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Wastewaters co-produced with shale gas drive slight regional salinization of groundwater

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28 **Abstract**

29 While unconventional oil and gas (UOG) development is changing the world economy,  
30 processes that are used during UOG development such as high-volume hydraulic fracturing  
31 (“fracking”) have been linked with water contamination. Water quality risks include leaks of gas  
32 and salty fluids (brines) that are co-produced at wellpads. Identifying the cause of contamination  
33 is difficult, however, because UOG wells are often co-located with other contaminant sources.  
34 We investigated the world’s largest shale gas play with publicly accessible groundwater data  
35 (Marcellus Shale in Pennsylvania, U.S.A. with ~29,000 analyses) and discovered that  
36 concentrations of brine-associated barium ([Ba]) and strontium ([Sr]) show small regional  
37 increases within 1 km of UOG development. Higher concentrations in groundwaters are  
38 associated with greater proximity to and density of UOG wells. Concentration increases are even  
39 larger when considering associations with the locations of i) spill-related violations and ii) some  
40 wastewater impoundments. These statistically significant relationships persist even after  
41 correcting for other natural and anthropogenic sources of salts. The most likely explanation is  
42 that UOG development slightly increases salt concentrations in regional groundwaters not  
43 because of fracking but because of the ubiquity of wastewater management issues. These results  
44 emphasize the need for stringent wastewater management practices across oil and gas operations.

45

46 **Introduction**

47 Unconventional oil and gas (UOG) development has advanced United States (U.S.)  
48 energy independence but incited concerns surrounding potential environmental and human health  
49 impacts. UOG development involves horizontal drilling and high-volume high-pressure  
50 hydraulic fracturing to extract hydrocarbons from unconventional formations such as shales and

51 other rocks with low permeability. UOG development in one of the world's largest shale gas  
52 plays, the Marcellus Shale, produces ~30x more gas and ~10x more wastewater per well  
53 compared to drilling in conventional reservoirs.<sup>1</sup> While increased gas production from the  
54 Marcellus Shale has reduced emissions of CO<sub>2</sub> and some pollutants as power generation has  
55 shifted from coal to gas, the 570% increase in wastewater co-produced with natural gas  
56 accentuates the need for proper handling, recycling, and disposal of produced materials to avoid  
57 environmental impacts.<sup>1-4</sup> Analyses of publicly available data from regulatory agencies show  
58 that incidents such as well construction impairments or wastewater spills are reported at >2% of  
59 all UOG wells, creating potential for environmental degradation.<sup>5-7</sup> However, the extent to which  
60 issues such as compromised well integrity or improper waste handling translate to water quality  
61 impacts remains poorly understood.

62         Research into the impacts of UOG development on groundwater quality has extensively  
63 focused on methane, the primary constituent of natural gas and the most commonly cited  
64 contaminant during UOG development.<sup>8-14</sup> However, another commonly reported pollutant  
65 released during UOG development is wastewater, which can be spilled into soils or streams  
66 because of issues related to recovery, storage, or transportation.<sup>5,6,15</sup> These wastewaters can  
67 contain a variety of contaminants. In the first weeks following hydraulic fracturing, waters that  
68 are co-produced with the gas (produced waters) are termed flowback waters. Flowback is  
69 comprised largely of fluids injected during the hydraulic fracturing of the well.<sup>2</sup> During the  
70 production lifetime of the well, in contrast, the produced water that returns with gas derives  
71 largely from so-called formation waters, i.e., waters in the shale formation itself that are  
72 geochemically identical to basin brines.<sup>16,17</sup> Formation brines typically comprise 92-96% of the  
73 wastewater generated over a UOG well's production lifetime, and are generally sodium (Na)-

74 calcium (Ca)-chloride (Cl) brines with salinities up to 7x modern ocean water.<sup>18,19</sup> They also  
75 typically contain less common species such as barium, strontium, and bromide whose  
76 concentrations ([Ba], [Sr], [Br], respectively) can be used to fingerprint contamination related to  
77 produced waters.<sup>18,20</sup> The highly concentrated nature of many UOG wastewaters creates the  
78 potential for their salts, metals, organic species, and naturally-occurring radioactive materials to  
79 degrade water resources.<sup>21,22</sup>

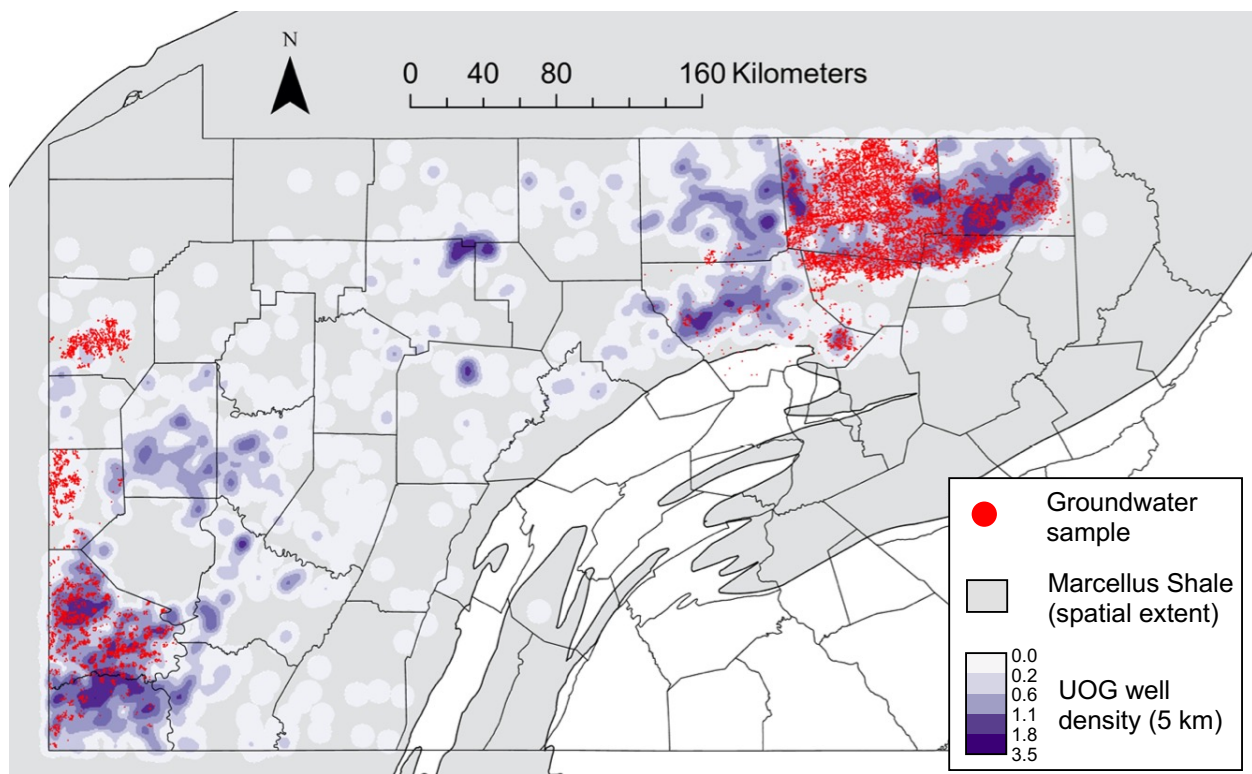
80 Many stakeholders including scientists, engineers, regulators, operators, and the public  
81 are interested in both why contamination occurs and how frequently it occurs during UOG  
82 development. The former question generally requires time-, money-, and fieldwork-intensive  
83 case studies in locations generally only accessed by landowners, regulators, or industry  
84 practitioners.<sup>10,13</sup> Determining the frequency of incidents typically requires statistical analyses of  
85 large regulatory and geochemical datasets.<sup>6,23</sup> Such analyses applied to the salinization of surface  
86 waters during UOG development shows that regional salt concentrations may be increasing very  
87 slightly in streams near UOG development and that local increases in [Ba], [Sr] and [Cl] in  
88 streams impacted by UOG development wastewater leaks or spills can persist for years.<sup>24-26</sup> A  
89 recent nationwide analysis of stream chemistry reported a significant increase in brine salt (Ba,  
90 Sr, Cl) concentrations in watersheds with higher UOG development density (i.e., the number of  
91 UOG wells).<sup>27</sup> A regional analysis in southwestern Pennsylvania also documented a significant  
92 increase in [Ba], [Sr], and [Cl] in groundwaters that correlate with the proximity and density of  
93 UOG development.<sup>28</sup> This regional increase was attributed to localized incidents or “hotspots”  
94 where brines had escaped into groundwater.<sup>28</sup>

95 Despite these studies, the actual causes of regional UOG impacts on water resources are  
96 difficult to identify because UOG development is broadly distributed across hydrocarbon basins

97 and includes many processes that could cause contamination ranging from drilling to “fracking”  
98 to waste disposal. Additionally, water quality prior to UOG development is not well-  
99 characterized in many basins, and UOG development often overlaps with road salting and  
100 longstanding forms of hydrocarbon extraction such as conventional oil and gas (COG)  
101 development or coal mining that have also been associated with groundwater impacts.<sup>28–30</sup> Many  
102 of the species most often associated with UOG development contamination, such as methane and  
103 brine salts, are also naturally present in groundwater and can be released by COG development  
104 as well.<sup>31–34</sup> Nevertheless, determining the extent to which UOG development may impact water  
105 supplies is important because in most locations of such development, local populations rely on  
106 domestic wells for drinking water.<sup>35,36</sup> Emerging studies that link proximity to UOG  
107 development to negative effects on human health have led to research into whether water  
108 supplies are an exposure pathway.<sup>37–40</sup>

109 In this study, we examined the concentrations of brine salt ions in groundwater to  
110 determine if they are impacted by specific processes during UOG development (e.g., well  
111 construction, wastewater management). Of the major shale plays under development worldwide,  
112 we are aware of only three U.S. states where the quantities and density of groundwater quality  
113 data readily available to the public are suitable for regional-scale analyses (Texas, Colorado,  
114 Pennsylvania).<sup>41</sup> To investigate the potential for groundwater impact, we therefore chose the  
115 state with the largest publicly accessible water quality database, Pennsylvania.<sup>42</sup> We emphasize  
116 UOG wells instead of COG wells because publicly available data suggest that UOG wells are  
117 responsible for 97.4% of the wastewater produced by oil and gas wells since the implementation  
118 of UOG development in our study area (Text S1).

