

1 This paper is a non-peer reviewed preprint submitted to *EarthArXiv*. The manuscript was
2 submitted to *Nature Geoscience* for peer review. Future updates on this manuscript will be
3 provided once it's peer-reviewed or accepted. Please feel free to contact me at:
4 shiche@oregonstate.edu if you have any questions or feedbacks.

5

Soil Organic Matter Composition Improves Predictions of Potential Soil Respiration across the Continental United States

Cheng Shi (shiche@oregonstate.edu)^a, Maruti Mudunuru (maruti@pnnl.gov)^b, Maggie Bowman (maggie.bowman6@gmail.com)^c, Qian Zhao (qian.zhao@pnnl.gov)^c, Jason Toyoda (Jason.Toyoda@pnnl.gov)^c, William Kew (william.kew@pnnl.gov)^c, Yuri Corilo (corilo@pnnl.gov)^c, Odeta Qafoku (Odeta.Qafoku@pnnl.gov)^c, John R. Bargar (john.bargar@pnnl.gov)^c, Satish Karra (karra@pnnl.gov)^c, & Emily Graham (emily.graham@pnnl.gov)^{d,e*}

^aOregon State University, Department of Biological & Ecological Engineering, Corvallis, OR, United States.

^bEnergy and Environment Directorate, Pacific Northwest National Laboratory, Richland, WA, United States.

^cEnvironmental Molecular Science Laboratory, Pacific Northwest National Laboratory, Richland, WA, United States.

^dEarth and Biological Sciences Directorate, Pacific Northwest National Laboratory, Richland, WA, United States.

^eSchool of Biological Sciences, Washington State University, Pullman, WA, United States.

co

*Corresponding author: emily.graham@pnnl.gov

Twitter handle: @mpchengsworld, @emilybonnell_g, @EMSLscience

Abstract

26 **Abstract**
27
28 Despite the importance of microbial respiration of soil organic matter (SOM) in regulating
29 carbon flux between soils and the atmosphere, soil carbon (C) cycling models remain primarily
30 based on climate and soil properties, leading to large uncertainty in their predictions. Molecular
31 data have long been proposed as a promising avenue for resolving modeling errors, but evidence
32 for improved predictions of soil C cycles with high-resolution measurements remains mixed.
33 With data from the 1000 Soils Pilot of the Molecular Observation Network (MONet), we
34 analyzed the molecular composition of water-extractable SOM from 66 soil cores across the
35 United States to address this knowledge gap. Our innovation lies in using machine learning (ML)
36 to distill the thousands of SOM formula that we detected per sample into tractable units. Then,
37 we compared ML predictions of measured potential soil respiration using (1) a suite of standard
38 soil physicochemical data, (2) ultrahigh-resolution SOM composition independently, and (3) in
39 combination with physicochemistry to assess the added value of molecular information to predict
40 soil respiration. In surface soils (0-10 cm), water-extractable SOM chemistry alone provided
41 better estimates of potential soil respiration than soil physicochemical factors alone, and using
42 the combined sets of predictors yielded the highest explanatory power of soil respiration rates. In
43 contrast, in subsoils (>10 cm), SOM composition did not improve ML-based respiration model
44 performance, possibly due to the greater importance of mineral-associated SOM below the
45 surface layer. Our results underscore a role of water-extractable SOM as a determinant of soil
46 respiration and a need to integrate SOM composition into carbon cycle modeling for enhanced
47 predictions of terrestrial-atmosphere climate feedback.

48 **Introduction**

49 Soil respiration is estimated to release 60-100 Gt of C to the atmosphere per year,^{1, 2} six to ten
50 times as much C as released by fossil fuel combustion (~10 Gt C³). Microbial respiration of soil
51 organic matter (SOM) is one of the most important contributors to soil carbon dioxide (CO₂)
52 emissions and a critical link in the global C cycle.⁴ With increasing temperatures under climate
53 change, soil C repositories are vulnerable to increased rates of microbial respiration,⁵⁻⁷ which can
54 lead to positive feedbacks in global CO₂ emissions and temperature rises.⁸ Despite decades of
55 research, soil C fluxes remain one of the largest uncertainties in global climate predictions.⁹⁻¹⁴
56 Novel molecular measurements have recently been applied to identify SOM composition in an
57 effort to understand molecular-scale processes that could improve model predictions of CO₂
58 fluxes.¹⁵⁻¹⁸ Despite these efforts, our attempts to improve soil C model predictions by refining
59 chemical pools have yielded mixed results.¹⁹⁻²¹

60

61 The interplay of factors such as soil moisture, pH, nutrients, mineralogy, and SOM concentration
62 and chemistry governs microbially-derived transformations of SOM;²²⁻²⁷ but these relationships
63 are difficult to constrain.^{4, 28} The most commonly used modeling approaches are based on
64 Raich's model, which estimates respiration primarily as a function of temperature and water
65 availability.^{29, 30} Newer process-based model formulations use an additional suite of physical and
66 biogeochemical measurements to represent microbial and mineral processes. They incorporate
67 SOM chemistry either through several discrete pools or through their thermodynamic
68 properties.^{21, 31-34} With large spatiotemporal heterogeneity and limited availability of
69 comprehensive and standardized measurements at regional-to-continental scales, accurate
70 predictions of microbial SOM decomposition across different ecosystems remain challenging.³⁵

71

72 A better understanding of SOM concentration, composition, and bioavailability may enhance our
73 ability to predict soil C cycling processes through their controls on soil respiration and related
74 enzymatic activities.^{21, 31-34} Variations in the bioavailability of chemical classes of SOM are
75 mediated by geochemical conditions and biophysical constraints, such as microbial biomass and
76 necromass, reactive metals and minerals, organic and mineral horizon thickness, and other
77 climate-related variables.³⁶ For example, coarse-textured soil is more conducive to
78 decomposition of chemically labile litter-derived C potentially due to higher fungal activity in

79 organic-rich horizons.^{37, 38} In addition, the interface between fresh litter inputs and soil minerals
80 can serve as a hotspot for microbial breakdown of C found in the litter, resulting in the formation
81 of soil aggregates and organo-mineral associations.³⁹ This variability underlines the essential
82 need to identify unique subsets of SOM formula that contribute more to soil respiration among
83 different ecosystems and soil depths.

84

85 The distillation of multidimensional SOM composition profiles into a tractable set of formula
86 that influence soil respiration is a key challenge in soil ecology.^{15, 28, 40-45} Unsupervised machine
87 learning models that summarize large data into a small number of significant features have been
88 widely used to study microbial communities, SOM composition, and other environmental
89 problems with multidimensional data.⁴⁶ Dimensionality reduction such as principal component
90 analysis (PCA)⁴⁷⁻⁴⁹ and clustering methods such as hierarchical clustering analysis⁵⁰⁻⁵² are the
91 most common tools to explore large molecular datasets. Although these tools are beginning to
92 be applied to determine the relationship between SOM composition and soil physicochemistry,⁴⁹
93 it is still challenging to extract a subset of SOM features associated with specific processes, like
94 soil respiration.

95

96 Although ultrahigh mass resolution measurements can provide unprecedented characterization of
97 the thousands of individual formulae that comprise SOM, the interpretation of these data types
98 largely remains guided by coarse chemical and ecological groupings. Here, we develop models
99 using semi-supervised machine learning (non-negative matrix factorization with custom k -means
100 clustering, NMF k) to reduce the complexity of molecular information into k distinct signatures of
101 water-extractable SOM chemistry at two depths in cores collected across the continental United
102 States. We then explore the extent to which these signatures and NMF k -enabled feature set can
103 provide additional insight into rates of soil respiration beyond variables that are more routinely
104 collected. By examining a multitude of physicochemical and SOM factors, our goal is to
105 elucidate the specific aspects of SOM chemistry that may be vital to understanding and
106 predicting below-ground C storage.

107

108 **Methods**

109

110 *Soil sampling and characterization.*

111

112 As part of the 1000 Soils Pilot study for the Molecular Observation Network (MONet) program,
113 we collected 66 soils from across the continental US using standardized sampling procedures
114 described by Bowman et al.⁵³ (Figure S1). Two long cores (30 cm) and three short cores (10 cm)
115 were collected at each site. We also conducted field measurements, including soil temperature,
116 volumetric water content, vegetation type, and weather conditions. Cores were shipped on ice
117 overnight to the Pacific Northwest Laboratory for further analysis. A full description of sampling
118 and analytical methodologies is available in Bowman et al.⁵³

119

120 Briefly, once soil cores were delivered to the lab, we divided the 30-cm cores into 10 cm depth
121 intervals, where only the top (hereafter, surface or surficial soil) and bottom (hereafter, subsoil)
122 sections were used for further analysis. We mixed the top sections with three short cores to
123 homogenize the local variation. The soils were then sieved through 4 mm sieves separately to
124 remove rocks and root structures. We measured gravimetric water content (GWC) by drying 10 g
125 of soil for 24 hours in a drying oven at 100 °C. We measured soil pH by mixing 20 g of dry soil
126 with 20 mL of DI water (1000 rpm on reciprocating shaker for 15 minutes), and tested with a
127 calibrated pH probe. Soil microbial biomass C and nitrogen (N) content were measured via
128 chloroform fumigation.⁵⁴⁻⁵⁶ We extracted phosphorus contents using Bray (pH < 7) or Olsen
129 extractions (pH > 7),^{57, 58} and extracted nitrate and ammonium with 0.5M K₂SO₄ and tested by
130 colorimetric methods. Ion concentrations of potassium (K), calcium (Ca), magnesium (Mg), and
131 sodium (Na) from 1:10 ammonium acetate extraction, Zinc (Zn), manganese (Mn), copper (Cu),
132 iron (Fe), boron (B), and sulfate (SO₄²⁻) from 1:2 soil to diethylenetriaminepentaacetic acid
133 (DPTA) extraction were measured using Inductively coupled plasma mass spectrometry (ICP-
134 MS). We measured total C and N using the AOAC official methods 972.43.⁵⁹ Soil texture was
135 measured by hydrometer analysis. Finally, we assessed potential soil respiration using the CO₂
136 burst method with 24 hours of incubation at 24 °C.⁵³

137

138 *Water extractable SOM characterization.*

139

140 We extracted water-soluble SOM by mixing 6 g of dry soil with 30 ml DI water in triplicates,
141 shaken for 2 hours at 800 rpm, and centrifuged at 6,000 rpm for 8 minutes. 5 ml of supernatant
142 was acidified with 2 μ l concentrated phosphoric acid (37%), and then loaded onto Agilent Bond
143 Elut PPL solid phase extraction (SPE) cartridges⁶⁰ with Gilson ASPEC® SPE system. A Bruker
144 7 T Fourier transform ion cyclotron resonance mass spectrometry (FTICR MS) at the
145 Environmental Molecular Sciences Laboratory (EMSL) in Richland, WA, was used to analyze
146 SOM composition, with a negative ionization mode and ion accumulation time at 0.01 or 0.025
147 seconds (depending on dissolved organic C concentration). The measured mass accuracy was
148 typically within 1 ppm. One lab blank and one Suwannee River Fulvic Acid (SRFA) sample (20
149 ppm) were tested every 30 soils to evaluate instrument performance.

150
151 Raw FTICR MS data was processed with CoreMS (Python package, installed on 2022/11/22),⁶¹
152 including signal processing, peak detection, and molecular formula assignment. Noise
153 thresholding was performed with signal-to-noise threshold (5 std.), mass error (0.3 ppm), and
154 stoichiometric limits from domain knowledge (supporting information). Suwannee River fulvic
155 acid (SRFA) standards were used to set a calibration threshold for all soils in the same batch.
156 Molecular formula was assigned based on both accurate mass and filtered by their confidence
157 score from CoreMS. After calibration and formulae assignment, we filtered the assigned peaks
158 by m/z between 200 to 1,000, present in at least 2 out of 3 replicates, not present in two or more
159 lab blanks, and with formulae confidence scores (combines m/z error and isotopic pattern)⁶¹
160 above 0.7. We predicted compound classes of the filtered formulae based on O/C and H/C ratios
161 of van Krevelen classes.^{62, 63} The suffix “-like” in chemical classes indicates the uncertainty of
162 the van Krevelen classification method.⁶³ We converted the peak intensity values to
163 present/absent (1/0) and separated the final dataset by soil depth (surface vs. subsoil) for
164 statistical analysis. Alpha diversity was calculated as the total number of SOM formulae
165 identified in each sample.

166

167 *Data analysis and machine learning methods.*

168

169 We used linear regression models to evaluate the relationship between soil potential respiration
170 and soil physicochemical variables. To avoid the impacts of different magnitudes of the data that

171 might lead to biased relationships, we performed \log_{10} transformation on potential respiration
172 rates, total C, total N, total sulfur, and Mn concentration. *stats.linregress* function from *scipy*
173 package (v 1.11.4) in Python (v 3.7.1) was applied to calculate the fitted line, r^2 value (*rvalue*²,
174 Pearson correlation), and p-value (*pvalue*). Pairwise plots with regression fitting were generated
175 by the *pairplot* function from the *seaborn* package (v 0.12.1) in Python.

176
177 We used non-negative matrix factorization (NMF)⁶⁴ with custom k-means clustering (NMFk)⁶⁵
178 to identify signature components from the 7,312 and 5,515 SOM molecular formula (for surface
179 and subsoil, respectively) we detected (i.e., N formulae in m soils) with *pyNMFk* package
180 (Python, <https://github.com/lanl/pyDNMFk>, Figure 1). More details on NMFk assumptions,
181 model settings, and model robustness are in the supporting information. Briefly, NMFk tends to
182 be more successful at extracting explainable basis or signatures from large multivariate datasets,
183 compared to other dimensionality reduction tools such as principal component analysis.^{64, 66} As
184 applied here, NMFk summarizes data into discrete signatures that contain weights for each SOM
185 formulae detected by FTICR-MS for each soil layer independently (i.e., a separate set of
186 signatures was generated to summarize surface versus subsoils, allowing us to explore depth-
187 specific relationships with potential soil respiration). The optimal number of signatures was
188 determined from silhouette coefficients of different NMFk models. A W-matrix with the weights
189 of each SOM formulae (N) to each extracted signature (k), and an H-matrix with the contribution
190 of each signature (k) to each soil sample (m) were generated from NMFk. To visualize the
191 composition of each NMFk signatures (W-matrix), we generated a heatmap of SOM formula
192 with normalized weights (0-1) >0.5 in at least one NMFk, clustered by van Krevelen class
193 assignment (*clustermap* function from *seaborn* package). Within each inferred chemical class of
194 SOM formula, we further clustered formula using the “linkage” method from the *scipy* package
195 (“ward” method with “Euclidean” distance) to illustrate the difference between NMFk
196 signatures.

197
198 To define groups of soils with high, medium, or low rates of potential respiration, we used k -
199 means clustering on potential soil respiration with the elbow method to select the number of
200 groups (*KMeans* from *scikit-learn* package).⁶⁷

201

202 Then, we mapped the extracted k signatures to soil respiration using supervised machine
203 learning. To evaluate the potential value of NMF k -extracted SOM signatures for explaining soil
204 respiration, we conducted three sets of machine learning models: (1) selected environmental
205 parameters alone (i.e., variables with $R^2 > 0.2$ in individual regressions, Figure 2, Table S1), (2)
206 SOM composition alone (NMF k weights from H-matrix), and (3) environmental and SOM
207 composition in combination. All machine learning models were built using gradient boosting
208 regression (GBR) from *scikit-learn* package (v. 0.24, Python). Model hyperparameters were
209 tuned first with 5-fold cross validation on 80% of each dataset (*train_test_split* in scikit-learn,
210 with the same *random_state* for models in the same layer) using *RandomizedSearchCV* function
211 from scikit-learn. We then used the best-tuned parameters with 80% of soils to build the finalized
212 model. Root means square error (RMSE) was used to evaluate the error of models. More details
213 on hyperparameter grids can be found in supporting information. All the models were then tested
214 with the other 20% of soils to compare their performance. The most important predictors for the
215 models with the best performance were then determined using MDI importance and/or mean
216 decrease in impurity to infer potential relationships between soil environmental parameters,
217 SOM composition, and potential soil respiration. Partial dependence plots were used to evaluate
218 the response of potential respiration to the selected important features.

219

220 **Results**

221 *Soil physicochemistry and potential respiration*

222

223 Overall, many soil parameters, including potential soil respiration, tended to be higher in surface
224 soils than in subsoils. Significant differences ($p < 0.05$) between surface soils and subsoils in total
225 C, total N, total sulfur, C/N ratio, and other factors are shown Figure S3. In particular, surface
226 soils had higher potential respiration rates (median: 72.6 $\mu\text{g CO}_2/\text{g soil/day}$) than subsoils
227 (median: 21.9 $\mu\text{g CO}_2/\text{g soil/day}$) (Mann–Whitney $U = 3022.5$, $N_{\text{surface}} = 63$, $N_{\text{subsoil}} = 61$, $P <$
228 0.05).

229

230 We grouped potential soil respiration into 3 levels corresponding to low, medium, and high
231 respiration in each soil layer using k -means clustering (Figure S2). For both surface and subsoils,
232 soil with high potential respiration tended to be sourced from the Midwestern and Northeastern

233 United States. (Figure 3, Figure S5). In surface soil, high potential respiration was associated
234 with five soils collected in Utah, Wyoming, and Virginia (within temperate conifer forest and
235 temperate broadleaf & mixed forest biomes, Figure 3, Figure S1). In subsoils, high respiration
236 was associated with three soils from Utah and Maryland (temperate conifer forests and broadleaf
237 & mixed forests biomes). Desert soils had the lowest respiration in both layers (Figure S1).

