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2 **Occurrences, distribution and sources of Pharmaceuticals and Personal Care Products**
3 **(PPCPs) in the mangrove dominated estuaries in the central part of the Indian**
4 **Sundarbans, a world heritage site.**

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27

28 **Abstract**

29 This study aimed to provide the baseline report on the occurrence of five emerging
30 contaminants namely Phenols, Triclosan, NSAIDs, Phthalate and Parabens under
31 pharmaceuticals and personal care products (PPCPs) in the surface water of Indian Sundarbans
32 estuarine system. The emerging contaminants of PPCPs were investigated from the two
33 estuaries in the central of Indian Sundarbans namely Thakuran and Matla covering entire
34 stretch during post monsoon season using gas chromatography mass spectrometry (GC-MS).
35 Among the different groups of PPCPs the most abundant found to be Phenolics with mean
36 concentration ranging from 230 to 4586 ng/L following Triclosan with mean concentration for
37 both Thakuran and Matla ranging from 96 to 1885 ng/L. The surface water concentration of
38 remaining PPCPs ranged widely; like for NSAIDs (290 to 791 ng/L), Phthalates (39-302 ng/L)
39 and Parabens (21-263 ng/L). Among the different Phenolics derivatives the highest
40 concentration was found for octyl phenol ranging 282.2 to 17590.1 ng/L followed by
41 Ketoprofen ranging 614.6 to 2333.7 ng/L and both derivatives observed to be high in the
42 seafront side of Matla estuary. For the Phthalates derivatives high concentration witnessed in
43 upstream of Thakuran and Matla i.e., for Di-2-ethylhexylphthalate ranging from 60.5 to 494.8
44 ng/L while for Parabens derivatives high concentration observed in downstream of Thakuran
45 i.e., Ethyl paraben ranging from 23.7 to 598.7 ng/L. The correlation of water quality parameters
46 with different PPCPs helps to elucidate the source as Phthalates are positively correlated with
47 DIN and DIP which clarifies its typical anthropogenic source from catchment areas, run off

48 from domestic sewage and effluents. Significant spatial variation observed for PPCPs
49 indicating persistence and long residence of these xenobiotics in the mangrove dominated
50 estuaries of Indian Sundarbans.

51 **Keywords:** Emerging contaminants, GC-MS, distribution, estuaries, mangroves,
52 environmental parameters

53

54

55 **1. Introduction**

56 Pharmaceuticals and personal care products (PPCPs) are rapidly been observed in aquatic
57 environments all over the World due to its wide applicability as daily usage and their growing
58 concern as these compounds showed potentially adverse effects on humans and biota
59 (Mimeault et al 2005; Shanmugam et al 2010; Ebele 2017; Ramaswamy et al 2011; 2018). This
60 diverse group of compounds are observed in the antibiotics, analgesics, steroids,
61 antidepressants, antipyretics, stimulants, antimicrobials, disinfectants, fragrances, cosmetics,
62 toothpastes, soaps, shampoos, detergents, lotions, sunscreens and many other chemicals (Wang
63 et al 2013; Liu et al., 2013., Sui et al 2015., Cui et al 2019., Zhang et al., 2020). PPCPs are
64 widely distributed in different compartments of the aquatic systems such as sediment, water
65 and ground water as they persist for longer time period, can be replenished easily, have
66 chemical properties which prevent them from rapid degradation and readily adsorbs to the
67 particle and accumulates in sediment (Wilson et al 2008; Sun et al 2016, Pinckney et al 2017;
68 Zhao et al 2019; Liao et al 2019). In recent years these emerging contaminants (PPCPs) in
69 marine estuaries, rivers and lakes have gained attention across the globe as they are Endocrine-
70 disrupting compounds (EDCs) that interfere with hormonal system, physiology of humans and
71 reproductive systems (Diamanti-Kandarakis, 2009; Schug et al.,2011; WHO/UNEP, 2013;
72 Zhao et al 2019) and due to the scarcity of study on the PPCPs distribution across the aquatic
73 ecosystem.

74 Transitional environments such as rivers, lagoons and estuaries are unique habitat for study as
75 are situated between land and sea harbouring many aquatic and marine organisms. There are
76 plenty of studies based on pollutants such as metals, dissolved nutrients, greenhouse gases,
77 pesticides, pathogens and biomonitoring using biological organisms as tracers and proxies
78 (Mukhopadhyay et al 2002., Na et al 2018., Dutta et al 2019., Mukherjee et al., 2020). The
79 class of emerging contaminants which have recently gained attention are Parabens (alkyl esters
80 of *p*-hydroxybenzoic acid) ranging from methyl to butyl or benzyl groups; Triclosan (5-chloro-

81 2 (2,4-dichlorophenoxy) phenol-TCS) are biphenyl ether; Phthalates (plasticizers) ranging
82 from dimethyl to di-n-butyl, benzyl, ethylhexyl and di-n-octyl phthalate; Phenols (phenolics of
83 nonyl, octyl, cumyl and bisphenol A,S,F); NSAIDs (non-steroidal anti-inflammatory drugs)
84 such as Ibuprofen, Ketoprofen, Diclofenac, Naproxen and etc. These emerging contaminants
85 are understudied in transitional environments such as estuaries which links the source to fate
86 of these pollutants. The factors influencing the distribution of PPCPs in highly dynamic
87 environments such as marginal marine habitats are understudied as this could provide the
88 understanding to the insights of potential risk associated to flora and fauna of the ecosystem.
89 The study of spatial or seasonal distribution of PPCPs profile at local sites to gain their local
90 specificities can help to enhance future environmental management programs concerning the
91 health of the ecosystem. This study of PPCP in Indian Sundarbans will be first-hand report
92 providing the baseline information to understand the contamination level and anthropogenic
93 impacts in these environments. The objectives of the present study to determine the occurrences
94 of different groups of compounds under PPCPs such as parabens, phenolics, phthalates,
95 NSAIDs and Triclosan which are polluting the pristine mangrove dominated estuarine
96 environment of Indian Sundarbans with an emphasis to identify their sources relating with the
97 environmental factors influencing their distribution in spatial scale.

98 **2. Materials and Methods**

99 **2.1. Study area**

100 The world's largest contiguous mangrove forest the Sundarbans stretches along the coasts of
101 Bangladesh (6017 km²) and India (4000 km²) which extends in the southern part of the state of
102 West Bengal in India. The Sundarbans declared as Ramsar site under Ramsar convention in
103 1992 (<https://rsis.ramsar.org/ris>) and World heritage site by UNESCO in
104 1997(https://www.sundarbanaffairswb.in/home/page/sundarban_biosphere) due to its unique
105 ecosystem services by providing the breeding and nursing habitats for diverse marine
106 organisms, houses endangered species such as Royal Bengal Tiger (*Panthera tigris*) and
107 Ganges river dolphins (*Platanista gangetica*), hotspot of flora and fauna diversity, mangroves
108 are the coastal foundations providing barriers against erosions and severe disaster such as
109 tsunamis and cyclones and lastly provides economy to 3.5 million of livelihoods which sustains
110 on resources and tourists attractions (Ellison 2000; Giri et al 2008; Cavanaugh et al 2015;
111 Kathiresan 2018) .

112 The Sundarbans ecosystems due to its huge biodiversity and ecological services are severely
113 affected due to pressures by humans as increasing the anthropogenic activities such as over-
114 exploitations of resources by practising pisciculture, aquaculture and salt farming, exploitation
115 of marine organism by fishing, cattle grazing, building dams, bridges and roads for various
116 purposes for example tourism, siltation, introducing mining and refineries, coastal pollutions
117 due to regular oil spills from huge ships and boats, constant change in hydrological regimes,
118 climate change and loss of biodiversity by human intrusions. Apart from anthropogenic
119 activities the ecosystem suffers due to natural stressors such as an increase in sediment salinity,
120 increase level of sulphide in sediment, anaerobic condition due to sea level rise, nutrient
121 limitations and continuous erosion by tides and total forest destruction by tsunamis and
122 cyclones. Sundarbans is highly irregular and is crisscrossed by numerous rivers and waterways
123 forming channels and creeks throughout the intertidal zone. The estuary having the main
124 connective of fresh water from Hooghly river which is called the artery of Sundarbans forming
125 swampy terrace characterized by fluvial marine deposits. The Indian Sundarbans are situated
126 at the land ocean boundary of the world's largest Ganga-Brahmaputra delta and the estuarine
127 phase is crisscrossed by several small and large rivers starting from west to east part of Indian
128 Sundarbans namely Mooriganga, Saptamukhi, Thakuran, Matla, Bidya, Gosaba and
129 Haribhanga to the extreme east side of the mangrove forest towards the boundary of India and
130 Bangladesh. In the Indian part of Sundarbans, the two major and extensive rivers are Matla and
131 Thakuran which are situated in mid of the ecosystem covering the latitudinal part from head to
132 sea mouth.

133 **2.2. Sampling**

134 Water samples for measuring levels PPCPs and other environmental parameters were collected
135 using *Niskin* water sampler (5L, General Oceanics) from stations located from head to sea end
136 of the two major estuaries of Indian Sundarbans namely, Matla and Thakuran. The stations
137 were located in transect wise to get the salinity variation from upper, middle and lower part of
138 the two estuaries (Fig. 1). The total number of stations in the Matla estuary were six and in the
139 Thakuran estuary were three covering their upper, middle and lower part. The samples were
140 collected during post monsoon months (December 2018 and February 2019) to avoid any
141 dilution with rainwater during monsoon and stable condition of estuaries during post monsoon
142 season as compared to pre-monsoon and monsoon.

