

## **Controls on coastal saline groundwater across North America**

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## 33 **Abstract**

34 Groundwater is crucial to sustaining coastal freshwater needs. About 32 million people in the  
35 coastal USA rely on groundwater as their primary water source. With rapidly growing coastal  
36 communities and increasing demands for fresh groundwater, understanding controls of continental-  
37 scale coastal groundwater salinity is critical. To investigate what hydrogeological factors (e.g.,  
38 topography, hydraulic conductivity) control coastal saline groundwater at continental scales, we  
39 have simulated variable-density groundwater flow across North America with the newly developed  
40 Global Gradient-based Groundwater Model with variable Densities (G<sup>3</sup>M-D). The simulation  
41 results suggest that under a steady climate and pre-development conditions (i.e., steady 30-year  
42 mean groundwater recharge, no withdrawals nor sea level rise) saline groundwater is present in  
43 18.6% of North America's coastal zone, defined as up to 100 km inland and up to 100 m above  
44 mean sea level. We find that the coastal zone is particularly vulnerable to containing saline  
45 groundwater at low hydraulic gradients ( $<10^{-4}$ ) and large hydraulic conductivities ( $>10^{-2}$  m day<sup>-1</sup>).  
46 To analyze model parameter sensitivities, i.e., which parameters control the resulting distribution  
47 of saline groundwater, we utilize the inherent spatial model variability. We find that hydraulic  
48 gradient, topographic gradient, hydraulic conductivity, and aquifer depth are important controls in  
49 different places. However, no factor controls coastal groundwater salinization alone, suggesting  
50 that parameter interactions are important. Using G<sup>3</sup>M-D based on G<sup>3</sup>M, a model that previous work  
51 found to be strongly controlled by topography, we find no controlling influence of recharge  
52 variability on the saline groundwater distribution in North America. Despite a likely overestimation  
53 of saline interface movement, the model required 492 000 years to reach a near-steady state,  
54 indicating that the saline groundwater distribution in North America has likely been evolving since  
55 before the end of the last ice age, approximately 20 000 years ago.

## 56 **1. Introduction**

57 Coastal groundwater is vital to sustaining coastal freshwater consumption and agricultural activities  
58 in the US (Barlow and Reichard, 2010) and other countries worldwide (Custodio, 2010; Shi and  
59 Jiao, 2014; Manivannan and Elango, 2019). About half of all coastal counties (143 of 297) in the  
60 US, home to 32 million people, rely on groundwater as their main water supply (Dieter et al., 2018).  
61 Between 1960 and 2008, the population in coastal counties in the US grew by over 80%, 20% more  
62 than non-coastline counties (Wilson and Fischetti, 2010). Over a similar period, from 1950 to 2015,  
63 the growing demand for freshwater led to a doubling of groundwater withdrawal in the US (Dieter

64 et al., 2018), causing hydraulic gradients at the coast to reduce and even turn landward (Jasechko  
65 et al., 2020).

66 Where hydraulic gradients at the coast decline (e.g., due to a drop in the groundwater table or  
67 relative to sea level rise), saline ocean water may intrude into the groundwater system and salinize  
68 freshwater aquifers. In addition to growing water demand, storm surges and sea-level changes may  
69 exacerbate seawater intrusion (Post et al., 2018). Seawater intrusion has already affected coastal  
70 groundwater across North America (Barlow and Reichard, 2010). Worldwide, nearly a third of all  
71 coastal metropolises are threatened by seawater intrusion (Cao et al., 2021). However, our  
72 understanding of the rapidly changing coastal groundwater lacks predictive capabilities  
73 (Richardson et al., 2024), which is why we need to better understand dominant controls of coastal  
74 saline groundwater and how these vary along coastlines.

75 Several continental and global studies have addressed the issue of seawater intrusion. In a study of  
76 the US coast, Ferguson & Gleeson (2012) show that groundwater withdrawal in coastal regions is  
77 a greater control on horizontal seawater intrusion than sea level rise or changes in groundwater  
78 recharge. Based on estimated submarine groundwater discharge and groundwater withdrawals, 9%  
79 of the contiguous United States coastline are vulnerable to seawater intrusion (Sawyer et al., 2016).  
80 Resilience against seawater intrusion driven by sea level rise is higher when groundwater levels  
81 within aquifers can shift upwards, balancing the gradient change induced by sea level rise (Michael  
82 et al., 2013). In other words, aquifers are more resilient where the topographic gradient to the coast  
83 remains larger than the hydraulic gradient to the coast as sea-level rise progresses. Similarly,  
84 groundwater simulations along the coast of California show that coastal topography controls  
85 seawater intrusion and overland flooding due to sea-level rise (Befus et al., 2020). Recently  
86 published results from groundwater models in 1 200 coastal regions around the world suggest that  
87 coastal fresh groundwater volumes will decrease by about 5% until 2100 due to sea-level rise  
88 (Zamrsky et al., 2024). They confirmed previous findings showing higher resilience against  
89 seawater intrusion in regions with higher topographic gradients, often aligning with steeper  
90 hydraulic groundwater gradients.