119 Pennsylvania (PA) is also a good test case because of the size of the gas play as well as  
120 the observation that spill rates in PA are generally comparable to other major gas-producing  
121 states.<sup>6,15</sup> In addition, much of the information about such incidents is publicly accessible for  
122 PA,<sup>6</sup> enabling a large-scale investigation of impacts on groundwater with an objective of  
123 elucidating relevant processes in many other major shale plays where such an investigation is not  
124 feasible.<sup>42</sup> The two most heavily drilled parts of this region are northeastern and southwestern  
125 PA (northeastern PA and southwestern PA, respectively, Figure 1). Northeastern PA is  
126 characterized by greater topographic relief but far more limited legacy hydrocarbon extraction  
127 (coal mining, conventional oil and gas) compared to southwestern PA.<sup>28,43</sup>



128  
129 Figure 1: Locations of the 28,609 sampled groundwaters indicated on a map showing the average  
130 density of UOG wells within a 5 km radius in Pennsylvania (calculated as the 5 km kernel  
131 density using 500 m bins). For closeups of western PA and northeastern PA, see Figures S2 and  
132 S3 and for locations of UOG wells, COG wells, and coal mining see Figure S4.

133 **Materials and Methods**

134 Our dataset consists of 28,609 groundwater analyses from the Shale Network database  
135 (available at <https://doi.org/10.26208/DT5Y-5B37> and [https://doi.org/10.4211/his-data-](https://doi.org/10.4211/his-data-shalenetwork)  
136 [shalenetwork](https://doi.org/10.4211/his-data-shalenetwork)), spanning the Marcellus Shale region of PA (Figure 1).<sup>44</sup> These samples were  
137 predominantly collected between April 2008 and April 2020, with the majority of samples  
138 collected pre-2014 (for more about the dataset see Figure S1, Text S2). We examined  
139 relationships among groundwater chemistry and the locations of UOG wells, UOG  
140 impoundments, and UOG-related violations documented by the state regulator, the Pennsylvania  
141 Department of Environmental Protection (PADEP). This is the most complete database of  
142 incidents during UOG development that we are aware of for the study area, although additional  
143 incidents not cited in violations could potentially occur. In addition to regulatory data, we  
144 obtained the locations of impoundments associated with UOG development identified from 2010  
145 satellite imagery by Skytruth.<sup>45</sup> The locations of these impoundments were determined by  
146 Skytruth from USDA aerial survey photography following outlined methods and QAQC  
147 protocols.<sup>46</sup> Impoundments are often used to store fresh water for UOG wellpads in PA, but prior  
148 to 2016 the storage of UOG wastewaters in impoundments was less strictly regulated and led to  
149 putative issues with wastewater leakage.<sup>5,47</sup>

150 We analyzed 3 metrics to understand relationships between groundwater samples and  
151 UOG activities: land usage (i.e., whether UOG activities were occurring within a specific radius  
152 of each sample), distance (i.e., the distance between the sample and nearest UOG activity), and  
153 density (i.e., the number of UOG activities within a specific radius of each sample). Each  
154 calculation only considered UOG activities which occurred before a respective water sample was  
155 collected (Text S3). We examined land usage and UOG development density within a buffer

156 radius around sample sites of both 1 km and 3 km. We emphasize the smaller radius in the main  
157 text because 1 km is in best agreement with physics-based models analyzing the distance that  
158 groundwater may travel from UOG wellpads to domestic wells in the study area.<sup>48</sup> However,  
159 case studies have demonstrated fracture-mediated migration of UOG contaminants up to 3 km  
160 from a wellpad, and thus we conducted tests with a larger radius that are summarized in the  
161 SI.<sup>5,10,13</sup>

162 We focused intensively on two cationic species, barium (Ba) and strontium (Sr), both of  
163 which are widely analyzed and are present at characteristically high concentrations in  
164 Appalachian Basin brines.<sup>17</sup> For example, median [Ba] and [Sr] in produced waters from the  
165 Marcellus Shale (1125 and 1380 mg/L, respectively) are over 3 orders of magnitude greater than  
166 those reported for shallow groundwater in the region.<sup>23,49</sup> Ba is derived from rock dissolution but  
167 is found in generally low concentrations in uncontaminated surface and groundwaters in PA  
168 relative to oil and gas inputs.<sup>50</sup> While also derived from dissolution of the carbonate rocks that  
169 are common in hydrocarbon basins,<sup>51</sup> Sr can also serve as an identifier for UOG wastewater  
170 contamination.<sup>24,25</sup> While neither species can be considered truly conservative (i.e., they may  
171 adsorb or react as they migrate through an aquifer), both Ba and Sr have previously been  
172 identified as an effective tracer for wastewater leakage during oil and gas development in the  
173 Marcellus and nationwide.<sup>27,50</sup>

174 Before selecting Ba and Sr as foci, and to exclude species that are greatly influenced by  
175 overlapping sources such as coal mining or road salting, we examined how median  
176 concentrations of Ba, Ca, Cl, Na, Sr, and sulfate (SO<sub>4</sub>) varied across different hydrocarbon-  
177 related land uses. This comparison suggests that Ba, Sr, and Cl are perhaps the best tracers for  
178 UOG impacts, and supports more widespread impacts of UOG vs. COG wells (Text S4, Table



179 S1). Of those three analytes, we emphasized Ba and Sr on the basis that both are widely analyzed  
180 ( $n = 25,878$  and  $17,649$ , respectively) and are generally detected above reporting limits ( $24,917$   
181 and  $16,463$ , respectively) in our dataset. In contrast, Cl is more abundant in road salt, which  
182 commonly impacts groundwater in PA, and Cl is more frequently censored in our dataset (i.e.,  
183 present below reporting limits).<sup>52,53</sup> In particular, Cl is only reported to be above reporting limits  
184 in  $21,584$  out of  $27,599$  analyses. As a check, we used specialized methods for highly censored  
185 data (Text S5) to validate our key conclusions using Cl (Text S6).

186 We assessed relationships between Ba and Sr and UOG wells both by comparing median  
187 concentrations in samples within the buffer radius to concentrations in samples outside the  
188 buffer, as well as with regression modeling comparing ion concentrations to the proximity and  
189 density of UOG wells. Given the skew in concentration distributions, we consider medians as  
190 opposed to means and, in our regressions, relationships between log concentrations and linear  
191 UOG metrics (Text S3).

192 Additionally, we assessed relationships with three specific activities associated with  
193 UOG: problems surrounding the casing and cementing of wells, impoundment of wastes, and  
194 spills of wastes. To assess associations with specific activities, we analyzed violations  
195 documented by the PADEP for casing and cementing impairments, impoundment-related issues,  
196 and pollution incidents (e.g., spills or leaks). We classified relevant violations in the PADEP Oil  
197 and Gas Compliance database into these three categories after slightly modifying a published  
198 scheme (Table S2).<sup>5,54</sup> While casing or cementing problems are known to sometimes allow gas  
199 leakage into groundwater, pollution incidents involving leaks from faulty impoundments or spills  
200 could enter either surface or groundwaters.

201 To investigate methods to account for the influence of background geologic and  
202 anthropogenic processes on our analyses, we also utilized a fixed effects regression model to  
203 better account for potentially confounding overlap with other sources of geogenic or  
204 anthropogenic salt. This regression includes binary, “dummy” variables reflecting a water  
205 sample’s proximity to anthropogenic and geologic sources of salt ions, as well as the bedrock  
206 lithology and season of sample collection (Text S3). These variables subsequently group the  
207 samples based on shared land use characteristics, and the fixed effects models only use the  
208 variation within groups to estimate the effect of the predictors. This method allows a better  
209 analysis of within-group variation, which can help reduce omitted variable bias from  
210 confounding factors (see Text S3 for a full description). We included additional tests to account  
211 for the small portion of censored concentration data (Text S5). For full details on the methods,  
212 see Text S3.

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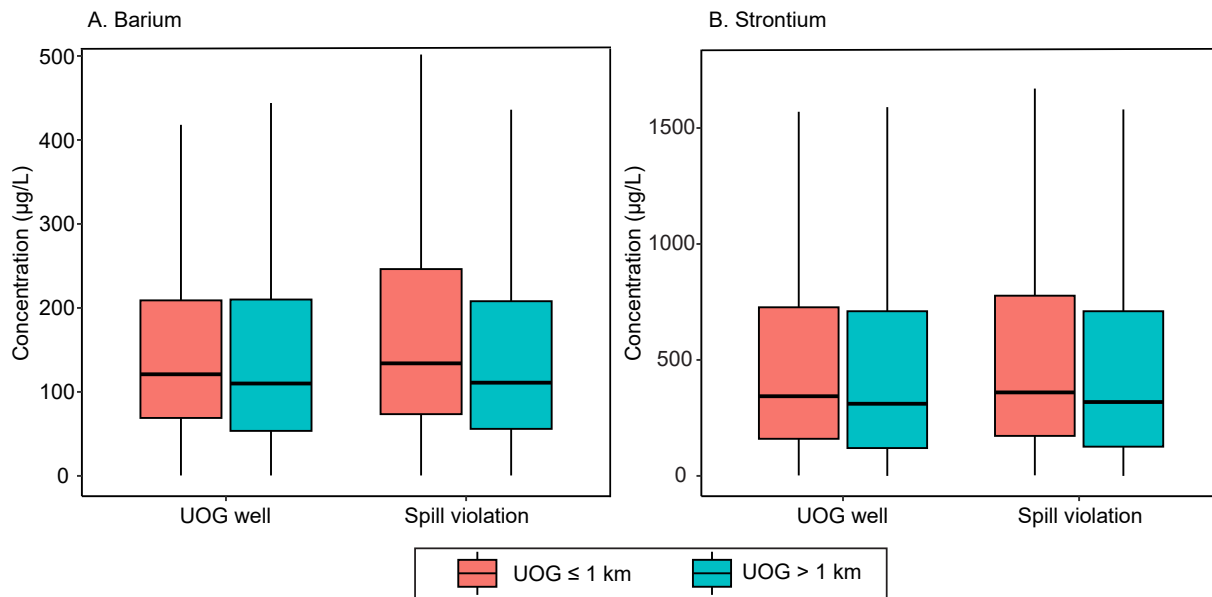
## 214 **Results**

### 215 *Barium and strontium concentrations increase with proximity to UOG wells and spills*

216 Throughout the paper we use statistical analyses to look for associations, and we define  
217 significance as referring to p-values  $< 0.05$ . We first considered whether median concentrations  
218 of brine salts are elevated in samples near UOG development. For this comparison we use the  
219 Wilcoxon-Mann-Whitney (WMW) test, as well as the more stringent Brunner-Munzel test, both  
220 of which are well-suited for non-parametric data.