238
239 We found relationships between soil respiration and many variables that supported prevailing
240 paradigms. A full correlation table of associations between different soil properties is available in
241 the SI (Table S1). Briefly, potential respiration rates in both surface and subsoils were positively
242 correlated with gravimetric water content (GWC) (r^2 : 0.246 and 0.225, $p < 0.05$) and cation
243 exchange capacity (CEC, r^2 : 0.405 and 0.354, $p < 0.05$, Figure 2). They were also positively
244 correlated with total C and total N content, with stronger relationships in surface soils (r^2 : 0.487
245 v.s. 0.268 for total C, r^2 : 0.439 v.s. 0.248 for total N, $p < 0.05$). Total bases and magnesium (Mg)
246 concentrations had a higher correlation to respiration in subsoils than surface soils (r^2 : 0.227 v.s.
247 0.146 and 0.287 v.s. 0.160, $p < 0.05$, Figure 2), while manganese (Mn) concentrations were
248 correlated to respiration in surface soils (r^2 : 0.324, $p < 0.05$, Figure 2).

249
250 *SOM composition and NMFk partitioning of SOM.*

251
252 Across all soils, the most common chemical classes of SOM were lignin-, condensed
253 hydrocarbon-, and tannin-like formula. Most formula in these classes were present in both
254 surface and subsoils (i.e., 'shared' formula). However, surface soils contained more unique
255 formula than subsoils for all compound classes (Figure 3b). In particular, many protein-, amino
256 sugar-, and lipid-like compounds were identified in surface soils only, with very few compounds
257 in these classes being unique to subsoils. Because SOM consists of thousands of different
258 compounds, we also used alpha diversity to represent the SOM richness per sample (Figure 3).
259 Soils from the Midwestern U.S. and the West Coast had relatively higher alpha diversity than
260 soils from other regions.

261
262 We used NMFk to summarize SOM composition into 7 and 5 NMFk signatures, respectively, for
263 surface and subsoils (Figure 4). Geographic patterns in SOM signatures are displayed in Figure

264 S6-7, with more geographic clustering of NMFs in surface soils than in subsoils. For surface
265 soils, NMF3 presented as the largest relative contributor to SOM composition in 20 soils across
266 all biomes (i.e., highest weighting in H-matrix, hereafter, 'dominant signature', Figure S6).
267 NMF2, NMF5, and NMF7 served as the dominant signature in at least 9 soils each. For subsoils,
268 NMF5 was the dominant signature in 27 soils distributed across all biomes in the continental
269 United States. NMF2 appeared to be the second dominant SOM signature in subsoils with the
270 highest weights in 16 soils. There was no single NMF signature that could exclusively represent
271 SOM composition of all sites in the same region for either surface or subsoils, suggesting that
272 SOM composition at local sites is best summarized by a combination of multiple NMFs.

273
274 The most important formula contributing to the composition of each NMF (i.e., formula with
275 normalized weights >0.5 in W-matrix) are shown in Figure 4a-b. For surface soils, NMF1, 4, 6,
276 and 7 had a relatively higher number of important compounds identified as lignin-like. NMF6
277 and 7 had larger contributions of condensed hydrocarbon-like formula. NMF1 had higher
278 contribution from protein-like and amino sugar-like compounds, while NMF3 and 5 had the
279 lowest contribution from protein-like, amino sugar-like, and lipid-like compounds, suggesting
280 their low microbial activities. NMF4 had the largest number of lipid-like compounds as
281 important features. In subsoil samples, important formula for all NMFs tended to be classified as
282 lignin-, tannin-, and/or condensed hydrocarbon-like. NMF1 and NMF5 had most important
283 features identified as lignin-like and some tannin-like compounds. NMF2 had the largest fraction
284 of condensed hydrocarbon-like compounds. NMF4 had larger contributions of protein-like and
285 amino sugar-like formula (Figure S8).

286
287 We also compared if formula contributing to NMF signatures tended to be similar among surface
288 and subsoils by assessing shared vs unique formula. NMF-selected formula (weights >0.5 in W-
289 matrix) followed the same general patterns as the overall SOM pool but showed amplified
290 relationships (Figure 4c). Most shared formula belonged to lignin-, tannin-, and/or condensed
291 hydrocarbon-like chemical classes. Very few NMF-selected formula were unique to subsoils,
292 with lipid-, amino sugar-, and especially protein-like important formula unique to surface soils.
293

294 We also observed differences in dominant NMF signatures across high-, medium-, and low-
295 respiration soils, particularly in surface soils (Figure 4d-e). High respiration surface soils were
296 characterized by five NMF signatures (1, 2, 3, 6, and 7), with the largest contribution from
297 NMF6. Low respiration surface soils, in contrast, uniquely contained NMF5, and they did not
298 have any contribution from NMF6. In subsoils, high respiration soils consisted of NMF 1, 2 and
299 4, while low respiration soils consisted of NMF1, 2, and 5. NMF5 had a larger contribution in
300 low-respiration soils from both temperate forests and grasslands/shrublands. NMF5 had higher
301 weights in low-respiration soils, and NMF3 and NMF4 had lower weights in low-respiration
302 soils.

303

304 *Relative importance of physicochemistry and SOM composition in potential soil respiration*
305 *models*

306

307 We developed gradient-boosting regression models to predict potential soil respiration with (1)
308 physicochemical variables, (2) SOM composition represented by NMF signatures, and (3) both
309 of them combined. Model performances are summarized in Table 1 and Figure 5.

310

311 Selected physicochemical variables (consisting of total C, total N, CEC, moisture, Mn (surface),
312 total base (subsoil), and Mg (subsoil) concentration) had significant independent Pearson's
313 correlation to respiration w/ $p < 0.05$ and $r^2 > 0.2$ (Table S1). Physicochemical variables
314 predicted potential respiration rates in surface and subsoils equally well ($R^2 = 0.44$ and 0.43
315 respectively for testing data). In surface soils, total C, total N, and cation exchange capacity
316 (CEC) were identified as the top 3 most important predictors, followed by Mn concentration and
317 soil moisture (Figure 4). In subsoils, CEC, total N, and soil moisture were the most important
318 predictor, and total C was the least important predictor (Figure S9).

319

320 Using SOM composition (NMF signatures) as predictors, we had better model performance in
321 surface soils than in subsoils (testing $R^2 = 0.54$ vs. 0.08), and SOM composition alone predicted
322 more slightly variation in potential respiration rates than physicochemical variables alone in
323 surface soils (testing $R^2 = 0.54$ vs. 0.44), even when controlling for an equal number of

324 predictors (testing $R^2 = 0.48$ vs. 0.44). NMF3, NMF5, and NMF2 were the most important SOM
325 composition variables for explaining soil respiration in surface soils (Figure 5).

326
327 When we combined both physicochemical variables and SOM composition into a single
328 predictor set, we obtained better respiration model performance ($R^2 = 0.62$) compared to models
329 with environmental variables or SOM composition in surface soils only. However, the model
330 describing potential respiration rates in subsoil was worse ($R^2 = 0.36$) when compared to models
331 based on physicochemical variables only. In surface soils, the 3 most important variables were
332 the same as the physicochemical model (Figure 5). NMF6 was identified as the most important
333 SOM variable, followed by NMF3, NMF2, and NMF5 (Figure 5). In subsoils, total N and Mg
334 concentration were the most important variables, followed by NMF5, total C and CEC.

335

336 **Discussion**

337 *Soil respiration and physicochemistry*

338 Soil moisture, total C, and total N appeared to regulate soil respiration in both surface soil and
339 subsoil, as evidenced by positive correlations of total C, N, and moisture with potential soil
340 respiration (Figure 2). This is consistent with previous work describing relationships between
341 these properties and soil respiration, as well as other factors that we observed to be correlated
342 with respiration including pH and CEC.^{33, 68-70} Soil physical properties (e.g, moisture and pore
343 space connectivity) can constrain microbial access to SOM molecules and nutrients isolated in
344 soil pore networks, thereby regulating microbial respiration of SOM.^{22, 33, 71-73} Additionally, C
345 and N can limit soil respiration through stoichiometric constraints on biomass production.^{4, 74-76}

346

347 We propose that differences in potential respiration between surface and subsoil may be related
348 to variation in soil C composition and stabilization mechanisms across soil layers. We observed a
349 steeper correlation between total C and potential soil respiration in surface soils than in subsoils,
350 despite similar slopes for relationships of N and moisture with respiration at both depths. While
351 we anticipated that microbial respiration would decrease significantly with soil depth,⁷⁷ the
352 change in the nature of the relationship between C and respiration suggests that differences in
353 SOM composition or microbial access to C substrates could be associated with potential rates of

354 respiration. Surface soils are generally rich in relatively bioavailable water-extractable organic
355 matter and contain higher proportions of microbial biomass in contrast to subsoils that are more
356 mineral with lower pore space connectivity and larger pools of mineral-associated organic
357 matter.⁷⁸ Given previously observed differences in SOM composition and soil structure, we
358 hypothesize that factors including oxygen availability and alternative electron acceptors may
359 influence heterotrophic respiration to a greater degree than soil C as depth increases.

360
361 We also found a suite of correlations between elements and potential soil respiration that may
362 reflect the influence of vegetation across rooting profiles; however, associations between
363 inorganic nutrients (NH₄⁺, NO₃⁻, PO₄³⁻) and respiration were conspicuously absent ($p > 0.05$,
364 Table S1).⁷⁹⁻⁸² Mg, Mn, Zn, and sulfate were correlated to potential soil respiration and are
365 known to have strong impacts on plant productivity that provides chemically labile C sources for
366 microbial respiration.⁸³⁻⁸⁵ Mn can also influence soil respiration by regulating the activities of
367 Mn peroxidase enzyme, a lignin-degrading enzyme produced by fungi and *Actinobacteria*.⁸⁶⁻⁹⁰
368 Because total N corresponded to potential soil respiration, the lack of relationship between
369 respiration and inorganic nutrients may indicate organic nutrients as key drivers of soil
370 respiration. Alternatively, inorganic nutrient limitations that vary tremendously through space
371 and time may not be observable across different ecosystems at the continental scale.^{91, 92}

372
373 In addition to patterns in soil physicochemistry, we observed geographic patterns in potential
374 soil respiration that contrasted with some previous estimates,⁶ including high rates of potential
375 soil respiration in the midwestern and mid-Atlantic regions, and at high elevations (Figure 3). A
376 notable difference between Nissan et al. and the current study is that Nissan et al. report
377 simulated mean annual values of heterotrophic respiration in soils, while the current study
378 reports the measured potential respiration rates of sieved soils collected during the summer
379 months. Because high latitude and high elevation ecosystems can exhibit intense, short-lived
380 peaks of biomass during summertime,⁹³ soils collected during this period may have relatively
381 extreme rates of potential respiration that are averaged out at the annual scale. Another
382 interpretation for higher potential soil respiration at high elevation is that relative humidity
383 typically increases with elevation and thus can stimulate higher microbial activities and SOM
384 decomposition.⁹⁴ In contrast, comparatively low potential soil respiration recorded in the

385 Southeastern United States could also reflect the comparatively low C content of these soils that
386 has been associated with faster turnover rates and high year-round temperatures.⁹⁵

387 *Depth partitioning in relationships between SOM composition and potential soil respiration*

388 Differences in SOM composition with soil depth and across the continental United States were
389 associated with potential soil respiration, supporting previous studies showing relationships
390 between SOM composition and soil respiration rates (Figure 3).^{8, 26, 27} Regardless of depth or
391 geographic location, the diversity of water-extractable SOM compounds appeared to be a
392 common factor in regulating potential soil respiration — soils with higher potential respiration
393 generally had more diverse pools of water-extractable SOM (Figure 3d-e).

394
395 Our results were consistent with a paradigm in which chemically bioavailable, plant-derived
396 molecules including proteins and amino sugars are degraded through soil profiles and
397 transformed into microbially-derived byproducts that are stabilized via organo-mineral
398 associations;⁹⁶⁻⁹⁸ whereas more chemically recalcitrant compounds (e.g., lignins and tannin) are
399 preserved due to their lower thermodynamic bioavailability.⁹⁹⁻¹⁰¹ Coincident decreases in SOM
400 diversity from surface to subsoils were also associated with decreases in potential soil respiration
401 (Figure 3b-c), further supporting a link between SOM pool composition and microbial
402 decomposition.^{101, 102} The comparatively diverse SOM pools in surface soils contained more
403 bioavailable compounds than subsoils, including protein-, amino sugar-, and lipid-like
404 compounds.^{103, 104} The number of formulae in these chemical classes declined with depth, and
405 formula that were common to both soil layers primarily included chemical classes with low
406 putative bioavailability such as lignin-, tannin-, and condensed hydrocarbon-like compounds.¹⁰⁴

407
408 Given that not all chemical constituents of SOM contribute to soil respiration and that surface
409 and subsoils differ substantially in mineralogy and structure, we hypothesized that distinct
410 subsets of SOM would contribute to respiration in surface vs. subsoils. There was no single NMF
411 that dominated low- vs. high-potential respiration soils in either layer, however, NMF weightings
412 varied substantially across soils with different rates of potential respiration in both layers (Figure
413 4d-e). This suggests that different subsets of SOM were disproportionately associated with soils
414 exhibiting high vs. low potential respiration rates. While patterns in SOM chemical across

415 geographic regions were difficult to disentangle, the spatial distribution of NMF types suggested
416 local similarity in SOM composition in both layers (Figure S6-7), likely reflecting similar
417 underlying chemistry, mineralogy, and/or biogeochemical processes.⁹⁵

418
419 The SOM formula within NMFs that correspond to changes in soil respiration may represent a
420 key step forward in understanding the chemical bioavailability of water-extractable organic
421 matter in soils. In surface soils, NMF6 displayed a dramatic increase in weightings from low-to-
422 high respiration soils. It contained a diverse suite of compounds including protein-, (soluble)
423 lipid-, and amino sugar-like formula that can be rapidly used as microbial substrate. Proteins and
424 amino sugars can fuel microbial metabolism of SOM,^{105, 106} thus the prevalence of these
425 compounds within NMF6 may support high potential rates of soil respiration. NMF1 and NMF7
426 in surface soils contained a diverse mixture of compounds and also increased from low-to-high
427 respiration soils, supporting a possible relationship between SOM pool diversity and microbial
428 respiration (see previous section). In contrast, surface NMF2, NMF3 and NMF5 decreased in
429 importance from low-to-high respiration soils and primarily consisted of a small but unique
430 subset of lignin- and tannin-like compounds (Figure 4a). This is consistent with low
431 bioavailability of its chemical constituents suppressing microbial respiration.^{100, 104} It suggests
432 that despite the often-inferred high bioavailability of water-extractable SOM,^{41, 107} there may be a
433 significant fraction of water-extractable SOM that is chemically protected from microbial
434 decomposition.^{40, 41, 106} Interestingly, NMF4 in surface soils — which contained the greatest
435 number of lipid-like formula (Figure 4a) and had a comparatively large fraction of protein-like
436 formula — was not present in any high-respiration soils. We therefore suggest that NMF4 may be
437 an indicator of non-living microbial biomass (i.e., necromass) which is disproportionately
438 comprised of lipids (microbial cell wall remnants) and amino sugars and proteins (the basis of
439 intracellular materials)^{108, 109}. Alternatively, the large number of lipid-like compounds in NMF4
440 could represent plant-derived lipids that are thought to be resistant to decomposition.¹¹⁰

441
442 The comparatively weak relationship between subsoil water-extractable SOM and potential soil
443 respiration as compared to surface soils highlights recent work emphasizing the importance of
444 mineral-associated organic matter in soil C storage.¹¹¹⁻¹¹³ In subsoils, NMF4 (associated with
445 high-respiration soils) and NMF5 (associated with low-respiration soils) had the largest

446 disparities in weighting across subsoils (Figure 4e). Consistent with observations from surface
447 soils, subsoil NMF4 contained the largest proportion of amino sugar- and protein-like formula
448 compared to other subsoil NMFs, while NMF5 was almost entirely composed of lignin- and
449 tannin-like compounds.¹⁰⁴ The composition of water-extractable SOM in mineral subsoils is an
450 emerging area of research, and it remains unclear how different SOM chemistries contribute to
451 subsoil respiration.⁹⁹ Our results suggest some consistencies in the chemical mechanisms of
452 SOM bioavailability across soil horizons. However, one subsoil NMF (NMF2) had unexpectedly
453 large weightings in high respiration subsoils despite low bioavailability typically associated with
454 its chemical constituents.^{104, 114} The remaining subsoil NMFs (1 and 3) were present in both low-
455 and high-respiration subsoils. This denotes that factors beyond chemical recalcitrance or beyond
456 the most commonly measured (water-extractable) SOM pool are critical to understanding
457 belowground C cycling.^{109, 115}

458

459 *Relative importance of physicochemistry and SOM composition in predicting potential soil*
460 *respiration*

461

462 By developing machine learning models to predict respiration with soil physicochemistry and
463 SOM composition (NMFs) separately and in combination, we were able to distinguish the
464 contributions of each set of factors for predicting soil potential respiration. The models based on
465 physicochemistry alone explained a modest amount of variation in soil respiration (44% and
466 43% in surface and subsoils, respectively), in line with the range of explanatory power observed
467 in other works.^{116, 117} The most important predictors identified by the physicochemical models
468 (total C, total N, and CEC for surface soils, CEC for subsoils) were consistent with the variables
469 with the highest independent correlations to potential soil respiration (Table S1).

470

471 For surface soils, models based on SOM composition alone (54% variation explained) and both
472 physicochemical factors and SOM composition combined (62% variation explained) suggest
473 that SOM composition (1) can predict soil respiration at least as well as commonly measured
474 physicochemical variables and (2) explains some portion of soil respiration that is not captured
475 by physicochemistry. In models based on SOM composition alone, NMF3 (which was mainly
476 in low-respiration soil and was comprised of lignin- and tannin-like formula, see previous

477 sections) was the strongest predictor of soil respiration followed by NMF2 and NMF5. The
478 relative chemical recalcitrance of the most important predictors of respiration may suggest that
479 the proportion of thermodynamically unfavorable formula in water-extractable SOM has a direct
480 inhibitory effect on soil metabolism. Indeed, thermodynamic regulation of organic C
481 composition can be a key control for the rate of respiration in ecosystems.^{40, 41} Therefore, the
482 inclusion of SOM composition in more mechanistic modeling approaches may be able to
483 improve predictions of soil respiration rates.