91 However, previous simulations of coastal groundwater share a major limitation: their model extent  
92 is limited landward, thus requiring assumptions about the landward boundary condition (Michael  
93 et al., 2013; Zamrsky et al., 2024), which can strongly impact the results of seawater intrusion  
94 simulations (Werner and Simmons, 2009; Ketabchi et al., 2016). Michael et al. (2013) simulated a

95 theoretical aquifer to analyze the effect of changing groundwater recharge, hydraulic conductivity,  
96 and anisotropy on the saltwater distribution in the aquifer. However, since the changes were applied  
97 one at a time, their combined effects were not simulated. Further, the only global assessment of  
98 seawater intrusion (Zamrsky et al., 2024) was limited to a quarter of the global coastline with  
99 permeable unconsolidated sedimentary formations. Hence, wide parameter ranges and  
100 combinations remain unexplored. Table S1 shows a comparison of continental and global coastal  
101 groundwater models.

102 Here, we use a Darcy approach (Reinecke et al., 2019b) to simulate groundwater flow of the entire  
103 North American continent under a steady climate (e.g., steady groundwater recharge) and natural,  
104 pre-pumping conditions (i.e., without withdrawals). The density zones are simulated with a SWI2-  
105 like variable density routine (Bakker et al., 2013). Like the problem described by Henry (1964),  
106 the entire groundwater system is fresh in its initial state, ensuring that the ocean is the only source  
107 of saltwater (which is a simplification as saline groundwater can have multiple other sources). As  
108 the over 450 000 model cells were parameterized individually, the model incorporates all  
109 combinations of input parameters existing at the simulated resolution of 5 arcminutes (roughly 9.2  
110 km at the Equator). This allows us to assess which of the impact factors, topographic gradient (dT),  
111 hydraulic gradient (dH), hydraulic conductivity (K), aquifer depth ( $D_{\text{aqu}}$ ), and groundwater  
112 recharge (GWR) control the simulated distribution of saline groundwater.

## 113 2. Methods

### 114 2.1 The global gradient-based groundwater model

115 The global gradient-based groundwater model, G<sup>3</sup>M (Reinecke et al. 2019a; Reinecke et al.,  
116 2019b), was inspired by concepts of MODFLOW-2005 (Harbaugh et al., 2005) and built to be  
117 coupled with global hydrological models. To facilitate the assessment of groundwater at the global  
118 scale, hydraulic gradients between grid cells drive the flow between the cells. The three-  
119 dimensional flow of groundwater is described by a partial differential equation (Harbaugh et al.,  
120 2005):

$$121 \quad \frac{\partial}{\partial x} \left( K_{xx} \frac{\partial h}{\partial x} \right) + \frac{\partial}{\partial y} \left( K_{yy} \frac{\partial h}{\partial y} \right) + \frac{\partial}{\partial z} \left( K_{zz} \frac{\partial h}{\partial z} \right) + W = S_s \frac{\partial h}{\partial t}$$

122 where  $K_{xx}$ ,  $K_{yy}$ , and  $K_{zz}$  [ $L^2T^{-1}$ ] are the hydraulic conductivity along the x, y, and z axes between  
123 the cells with sizes  $\Delta x$ ,  $\Delta y$ , and  $\Delta z$  [L];  $S_s$  [ $L^{-1}$ ] is the specific storage;  $h$  [L] is the hydraulic head.

124 W [ $T^{-1}$ ] incorporates the flows into and out of each cell, such as groundwater recharge, surface  
125 water bodies (i.e., rivers, lakes, and wetlands), or the ocean. Just like the flows between the cells,  
126 the flow from a cell to a river, lake, wetland, or ocean depends on the respective heads. Thus, each  
127 cell can receive/give water from/to neighboring cells and additionally from/to rivers, lakes,  
128 wetlands, and the ocean. In a coupled state these surface water bodies are updated by the  
129 hydrological model. In this study surface water heads are kept at their initial elevation (30<sup>th</sup>  
130 percentile of a 30 arcsecond digital elevation model; Reinecke et al., 2019b).

