

1 **Inverse modeling of satellite observations shows that the wet tropics drive the 2010-2022**
2 **methane increase**

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Abstract

Atmospheric methane concentrations rose rapidly over the past decade and surged in 2020-2022 but the causes are unclear. We find from inverse analysis of GOSAT satellite observations that global methane emissions increased from 500 to 550 Tg a⁻¹ from 2010 to 2019 and surged to 570-590 Tg a⁻¹ in 2020-2022. Concentrations of tropospheric OH (the main methane sink) show no long-term trend over 2010-2019, but a decrease over 2020-2022 that explains 28% of the methane surge. The methane emission increase over 2010-2022 is mainly from the wet tropics with dominant anthropogenic and wetland contributions from Africa (43% of the global emission increase), South America (18%), Equatorial Asia (18%), and India and Pakistan (12%). Emissions from the US and Russia decreased slightly over the period. The 2020-2022 emission surge is consistent with the terrestrial water storage increase due to tropical inundation in Africa and Equatorial Asia associated with La Niña conditions.

Introduction

Methane (CH₄) is the second most important anthropogenic greenhouse gas after CO₂. Its concentration has tripled in the atmosphere since preindustrial time and has contributed a 0.6°C increase in global mean surface air temperature [Szopa *et al.*, 2021]. The global mean atmospheric methane concentration increased at a rate of 5 – 13 ppbv a⁻¹ from 2007 to 2019 after having stayed flat at 1770-1780 ppbv during 2000-2007. 2020-2021 saw unprecedented increase at a rate of 15-18 ppbv a⁻¹ [NOAA, 2023]. Despite extensive characterization and quantification of the global atmospheric methane budget, the drivers of these increases have remained uncertain.

Major sources of methane include fossil fuels, livestock, rice, wastewater, wetlands, and open fires [Saunio *et al.*, 2020]. The dominant loss is tropospheric oxidation by the hydroxyl radical (OH). Proposed explanations for the methane rise since 2007 have included increase in oil and gas emissions [Hausmann *et al.*, 2016; Franco *et al.*, 2016; Worden *et al.*, 2017], increase in livestock and wetland emissions [Schaefer *et al.*, 2016; Nisbet *et al.*, 2016; Y. Zhang *et al.*, 2021; Worden *et al.*, 2023], and decrease in OH concentrations [Turner *et al.*, 2017; Rigby *et al.*, 2017]. Atmospheric ¹³C-CH₄ isotope (δ¹³C_{CH₄}) trends suggest an increase in microbial methane emissions [Nisbet *et al.*, 2016, 2019; Lan *et al.*, 2021 a,b; Basu *et al.*, 2022; Oh *et al.*, 2022]. The surge in 2020 has been attributed to wetland emissions [Qu *et al.*, 2022; Feng *et al.*, 2022a; Drinkwater *et al.*, 2023] or to a decrease in OH resulting from air pollutant reductions during the COVID-19 shutdown [Laughner *et al.*, 2021; Stevenson *et al.*, 2022; Peng *et al.*, 2022]. The sustained surge in 2021-2022 has received little investigation so far.

Here, we use a global analytical inversion of GOSAT satellite observations of atmospheric methane to determine the factors driving methane changes from 2010 to 2022. GOSAT has been providing stable high-quality retrievals of atmospheric methane since 2009 [Buchwitz *et al.*, 2015; Kuze *et al.*, 2009, 2016; Parker *et al.*, 2020] and has been used extensively in global inverse analyses of the methane budget [Monteil *et al.*, 2013; Cressot *et al.*, 2014; Alexe *et al.*, 2015; Pandey *et al.*, 2016; Janardanan *et al.*, 2020; Stanevich *et al.*, 2021]. The analytical inversion optimizes methane emissions to fit the GOSAT observations in a Bayesian framework with closed-form expressions of information content and straightforward generation of inversion ensembles [Maasackers *et al.*, 2019; Y. Zhang *et al.*, 2021; Lu *et al.*, 2021; Qu *et al.*, 2021].

