



Peer review status:

This is a non-peer-reviewed preprint submitted to EarthArXiv.

# **Ecological Ramifications of Phosphate Extraction and Use: An In-Depth Analysis of the Environmental Impacts and Systemic Consequences**

Ahmed Bekari

[ahmedbekari007@gmail.com](mailto:ahmedbekari007@gmail.com)

## **Abstract :**

Phosphorus (P) is a finite resource located within certain geologic reserves around the world.

Morocco and the Western Sahara together account for roughly 75% of reserves, raising questions of how to more sustainably use this precious resource. Phosphorus is mined from rock and eventually converted into usable fertilizer, which is applied to croplands. This study aims, by adopting a global lens to explore the environmental consequences stemming from phosphate extraction and transportation, emphasizing the systemic impacts of these processes.

## **Author's Note :**

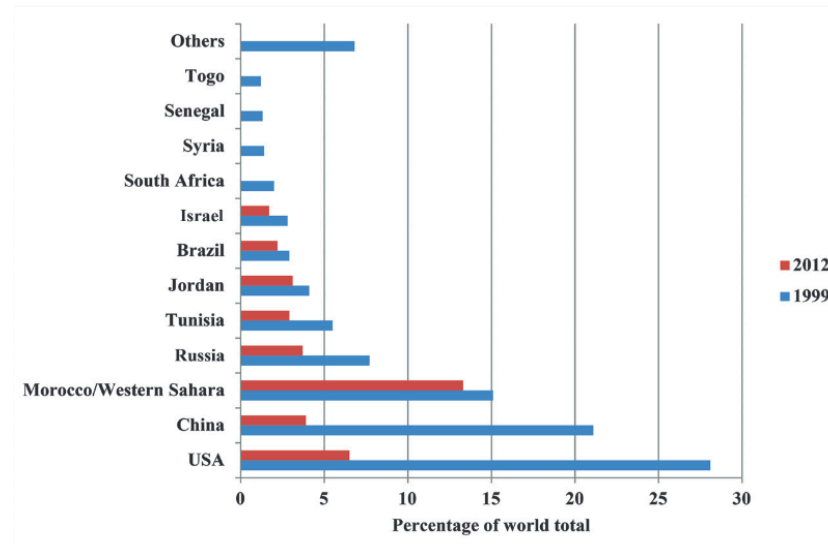
My personal interest in this stemmed from a discussion with one of the members of my nonprofit. She had mentioned that despite living in Morocco, we never thought of phosphorus as a danger for our environment, and that sparked an interest in this subject. I am particularly interested in anything in the sustainability realm that deals with the environment or climate change.

**Keywords** : Phosphorus, Environment, Mining, Transportation.

## 1. Introduction

Phosphate rock (PR) is a naturally occurring phosphorus- (P) rich rock with a P content of about 5%–13% ([Jasinski 2009](#); [Cordell And White 2011](#)). Global food production is dependent on PR because P is a limiting factor in crop production. PR is mainly used in agriculture as phosphate ( $\text{PO}_4^{3-}$ ) fertilizer and animal feed, but is also used in the manufacture of detergents, food additives, and metal stabilization ([Ibrahim et al. 2010](#)). Increasing population growth and the need to increase food production may also be responsible for the increasing demand for  $\text{PO}_4^{3-}$  fertilizers. Based on estimates of current PR reserves, studies indicate that a depletion of global reserves is not likely to occur within this century ([Heckenmüller et al. 2014](#)). The International Fertilizer Development Center (IFDC) has also estimated that world PR reserves will be able to supply needed PR far into the future ([VanKauwenbergh 2010](#)). However, to prolong the life of this limited resource and prevent environmental pollution, the efficient use of P in crop production is important ([Selles et al. 2011](#)).

**Fig. 1.** World phosphate rock production, 1999. (Source: [FAO 2004](#); [Mew 2000](#)). World phosphate rock production, 2012 (Source: [Hecken Müller et al. 2014](#); [Dennis 2013](#); [ICL 2013](#)).



Although very important in food production, water-soluble  $\text{PO}_4^{3-}$  fertilizers manufactured from PR are susceptible to leaching and can contribute to P loading of water bodies, leading to eutrophication of surface waters and pollution of groundwater (Leone et al. 2008). Direct application of PR reduces leaching losses of P and is cost effective. However, direct application of PR has limitations such as low solubility and low reactivity, which reduces its agronomic efficiency (Husnain et al. 2014). This has discouraged its adoption by many farmers despite efforts to improve its effectiveness. The application of  $\text{PO}_4^{3-}$  fertilizers is also an environmental concern because of the presence of toxic metals/metalloids and radionuclides in PR. The use of  $\text{PO}_4^{3-}$  fertilizers with toxic metals/metalloids such as arsenic (As), cadmium (Cd), and lead (Pb) can contaminate the food chain and greatly affect public health. Toxic Metals/metalloids are considered grave pollutants due to their persistence, toxicity, and non-degradability in the environment. Cd can cause demineralization of bone and can harm the kidneys (Bernard 2008). It has been classified as carcinogenic because it can cause cancer (Ferreccio et al. 2013). Pb can cause damage to the central nervous system and trigger learning problems in children (Hou et al. 2013). Radionuclides like uranium (U) may cause lung cancer and tumors of the lymphatic and hemato-poietic tissues (USEPA 2000; Murakami and Oki 2015). It is important to know the potential of PR to increase the levels of these hazardous elements in the environment. The debate on the future availability of PR and controversy regarding its contamination potential has raised several questions through the years.