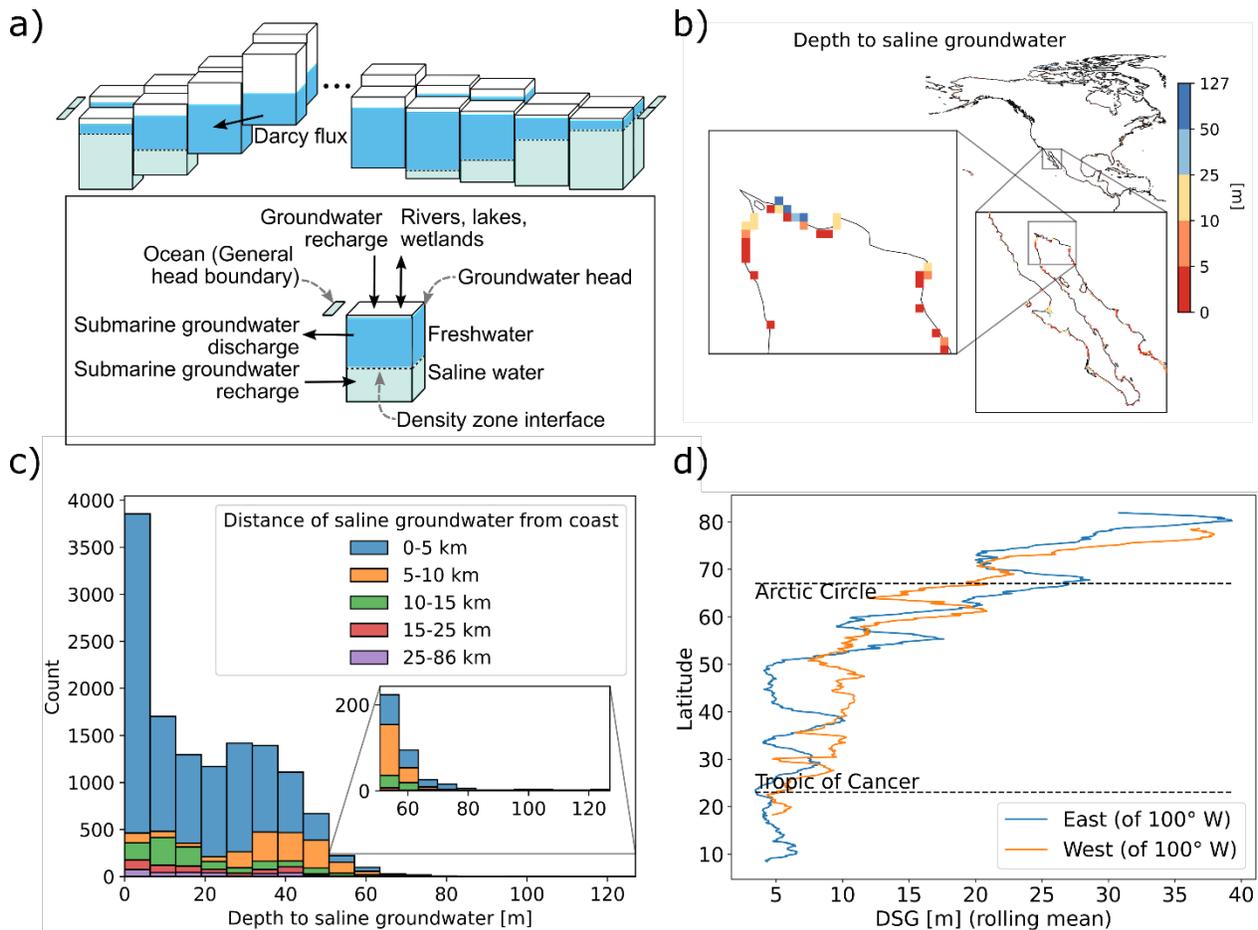
## 131 **2.2 The added variable density routine**

132 Freshwater has a lower density than water containing salt. In the newly developed Global Gradient-  
133 based Groundwater Model with variable Densities ( $G^3M-D$ ), sharp interfaces lie between density  
134 zones representing salinity levels. The height of these interfaces is simulated similarly to the  
135 Saltwater Intrusion package (SWI2) developed for MODFLOW (Bakker et al., 2013). A SWI2-like  
136 routine was implemented due to its wide range of applications and low simulation cost, which are  
137 essential in developing large-scale models. Compared to  $G^3M$ , which simulates groundwater heads,  
138  $G^3M-D$  has an additional density interface routine. The groundwater head routine accounts for the  
139 density zone volumes in each cell before solving the variable density equations in the separate  
140 density interface routine. Hence, the mass balance equation (used with constant density) is replaced  
141 by a volume balance equation when simulating variable densities (Text S1 and Bakker et al., 2013).  
142 As density interfaces may need many time steps to develop, multiple shorter variable density time  
143 steps can be simulated per groundwater flow step to reduce simulation time.

### 144 **2.2.1 Density zones and density interfaces**

145 In  $G^3M-D$ , like in SWI2, density zones in each cell are stacked vertically (see Fig 1 a)). The model  
146 calculates the height of horizontal sharp density interfaces, representing the limits of density zones.  
147 Each zone is constant in density (i.e., this corresponds to the discontinuous option in SWI2). In a  
148 setup with one density interface between the two density zones of fresh water and seawater (used  
149 in this study), the density interface represents the approximate location of 50 percent seawater in  
150 the aquifer, neglecting the effects of dispersion and diffusion. In other words, density interface  
151 heights change when the proportions of density zones within a cell change, without simulating a  
152 mixing of density zones. Another limitation is that density can only be inverted between model  
153 layers (i.e., aquifers) but not within the same model layer. While inputs of saline water (i.e., inflow

154 from a neighboring cell) may cause a density interface to rise, freshwater inputs (i.e., from  
 155 groundwater recharge, rivers, or neighboring cells) may induce groundwater flow out of the cell,  
 156 potentially lowering the interface height. At each density time step, new interface heights are  
 157 computed iteratively for all cells with saline water, followed by interface adjustments. These  
 158 adjustments allow the horizontal movement of saline water from a cell with saline water to an  
 159 entirely fresh neighboring cell (e.g., when the slope between an interface height and the  
 160 neighboring cell bottom is above a threshold). For equations and subroutines of the density routine,  
 161 please refer to Text S1 and Bakker et al. (2013).



162  
 163 **Figure 1** – Panel showing a) the concept of the variable density groundwater model (G<sup>3</sup>M-D), b)  
 164 simulated depth to saline groundwater (DSG) at Baja California (for the entire map of North  
 165 America, see Figure S1), c) a histogram of simulated depths to saline groundwater, and d) the  
 166 moving average of DSG on the east and west coast of North America by latitude (separation at  
 167 100° W). The model cell size is 5 arcminutes (~ 9.2 km at the equator).