221 We observed significantly higher median [Ba] and [Sr] in samples located within 1 km of  
222 a UOG well across PA (Figure 2). Median [Ba] and [Sr] in samples within 1 km of a UOG well  
223 are 11  $\mu\text{g/L}$  and 32  $\mu\text{g/L}$  higher, respectively, than samples farther than 1 km. UOG development

224 within 1 km thus corresponds to 12.2% higher median [Ba] and 10.5% higher median [Sr]. These  
 225 comparisons remain statistically significant using the Brunner-Munzel statistical test (Table S3,  
 226 Table S4).  
 227  
 228



229  
 230 Figure 2: Box and whisker plots summarizing statewide (A) barium and (B) strontium  
 231 concentrations for Pennsylvania samples  $\leq 1$  km from locations of UOG development (red) and  
 232  $> 1$  km from UOG development (aqua) for UOG wells and spill violations. The bounds of the  
 233 boxes depict Q1 and Q3 concentrations, while the thick center line displays the median  
 234 concentration. All comparisons shown found a statistically significant difference in median  
 235 concentrations (see Figure S5 for the comparison of all UOG attributes considered). Outlier data  
 236 (defined as  $> Q3 + 1.5 * IQR$  or  $< Q1 - 1.5 * IQR$ , where Q1 and Q3 are the first and third  
 237 quartiles and IQR is the interquartile range) are not plotted due to the large right skew in the  
 238 data. Calculations only include UOG wells spudded before water sample collection.  
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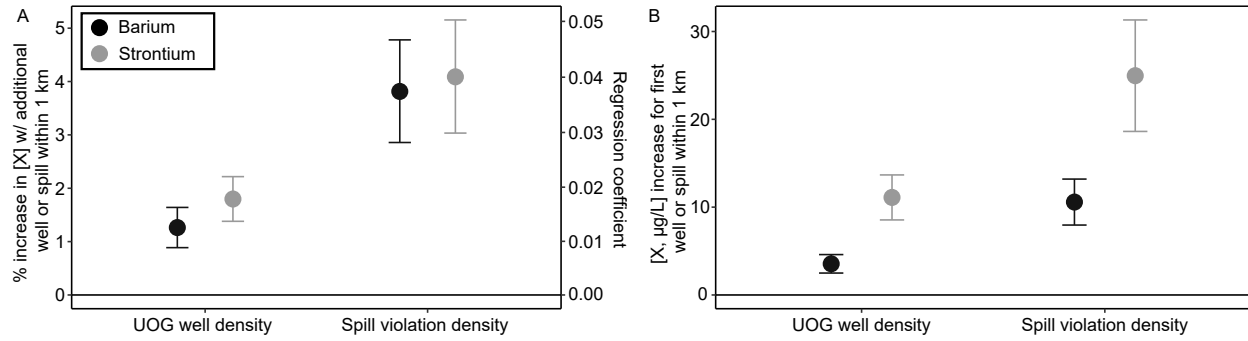
240           Next, we investigated whether these increases persist when considering specific UOG  
241 processes as documented by violations at UOG wellpads in the PADEP compliance database.  
242 Median [Ba] and [Sr] are significantly higher within 1 km of a pollution violation (“spill”) across  
243 the state as compared to samples >1 km (Table S3, Table S4). Furthermore, the magnitude of the  
244 increase within 1 km of spills is larger than the increase within 1 km of a UOG wellpad (Figure  
245 2). Once again, these relationships remain statistically significant using the Brunner-Munzel test  
246 (Table S3, Table S4). In contrast, we observe no significant increase in median [Ba] or [Sr]  
247 within 1 km of wellpads cited for violations related to impoundment or casing/cementing  
248 violations (Table S3, Table S4).

249

#### 250 *Brine salt concentrations increase with density of UOG wells*

251           Given observed statewide increases in median [Ba] and [Sr] within 1 km of UOG  
252 wellpads, we investigated whether [Ba] and [Sr] also show significant increases associated with  
253 higher density of UOG wells. We identified small but statistically significant relationships  
254 between [Ba] and [Sr] and the density of UOG wells within 1 km (Figure 3A, Table S5).  
255 Regressions calculated using a radius of 3 km rather than 1 km typically revealed smaller  
256 regression coefficients (e.g., a smaller magnitude of impacts) but strengthened significance,  
257 where the latter result is likely related to the larger number of samples within the 3 km buffer  
258 (Table S5). Both [Ba] and [Sr] also significantly increase with proximity to the nearest UOG  
259 well (Table S5). In sum, these data are consistent with increases in groundwater [Ba] and [Sr]  
260 with UOG development.

261



262

263 Figure 3: Percent increases in the concentration of barium and strontium associated with  
 264 increasing UOG development density within 1km, and their associated regression coefficients

265 (A). These values were calculated using Equation 1 for the full statewide dataset using  
 266 regressions analyzing the relationship between log[Barium] and log[Strontium] and UOG well  
 267 density or pollution violation (“spill”) density within a 1 km radius of water samples. The  
 268 corresponding average increase in ion concentrations for the first well spudded or spill within 1  
 269 km are shown in (B), calculated using Equation 2 with mean concentrations. Error bars show  
 270 standard error. All regressions yielded statistically significant ( $p < 0.05$ ) correlations.

271 Calculations only include UOG wells spudded before water sample collection.

272

273

274 We next used our regressions to quantify the magnitude of increase in concentration (e.g.,  
 275 in  $\mu\text{g/L}$ ) per additional UOG well within 1 km. In a log-linear model regression such as used  
 276 here, the regression coefficient,  $\beta$ , calculated for relationships between log concentrations and  
 277 UOG well density cannot be directly interpreted as the increase in concentration (e.g., in  $\mu\text{g/L}$ )  
 278 per additional well. However, the percent increase in concentration for every additional well can  
 279 be calculated as:

280

$$\% \text{ increase} = e^{\beta} - 1 \quad \text{Equation 1}$$

281 where  $\beta$  is the calculated regression coefficient. Following equation 1, [Ba] increases by 1.27%  
282 and [Sr] by 1.80% for every additional UOG well within 1 km. We can corroborate these values  
283 by also assessing the increase in concentration for every additional well using an estimate of the  
284 Akritas-Theil-Sen slope (Text S7). Using this alternate regression calculation which can handle  
285 non-parametric, censored data, we estimate a 2.2 - 2.6  $\mu\text{g/L}$  increase in [Ba] and 6.1 - 8.2  $\mu\text{g/L}$   
286 increase in [Sr] per additional UOG well within 1 km. These increases represent up to 2.3% and  
287 2.5% of the median [Ba] and [Sr], respectively, and 0.92% and 1.32% of the mean [Ba] and [Sr]  
288 in our groundwater dataset.

289 From our regression coefficients, we can also estimate the average  $\mu\text{g/L}$  increase in [Ba]  
290 or [Sr] ( $\Delta C_{\text{avg}}$ ) from UOG well density (#UOG1km) across the entire study area as:

$$291 \quad \Delta C_{\text{avg}} = C_{\text{avg}} * (e^{\beta * \#UOG1\text{km}} - 1) \quad \text{Equation 2}$$

292 Here  $C_{\text{avg}}$  is the mean concentration of Ba or Sr across the region of interest ( $\mu\text{g/L}$ ), #UOG1 km  
293 is the number of UOG wells within 1 km (density), and  $\beta$  is the regression coefficient. Based on  
294 the calculated regression coefficients and mean [Ba] and [Sr] (283  $\mu\text{g/L}$  and 623  $\mu\text{g/L}$ ,  
295 respectively), we calculate that the first UOG well spudded within 1 km (#UOG1km = 1)  
296 increases [Ba] and [Sr] by 3.6  $\mu\text{g/L}$  and 11.2  $\mu\text{g/L}$ , respectively (Figure 3B). Using instead the  
297 mean #UOG1km (0.72 UOG wells within 1 km) for groundwater samples in the full statewide  
298 dataset, the average concentration increases attributed to UOG development are 2.58  $\mu\text{g/L}$  (Ba)  
299 and 8.04  $\mu\text{g/L}$  (Sr). At the highest density of UOG wells within 1 km of a water sample in PA (n  
300 = 21 UOG wells), this corresponds to an 85.7  $\mu\text{g/L}$  increase in [Ba] and a 282.4  $\mu\text{g/L}$  increase in  
301 [Sr].

302

303 *Potential sources of UOG wastewater releases*

304 Across PA, [Ba] and [Sr] also show a statistically significant increase with the number of  
305 pollution violations (i.e., spills) within 1 km (Table S6, Table S7). Given both UOG well density  
306 and spill density are expressed as the number of UOG wells or spills, respectively, within 1 km,  
307 we compared regression coefficients to understand the relative impacts of UOG wells versus  
308 spills. One additional spill within 1 km has a greater impact on concentration compared to one  
309 additional UOG well (Figure 3). For example, we calculate 3.8% and 4.1% increases in [Ba] and  
310 [Sr], respectively, for every additional spill within 1 km using Equation 1 and the coefficients  
311 from our regression analyses (Figure 3A, Table S6, Table S7). Following Equation 2, we  
312 calculate the average effect of the first spill within 1 km to be a 10.8  $\mu\text{g/L}$  increase in [Ba] and a  
313 25.5  $\mu\text{g/L}$  increase in [Sr] (Figure 3B). Estimates of the Akritas-Theil-Sen slope are also  
314 consistent with 2-3x greater increases in [Ba] and [Sr] associated with an increasing number of  
315 spills within 1 km compared to all UOG wells (Text S7). Consistent with the trends we observed  
316 in median concentrations, the other violations we considered were not associated with significant  
317 increases in [Ba] and [Sr] (Table S6, Table S7).

318 When [Ba] and [Sr] are evaluated relative to distance rather than density of UOG  
319 development metrics statewide, we identify significant relationships indicating increasing salt  
320 concentrations closer to UOG development for all metrics except [Sr] and impoundment  
321 violations (Table S6, Table S7).

322

323 *Statistically significant relationships persist when accounting for overlapping sources*

324 As discussed previously, UOG development overlaps with other sources of salt ions in  
325 groundwater and other features that could obscure contamination. These factors include legacy  
326 hydrocarbon extraction (e.g., conventional oil and gas brines and coal mining), structural features

327 conducive to migration of natural basin brines (e.g., along faults or channelized by anticlinal  
328 folding), and road salting. When we implement a fixed effects regression that better accounts for  
329 these features (Text S3), relationships between salt ions and UOG well density and distance  
330 remain statistically significant (Table S8, Table S9, Figure S6).

331

### 332 *Relationships between UOG development and brine salt ion concentrations in subregions of PA*

333 To understand what causes statewide increases in salt ion concentrations in groundwater  
334 and to investigate why a few regressions do not yield significant correlations, we also examined  
335 whether confounding variables may affect statewide relationships by investigating two  
336 subregions of the state separately (northeastern PA and southwestern PA). The subregions are  
337 characterized by the highest density of UOG development but differ with respect to land use and  
338 geology (Text S2).