## 2. Environmental impacts of the uses and mining of PR

### 2.1. Water pollution

Leaching losses of applied P from fertilized farmlands can lead to pollution of groundwater, while runoff could pollute surface waters and lead to eutrophication. Eutrophication is the excessive loading of nutrients in surface waters resulting in excessive growth and accumulation of green and blue-green algae (Hodgkin and Hamilton 1993). This can lead to reduced oxygen levels in the water which causes declines in population of aquatic organisms, reduced biodiversity, and imbalanced trophic food webs (Kennish and de Jonge 2011; Scavia et al. 2014). P is generally the limiting nutrient for algal growth, and leaching of P from application of  $\text{PO}_4^{3-}$  fertilizers to sandy soils is the main source of P (Hodgkin and Hamilton 1993). Although both N and P are major plant nutrients that can affect water quality, whole-lake experiments have shown that regulation of P alone can effectively mitigate eutrophication (Wang and Wang 2009). PR can increase P loading in surface waters through leaching of P from waste accumulated from mining operations, leaching and runoff from  $\text{PO}_4^{3-}$  -fertilized soils, leaching of P from farmlands receiving animal manure, and use of P as detergents. Vandenhove et al. (2015) determined the effect of  $\text{PO}_4^{3-}$  mining on aquatic ecosystems in five countries (Belgium, Spain, Syria, Egypt, Brazil). They predicted that the aquatic ecosystems around the  $\text{PO}_4^{3-}$  fertilizer plants in Belgium, Spain, and Brazil are potentially most at risk. Bulut and Aksoy (2008) investigated the role of fertilizer application rate on P transport and loading to Lake Uluabat in Turkey.

Their results showed that, when the amount of fertilizer applied to agricultural lands was doubled, the P load to Lake Uluabat increased by 32%. P transported to Lake Uluabat declined by about 6%, 10%, and 16% when the fertilizer usage on the agricultural fields decreased by 20%, 30%, and 50%, respectively. Animal manure is a source of P because the animals are given feed containing P from PR.

PO<sub>4</sub><sup>3-</sup> fertilizers may also increase the mobility of As in As-contaminated soils, leading to As contamination of groundwater. P and As are both Group V elements with similar chemical properties and they both compete for exchange sites in the soil. [Castlehouse et al. \(2010\)](#) studied As mobilization using a natural As-contaminated soil and suggested that P may increase As mobility by increasing the concentration of non-specifically sorbed As. In another study, addition of PO<sub>4</sub><sup>3-</sup> fertilizers significantly increased the amount of As leached from lead arsenate-contaminated soils ([Davenport and Peryea 1991](#)). Because P addition leads to As mobilization in the soil, PO<sub>4</sub><sup>3-</sup> fertilizers have been used to increase As availability to As hyperaccumulating plants used in phytoremediation of As-contaminated soils ([X. Cao et al. 2003](#); [Fayiga and Ma 2006](#); [Lessl et al. 2014](#)). In the absence of As hyperaccumulators, compost and PO<sub>4</sub><sup>3-</sup> amendments increased As leaching in chromated copper arsenate- (CCA) contaminated soils ([X. Cao et al. 2003](#)).

Another important use of PO<sub>4</sub><sup>3-</sup> compounds is in making detergents. The main P compound in detergents is sodium tripolyphosphate (Na<sub>5</sub> P<sub>3</sub> O<sub>10</sub>; STPP) which is prepared from H<sub>3</sub> PO<sub>4</sub> by neutralization with soda ash (sodium oxide) to form sodium hydrogen phosphates ([Glennie et al. 2002](#)). PO<sub>4</sub><sup>3-</sup> in the form of STPP is the most commonly used builder which forms complexes with Ca<sup>2+</sup> and Mg<sup>2+</sup> ions, creating favorable conditions for detergent action by reducing water hardness ([Kundu et al. 2015](#)). It has been reported that where STPP is used as builder in household detergents, it contributes up to 50% of soluble P in municipal wastewater; therefore a reduction in the use of PO<sub>4</sub><sup>3-</sup>-based detergents is expected to reduce P loading of surface water bodies ([European Commission 2015](#)). Many suggested alternatives such as sodium citrates, ethylenediamine-tetraacetic acid (EDTA), and nitrilotriacetic acid (NTA) have been found to be less effective and less cost-efficient than PO<sub>4</sub><sup>3-</sup> in detergents ([Kundu et al. 2015](#)). The only alternative that may be effective is Zeolite A, which is comparatively inert and derived from aluminium oxide ([Landbank 1994](#)). Measures to reduce the use of STPP-based detergents in the European Union (EU) include the introduction of laws or voluntary agreements to change to Zeolite A as the builder for household laundry detergents. STPP consumption has decreased substantially since the early 1980s due to bans or restrictions on its use in many developed countries ([European Commission 2015](#)).

In summary, waste from PR industries and PO<sub>4</sub><sup>3-</sup> fertilizers has contributed to water pollution which could lead to eutrophication. It has been shown that reducing the amount of PO<sub>4</sub><sup>3-</sup> fertilizer applied to soils leads to a reduction of P loading in surface waters. Even though the use of PR products like PO<sub>4</sub><sup>3-</sup> fertilizers and detergents has led to the pollution of surface and groundwater bodies, regulation of these products has reduced the extent of environmental contamination.