## 168 **2.2.2 Testing the implementation**

169 The newly implemented variable density routine was tested using Examples 1 to 3 from the SWI2  
170 documentation (Bakker et al., 2013). Each example tests different parts of our implementation (see  
171 Fig S2). Our results in Example 1 show that G<sup>3</sup>M-D can accurately simulate the height change of  
172 one density interface in an aquifer with an inflow of saline water. Example 2 shows that more than  
173 one interface can be simulated correctly, simulating two interfaces rotating around the brackish  
174 zone they enclose. Including three different aquifer layers with changing hydraulic conductivities,  
175 Example 3 demonstrates that the movement of a density interface between layers is also accurate.

## 176 **2.3 The variable-density groundwater model of North America**

### 177 **2.3.1 Model setup**

178 Several global datasets are used in the G<sup>3</sup>M-D model setup of the North American continent,  
179 including the entire inland. Elevation and surface water bodies (i.e., rivers, lakes, and wetlands) are  
180 parameterized as in Reinecke et al. (2019b). GLHYMPS 2.0 (Huscroft et al., 2018) provided  
181 hydraulic conductivity (mean of model cells: 8.57 m day<sup>-1</sup>) and effective porosity (mean of model  
182 cells: 0.047). Cells with an effective porosity of 0 (56% of the model cells, 20.4% at the coastline)  
183 are excluded from the variable density routine, meaning they cannot hold saline water. The model  
184 does not represent conduits (e.g., in karstic or volcanic aquifers). The groundwater recharge input  
185 was calculated as the 1987-2016 mean from a WaterGAP (Müller Schmied et al., 2020) simulation  
186 using WFDEI (Weedon et al., 2014) as meteorological forcing (mean of model cells: 0.187 mm  
187 day<sup>-1</sup>). The thickness of the single aquifer layer was defined using depth to bedrock data (mean of  
188 model cells: 24.26 m) by Shangguan et al. (2017). This entails that no aquitards or deep confined  
189 aquifers are represented in the model. The input data for elevation, groundwater recharge, effective  
190 porosity, hydraulic conductivity, and aquifer thickness are displayed in Figure S3.

191 A general head boundary (GHB) (Harbaugh, 2005) represents the ocean at all coastline cells and is  
192 set to a constant elevation of 0 m. The coastal shoreline permeability was retrieved from the global  
193 coastal permeability dataset (CoPerm) (Moosdorf et al., 2024) and used to parameterize the GHB  
194 conductance. No groundwater pumping was included in the simulation to assess the coastal saline  
195 groundwater under naturalized conditions. The assumption that the ocean is the only source of  
196 saltwater entails that existing saline groundwater deposits in large parts of North America are  
197 omitted (Feth, 1965; Reilly et al., 2008). Assuming a constant groundwater temperature of 12°C

198 for the entire North American continent, freshwater (salinity: 0 parts per thousand) was assigned  
199 the density of  $999.5 \text{ kg/m}^3$ , and ocean water (salinity: 35 parts per thousand) was assigned the  
200 density of  $1\,026.6 \text{ kg/m}^3$ . A comparison to other continental or global studies on coastal saline  
201 groundwater is shown in Table S1.

### 202 **2.3.2 Finding stable interface positions**

203 At the start of the simulation, all groundwater in the system was fresh. Over time, saline ocean  
204 water intruded the simulated system through the general head boundary. Since groundwater density  
205 develops significantly slower than the groundwater head, the model was run under pseudo-steady  
206 state conditions (Bakker et al., 2013), i.e., with steady sea level, coastline, and recharge, while  
207 computing changes in density interface heights. Further, one thousand annual density time steps  
208 were simulated for each groundwater flow time step of thousand years. The simulation was run  
209 until the interface heights were stable, i.e., the following conditions were satisfied: in two  
210 consecutive time steps of thousand years the interface height change (a) in 95% of the cells with  
211 saline water is below 0.05 m and (b) in 99% of the cells with saline water was below 0.1 m. This  
212 was the case after 492 time steps (i.e., 492 000 years).

### 213 **2.4 Utilizing spatial variability of inputs and outputs to understand process controls**

214 The groundwater model of North America simulates heads and interface heights in 452 736 cells,  
215 of which 18 808 are coastline cells (i.e., cells with at least one side facing the ocean). We use the  
216 intrinsic spatial variability of inputs and outputs in our evaluation to analyze the factors that control  
217 coastal saline groundwater (similar to Gnann et al., 2023). We consider three aquifer properties:  
218 aquifer depth, hydraulic conductivity, and topographic gradient, as well as two hydrologic  
219 characteristics: groundwater recharge and the hydraulic gradient (resulting from the groundwater  
220 head routine).

