

## Harmonised airborne laser scanning products can address the limitations of large-scale spaceborne vegetation mapping

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## **Abstract**

Vegetation structure data are essential for understanding the functioning of terrestrial ecosystems and for informing various science-policy interfaces. Recent years have seen a growing demand for high-resolution data on vegetation structure, driving the prediction of such metrics at fine resolutions (1 m - 30 m) at state, continental, and global scales by combining satellite data with machine learning. As these initiatives expand, it is crucial for the remote sensing and ecological communities to actively discuss the quality and usability of these products. Here, we (i) provide a brief overview of space-borne lidar missions measuring vegetation structure; (ii) using global canopy height models (CHMs) as an example, we demonstrate that predicted products exhibit significant errors exceeding natural changes in canopy height observed over a 10-year period, indicating that even a 10-year-old CHM derived from airborne laser scanning (ALS) is superior to currently available predicted CHMs; therefore, (iii) we recommend that regions with abundant ALS data prioritize harmonizing ALS-based vegetation metrics rather than relying solely on much less accurate predicted products derived from satellite data. We investigated the availability of ALS data in Europe and found that they are available for 26 countries, collected mostly between 2009 and 2024. We argue that, despite variations in data characteristics, including temporal inconsistencies and differences in point density and classification accuracy, the production of vegetation structure metrics, particularly CHMs, in raster format at fine resolution is both necessary and feasible. As new acquisitions are planned or underway, it is important to coordinate efforts to facilitate harmonization, develop continent-wide products, and ensure free access for research and policy communities. Beyond numerous ecological applications, such consistent benchmark datasets are crucial for calibrating future Earth Observation missions, making them essential for producing truly global, fine-resolution vegetation structure data.

## **Keywords**

Canopy, Earth observation, Forestry, Lidar, Validation, Vegetation structure

# 1. Introduction

Ecosystem structure – the spatial arrangement of biotic and abiotic elements that make up an ecosystem – is an Essential Biodiversity Variable (EBV) considered critical for understanding ecosystem function (Skidmore et al. 2021). In terrestrial ecosystems, the vegetation structure – the horizontal and vertical distribution of vegetation – is the key component. Vegetation structure plays a crucial role in modulating multiple ecosystem processes. In particular, it regulates energy flow, water cycling, carbon sequestration, and primary productivity (Murphy et al. 2022; LaRue et al. 2023a; Li et al. 2024). Furthermore, vegetation structure creates unique habitats that support species coexistence across different layers of vertical profile of the vegetation (Davies and Asner 2014; Moudrý et al. 2021; Wildermuth et al. 2023; Kempainen et al. 2024). The prevailing theory is that structurally complex vegetation stands are more effective at optimizing the incoming light and water resources, leading to better carbon assimilation (Atkins et al. 2018; Seidel and Ammer 2023), and that they provide a greater number of ecological niches, thereby enhancing biodiversity (Tews et al. 2004; Stein et al. 2014; Torresani et al. 2020; Coverdale and Davies 2023).

Remote sensing technologies such as Light Detection and Ranging (lidar) have played a key role in addressing knowledge gaps, providing a way to accurately map vegetation structure from local to global scales (Herold et al. 2019; Valbuena et al. 2020; Jutras-Perreault et al. 2023; Liu et al. 2023; Moudrý et al. 2023; Rosen et al. 2024). Particularly, lidar sensors onboard planes (i.e., airborne laser scanning; ALS) have considerably advanced national monitoring programs (e.g., Assmann et al. 2022; Kissling et al. 2023) and robust approaches to convert ALS data into structural metrics are available (Fischer et al. 2024). However, while the costs of ALS have decreased in recent years, large continuous coverage exists only in a few regions, mostly in Europe, North America, and Australia, as well as countries such as New Zealand and Japan, and only a few countries have mapped their entire territory more than once.

