bounded by the water sub-boundaries due to feedbacks between the water cycle
770 and other components of the Earth System. This further highlights the need for iterative approaches
771 such as adaptive management which provide the opportunity to regularly update management
772 strategies and targets. In addition to its benefits, the planetary boundaries perspective may highlight
773 new governance needs at both local and global scales (Biermann, 2012; Galaz et al., 2012b, 2012a), for
774 instance fundamental incompatibilities between existing economic and environmental treaties
775 (Biermann et al., 2012). But many of the integrated assessment models currently used for resource use
776 assessments are essentially “black box” tools, where assumptions about human behaviour, social
777 structures, and economic priorities are invisible to the user. From an evaluation perspective, as more
778 societal actors become engaged in water management, care needs to be taken to avoid ‘greenwashing’ -
779 creating the perception of water-friendly practices without having tangible impacts on the control
780 variable of interest. Here, the harmonisation of fair shares and local safe operating space approaches
781 can help avoid greenwashing, by providing ways to evaluate local management based on quantified
782 changes in the value of a control variable, rather than practices intended or believed to have a positive
783 benefit.

784 6. Conclusions

785 The planetary boundaries are a useful framework for defining the global safe operating space for
786 humanity. However, the use of the planetary boundary for water management and governance at sub-
787 global scales encounters major challenges in reconciling the global-scale definition of the planetary
788 boundary with the smaller scales in which water decisions are made. Previous work to translate the
789 water planetary boundary to local contexts has primarily adopted either a fair shares approach, which
790 calculates the maximum allowable local contribution to the global planetary boundary and quantifies
791 the global responsibility (contribution) of the local water use; or a local safe operating space approach,
792 which uses the principles of the planetary boundary framework to define locally-relevant boundaries.

793 We present a harmonized approach to local use of the water planetary boundary which combines the
794 advantages of the fair shares approach (global responsibility) and the local safe operating space
795 approach (local relevance). This approach, based on the precautionary principle, can be used in both
796 socially-defined contexts (cities, nations, intergovernmental or companies or industries) and physically-

797 defined contexts (watersheds, aquifers, continents). Using these harmonized water sub-boundaries will
798 ensure that actions in a local context are contributing to water sustainability at all scales from local to
799 global. Integrating the water planetary boundary with existing water management and governance
800 approaches provides a framework to incorporate effects on water systems beyond the local, national, or
801 watershed context; integrates socio-economic and ethical considerations with biophysical constraints;
802 and provides a consistent approach for inter-regional comparisons. Furthermore, the water planetary
803 boundary further highlights the need for adaptive water management approaches which can respond to
804 complex, nonlinear changes in Earth System processes.

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811

812 **Tables**

813 **Table 1. Proposed water planetary sub-boundaries from Gleeson et al. (2019a, 2019b) which correspond**
 814 **to water stores and functions. Note that control variables are not defined.**

Water Store	Core function of this water store in Earth System	Possible response variable(s)	Possible control variable(s)	Cross-scale interaction not considered in traditional water management approaches
Atmospheric water	Hydroclimatic regulation	Climate pattern stability or land-atmosphere coupling	Land area with evaporation change	Changes in precipitation due to upwind changes in land use or irrigation (Keys et al., 2017)
	Hydroecological regulation	Terrestrial biosphere integrity	Land area with precipitation change	
Soil moisture	Hydroclimatic regulation	Carbon uptake or net primary productivity	Global root zone storage capacity	Trade-offs between global CO ₂ budget and local water availability (Heck et al., 2018b)
Surface water	Hydroecological regulation	Aquatic biosphere integrity	Watersheds or total river length within environmental flow limits	Importance of local aquatic systems to global biodiversity pool (Mace et al., 2014)
Groundwater	Storage	Terrestrial or aquatic biosphere integrity	Watersheds with sufficient low flows	Groundwater coupling with global climate system (Cuthbert et al., 2019)
Frozen water	Storage	Sea level rise	Volume of ice melt	Local responsibility for global sea level rise (Hardy & Nuse, 2016) and knock-on impacts on other planetary boundaries (Crépin et al., 2017)

815

816 Table 2. Previous use of the planetary boundaries in local contexts. Shaded rows indicate studies which
 817 did not define a water boundary.

Study	Context	Approach (all planetary boundaries)	Details related to water planetary boundary
Nykvist et al. (2013)	National (Sweden)	Fair shares: compare Sweden's performance to Rockström et al. (2009) boundaries using per-capita downscaling	<i>Control variable:</i> Consumptive blue water use <i>Response variable:</i> Not stated <i>Details:</i> Argue that fair shares approach not relevant for water since it does not account for regional variation in water availability
Cole et al. (2014)	National (South Africa)	Local safe operating space: develop local indicators and boundaries based on primary environmental concerns in South Africa	<i>Control variable:</i> Consumptive blue water use <i>Response variable:</i> Biosphere integrity (implicitly by defining boundary based on local environmental flow needs) <i>Details:</i> Quantified surface water availability after accounting for environmental flow needs, added estimated safe groundwater yield
Dearing et al. (2014)	Sub-national (two regions within China)	Local safe operating space: develop local control and response variables based on time-series data for primary environmental concerns	Water boundary was not defined - primary regional issues are water quality, air quality, and sedimentation.
Hoff et al. (2014)	National (all nations within European Union)	Fair shares: compare national performance to Rockström et al. (2009) boundaries using per-capita downscaling	<i>Control variable:</i> Consumptive blue water use <i>Response variable:</i> Not stated <i>Details:</i> Acknowledges limited utility of per capita downscaling, suggest future work should use context-specific factors such as environmental flows and also green water
Dao et al. (2015)	National (Switzerland)	Fair shares: compare national performance to Rockström et al. (2009) boundaries using per-capita downscaling with consideration of past and future generations	Water boundary was not defined - authors argue that only regional limits exist and global threshold not relevant.
Kahiluoto et al. (2015)	National (Sweden and Ethiopia)	Local safe operating space: estimated local biochemical flows boundary based on historical data Fair shares: per-capita downscaling of Carpenter & Bennett (2011) and Steffen et al. (2015) boundaries	Water boundary was not studied.