484
485 However, models for subsoils displayed different dynamics. In the subsoil model based on
486 physicochemical variables alone, total C was the least important predictor (vs. the most
487 important predictor for surface soils), and the model containing SOM composition did not yield
488 high predictive power. We also observed a similar pattern in the partial dependence of soil
489 potential respiration to soil total C across the layers (Figure S10). The marginal effect of total C
490 to surface soil respiration was stronger than the effect on subsoil respiration, supporting a
491 stronger association between total C and potential respiration in surface soil vs. subsoil. The low
492 predictive power of total C relative to other physicochemical factors could explain why SOM
493 composition did not add predictive power to potential respiration in subsoils. Since more total
494 and organic C is stored in surface soils, resolution into the water-extractable SOM pool (reflected
495 here by NMFs) might be a more significant factor for predicting surface soil respiration than in
496 subsoils that are characterized by lower total C and more mineral-associated SOM.⁹⁹

497
498 Our results suggest that NMF-extracted signatures of SOM composition are able to improve
499 surface soil model performance by integrating fundamental molecular information into soil
500 respiration models across very different soil ecosystems at the continental scale. NMF6, which
501 was the most important NMF signature in combined models of surface respiration, consisted of
502 diverse chemically-bioavailable compounds, and it mainly existed in high-respiration soils (see
503 previous sections).¹⁰⁴ We therefore suggest that chemically-bioavailable compounds in water-
504 extractable SOM pools may provide the greatest complementary explanatory power to
505 physicochemical factors in respiration predictions. Because SOM pools vary tremendously at the
506 continental-scale, refined regional or local studies that encompass lower-variability parameter
507 spaces may yield even more value of SOM molecular data to soil C modeling.

508
509 We note that physicochemical predictors were stronger predictors of soil respiration than SOM
510 composition in the combined surface soil models. However, the inclusion of SOM composition
511 improved physicochemistry-only models by 18%, indicating that it may significantly impact our
512 ability to predict the rate of soil C cycling processes. Future modeling with carefully applied
513 machine learning approaches may open up new avenues for further extracting the relevant
514 portions of SOM pools for inclusion in climate models.

515

516 **Conclusion**

517 Leveraging molecular information of SOM chemistry to improve conceptualizations and models
518 of soil C cycling is a pressing challenge for global biogeochemical and climate predictions. In
519 this study, we use machine learning (NMF k) to distill the thousands of SOM molecules detected
520 by ultrahigh resolution mass spectrometry in soil cores across the continental United States into
521 tractable units. We disentangle these signatures of SOM composition into compounds that are
522 associated with soils exhibiting low versus high rates of potential respiration. These compounds
523 are consistent with prevailing understandings of SOM bioavailability and further suggest
524 chemical recalcitrance as an important mechanism of soil C stabilization in surface soils.
525 Additionally, SOM chemistry (as summarized by NMF k) explained a greater proportion of
526 potential soil respiration than commonly measured physicochemical factors, and provided
527 additional explanatory power beyond these factors in combined models. Our results provide a
528 basis for molecular information to spur the development of new process-based representations of
529 soil C cycles and underscore the role of specific chemical constituents within the water-
530 extractable SOM as a determinant of soil respiration.

531

532 **Acknowledgement**

533 Soil data were provided by the Molecular Observation Network (MONet) at the Environmental
534 Molecular Sciences Laboratory (<https://ror.org/04rc0xn13>), a DOE Office of Science user
535 facility sponsored by the Biological and Environmental Research program under Contract No.
536 DE-AC05-76RL01830. The work (proposal: 10.46936/10.25585/60008970) conducted by the
537 U.S. Department of Energy, Joint Genome Institute (<https://ror.org/04xm1d337>), a DOE Office

538 of Science user facility, is supported by the Office of Science of the U.S. Department of Energy
539 operated under Contract No. DE-AC02-05CH11231.

540 The Molecular Observation Network (MONet) database is an open, FAIR, and publicly available
541 compilation of the molecular and microstructural properties of soil. Data in the MONet open
542 science database can be found at <https://www.emsl.pnnl.gov/monet>.

543 References

544

- 545 1. Giardina, C. P.; Litton, C. M.; Crow, S. E.; Asner, G. P., Warming-related increases in soil CO₂ efflux are
546 explained by increased below-ground carbon flux. *Nature Climate Change* **2014**, *4*, (9), 822-827.
- 547 2. Jian, J.; Vargas, R.; Anderson-Teixeira, K.; Stell, E.; Herrmann, V.; Horn, M.; Kholod, N.; Manzon, J.;
548 Marchesi, R.; Paredes, D.; Bond-Lamberty, B., A restructured and updated global soil respiration database (SRDB-
549 V5). *Earth Syst. Sci. Data* **2021**, *13*, (2), 255-267.
- 550 3. Friedlingstein, P.; O'Sullivan, M.; Jones, M. W.; Andrew, R. M.; Gregor, L.; Hauck, J.; Le Quéré, C.;
551 Lujckx, I. T.; Olsen, A.; Peters, G. P.; Peters, W.; Pongratz, J.; Schwingshackl, C.; Sitch, S.; Canadell, J. G.; Ciais,
552 P.; Jackson, R. B.; Alin, S. R.; Alkama, R.; Arneth, A.; Arora, V. K.; Bates, N. R.; Becker, M.; Bellouin, N.; Bittig,
553 H. C.; Bopp, L.; Chevallier, F.; Chini, L. P.; Cronin, M.; Evans, W.; Falk, S.; Feely, R. A.; Gasser, T.; Gehlen, M.;
554 Gkritzalis, T.; Gloege, L.; Grassi, G.; Gruber, N.; Gürses, Ö.; Harris, I.; Hefner, M.; Houghton, R. A.; Hurtt, G. C.;
555 Iida, Y.; Ilyina, T.; Jain, A. K.; Jersild, A.; Kadono, K.; Kato, E.; Kennedy, D.; Klein Goldewijk, K.; Knauer, J.;
556 Korsbakken, J. I.; Landschützer, P.; Lefèvre, N.; Lindsay, K.; Liu, J.; Liu, Z.; Marland, G.; Mayot, N.; McGrath, M.
557 J.; Metz, N.; Monacci, N. M.; Munro, D. R.; Nakaoka, S. I.; Niwa, Y.; O'Brien, K.; Ono, T.; Palmer, P. I.; Pan, N.;
558 Pierrot, D.; Pocol, K.; Poulter, B.; Resplandy, L.; Robertson, E.; Rödenbeck, C.; Rodriguez, C.; Rosan, T. M.;
559 Schwingler, J.; Séférian, R.; Shutler, J. D.; Skjelvan, I.; Steinhoff, T.; Sun, Q.; Sutton, A. J.; Sweeney, C.; Takao, S.;
560 Tanhua, T.; Tans, P. P.; Tian, X.; Tian, H.; Tilbrook, B.; Tsjino, H.; Tubiello, F.; van der Werf, G. R.; Walker, A.
561 P.; Wanninkhof, R.; Whitehead, C.; Willstrand Wranne, A.; Wright, R.; Yuan, W.; Yue, C.; Yue, X.; Zaehle, S.;
562 Zeng, J.; Zheng, B., Global Carbon Budget 2022. *Earth Syst. Sci. Data* **2022**, *14*, (11), 4811-4900.
- 563 4. Graham, E. B.; Hofmockel, K. S., Ecological stoichiometry as a foundation for omics-enabled
564 biogeochemical models of soil organic matter decomposition. *Biogeochemistry* **2022**, *157*, (1), 31-50.
- 565 5. Lei, J.; Guo, X.; Zeng, Y.; Zhou, J.; Gao, Q.; Yang, Y., Temporal changes in global soil respiration since
566 1987. *Nature communications* **2021**, *12*, (1), 403.
- 567 6. Nissan, A.; Alcolombri, U.; Peleg, N.; Galili, N.; Jimenez-Martinez, J.; Molnar, P.; Holzner, M., Global
568 warming accelerates soil heterotrophic respiration. *Nature Communications* **2023**, *14*, (1), 3452.
- 569 7. Melillo, J. M.; Frey, S. D.; DeAngelis, K. M.; Werner, W. J.; Bernard, M. J.; Bowles, F. P.; Pold, G.;
570 Knorr, M. A.; Grandy, A. S., Long-term pattern and magnitude of soil carbon feedback to the climate system in a
571 warming world. *Science* **2017**, *358*, (6359), 101-105.
- 572 8. Bond-Lamberty, B.; Thomson, A., Temperature-associated increases in the global soil respiration record.
573 *Nature* **2010**, *464*, (7288), 579-582.
- 574 9. Todd-Brown, K. E.; Randerson, J. T.; Post, W. M.; Hoffman, F. M.; Tarnocai, C.; Schuur, E. A.; Allison, S.
575 D., Causes of variation in soil carbon simulations from CMIP5 Earth system models and comparison with
576 observations. *Biogeosciences* **2013**, *10*, (3), 1717-1736.
- 577 10. Todd-Brown, K.; Randerson, J.; Hopkins, F.; Arora, V.; Hajima, T.; Jones, C.; Shevliakova, E.; Tjiputra, J.;
578 Volodin, E.; Wu, T., Changes in soil organic carbon storage predicted by Earth system models during the 21st
579 century. *Biogeosciences* **2014**, *11*, (8), 2341-2356.
- 580 11. Bradford, M. A.; Wieder, W. R.; Bonan, G. B.; Fierer, N.; Raymond, P. A.; Crowther, T. W., Managing
581 uncertainty in soil carbon feedbacks to climate change. *Nature Climate Change* **2016**, *6*, (8), 751-758.
- 582 12. Davidson, E. A.; Janssens, I. A., Temperature sensitivity of soil carbon decomposition and feedbacks to
583 climate change. *Nature* **2006**, *440*, (7081), 165-173.
- 584 13. Warner, D.; Bond-Lamberty, B.; Jian, J.; Stell, E.; Vargas, R., Spatial predictions and associated
585 uncertainty of annual soil respiration at the global scale. *Global Biogeochemical Cycles* **2019**, *33*, (12), 1733-1745.
- 586 14. Crowther, T. W.; Todd-Brown, K. E. O.; Rowe, C. W.; Wieder, W. R.; Carey, J. C.; Machmuller, M. B.;
587 Snoek, B. L.; Fang, S.; Zhou, G.; Allison, S. D.; Blair, J. M.; Bridgham, S. D.; Burton, A. J.; Carrillo, Y.; Reich, P.
588 B.; Clark, J. S.; Classen, A. T.; Dijkstra, F. A.; Elberling, B.; Emmett, B. A.; Estiarte, M.; Frey, S. D.; Guo, J.;
589 Harte, J.; Jiang, L.; Johnson, B. R.; Kröel-Dulay, G.; Larsen, K. S.; Laudon, H.; Lavallee, J. M.; Luo, Y.; Lupascu,
590 M.; Ma, L. N.; Marhan, S.; Michelsen, A.; Mohan, J.; Niu, S.; Pendall, E.; Peñuelas, J.; Pfeifer-Meister, L.; Poll, C.;
591 Reinsch, S.; Reynolds, L. L.; Schmidt, I. K.; Sistla, S.; Sokol, N. W.; Templer, P. H.; Treseder, K. K.; Welker, J. M.;
592 Bradford, M. A., Quantifying global soil carbon losses in response to warming. *Nature* **2016**, *540*, (7631), 104-108.
- 593 15. Billings, S. A.; Lajtha, K.; Malhotra, A.; Berhe, A. A.; de Graaff, M. A.; Earl, S.; Fraterrigo, J.; Georgiou,
594 K.; Grandy, S.; Hobbie, S. E., Soil organic carbon is not just for soil scientists: measurement recommendations for
595 diverse practitioners. *Ecological Applications* **2021**, *31*, (3), e02290.

- 596 16. Liang, C.; Amelung, W.; Lehmann, J.; Kästner, M., Quantitative assessment of microbial necromass
597 contribution to soil organic matter. *Global change biology* **2019**, *25*, (11), 3578-3590.
- 598 17. Sanderman, J.; Baldock, J. A.; Dangal, S. R. S.; Ludwig, S.; Potter, S.; Rivard, C.; Savage, K., Soil organic
599 carbon fractions in the Great Plains of the United States: an application of mid-infrared spectroscopy.
600 *Biogeochemistry* **2021**, *156*, (1), 97-114.
- 601 18. Bahureksa, W.; Tfaily, M. M.; Boiteau, R. M.; Young, R. B.; Logan, M. N.; McKenna, A. M.; Borch, T.,
602 Soil organic matter characterization by Fourier transform ion cyclotron resonance mass spectrometry (FTICR MS):
603 A critical review of sample preparation, analysis, and data interpretation. *Environmental science & technology* **2021**,
604 *55*, (14), 9637-9656.
- 605 19. Cotrufo, M. F.; Wallenstein, M. D.; Boot, C. M.; Deneff, K.; Paul, E., The Microbial Efficiency-Matrix S
606 tabilization (MEMS) framework integrates plant litter decomposition with soil organic matter stabilization: Do labile
607 plant inputs form stable soil organic matter? *Global change biology* **2013**, *19*, (4), 988-995.
- 608 20. Sulman, B. N.; Phillips, R. P.; Oishi, A. C.; Shevliakova, E.; Pacala, S. W., Microbe-driven turnover offsets
609 mineral-mediated storage of soil carbon under elevated CO₂. *Nature Climate Change* **2014**, *4*, (12), 1099-1102.
- 610 21. Robertson, A. D.; Paustian, K.; Ogle, S.; Wallenstein, M. D.; Lugato, E.; Cotrufo, M. F., Unifying soil
611 organic matter formation and persistence frameworks: the MEMS model. *Biogeosciences* **2019**, *16*, (6), 1225-1248.
- 612 22. Falloon, P.; Jones, C. D.; Ades, M.; Paul, K., Direct soil moisture controls of future global soil carbon
613 changes: An important source of uncertainty. *Global Biogeochemical Cycles* **2011**, *25*, (3).
- 614 23. Ciais, P.; Sabine, C.; Bala, G.; Bopp, L.; Brovkin, V.; Canadell, J.; Chhabra, A.; DeFries, R.; Galloway, J.;
615 Heimann, M., Carbon and other biogeochemical cycles. In *Climate change 2013: the physical science basis. Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*, Cambridge University Press: 2014; pp 465-570.
- 616 24. Amador, J.; Jones, R. D., Nutrient limitations on microbial respiration in peat soils with different total
617 phosphorus content. *Soil Biology and Biochemistry* **1993**, *25*, (6), 793-801.
- 618 25. Malik, A. A.; Puissant, J.; Buckeridge, K. M.; Goodall, T.; Jehmlich, N.; Chowdhury, S.; Gweon, H. S.;
619 Peyton, J. M.; Mason, K. E.; van Agtmaal, M., Land use driven change in soil pH affects microbial carbon cycling
620 processes. *Nature communications* **2018**, *9*, (1), 3591.
- 621 26. Fang, C.; Moncrieff, J. B., The variation of soil microbial respiration with depth in relation to soil carbon
622 composition. *Plant and Soil* **2005**, *268*, 243-253.
- 623 27. Curiel Yuste, J.; Baldocchi, D.; Gershenson, A.; Goldstein, A.; Misson, L.; Wong, S., Microbial soil
624 respiration and its dependency on carbon inputs, soil temperature and moisture. *Global Change Biology* **2007**, *13*,
625 (9), 2018-2035.
- 626 28. Billings, S. A.; Ballantyne IV, F., How interactions between microbial resource demands, soil organic
627 matter stoichiometry, and substrate reactivity determine the direction and magnitude of soil respiratory responses to
628 warming. *Global Change Biology* **2013**, *19*, (1), 90-102.
- 629 29. Raich, J. W.; Potter, C. S., Global patterns of carbon dioxide emissions from soils. *Global biogeochemical*
630 *cycles* **1995**, *9*, (1), 23-36.
- 631 30. Raich, J. W.; Potter, C. S.; Bhagawati, D., Interannual variability in global soil respiration, 1980–94.
632 *Global Change Biology* **2002**, *8*, (8), 800-812.
- 633 31. Kyker-Snowman, E.; Wieder, W. R.; Frey, S. D.; Grandy, A. S., Stoichiometrically coupled carbon and
634 nitrogen cycling in the Microbial-Mineral Carbon Stabilization model version 1.0 (MIMICS-CN v1.0).
635 *Geoscientific Model Development* **2020**, *13*, (9), 4413-4434.
- 636 32. Wieder, W. R.; Hartman, M. D.; Sulman, B. N.; Wang, Y. P.; Koven, C. D.; Bonan, G. B., Carbon cycle
637 confidence and uncertainty: Exploring variation among soil biogeochemical models. *Global change biology* **2018**,
638 *24*, (4), 1563-1579.
- 639 33. Waring, B. G.; Sulman, B. N.; Reed, S.; Smith, A. P.; Averill, C.; Creamer, C. A.; Cusack, D. F.; Hall, S.
640 J.; Jastrow, J. D.; Jilling, A., From pools to flow: The PROMISE framework for new insights on soil carbon cycling
641 in a changing world. *Global Change Biology* **2020**, *26*, (12), 6631-6643.
- 642 34. Song, H.-S.; Stegen, J. C.; Graham, E. B.; Lee, J.-Y.; Garayburu-Caruso, V. A.; Nelson, W. C.; Chen, X.;
643 Moulton, J. D.; Scheibe, T. D., Representing organic matter thermodynamics in biogeochemical reactions via
644 substrate-explicit modeling. *Frontiers in Microbiology* **2020**, *11*, 531756.
- 645 35. Bradford, M. A.; Wood, S. A.; Addicott, E. T.; Fenichel, E. P.; Fields, N.; González-Rivero, J.; Jevon, F.
646 V.; Maynard, D. S.; Oldfield, E. E.; Polussa, A.; Ward, E. B.; Wieder, W. R., Quantifying microbial control of soil
647 organic matter dynamics at macrosystem scales. *Biogeochemistry* **2021**, *156*, (1), 19-40.
- 648 36. Hall, S. J.; Ye, C.; Weintraub, S. R.; Hockaday, W. C., Molecular trade-offs in soil organic carbon
649 composition at continental scale. *Nature Geoscience* **2020**, *13*, (10), 687-692.