We perform independent Bayesian analytical inversions of GOSAT methane observations for each individual year over 2010 – 2022. For each year, the inversion optimizes a set of variables (henceforth collectively referred to as the “state vector”) consisting of (1) annual mean non-wetland methane emissions in land-containing $4^{\circ}\times 5^{\circ}$ grid cells (1009 elements), (2) monthly wetland methane emissions in 14 subcontinental regions (168 elements), and (3) annual hemispheric methane loss frequency against oxidation by tropospheric OH (2 elements). We use the GEOS-Chem global chemical transport model with $4^{\circ}\times 5^{\circ}$ horizontal resolution as the forward model in the inversion to relate the state vector elements to the dry column mixing ratios (X_{CH_4}) observed by GOSAT. Starting from prior estimates for all state vector elements, we produce optimized posterior estimates by drawing information from the observations following Bayes’ theorem (see Methods for details). We estimate the errors in the posterior estimates for each year from the spread of a 10-member inversion ensemble with varying inversion parameters. The prior emission estimates are 2010-2019 mean values from the Global Fuel Exploitation Inventory (GFEI) version 2.0 [Scarpelli *et al.*, 2022] for oil, gas, and coal exploitation; the EDGAR v4.3.2 inventory [Janssens-Maenhout *et al.*, 2019] for other anthropogenic sources; and the nine highest-performance members of the WetCHARTs v1.3.1 wetlands inventory ensemble [Ma *et al.*, 2021] for individual months. Loss of methane from oxidation by tropospheric OH is calculated with archived 3-D climatological monthly fields of OH concentrations from a GEOS-Chem full-chemistry simulation [Wecht *et al.*, 2014] and accounts for 86% of the total atmospheric loss of methane (additional minor sinks include oxidation by Cl atoms, stratospheric oxidation, and soil uptake). The prior estimates include no 2010-2022 trends in emissions or sinks, so that all trend information is from the observations. The GEOS-Chem inversion is initialized on January 1 of each year with GOSAT observations so that subsequent discrepancies with observations reflect emissions for that year. See Methods for more details.

Results

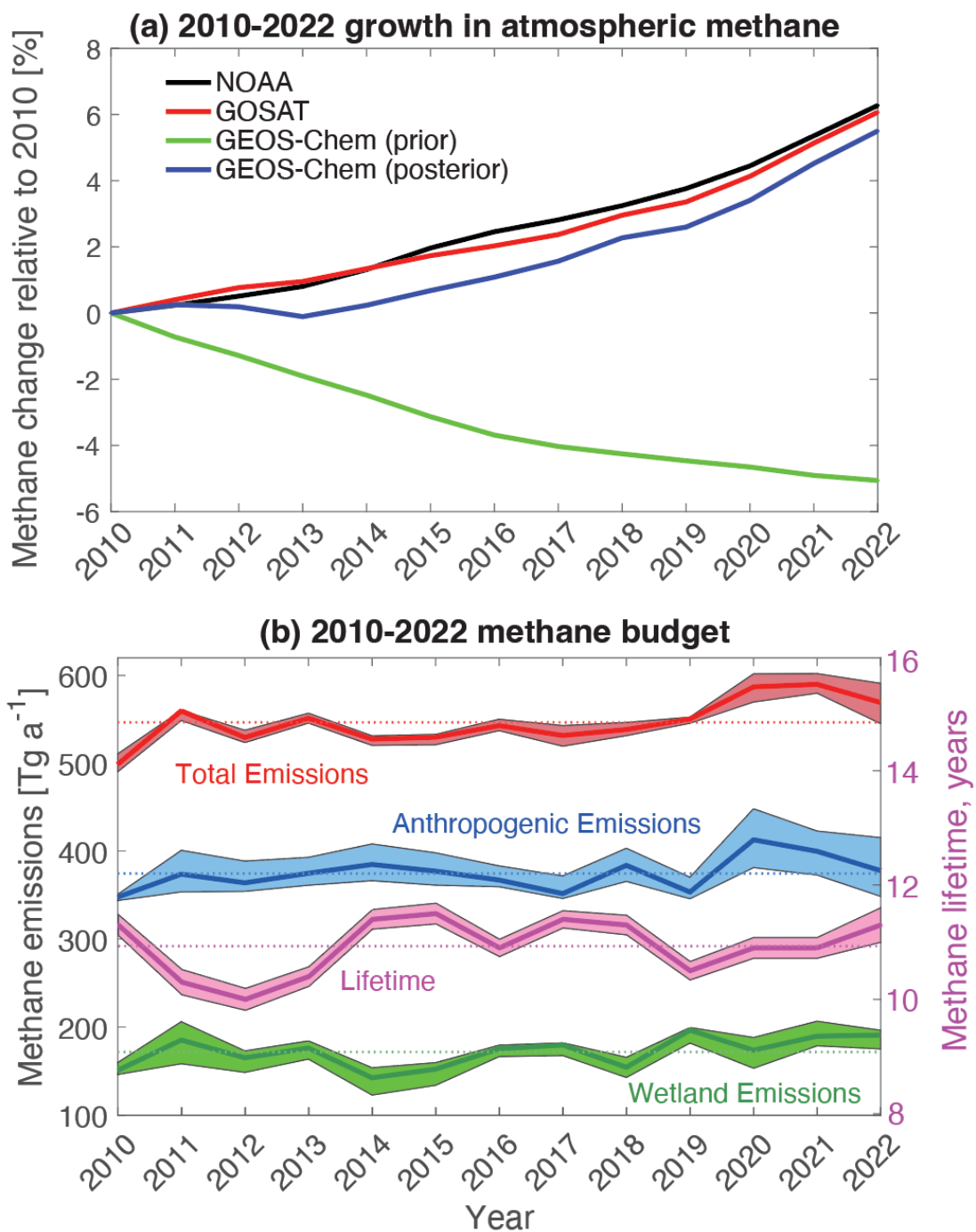
2010-2022 global trends of methane emissions and tropospheric OH