## 2.2 The case study region: Eutrophication in Nigeria

Eutrophication is a threat to the water quality of rivers, lakes and reservoirs, hence their classification into oligotrophic, mesotrophic and eutrophic based on level of eutrophication. Recently in Nigeria eutrophication rates in water bodies have increased dramatically as a result of alterations in nutrient cycles related to land use. The increasing demand for food and food insecurity which has the increased use of fertilizers on farmland has been pointed out as contributing over 80% of eutrophication of water bodies worldwide. [Adeyemo et al](#) observes that land use around riverine areas in Nigeria is predominantly for farmland and this could be explanation for the high level of phosphate from run-off during the rainy season in Ibadan River. [Nweze and Onyishi](#) indicated that indiscriminate use of fertilizers around Nike Lake in Enugu, Nigeria is the major source of pollution causing Cyanophycean blooms and ecological disaster. Since the population boom in Nigeria in the 80's the use of N.P.K fertilizer has increased and led to the treat observed in water bodies located close to these farmland. When washed as a result of erosion, flooding increased the nutrient load of these water bodies. Other anthropogenic activities that can cause eutrophication in water bodies include increased silt, deforestation, lumbering activities and other land use perturbation.

## 2.3. Deforestation

A new WWF report, "Extracted Forests: Unearthing the role of mining-related deforestation as a driver of global deforestation", sheds light on how the impacts of mining, currently the fourth largest driver of deforestation, increases significantly, affecting up to 1/3 of the world's forest ecosystems, when indirect impacts are taken into account.

These indirect impacts, which include mining related infrastructure, settlements, agriculture through settlement, water and soil contamination and illegal logging, explain why demand by just six countries (often located far from the mines) are responsible for 51% of mining-related deforestation. 85% of the deforestation footprint of the EU, for instance, is located outside of the region.

More than 80% of direct mining-related deforestation takes place in just 10 countries, with tropical rainforests suffering the most damage – this most affected biome contains 29% of mining sites but 62% of mining-related deforestation.

Negative consequences for wildlife and ecosystems have already been recorded: the numbers of Indochinese tigers in the Greater Mekong region have decreased drastically due to habitat fragmentation from infrastructure development, and gold mining in the Amazon Basin has led to increased mercury levels in endangered species such as the Tucuxi river dolphin.

Mining activities have seen an alarming acceleration – more than 1/3 of the mining-related deforestation seen in the last 20 years occurred in just the last five years – and are tipped to increase in coming years.

Bad mining practices do not only lead to negative environmental impacts, they also result in negative social impacts and human rights violations that can only worsen if not addressed.

The report highlights how the rights of Indigenous communities are at risk in certain regions where mining expansion or processing of ores destroy ancestral territories, affecting and violating community interests.

By highlighting at-risk regions and both direct and indirect impacts that mining can have on forest ecosystems, the report provides governments and companies with an in-depth understanding of the challenges that need to be tackled.

“Metals are important for the development of human civilization and the life we live today. But the footprint left by the extraction of these commodities has a heavy toll on ecosystems,” said Tobias Kind-Rieper, Global Lead Mining & Metals, WWF Germany.

## **2.5. Greenhouse Gas Emissions**

It is also important to consider the impact of land use change in the context of greenhouse gas emissions. The destruction of vegetation and soils when land is cleared for mining results in the release of carbon dioxide and other greenhouse gases. Another important consideration relates to the quantity of greenhouse gases released per unit mass of mined material, as some less concentrated mineral deposits require proportionally higher energy usage. For example, mining a kilogram of diamond produces around [800,000 kg CO<sub>2</sub>e](#) compared to a kilogram of a highly abundant mineral such as iron which produces only about [2 kg CO<sub>2</sub>e](#).

The creation of products from mined materials uses high amounts of energy throughout the different stages of the production chain and most of this energy is currently sourced from the burning of fossil fuels.

Reducing reliance on fossil fuels in the mining process by electrifying the technology and running it off a green energy grid is a key aim to allow mining to continue along a more sustainable path. Automation of many of the stages of mining is also another vital change that will not only improve safety but also increase efficiency and cut energy costs. However, it will remain difficult to swiftly transition the mining industry into becoming a net zero emitter, and with the short supply of many rare earth metals, it is crucial to reuse and recycle mined materials wherever possible.

## 2.6. Soil pollution and Land occupation

### 2.6.1. Toxic metals/metalloid

The presence of hazardous elements such as toxic metals or metalloids in PR makes it a potential source of contamination, especially on agricultural soils receiving PO<sub>4</sub><sup>3-</sup> fertilizers. There are conflicting reports about the effect of PO<sub>4</sub><sup>3-</sup> fertilizers on the metal content of the soil. Some scientists suggest that application of PO<sub>4</sub><sup>3-</sup> fertilizer does not increase metal content of soils, while others have reported otherwise. [Ogunleye et al. \(2002\)](#) reported that the average concentrations of toxic elements (As, antimony (Sb), Cr, Zn) in PR are not appreciably different from those in agricultural soils, and [da Conceição and Bonotto \(2006\)](#) reported that the toxic metals in PO<sub>4</sub><sup>3-</sup> fertilizers applied to Brazilian crops did not raise their concentration in soil to harmful levels.