- 652 37. Huys, R.; Poirier, V.; Bourget, M. Y.; Roumet, C.; Hättenschwiler, S.; Fromin, N.; Munson, A. D.;
653 Freschet, G. T., Plant litter chemistry controls coarse-textured soil carbon dynamics. *Journal of Ecology* **2022**, *110*,
654 (12), 2911-2928.
- 655 38. Scott, N. A.; Cole, C. V.; Elliott, E. T.; Huffman, S. A., Soil textural control on decomposition and soil
656 organic matter dynamics. *Soil Science Society of America Journal* **1996**, *60*, (4), 1102-1109.
- 657 39. Witzgall, K.; Vidal, A.; Schubert, D. I.; Höschen, C.; Schweizer, S. A.; Buegger, F.; Pouteau, V.; Chenu,
658 C.; Mueller, C. W., Particulate organic matter as a functional soil component for persistent soil organic carbon.
659 *Nature Communications* **2021**, *12*, (1), 4115.
- 660 40. Tureçcaia, A. B.; Garayburu-Caruso, V. A.; Kaufman, M. H.; Danczak, R. E.; Stegen, J. C.; Chu, R. K.;
661 Toyoda, J. G.; Cardenas, M. B.; Graham, E. B., Rethinking Aerobic Respiration in the Hyporheic Zone under
662 Variation in Carbon and Nitrogen Stoichiometry. *Environmental Science & Technology* **2023**, *57*, (41), 15499-
663 15510.
- 664 41. Garayburu-Caruso, V. A.; Stegen, J. C.; Song, H.-S.; Renteria, L.; Wells, J.; Garcia, W.; Resch, C. T.;
665 Goldman, A. E.; Chu, R. K.; Toyoda, J., Carbon limitation leads to thermodynamic regulation of aerobic
666 metabolism. *Environmental Science & Technology Letters* **2020**, *7*, (7), 517-524.
- 667 42. Hodgkins, S. B.; Tfaily, M. M.; McCalley, C. K.; Logan, T. A.; Crill, P. M.; Saleska, S. R.; Rich, V. I.;
668 Chanton, J. P., Changes in peat chemistry associated with permafrost thaw increase greenhouse gas production.
669 *Proceedings of the National Academy of Sciences* **2014**, *111*, (16), 5819-5824.
- 670 43. Li, H.; Bölscher, T.; Winnick, M.; Tfaily, M. M.; Cardon, Z. G.; Keiluweit, M., Simple plant and microbial
671 exudates destabilize mineral-associated organic matter via multiple pathways. *Environmental science & technology*
672 **2021**, *55*, (5), 3389-3398.
- 673 44. Graham, E. B.; Crump, A. R.; Kennedy, D. W.; Arntzen, E.; Fansler, S.; Purvine, S. O.; Nicora, C. D.;
674 Nelson, W.; Tfaily, M. M.; Stegen, J. C., Multi'omics comparison reveals metabolome biochemistry, not
675 microbiome composition or gene expression, corresponds to elevated biogeochemical function in the hyporheic
676 zone. *Science of the total environment* **2018**, *642*, 742-753.
- 677 45. Graham, E. B.; Tfaily, M. M.; Crump, A. R.; Goldman, A. E.; Bramer, L. M.; Arntzen, E.; Romero, E.;
678 Resch, C. T.; Kennedy, D. W.; Stegen, J. C., Carbon inputs from riparian vegetation limit oxidation of physically
679 bound organic carbon via biochemical and thermodynamic processes. *Journal of Geophysical Research:*
680 *Biogeosciences* **2017**, *122*, (12), 3188-3205.
- 681 46. Sonnwald, M.; Dutkiewicz, S.; Hill, C.; Forget, G., Elucidating ecological complexity: Unsupervised
682 learning determines global marine eco-provinces. *Science Advances* **2020**, *6*, (22), eaay4740.
- 683 47. Nishiyama, E.; Higashi, K.; Mori, H.; Suda, K.; Nakamura, H.; Omori, S.; Maruyama, S.; Hongoh, Y.;
684 Kurokawa, K., The Relationship Between Microbial Community Structures and Environmental Parameters
685 Revealed by Metagenomic Analysis of Hot Spring Water in the Kirishima Area, Japan. *Frontiers in bioengineering*
686 *and biotechnology* **2018**, *6*, 202.
- 687 48. Ortner, M.; Seidel, M.; Semella, S.; Udelhoven, T.; Vohland, M.; Thiele-Bruhn, S., Content of soil organic
688 carbon and labile fractions depend on local combinations of mineral-phase characteristics. *SOIL* **2022**, *8*, (1), 113-
689 131.
- 690 49. Ukalska-Jaruga, A.; Smreczak, B.; Klimkowicz-Pawlas, A., Soil organic matter composition as a factor
691 affecting the accumulation of polycyclic aromatic hydrocarbons. *Journal of Soils and Sediments* **2019**, *19*, (4), 1890-
692 1900.
- 693 50. Shigyo, N.; Furusawa, H.; Yamashita, N.; Nagakura, J.; Manaka, T.; Yamada, T.; Hirai, K., Slope-induced
694 factors shape bacterial communities in surface soils in a forested headwater catchment. *CATENA* **2022**, *214*, 106253.
- 695 51. Massoni, J.; Bortfeld-Miller, M.; Widmer, A.; Vorholt, J. A., Capacity of soil bacteria to reach the
696 phyllosphere and convergence of floral communities despite soil microbiota variation. *Proceedings of the National*
697 *Academy of Sciences* **2021**, *118*, (41), e2100150118.
- 698 52. Simister, R. L.; Iulianella Phillips, B. P.; Wickham, A. P.; Cayer, E. M.; Hart, C. J. R.; Winterburn, P. A.;
699 Crowe, S. A., DNA sequencing, microbial indicators, and the discovery of buried kimberlites. *Communications*
700 *Earth & Environment* **2023**, *4*, (1), 387.
- 701 53. Bowman, M. M.; Heath, A. E.; Varga, T.; Battu, A. K.; Chu, R. K.; Toyoda, J.; Cheeke, T. E.; Porter, S. S.;
702 Moffett, K. B.; LeTendre, B.; Qafoku, O.; Bargar, J. R.; Mans, D. M.; Hess, N. J.; Graham, E. B., One thousand
703 soils for molecular understanding of belowground carbon cycling. *Frontiers in Soil Science* **2023**, *3*.
- 704 54. Brookes, P.; Landman, A.; Pruden, G.; Jenkinson, D., Chloroform fumigation and the release of soil
705 nitrogen: a rapid direct extraction method to measure microbial biomass nitrogen in soil. *Soil biology and*
706 *biochemistry* **1985**, *17*, (6), 837-842.

- 707 55. Witt, C.; Gaunt, J. L.; Galicia, C. C.; Ottow, J. C.; Neue, H.-U., A rapid chloroform-fumigation extraction
708 method for measuring soil microbial biomass carbon and nitrogen in flooded rice soils. *Biology and Fertility of Soils*
709 **2000**, *30*, 510-519.
- 710 56. Zhao, Q.; Thompson, A. M.; Callister, S. J.; Tfaily, M. M.; Bell, S. L.; Hobbie, S. E.; Hofmockel, K. S.,
711 Dynamics of organic matter molecular composition under aerobic decomposition and their response to the nitrogen
712 addition in grassland soils. *Science of the Total Environment* **2022**, *806*, 150514.
- 713 57. Corbridge, D. E. C., *Phosphorus. An outline of its chemistry, biochemistry, and technology*. Elsevier
714 Scientific Co.: 1980.
- 715 58. Bray, R. H.; Kurtz, L. T., Determination of total, organic, and available forms of phosphorus in soils. *Soil*
716 *science* **1945**, *59*, (1), 39-46.
- 717 59. AOAC, I., AOAC Official Method 972.43, Microchemical determination of carbon, hydrogen, and
718 nitrogen, automated method. *Official Methods of Analysis of AOAC International*. AOAC International,
719 Gaithersburg, MD **2006**, 5-6.
- 720 60. Dittmar, T.; Koch, B.; Hertkorn, N.; Kattner, G., A simple and efficient method for the solid-phase
721 extraction of dissolved organic matter (SPE-DOM) from seawater. *Limnology and Oceanography: Methods* **2008**, *6*,
722 (6), 230-235.
- 723 61. Corilo, Y.; Kew, W.; McCue, L., EMSL-Computing/CoreMS: CoreMS 1.0. 0 (v1. 0.0). *Zenodo* **2021**,
724 *5281*.
- 725 62. Kim, S.; Kramer, R. W.; Hatcher, P. G., Graphical method for analysis of ultrahigh-resolution broadband
726 mass spectra of natural organic matter, the van Krevelen diagram. *Analytical chemistry* **2003**, *75*, (20), 5336-5344.
- 727 63. Tfaily, M. M.; Chu, R. K.; Tolić, N.; Roscioli, K. M.; Anderton, C. R.; Paša-Tolić, L.; Robinson, E. W.;
728 Hess, N. J., Advanced solvent based methods for molecular characterization of soil organic matter by high-
729 resolution mass spectrometry. *Analytical chemistry* **2015**, *87*, (10), 5206-5215.
- 730 64. Lee, D.; Seung, H. S., Algorithms for non-negative matrix factorization. *Advances in neural information*
731 *processing systems* **2000**, *13*.
- 732 65. Bhattarai, M.; Chennupati, G.; Skau, E.; Vangara, R.; Djidjev, H.; Alexandrov, B. S. In *Distributed Non-*
733 *Negative Tensor Train Decomposition*, 2020 IEEE High Performance Extreme Computing Conference (HPEC), 22-
734 24 Sept. 2020, 2020; 2020; pp 1-10.
- 735 66. Devarajan, K., Nonnegative Matrix Factorization: An Analytical and Interpretive Tool in Computational
736 Biology. *PLOS Computational Biology* **2008**, *4*, (7), e1000029.
- 737 67. Bholowalia, P.; Kumar, A., EBK-means: A clustering technique based on elbow method and k-means in
738 WSN. *International Journal of Computer Applications* **2014**, *105*, (9).
- 739 68. Riaz, M.; Marschner, P., Sandy Soil Amended with Clay Soil: Effect of Clay Soil Properties on Soil
740 Respiration, Microbial Biomass, and Water Extractable Organic C. *Journal of Soil Science and Plant Nutrition*
741 **2020**, *20*, (4), 2465-2470.
- 742 69. Lee, K.-H.; Jose, S., Soil respiration, fine root production, and microbial biomass in cottonwood and
743 loblolly pine plantations along a nitrogen fertilization gradient. *Forest Ecology and Management* **2003**, *185*, (3),
744 263-273.
- 745 70. Chen, S.; Zou, J.; Hu, Z.; Chen, H.; Lu, Y., Global annual soil respiration in relation to climate, soil
746 properties and vegetation characteristics: Summary of available data. *Agricultural and Forest Meteorology* **2014**,
747 *198-199*, 335-346.
- 748 71. Orchard, V. A.; Cook, F., Relationship between soil respiration and soil moisture. *Soil Biology and*
749 *Biochemistry* **1983**, *15*, (4), 447-453.
- 750 72. Xu, L.; Baldocchi, D. D.; Tang, J., How soil moisture, rain pulses, and growth alter the response of
751 ecosystem respiration to temperature. *Global Biogeochemical Cycles* **2004**, *18*, (4).
- 752 73. Moyano, F. E.; Manzoni, S.; Chenu, C., Responses of soil heterotrophic respiration to moisture availability:
753 An exploration of processes and models. *Soil Biology and Biochemistry* **2013**, *59*, 72-85.
- 754 74. Wang, Y.-P.; Houlton, B. Z., Nitrogen constraints on terrestrial carbon uptake: Implications for the global
755 carbon-climate feedback. *Geophysical Research Letters* **2009**, *36*, (24).
- 756 75. Elser, J.; Sterner, R.; Gorokhova, E. a.; Fagan, W.; Markow, T.; Cotner, J.; Harrison, J.; Hobbie, S.; Odell,
757 G.; Weider, L., Biological stoichiometry from genes to ecosystems. *Ecology letters* **2000**, *3*, (6), 540-550.
- 758 76. Soong, J. L.; Fuchsluger, L.; Marañon-Jimenez, S.; Torn, M. S.; Janssens, I. A.; Penuelas, J.; Richter, A.,
759 Microbial carbon limitation: The need for integrating microorganisms into our understanding of ecosystem carbon
760 cycling. *Global change biology* **2020**, *26*, (4), 1953-1961.
- 761 77. Fang, C.; Moncrieff, J. B., The variation of soil microbial respiration with depth in relation to soil carbon
762 composition. *Plant and Soil* **2005**, *268*, (1), 243-253.

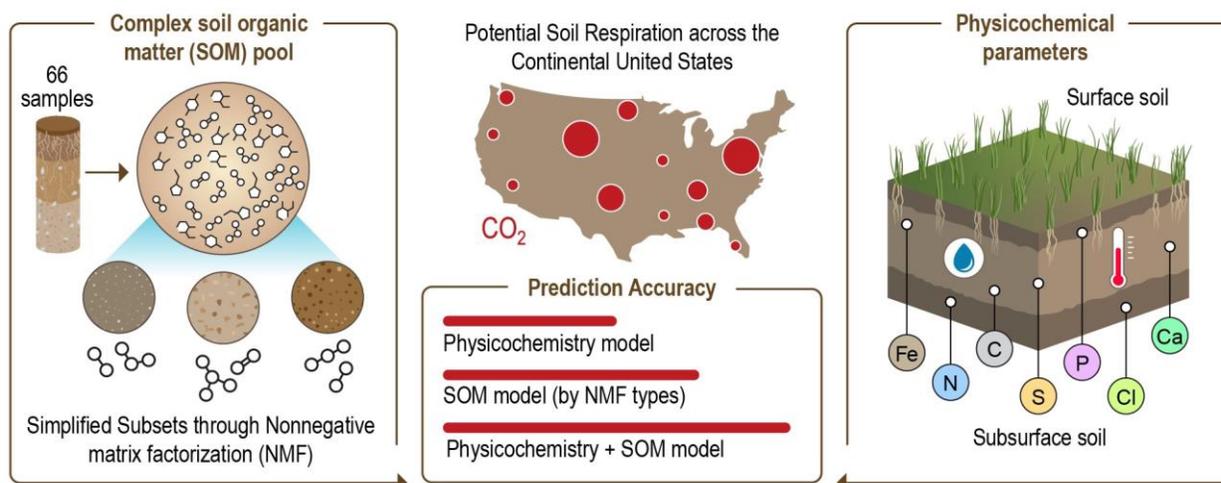
- 763 78. Schimel, J., The Democracy of dirt: relating micro-scale dynamics to macro-scale ecosystem function.
764 *Microbes: The foundation stone of the biosphere* **2021**, 89-102.
- 765 79. Nicolás, C.; Martín-Bertelsen, T.; Floudas, D.; Bentzer, J.; Smits, M.; Johansson, T.; Troein, C.; Persson,
766 P.; Tunlid, A., The soil organic matter decomposition mechanisms in ectomycorrhizal fungi are tuned for liberating
767 soil organic nitrogen. *The ISME Journal* **2019**, *13*, (4), 977-988.
- 768 80. Subedi, P.; Jokela, E. J.; Vogel, J. G.; Bracho, R.; Inglett, K. S., The effects of nutrient limitations on
769 microbial respiration and organic matter decomposition in a Florida Spodosol as influenced by historical forest
770 management practices. *Forest Ecology and Management* **2021**, *479*, 118592.
- 771 81. Mori, T.; Lu, X.; Aoyagi, R.; Mo, J., Reconsidering the phosphorus limitation of soil microbial activity in
772 tropical forests. *Functional Ecology* **2018**, *32*, (5), 1145-1154.
- 773 82. Fan, B.; Yin, L.; Dijkstra, F. A.; Lu, J.; Shao, S.; Wang, P.; Wang, Q.; Cheng, W., Potential gross nitrogen
774 mineralization and its linkage with microbial respiration along a forest transect in eastern China. *Applied Soil*
775 *Ecology* **2022**, *171*, 104347.
- 776 83. Gransee, A.; Führs, H., Magnesium mobility in soils as a challenge for soil and plant analysis, magnesium
777 fertilization and root uptake under adverse growth conditions. *Plant and Soil* **2013**, *368*, (1), 5-21.
- 778 84. Opfergelt, S.; Cornélis, J. T.; Houben, D.; Givron, C.; Burton, K. W.; Mattielli, N., The influence of
779 weathering and soil organic matter on Zn isotopes in soils. *Chemical Geology* **2017**, *466*, 140-148.
- 780 85. Chao, L.; Liu, Y.; Freschet, G. T.; Zhang, W.; Yu, X.; Zheng, W.; Guan, X.; Yang, Q.; Chen, L.; Dijkstra,
781 F. A., Litter carbon and nutrient chemistry control the magnitude of soil priming effect. *Functional Ecology* **2019**,
782 *33*, (5), 876-888.
- 783 86. Neupane, A.; Herndon, E. M.; Whitman, T.; Faiia, A. M.; Jagadamma, S., Manganese effects on plant
784 residue decomposition and carbon distribution in soil fractions depend on soil nitrogen availability. *Soil Biology and*
785 *Biochemistry* **2023**, *178*, 108964.
- 786 87. Kranabetter, J. M.; Philpott, T.; Dunn, D., Manganese limitations and the enhanced soil carbon
787 sequestration of temperate rainforests. *Biogeochemistry* **2021**, *156*, (2), 195-209.
- 788 88. Santos, F.; Herndon, E., Plant-Soil Relationships Influence Observed Trends Between Manganese and
789 Carbon Across Biomes. *Global Biogeochemical Cycles* **2023**, *37*, (1), e2022GB007412.
- 790 89. Li, H.; Santos, F.; Butler, K.; Herndon, E., A critical review on the multiple roles of manganese in
791 stabilizing and destabilizing soil organic matter. *Environmental science & technology* **2021**, *55*, (18), 12136-12152.
- 792 90. Whalen, E. D. Manganese Limitation as a Mechanism for Reduced Decomposition in Soils under Long-
793 Term Atmospheric Nitrogen Deposition. University of New Hampshire, 2017.
- 794 91. Taylor, P. G.; Townsend, A. R., Stoichiometric control of organic carbon–nitrate relationships from soils to
795 the sea. *Nature* **2010**, *464*, (7292), 1178-1181.
- 796 92. Zhang, M.; Zhang, X.; Zhang, L.; Zeng, L.; Liu, Y.; Wang, X.; He, P.; Li, S.; Liang, G.; Zhou, W.; Ai, C.,
797 The stronger impact of inorganic nitrogen fertilization on soil bacterial community than organic fertilization in
798 short-term condition. *Geoderma* **2021**, *382*, 114752.
- 799 93. Siles, J. A.; Cajthaml, T.; Filipová, A.; Minerbi, S.; Margesin, R., Altitudinal, seasonal and interannual
800 shifts in microbial communities and chemical composition of soil organic matter in Alpine forest soils. *Soil Biology*
801 *and Biochemistry* **2017**, *112*, 1-13.
- 802 94. Berryman, E. M.; Marshall, J. D.; Kavanagh, K., Decoupling litter respiration from whole-soil respiration
803 along an elevation gradient in a Rocky Mountain mixed-conifer forest. *Canadian Journal of Forest Research* **2014**,
804 *44*, (5), 432-440.
- 805 95. Brye, K. R.; McMullen, R. L.; Silveira, M. L.; Motschenbacher, J. M. D.; Smith, S. F.; Gbur, E. E.; Helton,
806 M. L., Environmental controls on soil respiration across a southern US climate gradient: a meta-analysis. *Geoderma*
807 *Regional* **2016**, *7*, (2), 110-119.
- 808 96. Kallenbach, C. M.; Frey, S. D.; Grandy, A. S., Direct evidence for microbial-derived soil organic matter
809 formation and its ecophysiological controls. *Nature Communications* **2016**, *7*, (1), 13630.
- 810 97. Roth, V.-N.; Lange, M.; Simon, C.; Hertkorn, N.; Bucher, S.; Goodall, T.; Griffiths, R. I.; Mellado-
811 Vázquez, P. G.; Mommer, L.; Oram, N. J.; Weigelt, A.; Dittmar, T.; Gleixner, G., Persistence of dissolved organic
812 matter explained by molecular changes during its passage through soil. *Nature Geoscience* **2019**, *12*, (9), 755-761.
- 813 98. Zhao, Q.; Callister, S. J.; Thompson, A. M.; Kukkadapu, R. K.; Tfaily, M. M.; Bramer, L. M.; Qafoku, N.
814 P.; Bell, S. L.; Hobbie, S. E.; Seabloom, E. W.; Borer, E. T.; Hofmockel, K. S., Strong mineralogic control of soil
815 organic matter composition in response to nutrient addition across diverse grassland sites. *Science of The Total*
816 *Environment* **2020**, *736*, 137839.
- 817 99. Rumpel, C.; Kögel-Knabner, I., Deep soil organic matter—a key but poorly understood component of
818 terrestrial C cycle. *Plant and soil* **2011**, *338*, 143-158.