Figure 1 (a) shows the relative changes in global mean methane concentration from 2010 to 2022 as measured by GOSAT (dry column mixing ratios X_{CH_4}) and by the surface sites reporting to NOAA’s Earth System Research Laboratory [NOAA, 2023]. The GOSAT and NOAA increasing trends are consistent. The prior GEOS-Chem simulation shows a decrease during that period because the balance between prior emissions and sinks is negative. The posterior simulation with optimized methane emissions and OH concentrations effectively corrects the trend. It does not exactly match the observed trend because we optimize 1179 independent state vector elements, which provides a wealth of information for trend attribution but leads to some smoothing with the prior estimates. But the difference is sufficiently small, particularly in the second half of the period, that we can usefully exploit the posterior estimates to interpret the observed methane trend.

Figure 1(b) shows the posterior estimates of global methane emissions and OH oxidation lifetimes in individual years. These are the two variables driving the posterior trend in global methane concentration (Figure 1(a)). The narrow spread of the inversion ensemble suggests that we can separate the role of emissions and tropospheric OH in contributing to the trend. Global methane emission increased from 500 Tg a^{-1} in 2010 to 550 Tg a^{-1} in 2019 and then more steeply to $570\text{-}590 \text{ Tg a}^{-1}$ in 2020-2022. Anthropogenic emissions were stable at $350\text{-}380 \text{ Tg a}^{-1}$ in 2010-2019, increased to $400\text{-}410 \text{ Tg a}^{-1}$ in 2020-2021, and dropped back to 380 Tg a^{-1} in 2022. Wetland emissions increased from $150\text{-}180 \text{ Tg a}^{-1}$ in 2010-2018 to $170\text{-}200 \text{ Tg a}^{-1}$ in 2019-2022. This increase in wetland emission is consistent with the $13\text{-}26 \text{ Tg}$ increase in 2020-2021 relative to 2000-2006 in bottom-up wetland emission model estimates [Z. Zhang *et al.*, 2023].

125 The posterior estimate of methane lifetime against oxidation by tropospheric OH has no significant
126 long-term trend over 2010-2022. It has a mean value of 10.9 years, 7% longer than the prior estimate
127 of 10.2 years and consistent with the 11.2 ± 1.3 year estimate from the methyl chloroform proxy
128 [Prather *et al.*, 2012]. Methane lifetime has -10%/+5% interannual variability that drives variability
129 in the methane trend, including a 6% decrease in OH (increased lifetime) from 2019 to 2022 that
130 would contribute to the methane surge.

131
132 Our attribution of the 2010-2019 methane concentration trend to an increase in emissions, with no
133 significant long-term trend in OH, is consistent with previous inversions of GOSAT data for that
134 period [Maasackers *et al.*, 2019; Lu *et al.*, 2021; Y. Zhang *et al.*, 2021; Yin *et al.*, 2021]. The largest
135 difference between methane sources and sinks is in 2020 and 2021 (a growth rate of 48 Tg a^{-1}),
136 corresponding to the 15-18 ppbv a^{-1} surge in methane concentration [NOAA, 2023]. Tropospheric
137 OH concentrations decreased by 3% in 2020 relative to 2019, while methane emissions increased
138 by 7%. Several studies suggested that reduced NO_x emissions during the COVID-19 shutdown in
139 2020 decreased tropospheric OH [Laughner *et al.*, 2021; Stevenson *et al.*, 2022; Peng *et al.*, 2022]
140 but we find that this is within OH interannual variability over the 2010-2022 record. The methane
141 surge also extended into 2021, which we explain by sustained high emissions. The global methane
142 increase slows down by 17% in 2022 [NOAA, 2023] because of decreasing anthropogenic methane
143 emissions offsetting a decrease in OH.
144



145
 146 **Figure 1.** (a) 2010-2022 trends in global annual mean methane concentrations, expressed as
 147 changes relative to 2010. Results are shown for surface methane concentrations as reported by the
 148 Global Monitoring Division of NOAA’s Earth System Research Laboratory [NOAA, 2023], dry
 149 methane column mixing ratios (X_{CH_4}) measured by the GOSAT satellite instrument, and GEOS-
 150 Chem simulations of X_{CH_4} using prior and posterior estimates of methane emissions and OH
 151 concentrations. (b) Posterior estimates of 2010-2022 trends in global methane emissions and
 152 lifetime against oxidation by tropospheric OH. Thick lines show results from our base inversion
 153 and shaded areas show the spread of the 10-member inversion ensemble with different inversion
 154 parameters (see Methods for details). The thin dotted lines are mean 2010-2022 values.
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2010-2022 regional trends of methane emissions

Figure 2(a) shows the global distribution of the methane emission growth rates over 2010-2022 as linear regressions on the posterior estimates. The global rate of increase over that period is 5 Tg a^{-2} . The largest increases are in Africa (3 Tg a^{-2}), mostly in the central region of the continent with high wetland and livestock emissions. We also find increasing trends in Equatorial Asia (1.0 Tg a^{-2}), South America (1.0 Tg a^{-2}), India and Pakistan (0.8 Tg a^{-2}), and the Middle East (0.3 Tg a^{-2}). We find decreasing emissions in the US (-0.3 Tg a^{-2}) and Russia (-0.2 Tg a^{-2}). China (0.09 Tg a^{-2}) shows opposite emission trends over the North China Plain (increase) and South China (decrease), consistent with a shift in coal mining activity from south to north [Y. Zhang *et al.*, 2022].