This was not the case in the UK, according to a report on the impact of long-term PO<sub>4</sub><sup>3-</sup> fertilizer applications on As concentrations in soil and herbage ([Hartley et al. 2013](#)). PO<sub>4</sub><sup>3-</sup> fertilizer used in the Park Grass Experiment between 1888 and 1947 added a substantial amount of As to the soil, resulting in a near doubling of the total As concentration in the topsoil. It has also been reported that PO<sub>4</sub><sup>3-</sup> fertilizer is a main source of As in areas of Sri Lanka where the population is affected by chronic kidney disease (Jayasumana et al. 2015). Analysis of fertilizers for their As content revealed that the highest amount was in TSP with a mean value of 31 mg·kg<sup>-1</sup> ([Jayasumana et al. 2015](#)), while PR had the next highest concentration (8.56 mg·kg<sup>-1</sup>).

To investigate the effects of long-term applications of PO<sub>4</sub><sup>3-</sup> fertilizers on the solubility of Cd and Zn, three PO<sub>4</sub><sup>3-</sup> fertilizers were applied to soils in a laboratory experiment using a single fertilizer addition equivalent to 15 years of application ([Lambert et al. 2007](#)). Three MAP fertilizers containing various amounts of metals were also applied on cultivated fields for three years at three different rates. PO<sub>4</sub><sup>3-</sup> fertilizer treatments increased the Zn solution concentration in 83% and 53% of treatments in the field and laboratory, respectively, while it increased the concentration of Cd in soil extracts by 87% and 80% of the treatments in field and laboratory experiments, respectively ([Lambert et al. 2007](#)).

These hazardous elements in PR and PO<sub>4</sub><sup>3-</sup> fertilizers may enter the food chain through contaminated food crops and cause severe health problems for the unsuspecting public. It is therefore necessary to protect the public from consumption of food grown with contaminated PO<sub>4</sub><sup>3-</sup> fertilizers. Regulation of PR and PO<sub>4</sub><sup>3-</sup> fertilizers based on their composition is necessary to reduce health risks caused by potential contamination of the food chain. Regulatory bodies have established tolerable limits specifying concentrations of metals allowable in PR or PO<sub>4</sub><sup>3-</sup> fertilizers.

[Gonçalves et al. \(2014\)](#) reported that some countries like the USA, Canada, Australia, Japan, China, Germany, Sweden, and Finland do have legislation to enforce tolerable limits and standards for PO<sub>4</sub><sup>3-</sup> fertilizers. According to the report, the tolerable limit can vary between regions within one country. For example, in the USA the limit for Cd concentration in PO<sub>4</sub><sup>3-</sup> fertilizers is 4 mg·kg<sup>-1</sup> in California but up to 165 mg·kg<sup>-1</sup> in Washington. In other countries, it is about 50 mg·kg<sup>-1</sup> in Switzerland and Finland, 300 mg·kg<sup>-1</sup> in Australia, 8 mg·kg<sup>-1</sup> in China, and as high as 343 mg·kg<sup>-1</sup> in Japan.



Among the hazardous elements in PO<sub>4</sub><sup>3-</sup> fertilizers, Cd is the most widely studied because of its relatively high frequency of occurrence and its toxicity to human health ([FAO 2004](#)). Cd is toxic at very low concentrations and its uptake and accumulation in the food chain makes it a public health concern. Presently, there is no cost effective way of removing Cd from PR ores, but it is possible to produce PO<sub>4</sub><sup>3-</sup> fertilizers with very low Cd concentrations by using PR with a low Cd content ([Van Kauwenbergh 2006](#)). Regulatory limits on the Cd content of PO<sub>4</sub><sup>3-</sup> fertilizers are necessary because fertilization increases the risk of transferring the heavy metal into the food chain. The limits established by the European Commission were based on the reports of an impact assessment study which showed that fertilizers containing 20 mg Cd·kg<sup>-1</sup> P<sub>2</sub>O<sub>5</sub> or less are not expected to result in long-term accumulation in the soil, while fertilizers containing 60 mg Cd·kg<sup>-1</sup> P<sub>2</sub>O<sub>5</sub> could result in long-term accumulation of Cd in the soil ([Roberts 2014](#)). Even though model simulation outcomes indicate that some of the existing fertilizer regulations are not strict enough to prevent significant accumulation of Cd in cropland soils ([Chen et al. 2007](#)), scientific risk assessments in the USA have shown that PO<sub>4</sub><sup>3-</sup> fertilizers containing relatively small amounts of Cd are safe and do not pose any risk to human health ([Roberts 2014](#)).

### 2.6.2. Radionuclides

There are conflicting reports on the effect of PO<sub>4</sub><sup>3-</sup> fertilizers on the radioactivity of receiving soils. It has been reported that the application of PR or soluble PO<sub>4</sub><sup>3-</sup> fertilizers to unfertile fields during farming could raise radioactivity levels in soils ([Akhtar et al. 2005](#); [Mwilo et al. 2007](#)). However, a systematic study of the radionuclide content of various Yugoslavian soils reported no statistical difference between soils unfertilized since 1948 and normal agricultural soils ([Baetsle 1991](#)). This did not agree with a study in Pakistan, where fertilized farmlands had higher activity concentrations of Ra-226 (65%–85%) and K-40 (34%–43%) than barren land ([Akhtar et al. 2012](#)).