143 The samples for PPCP were collected in pre-cleaned amber colored glass bottles and stored in
144 ice box for transporting to laboratory. The *in-situ* environmental parameters from each
145 sampling sites were measured on board like water temperature using mercury thermometer; 0-
146 50°C graduation, Zeal 76 MM immersion and pH using portable pH meter; Orion Star A211,
147 fitted with a Ross combination electrode calibrated on the US National Bureau of Standards
148 (NBS) scale with reproducibility of ± 0.005 pH units. Dissolved oxygen (DO) samples were
149 fixed in glass bottle of 125 ml using Winkler A (manganous sulphate 60%) and B (alkaline
150 iodide) reagents and analysed onboard following standard protocol (Grasshoff 1983). For
151 dissolved nutrients, water samples were collected in HDPE bottles after filtering through GF/F
152 filter paper (0.45 μm) and were stored in ice box for transporting to laboratory. Among the
153 dissolved nutrients nitrate, ammonia and phosphate were analysed. Moreover, salinity was
154 measured following Grasshoff (1983).

155 **2.3. Sample preparation and analysis**

156 All the reference standards of Phenolics, Triclosan, NSAIDs, Phthalates and Parabens and the
157 derivatizing reagent N-methyl-N-(trimethylsilyl) trifluoroacetamide (MSTFA) were purchased
158 from Sigma-Aldrich (USA). Phenanthrene-d10 was procured from Cambridge Isotope
159 Laboratories Inc. (USA). Acetone, n-hexane and ethyl acetate of HPLC grade were procured
160 from Qualigens Fine Chemicals (Mumbai, India). Sodium sulfate (anhydrous) obtained from
161 Hi Media Laboratory Pvt. Ltd. (Mumbai, India). Solid Phase Extraction (SPE) cartridges
162 (Strata C18-E, 1000 mg/6 mL) were from Phenomenex (Torrance, CA, USA). Glass vials of a
163 capacity of 2mL were purchased from Agilent Technologies (Waldbronn, Germany). The
164 Milli-Q water was obtained from Direct Q3-UV (Millipore, India). All glassware used for
165 sample analysis were precleaned with 10% Labolene solution, acid washed rinsed with
166 ultrapure water & acetone and dried in hot air oven. The individual standards stock solutions
167 of 100 $\mu\text{g/ml}$ or ppm of Phthalates, NSAIDs, Phenolics, Triclosan and Parabens were prepared
168 by dissolving 10 mg in 100mL of an acetone: ethyl acetate (1:1, v/v) solvent which were serially
169 further diluted to prepare the working standards in the range 10, 50, 100 ng/L. The working
170 standard solutions for calibration and recovery spike were stored at $-20\text{ }^{\circ}\text{C}$.

171 Five different PPCPs with different derivatives were analyzed in the present study. The
172 different class of PPCPs analyzed in present study were **Phenolics** (octyl phenol (OP), nonyl
173 phenol (NP), bisphenol A (BPA), BPS, BPF), **Triclosan**, **NSAIDS** (non-steroidal anti-
174 inflammatory drugs) (Ibuprofen, Naproxen, Ketoprofen, Diclofenac), **Phthalates** (Dimethyl

175 phthalate (DMP), Diethyl phthalate (DEP), Di-n-butyl phthalate (DBP); Butyl benzyl phthalate
176 (BBP) and Di-2-ethylhexylphthalate) (DEHP) and **Parabens** (methyl, ethyl, propyl and butyl
177 parabens). For the analysis of different PPCP different extraction procedure have been
178 followed. Triclosan extraction was done by liquid phase extraction following standard
179 protocol (Nishi et al 2008 and Ramaswamy et al 2011a). 1000 ml of water samples was taken
180 in pre-cleaned separating funnel of 2L and added with 10 g of NaOH and 50 ml of n-hexane.
181 The mixture was shaken well for 10 mins and allowed to stand for 30 minutes for separation
182 of layers. The n-hexane impurity layer was discarded and water phase was collected. The pH
183 of the water phase was adjusted to -2 by adding 6M HCL, following which 50 ml of n-hexane
184 was added and shaken again for 10 minutes. After the separation of n-hexane layer was
185 collected and pooled with the previous hexane extract. To the n-hexane extract 3g of Na₂SO₄
186 (anhydrous baked at 200°C overnight) was added for dehydration and left undisturbed for 30
187 minutes to remove moisture from the extract. The n-hexane layer was then transferred to
188 condensation flask and the extract was concentrated to 5ml by rotary evaporate 35 °C (BUCHI
189 R-210, Switzerland) and further dried to 1ml. The final extract is collected in sterilized glass
190 vials of 2ml capacity and stored in refrigerator at 4°C prior analysis in Gas Chromatography-
191 Mass Spectrophotometer (GC-MS)

192 For other PPCPs like phenolic, NSAIDs, Phthalates and parabens were analyzed following
193 solid phase extraction using selective cartridges like (OASIS HLB) and C1 8E. For phenolic,
194 Phthalates and parabens extraction cartridges were pre-conditioned passing 5ml ethyl
195 acetate, 5ml methanol, 5ml ultra-pure water at a constant rate of 3ml/min. While for NSAIDs
196 preconditioning was done using 3ml 1:1 (v/v) ethyl acetate and acetone; 3 ml methanol; 3 ml
197 ultra-pure water. 1000 ml of water samples was adjusted for pH to acidic range (2 to 3) for
198 phenolics and NSAIDs and to neutral for phthalates and parabens. Sample were passed
199 through the cartridges following which they were left for drying. for Cartridges were eluted
200 with 20ml of ethyl acetate and eluent was dehydrated with Na₂SO₄ (anhydrous baked at 200°C
201 overnight) The eluent was concentrated using rotavapor and transferred into sterilized glass
202 vials for derivatization. To the concentrated extract 25 µl of MSTFA (N-methyl-N-
203 (trimethylsilyl) trifluoroacetamide) (Ramaswamy et al.2011; Shanmugam et al. 2014, 2010;
204 Chakraborty et al 2018) was added and kept at 70°C for 30 minutes (for phenolics &
205 phthalates), 35 minutes (for NSAIDs) and 15 minutes (for Parabens). Derivatization is done to

206 make the compounds more suitable for GC-MS analysis as most compounds require
207 transformation into more volatile and thermally stable and achieved to sensitivity once
208 derivatized. The derivatized extract -injected into GC-MS using auto injector for qualitative
209 and quantitative analysis.

210 The identification of PPCPs was performed by gas chromatograph (GC-2010) interfaced with
211 a quadrupole mass spectrometer (QP-2010) (Shimadzu Corporation, Japan) where the
212 chromatographic separation of different PPCPs was achieved with DB-1 fused silica capillary
213 column (30m×0.32mm i.d., 0.25µm film thickness, J&W Scientific, Folsom, CA, USA).
214 Helium with a purity of 99.999% was used as the carrier gas at a flow rate of 2.25 mL min⁻¹.
215 An instrumental blank, procedural blank, blank spike, sample duplicate, and sample spike were
216 applied for each sampling. 1µL of derivatized extract was injected in split less mode using an
217 auto sampler. The injector port, interface, ion source and GC-MS temperature program are
218 mentioned in Ramaswamy et al 2011; Shanmugam et al 2010, 2013 and Selvaraj et al 2014. In
219 electron ionization (EI) mode at 70 eV the mass spectrometer was operated and at an emission
220 current of 60 µA, full scan data was obtained in a mass range of m/z 35–500. Scanning interval
221 and SIM (selected ion monitoring) sampling rate were 0.5 and 0.2 s with the mass selective
222 detector operated in SIM mode. The calibration of all the mentioned PPCPs such as Triclosan,
223 Phenols, Phthalates, NSAIDs and Parabens were done for three-point concentration such as
224 10, 50 and 100 ng/L. The calibration for all the compounds standards were achieved r²>0.99
225 with recovery percentage (R%) greater than 99% for mass ion fragments of PPCPs derivatives.
226 The detailed of the mass ion fragments with limit of detection (LOD) for each derivative of
227 PPCP standards are summarized in TS1 (supplementary material).

228 **2.5 Statistical analysis**

229 The box and whisker plot indicates the median and 25% to 75% of data variations. It gives the
230 overall range with minimum and maximum values indicated by whiskers of the five different
231 categories of PPCPs studied during post monsoon months for all sampling locations. To find
232 the relationship between PPCPs and the studied environmental parameters during the sampling
233 period Principal Component Analysis (PCA) was performed (Fig 4). In the PCA plot the first
234 axis itself explains 53.74% of data variability and the second axis i.e., Factor 2 represents 34.52
235 % of variability, thus the two axes together explain 88.26 % of total data variability. All
236 statistical analyses were performed using Statistica v. 7.0 platforms (StatSoft. Inc., 2004).