- 819 100. Kögel-Knabner, I., The macromolecular organic composition of plant and microbial residues as inputs to
820 soil organic matter. *Soil Biology and Biochemistry* **2002**, *34*, (2), 139-162.
- 821 101. Kramer, C.; Gleixner, G., Soil organic matter in soil depth profiles: Distinct carbon preferences of
822 microbial groups during carbon transformation. *Soil Biology and Biochemistry* **2008**, *40*, (2), 425-433.
- 823 102. Davenport, R.; Bowen, B. P.; Lynch, L. M.; Kosina, S. M.; Shabtai, I.; Northen, T. R.; Lehmann, J.,
824 Decomposition decreases molecular diversity and ecosystem similarity of soil organic matter. *Proceedings of the*
825 *National Academy of Sciences* **2023**, *120*, (25), e2303335120.
- 826 103. Jones, D. L., Amino acid biodegradation and its potential effects on organic nitrogen capture by plants. *Soil*
827 *biology and biochemistry* **1999**, *31*, (4), 613-622.
- 828 104. Marschner, B.; Kalbitz, K., Controls of bioavailability and biodegradability of dissolved organic matter in
829 soils. *Geoderma* **2003**, *113*, (3-4), 211-235.
- 830 105. Campbell, T. P.; Ulrich, D. E. M.; Toyoda, J.; Thompson, J.; Munsky, B.; Albright, M. B. N.; Bailey, V. L.;
831 Tfaily, M. M.; Dunbar, J., Microbial Communities Influence Soil Dissolved Organic Carbon Concentration by
832 Altering Metabolite Composition. *Frontiers in Microbiology* **2022**, *12*.
- 833 106. Hernández, D. L.; Hobbie, S. E., The effects of substrate composition, quantity, and diversity on microbial
834 activity. *Plant and Soil* **2010**, *335*, (1), 397-411.
- 835 107. Marschner, B.; Kalbitz, K., Controls of bioavailability and biodegradability of dissolved organic matter in
836 soils. *Geoderma* **2003**, *113*, (3), 211-235.
- 837 108. Camenzind, T.; Mason-Jones, K.; Mansour, I.; Rillig, M. C.; Lehmann, J., Formation of necromass-derived
838 soil organic carbon determined by microbial death pathways. *Nature Geoscience* **2023**, *16*, (2), 115-122.
- 839 109. Angst, G.; Mueller, K. E.; Nierop, K. G. J.; Simpson, M. J., Plant- or microbial-derived? A review on the
840 molecular composition of stabilized soil organic matter. *Soil Biology and Biochemistry* **2021**, *156*, 108189.
- 841 110. Angst, G.; Mueller, K. E.; Kögel-Knabner, I.; Freeman, K. H.; Mueller, C. W., Aggregation controls the
842 stability of lignin and lipids in clay-sized particulate and mineral associated organic matter. *Biogeochemistry* **2017**,
843 *132*, (3), 307-324.
- 844 111. Cotrufo, M. F.; Ranalli, M. G.; Haddix, M. L.; Six, J.; Lugato, E., Soil carbon storage informed by
845 particulate and mineral-associated organic matter. *Nature Geoscience* **2019**, *12*, (12), 989-994.
- 846 112. Benbi, D.; Boparai, A.; Brar, K., Decomposition of particulate organic matter is more sensitive to
847 temperature than the mineral associated organic matter. *Soil Biology and Biochemistry* **2014**, *70*, 183-192.
- 848 113. Lugato, E.; Lavallee, J. M.; Haddix, M. L.; Panagos, P.; Cotrufo, M. F., Different climate sensitivity of
849 particulate and mineral-associated soil organic matter. *Nature Geoscience* **2021**, *14*, (5), 295-300.
- 850 114. Lehmann, J.; Hansel, C. M.; Kaiser, C.; Kleber, M.; Maher, K.; Manzoni, S.; Nunan, N.; Reichstein, M.;
851 Schimel, J. P.; Torn, M. S.; Wieder, W. R.; Kögel-Knabner, I., Persistence of soil organic carbon caused by
852 functional complexity. *Nature Geoscience* **2020**, *13*, (8), 529-534.
- 853 115. Zhang, H.; Goll, D. S.; Wang, Y. P.; Ciais, P.; Wieder, W. R.; Abramoff, R.; Huang, Y.; Guenet, B.;
854 Prescher, A. K.; Viscarra Rossel, R. A., Microbial dynamics and soil physicochemical properties explain large-scale
855 variations in soil organic carbon. *Global Change Biology* **2020**, *26*, (4), 2668-2685.
- 856 116. Graham, E. B.; Wieder, W. R.; Leff, J. W.; Weintraub, S. R.; Townsend, A. R.; Cleveland, C. C.; Philippot,
857 L.; Nemergut, D. R., Do we need to understand microbial communities to predict ecosystem function? A
858 comparison of statistical models of nitrogen cycling processes. *Soil Biology and Biochemistry* **2014**, *68*, 279-282.
- 859 117. Allison, S., A trait-based approach for modelling microbial litter decomposition. *Ecology letters* **2012**, *15*,
860 (9), 1058-1070.

861

862

863



864

865 Figure 1. Proposed workflow: Machine learning models summarize molecular data to predict soil
 866 respiration. Non-negative matrix factorization (NMF k) extracts key SOM signatures from high
 867 resolution mass spectrometry measurements of SOM. Gradient boosting regression predicts soil
 868 respiration with physicochemistry, SOM signatures, and physicochemistry combined with SOM
 869 signatures.

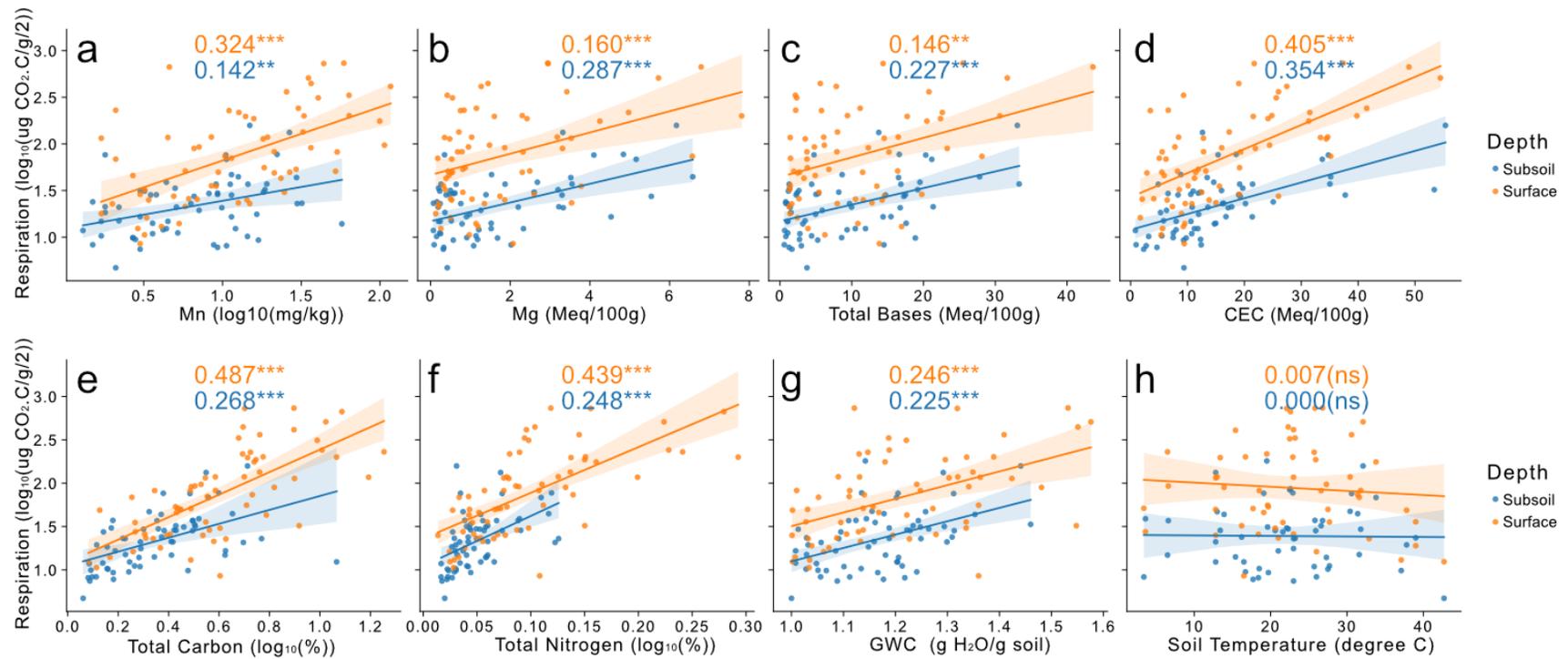


Figure 2. The relationship between soil characteristics and potential respiration. (a-h) show [Manganese(Mn), Magnesium(Mg), Total Bases, CEC, Total C, Total N, GWC, Soil Temperature], respectively. Orange represents surface soils and blue represents subsoils. Lines denote the fitted linear regression function. Numbers on each panel are r^2 value from linear regression, the stars behind represents statistical significance (***) ($p \leq 0.001$), ** ($p \leq 0.01$), ns ($p > 0.05$)).

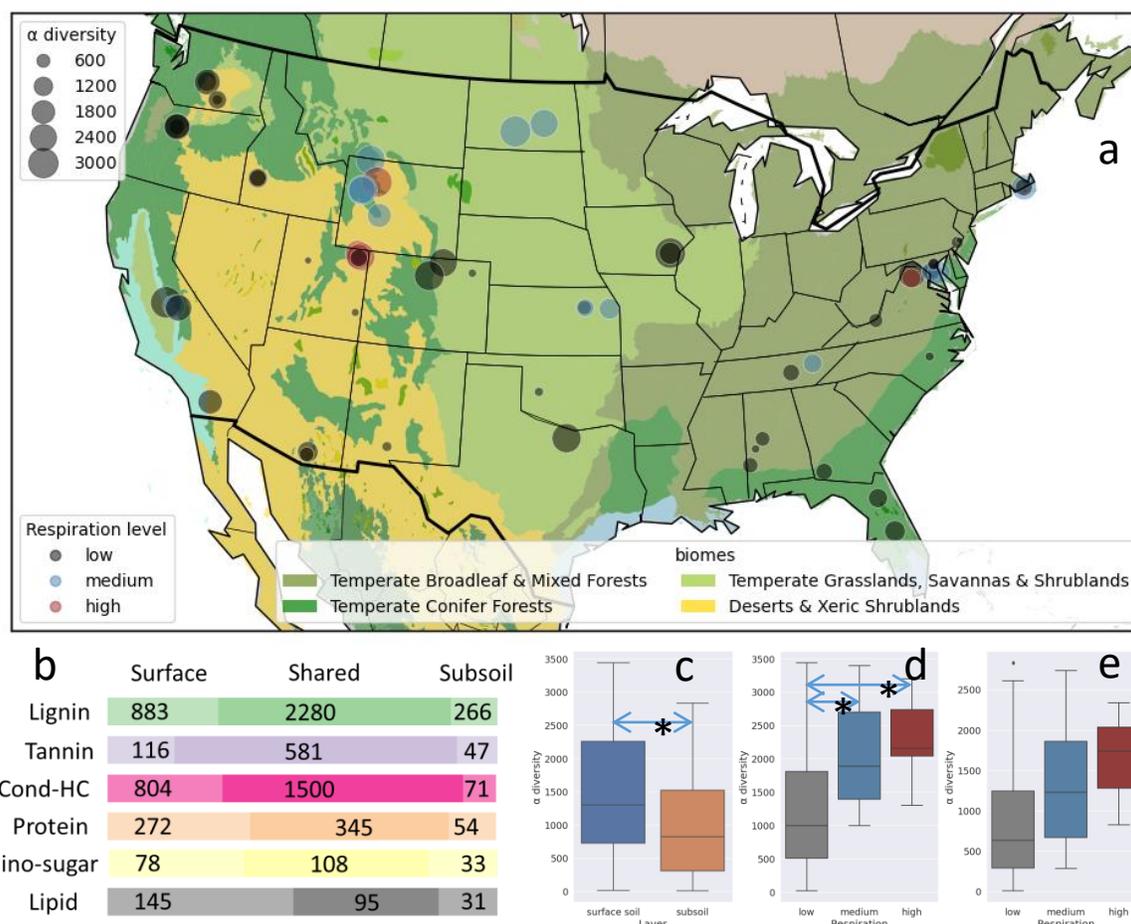
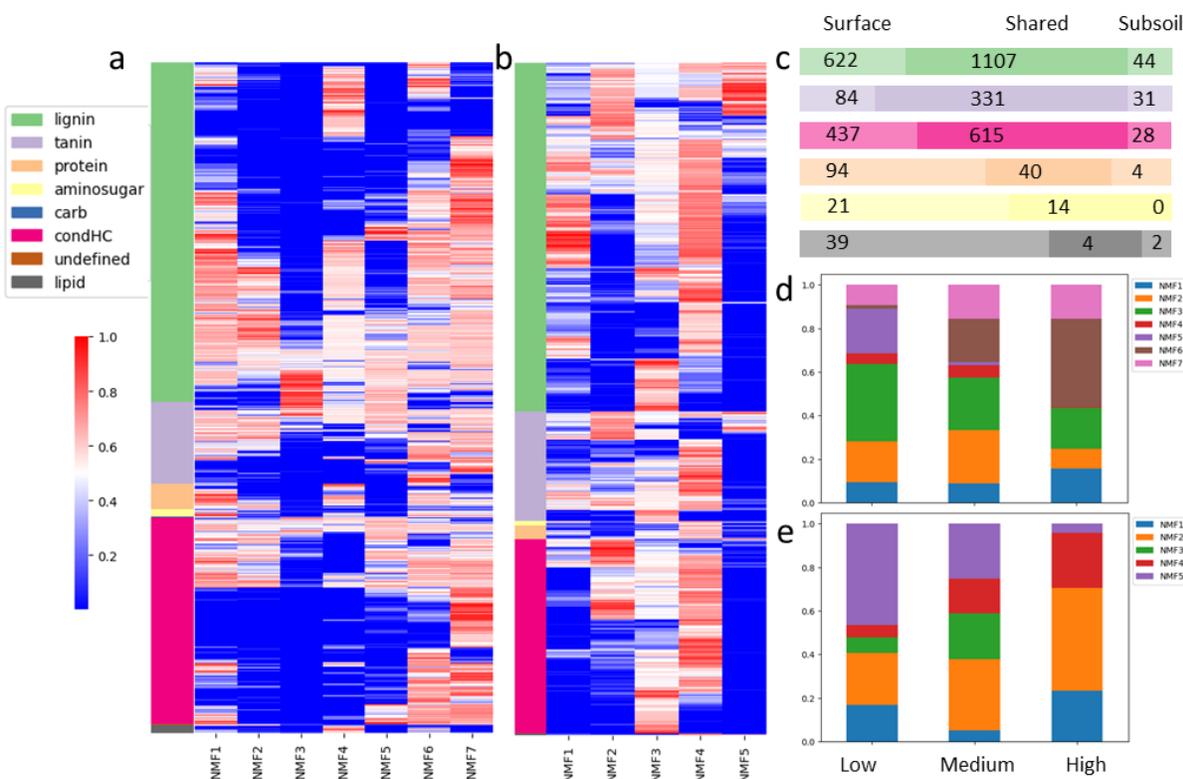