The extent of pollution and health risk associated with PR depends on factors such as the geochemical properties of the PR, method of mining and processing, soil characteristics, and environmental conditions ([FAO 2004](#)). Sometimes the composition of PR is different from that of fertilizers produced from it. For example, though high Ra-226 concentrations have been reported in Egyptian PR, the radionuclide content of PO<sub>4</sub><sup>3-</sup> fertilizers produced from Egyptian PR was found to be low ([Hassan et al. 2015](#)). The radionuclide concentration has also been found to vary with fertilizer type. Radiological analysis of PO<sub>4</sub><sup>3-</sup> fertilizers produced from PR showed that the activity concentration of U-238 was much higher in SSP (284 Bq·kg<sup>-1</sup>) and TSP (396 Bq·kg<sup>-1</sup>) than in DAP (<2.03 Bq·kg<sup>-1</sup>) and NPK fertilizer (NPK < 2.03 Bq·kg<sup>-1</sup>) ([Hameed et al. 2014](#)). Another study also measured the radionuclide content of different fertilizers in Egypt; results showed that SSP had the highest Ra-226 while NK fertilizer had the highest Th-232 and K-40 ([Uosif et al. 2014](#)).

Alpha track density can be used to determine the activity of radionuclides because they emit alpha particles. [Kumar and Chauhan \(2014\)](#) measured the alpha track density of plants grown with different fertilizer types. Results showed that the alpha track density of plants grown with PO<sub>4</sub><sup>3-</sup> fertilizers were higher than with organic and urea fertilizers. Studies evaluating the radiological impact of radionuclides in PO<sub>4</sub><sup>3-</sup> fertilizers have reported that the radium equivalent activity is below the permissible limit and does not pose any radiological risk ([Saueia and Mazilli 2006](#); [Khater and Al-Sewaidan 2008](#); [Bituh et al. 2009](#); [Cevik et al. 2010](#); [Hameed et al. 2014](#)).

Though many studies have reported the adverse effects of PO<sub>4</sub><sup>3-</sup> fertilizers on the environment, there is limited data on the effect of direct application of PR on the metal and radionuclide content of the soil, likely due to the low percentage of people involved in direct application.

In summary, although there are hazardous elements in PR, the application of PO<sub>4</sub><sup>3-</sup> fertilizers in farmlands may or may not cause soil pollution depending on several factors. Tolerable limits of toxic metals/metalloids for PO<sub>4</sub><sup>3-</sup> fertilizers have been established in different countries. The radionuclide content of different PO<sub>4</sub><sup>3-</sup> fertilizers varies widely, though PO<sub>4</sub><sup>3-</sup> fertilizers consistently have more radionuclides than organic and urea fertilizers. Risk assessments have shown that PO<sub>4</sub><sup>3-</sup> fertilizers containing small amounts of Cd in PR are safe, while radionuclides in PR do not pose any radiological risk.

### 3. Conclusion

All in all, the implications of phosphate extraction and utilization are of ecological significance with severe after-effects and far-reaching effects that should be closely watched. Our comprehensive analysis has revealed that processes of phosphate mining, extraction, and transport today represent acute ecosystems, water resources and human health threats.

The key environmental challenges brought about by phosphate production include habitat destruction, water pollution, soil erosion, and air pollution. In addition to this, the irreversible adverse effects of these activities include habitat loss, biodiversity extinction, and erosion that should be addressed through altered ground-up approaches and regulatory mechanisms. Addressing these challenges necessitates a multidisciplinary approach involving technological advances, better waste disposal systems, and the implementation of environmentally sound practices throughout the phosphate value chain. In addition, maintenance of relationships and transparent collaboration between stakeholders such as governments, industries, local communities should be forested for a meaningful fight against environmental impacts of phosphate extraction and use.

In facing the ensuing challenges fostered by the phosphate extraction and utilization, we need to strengthen sustainability, resilience, and our natural resource stewardship. Through an investment in research, innovation, and wholesome management strategies, we can endeavor to reduce environmental damage, protect biodiversity, and protect the welfare of the current and future generations. The only way we can manage to find solutions to the ecological impacts of phosphate extraction is through joint efforts and organized action that shall usher the world to a more consonant and balanced habitat with Mother Earth.