237 **3. Results**

238 **3.1. Spatial variation of PPCPs**

239 The different concentration of PPCPs for Thakuran and Matla estuaries are shown in box and
240 whisker plot (Fig 2) with median concentration and variations of different PPCP compounds.
241 Among the different PPCPs the most abundant was phenolics in the study area with mean
242 concentration ranging from 230 to 4586 ng/L following Triclosan with mean concentration for
243 both Thakuran and Matla ranging from 96 to 1885 ng/L. The remaining PPCPs in descending
244 range of concentrations are NSAIDs (290 to 791 ng/L), Phthalates (39-302 ng/L) and Parabens
245 (21-263 ng/L) (Fig 2). In the spatial distribution from head (i.e., upper) to mouth (i.e., lower)
246 of estuaries, Phenolics, Triclosan, NSAIDs and Parabens showed higher concentration in the
247 lower stretch of both Matla and Thakuran while only Phthalates showed high concentration in
248 upper part of both estuaries (Table-1). Among the five different PPCPs studied comparatively
249 higher concentration of Phenolics, Triclosan, NSAIDs and Phthalates are found in Matla as
250 compared to Thakuran while Parabens are higher in Thakuran estuary (Table-1 & 2). Phenolics
251 concentration for Matla lower (4586.26 ± 6607.29 ng/L) showed values four times higher than
252 upper and middle of Matla and Thakuran. Phenolics concentration for Thakuran showed
253 narrow variation in middle (1469.59 ± 1756.90 ng/L) and lower (1379.15 ± 1437.99 ng/L) part
254 of the estuary whereas the upper part showed very low concentration (169.81 ± 168.63 ng/L)
255 as compared other parts of both estuaries (Table-1). For Triclosan quite higher values are
256 observed in Matla lower (1885.22 ± 0 ng/L) and upper (1613.33 ± 354.92 ng/L) while lowest
257 value observed in Thakuran middle (95.65 ± 0 ng/L) and almost no variation in Thakuran upper
258 (356.52 ± 0 ng/L) and lower (395.65 ± 0 ng/L). For NASIDs higher values are found in Matla
259 lower (790.48 ± 902.80 ng/L) and Thakuran lower (617.33 ± 364.74 ng/L) with narrow
260 variation in concentration for upper and middle for both estuaries (Table-1). High concentration
261 of Phthalates is observed in Matla upper (233.59 ± 194.58 ng/L) and middle (237.10 ± 231.69
262 ng/L) with least value in Thakuran middle (39.10 ± 34.28 ng/L) and low value in Matla lower
263 (87.80 ± 70.45 ng/L). Parabens concentration in Matla showed low and narrow variation as
264 compared to Thakuran where highest values are observed in Thakuran lower (263.31 ± 294.15
265 ng/L) following Thakuran middle (215.44 ± 232.91 ng/L) and least in Thakuran upper (21.92
266 ± 9.28 ng/L) presented in Table-1.

267 Among the different derivatives of PPCPs the highest concentration was found for octyl phenol
268 ranging 282.2 to 17590.1 ng/L with highest values observed in Matla lower (17590.08 ng/L)
269 and least in Thakuran upper (282.19 ng/L) (Fig 3a). The next following derivative was
270 Ketoprofen ranging 614.6 to 2333.7 ng/L which showed narrow variations between Matla and

271 Thakuran upper and middle while the lower part of Matla (2333.75 ng/L) was almost double
272 in concentration to Thauran lower (1073.29 ng/L) (Fig 3b). Other derivatives of NASIDs did
273 not show huge variations between the two estuaries or in their spatial distributions from head
274 to mouth (Fig 3b). Among the Phthalates derivatives high concentration of Di-2-
275 ethylhexylphthalate (DEHP) was observed in Thakuran upper (691.79 ng/L) and Matla upper
276 (494.75 ng/L) while lowest concentration (3.77 ng/L) was found to be for benzyl butyl
277 phthalate (BBP) in Thakuran upper. High variations among all the Phthalate derivatives are
278 observed for Matla and Thaukuran middle with Matla middle Phthalate derivatives
279 concentration almost six times higher than Thakuran middle (Fig 3c). Among the Parabens
280 derivatives high concentration was observed for ethyl parabens (Eth-paraben) with marked
281 spatial variation in Thakuran estuaries ranging from upper (23.72 ng/L) to lower (598.75 n/L)
282 stations. For Matla and Thakuran upper stations did not show huge variation and the values
283 were low as compared to Thakuran middle and lower (Fig 3d). Though low concentration
284 among the Parabens derivatives were for methyl and propyl Parabens.

285 The average variation of five different PPCP compounds between two estuaries were compared
286 and it was observed that all the PPCPs except Parabens range was high in Matla as compared
287 to Thakuran estuary (Table-2). The marked variation between the two estuaries were mainly in
288 Triclosan and Parabens (Table-2). Very high concentration of Triclosan observed in Matla
289 (1237.83 ± 747.46) compared to Thakuran (282.61 ± 163.09) while opposite trend was observed
290 for Parabens where high the concentration found in Thakuran (166.89 ± 127.81) than Matla
291 (44.54 ± 24.63).

292 **3.2. Water quality parameters**

293 The environmental parameters of the two major estuaries of India Sundarbans i.e., Matla and
294 Thakuran show minor differences (Table-2). The surface water temperature ranged from 22 to
295 24 ° C during post-monsoon period. The variations between two estuaries for pH, DO, SPM
296 and DIP are not evident. However, dissolved inorganic nitrogen (DIN) for Matla (10.03 ± 5.27)
297 observed to be high as compared to Thakuran (7.44 ± 5.14) estuary while the Salinity found to
298 be high for Thakuran (25.01 ± 2.67) as compared to Matla (22.38 ± 3.43).

299 The PCA plot reveals Salinity, pH, DIN and DIP as the predominant elements of the
300 components for first axis while temperature mainly relates to the second axis. It has been
301 observed from the PCA plot that the Phthalates levels in the study period is mainly related to
302 DIN and DIP of water samples while parabens are positively related to salinity and pH of

303 estuarine water. The remaining PPCP i.e., Phenols, Triclosan and NSAIDs are related to
304 temperature but not with other environmental parameters (Fig 4).

305 **4. Discussion**

306 **4.1. Spatial distributions of Physico-chemical parameters and PPCPs: -**

307 Surface water temperature was typical for subtropical climate and salinity range is narrow
308 indicating polyhaline condition in these estuaries (Dutta et al., 2019). The oxygen level and its
309 saturation represent well oxygenated condition indicating lower waste water discharge into
310 these systems than urban estuaries. These estuaries within mangrove habitats exhibiting
311 representative pH values of seawater and nutrients level much lesser than fluvial estuaries in
312 the proximity like Hooghly estuary (Mukhopadhyay et al., 2006) indicating marine dominated
313 condition. Minor inter-estuarine variation of physico-chemical parameters represent their
314 similar and homogenous aquatic environmental settings, however lower salinity and higher
315 DIN values in Matla indicating higher amount of fresh water discharge into the system
316 primarily in the form of sewage and waste water.

317 By comparing the present data with national and international studies it has been observed that
318 the studied PPCPs such as Phenolics, Triclosan and NSAIDs are found to be moderate to high
319 concentration in Matla-Thakuran estuarine. Parabens and Phthalates in the present study
320 observed to be comparative lower ranges with national and international estuarine and river
321 systems (Table-3). Higher levels of Phenolics, Triclosan, NSAIDs and Parabens are at lower
322 stretch of the Matla-Thakuran estuarine system while Phthalates shows higher levels toward
323 the upper stretch of the estuary. High concentrations of these micro contaminants in the lower
324 stretch of estuary could be due to agricultural runoff, through natural processes such as
325 decomposition of organic matter and their high persistence level that these are transported
326 easily and are accumulated over time in such natural habitats (Bayen 2012; Bartons and
327 Peneulas 2017; Kinney and Heuvel 2020; Gan et al 2023)

328 **4.1.1. Phenolics**

329 Phenolics are Endocrine disrupting compounds primarily consists of 4-nonylphenol (4-NP), 4-
330 tert-octylphenol (4-t-OP) and bisphenol A (BPA), which are widely used as non-ionic
331 surfactants in industrial, agricultural and household applications (Diao et al 2017). Phenolics
332 levels in the present mangroves dominated estuaries is comparable to the other estuarine, lake
333 and river samples (Zhao et al 2009., Diao et al 2017 and Selvarja et al 2014) (Table-3). Few