339 Consistent with the statewide data, median [Ba] and [Sr] are higher within 1 km of UOG  
340 wells in both subregions (Table S3, Table S4). Additionally, median [Ba] and [Sr] are generally  
341 higher within 1 km of spills (Table S3, Table S4). The only exception is [Sr] in southwestern PA  
342 (Table S4). Median [Ba] is also significantly higher within 1 km of historical impoundments in  
343 southwestern PA (Table S3).

344 We also investigated correlations with respect to distance and density within these  
345 subregions. We observed relationships that were statistically significant for both analytes in both  
346 southwestern PA and northeastern PA with respect to distance to the nearest UOG well. In other  
347 words, both Ba and Sr increase in concentration closer to UOG wells in each subregion (Table  
348 S5). We also identify small, significant increases in both analytes with increased UOG well



349 density in southwestern PA, just as we observed in the statewide analysis (Table S5). However,  
 350 we did not observe this relationship with UOG well density in northeastern PA (Table S5).

351 Additionally, we observe significant increases in [Ba] and [Sr] in southwestern PA  
 352 associated with a higher density of spills within 1 km (Table S6). [Sr] in southwestern PA also  
 353 increases with greater density of casing/cementing violations (Table S6, Table S7). In contrast,  
 354 [Ba] and [Sr] are not significantly correlated with spill density within 1 km in northeastern PA  
 355 (Table S6, Table S7).

356 Most of the inconsistencies we observe between our statewide versus regional analyses  
 357 disappear after implementing fixed effects for other salt sources. For example, when we include  
 358 fixed effects, relationships among UOG well density and [Ba] and [Sr] are statistically  
 359 significant in both southwestern PA and northeastern PA (Figure S6, Table S8, Table S9).  
 360 Similarly, relationships between [Sr] and spill violation density are significant in both  
 361 southwestern PA and northeastern PA when fixed effects are implemented (Table S9).

362 In summary, we observed statistically significant relationships statewide between [Ba]  
 363 and [Sr] and UOG wells and spills across all methods of comparison (Table 1). These  
 364 relationships were often statistically significant within subregions southwestern PA or  
 365 northeastern PA as well, especially when fixed effects were included in regression analyses  
 366 (Table 1).

367

368 Table 1: p-values for the relationship between barium or strontium, and UOG development  
 369 variables across comparison of medians and regression analyses.

Species	UOG variable <sup>1</sup>	Comparison of medians (1 km) <sup>2</sup>	Density (within 1 km)	Distance	Density-with fixed effects	Distance-with fixed effects
<i>Full PA dataset</i>						
Barium	UOG wells	<0.001	<0.001	<0.001	<0.001	<0.001
	Spills	<0.001	<0.001	<0.001	<0.001	<0.001

Strontium	UOG wells	<b>&lt;0.001</b>	<b>&lt;0.001</b>	<b>&lt;0.001</b>	<b>&lt;0.001</b>	<b>&lt;0.001</b>
	Spills	<b>&lt;0.001</b>	<b>&lt;0.001</b>	<b>&lt;0.001</b>	<b>&lt;0.001</b>	<b>&lt;0.001</b>
<i>Northeastern PA</i>						
Barium	UOG wells	<b>&lt;0.001</b>	0.063	<b>&lt;0.001</b>	<b>0.001</b>	<b>&lt;0.001</b>
	Spills	<b>&lt;0.001</b>	0.198	<b>&lt;0.001</b>	0.088	<b>&lt;0.001</b>
Strontium	UOG wells	<b>&lt;0.001</b>	0.893	<b>&lt;0.001</b>	<b>0.002</b>	<b>&lt;0.001</b>
	Spills	<b>0.014</b>	0.114	<b>&lt;0.001</b>	<b>0.004</b>	<b>&lt;0.001</b>
<i>Southwestern PA</i>						
Barium	UOG wells	<b>&lt;0.001</b>	<b>&lt;0.001</b>	<b>&lt;0.001</b>	<b>&lt;0.001</b>	<b>&lt;0.001</b>
	Spills	<b>&lt;0.001</b>	<b>&lt;0.001</b>	<b>&lt;0.001</b>	<b>&lt;0.001</b>	<b>&lt;0.001</b>
Strontium	UOG wells	<b>&lt;0.001</b>	<b>0.004</b>	<b>&lt;0.001</b>	<b>0.004</b>	<b>0.025</b>
	Spills	0.127	<b>0.012</b>	<b>&lt;0.001</b>	<b>&lt;0.001</b>	<b>&lt;0.001</b>

- 370 1. Bolded values indicate statistically significant ( $p < 0.05$ ) correlations with the respective  
371 variable  
372 2. p-value is displayed for a two-sided WMW test, see tables S3/S4 for one-sided and BM results  
373

## 374 Discussion

375 *Brine salts increases likely because of wastewater mishandling*

376 Statewide, we observed significantly higher median [Ba] and [Sr] within 1 km of UOG  
377 wells, as well as significant increases in [Ba] and [Sr] with a higher density of UOG wells (Table  
378 1). When we repeat these analyses to instead consider only COG wells or all oil and gas wells  
379 (UOG + COG), we do not identify such consistent relationships (Text S8). Similar increases  
380 associated with UOG development have been reported for surface waters nationwide<sup>27</sup> and for  
381 groundwaters in southwestern PA,<sup>28</sup> but our study is the first to indicate a statewide increase in  
382 groundwater brine salt ion concentrations associated with UOG development. The coefficients  
383 we calculate for increases in [Ba] and [Sr] in groundwater per UOG well are ~25-50 times larger  
384 than observed for PA surface water,<sup>27</sup> consistent with greater dilution of surface waters by  
385 meteoric water as compared to groundwaters. The surface water trends are plausibly driven by  
386 groundwater contamination, especially considering that most streams in PA are gaining  
387 streams.<sup>55</sup>

388 We also observed statewide that median [Ba] and [Sr] were higher within 1 km of  
389 documented pollution violations, and we identified significant increases in the concentrations of  
390 these ions correlated with higher spill density. The increases in concentration associated with  
391 spills were typically larger than the increases calculated for regressions versus proximity to or  
392 number of UOG wells alone. From this we infer that a subset of UOG wells that experienced  
393 spills may drive the regional correlations with UOG wells. In other words, the small regional  
394 increases may be explained by problematic, isolated sites. We emphasize spills as the likeliest  
395 pathway for salts to reach groundwater because we observed consistently significant  
396 relationships across multiple tests: comparison of medians, regressions with UOG density and  
397 distance, and fixed effects analysis (Table 1). In contrast, violations pertaining to subsurface well  
398 integrity (i.e., casing/cementing violations) were not associated with significant concentration  
399 increases across these tests. In the case of both spills and well integrity issues, we acknowledge  
400 that not all incidents that may merit a violation are necessarily reported or documented.  
401 However, this observation nonetheless suggests that surface impacts rather than downhole  
402 problems are primarily responsible for slight groundwater salinization during UOG development  
403 at the well depths reflected in our dataset.

404 To further test whether a surface source is the best explanation for the impacts we  
405 observe, we repeated our analyses considering only UOG wellpads located at higher elevations  
406 than the respective water sample. We conducted this analysis because it is less likely that water  
407 samples could have been impacted by surface processes at a lower-elevation UOG wellpad due  
408 to the strong control of gravity on shallow groundwater flow in the Appalachian Basin.<sup>48,56</sup> When  
409 we consider only higher-elevation UOG wells, we observe that the effect of UOG development  
410 becomes even stronger, resulting in larger regression coefficients and increased significance for

411 relationships between ion concentrations and UOG well density (Table S10). When only higher-  
412 elevation UOG wells are included in the calculation, we calculate 3.88 and 12.32 µg/L average  
413 increases in Ba and Sr based on the average density of higher elevation UOG wells, and 139 and  
414 483 µg/L increases in Ba and Sr at the highest UOG well density. We similarly observe  
415 increased coefficients in regressions analyzing only higher-elevation spills relative to those  
416 analyzing all spills (Table S11). The strengthened relationship among UOG development and  
417 salt ion concentrations when only higher elevation wellpads are considered further supports a  
418 surface source of contamination. The lack of significant positive relationships with  
419 casing/cementing violations further supports that surface sources of brine, rather than subsurface  
420 activities such as hydraulic fracturing, explain increased [Ba] and [Sr] nearby UOG  
421 development.

422         To further investigate the hypothesis that spills could explain increases in brine salt ions,  
423 we examined waste production data from UOG wells in proximity to water samples. Our  
424 working hypothesis was a greater volume of produced water may create more potential for  
425 mishandling and larger volumes of spillage when problems occur. Regressing log concentrations  
426 against log production volumes prior to water sample collection, we identified a significant  
427 increase in [Sr] associated with larger volumes of produced water at UOG wells within 1 km of  
428 the respective water sample across PA and for [Ba] in southwestern PA (Text S9, Table S12).

429         While these associations point to spills as a likely mechanism for increased salt ion  
430 concentrations in groundwaters, most wellpad spills are very small in volume.<sup>6,15</sup> For example,  
431 regulatory data indicate reported spills in PA are typically 100 L-10,000 L in volume.<sup>15</sup> A mass  
432 balance calculation informed by geological observations reveals that only produced water spills  
433 near the upper range of reported spill volumes (e.g., > ~1000 L) are likely to explain the average

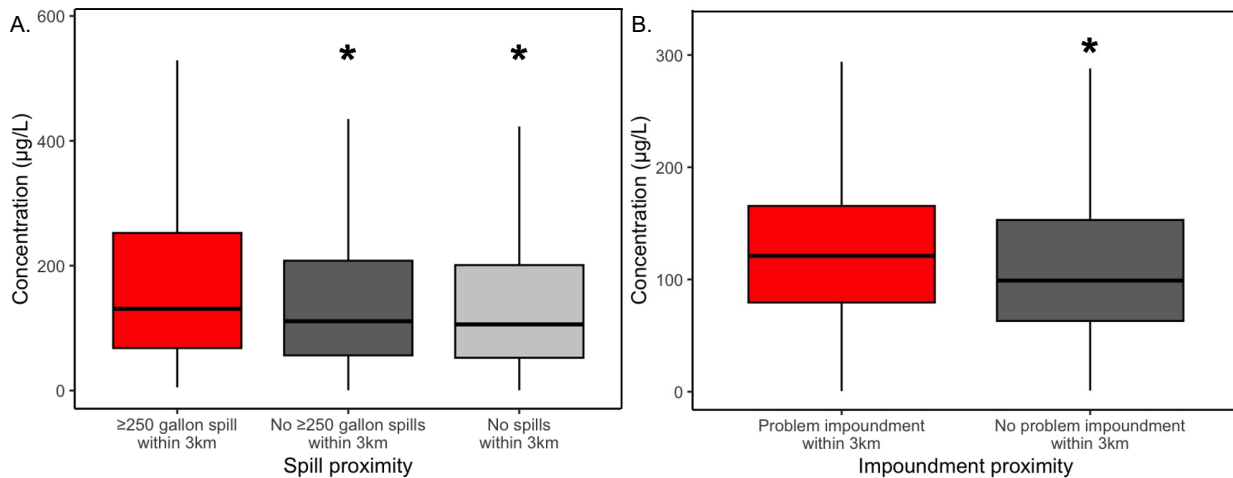
434 increase in [Ba] in groundwater we observe within 1 km of UOG wells (Text S10). The salt  
435 contamination we document is therefore most likely associated with the small number of isolated  
436 high-volume spills.