Figure 2(b) shows that Africa accounts for 43% of the total emission increase from 2010 to 2022, 34% of the global anthropogenic emission increase, and 53% of the global wetland emission increase. The inversion can separate the 2010-2022 changes in anthropogenic and wetland emissions in Africa with moderate error correlations of $R = -0.6$ (see Equation 3 in the Methods). Other drivers of the global methane increase over 2010-2022 include South America (18% of the global total emission increase, 16% anthropogenic, 20% wetland, $R = -0.6$), Equatorial Asia (18% total, 9% anthropogenic, 22% wetland, $R = -0.5$), India and Pakistan (12% total, 17% anthropogenic, 2% wetland, $R = -0.5$), and the Middle East (6% total, 9% anthropogenic, 1% wetland, $R = -0.5$).

Methane emission trends (2010-2022) [Tg a⁻²]

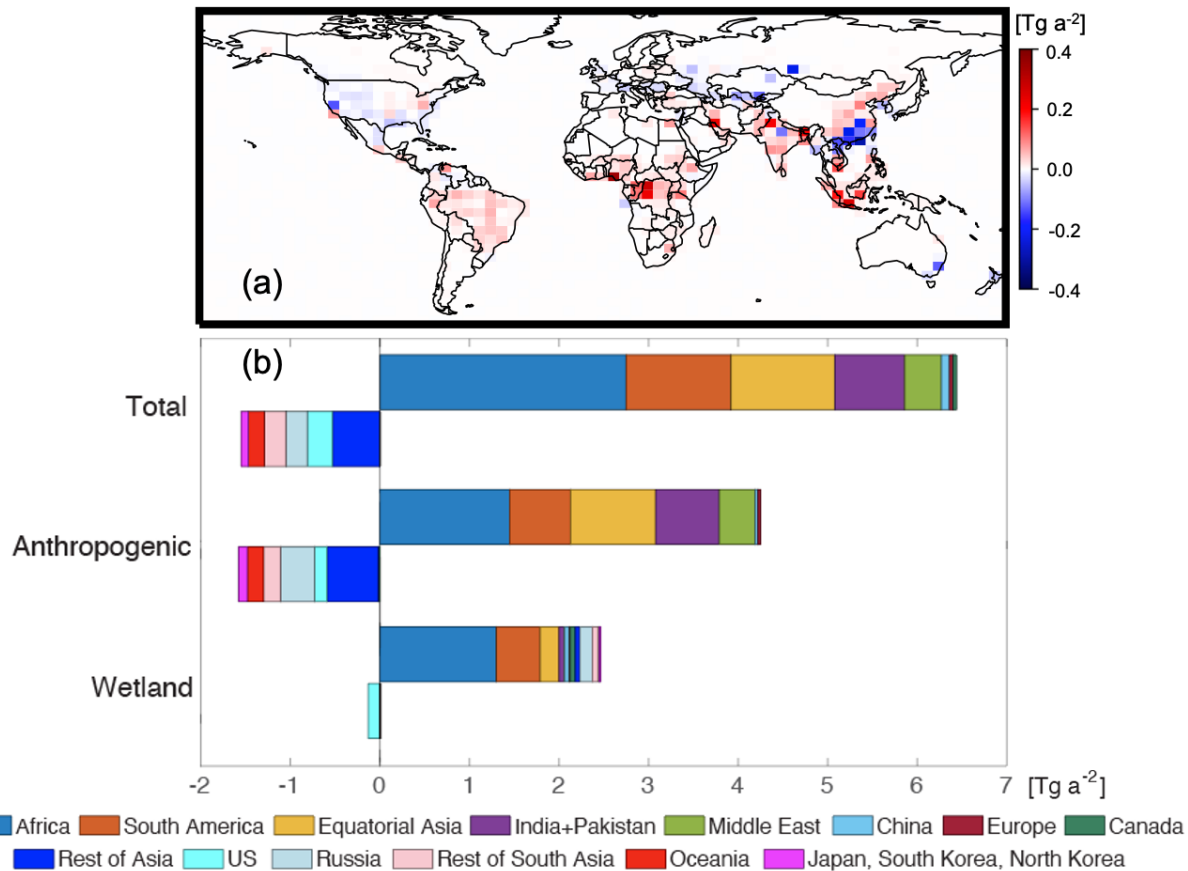


Figure 2. (a) Methane emission trends from inversion of GOSAT observations for 2010-2022. Values for each 4°×5° grid cell are fit by linear regression to the posterior emissions for individual years. (b) Emission growth rates by region for 2010-2022.