221 For all cells containing saline groundwater at the stable state, we evaluate three different aspects of  
222 coastal saline groundwater: Saline Groundwater Fraction (SGF), Thickness of Fresh Groundwater  
223 column (TFG), and Distance of saline groundwater from Coast (DC) (Table 1). We separately  
224 assess factor value distributions in cells with moderate and pronounced (1) saline groundwater  
225 fraction, (2) thickness of fresh groundwater column, and (3) distance of saline groundwater from  
226 the coast to assess which factors control the severity of saltwater occurrence in coastal groundwater  
227 (see Table 1). We repeated this evaluation with increased and decreased thresholds to assess the

228 sensitivities of the thresholds separating into moderate and pronounced aspects of coastal saline  
229 groundwater.

230 **Table 1** – Aspects of coastal saline groundwater with their respective abbreviations and  
231 explanations.

<b>Aspect of coastal saline groundwater</b>	<b>Abbreviation</b>	<b>Calculated as</b>	<b>Aspect pronounced if</b>
Saline Groundwater Fraction	SGF	Share of saline water in the groundwater column	$SGF > 0.5$
Thickness of Fresh Groundwater column	TFG	Groundwater head – Interface height	$TFG < 5 \text{ m}$
Distance of saline groundwater from coast	DC	Cell distance from the coastline in km	$DC > 10 \text{ km}$

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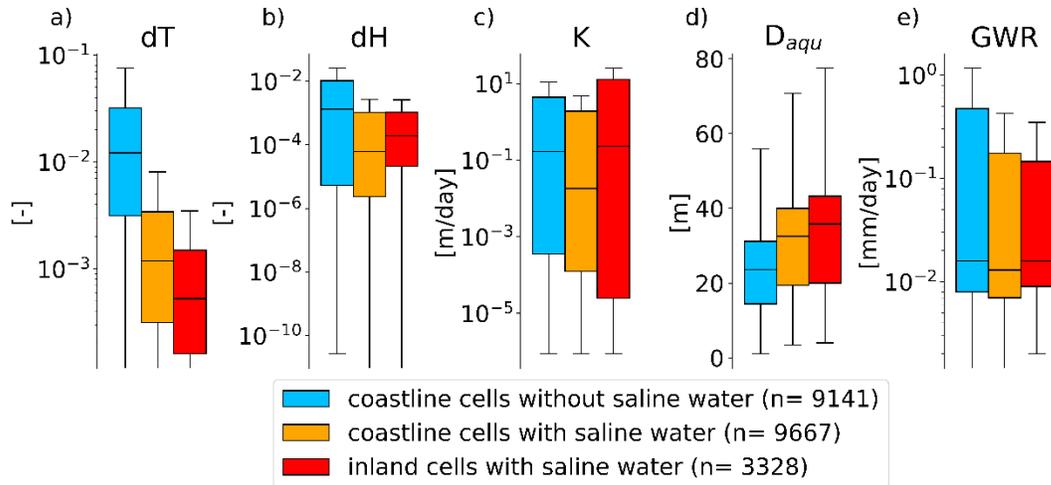
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### 234 3. Results

235 In the final (stable) state of the salinity interfaces, 12 995 (2.9%) of the 452 736 simulated cells  
236 contained saline groundwater (9 667 of which are coastline cells). The simulated state does not  
237 necessarily represent the current real-world situation. It evolved from an initial entirely fresh  
238 groundwater system and shows the potential spatial distribution of saline groundwater for steady  
239 groundwater recharge and sea level without groundwater pumping or historical marine brine  
240 deposits. The simulated depth to saline groundwater (DSG) in most cells (86%) containing saline  
241 water is less than 40 m (Fig 1 c)), with large regions of shallow saline groundwater in Alaska (US),  
242 Nunavut (Canada), and Oaxaca (Mexico) (Fig 1 b), Fig S3). In both the east and west of North  
243 America, depth to saline groundwater (DSG) reduces from about 50 m in the north to roughly 10  
244 m in the south (Fig 1 d), reflecting the aquifer thickness distribution (Fig S3). In the following, we  
245 examine the sensitivity to the possible impacting factors, i.e., topographic gradient (dT), hydraulic  
246 gradient (dH), hydraulic conductivity (K), aquifer depth ( $D_{\text{aqu}}$ ), and groundwater recharge (GWR)  
247 to find the dominant controls in coastal groundwater salinity on the continental scale.

#### 248 3.1 Topographic gradient and aquifer depth control incursion at the continental coastline

249 Roughly half (i.e., 9 667 of 18 808) of the North American coastline cells (i.e., cells with at least  
250 one side facing the ocean) contain saline water in the stable state, while the other half (9 141) stays  
251 entirely fresh. Figure 2 shows the factor distributions of (1) coastline cells without saline water  
252 (blue), (2) coastline cells with saline water (orange), and (3) inland cells with saline water (red).  
253 Fresh inland cells are omitted in Figure 2. The median topographic gradient (dT) in coastline cells  
254 without saline water (just over 0.02) is one order of magnitude larger than in coastline cells with  
255 saline water (just over 0.002) (Fig 2 a)), mainly because saline water can only enter a model cell if  
256 the sea level is above the aquifer bottom (applies to 66% of coastline cells). The median hydraulic  
257 gradient (dH) in coastline cells without saline water (roughly  $10^{-3}$ ) is one order of magnitude larger  
258 than in cells with saline groundwater at the coastline and inland (Fig 2 b)). Besides topographic  
259 gradient (dT), hydraulic conductivity (K) seems to control the distribution of saline water inland,  
260 since hydraulic conductivity is much higher in inland cells containing saline water (Fig 2 c)).  
261 Further, cells with saline water tend to have a larger aquifer depth (Fig 2 d)) and groundwater  
262 recharge (GWR) can be much higher in fresh coastline cells than in cells with saline groundwater  
263 (Fig 2 e)).