#### 4. References

- Science, New Series, Vol. 52, No. 1351 (Nov. 19, 1920), p. 484 (1 page)
- RCC Perspectives, No. 4, Big Country, Big Issues: Canada's Environment, Culture, and History (2011), pp. 86-100 (15 pages)
- Consilience, No. 19 (2018), pp. 17-35 (19 pages)
- Environmental Conservation, Vol. 15, No. 3 (Autumn 1988), pp. 269-270 (2 pages)
- Environmental Reviews, Vol. 24, No. 4 (December 2016), pp. 403-415 (13 pages)
- Jasinski, S.M. 2009. Phosphate Rock. Mineral Commodity Summaries, U.S. Geological Survey. Available from [http://Minerals.USGS.GOV/Minerals/Pubs/Commodity/Phosphate\\_Rock/MCS-2009-Phosp.Pdf](http://Minerals.USGS.GOV/Minerals/Pubs/Commodity/Phosphate_Rock/MCS-2009-Phosp.Pdf) [accessed 17 August 2015].
- Cordell, D., and White, S. 2011. Peak Phosphorus: Clarifying the Key Issues of a Vigorous Debate About Long-Term Phosphorus Security. *Sustainability*, 3: 2027–2049. doi:10.3390/su3102027.
- Ibrahim, S.S., El-Midany, A.A., and Boulos, T.R. 2010. Economic Preferences of Mechanical Activation Over Mineral Beneficiation for Phosphate Rock Direct Applications. *Physicochem. Probl. Miner. Process.* 44: 63–78.
- Cooper, J., Lombardi, R., Boardman, D., and Carliell-Marquet, C. 2011. The Future Distribution and Production of Global Phosphate Rock Reserves. *Resour.Conserv. Recycl.* 57: 78–86. doi:10.1016/j.resconrec.2011.09.009.
- Cordell, D., and White, S. 2011. Peak Phosphorus: Clarifying the Key Issues of a Vigorous Debate About Long-Term Phosphorus Security. *Sustainability*, 3:2027–2049. doi:10.3390/su3102027.
- Heckenmüller, M., Narita, D., and Klepper, G. 2014. Global Availability of Phosphorus and Its Implications for Global Food Supply: An Economic Overview. Kiel Working Paper No. 1897. Kiel Institute for The World Economy, Germany.
- Van Kauwenberg, S.J. 2010. World Phosphate Rock Reserves and Resources. International Fertilizer Development Centre. IFDC.
- Selles, F., Campbell, C.A., Zentner, R.P., Curtin, D., James, D.C., and Basnyat, P. 2011. Phosphorus use efficiency and long-term trends in soil available phosphorus in wheat production systems with and without nitrogen fertilizer. *Can. J. Soil Sci.* 91: 39–52, 10.4141/CJSS10049.
- Leone, A., Ripa, M.N., Boccia, L., and Lo Porto, A. 2008. Phosphorus export from agricultural land: a simple approach. *Biosyst. Eng.*101:270–280.doi:10.1016/j.biosystemseng.2008.07.005.
- Husnain, S.R., Sutriadi, T., Nassair, A., and Sarwani, M. 2014. Improvement of Soil Fertility and Crop Production through Direct Application of Phosphate Rock on Maize in Indonesia. *Procedia Eng.* 83: 336–343. doi:10.1016/j.proeng.2014.09.025.
- Bernard, A. 2008. Cadmium and its adverse effects on human health. *Indian J.Med. Res.* 128(4): 557–564. PMID:19106447.

- 
- Ferreccio, C., Smith, A., Duran, V., Barlaro, T., Benitez, H., Valdes, R., et al. 2013. Case-Control Study of Arsenic in Drinking Water and Kidney Cancer in Uniquely Exposed Northern Chile. *Am. J. Epidemiol.* 178(5): 813–818. doi:10.1093/aje/kwt059. PMID:23764934.
  - Hou, S., Yuan, L., Jin, P., Ding, B., Qin, N., Li, L., et al. 2013. A clinical study of the effects of lead poisoning on the intelligence and neurobehavioral abilities of children. *Theor. Biol. Med. Modell.* doi:10.1186/1742-4682-10-13.
  - USEPA. 2000. Radionuclides (including Radon, Radium and Uranium), Hazard Summary. Available from <https://www3.epa.gov/airtoxics/hlthef/radionuc.html> [accessed 15 June 2016].
  - Murakami, M., and Oki, T. 2015. Correction: Estimated dietary intake of radionuclides and health risks for the citizens of Fukushima City, Tokyo, and Osaka after the 2011 nuclear accident. *PLOS One*, 10(8): 0136223. doi:10.1371/journal.pone.0136223.
  - FAO. 2004. Use of Phosphate Rock for Sustainable Agriculture. In *FAO Fertilizer and Plant Nutrition Bulletin*. Edited by F. Zapata and R.N. Roy. A Joint Publication of the FAO Land and Water Development Division and International Atomic Energy Agency (IAEA).
  - Mew, M. 2000. Phosphate Rock. In *Metals and Mineral Annual Review*. London, The Mining Journal Ltd. pp. 110–122
  - Dennis, C. 2013. Arianne Resources Inc. – A Phosphate Rock Star. Toronto: Dundee Securities Ltd. Available from <http://Research.DundeeCapitalMarkets.Com/En//Media/Dcm/Publications/CoverageIsearch/2013/March/D/Dan031313.Ashx> [accessed 27 July 2015].
  - ICL. 2013. Website of Israel Chemicals Ltd. (ICL). Available from <http://www.Icl-Group.Com/AboutIcl-Segments/General/3df4fef5-4e7b-44b3-88b4-99f67dedd60a.AspX>.
  - Hodgkin, E.P., and Hamilton, B.H. 1993. Fertilizers and eutrophication in south-western Australia: Setting the scene. *Fertilizer Res.* 36: 95–103. doi:10.1007/BF00747579.
  - Kennish, M.J., and de Jonge, V.N. 2011. Chemical Introductions to the Systems: Diffuse and Nonpoint Source Pollution from Chemicals (Nutrients: Eutrophication). In *Treatise on Estuarine and Coastal Science*, Academic Press, Waltham, pp. 113–148.
  - Scavia, D., Allan, J., Arend, K., Bartell, S., Beletsky, D., Bosch, N., et al. 2014. Assessing and addressing the re-eutrophication of Lake Erie: Central basin hypoxia. *J. Great Lakes Res.* 40: 226–246. doi:10.1016/j.jglr.2014.02.004.
  - Hodgkin, E.P., and Hamilton, B.H. 1993. Fertilizers and eutrophication in southwestern Australia: Setting the scene. *Fertilizer Res.* 36: 95–103. doi:10.1007/BF00747579.
  - Wang, H., and Wang, H. 2009. Mitigation of lake eutrophication: Loosen nitrogen control and focus on phosphorus abatement. *Progr. Nat. Sci.* 19: 1445–1451. doi:10.1016/j.pnsc.2009.03.009.