437         Although we wanted to assess local contamination on a spill-by-spill basis, spill volumes  
438 are not widely reported for violations cited by the PADEP: only 232 / 1338 spills catalogued up  
439 to 2014 include volume estimates.<sup>6</sup> If we nonetheless investigate those reported incidents and  
440 define “large spills” as >250 gallons (~1000 L), we can calculate if large spills influenced [Ba] in  
441 the 102 or 1302 analyzed samples from nearby groundwater with respect to the two buffer  
442 distances, 1 km or 3 km respectively. We observed that the median [Ba] for samples within the  
443 buffer distance from large spills (137 µg/L for 1 km or 131 µg/L for 3 km) is ~23-24% higher  
444 than the median in samples over 1 km or 3 km from any reported spills (111 µg/L for 1 km and  
445 106 µg/L for 3 km) (Figure 4A, Table S13). Similar relationships were observed for the >500-  
446 gallon spills (Table S14) but the smaller number of documented >500-gallon spills (n = 63)  
447 yields statistical significance only for a buffer of 3 km (where a larger number of samples, n =  
448 902, are located within 3 km of a >500 gallon spill vs. n = 77 samples within 1 km).

449         The totality of these results leads us to attribute the slightly higher concentrations of brine  
450 salt ions in surface and ground waters near UOG development<sup>27,28</sup> to isolated incidents of spills  
451 and leaks on wellpads. Consistent with this possibility, wastewater spills and leaks in some  
452 locations have resulted in well-documented increases in salt ion concentrations in nearby surface  
453 waters.<sup>21,26,57</sup> However, this is the first published study to document a regional impact of UOG  
454 development on water resources where evidence for the specific cause has also been identified.

455  
456

457



458

459 Figure 4: Box and whisker plots displaying barium concentrations for (A) samples within 3 km  
460 of a large ( $\geq 250$  gallon or 964 L) spill vs. samples  $> 3$  km from a large spill and samples  $> 3$  km  
461 from any spill, and (B) southwestern PA samples within 3 km of an impoundment that was  
462 mandated to close, upgrade, or store only freshwater by the PADEP as compared to southwestern  
463 PA samples  $> 3$  km from these impoundments. In both A and B, a significant increase in median  
464 [Ba] was identified within 3 km of the spills/impoundments, where an asterisk (\*) denotes  
465 significant differences between sample groups relative to samples within 3 km.

466

467 *Wastewater impoundments may also release salt ions to groundwater*

468 A second kind of spill or leakage may also have been important early in UOG  
469 development in PA, namely, leakage of wastewaters from impoundments.<sup>5</sup> To investigate this,  
470 we considered correlations with the locations of wellpad impoundments identified using 2010  
471 satellite imagery (henceforth referred to as historical impoundments), which may have stored  
472 UOG wastewaters (Text S2).<sup>45</sup> These “historical impoundments” are potentially important

473 because after 2016, temporary storage of wastewaters in wellpad impoundments was  
474 discontinued in PA.<sup>47</sup>

475 In particular, we observe the strongest evidence for impacts from these impoundments in  
476 southwestern PA: in that area, [Ba] is significantly higher within 1 km of historical  
477 impoundments and [Ba] increases with greater density and proximity of these impoundments  
478 (Table S3, Table S6). The problematic nature of some impoundments has previously resulted in  
479 regulatory action in southwestern PA. Specifically, because of observed or inferred infractions,  
480 eight impoundments in southwestern PA (out of an estimated 500-600 operating statewide yearly  
481 before 2016) were ordered by the PADEP in 2014 to be i) fully shut down, ii) upgraded with  
482 respect to liners and systems for leak detection, or iii) limited to storage of only freshwater.<sup>45,58</sup>  
483 The USEPA also documented likely leakage of Cl from one of the eight impoundments into  
484 downgradient groundwater at a location where significant health impacts were alleged.<sup>59</sup>

485 When we compare median [Ba] between southwestern PA samples within 1 km of the  
486 estimated locations of these eight impoundments vs. samples >1 km away (Text S2), we find  
487 ~34% higher median [Ba] in samples within 1 km of these impoundments (134 vs. 100 µg/L)  
488 (Table S15). We observe a similar increase when we compare median [Ba] for samples within 3  
489 km of an impoundment (123 µg/L) vs. samples >3 km away (99 µg/L) (Figure 4B, Table S15).  
490 These differences are statistically significant within both 1 km and 3 km. In addition, one of  
491 these problematic impoundments is located within a previously identified subregion (“hotspot”)  
492 in southwestern PA where [Cl] increased with higher UOG well density.<sup>28</sup>

493

494 *Regional differences in hydrogeology and land usage complicate identification of impact*

495 We generally observed statistically significant relationships between [Ba] and [Sr] and  
496 UOG development density in southwestern PA but not in northeastern PA. This comparison of  
497 southwestern PA and northeastern PA is important not only because these subregions contain  
498 some of the highest density of UOG development in the world, but also because the data  
499 demonstrate how geology and land use combine to complicate the detection of contamination  
500 during UOG development. In particular, southwestern PA and northeastern PA differ with regard  
501 to the topographic relief (higher in northeastern PA) as well as the extent of prior hydrocarbon  
502 extraction (extensive legacy development in southwestern PA).<sup>43</sup>

503 The importance of topographic relief may explain why we observed increased  
504 significance in northeastern PA when we accounted for elevation or overlapping sources in our  
505 analyses. For example, when we investigated the association of salt ions with the density of  
506 UOG wells in northeastern PA, relationships were not statistically significant. However, when  
507 we considered only higher elevation UOG wells or implemented a fixed effects regression,  
508 increases in concentration associated with UOG density (Ba) and UOG distance (Ba and Sr)  
509 were of greater magnitude and statistically significant.

510 One explanation for these results is that strong topographic and geologic influence on  
511 brine salt occurrence in northeastern PA masks effects of UOG development in that region. In  
512 particular, where topographic relief is the highest (in northeastern PA), naturally elevated  
513 concentrations of species like Ba and Sr are generally observed in valley bottoms and other  
514 topographic lows.<sup>32</sup> This natural phenomenon has been attributed by some to natural upwelling  
515 of Appalachian Basin brines from deeper than a few hundred meters depth into valleys.<sup>32,60</sup> An  
516 alternative explanation is that these natural brines were forced to migrate upward during tectonic  
517 orogeny in the deep geologic past, and although these brines are no longer migrating, the salts in



518 the rock have not yet been completely flushed out yet.<sup>43</sup> Regardless of the explanation, natural  
519 brine migration may be particularly important in northeastern PA because of geologic features in  
520 that area such as anticlinal folding and faults.<sup>23,31</sup> While groundwater flow is still predominantly  
521 gravity-driven and brines can still occur at shallow depths in southwestern PA, topographic relief  
522 is smaller and the extent of surface faulting is more limited.<sup>43,56</sup> As a result, topographic forcing  
523 likely has a smaller influence on groundwater chemistry in southwestern PA, with less  
524 differentiation between fresher (e.g., Ca-HCO<sub>3</sub> type) waters at high elevation and saltier (e.g.,  
525 Na-Cl type) groundwaters at low elevations.<sup>28,43</sup> These hydrogeologic differences may serve to  
526 mask some of the impacts of brine spills on groundwater in northeastern PA compared to  
527 southwestern PA, as the strong topographic forcing on brine salt occurrence in northeastern PA  
528 may obscure any increases in salt ion concentrations from UOG development.

529         In addition to geogenic processes shaping groundwater chemistry, the long history of  
530 energy development in southwestern PA also complicates contaminant attribution. For waters  
531 sampled in southwestern PA that were >1 km from UOG wells and coal mining but located <1  
532 km from COG wells, we did not see significant differences in median [Ba] or [Sr] compared to  
533 samples >1 km from any hydrocarbon extraction (Table S1). From this we inferred the effects of  
534 these legacy COG wells on groundwater chemistry may be minor in southwestern PA. However,  
535 our dataset shows significant increases in [Sr] and decreases in [Ba] associated with coal mining  
536 (Table S1). The increase in [Sr] nearby coal mining is not surprising because of the ubiquity of  
537 acidic mine drainage in the area and the likelihood that acids dissolve local carbonate bedrock,  
538 releasing Sr incorporated in the carbonate lattice during dissolution.<sup>51</sup> Lower [Ba] nearby coal  
539 mining may be explained by a) significantly higher median [SO<sub>4</sub>] where coal mining is <1 km  
540 from the water sample (likely reflecting sulfate produced via sulfide mineral oxidation, the

541 driving force of acid mine drainage production) and b) the low solubility of Ba and SO<sub>4</sub> in co-  
542 solution.<sup>61</sup>

543 Despite such overlap, the significance of relationships between Ba and Sr and UOG  
544 development in southwestern PA persists after the implementation of fixed effects to control for  
545 overlapping anthropogenic sources of salts (Table S8, Table S9). In some cases, the impacts of  
546 UOG development on salt ion concentrations (particularly [Ba]) in groundwater appear strongest  
547 in southwestern PA, potentially implying that overlap with legacy hydrocarbon extraction may  
548 increase contamination during UOG development. However, our investigation also reveals that  
549 other attributes in southwestern PA (namely problematic impoundments) may explain why  
550 impacts sometimes appear greater in southwestern PA. As such, we cannot conclude that overlap  
551 between UOG development and other forms of hydrocarbon extraction increases the frequency  
552 of contamination.