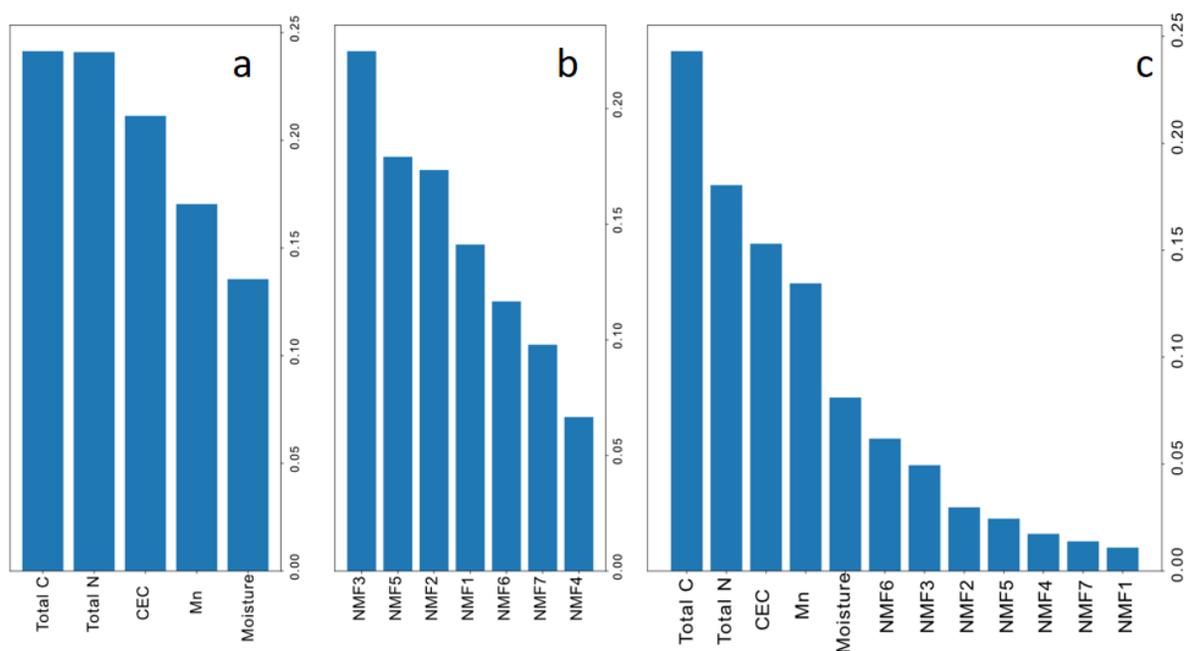
Figure 3. (a) Spatial distribution of soil respiration levels (labeled by colors) and alpha diversity of each sample (sizes). Soil respiration levels are determined by *k*-means clustering on soil respiration rates ($\mu\text{g CO}_2/\text{g soil/day}$). Soils from temperate conifer forests and temperate grasslands, savannas & shrublands have relatively higher respiration rates compared to other biomes (Figure S1). (b) The number of shared and unique SOM compound classes identified between surface and subsoils. The classes were suggested by van-Krevelen plot. (c) The difference of alpha diversity in surface and subsoil soils ($p < 0.05$ from ANOVA, *: $p < 0.05$ from Tukey's HSD test) (d) the difference of alpha diversity in surface soils with different levels of potential respiration ($p < 0.05$ from ANOVA, *: $p < 0.05$ from Tukey's HSD test) (e) the difference of alpha diversity in subsoils with different levels of potential respiration ($p < 0.05$ from ANOVA).

1

2
3

4 Figure 4. NMFk partitioning of SOM composition. (a-b) Relative contribution of organic formula
 5 to each SOM signatures identified by NMFk in a) surface and b) subsoils. The color in each cell
 6 represents the normalized (0 to 1) relative contribution for each SOM feature (row) to each
 7 NMFk signature (column), red indicates the most important contributor, and blue indicates the
 8 least. The side bar indicates the compound class of each SOM feature. (c) The number of shared
 9 and unique formula identified as important (normalized weights >0.5) by NMFk in surface and
 10 subsoils. (d-e) The relative contribution of NMFk signatures to each level of respiration rates in
 11 both d) surface and e) subsoils. Surface soils: low respiration level (N = 44), medium respiration
 12 level (N = 14), high respiration level (N = 5, UT12, UT23, UT24, WY03, Temperate Conifer
 13 Forests, SCBI Temperate Broadleaf & Mixed Forests). Subsoils: low respiration level (N = 48),
 14 medium respiration level (N = 10), high respiration level (N = 3, T12, UT19, Temperate Conifer
 15 Forests, WLOO, Temperate Broadleaf & Mixed Forests).

16



17
 18 Figure 5. Relative importance of each predictor in surface soil potential respiration machine
 19 learning models. a) Physicochemistry model, with physicochemical variables only. b) SOM
 20 model, with SOM signatures represented by NMFs only. c) Physicochemistry & SOM model
 21 with both physicochemical variables and SOM signatures.

22

23 Table 1. Model performance for predictions of potential soil respiration with physicochemical
 24 variables (Physiochemistry model), SOM by NMF k signatures (SOM_model), and combined
 25 physicochemical variables and SOM variables (Physiochemistry &SOM_model) for average 5-
 26 fold cross-validation accuracies (training soils, RMSE), and testing sample accuracies (RMSE,
 27 R2).

28

| | Physiochemistry Model | SOM_model | Physiochemistry &SOM_model |
|-------------------|--------------------------|-----------|-------------------------------|
| Surface_CV | 0.80 | 1.05 | 0.82 |
| Surface_test | 0.98 | 0.89 | 0.82 |
| Surface_test (R2) | 0.44 | 0.54 | 0.62 |
| Subsoil_CV | 0.60 | 0.82 | 0.67 |
| Subsoil_test | 0.46 | 0.80 | 0.49 |
| Subsoil_test (R2) | 0.43 | 0.08 | 0.36 |

29

30

31 **Supporting Information of**
32 **Soil Organic Matter Composition Improves Predictions of Potential Soil**
33 **Respiration across the Continental United States**

34
35 Cheng Shi^a, Maruti Mudunuru^b, Maggie Bowman^c, Qian Zhao^c, Jason Toyoda^c, Will Kew^c, Yuri Corilo^c, Odeta
36 Qafoku^c, John R. Bargar^c, Satish Karra^c, & Emily Graham^{d,e*}

37
38 ^aOregon State University, Department of Biological & Ecological Engineering, Corvallis, OR, United States.

39 ^bEnergy and Environment Directorate, Pacific Northwest National Laboratory, Richland, WA, United States.

40 ^cEnvironmental Molecular Science Laboratory, Pacific Northwest National Laboratory, Richland, WA, United
41 States.

42 ^dEarth and Biological Sciences Directorate, *Pacific Northwest National Laboratory, Richland, WA, United*
43 *States.*

44 ^e*School of Biological Sciences, Washington State University, Pullman, WA, United States.*

45

46 **Corresponding author: emily.graham@pnnl.gov*

47

48

49 *NMFk model assumption and robustness*

50 NMFk model was selected to decompose the SOM composition matrix into multiple basis signatures, due
51 to its ability to capture unique and sparse characteristics or data patterns ¹. The underlying assumption of NMFk is
52 that there are similar distributions of variables across samples such that the main characteristics of each sample can
53 be represented by the combination of a limited number of non-negative additive components (signatures) ². It has
54 also been widely used in environmental forensics ^{3,4}, text mining ⁵, face recognition ⁶. Vesselinov et al. used NMFk
55 to identify unknown recharge sources of groundwater driven by various physical and chemical processes ⁷. Cai et al.
56 used NMF to extract key features and reveal temporal changes in microbial communities ⁸. Instead of linear
57 transformation of the original dataset by correlations like principal components analysis (PCA), NMFk uses non-
58 negativity constraints that makes it better suited to identify representative SOM signatures and evaluate their
59 distribution in different samples. Furthermore, the additive fashion of extracted signatures by different weights in
60 NMFk fit the intuition of different pools of SOM molecules combined into the mixture of SOM in a certain sample.
61 Therefore, the NMFk extracted SOM signatures are more explainable compared to PCA or other ordination
62 techniques.

63 The number of dominant types (k) was determined by silhouette coefficient with a threshold of 0.5 to test
64 model stability ^{9,10}. The last model above the threshold (> 0.5) is selected as the final model. This is because the
65 selected model should have good separation between different non-negative signatures but also a stable solution at
66 the same time.

68 *Gradient Boosting regression models*

69 Gradient boosting is a machine learning algorithm that combines multiple weak models, such as decision
70 trees, into a stronger model iteratively, where each weak model learns from the residual error from the previous
71 model. ¹¹ It is one of the most powerful and effective machine learning models that is widely used in many different
72 areas. Gradient boosting regression is an ensemble model that iteratively learns from the error of previous model.
73 Using ensemble, it is capable to generate predictions from multiple decision tree models and thus provide a more
74 robust prediction. It usually has better performance with smaller dataset, because it tends less overfit the data ¹².
75 Therefore, it is suitable for predicting soil respiration with physicochemistry and SOM types.

76 We performed feature selection for physicochemical factors by statistical relevance (Table S1), to remove
77 irrelevant features that likely introduce noise and leads to overfitting of the model. ^{13,14} Total C, total N, CEC, Mn
78 and soil moisture were selected as predictors for surface soil models. Total C, total N, total base, CEC, Mg and soil
79 moisture were selected for subsoil models. The detailed settings of hyperparameter dictionary for
80 *RandomizedSearchCV* function and the tuned parameter set used for the final model is in Table S2. To avoid the
81 impacts of the increased number of predictors on improved model performance for surface respiration model
82 (physicochemistry model: n = 5, SOM model: n = 7), we developed another version of SOM model without the two
83 least important predictors (NMF7, NMF4). The model performance was still better (testing R² = 0.48 vs. 0.44)
84 compared to the physicochemistry model with the same number of predictors (n = 5).

85

86

87

88 *Supporting Tables*

89 Table S1. Coefficient of Determination between soil respiration and soil biogeochemistry

90 (Pearson's correlation R-square)

| | Surface R ² | Surface p-value | Subsoil R ² | Subsoil p-value |
|------------|---------------------------|--------------------|---------------------------|--------------------|
| Mn | 0.324 | 0.000 | 0.142 | 0.003 |
| Mg | 0.160 | 0.001 | 0.287 | 0.000 |
| K | 0.004 | 0.638 | 0.053 | 0.071 |
| Na | 0.005 | 0.577 | 0.026 | 0.211 |
| B | 0.119 | 0.006 | 0.018 | 0.295 |
| Zn | 0.173 | 0.001 | 0.102 | 0.011 |
| Fe | 0.089 | 0.017 | 0.043 | 0.106 |
| Cu | 0.092 | 0.016 | 0.133 | 0.004 |
| Total Base | 0.146 | 0.002 | 0.227 | 0.000 |
| CEC | 0.405 | 0.000 | 0.354 | 0.000 |
| Total C | 0.487 | 0.000 | 0.268 | 0.000 |
| Total N | 0.439 | 0.000 | 0.248 | 0.000 |
| Total S | 0.080 | 0.028 | 0.036 | 0.160 |
| GWC | 0.246 | 0.000 | 0.225 | 0.000 |
| Soil T | 0.007 | 0.545 | 0.000 | 0.919 |
| pH | 0.116 | 0.004 | 0.007 | 0.513 |
| SO4 | 0.172 | 0.001 | 0.002 | 0.759 |
| P | 0.001 | 0.855 | 0.003 | 0.695 |
| NH4 | 0.002 | 0.761 | 0.000 | 0.992 |
| NO3 | 0.004 | 0.634 | 0.004 | 0.634 |

| | | | | |
|-----------|--------------|--------------|--------------|--------------|
| Sand% | 0.140 | 0.001 | 0.176 | 0.000 |
| Silt% | 0.081 | 0.017 | 0.077 | 0.022 |
| Clay% | 0.157 | 0.001 | 0.182 | 0.000 |
| Elevation | 0.136 | 0.006 | 0.090 | 0.029 |
| alpha_div | 0.159 | 0.001 | 0.143 | 0.003 |

91
92
93

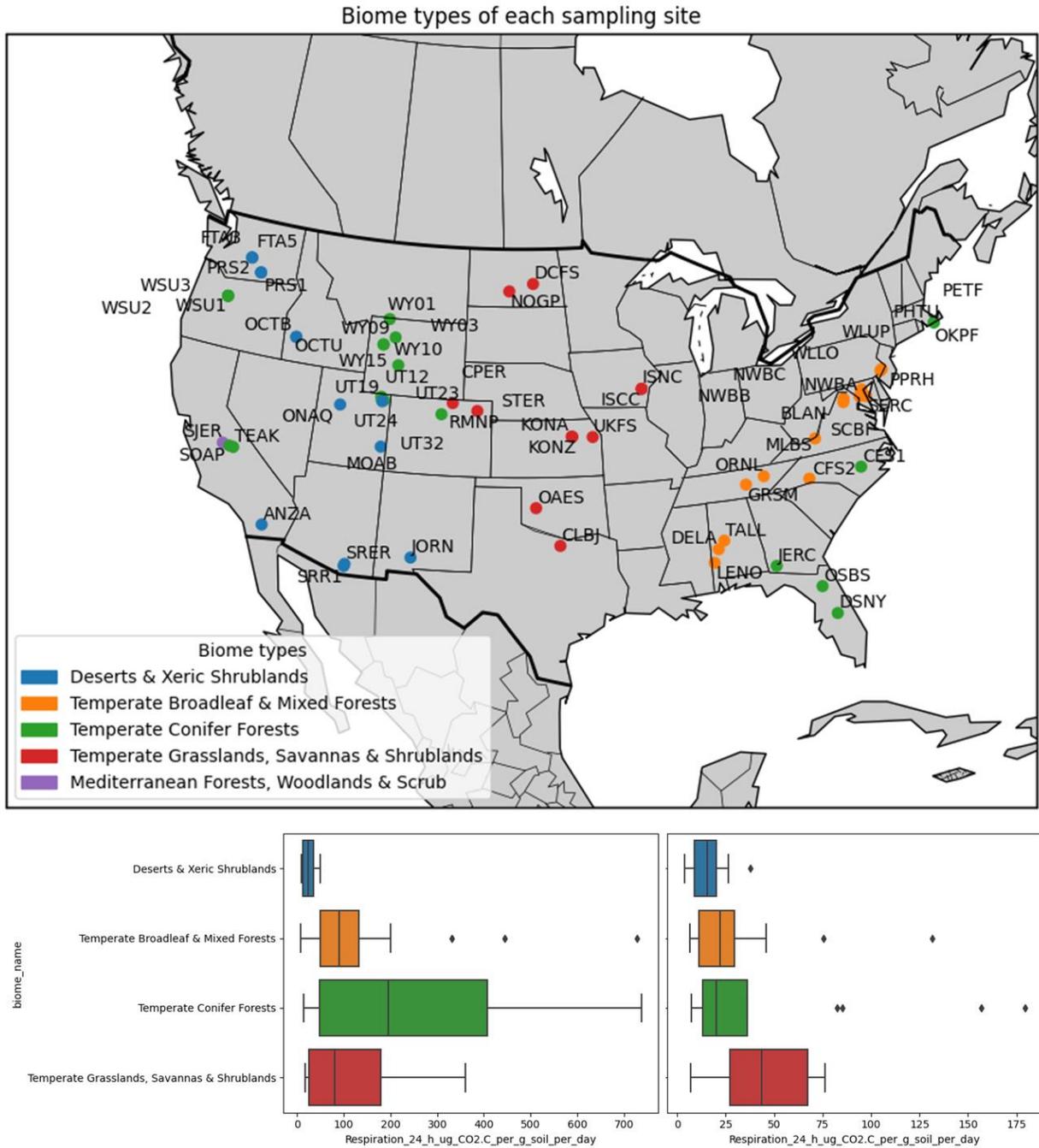
94 Table S2. Hyperparameter tuning settings and the tuned hyperparameters used in each model.

| Hyperparameter name | param_distributions | Physicochemistry Model | | SOM Model | | Physicochemistry & SOM Model | |
|---------------------|---------------------------------|------------------------|------------------|------------------|-----------------|------------------------------|-----------------|
| | | Surface | Subsoil | Surface | subsoil | surface | subsoil |
| n_estimators | randint(50,5000) | 1213 | 1722 | 422 | 636 | 1392 | 351 |
| max_depth | randint(2,60) | 31 | 58 | 14 | 7 | 40 | 16 |
| max_features | randint(1, X.shape[1]) | 1 | 6 | 2 | 5 | 3 | 7 |
| min_samples_split | randint(2, 10) | 6 | 6 | 4 | 6 | 7 | 9 |
| learning_rate | [0.0001, 0.001, 0.01, 0.1, 1.0] | 0.01 | 0.01 | 0.1 | 0.001 | 0.1 | 0.1 |
| ccp_alpha | expon(scale=0.1) | 0.000941 9401 | 0.017319 5734 | 0.043552 4849 | 0.00177 8767 | 1.867313 65e-05 | 0.00065 9532 |

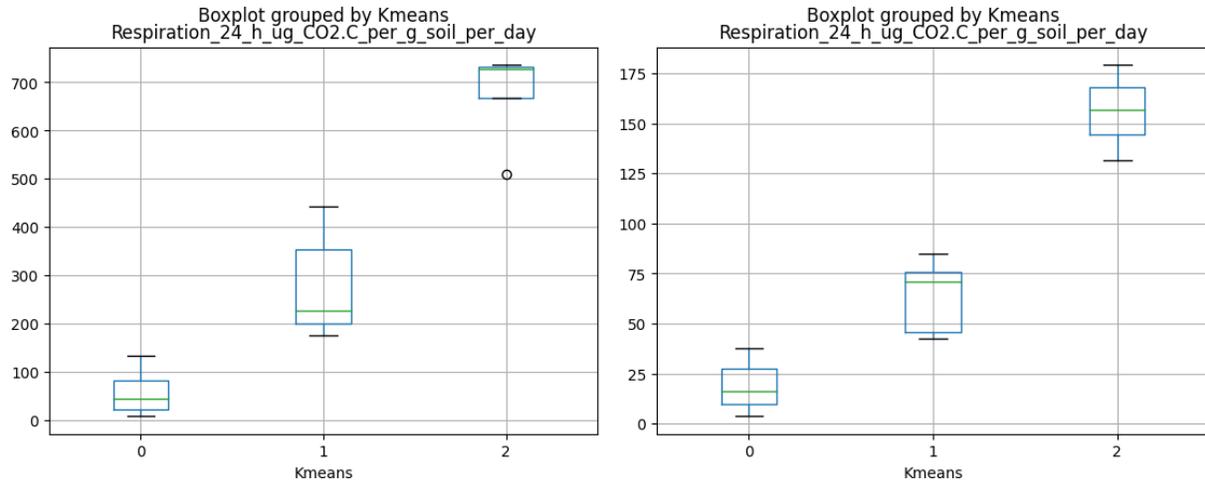
95

96

97 *Supporting Figures*



98
 99 Figure S1. Sampling locations, sample names, and their biome types obtained from WWF
 100 terrestrial ecoregions (a). Difference of soil potential respiration by biomes in b) surface and c)
 101 subsoil.