334 other studies are also observed with 3-4 folds higher values of phenolics than the present study
335 such as the study of Wang et al (2013) who reported the total mean concentration of phenolics
336 ranging from 248-4650 ng/L with average concentration of 1384 ng/L in 22 river estuaries
337 around Dianchi lake, China. High concentration compared to the present study was also
338 reported from Pearl river estuaries China where the mean concentration of Phenolics in water
339 samples ranged from 233.04 to 3352.86 ng/L (Diao et al 2017). There are many other reported
340 values of phenolics much higher than the present studies from river for e.g., the study from
341 Spain where mean concentration reported 37300 ng/L (Cespedes et al 2005) and from Pearl
342 river estuaries, China with concentration of phenolics observed to be 11300 ng/L (Zhao et al
343 2009). There are few national studies which also reported higher concentration of phenols in
344 Indian rivers namely Kaveri, Vellar and Tamiraparani rivers where the most abundant
345 phenolics observed were NP ranged from ND to 2200 ng/L and OP ranging from ND to 16.3
346 ng/L (Selvaraj et al 2014). Previous national studies from Hooghly river-estuarine regions
347 reported higher values of BPA which showed 15 times higher values than the present BPA
348 concentration in Malta-Thakuran estuaries (Tabl-3). Among the different derivatives in the
349 present study high levels of OP is been observed following BPF compared to the global reports.
350 The overall trend of phenolics showed high concentration in Matla than Thakuran estuary
351 which might be associated with discharge of waste water as point sources and canal effluents
352 to Baidydhari river. There is always a huge difference between available treatment facilities of
353 the cities and generated waste water. By the reports of 2013 from Central Pollution Control
354 Board, 49% of waste water from four major districts namely Kolkata, North 24 Parganas,
355 Hooghly and Howrah are been directly discharged without treatment to rivers and estuaries.
356 Apart from the direct discharge of sewage and waste water effluents to estuaries and rivers the
357 recreational sites, tourists' boats, fishing trawlers and commercial fishing activities are also
358 adding to the pollution loads to such rural ecosystems (Mansson et al 2008 and Selvaraj et al
359 2014). In the present study the average concentration of phenolics derivatives in the estuarine
360 water samples ranged in order of OP > BPA > BPF > NP > BPS (Fig. 3a and Table 3). The
361 high concentration of OP in the water is also observed in river water of China (Wang et al
362 2013). There was no significant variation ($p= 0.33$) found in phenolics across the sampling
363 location from upper to lower stretch of estuaries as tested by Kruskal-Wallis test of analysis of
364 variance (Fig. 3) which might indicate the persistence nature in the system for the compound.
365 Many other studies state the higher concentration of Phenolics in estuarine water samples are
366 not only by general anthropogenic discharges such as sewage treatment plants, urban city
367 wastes or industrial effluents but might be due to some other factors such as sorption/desorption

368 by sediment, long time for biotransformation of phenolics (Wang et al 2013., Diao et al 2017)
369 and other the natural sources such as plant litter leaching as source of Phenolics (Sanyal et al
370 2020). Other studies reported that the half-life period of transformed phenolics group such as
371 nonyl phenol are more than months so are easily transported and persistent a long way by river
372 flow which might also results observing high concentration of phenolics in the present estuarine
373 water (Ying et al., 2002; L and Li, 2003, Xu et al 2006). The highest and abundant compound
374 observed in the present study is phenolics among other PPCPs indicates its long-term
375 biodegradation and persistent nature along with some natural sources as the decomposition of
376 organic materials of dead plants apart from only anthropogenic. In the present study the natural
377 sources can be from huge mangrove litter leaching in the present study as evident from optical
378 proxies ($S_{275-295}$ and $SUVA_{254}$) of more terrigenous type chromophoric dissolved organic
379 matter (CDOM) in Indian Sundarbans (Sanyal et al 2020), due to the paucity of the studies
380 regarding the mangrove litter leaching or the natural sources of phenolics into estuarine systems
381 it is difficult to conclude the dominant or exact natural source of phenolics in Indian
382 Sundarbans.

383 **4.1.2 Triclosan**

384 Triclosan are used as antimicrobial agents and are also used by many textiles industries as
385 additive to prevent odour, used in plastics as antimicrobial additive to protect from
386 deterioration, decolourisation and odour, used to eliminate house dust mites and by many other
387 industries to prevent the growth of bacteria and fungi (Sabaliunas et al 2003; Ramaswamy et
388 al 2011; Pintado-herrera et al 2014). The next highly abundant PPCP observed in the present
389 study is Triclosan (Fig. 2). Hence Triclosan has no natural sources and are added to the rivers
390 or estuaries by sewage, hospitals and waste water effluents (Wilson et al 2008; Ramaswamy et
391 al 2011; Wijnen et al 2018). The present study of Triclosan concentrations observed to be high
392 (up to 1885 ng/L) as compared to many global studies from estuaries and river (Bester 2005;
393 Hua et al 2005; Wu et al 2007; Pintado-herrera et al 2014) which reported concentration
394 between <1-300 ng/L. Among the different international studies till date highest reported levels
395 of Triclosan is 2300 ng/L from 139 sites of streams surface waters in United States (Kolpin et
396 al 2002). In comparison with the national studies the highest known concentration in natural
397 water bodies reported from Tamiraparani in South Indian state of Tamil Nadu where
398 concentration reported up to 51600 ng/L at Cheranmahadevi following 3800 ng/L at
399 Tirunelveli surface river water (Ramaswamy et al 2011). The reason for high concentration of
400 Triclosan in the study of Ramaswamy et al (2011) was from household or sewage effluents,

401 STP effluents, hospitals discharge and industrial activities. Triclosan levels are also higher in
402 Matla than Thakuran estuary alike with the phenolics trend across the estuary which clarifies
403 the fact that more sewage and urban wastes are directly added through canals and point sources
404 such as direct addition of untreated sewage waste water, tourists' activities and recreational
405 sites without treatment to the Bidyadhari river and furthermore it gets mixed with Matla
406 estuary. In the present study the Triclosan sources might be significantly from STP effluents
407 and from community sewers rather than direct outputs from large industries as the upper stretch
408 of the estuary do not show spike in concentration. Matla-Thakuran estuaries is surrounded by
409 densely populated districts with total population of Kolkata and North 24 parganas sum up to
410 25.5 million and 86 lakhs population of south 24 parganas which uses the ferry connections
411 along the river, supports fisherman livelihood and this estuarine region is a unique breeding
412 and nursery ground for fishes (Mandal et al 2019). The domestic and small-scale industries
413 effluents along with the southern Kolkata sewage waste water from treatment plants directly
414 gets dumped to the estuarine rivers and ultimately drains to Bay of Bengal. The higher values of
415 Triclosan in lower stretch of both the estuaries shows its high residence time with persistence
416 nature which gets accumulated and gets transported to longer distance.

417 **4.1.3 Non-steroidal anti-inflammatory drugs (NSAIDs)**

418 NSAIDs belongs to the mostly used pharmaceuticals in the humans and veterinary medicine.
419 The four types of NSAIDs in the present study are analgesics commonly known as pain killers
420 and are widely used as over the counter drugs (Hudec et al 2011). Each group of compounds
421 have varying degree of analgesics and anti-inflammatory properties depending upon their high
422 over the counter sales. The present study mean concentration of NSAIDs ranged up to 791 ng/L
423 (Fig. 2) which is comparable to most of the rivers and estuarine such as Togola and Budzinski
424 (2007); Silva et al (2011); Daneshvar et al (2012) and Shanmugam et al (2013), reported
425 concentrations from ND to highest 1060 ng /L. There were studies from rivers of Canada and
426 Taiwan which reported concentration higher up to 6400 ng/L (Brun et al. 2006; Selke et al.
427 2010; Lin et al 2010; Aydin and Talinli 2013; Marsik et al 2017) (Table-3). Highest recorded
428 NSAIDs concentration in natural water bodies is from Nairobi river basin Kenya which
429 reported Ibuprofen concentration varying from 10000 to 30000 ng/L (Koreje et al 2012). In
430 the present study the most abundant NSAIDs is Ketoprofen following ibuprofen, diclofenac
431 and naproxen (Fig 4), similar high concentration of Ketoprofen and Ibuprofen concentrations
432 are reported in Ebro river basin, Spain (Silva et al 2011) and from national study from Kaveri,
433 Vellare and Tamiraparani, India (Shanmugam et al 2013). In many previous studies the high

434 concentration of these drugs is not only due to higher usage but lack of efficiency or functional
435 unit of STP in urban areas near the estuaries or rivers and difficult in removal of ketoprofen
436 and ibuprofen during sewage treatments. The mean concentration of NSAIDs from head to
437 mouth of the estuaries of both Matla and Thakuran also did not show wide variation. Though
438 slightly higher values are observed at the lower stretch of both the estuaries likely due to the
439 high residence time of these compounds and is readily not removed by physical factors such as
440 oxidation or irradiation.

441 **4.1.4 Phthalates**

442 Esters of phthalic acid are of environmental relevance due to their ecotoxicological potential
443 and high production rates as they are used in many chemical industries and representatives of
444 high production volume chemicals (Clara et al 2010; ESIS, 2019). Phthalates are plasticizers
445 which are mainly used as polymers in production of plastic products like polyvinyl chloride
446 (PVC), vinyl flooring, adhesives, lubricants, varnishes and paints, gelling agents, dispersants,
447 packing materials of toys, textiles, detergents, medical devices, stabilizers, binders, lubricating
448 and emulsifying agents, etc. (Latini 2005; EU-RAR 2008; USEPA 2012, Selvaraja et al 2014).
449 The results of phthalate mean concentration in Indian Sundarbans mangrove ecosystem ranges
450 from 39 to 302 ng/L (Fig. 2) while the upper part of both the estuaries i.e., Matla and Thakuran
451 showed high concentration of phthalates than the lower part of estuaries (Table- 1 and 2). The
452 range of phthalate in the present study is found to be low as compared to the national study by
453 Selvaraja et al (2014), reported phthalates from Kaveri river, India ranging from ND to 822
454 ng/L while the present range of phthalate concentration are found to be low for certain
455 derivatives as compared with the other international studies such as Dargnat et al. 2009, from
456 Seine river, France; He et al (2011) from Yangtze, China; Santhi and Mustafa (2013) from
457 Selangor river, Malaysia (Table-3). Though many global riverine concentrations of phthalates
458 are reported high up to 5000 ng/L (Vethaak et al 2005) in Dommel river, Netherlands, from
459 Spain rivers namely Muga, Fluvia, Ter, Besos, Llobregat, Ebro concentration reported as high
460 as 4980 ng/L (Sanchez-Avila et al. 2012) and from Eleven point white river, USA reported
461 concentration up to 4140 ng/L (Solís et al. 2007). Few studies have reported low concentration
462 of phthalates ranging from 10 to 60 ng/L in Hoje Sweden by Bendz et al. (2005) (Table-3).
463 High concentrations of phthalate in the present study are DEHP following DBP, DEP and least
464 observed was DMP (Fig 3c and Table-3) which is comparable with the study of Hooghly river
465 estuary where highest levels was reported for DEHP followed by DEP and DBP
466 (Mukhopadhyay and Chakraborty 2021). The most abundant and ubiquitous phthalate in the