Sandin et al. (2015)	Industry (clothing in Sweden)	Fair shares: compare industry performance to Rockström et al. (2009) boundary using four ethical approaches	<p><i>Control variable:</i> Consumptive blue water use</p> <p><i>Response variable:</i> Not stated</p> <p><i>Details:</i> Compared four different ethical approaches to apportioning impacts. Note that framework could be adopted for regional, context-specific analysis.</p>
Fanning & O'Neill (2016)	National (Canada and Spain) and sub-national (Nova Scotia and Andalusia)	Local safe operating space: calculate carbon, nutrient, water, and land footprint relative to local thresholds	<p><i>Control variable:</i> Consumptive blue water use</p> <p><i>Response variable:</i> Biosphere integrity, implicitly by defining boundary based on local environmental flow needs following Steffen et al. (2015).</p> <p><i>Details:</i> Use threshold of blue water consumption exceeding blue water availability for >3 months annually</p>
Teah et al. (2016)	Sub-national (Middle reaches of Heihe River, China)	Local safe operating space: develop local indicators and boundaries based on primary environmental concerns in South Africa	<p><i>Control variable:</i> Consumptive blue water use</p> <p><i>Response variable:</i> Not stated</p> <p><i>Details:</i> Separately calculated low-risk and high-risk boundaries.</p>
Brejnrod et al. (2017)	Individual building (comparison of two types of houses)	Fair shares: calculate performance of a single building relative to the per-capita carrying capacity	<p><i>Control variable:</i> Consumptive blue water use</p> <p><i>Response variable:</i> Not stated</p> <p><i>Details:</i> Primary water savings could be obtained by reducing living area per person.</p>
Wolff et al. (2017)	Company (mass-market retailer)	Fair shares: Assess only one step in supply chain (agricultural production for food portfolio); full assessment of entire supply chain not feasible.	Water boundary was not defined - case study focused on biodiversity only.
Häyhä et al. (2018)	National (all nations within European Union)	Fair shares: compare per-capita and consumption-based downscaling of Rockström et al. (2009) boundaries	<p><i>Control variable:</i> Consumptive blue water use</p> <p><i>Response variable:</i> Not stated</p> <p><i>Details:</i> Suggest future work constrain freshwater use based on regional environmental flow limits</p>
O'Neill et al. (2018)	National (all nations)	Fair shares: compare national performance to Rockström et al. (2009) boundaries using per-capita downscaling	<p><i>Control variable:</i> Consumptive blue water use</p> <p><i>Response variable:</i> Not stated</p>

			<i>Details:</i> Acknowledge limitations of per-capita downscaling
Ryberg et al. (2018)	Industry (laundry washing within European Union)	Fair shares: compare national performance to Rockström et al. (2009) boundaries using four downscaling approaches	<i>Control variable:</i> Consumptive blue water use <i>Response variable:</i> Not stated <i>Details:</i> Downscaling approaches based on economic indicators (consumption expenditure and gross value added). Choice of downscaling approach significantly affected results.

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819

820 Table 3. Methods by which the water planetary boundary complements typical water management
 821 approaches in different contexts.

Context	Typical Management Approaches or Metrics	Value added by fair shares approach	Value added by local safe operating space approach
Watershed or aquifer management (single jurisdiction or transboundary)	Flow metrics, groundwater levels, allocations, IWRM	<ul style="list-style-type: none"> Integrating water with other Earth System functions, socio-economic, and ethical considerations Account for impacts outside the local context (global citizenship). 	<ul style="list-style-type: none"> Integrating water with other Earth System functions, socio-economic, and ethical considerations Considering water fluxes beyond traditional water system boundaries
Political jurisdiction (state, national, multi-national)	Water policy and regulations, trade agreements, treaties	<ul style="list-style-type: none"> Integrating water with other Earth System functions, socio-economic, and ethical considerations Account for impacts outside the local context (global citizenship). Provides consistency for comparing different countries or members 	<ul style="list-style-type: none"> Integrating water with other Earth System functions, socio-economic, and ethical considerations Considering water fluxes beyond jurisdiction boundaries
Commercial organization (corporation, industry)	Life cycle analysis, industry standards, water footprinting	<ul style="list-style-type: none"> Demonstrate commitment to global sustainability Provides consistency for comparing different companies or regions 	<ul style="list-style-type: none"> Evaluate resilience of supply chain

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