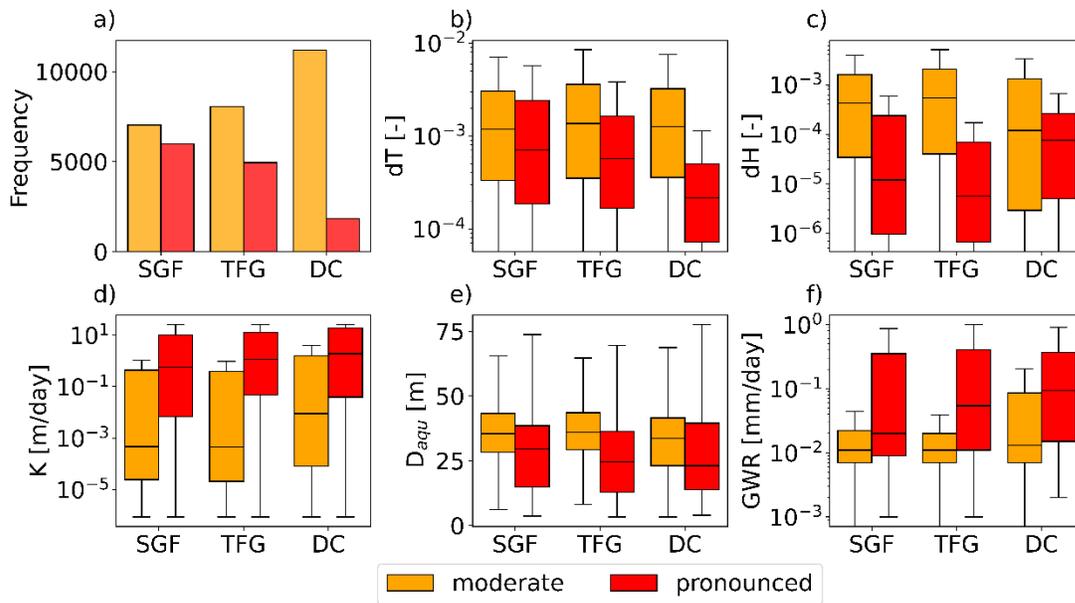


264  
 265 **Figure 2** – Boxplots of a) topographic gradient (dT), b) hydraulic gradient (dH), c) hydraulic  
 266 conductivity (K), d) aquifer depth ( $D_{aqu}$ ), and e) groundwater recharge (GWR) in coastline cells  
 267 (i.e., cells with one side facing an ocean) without saline groundwater (blue), coastline cells with  
 268 saline groundwater (orange), and inland cells with saline groundwater (red). Subplots a), b), c) and  
 269 e) are in logarithmic scale and hence do not show 0 on the y-axis (see Fig S5 for plot without  
 270 logarithmic scales).

271 **3.2 Several factors control coastal groundwater salinity at the continental scale**

272 We apply thresholds (Table 1) to categorize saline groundwater fraction (SGF), the thickness of  
 273 fresh groundwater (TFG) column, and the distance of saline groundwater from the coast (DC) into  
 274 moderate and pronounced to assess which parameters control the severity of an aspect of coastal  
 275 saline groundwater in a cell. Pronounced salinization appears in 14-46% of saline cells (Fig 3 a)).  
 276 Lower topographic gradients (dT) allow saline groundwater to intrude farther from the coast (DC)  
 277 (Fig 3 b)). Cells with lower hydraulic gradients (dH) are more often exposed to higher saline  
 278 groundwater fractions (SGF) and lower thicknesses of fresh groundwater columns (TFG) (Fig 3  
 279 c)). Higher hydraulic conductivity (K) increases the exposure to all three aspects of saline  
 280 groundwater, illustrated by the approximately two orders of magnitude between the median  
 281 hydraulic conductivity (K) in cells with moderate and pronounced aspects (Fig 3 d)). The  
 282 distributions of aquifer depth ( $D_{aqu}$ ) in cells with moderate and pronounced aspects of saline  
 283 groundwater are similar (Fig 3 e)). Groundwater recharge (GWR) values are higher in cells with  
 284 pronounced aspects of saline groundwater (Fig 3 f)). The usage of thresholds other than

285 those described in Table 1 leads to similar results (Fig S6 and Fig S7), and a scatterplot version of  
 286 Figure 3 can be found in Figure S8.



287  
 288 **Figure 3** – Subplot a) shows the number of cells with moderate (orange) and pronounced (red)  
 289 Saline Groundwater Fraction (SGF), Thickness of Fresh Groundwater (TFG), and Distance of  
 290 saline groundwater from Coast (DC). Aspects of saline groundwater in cells are categorized as  
 291 pronounced if  $SGF > 0.5$ ,  $TFG < 5$  m, and  $DC > 10$  km. The remaining subplots show boxplots of  
 292 cells with moderate and pronounced SGF, TFG, and DC for b) topographic gradient (dT), c)  
 293 hydraulic gradient (dH), d) hydraulic conductivity (K), e) aquifer depth ( $D_{aqu}$ ), and f) groundwater  
 294 recharge (GWR).