Figure 3 shows 2010-2022 emission trends for the four regions driving the global increase. 49% of the 2020-2022 surge is driven by Africa (Figure S1). Our inversion indicates a total (anthropogenic + wetland) emission increase in Africa of 20 Tg a⁻¹ from 2019 to 2021-2022, comparable with the 9-year increase of 23 Tg a⁻¹ from 2010 to 2019. Anthropogenic emissions in Africa have been increasing steadily over 2010-2022. This may be explained by rapid growth in livestock [Y. Zhang *et al.*, 2021; Worden *et al.*, 2023] and rice cultivation [Chen *et al.*, 2023]. Wetland emission increase in Africa has accelerated, from 3 Tg a⁻¹ over 2010-2017 to 14 Tg a⁻¹ over 2017-2022. Regions with the largest emission increase in Figure 2 (a) (East Africa, the Democratic Republic of the Congo, and Nigeria) represent 78% of wetland methane sources in Africa [Ma *et al.*, 2021]. Previous inverse studies of satellite methane data have pointed to an increase in African wetland emissions driven by increased inundation [Lunt *et al.*, 2021; Feng *et al.*, 2022a, b; Qu *et al.*, 2022]. Africa wetland methane emission trends are also consistent with observed terrestrial water storage trends from the Gravity Recovery and Climate Experiment satellites (GRACE and GRACE/FO; Figure S2), which provides a strong proxy for anaerobic methane production processes in wetland ecosystems [Bloom *et al.*, 2010, 2012].

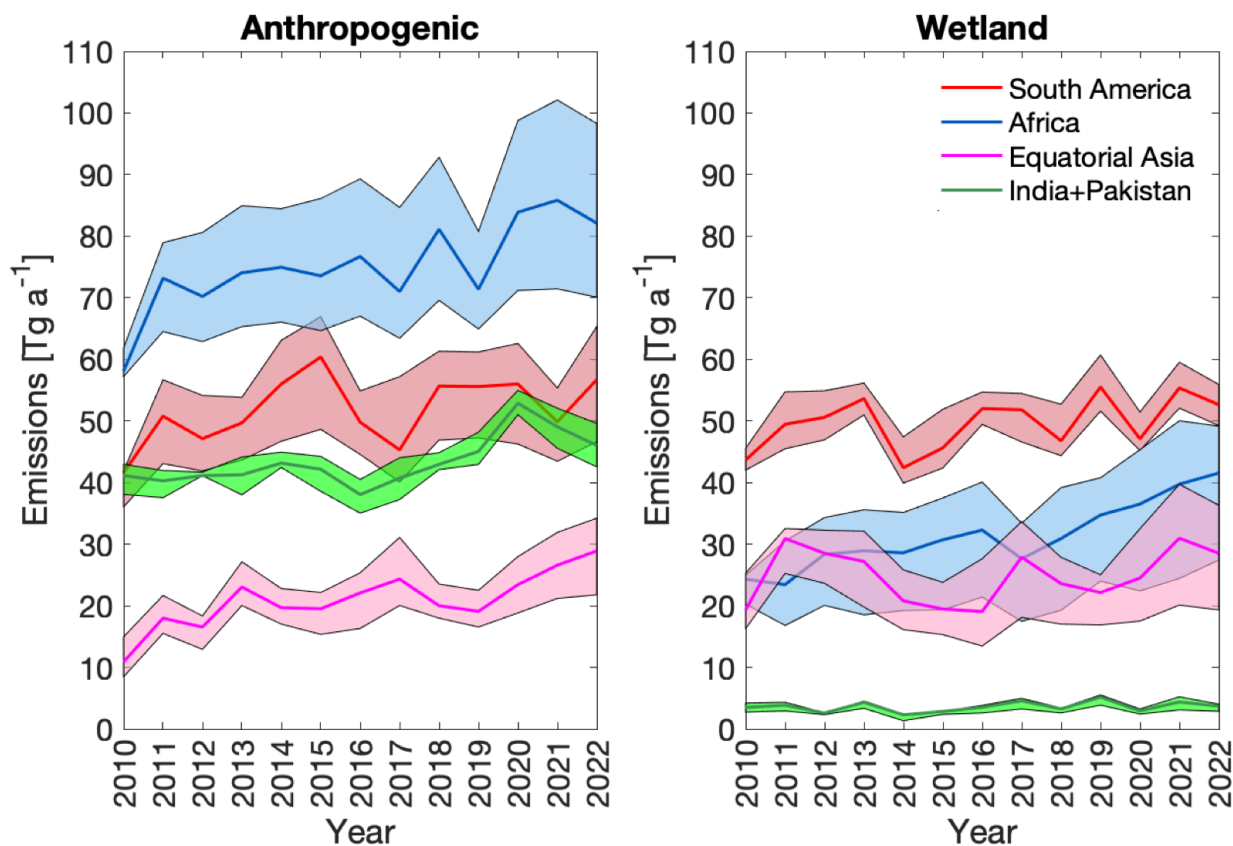


Figure 3. Regional trends of methane emissions informed by GOSAT for 2010-2022. Results are shown for the four regions driving most of the global emission increase. Lines show results from our base inversion and shaded areas show the spread of the 10-member inversion ensemble with different inversion parameters.