553

#### 554 *Environmental implications*

555 Across the largest shale gas play with public access to high-density groundwater data in  
556 the world, UOG development is associated with slightly increased concentrations of brine salt  
557 ions in groundwater (this study) and surface waters (Bonetti et al.).<sup>27</sup> Our results also suggest  
558 these regional impacts are best explained by a small subset of large spills or leaks that occurred  
559 at wellpads and impoundments. These incidents likely produce “hotspots” where concentrations  
560 of brine species increase nearby UOG development, explaining the regional effects.<sup>28</sup>

561 Our estimates suggest the average increases in [Ba] and [Sr] associated with UOG  
562 development should not exceed 15% of the USEPA’s recommended levels for either Ba or Sr (2  
563 mg/L and 4 mg/L, respectively). However, the occurrence of relatively elevated Ba and Sr in

564 groundwaters near UOGD highlights the potential for the presence of more hazardous species in  
565 brines that are not widely monitored or only reportable at very high concentrations. These  
566 include toxic trace elements such as thallium, arsenic, and cadmium, and the species responsible  
567 for most of the radioactivity in the brines (radium). To investigate this, we calculated the  
568 statewide medians for species concentrations in the USGS Produced Water database, including  
569 trace metals, organics, and radioactive species. We then assumed that the statewide median mass  
570 ratios of [X] to [Ba] or [Sr] (where X is one of the species measured in produced waters) in the  
571 produced water could be used to estimate [X] as a function of our data for [Ba] and [Sr]. The  
572 increases we calculate statewide represent, on average, very small portions of brine mixing into  
573 water samples. For example, our calculated increases in [Ba] and [Sr] are always <150 µg/L (Ba)  
574 and <500 µg/L (Sr) even at the highest UOG well density, which would represent mixing of  
575 <0.04% brine based on median [Ba] and [Sr] reported for Marcellus produced waters. Based on  
576 these calculations, we estimate that other potentially hazardous species are also not likely to  
577 exceed USEPA limits when considered on a regional basis (Table S16).

578         However, these statewide estimates do not exclude the possibility of localized risks  
579 because the regional concentration effects we have documents are likely caused by localized  
580 contamination incidents. This is in-line with previous work that found increases in [Cl]  
581 calculated per UOG well in southwestern PA were over 10x greater in some geospatially-  
582 identified hotspots than calculated regionwide.<sup>28</sup> Mixing of just 0.2% - 0.5% brine in could drive  
583 the concentrations of species including radium to exceed EPA limits based on average produced  
584 water compositions in PA.

585         We show that brine contamination has likely affected groundwaters in the largest shale  
586 gas play in the world where water quality data are publicly available. The high production

587 volumes and salinity of produced waters in other major shale gas plays<sup>62</sup> and relative ubiquity of  
588 spills<sup>6,15</sup> leads to the conclusion that similar impacts should be studied in other shale gas plays,  
589 and especially where very large spills have occurred (Text S11).<sup>15,26</sup> In some shale plays,  
590 produced water volumes exceed recycling and re-injection capabilities and this is projected to  
591 increase worldwide into the future.<sup>63</sup> Further, while UOG wells generate greater brine volumes  
592 than COG wells on a per-well basis, problems surrounding wastewater storage also occur and  
593 can contaminate groundwater during COG development. Our results emphasize the need for  
594 stringent management of oil and gas wastewaters to protect water resources.

595

596 **Supporting Information:**

597 Estimates of wastewater volumes produced from unconventional vs. conventional wells  
598 (Text S1), additional dataset details (Text S2), additional details on analysis methods (Text S3),  
599 comparison of medians across land uses (Text S4), treatment of censored data (Text S5),  
600 analyses with chloride (Text S6), estimates of Akritas-Theil-Sen slopes (Text S7), analyses for  
601 conventional and all oil and gas wells (Text S8), relationships with wastewater production  
602 volumes (Text S9), spill mixing calculation details (Text S10), implications across shale gas  
603 basins (Text S11), and potential explanations for inverse relationships (Text S12). Sample  
604 collection dates (Figure S1), insets of Figure 1 (Figures S2-3), map of oil and gas wells and coal  
605 mining locations in PA (Figure S4), full comparison of medians (Figure S5), regression  
606 coefficient and p-values with and without fixed effects (Figure S6), fixed effects sensitivity  
607 analyses (Figure S7). Comparison of median concentrations across land uses (Table S1),  
608 violation classification scheme (Table S2), comparison of median [Ba] and [Sr] (Tables S3-S4).  
609 Results for regression analyses considering UOG wells (Table S5), UOG violations (Tables S6-

610 S7), with fixed effects (Tables S8-S9), only higher-elevation wells/spills (Tables S10-S11), and  
611 waste production volumes (Table S12). Comparison of median [Ba] relative to locations of large  
612 spills and impoundments (Tables S13-S15), median ratios of additional species to [Ba] or [Sr] in  
613 PA produced waters (Table S16), summary statistics for species in the dataset (Table S17),  
614 coordinates estimated for impoundments (Table S18), summary statistics of sample proximity to  
615 relevant geologic and anthropogenic features (Table S19), number of samples nearby relevant  
616 features (Table S20), summary statistics for sample proximity to violations (Table S21), analyses  
617 using a 3 km radius (Tables S22-24), analyses using a tobit regression (Tables S25-S28),  
618 Akritas-Theil-Sen slope estimates (Tables S29-S31), and analyses considering conventional  
619 wells and all oil and gas wells (Tables S32-S34).

620

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631

632

633 **References**

- 634 (1) Lutz, B. D.; Lewis, A. N.; Doyle, M. W. Generation, Transport, and Disposal of Wastewater  
635 Associated with Marcellus Shale Gas Development. *Water Resour. Res.* **2013**, *49* (2), 647–  
636 656. <https://doi.org/10.1002/wrcr.20096>.
- 637 (2) Vidic, R. D.; Brantley, S. L.; Vandenbossche, J. M.; Yoxtheimer, D.; Abad, J. D. Impact of  
638 Shale Gas Development on Regional Water Quality. *Science* **2013**, *340* (6134), 1235009.  
639 <https://doi.org/10.1126/science.1235009>.
- 640 (3) Chen, L.; Miller, S. A.; Ellis, B. R. Comparative Human Toxicity Impact of Electricity  
641 Produced from Shale Gas and Coal. *Environ. Sci. Technol.* **2017**, *51* (21), 13018–13027.  
642 <https://doi.org/10.1021/acs.est.7b03546>.
- 643 (4) Jiang, M.; Griffin, W. M.; Hendrickson, C.; Jaramillo, P.; VanBriesen, J.; Venkatesh, A. Life  
644 Cycle Greenhouse Gas Emissions of Marcellus Shale Gas. *Environ. Res. Lett.* **2011**, *6* (3),  
645 034014. <https://doi.org/10.1088/1748-9326/6/3/034014>.
- 646 (5) Brantley, S. L.; Yoxtheimer, D.; Arjmand, S.; Grieve, P.; Vidic, R.; Pollak, J.; Llewellyn, G.  
647 T.; Abad, J.; Simon, C. Water Resource Impacts during Unconventional Shale Gas  
648 Development: The Pennsylvania Experience. *Int. J. Coal Geol.* **2014**, *126*, 140–156.  
649 <https://doi.org/10.1016/j.coal.2013.12.017>.
- 650 (6) Patterson, L. A.; Konschnik, K. E.; Wiseman, H.; Fargione, J.; Maloney, K. O.; Kiesecker,  
651 J.; Nicot, J.-P.; Baruch-Mordo, S.; Entrekin, S.; Trainor, A.; Saiers, J. E. Unconventional  
652 Oil and Gas Spills: Risks, Mitigation Priorities, and State Reporting Requirements.  
653 *Environ. Sci. Technol.* **2017**, *51* (5), 2563–2573. <https://doi.org/10.1021/acs.est.6b05749>.
- 654 (7) Lackey, G.; Rajaram, H.; Bolander, J.; Sherwood, O. A.; Ryan, J. N.; Shih, C. Y.; Bromhal,  
655 G. S.; Dilmore, R. M. Public Data from Three US States Provide New Insights into Well  
656 Integrity. *Proc. Natl. Acad. Sci. U.S.A.* **2021**, *118* (14), e2013894118.  
657 <https://doi.org/10.1073/pnas.2013894118>.
- 658 (8) Osborn, S. G.; Vengosh, A.; Warner, N. R.; Jackson, R. B. Methane Contamination of  
659 Drinking Water Accompanying Gas-Well Drilling and Hydraulic Fracturing. *Proc. Natl.*  
660 *Acad. Sci. U.S.A.* **2011**, *108* (20), 8172–8176. <https://doi.org/10.1073/pnas.1100682108>.
- 661 (9) Jackson, R. B.; Vengosh, A.; Darrah, T. H.; Warner, N. R.; Down, A.; Poreda, R. J.; Osborn,  
662 S. G.; Zhao, K.; Karr, J. D. Increased Stray Gas Abundance in a Subset of Drinking Water  
663 Wells near Marcellus Shale Gas Extraction. *Proc. Natl. Acad. Sci. U.S.A.* **2013**, *110* (28),  
664 11250–11255. <https://doi.org/10.1073/pnas.1221635110>.
- 665 (10) Llewellyn, G. T.; Dorman, F.; Westland, J. L.; Yoxtheimer, D.; Grieve, P.; Sowers, T.;  
666 Humston-Fulmer, E.; Brantley, S. L. Evaluating a Groundwater Supply Contamination  
667 Incident Attributed to Marcellus Shale Gas Development. *Proc. Natl. Acad. Sci. U.S.A.*  
668 **2015**, *112* (20), 6325–6330. <https://doi.org/10.1073/pnas.1420279112>.
- 669 (11) Siegel, D. I.; Azzolina, N. A.; Smith, B. J.; Perry, A. E.; Bothun, R. L. Methane  
670 Concentrations in Water Wells Unrelated to Proximity to Existing Oil and Gas Wells in  
671 Northeastern Pennsylvania. *Environ. Sci. Technol.* **2015**, *49* (7), 4106–4112.  
672 <https://doi.org/10.1021/es505775c>.
- 673 (12) Barth-Naftilan, E.; Sohng, J.; Saiers, J. E. Methane in Groundwater before, during, and after  
674 Hydraulic Fracturing of the Marcellus Shale. *Proc. Natl. Acad. Sci. U.S.A.* **2018**, *115* (27),  
675 6970–6975. <https://doi.org/10.1073/pnas.1720898115>.
- 676 (13) Woda, J.; Wen, T.; Oakley, D.; Yoxtheimer, D.; Engelder, T.; Castro, M. C.; Brantley, S.  
677 Detecting and Explaining Why Aquifers Occasionally Become Degraded near