467 present study is DEHP, DEP and DBP showed comparable range of concentrations reported
468 from other national and international studies such as from Yangtze river, China (He et al. 2011);
469 Rhine, Elbe, Ruhr, Mosel, Havel, Spree, Oder rivers, Germany (Fromme et al. 2002) and
470 Kaveri river, India (Selvaraja et al. 2014). The stations which are located in the upper part of
471 the estuaries of Matla and Thakuran showed higher levels of phthalates than the stations
472 situated close to the sea mouth (Fig 3) which indicates that the main sources of Phthalates are
473 from sewage effluents of urban, household, stormwater runoff and agricultural runoff rather
474 than any point or local sewer sources with low chance of getting diluted with marine water.

475 **4.1.5 Parabens**

476 Parabens are ubiquitously found in estuarine and river waters as these are present in daily usage
477 products such as shampoos, shower gels, toothpastes, creams, lotions, soaps, fragrances, UV-
478 filters, preservative chemicals in food and beverages due to their antifungal and antibacterial
479 potentials (Darbre et al 2004; Zhang et al 2011; Ye et al 2008; Alan 2008; Blanco et al 2009;
480 Liao et al., 2013; Zhao et al 2019). Since the low cost and wide applicability of Parabens
481 containing products on nails, skin, scalp, lips, mucosae and in food preservations lead to
482 continuous exposure of human and are directly or indirectly are discharged to natural water
483 bodies such as rivers, lakes, estuaries and ultimately to sea. In the present study amongst the
484 different groups of PPCPs the lowest concentration observed is for Parabens. In the present
485 study parabens ranges from 22 to 263 ng/L from upper to lower stretch of both the estuaries of
486 Indian Sundarbans (Table 1 and Table 2). The concentration of Parabens towards the lower
487 stretch of the estuary was found to be more and the range of the present study is slightly higher
488 as comparable to the national study of Ramaswamy et al (2011) from Kaveri, Vellare,
489 Tamiraparani rivers and Pichavaram mangrove, south eastern part of India and in comparable
490 range with international study of Hasegawa et al. (2016) Japan and from Urban streams in
491 Tokushima and Osaka, Japan (Yamamoto et al, 2011). Highest concentration (up to 52100
492 ng/L) of parabens till date is reported from Mogi Guaçu River, Brazil which is almost 20 folds
493 higher than the present study following the study of Peng et al. (2007) from Pearl river delta,
494 South China reported concentration up to 3142 ng/L. Lower concentration of Parabens i.e. <
495 50 ng/L are also reported from the rivers of South Wales, UK (Kasprzyk-Hordern et al. 2008);
496 North-eastern Switzerland (Jonkers et al.2010); Galicia, Spain (Gonzalez-Mariño et al., 2009);
497 Greater Pittsburgh area, USA (Renz et al., 2013); Pearl river estuary (Zhao et al 2019). Many
498 similar studies with high and low levels of Parabens with their different derivatives are given
499 in Table-3 and in comparison, Parabens levels in Matla-Thakuran estuary is found to be high

500 for its some derivates. (Fig 3d). The most abundant Parabens found in many studies is ethyl
501 parabens which is also observed in all sampling locations of the present study (Fig 3d). The
502 level of Parabens concentration shows higher values in the lower stretch of both the estuaries
503 and changes sharply for Thakuran estuary lower (21.91 to 263.30 ng/L) than Matla (37.98 to
504 48.14 ng/L) in Table-1 which might be due to point source of effluents to Thakuran lower. The
505 high values of Parabens in Thakuran can be due to the dumping of untreated wastes of
506 household sewage from nearby villages or any local point source which results in huge
507 variation and higher concentration. Though higher values are witnessed in the lower stretch
508 which might indicate the persistence nature with longer residence time and not easily
509 biodegraded but transported without removal to longer stretch of the estuary.

510 **4.2. PPCP relationship with environmental parameters**

511 Very few studies have analyzed the relationship of PPCPs with water quality parameters such
512 as temperature, salinity, pH, dissolved oxygen, dissolved nutrients, suspended particulate
513 matters and etc. (Glassmeyer 2005; Yang et al 2013; Lv et al 2014; Sun et al 2016). The
514 relationship of these PPCPs with physic-chemical parameters of estuarine water helps to
515 understand the source and nature of these compounds and can be used as sewage markers for
516 the ecosystem (Sun et al 2016). From the present analysis it has been observed that Phenols,
517 Triclosan and NSAIDs are positive and significantly correlated with temperature while
518 phthalates are positively correlated with dissolved inorganic nitrogen and phosphate. The
519 relationship with PPCPs with temperature has previously been reported as key factors in
520 degradation of these long chain organic compounds (Xu et al. 2006; Cailleaud et al. 2007). The
521 positive relationship of Phthalates with DIN and DIP may clearly show the source as
522 allochthonous such as effluents, domestics and run off from catchment areas. As in case of
523 Phthalates high levels are observed at the upper stretch of the estuaries which indicates its total
524 anthropogenic inputs and readily does not mix or transported to longer stretch of estuaries. The
525 relationship of Parabens with pH and salinity might indicate the mixing and dilution with sea
526 water and their persistence capacity in the system for longer period. Studies have reported about
527 the biodegradation process as primary mechanisms for removal of such derivatives depending
528 upon the pH and dissolved oxygen concentration of water sample which accelerates the
529 metabolic breakdown of these compounds. Water samples with low pH could make difficulties
530 in the hydrolytic transformation of PPCPs while aerobic conditions facilitate easier
531 biotransformation than anaerobic condition (Joss et al. 2005; Lonappan et al. 2016; Marsik et
532 al 2017). The positive relation of Parabens with salinity and pH in the present study might

533 indicate its longer residence time which do not get easily degraded and persists in the
534 environments. Salinity shows positive significant correlation with Phenolics and NSAIDs
535 which might be due to the change in solubility of these compounds in higher salinity and
536 alkaline nature of seawater. Triclosan shows negative significant correlation with salinity
537 which might be due to “the salting out” effect where the aqueous solubility of neutral non polar
538 compounds decreases in presence of major sea water ions such as Na^+ , K^+ , Cl^- and SO_4^{2-} (Zhou
539 and Liu, 2000).

540

541 **5. Conclusion**

542 The study presents the first-time report on the occurrences, distribution and sources of
543 pharmaceuticals and personal care products (PPCPs) from the estuarine water of Indian
544 Sundarbans. Results showed comparative ranges with some of the international river and
545 estuarine systems along with national studies. The most abundant PPCP found in the present
546 study is phenolics which showed persistent nature with longer residence time and high
547 accumulation in the natural habitat as higher values observed in lower stretch of estuaries and
548 few stations located at the upper stretch of estuary. The spatial distribution of PPCPs along the
549 major estuaries of Indian Sundarbans showed significant variation for parabens while other
550 PPCPs such as phenolics, triclosan and NSAIDs did not show marked variation or trend from
551 head to mouth of the estuary. Though decreasing trend in concentration for phthalates were
552 observed from upper to lower stretch of the estuary. Phthalates shows positive relation with
553 DIN and DIP which elucidates its typical anthropogenic sources from catchment areas, direct
554 runoffs of plastic wastes from domestic and small industrial effluents along with storm runoff
555 to the estuaries. The positive relationship of parabens with pH and salinity might states that
556 this compound is difficult to remove by physical or biological processes or sea water dilution
557 and can be transported to long way which persists in the environment due to its longer residence
558 and higher accumulation nature. Thus, these compounds can be used as specific markers for
559 sewage and effluents as these markers can help to monitor the ecological status of the
560 ecosystem and could delineate the fact of improper function of STPs of the cities. Further
561 studies are needed to understand the specific discharge of these contaminants from urban pool
562 mainly from sewage treatment plants to test their efficiency in successful removal of these
563 micropollutants which gets diluted to the estuaries and rives or are adsorbed to sediments
564 during transportation.