Equatorial Asia accounts for 41% of the methane emission surge in 2020-2022. Wetland emissions contribute to 50-65% of methane sources in this region according to our posterior estimates and have increased by 5 Tg a⁻¹ (47% of the anthropogenic + wetland emission increase) from 2010-2019 to 2019-2022. Similarly to the Africa wetland regions, we find a correlated trends in GRACE and GRACE/FO observed terrestrial water storage with peaks in 2011-2012, 2017, and 2021 that are consistent with wetland methane emissions (Figure S2). In 2020 and 2021, Equatorial Asia witnessed the highest precipitation over the past decade accompanied by devastating inundations [Hermawan *et al.*, 2022; World Bank, 2023a, b]. The bottom-up WetCHARTs methane emissions also show a 16% increase in wetland methane emissions in Equatorial Asia from 2019 to 2020. Anthropogenic emissions steadily increased over 2010-2019 with acceleration over 2020-2022. Rice cultivation is the largest contributor to the rise of anthropogenic emissions from 2010-2019 to 2020-2022 in Equatorial Asia (30%), followed by waste (15%), oil and gas (10%), and livestock (8%). Increasing rice emissions is consistent with the increase in inundation [Jain *et al.*, 2000].

India and Pakistan show no trend in methane emissions over 2010-2019 but an 8 Tg a⁻¹ surge in 2020 making that region an important contributor to the global methane surge. Posterior estimates show that livestock (50%), rice cultivation (25%), and waste (13%) are the drivers of the 2020 surge. Livestock and waste have been driving the 2009-2018 methane emissions in this region [Worden *et al.*, 2023]. Emission increase from rice cultivation may again be driven by its sensitivity to inundation. Variability in precipitation is a key factor governing rice emissions in India and Indonesia [B. Zhang *et al.*, 2016]. The 2019-2021 precipitations in India were the highest of the past decade [Statista, 2023], leading to floods across the country in 2020 [Gaon Connection, 2020;

233 *Reliefweb*, 2020]. Pakistan experienced its largest precipitation over the past decade in 2020,
234 including heavy flooding and stagnant water in Sindh, a major rice cultivation area [*USDA*, 2023;
235 *Dartmouth Flood Observatory*, 2020].

236
237 The heavy precipitation in these wetland and rice cultivation regions is likely associated with the
238 La Niña conditions from 2020 to early 2023. El Niño-Southern Oscillation (ENSO) can explain
239 much of the interannual variation of tropical wetland methane emissions [*Hodson et al.*, 2011; *Zhu*
240 *et al.*, 2017; *Z. Zhang et al.*, 2018]. *Hodson et al.* [2011] analyzed the correlation between their
241 wetland methane emission model and the ENSO index and found that La Niña events over 1950–
242 2005 led to maximum and mean increases of 14 Tg a⁻¹ and 8 Tg a⁻¹ of global wetland methane
243 emissions compared to the 1950-2005 mean value. These increases are smaller than our posterior
244 estimate of a 20 Tg a⁻¹ global increase in wetland emissions from 2010-2018 to 2019-2022. *Hodson*
245 *et al.* [2011] also found that the tropics (44%) and northern temperate regions (27%) are the largest
246 contributors to the global wetland interannual variability, whereas boreal wetland only accounts for
247 12%, consistent with our results.

248
249 South America is the largest contributor to the global methane emission increases prior to 2020 (26
250 Tg a⁻¹ total increase from 2010 to 2019) but does not make a specific contribution to the 2020-2022
251 methane surge. Livestock is the largest anthropogenic methane source in that region (58%). The
252 pattern of growth in Figure 2 is consistent with livestock inventories [*Chang et al.*, 2019; *Crippa et*
253 *al.*, 2019; *Janssens-Maenhout et al.*, 2019; *FAOSTAT*, 2023] and previous inversions [*Y. Zhang et*
254 *al.*, 2021]. The increases in wetland emissions are mainly from Amazon and north Peru.