- 678       Hydraulically Fractured Shale Gas Wells. *Proc. Natl. Acad. Sci. U.S.A.* **2018**, *115*,  
679       201809013. <https://doi.org/10.1073/pnas.1809013115>.
- 680 (14) Li, Y.; Thelemaque, N. A.; Siegel, H. G.; Clark, C. J.; Ryan, E. C.; Brenneis, R. J.;  
681       Gutchess, K. M.; Soriano, M. A.; Xiong, B.; Deziel, N. C.; Sainers, J. E.; Plata, D. L.  
682       Groundwater Methane in Northeastern Pennsylvania Attributable to Thermogenic Sources  
683       and Hydrogeomorphologic Migration Pathways. *Environ. Sci. Technol.* **2021**, *55* (24),  
684       16413–16422. <https://doi.org/10.1021/acs.est.1c05272>.
- 685 (15) Maloney, K. O.; Baruch-Mordo, S.; Patterson, L. A.; Nicot, J.-P.; Entekin, S. A.; Fargione,  
686       J. E.; Kiesecker, J. M.; Konschnik, K. E.; Ryan, J. N.; Trainor, A. M.; Sainers, J. E.;  
687       Wiseman, H. J. Unconventional Oil and Gas Spills: Materials, Volumes, and Risks to  
688       Surface Waters in Four States of the U.S. *Sci. Total Environ.* **2017**, *581–582*, 369–377.  
689       <https://doi.org/10.1016/j.scitotenv.2016.12.142>.
- 690 (16) Chapman, E. C.; Capo, R. C.; Stewart, B. W.; Kirby, C. S.; Hammack, R. W.; Schroeder, K.  
691       T.; Edenborn, H. M. Geochemical and Strontium Isotope Characterization of Produced  
692       Waters from Marcellus Shale Natural Gas Extraction. *Environ. Sci. Technol.* **2012**, *46* (6),  
693       3545–3553. <https://doi.org/10.1021/es204005g>.
- 694 (17) Haluszczak, L. O.; Rose, A. W.; Kump, L. R. Geochemical Evaluation of Flowback Brine  
695       from Marcellus Gas Wells in Pennsylvania, USA. *Appl. Geochem.* **2013**, *28*, 55–61.  
696       <https://doi.org/10.1016/j.apgeochem.2012.10.002>.
- 697 (18) Vengosh, A.; Jackson, R. B.; Warner, N.; Darrah, T. H.; Kondash, A. A Critical Review of  
698       the Risks to Water Resources from Unconventional Shale Gas Development and Hydraulic  
699       Fracturing in the United States. *Environ. Sci. Technol.* **2014**, *48* (15), 8334–8348.  
700       <https://doi.org/10.1021/es405118y>.
- 701 (19) Kondash, A. J.; Albright, E.; Vengosh, A. Quantity of Flowback and Produced Waters from  
702       Unconventional Oil and Gas Exploration. *Sci. Total Environ.* **2017**, *574*, 314–321.  
703       <https://doi.org/10.1016/j.scitotenv.2016.09.069>.
- 704 (20) Barbot, E.; Vidic, N. S.; Gregory, K. B.; Vidic, R. D. Spatial and Temporal Correlation of  
705       Water Quality Parameters of Produced Waters from Devonian-Age Shale Following  
706       Hydraulic Fracturing. *Environ. Sci. Technol.* **2013**, *47* (6), 2562–2569.  
707       <https://doi.org/10.1021/es304638h>.
- 708 (21) Lauer, N. E.; Harkness, J. S.; Vengosh, A. Brine Spills Associated with Unconventional Oil  
709       Development in North Dakota. *Environ. Sci. Technol.* **2016**, *50* (10), 5389–5397.  
710       <https://doi.org/10.1021/acs.est.5b06349>.
- 711 (22) Orem, W.; Varonka, M.; Crosby, L.; Haase, K.; Loftin, K.; Hladik, M.; Akob, D. M.; Tatu,  
712       C.; Mumford, A.; Jaeschke, J.; Bates, A.; Schell, T.; Cozzarelli, I. Organic Geochemistry  
713       and Toxicology of a Stream Impacted by Unconventional Oil and Gas Wastewater Disposal  
714       Operations. *Appl. Geochem.* **2017**, *80*, 155–167.  
715       <https://doi.org/10.1016/j.apgeochem.2017.02.016>.
- 716 (23) Wen, T.; Niu, X.; Gonzales, M.; Zheng, G.; Li, Z.; Brantley, S. L. Big Groundwater Data  
717       Sets Reveal Possible Rare Contamination Amid Otherwise Improved Water Quality for  
718       Some Analytes in a Region of Marcellus Shale Development. *Environ. Sci. Technol.* **2018**,  
719       *52* (12), 7149–7159. <https://doi.org/10.1021/acs.est.8b01123>.
- 720 (24) Warner, N. R.; Christie, C. A.; Jackson, R. B.; Vengosh, A. Impacts of Shale Gas  
721       Wastewater Disposal on Water Quality in Western Pennsylvania. *Environ. Sci. Technol.*  
722       **2013**, *47* (20), 11849–11857. <https://doi.org/10.1021/es402165b>.

- 723 (25) Akob, D. M.; Mumford, A. C.; Orem, W.; Engle, M. A.; Klinges, J. G.; Kent, D. B.;  
724 Cozzarelli, I. M. Wastewater Disposal from Unconventional Oil and Gas Development  
725 Degrades Stream Quality at a West Virginia Injection Facility. *Environ. Sci. Technol.* **2016**,  
726 *50* (11), 5517–5525. <https://doi.org/10.1021/acs.est.6b00428>.
- 727 (26) Cozzarelli, I. M.; Skalak, K. J.; Kent, D. B.; Engle, M. A.; Benthem, A.; Mumford, A. C.;  
728 Haase, K.; Farag, A.; Harper, D.; Nagel, S. C.; Iwanowicz, L. R.; Orem, W. H.; Akob, D.  
729 M.; Jaeschke, J. B.; Galloway, J.; Kohler, M.; Stoliker, D. L.; Jolly, G. D. Environmental  
730 Signatures and Effects of an Oil and Gas Wastewater Spill in the Williston Basin, North  
731 Dakota. *Sci. Total Environ.* **2017**, *579*, 1781–1793.  
732 <https://doi.org/10.1016/j.scitotenv.2016.11.157>.
- 733 (27) Bonetti, P.; Leuz, C.; Michelon, G. Large-Sample Evidence on the Impact of  
734 Unconventional Oil and Gas Development on Surface Waters. *Science* **2021**, *373* (6557),  
735 896–902. <https://doi.org/10.1126/science.aaz2185>.
- 736 (28) Shaheen, S. W.; Wen, T.; Herman, A.; Brantley, S. L. Geochemical Evidence of Potential  
737 Groundwater Contamination with Human Health Risks Where Hydraulic Fracturing  
738 Overlaps with Extensive Legacy Hydrocarbon Extraction. *Environ. Sci. Technol.* **2022**, *56*  
739 (14), 10010–10019. <https://doi.org/10.1021/acs.est.2c00001>.
- 740 (29) Sherwood, O. A.; Rogers, J. D.; Lackey, G.; Burke, T. L.; Osborn, S. G.; Ryan, J. N.  
741 Groundwater Methane in Relation to Oil and Gas Development and Shallow Coal Seams in  
742 the Denver-Julesburg Basin of Colorado. *Proc. Natl. Acad. Sci. U.S.A.* **2016**, *113* (30),  
743 8391–8396. <https://doi.org/10.1073/pnas.1523267113>.
- 744 (30) Siegel, H. G.; Soriano, M. A.; Clark, C. J.; Johnson, N. P.; Wulsin, H. G.; Deziel, N. C.;  
745 Plata, D. L.; Darrah, T. H.; Saiers, J. E. Natural and Anthropogenic Processes Affecting  
746 Domestic Groundwater Quality within the Northwestern Appalachian Basin. *Environ. Sci.*  
747 *Technol.* **2022**, *56* (19), 13761–13773. <https://doi.org/10.1021/acs.est.2c04011>.
- 748 (31) Kreuzer, R. L.; Darrah, T. H.; Grove, B. S.; Moore, M. T.; Warner, N. R.; Eymold, W. K.;  
749 Whyte, C. J.; Mitra, G.; Jackson, R. B.; Vengosh, A.; Poreda, R. J. Structural and  
750 Hydrogeological Controls on Hydrocarbon and Brine Migration into Drinking Water  
751 Aquifers in Southern New York. *Groundwater* **2018**, *56* (2), 225–244.  
752 <https://doi.org/10.1111/gwat.12638>.
- 753 (32) Warner, N.; Jackson, R.; Darrah, T.; Osborn, S.; Down, A.; Zhao, K.; White, A.; Vengosh,  
754 A. Geochemical Evidence for Possible Natural Migration of Marcellus Formation Brine to  
755 Shallow Aquifers in Pennsylvania. *Proc. Natl. Acad. Sci. U.S.A.* **2012**, *109*, 11961–11966.  
756 <https://doi.org/10.1073/pnas.1121181109>.
- 757 (33) Whittemore, D. O. Fate and Identification of Oil-Brine Contamination in Different  
758 Hydrogeologic Settings. *Appl. Geochem.* **2007**, *22* (10), 2099–2114.  
759 <https://doi.org/10.1016/j.apgeochem.2007.04.002>.
- 760 (34) McMahan, P. B.; Thomas, J. C.; Crawford, J. T.; Dornblaser, M. M.; Hunt, A. G. Methane  
761 in Groundwater from a Leaking Gas Well, Piceance Basin, Colorado, USA. *Sci. Total*  
762 *Environ.* **2018**, *634*, 791–801. <https://doi.org/10.1016/j.scitotenv.2018.03.371>.
- 763 (35) Jasechko, S.; Perrone, D. Hydraulic Fracturing near Domestic Groundwater Wells. *Proc.*  
764 *Natl. Acad. Sci. U.S.A.* **2017**, *114* (50), 13138–13143.  
765 <https://doi.org/10.1073/pnas.1701682114>.
- 766 (36) Soriano Jr., M. A.; Warren, J. L.; Clark, C. J.; Johnson, N. P.; Siegel, H. G.; Deziel, N. C.;  
767 Saiers, J. E. Social Vulnerability and Groundwater Vulnerability to Contamination From