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575

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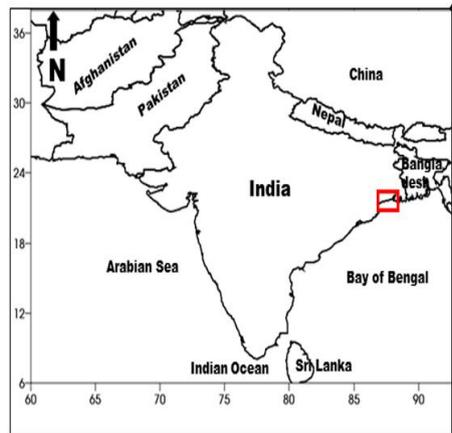
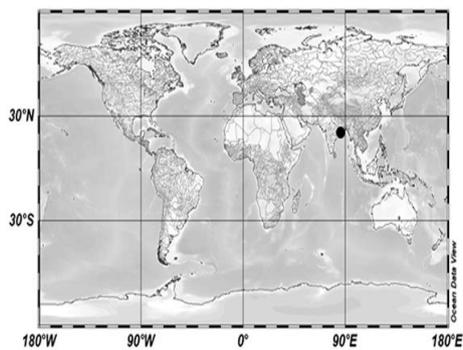
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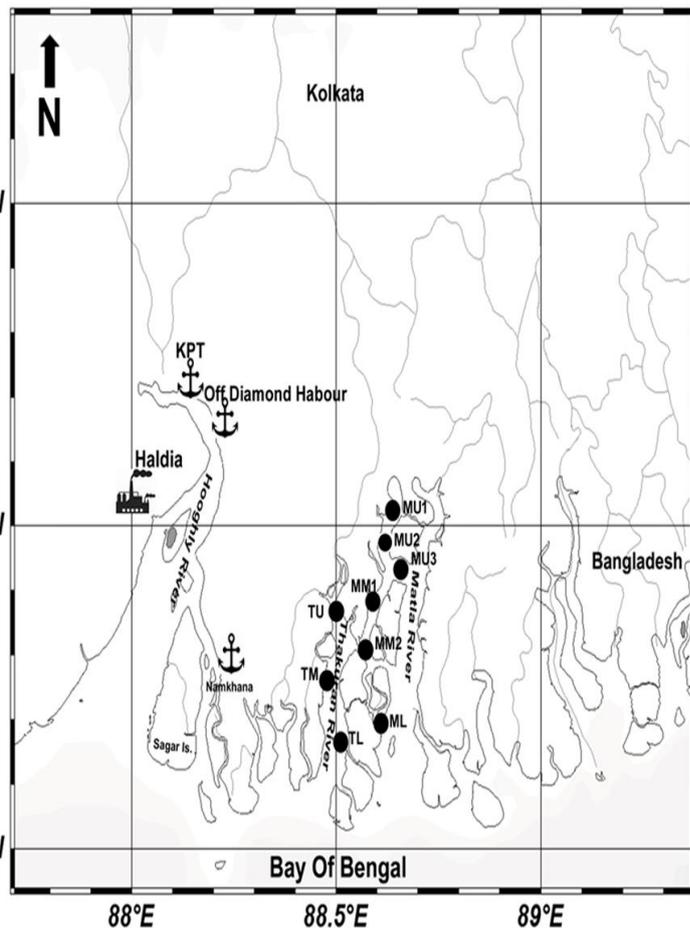
590 **Fig. 1** Map of sampling location in Indian Sundarbans mangrove ecosystem.



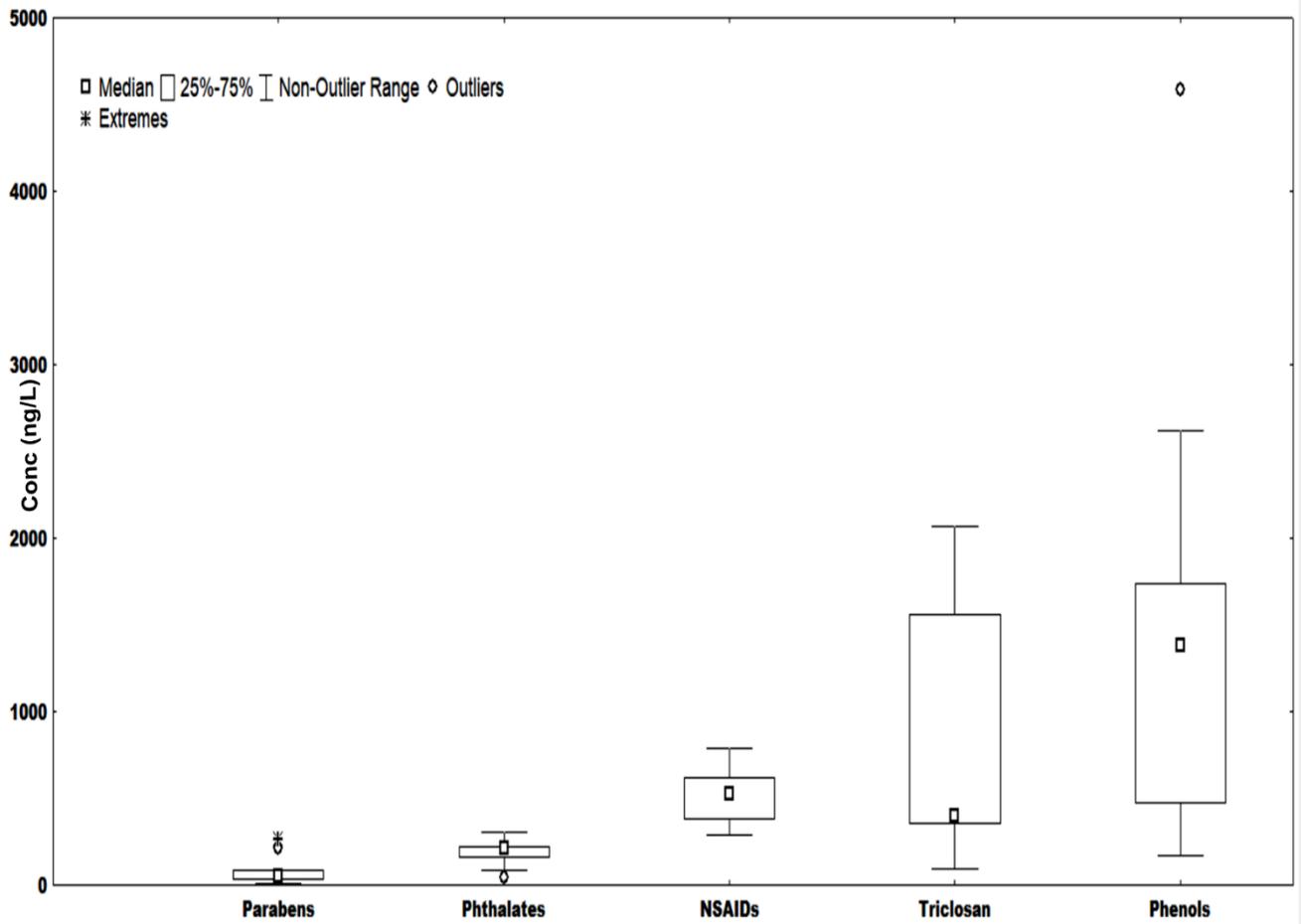
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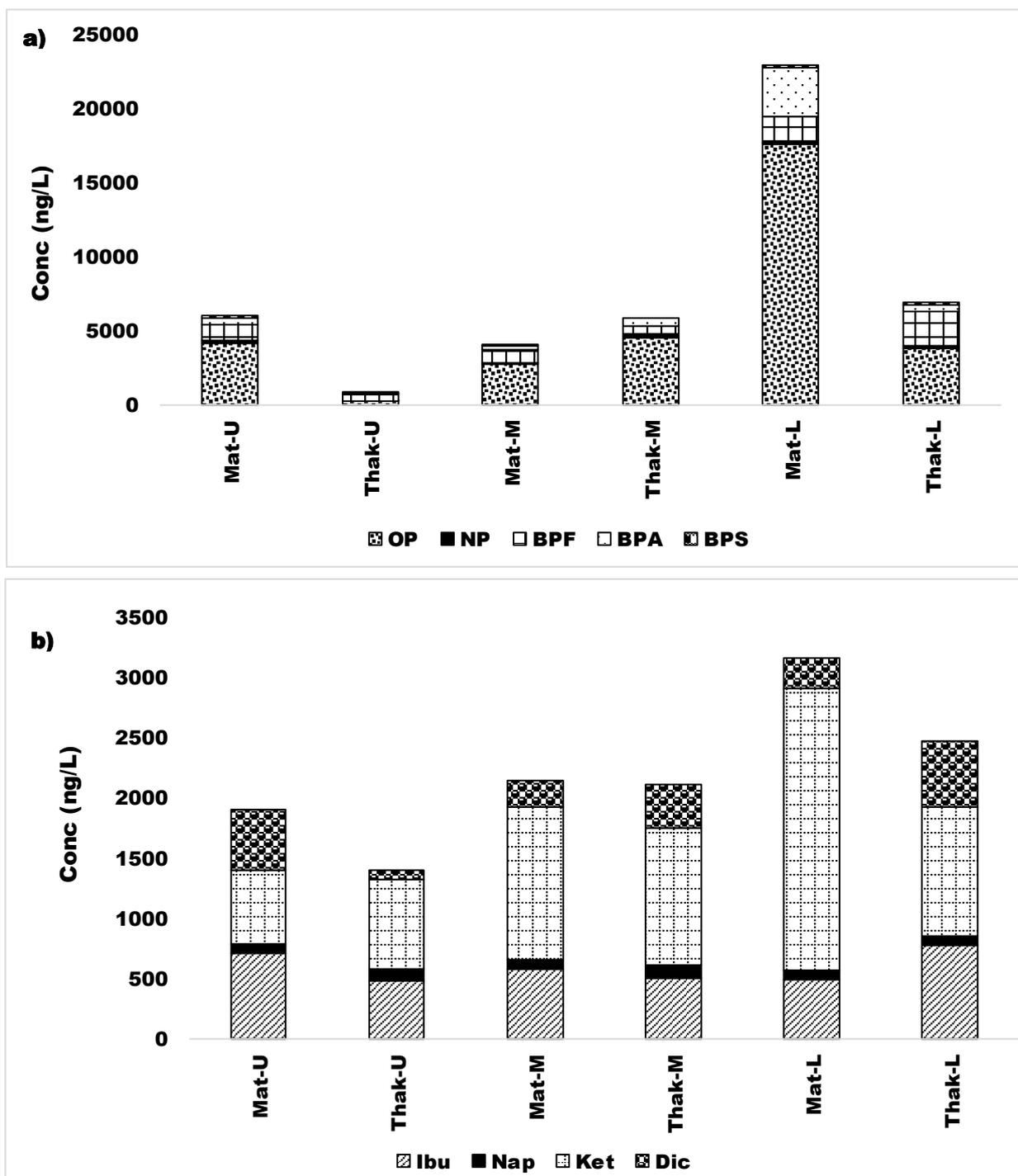


592 **Fig 2.** Box and whisker plot of five different studied PPCPs. The point inside the box denotes
593 the median concentration of PPCP of each group, the box denotes the 25 and 75 percentiles
594 of data.