#### 295 4 Discussion

296 Simulating the North American groundwater salinity distribution, we find results that align with  
 297 the literature, identifying aquifers with lower topographic gradient to the coast (Michael et al.,  
 298 2013), lower hydraulic gradient (Ferguson and Gleeson, 2012), and larger aquifer thickness (Mazi  
 299 et al., 2013) as more vulnerable towards containing saline groundwater at shallower depth and  
 300 further from the coast. Figure 3 suggests that while saline groundwater can be expected in regions  
 301 with hydraulic gradients below  $10^{-3}$ , which has been used by Ferguson and Gleeson (2012), high  
 302 exposure to saline groundwater can be expected at hydraulic gradients below  $10^{-4}$ . Of the North  
 303 American coastline cells classified by Michael et al. (2013) as topography-limited and thus

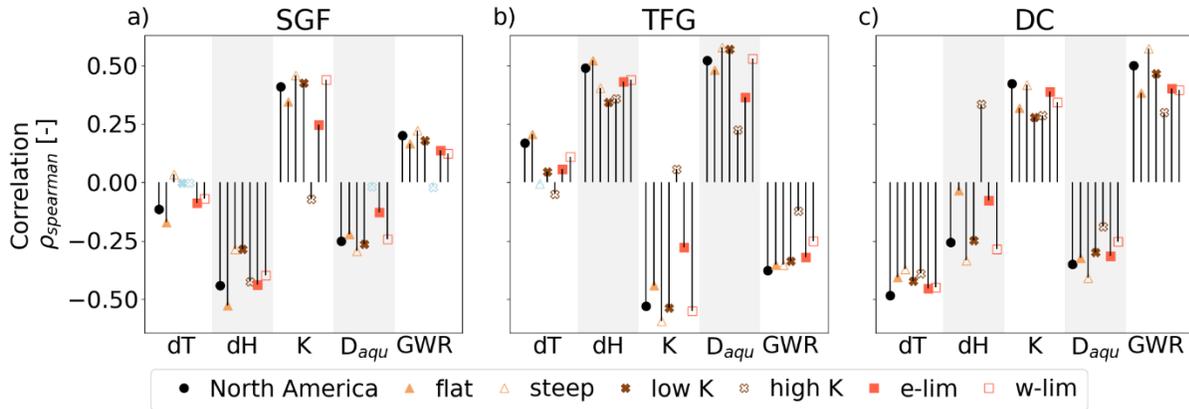
304 particularly vulnerable to seawater intrusion from sea level rise, 63% contain saline water in the  
305 presented simulation. In comparison, only 47% of cells in recharge-limited regions contain saline  
306 groundwater, which indicates that topography-limited cells may already (i.e., without sea level rise)  
307 be more likely to contain saline groundwater due to their relatively flat topography.

308 Our results show that inland cells with hydraulic conductivity ( $K$ ) above  $10^{-2}$  m day<sup>-1</sup> are  
309 particularly vulnerable to containing saline groundwater. This is consistent with our understanding  
310 that saline groundwater can be found where hydraulic conductivity is high enough for substantial  
311 groundwater flows (e.g., Shi and Jiao, 2014; Deng et al., 2017; Costall et al., 2020). However, the  
312 model does not contain conduits (e.g., in karstic or volcanic aquifers) or related focused  
313 groundwater exchanges between aquifers and the ocean (Kreyns et al., 2020), limiting its  
314 applicability in regions with such lithology. Surprisingly, groundwater recharge (GWR) values are  
315 higher in cells with pronounced aspects of saline groundwater (Fig 3 f)), indicating that more saline  
316 water spreads into regions with higher groundwater recharge. Such behavior has been reported for  
317 groundwater recharge below 100 mm yr<sup>-1</sup> (Michael et al., 2013). In cells with groundwater recharge  
318 above 100 mm yr<sup>-1</sup>, resilience against saline groundwater (i.e., SGF, TFG and DC) increases with  
319 increasing groundwater recharge (GWR) (Fig S9). However, the influence of groundwater recharge  
320 on the aspects of saline groundwater is low, potentially because topography is the main control in  
321 the applied groundwater model (Reinecke et al., 2024).

322 Across the North American continent, we identify coastal saline groundwater, in particular in  
323 Florida (US), along the US East Coast, and in Mexico (Fig S3), where issues with SWI have been  
324 reported (Barlow and Reichard, 2010). Additionally, we identified regions prone to containing  
325 saline groundwater, which have hardly been studied, in Alaska (US), Nunavut (Canada), and  
326 Oaxaca (Mexico). Using simple assumptions to estimate the vulnerability towards containing saline  
327 water (see Text S3), find that 23%/49% of the coastal area could be vulnerable due to low  
328 hydraulic/topographic gradients, while 68% of the coastal area could be vulnerable due to high  
329 hydraulic conductivities. However, parameter interactions limit the simulated area with saline  
330 groundwater to 18.6% (520 122 km<sup>2</sup>) of the coastal zone.