256 Discussion

257
258 In summary, we find from the inversion of 2010-2022 GOSAT satellite data for individual years
259 that global methane emission increased from 500 Tg a⁻¹ to 550 Tg a⁻¹ over 2010-2019 followed by
260 a surge to 570-590 Tg a⁻¹ in 2020-2022. The methane sink from oxidation by tropospheric OH does
261 not contribute to the 2010-2019 trend, but decreasing OH contributes 28% of the observed methane
262 surge in 2020-2022. The largest growth rate in emissions over the 2010-2022 period was from
263 Africa (3 Tg a⁻², accounting for 43% of the global emission increase), followed by South America
264 (1.0 Tg a⁻², 18%), Equatorial Asia (1.0 Tg a⁻², 18%), and India and Pakistan (0.8 Tg a⁻², 12%).
265 Emissions in China, the US, and Russia were relatively flat. The emission surge in 2020-2022 is
266 mostly from Africa (49% of the surge) and Equatorial Asia (41%). We attribute the surge mainly
267 to emission increases from tropical wetlands and rice cultivation as a result of inundation associated
268 with La Niña conditions. One possible explanation for such an emission increase is the precipitation
269 changes associated with climate change [*Putnam and Broecker*, 2017; *Y. Zhang and Fueglistaler*,
270 2019]. Improved observations of wetland methane emissions [*Bloom et al.*, 2016] and process-
271 based attribution of methane responses to climate change [e.g. *Ma et al.*, 2023] would help
272 quantitatively resolve the role of changing precipitation patterns on tropical wetland ecosystems.

275 Materials and Methods

276
277 We use GOSAT 2010-2021 methane dry column mixing ratios (X_{CH_4}) from the University of
278 Leicester proxy product version 9.0 [*Parker and Boesch*, 2020]. GOSAT is in a polar sun-
279 synchronous low-Earth orbit. X_{CH_4} is retrieved using solar backscatter in the 1.65 μm shortwave
280 infrared absorption band. We include retrievals over land and glint retrievals over the ocean but
281 exclude data poleward of 60° because of the seasonal sampling bias, extensive cloudiness, and

282 greater sensitivity to variability in the stratosphere. We filter out low-quality retrievals using
283 “xch4_quality_flag” = 0.

284
285 We use the GEOS-Chem chemical transport model version 12.5.0 (10.5281/zenodo.3403111) at
286 $4^\circ \times 5^\circ$ grid resolution to relate methane emissions and OH concentrations to X_{CH_4} using the
287 averaging kernel and prior vertical profiles in the GOSAT retrieval. GEOS-Chem is driven by the
288 Modern-Era Retrospective analysis for Research and Applications, Version 2 (MERRA-2)
289 meteorological fields from the NASA Global Modeling and Assimilation Office (GMAO), which
290 use consistent model physics and data assimilation for 2010-2022. The methane simulation in
291 GEOS-Chem is described in *Wecht et al.* [2014]. Loss of methane from oxidation by tropospheric
292 OH is calculated with archived 3-D climatological monthly fields of OH concentrations from a full-
293 chemistry simulation [*Wecht et al.*, 2014]. Initial concentrations of methane on January 1 of each
294 year are adjusted to match GOSAT X_{CH_4} with a globally uniform scaling factor calculated using the
295 ratio of observed and simulated X_{CH_4} averaged from Jan 1 to Jan 10. This adjustment is not applied
296 to the results of Figure 1, which shows the 2010-2022 trends of GEOS-Chem with prior and
297 posterior budget terms in free-running simulations.

298
299 We use the same anthropogenic and wetland emissions as prior estimates for all years in the 2010-
300 2022 period. Anthropogenic emissions are 2010-2019 means from the Global Fuel Exploitation
301 Inventory (GFEI) version 2.0 [*Scarpelli et al.*, 2022] for oil, gas, and coal, and from the global
302 EDGAR v4.3.2 inventory [*Janssens-Maenhout et al.*, 2019] for other sources. Emissions for the
303 contiguous US are from the gridded version of the US Environmental Protection Agency (EPA)
304 greenhouse gas inventory [*Maasackers et al.*, 2016]. The seasonality of rice emissions from *B.*
305 *Zhang et al.* [2016] and of manure emissions from *Maasackers et al.* [2016] are applied locally to
306 the EDGAR annual totals. We use monthly wetland emissions from the nine highest-performance
307 members of the WetCHARTs v1.3.1 inventory ensemble [*Ma et al.*, 2021]. Monthly open fire
308 emissions averaged over 2010-2019 are from the Global Fire Emissions Database version 4
309 (GFED4) [*van der Werf et al.*, 2017]. Termite emissions are from *Fung et al.* [1991]. Geological
310 seepage emissions are from *Etiopie et al.* [2019] and are scaled to a global magnitude of 2 Tg a⁻¹
311 based on *Hmiel et al.* [2020].

312
313 The loss of methane from reaction with tropospheric OH has a corresponding methane lifetime of
314 10.2 years in the prior simulation. We also use archived Cl concentration fields from *Wang et al.*
315 [2019] for methane oxidation by tropospheric Cl atoms, 2-D monthly loss frequencies from the
316 NASA Global Modeling Initiative model [*Murray et al.*, 2012] for stratospheric oxidation, and the
317 Methanotrophy Model [*Murguia-Flores et al.*, 2018] for methane soil uptake.