- 768 Unconventional Hydrocarbon Extraction in the Appalachian Basin. *GeoHealth* **2023**, 7 (4),  
769 e2022GH000758. <https://doi.org/10.1029/2022GH000758>.
- 770 (37) Deziel, N. C.; Brokovich, E.; Grotto, I.; Clark, C. J.; Barnett-Itzhaki, Z.; Broday, D.; Agay-  
771 Shay, K. Unconventional Oil and Gas Development and Health Outcomes: A Scoping  
772 Review of the Epidemiological Research. *Environ. Res.* **2020**, 182, 109124.  
773 <https://doi.org/10.1016/j.envres.2020.109124>.
- 774 (38) Clark, C. J.; Xiong, B.; Soriano, M. A.; Gutchess, K.; Siegel, H. G.; Ryan, E. C.; Johnson,  
775 N. P.; Cassell, K.; Elliott, E. G.; Li, Y.; Cox, A. J.; Bugher, N.; Glist, L.; Brenneis, R. J.;  
776 Sorrentino, K. M.; Plano, J.; Ma, X.; Warren, J. L.; Plata, D. L.; Saiers, J. E.; Deziel, N. C.  
777 Assessing Unconventional Oil and Gas Exposure in the Appalachian Basin: Comparison of  
778 Exposure Surrogates and Residential Drinking Water Measurements. *Environ. Sci. Technol.*  
779 **2022**, 56 (2), 1091–1103. <https://doi.org/10.1021/acs.est.1c05081>.
- 780 (39) Clark, C. J.; Johnson, N. P.; Soriano, M.; Warren, J. L.; Sorrentino, K. M.; Kadan, -Lottick  
781 Nina S.; Saiers, J. E.; Ma, X.; Deziel, N. C. Unconventional Oil and Gas Development  
782 Exposure and Risk of Childhood Acute Lymphoblastic Leukemia: A Case–Control Study in  
783 Pennsylvania, 2009–2017. *Environ. Health Perspect.* **2022**, 130 (8), 087001.  
784 <https://doi.org/10.1289/EHP11092>.
- 785 (40) Hill, E. L.; Ma, L. Drinking Water, Fracking, and Infant Health. *J. Health Econ.* **2022**, 82,  
786 102595. <https://doi.org/10.1016/j.jhealeco.2022.102595>.
- 787 (41) Wen, T.; Liu, M.; Woda, J.; Zheng, G.; Brantley, S. L. Detecting Anomalous Methane in  
788 Groundwater within Hydrocarbon Production Areas across the United States. *Water Res.*  
789 **2021**, 117236. <https://doi.org/10.1016/j.watres.2021.117236>.
- 790 (42) Brantley, S. L.; Vidic, R. D.; Brasier, K.; Yoxtheimer, D.; Pollak, J.; Wilderman, C.; Wen,  
791 T. Engaging over Data on Fracking and Water Quality. *Science* **2018**, 359 (6374), 395–397.  
792 <https://doi.org/10.1126/science.aan6520>.
- 793 (43) Siegel, D. I.; Smith, B.; Perry, E.; Bothun, R.; Hollingsworth, M. Pre-Drilling Water-  
794 Quality Data of Groundwater Prior to Shale Gas Drilling in the Appalachian Basin:  
795 Analysis of the Chesapeake Energy Corporation Dataset. *Appl. Geochem.* **2015**, 63, 37–57.  
796 <https://doi.org/10.1016/j.apgeochem.2015.06.013>.
- 797 (44) Brantley, S. L. Shale Network Database. *Consortium for Universities for the Advancement*  
798 *of Hydrologic Sciences, Inc. (CUAHSI)* **2011**. [https://doi.org/10.4211/his-data-](https://doi.org/10.4211/his-data-shalenetwork)  
799 [shalenetwork](https://doi.org/10.4211/his-data-shalenetwork).
- 800 (45) SkyTruth. *SkyTruth Releases Map of Drilling-Related Impoundments Across PA*. SkyTruth.  
801 <https://skytruth.org/2014/10/pa-drilling-impoundments-2005-2013/> (accessed 2023-01-13).
- 802 (46) SkyTruth. SkyTruth FrackFinder PA 2005-2013 Methodology, 2014.  
803 [https://skytruth.org/wp-content/uploads/2014/10/SkyTruth\\_FrackFinder\\_PA\\_2005-](https://skytruth.org/wp-content/uploads/2014/10/SkyTruth_FrackFinder_PA_2005-2013_Methodology.pdf)  
804 [2013\\_Methodology.pdf](https://skytruth.org/wp-content/uploads/2014/10/SkyTruth_FrackFinder_PA_2005-2013_Methodology.pdf).
- 805 (47) *Title 25 Pa. Code Chapter 78a - Unconventional Wells*; Vol. 78a.56(d).  
806 [https://www.pacodeandbulletin.gov/Display/pacode?file=/secure/pacode/data/025/chapter7](https://www.pacodeandbulletin.gov/Display/pacode?file=/secure/pacode/data/025/chapter78/chap78toc.html)  
807 [8/chap78toc.html](https://www.pacodeandbulletin.gov/Display/pacode?file=/secure/pacode/data/025/chapter78/chap78toc.html) (accessed 2023-09-21).
- 808 (48) Soriano, M. A.; Siegel, H. G.; Gutchess, K. M.; Clark, C. J.; Li, Y.; Xiong, B.; Plata, D. L.;  
809 Deziel, N. C.; Saiers, J. E. Evaluating Domestic Well Vulnerability to Contamination From  
810 Unconventional Oil and Gas Development Sites. *Water Resour. Res.* **2020**, 56 (10),  
811 e2020WR028005. <https://doi.org/10.1029/2020WR028005>.
- 812 (49) Blondes, M. S.; Gans, K. D.; Engle, M. A.; Kharaka, Y. K.; Reidy, M. E.; Saraswathula, V.;  
813 Thordsen, J. J.; Rowan, E. L.; Morrissey, E. A. U.S. Geological Survey National Produced

- 814 Waters Geochemical Database v2.3 [Data Set]. *U.S. Geological Survey Data Release* **2019**.  
815 <https://doi.org/10.5066/F7J964W8>.
- 816 (50) Niu, X.; Wendt, A.; Li, Z.; Agarwal, A.; Xue, L.; Gonzales, M.; Brantley, S. L. Detecting  
817 the Effects of Coal Mining, Acid Rain, and Natural Gas Extraction in Appalachian Basin  
818 Streams in Pennsylvania (USA) through Analysis of Barium and Sulfate Concentrations.  
819 *Environ. Geochem. Health* **2018**, *40* (2), 865–885. [https://doi.org/10.1007/s10653-017-](https://doi.org/10.1007/s10653-017-0031-6)  
820 [0031-6](https://doi.org/10.1007/s10653-017-0031-6).
- 821 (51) Musgrove, M. The Occurrence and Distribution of Strontium in U.S. Groundwater. *Appl.*  
822 *Geochem.* **2021**, *126*, 104867. <https://doi.org/10.1016/j.apgeochem.2020.104867>.
- 823 (52) Epuna, F.; Shaheen, S. W.; Wen, T. Road Salting and Natural Brine Migration Revealed as  
824 Major Sources of Groundwater Contamination across Regions of Northern Appalachia with  
825 and without Unconventional Oil and Gas Development. *Water Res.* **2022**, *225*, 119128.  
826 <https://doi.org/10.1016/j.watres.2022.119128>.
- 827 (53) Titler, R. V.; Curry, P. *Chemical Analysis of Major Constituents and Trace Contaminants*  
828 *of Rock Salt*; Pennsylvania Department of Environmental Protection, Bureau of Water  
829 Standards and Facility Regulation, 2011.
- 830 (54) PADEP. *Oil and Gas Compliance - Report Extracts*. PA Department of Environmental  
831 Protection. <https://greenport.pa.gov/ReportExtracts/OG/OilComplianceReport> (accessed  
832 2023-01-12).
- 833 (55) Risser, D. W.; Conger, R. W.; Ulrich; Asmussen, M. P. *Estimates of Ground-Water*  
834 *Recharge Based on Streamflow-Hydrograph Methods: Pennsylvania*; U.S. Geological  
835 Survey Open-File Report; 2005–1333; 2005; p 30.
- 836 (56) Soriano, M. A.; Deziel, N. C.; Saiers, J. E. Regional Scale Assessment of Shallow  
837 Groundwater Vulnerability to Contamination from Unconventional Hydrocarbon  
838 Extraction. *Environ. Sci. Technol.* **2022**, *56* (17), 12126–12136.  
839 <https://doi.org/10.1021/acs.est.2c00470>.
- 840 (57) Agarwal, A.; Wen, T.; Chen, A.; Zhang, A. Y.; Niu, X.; Zhan, X.; Xue, L.; Brantley, S. L.  
841 Assessing Contamination of Stream Networks near Shale Gas Development Using a New  
842 Geospatial Tool. *Environ. Sci. Technol.* **2020**, *54* (14), 8632–8639.  
843 <https://doi.org/10.1021/acs.est.9b06761>.
- 844 (58) Pennsylvania Department of Environmental Protection. Pennsylvania DEP Fines Range  
845 Resources \$4.15 Million for Violating Environmental Regulations, 2014.  
846 [https://www.prnewswire.com/news-releases/pennsylvania-dep-fines-range-resources-415-](https://www.prnewswire.com/news-releases/pennsylvania-dep-fines-range-resources-415-million-for-violating-environmental-regulations-275638111.html)  
847 [million-for-violating-environmental-regulations-275638111.html](https://www.prnewswire.com/news-releases/pennsylvania-dep-fines-range-resources-415-million-for-violating-environmental-regulations-275638111.html) (accessed 2023-09-04).
- 848 (59) U.S.E.P.A. EPA’s Hydraulic Fracturing Study: Retrospective Case Studies, 2016.
- 849 (60) Llewellyn, G. T. Evidence and Mechanisms for Appalachian Basin Brine Migration into  
850 Shallow Aquifers in NE Pennsylvania, USA. *Hydrogeol. J.* **2014**, *22* (5), 1055–1066.  
851 <https://doi.org/10.1007/s10040-014-1125-1>.
- 852 (61) Cravotta, C. A. Dissolved Metals and Associated Constituents in Abandoned Coal-Mine  
853 Discharges, Pennsylvania, USA. Part 2: Geochemical Controls on Constituent  
854 Concentrations. *Appl. Geochem.* **2008**, *23* (2), 203–226.  
855 <https://doi.org/10.1016/j.apgeochem.2007.10.003>.
- 856 (62) Scanlon, B. R.; Reedy, R. C.; Xu, P.; Engle, M.; Nicot, J. P.; Yoxtheimer, D.; Yang, Q.;  
857 Ikonnikova, S. Can We Beneficially Reuse Produced Water from Oil and Gas Extraction in  
858 the U.S.? *Sci. Total Environ.* **2020**, *717*, 137085.  
859 <https://doi.org/10.1016/j.scitotenv.2020.137085>.

860 (63) Scanlon, B. R.; Ikonnikova, S.; Yang, Q.; Reedy, R. C. Will Water Issues Constrain Oil and  
861 Gas Production in the United States? *Environ. Sci. Technol.* **2020**, *54* (6), 3510–3519.  
862 <https://doi.org/10.1021/acs.est.9b06390>.  
863