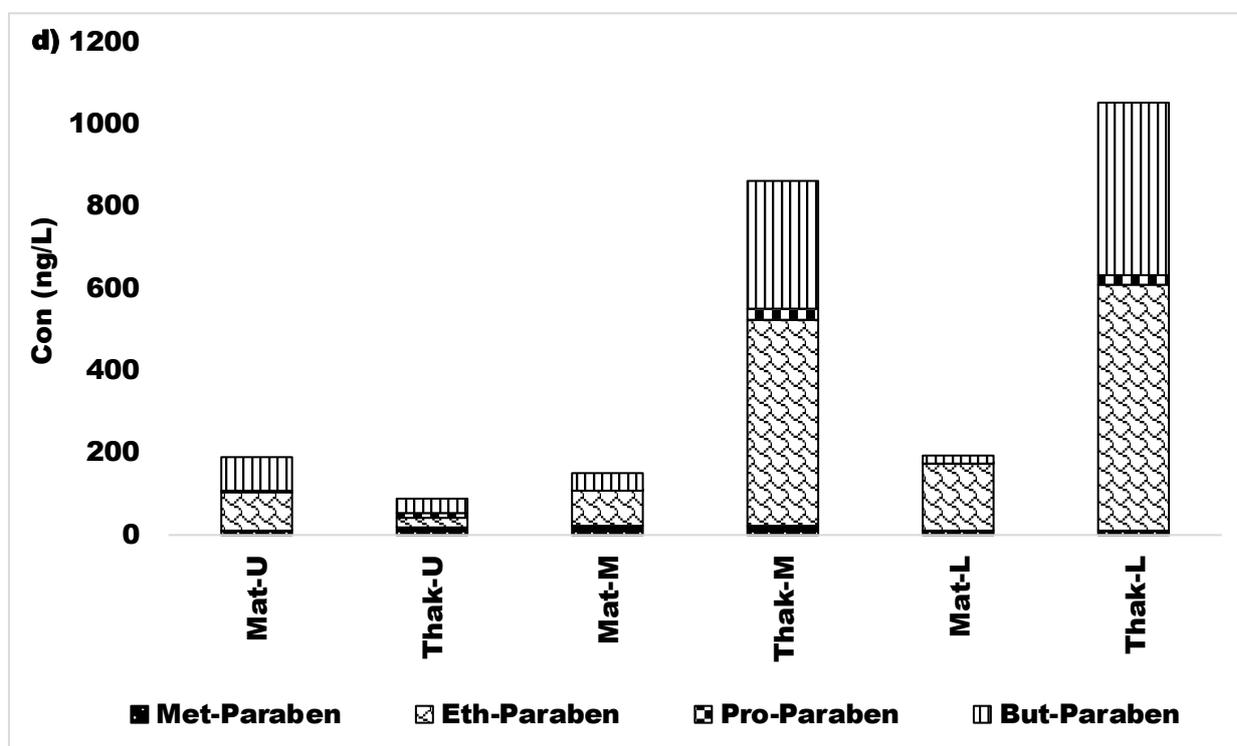
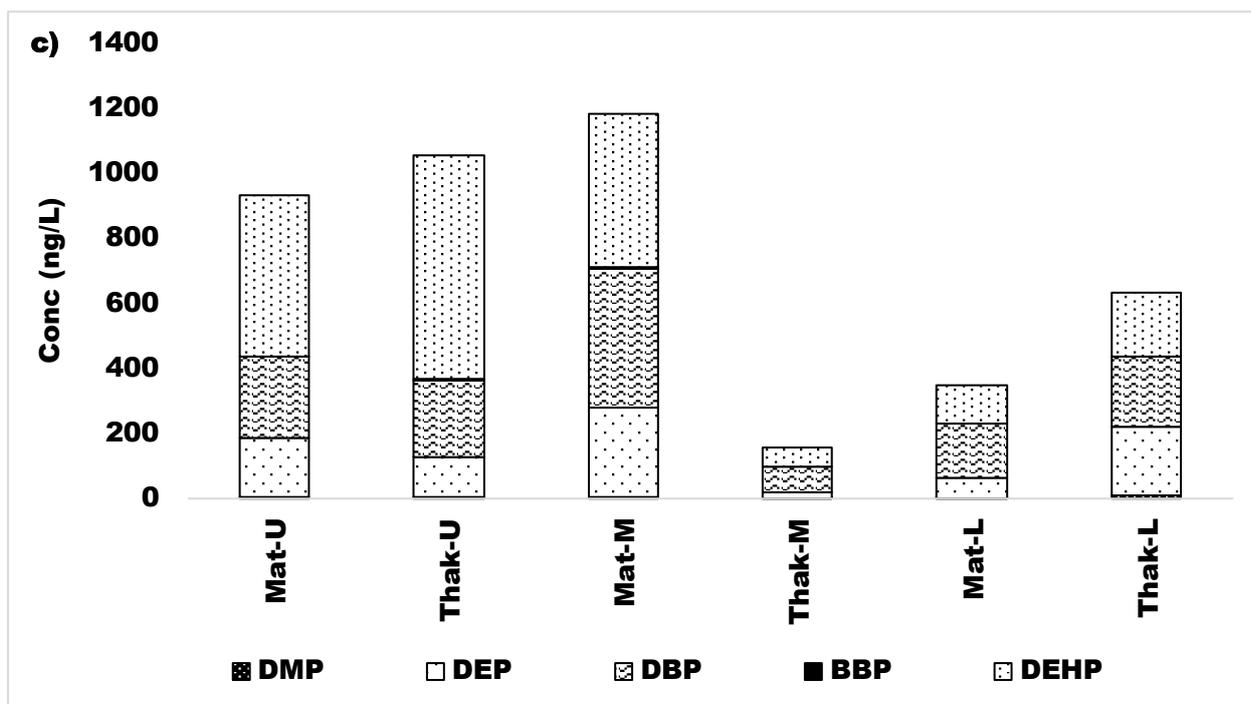


595 **Fig 3.** Concentration levels of four different derivatives of PPCPs: a) Phenols; b) NASIDs; c)
 596 Phthalates and d) Parabens in Indian Sundarbans