102

103 Figure S2. k-means clustering of soil respiration rates at different depths (a: surface soils, b:

104 subsoils). 3 levels of respiration were determined for both surface and subsoil.

105

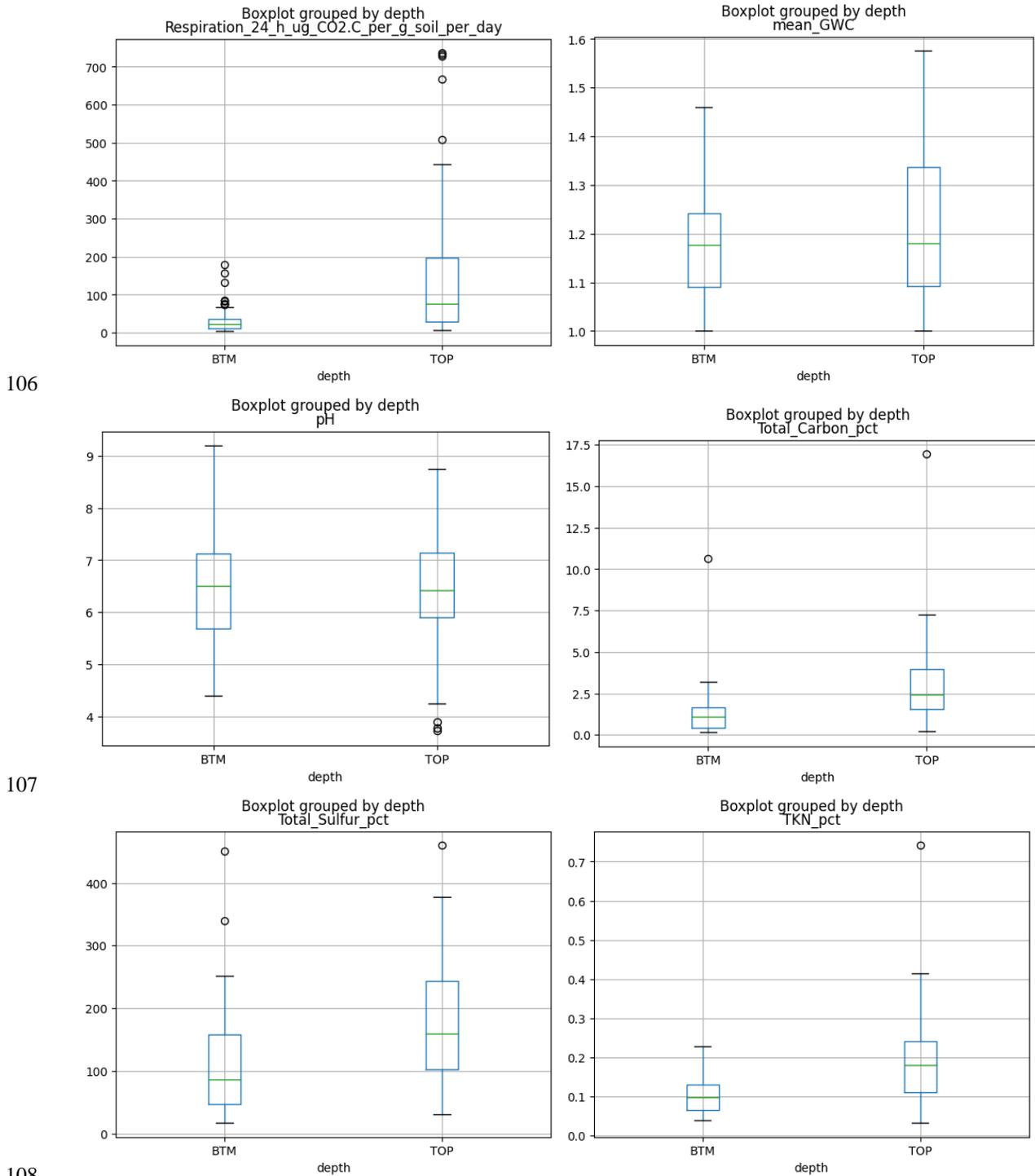
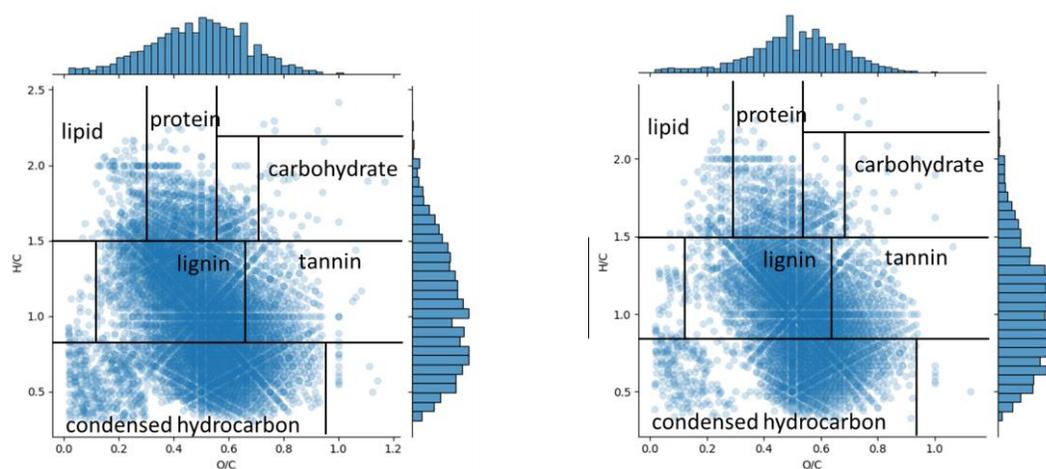
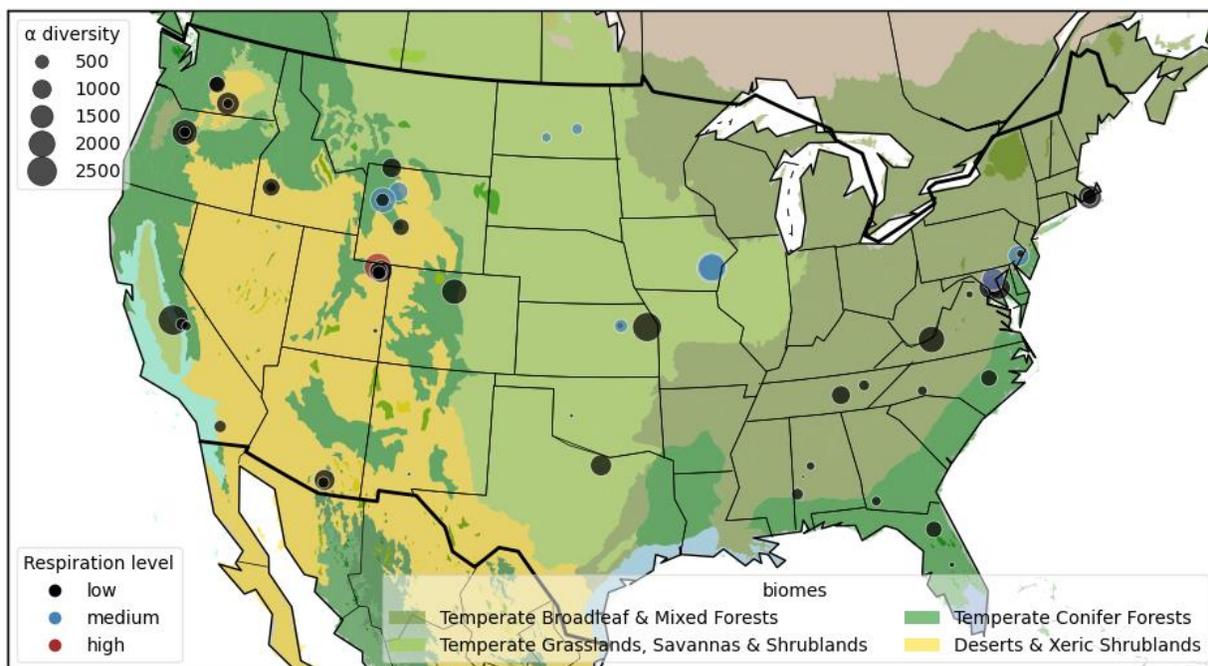


Figure S3. Boxplots of difference in soil biogeochemistry between surface and subsoils. a) potential respiration, b) moisture content, c) pH, d) total C, e) total S, f) total N.

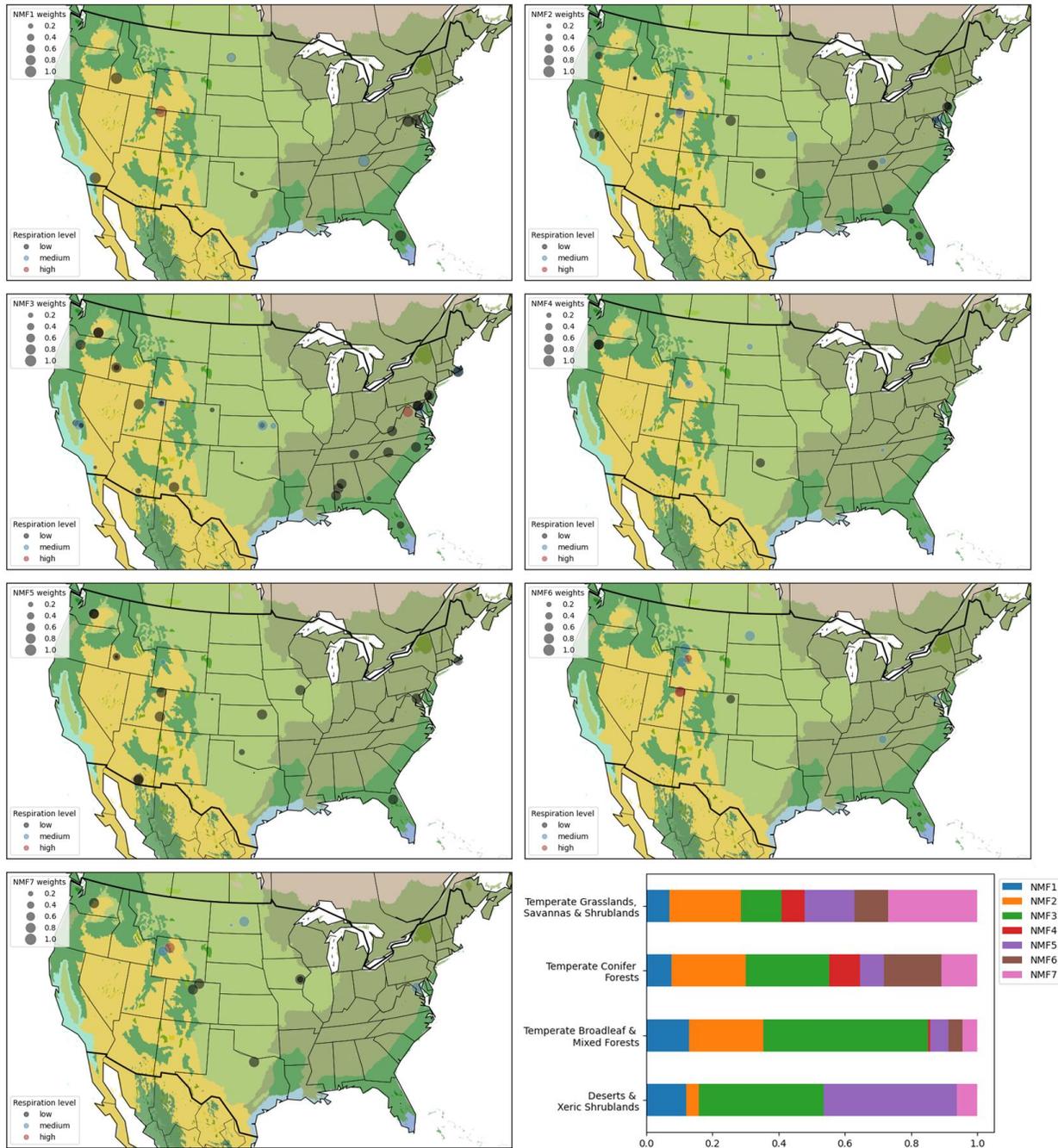


113
114 Figure S4 Van Krevelen Diagram of SOM formula identified in a) surface b) subsoils.

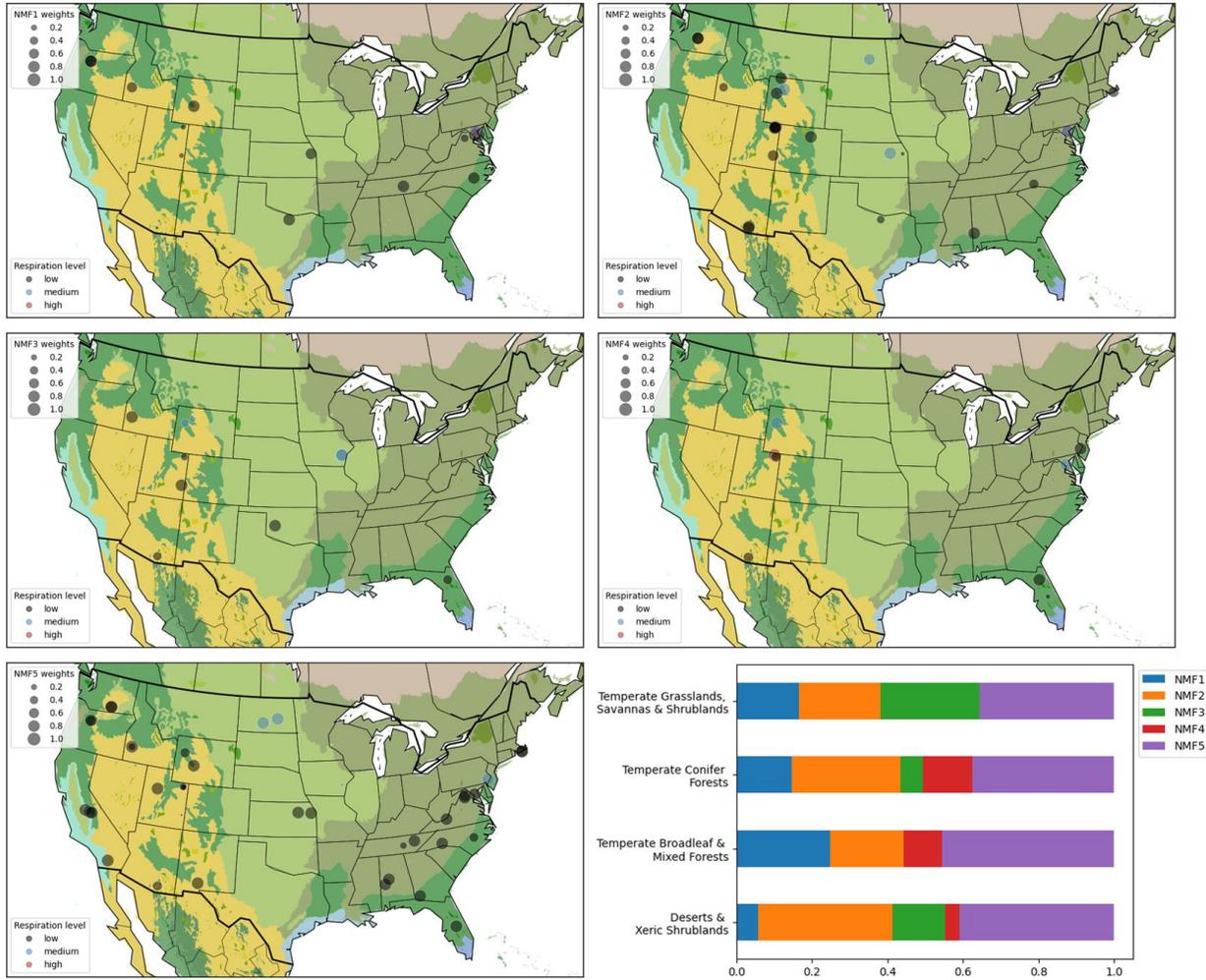
115
116



117
118 Figure S5. Spatial distribution of subsoil respiration levels (labeled by colors) and alpha diversity
119 of each sample (sizes). Soil respiration levels are determined by K-means clustering on soil
120 respiration rates ($\mu\text{g CO}_2/\text{g soil/day}$)