331 Although the influence of factors on the distribution of saline groundwater is evident, it does not  
332 show the full picture. Computing Spearman rank correlations of the factors with aspects of saline  
333 groundwater shows that weak to moderate monotonic relationships exist between most factors and  
334 aspects of saline groundwater (Fig 4). Scatter plots indicate non-monotonic relationships of factor

335 values with aspects of groundwater salinity (see Fig S8, Fig S10-S12), likely caused by factor  
 336 interactions not captured by Spearman rank correlations. Text S2 provides a detailed description of  
 337 Figure 4.



338  
 339 **Figure 4** – Spearman rank correlation of a) Saline Groundwater Fraction (SGF), b) Thickness of  
 340 Fresh Groundwater column (TFG), c) Distance of saline groundwater from coast (DC) with  
 341 topographic gradient (dT), hydraulic gradient (dH), hydraulic conductivity (K), aquifer depth  
 342 ( $D_{\text{aqu}}$ ) and groundwater recharge (GWR). Thresholds for the delineation of flat/steep, low/high K,  
 343 energy-/water-limited (e-/w-lim) regions are given in Table S2. Insignificant correlations are  
 344 shown in light blue. All correlations and p-values are displayed in Tables S4 to S6.

345 For saline water transport, the model’s spatial resolution of 5 arcmin is very coarse. Due to the  
 346 horizontal sharp interfaces applied, saline water entering a model cell on one side may cause the  
 347 horizontal interface to lift up. This shift in interface height entails that the saline water entering at  
 348 one time step may be transferred further to another neighboring cell in the next simulation step,  
 349 enabling saline water transport of several kilometers in just a year. Thus, the applied model likely  
 350 overestimates saline groundwater movement compared to real-world dispersive transfer. Despite  
 351 the likely faster movement of the saline interface, the model required 492 000 years to reach a state  
 352 of very slow interface movement, indicating that the saline groundwater distribution in North  
 353 America has been evolving since before the end of the last ice age, approximately 20 000 years  
 354 ago.

## 355 5 Conclusions

356 Given rapidly evolving coastal communities and growing demand for fresh groundwater in large  
 357 parts of North America, improving our understanding of continental coastal groundwater salinity

358 is pivotal. To assess the dominant controls of coastal saline groundwater occurrence and incursion  
359 at the continental scale, we have simulated variable density groundwater flow in North America  
360 until the sharp interface between fresh and saline water was stable under steady climatic forcing.  
361 Assessing the parameter values of fresh and saline cells at the coastline, we find that low  
362 topographic gradients and high aquifer depths enable saltwater to enter coastal aquifers. We show  
363 that coastline and inland cells are more vulnerable to containing saline groundwater if topographic  
364 gradients are lower and hydraulic conductivities are higher. Focusing on three aspects of coastal  
365 groundwater salinity, we show that under steady inputs, hydraulic gradient, topographic gradient,  
366 hydraulic conductivity, and aquifer depth control the salinity of coastal and inland cells. The impact  
367 of groundwater recharge seems to be limited in G<sup>3</sup>M-D. Our model results align with previous  
368 results identifying hydraulic conductivity as control in saline groundwater distribution. With  
369 hydraulic conductivities over  $10^{-2}$  m day<sup>-1</sup>, 68% of the North American coastal zone (i.e., up to 100  
370 km onshore and up to 100 m elevation) is, in principle, likely to carry saline groundwater. However,  
371 parameter interactions limit the simulated area with saline water to 18.6% of the North American  
372 coastal zone. Future research should assess the parameter interactions and use transient simulations  
373 to examine how changes in groundwater recharge and sea level rise impact seawater intrusion,  
374 particularly in regions with high hydraulic conductivities and low elevation.

### 375 **Data and Code availability**

- 376 • The code of G<sup>3</sup>M and G<sup>3</sup>M-D is available at: [https://github.com/rreinecke/global-gradient-](https://github.com/rreinecke/global-gradient-based-groundwater-model)  
377 [based-groundwater-model](https://github.com/rreinecke/global-gradient-based-groundwater-model)
- 378 • The North America model of G<sup>3</sup>M-D is available at:  
379 <https://github.com/EarthSystemModelling/3GM-D-NorthAmerica> (includes the code of  
380 G<sup>3</sup>M-D as a git submodule)
- 381 • The elevation data by Lehner et al. (2008) is available at:  
382 <https://www.hydrosheds.org/products/hydrosheds>
- 383 • The groundwater recharge data by Müller Schmied et al. (2020) is available at:  
384 <https://doi.pangaea.de/10.1594/PANGAEA.918447>
- 385 • The GLHYMPS 2.0 data (including hydraulic conductivity and effective porosity) by  
386 Huscroft et al. (2018) is available at:  
387 <https://borealisdata.ca/dataset.xhtml?persistentId=doi%3A10.5683/SP2/TTJNIU>

- 388 • The CoPerm data (used to set the hydraulic conductivity of the general head boundary) by  
389 Moosdorf et al. (2024) is available at: <https://doi.pangaea.de/10.1594/PANGAEA.958901>
- 390 • The depth to bedrock data (used to set aquifer depth) by Shangguan et al. (2017) is available  
391 at: <http://globalchange.bnu.edu.cn/research/dtb.jsp>
- 392 • The groundwater heads and interface heights of the final time step, which are evaluated in  
393 this study, are available at: [10.5281/zenodo.13928185](https://zenodo.org/doi/10.5281/zenodo.13928185)

394

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