318
319 We use water storage data from the GRACE and GRACE/FO Mascon version RL06.2 [*Save et al.*,
320 2016; *Save*, 2020], with all recommended corrections applied. Ocean water storage anomalies were
321 removed from regionally averaged GRACE time series over Equatorial Asia and Africa.

322 323 **Atmospheric Inverse Analysis**

324
325 The state vector \mathbf{x} of the inversion consists of 1179 elements for each year, including (1) annual
326 mean non-wetland methane emissions for land-containing $4^\circ \times 5^\circ$ grid cells (1009 elements), (2)
327 monthly wetland methane emissions for 14 subcontinental regions (168 elements), and (3) annual
328 hemispheric methane loss frequency against oxidation by tropospheric OH (2 elements). For each
329 year’s inversion, we perturb each of the state vector elements in 1179 GEOS-Chem simulations to
330 construct the full Jacobian matrix \mathbf{K} , which describes the sensitivity of the methane observations to

331 the state vector as simulated by GEOS-Chem. The posterior estimate with Gaussian error statistics
332 is obtained by minimizing the scalar cost function $J(\mathbf{x})$:

$$333 \quad J(\mathbf{x}) = (\mathbf{x} - \mathbf{x}_a)^T \mathbf{S}_a^{-1} (\mathbf{x} - \mathbf{x}_a) + \gamma (\mathbf{y} - \mathbf{K}\mathbf{x})^T \mathbf{S}_o^{-1} (\mathbf{y} - \mathbf{K}\mathbf{x}), \quad (1)$$

334 where \mathbf{x}_a is the prior estimate of the state vector, \mathbf{S}_a is the prior error covariance matrix, \mathbf{S}_o is the
335 observational error covariance matrix assumed to be diagonal, and γ is a regularization parameter
336 that accounts for the effect of unresolved correlation in the observational error. We construct \mathbf{S}_a ,
337 \mathbf{S}_o , and γ following *Qu et al.* [2022] and *Y. Zhang et al.* [2021]. By using the same prior estimate
338 for individual years (as opposed to a Kalman filter using the posterior estimate from the previous
339 year), we ensure that observations from each year have equal weight in the inversion and we make
340 no assumption of trend persistence from one year to the next.

341 The best posterior estimate is given by [Rodgers, 2000]:

$$342 \quad \hat{\mathbf{x}} = \mathbf{x}_a + (\gamma \mathbf{K}^T \mathbf{S}_o^{-1} \mathbf{K} + \mathbf{S}_a^{-1})^{-1} \gamma \mathbf{K}^T \mathbf{S}_o^{-1} (\mathbf{y} - \mathbf{K}\mathbf{x}_a). \quad (2)$$

343 with posterior error covariance matrix $\hat{\mathbf{S}}$:

$$344 \quad \hat{\mathbf{S}} = (\gamma \mathbf{K}^T \mathbf{S}_o^{-1} \mathbf{K} + \mathbf{S}_a^{-1})^{-1}. \quad (3)$$

345 This expression for the posterior error covariance matrix assumes that the inversion parameters are
346 correct but there is also uncertainty associated with those. These parameters in the base inversion
347 include 50% error standard deviation for anthropogenic emissions on the $4^\circ \times 5^\circ$ grid with no off-
348 diagonal correlations, 50% error standard deviation for wetland emissions with off-diagonal
349 correlations calculated from the WetCHARTs ensemble [Bloom et al., 2017; Y. Zhang et al., 2021],
350 10% error standard deviation for hemispheric annual OH concentrations, and $\gamma = 0.1$ to match the
351 expected chi-squared value for the contribution of the prior estimate terms to the posterior cost
352 function [Lu et al., 2021]. As an alternative approach to estimate errors, we conducted inversion
353 ensembles for each year varying parameters from our base inversion one at a time, including prior
354 error standard deviations for anthropogenic emissions (20%, 70%), wetland emissions (20%, 70%),
355 and hemispheric OH concentrations (5%, 20%), and regularization parameters ($\gamma = 0.05, 0.5,$ and
356 1). This results in a 10-member inversion ensemble including the base inversion. The spread of
357 results for that ensemble is larger than the posterior error standard deviation and we use it as a more
358 conservative error estimate on our inversion results [Qu et al., 2022; Chen et al., 2022].

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687 **Data and materials availability:** The GOSAT methane retrieval is available at
688 [https://www.leos.le.ac.uk/data/GHG/GOSAT/v9.0/CH4_GOS_OCPR_v9.0_final_nceo_2009_20](https://www.leos.le.ac.uk/data/GHG/GOSAT/v9.0/CH4_GOS_OCPR_v9.0_final_nceo_2009_2022.tar.gz)
689 [22.tar.gz](https://www.leos.le.ac.uk/data/GHG/GOSAT/v9.0/CH4_GOS_OCPR_v9.0_final_nceo_2009_2022.tar.gz) (last accessed Nov 19, 2023). The data that support the findings of this study are openly
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693 (last access Dec 1, 2023). All data are available in the main text or the supplementary materials.
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