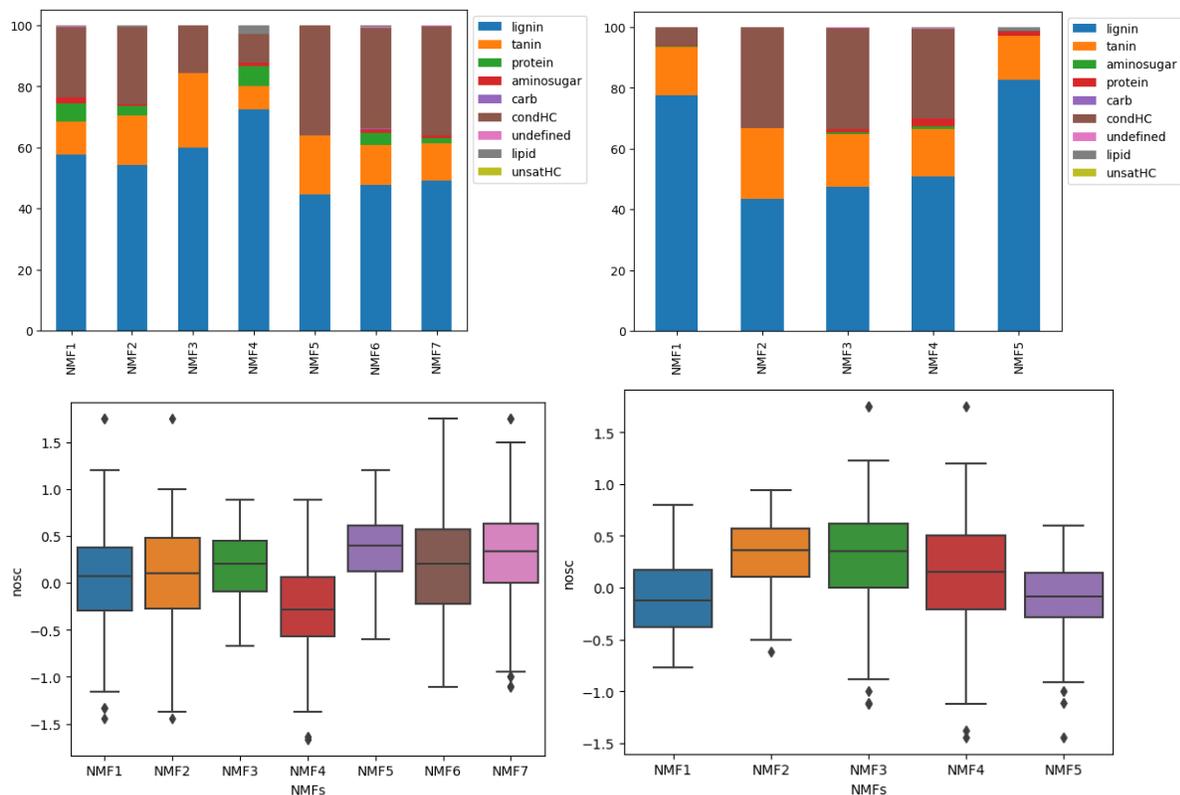


121
 122 Figure S6. The weights of 7 surface soil SOM types in all samples identified by NMF k using
 123 SOM composition data obtained from FT-ICR MS, and the relative contribution of the 7 types in
 124 each biome. Deserts & Xeric Shrublands (N = 13), Temperate Broadleaf & Mixed Forests (N =
 125 17), Temperate Conifer Forests (N = 21), Temperate Grasslands, Savannas & Shrublands (N =
 126 11).



127
 128 Figure S7. The weights of 5 subsoil SOM types in all samples identified by NMFk using SOM
 129 composition data obtained from FT-ICR MS, and the relative contribution of the 5 types in each
 130 biome. Deserts & Xeric Shrublands (N = 13), Temperate Broadleaf & Mixed Forests (N = 17),
 131 Temperate Conifer Forests (N = 21), Temperate Grasslands, Savannas & Shrublands (N = 9).

132
 133
 134

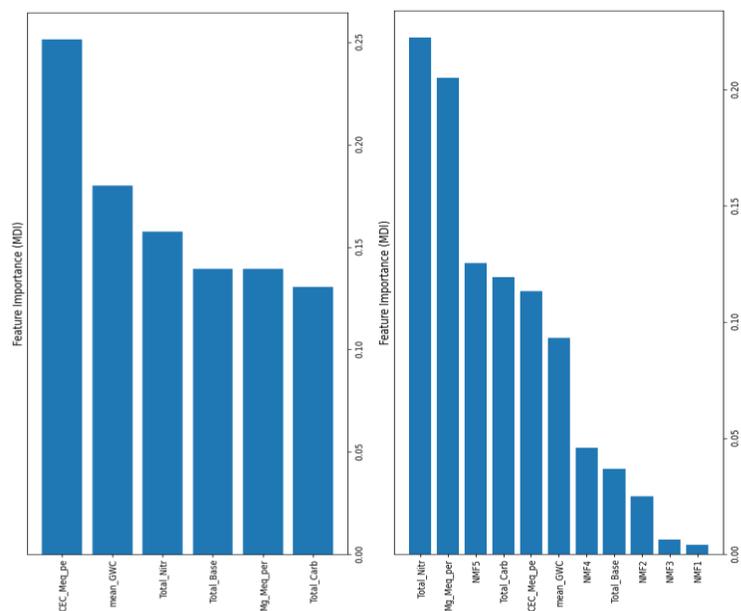


135

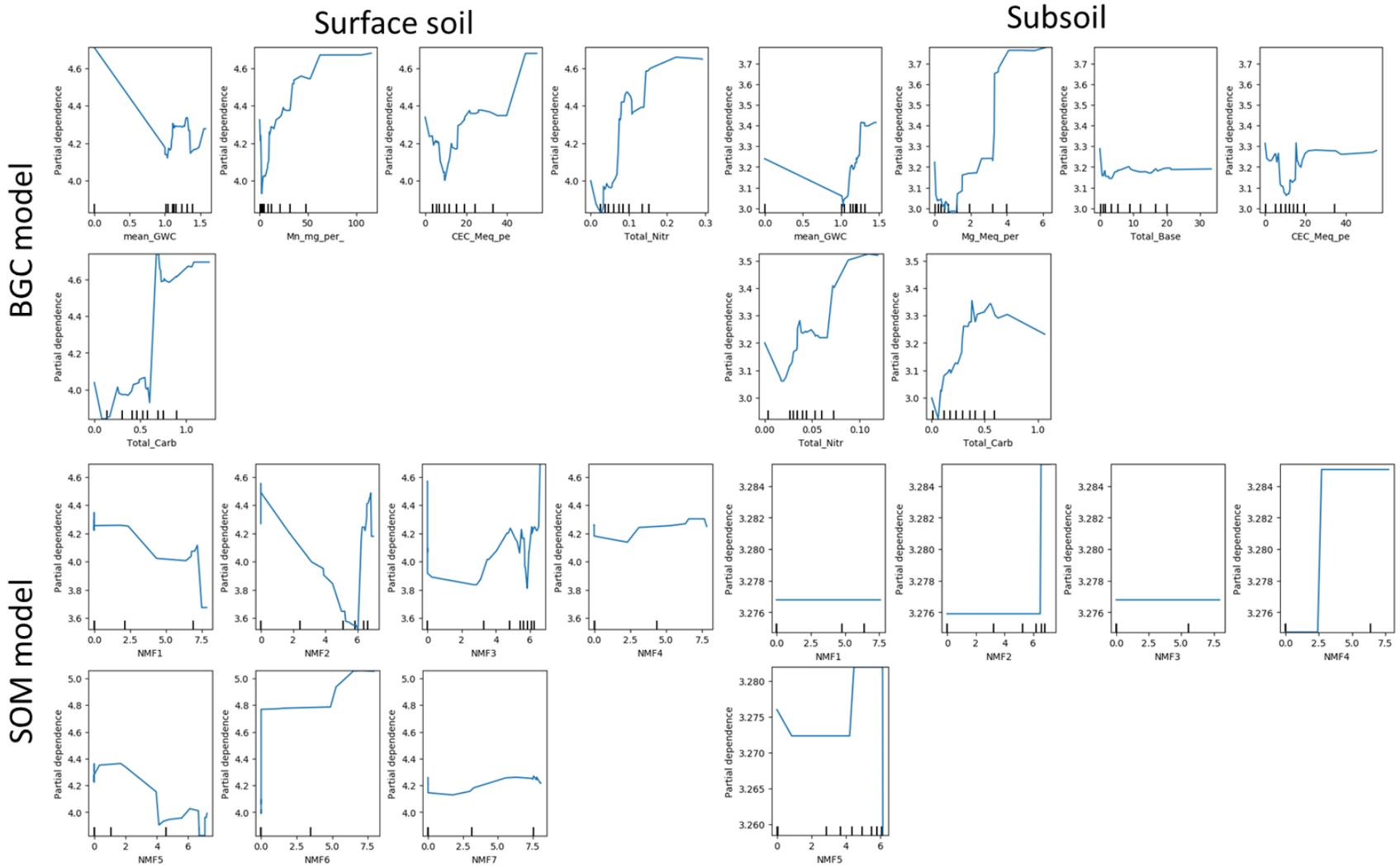
136

137 Figure S8. Relative contribution of each compound class to each NMF type for important
 138 features with normalized weights of greater than 0.5 in a) surface soil and b) subsoil. Boxplot
 139 shows the difference of Nominal Oxidation State of Carbon (NOSC) Values for the important
 140 compounds ($w > 0.5$) for each NMF in c) surface soil and d) subsoil.

141



142
 143 Figure S9. Relative importance of each predictor in subsoil potential respiration models. a) Physicochemical model,
 144 with biogeochemical variables only. b) Physicochemistry & SOM_model with both physicochemical variables and
 145 SOM types. (SOM model for subsoil has bad performance (Table 1) and therefore feature importance is not reported
 146 here).
 147



BGC&SOM model

Surface soil

Subsoil

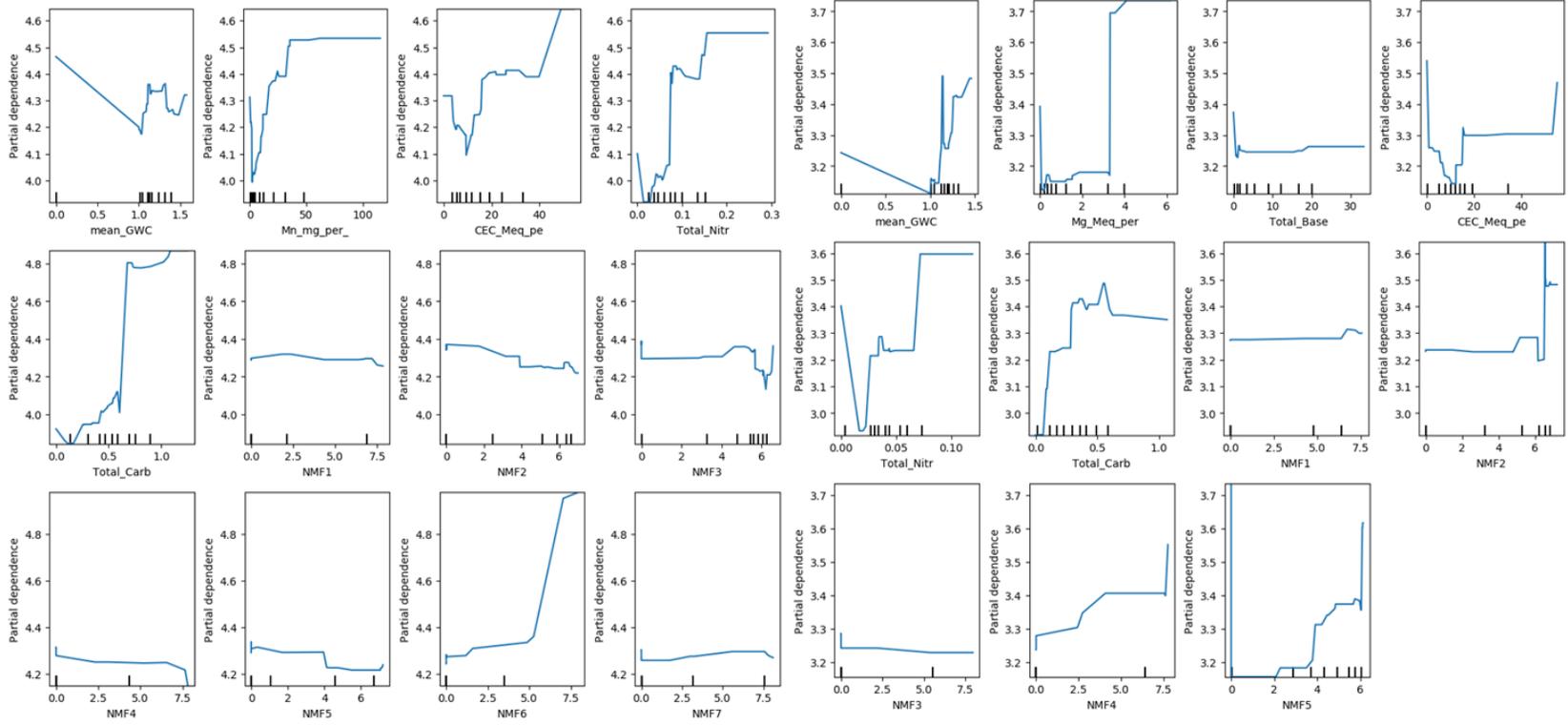


Figure S10. Partial dependence of potential respiration to predictors of soil biogeochemistry and/or SOM composition in surface and subsoil models. a) BGC model with biogeochemical variables for surface soil, b) BGC model with biogeochemical variables for subsoil, c) SOM model with SOM variables for surface soil, d) SOM model with SOM variables for subsoil (bad model performance), e) BGC&SOM model with both biogeochemical and SOM variables for surface soil, f) BGC&SOM model with both biogeochemical and SOM variables for subsoil.

References

1. Lee, D.; Seung, H. S., Algorithms for non-negative matrix factorization. *Advances in neural information processing systems* **2000**, *13*.
2. Paatero, P.; Tapper, U., Positive matrix factorization: A non-negative factor model with optimal utilization of error estimates of data values. *Environmetrics* **1994**, *5*, (2), 111-126.
3. Johnson, G. W.; Ehrlich, R.; Full, W.; Ramos, S., Principal components analysis and receptor models in environmental forensics. In *Introduction to environmental forensics*, Elsevier: 2015; pp 609-653.
4. Rodenburg, L. A.; Du, S.; Xiao, B.; Fennell, D. E., Source apportionment of polychlorinated biphenyls in the New York/New Jersey Harbor. *Chemosphere* **2011**, *83*, (6), 792-798.
5. Pauca, V. P.; Shahnaz, F.; Berry, M. W.; Plemmons, R. J. In *Text mining using non-negative matrix factorizations*, Proceedings of the 2004 SIAM international conference on data mining, 2004; SIAM: 2004; pp 452-456.
6. Guillaumet, D.; Vitria, J. In *Non-negative matrix factorization for face recognition*, Catalanian Conference on Artificial Intelligence, 2002; Springer: 2002; pp 336-344.
7. Vesselinov, V. V.; Alexandrov, B. S.; O'Malley, D., Contaminant source identification using semi-supervised machine learning. *Journal of Contaminant Hydrology* **2018**, *212*, 134-142.
8. Cai, Y.; Gu, H.; Kenney, T., Learning Microbial Community Structures with Supervised and Unsupervised Non-negative Matrix Factorization. *Microbiome* **2017**, *5*, (1), 110.
9. Vangara, R.; Bhattarai, M.; Skau, E.; Chennupati, G.; Djidjev, H.; Tierney, T.; Smith, J. P.; Stanev, V. G.; Alexandrov, B. S., Finding the Number of Latent Topics With Semantic Non-Negative Matrix Factorization. *IEEE Access* **2021**, *9*, 117217-117231.
10. Bhattarai, M.; Chennupati, G.; Skau, E.; Vangara, R.; Djidjev, H.; Alexandrov, B. S. In *Distributed Non-Negative Tensor Train Decomposition*, 2020 IEEE High Performance Extreme Computing Conference (HPEC), 22-24 Sept. 2020, 2020; 2020; pp 1-10.
11. Friedman, J. H., Greedy function approximation: a gradient boosting machine. *Annals of statistics* **2001**, 1189-1232.
12. Hastie, T.; Tibshirani, R.; Friedman, J. H.; Friedman, J. H., *The elements of statistical learning: data mining, inference, and prediction*. Springer: 2009; Vol. 2.
13. Christ, M.; Braun, N.; Neuffer, J.; Kempa-Liehr, A. W., Time Series Feature Extraction on basis of Scalable Hypothesis tests (tsfresh – A Python package). *Neurocomputing* **2018**, *307*, 72-77.
14. Yuan, B.; Tan, Y. J.; Mudunuru, M. K.; Marcillo, O. E.; Delorey, A. A.; Roberts, P. M.; Webster, J. D.; Gammans, C. N. L.; Karra, S.; Guthrie, G. D.; Johnson, P. A., Using Machine Learning to Discern Eruption in Noisy Environments: A Case Study Using CO₂-Driven Cold-Water Geyser in Chimayó, New Mexico. *Seismological Research Letters* **2019**, *90*, (2A), 591-603.