- 
- Vandenhove, H., Batlle, J.V., and Sweeck, L. 2015. Potential radiological impact of the phosphate industry on wildlife. *J. Environ. Radioact.* 141: 14–23. doi:10.1016/j.jenvrad.2014.11.001. PMID:25500062.
  - Aoun, M., Arnaudguilhem, C., El Samad, O., Khozam, R.B., and Lobinski, R. 2015. Impact of a phosphate fertilizer plant on the contamination of marine biota by heavy elements. *Environ. Sci. Pollut. Res.* 22(19):14940–14949. doi:10.1007/s11356-015-4691-4. PMID:26002362.
  - Bulut, E., and Aksoy, A. 2008. Impact of fertilizer usage on phosphorus loads to Lake Uluabat. *Desalination*, 226(1–3): 289–297. doi:10.1016/j.desal.2007.02.112.
  - Castlehouse, H., Osborn, A., and Cripps, J. 2010. Modeling the chemical influences on bioavailability of geogenic arsenic in soils. In 19th World Congress of Soil Science, Australia.
  - Davenport, J., and Peryea, F. 1991. Phosphate Fertilizers Influence Leaching of Lead and Arsenic in a Soil Contaminated with Lead Arsenate. *Water, Air, Soil Pollut.* 57–58: 101–110. doi:10.1007/BF00282873.
  - Cao, X., Ma, L.Q., and Shiralipour, A. 2003. Effects of compost and phosphate amendments on arsenic mobility in soils and arsenic uptake by the hyperaccumulator, *Pteris vittata* L. *Environ. Pollut.* 26:157–167. doi:10.1016/S02697491(03)00208-2. PMID:12927487.
  - Lessl, J.T., Luo, J., and Ma, L.Q. 2014. *Pteris vittata* continuously removed arsenic from non-labile fraction in three contaminated-soils during 3.5 years of phytoextraction. *J. Hazard. Mater.* 279: 485–492. doi:10.1016/j.jhazmat.2014.06.056. PMID:25108101.
  - Fayiga, A.O., and Ma, L.Q. 2006. Using Phosphate Rock to Immobilize Metals in Soil and Increase Arsenic Uptake to Hyperaccumulate *Pteris vittata*. *Sci. Total Environ.* 359: 17–25. doi:10.1016/j.scitotenv.2005.06.001. PMID:15985282.
  - Glennie, E.B., Littlejohn, C., Gendebien, A., Hayes, A., Palfrey, R., Sivil, D., and Wright, K. 2002. Phosphates and alternative detergent builders – Final report. EU Environment Directorate.
  - Kundu, S., Coumar, M., Rajendiran, S., and Rao, S. 2015. Phosphates from detergents and eutrophication of surface water ecosystem in India. *Curr. Sci.* 108(7).
  - European commission. 2015. Phosphates and Alternative Detergent Builders. Available from [http://ec.europa.eu/environment/water/pollution/phosphates/index\\_en.htm](http://ec.europa.eu/environment/water/pollution/phosphates/index_en.htm) [accessed 26 March 2016].
  - Landbank. 1994. The Phosphate Report. Landbank Environmental Research and Consulting, London.
  - [https://wwf.panda.org/wwf\\_news/?8455466/Mining-impacts-affect-up-to-13-of-global-forest-ecosystems-and-tipped-to-rise-with-increased-demand-for-metals](https://wwf.panda.org/wwf_news/?8455466/Mining-impacts-affect-up-to-13-of-global-forest-ecosystems-and-tipped-to-rise-with-increased-demand-for-metals)
  - <https://earth.org/environmental-problems-caused-by-mining/>
  - Ogunleye, P.O., Mayaki, M.C., and Amapu, I.Y. 2002. Radioactivity and Heavy Metal Composition of Nigerian Phosphate Rocks: Possible Environmental Implications. *J. Environ. Radioact.* 62(1): 39–48. doi:10.1016/S0265-931X(01)00149-7. PMID:12141606.