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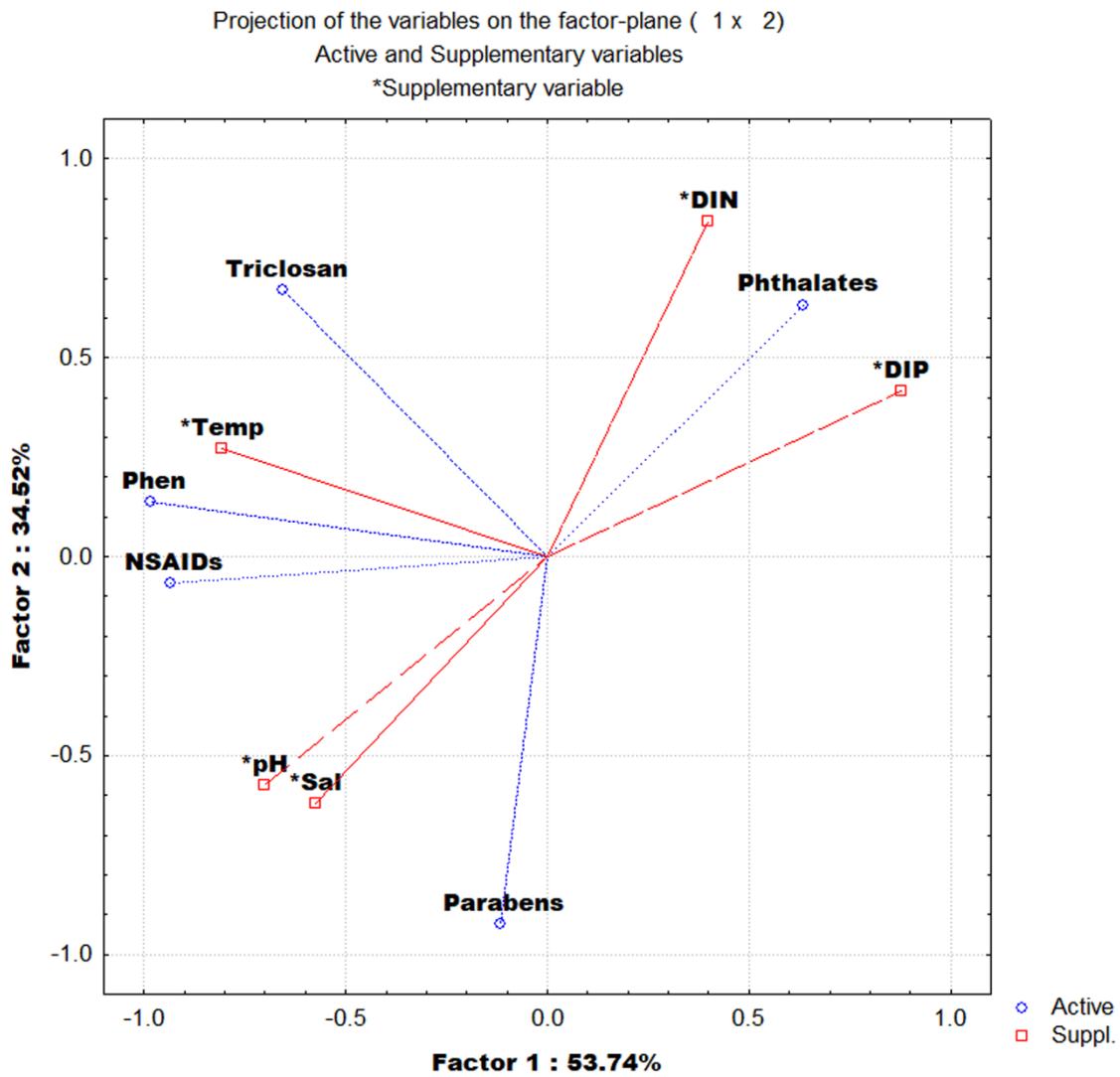


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601 **Fig 4.** Principal component analysis based on environmental parameters as supplementary
602 variables and PPCPs as active variables

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632 **Table 1**-Average values with standard deviation of PPCP concentration transect wise in
 633 Matla and Thakuran estuaries

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PPCPs	Estuary transects	Matla	Thakuran
Phenolics (ng/L)	Upper	1206.86 ± 1514.69	169.81 ± 168.63
	Middle	814.11 ± 985.48	1469.59 ± 1756.90
	Lower	4586.26 ± 6607.29	1379.15 ± 1437.99
Triclosan (ng/L)	Upper	1613.33 ± 354.92	356.52 ± 0
	Middle	780.87 ± 7.99	95.65 ± 0
	Lower	1885.22 ± 0	395.65 ± 0
NSAIDs (ng/L)	Upper	475.15 ± 243.66	349.36 ± 281.93
	Middle	536.58 ± 459.87	528.02 ± 376.24
	Lower	790.48 ± 902.80	617.33 ± 364.74
Phthalates (ng/L)	Upper	233.59 ± 194.58	211.63 ± 285.23
	Middle	237.10 ± 231.69	39.10 ± 34.28
	Lower	87.80 ± 70.45	158.4 ± 99.94
Parabens (ng/L)	Upper	47.70 ± 61.63	21.92 ± 9.28
	Middle	37.98 ± 48.55	215.44 ± 232.91
	Lower	48.14 ± 77.26	263.31 ± 294.15

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652 **Table 2**-Average values with standard deviation of environmental parameters and PPCP
 653 concentration of Matla and Thakuran estuaries of post monsoon month.

Parameters	Matla	Thakuran
Water TC	23.13±1.04	22.77±0.29
Sal	22.38±3.43	25.01±2.67
pH	8.13±0.21	8.21±0.13
DO (mg/L)	7.57±0.60	7.72±0.65
SPM (mg/L)	52.23±21.84	54.44±4.02
DIN (µM)	10.03±5.27	7.44±5.14
DIP (µM)	0.89±0.29	0.80±0.28
Phenolics (ng/L)	1704.85±1677.80	1006.18±725.73
Triclosan (ng/L)	1237.83±747.46	282.61±163.09
NSAIDs (ng/L)	548.18±198.89	498.24±136.44
Phthalates (ng/L)	217.64±79.05	136.38±88.34
Parabens (ng/L)	44.54±24.63	166.89±127.81

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663 **Table 3**-Comparision of different derivatives of PPCP with other estuaries and rivers of the
 664 world

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Study location	Emerging contaminants in water samples					Reference
Phenolics	OP	NP	BPA	BPS	BPF	
Pearl river estuary, South China	3.3-259	92-3950	24.9-811	NP	NP	Zhao et al (2009)
Dianchi lake, China	2.7-21.3	12.5-45.2	50.6-530.3	NP	NP	Wang et al (2012)
Kaveri, Vellar and Tamiraparani rivers, South India	ND-16.3	ND-2200	6.6-136	NP	NP	Selvaraj et al (2014)
Pearl river estuary, China	1.2-3.9	233-3352.8	12.4-62.7	NP	NP	Diao et al (2017)
Matla and Thakuran, Indian Sundarbans	282.2-17590.1	16.6-330.9	61.2-585.1	40.7-187.3	448.4-2288.5	Our study
Triclosan						
139 streams in US	140-2300					Kolpin et al (2002)
Hong Kong, China	15-110					Wu et al (2007)
Mississippi River, U.S.	8.8-34.9					Zhang et al 2007
Pearl river estuary, South China	8.1-247					Zhao et al (2009)
Guadalete river estuary, South-West Spain	27-310					Pintado-herrera et al (2014)
Tamiraparani, Cheranmahadevi and Tirunelveli, Tamil Nadu, South India	3800-5160					Ramaswamy et al (2011)
Matla and Thakuran, Indian Sundarbans	95.6-1885.2					Our study
NSAIDs	Naproxen	Diclofenac	Ibuprofen	Ketoprofen		
Pearl river estuaries, China	ND-328	8.3-114	ND-113	NP		Peng et al (2008) and Zhao et al (2010)
Sindian river, Taiwan	35.2-270	ND-56.5	ND-4350	ND-45		Lin et al (2010)
Rivers, Canada	ND-4500	ND-89	ND-6400	ND-79		Brun et al (2006)
Seine river estuary, France	<2.6-275	71-172	<2-610	<2.4-33.2		Togola and Budzinski (2007)
River, Poland	ND-753	ND-429	ND	ND-258		Baranowska and Kowalski (2011)

Ebro river basin, Spain	ND-109	ND-148	ND-541	ND-1060		Silva et al (2011)
Fyris river, Sweden	447	286	818	364		Daneshvar et al. (2012)
Tiber river, Italy	200-264	Nd-120	95-210	Nd-150		Patrolecco et al. (2013)
Vellar, Kaveri and Tamiraparani river, India	ND-28	ND-103	ND-200	ND-100		Shanmugam et al (2003)
Matla and Thakuran, Indian Sundarbans	73.9-114.4	71.2-543.4	476.2-776.8	614.6-2333.7		Our study
Phthalates	DMP	DEP	DBP	BBP	DEHP	
Dommel river, Netherlands	10-190	70-2300	70-3100	10-1800	900-5000	Vethaak et al (2005)
Eleven point, White, USA	NP	NP	140-4140	40-350	NP	Soli's et al. (2007)
Muga, Fluvia, Ter, Besos, Llobregat, Ebro, Spain	NP	50-280	NP	<LOQ-20	120-4980	Sanchez-Avila et al. (2012)
Yangtze, China	<10-25	<10-211	105-286	10-21	<10-836	He et al (2011)
Seine, France	26-184	71-181	67-319	ND	160-314	Dargnat et al. (2009)
Kaveri river, India	ND-94	36-520	ND-372	5.4-145	ND-822	Selvaraj et al 2014
Matla and Thakuran, Indian Sundarbans	1.2-9.3	20-275.1	74.7-425.9	ND-6.1	60.5-494.8	Our study
Parabens	MeP	EtP	PrP	BuP		
South Wales, United Kingdom	0.3-400	0.5-15	0.2-24	0.3-52		Kasprzyk-Hordern et al (2008)
Pearl River Delta, South China	0-1062	NP	0-3142	ND		Peng et al (2008)
Sea near outfalls, Ria de Aveiro, Portugal	5.1-62	0.3-15	0.5-6.1	0.2-7.1		Jonkers et al (2010)
Beijing area, China	0.8-920	ND-294	ND-565	ND-41.5		Li et al (2016)
Jiulong River estuary, China	1.6-68.8	NP	0.4-69.9	ND		Sun et al (2016)
Kenting National Park, Hengchun Peninsula, Taiwan.	6.7-104	ND-2.8	0.7-56.2	ND-1.8		Kung et al (2018)

Pearl river Estuary, China	0.2-4.8	0.5-1.2	0.6-1.9	0.1-0.5	Zhao et al (2019)
Matla and Thakuran, Indian Sundarbans	9.5-23.8	23.7-501.1	0.1-27	19.6-421.7	Our study

666 **NP** not performed, **ND** not detected, **BDL** below detection limit, **LOQ** limit of quantification

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669 **Fig. 1** Map of sampling location in Indian Sundarbans mangrove ecosystem.

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