- 
- da Conceição, F.T., and Bonotto, D.M. 2006. Radionuclides, heavy metals and fluorine incidence at Tapira phosphate rocks, Brazil and their industrial (by) products. *Environ. Pollut.* 139: 232–243. doi:10.1016/j.envpol.2005.05.014. PMID:16099562.
  - Hartley, T.N., Macdonald, A.J., McGrath, S.P., and Zhao, F. 2013. Historical arsenic contamination of soil due to long term phosphate fertilizer applications. *Environ. Pollut.* 180: 259–264. doi:10.1016/j.envpol.2013.05.034. PMID:23792386.
  - Jayasumana, C., Fonseka, S., Fernando, A., Jayalath, K., Amarasinghe, M., Siribaddana, S., et al. 2015. Phosphate fertilizer is a main source of arsenic in areas affected with chronic kidney disease of unknown etiology in Sri Lanka. *SpringerPlus* 4: 90. doi:10.1186/s40064-015-0868-z.
  - Lambert, R., Grant, C., and Sauvé, S. 2007. Cadmium and zinc in soil solution extracts following the application of phosphate fertilizers. *Sci. Total Environ.* 378(3): 293–305. doi:10.1016/j.scitotenv.2007.02.008. PMID:17400282.
  - Landbank. 1994. *The Phosphate Report*. Landbank Environmental Research and Consulting, London.
  - Gonçalves, A.C., Jr., Nacke, H., Schwantes, D., and Coelho, G.F. 2014. Heavy metal contamination in Brazilian agricultural soils due to application of fertilizers. In *Environmental risk assessment of soil contamination*.
  - FAO. 2004. *Use of Phosphate Rock for Sustainable Agriculture*. In *FAO Fertilizer and Plant Nutrition Bulletin*. Edited by F. Zapata and R.N. Roy. A Joint Publication of the FAO Land and Water Development Division and International Atomic Energy Agency (IAEA).
  - Van Kauwenbergh, S.J. 2006. *Fertilizer Raw Material Resources of Africa*. Africa Fertilizer Summit. Abuja, Nigeria. IFDC R-16.
  - Roberts, T.L. 2014. Cadmium and Phosphorus Fertilizers: The Issues and the Science. *Procedia Eng.* 83: 52–59. doi:10.1016/j.proeng.2014.09.012.
  - Chen, W., Chnag, A.C., and Wu, L. 2007. Assessing long term environmental risks of trace elements in phosphate fertilizers. *Ecotoxicol. Environ. Safety*, 67: 48–58. doi:10.1016/j.ecoenv.2006.12.013.
  - Akhtar, N., Tufail, M., and Ashraf, M. 2005. Natural Environmental Radioactivity and Estimation of Radiation Exposure from Saline Soils. *Int. J. Environ. Sci. Technol.* 1(4): 279–285. doi:10.1007/BF03325843.
  - Mlwilo, N.A., Mohammed, N.K., and Spyrou, N.M. 2007. Radioactivity Levels of Staple Foodstuffs and Dose Estimates for Most of the Tanzanian Population. *J. Radiol. Prot.* 27: 471–480. doi:10.1088/0952-4746/27/4/008.
  - Baetsle, L.H. 1991. *Study of the Radio-Nuclides Contained in Wastes Produced by the Phosphate Industry and Their Impact on the Environment*. European Atomic Energy Communities. Nuclear Science and Technology. Final Report. Eur 13262 En.

- 
- Hameed, P.S., Pillai, G.S., and Mathiyarasu, R. 2014. A Study on the Impact of Phosphate Fertilizers on the Radioactivity Profile of Cultivated Soils in Srirangam (Tamil Nadu, India). *J. Radiat. Res. Appl. Sci.* 7: 463–471. doi:10.1016/j.jrras.2014.08.011.
  - FAO. 2004. Use of Phosphate Rock for Sustainable Agriculture. In *FAO Fertilizer and Plant Nutrition Bulletin*. Edited by F. Zapata and R.N. Roy. A Joint Publication of the FAO Land and Water Development Division and International Atomic Energy Agency (IAEA).
  - Hassan, N.M., Mansour, N.A., Fayed-Hassan, M., and Sedqy, E. 2015. Assessment of natural radioactivity in fertilizers and phosphate ores in Egypt. *J. Taibah Univ. Sci.* 10(2): 296–306. doi:10.1016/j.jtusci.2015.08.009.
  - Hameed, P.S., Pillai, G.S., and Mathiyarasu, R. 2014. A Study on the Impact of Phosphate Fertilizers on the Radioactivity Profile of Cultivated Soils in Srirangam (Tamil Nadu, India). *J. Radiat. Res. Appl. Sci.* 7: 463–471. doi:10.1016/j.jrras.2014.08.011.
  - Uosif, M., Mostafa, A., Elsaman, R., and Moustafa, E. 2014. Natural radioactivity levels and radiological hazards indices of chemical fertilizers commonly used in Upper Egypt. *J. Radiat. Res. Appl. Sci.* 7(4): 430–437. doi:10.1016/j.jrras.2014.07.006.
  - Kumar, A., and Chauhan, R.P. 2014. Measurement of alpha activity from leaves and roots of radish plant enhanced by fertilizers. *J. Rad. Res. Appl. Sci.* 7(4): 454–458. doi:10.1016/j.jrras.2014.07.009.
  - Saueia, C.H., and Mazzilli, B.P. 2006. Distribution of Natural Radionuclides in the Production and Use of Phosphate Fertilizers in Brazil. *J. Environ. Radioact.* 89: 229–239. doi:10.1016/j.jenvrad.2006.05.009. PMID:16849030.
  - Khater, A.E., and Al-Sewaidan, H. 2008. Radiation exposure due to agricultural uses of phosphate fertilizers. *Radiat. Meas.* 43(8): 1402–1407. doi:10.1016/j.radmeas.2008.04.084.
  - Bituh, T., Marovic, G., Franic, Z., Sencar, J., and Bronzovic, M. 2009. Radioactive contamination in Croatia by phosphate fertilizer production. *J. Haz. Mater.* 162(2–3): 1199–1203. doi:10.1016/j.jhazmat.2008.06.005.
  - Cevik, U., Baltas, H., Tabak, A., and Damla, N. 2010. Radiological and chemical assessment of phosphate rocks in some countries. *J. Haz. Mater.* 182(1–3): 531–535. doi:10.1016/j.jhazmat.2010.06.064.
  - Hameed, P.S., Pillai, G.S., and Mathiyarasu, R. 2014. A Study on the Impact of Phosphate Fertilizers on the Radioactivity Profile of Cultivated Soils in Srirangam (Tamil Nadu, India). *J. Radiat. Res. Appl. Sci.* 7: 463–471. doi:10.1016/j.jrras.2014.08.011.