Recent advances in space-borne lidar missions, such as Global Ecosystem Dynamics Investigation (GEDI) and the Ice, Cloud, and land Elevation Satellite-2 (ICESat-2), help address the spatial and temporal limitations of ALS data (Markus et al. 2017; Dubayah et al. 2020). These missions provide free data that have enabled the creation of global models of vegetation structure (e.g., Mulverhill et al. 2022; Burns et al. 2024), supporting innovative and impactful research. For instance, vegetation structure products derived from space-borne lidar data have been used to monitor forest/woodland structure and regrowth (Milenković et al. 2022; Jucker et al. 2023; Stritih et al. 2023), track carbon losses from disturbances (Holcomb et al. 2024), evaluate the effectiveness of protected areas from the perspective of carbon stocks and vegetation structure (Ceccherini et al. 2023; Lang et al. 2023;

34 [Liang et al. 2023](#), [Brodie et al. 2023](#)), and to assess species diversity and species-environment  
35 relationships ([Marselis et al. 2022](#); [Smith et al. 2022](#); [Vogeler et al. 2023](#); [Xu et al. 2024](#)). However,  
36 space-borne lidar data are spatially and temporally limited, and their derived products, such as global  
37 canopy height models (CHMs), suffer from accuracy issues ([Mandl et al. 2023](#); [Moudry et al. 2024](#)),  
38 which impair their applicability ([Hakkenberg et al. 2023](#)).

39 Consistent data on vegetation structure is essential for informing multiple science-policy interfaces.  
40 For instance, the Ecosystem Vertical Profile, which refers to the vertical distribution of biomass, is one  
41 of the 21 EBVs defined by the Group on Earth Observations Biodiversity Observation Network ([Pereira  
42 et al. 2013](#)), setting key data requirements for global, UN-level, biodiversity change monitoring and  
43 reporting (e.g., in the scope of Global Biodiversity Observing System, [GEO BON 2022](#)). Vegetation  
44 structure data also support the United Nations' System of Environmental-Economic Accounting  
45 ([United Nations 2021, 2022](#)) and play a role in tracking progress toward Global Biodiversity Framework  
46 targets ([Skidmore et al. 2021](#)). Additionally, such data are vital for assessing restoration success, as  
47 emphasized in the Nature Restoration Law and EU Forest Strategy for 2030 ([Ruiz-Jaen and Adie 2005](#);  
48 [LaRue et al. 2023b](#)).

49 As new mapping and modeling initiatives develop data products on vegetation structure to meet the  
50 monitoring needs of emerging science-policy interfaces, it is crucial for the remote sensing and  
51 ecological communities to actively discuss the usability of these products and establish plans to  
52 measure, avoid, and correct their limitations. In this article, we aim to (i) provide a brief overview of  
53 space-borne lidar missions that measure vegetation structure, (ii) examine the limitations of these  
54 data, using CHMs as a case study, and (iii) propose a path toward improving continental and global  
55 CHMs through the harmonisation of airborne laser scanning products.

## 56 **2. Mapping vegetation structure with space-borne lidar**

57 Details and examples of the usability and advantages of lidar remote sensing for mapping vegetation  
58 structure can be found in [Lefsky et al. \(2002\)](#), [Bergen et al. \(2009\)](#), and [Moudry et al. \(2023\)](#). Simply  
59 put, lidar is ideal for measuring vegetation structure because it can penetrate through the gaps in the  
60 vegetation, capturing its vertical structure as well as the shape of the terrain underneath. Lidar sensors  
61 can be installed on various platforms, including tripods, backpacks, cars, drones, helicopters, planes,  
62 and satellites. Notably, space-borne lidar is especially valuable for large-scale mapping due to its  
63 consistent and extensive global coverage. Satellite sensors are expected to become the primary data  
64 source for mapping the vegetation structure in response to international monitoring requirements.

65 Yet, technical issues concerning data coverage and accuracy persist ([Hancock et al. 2021](#); [Liu et al.](#)  
66 [2021](#)).

67 The first global dataset characterizing canopy structure was obtained from the Geoscience Laser  
68 Altimeter System (GLAS) onboard the Ice, Cloud, and land Elevation Satellite (ICESat), operational from  
69 2003 to 2009 ([Abshire et al. 2005](#)). This mission was primarily intended for measuring polar ice caps  
70 but it also enabled the development of datasets for ground elevation and canopy height ([Schutz et al.](#)  
71 [2005](#); [Simard et al. 2011](#)). In 2018, NASA launched two space-borne lidar missions, ICESat-2 ([Markus](#)  
72 [et al. 2017](#)) and GEDI ([Dubayah et al. 2020](#)), aimed among other objectives at providing global data on  
73 terrain and canopy height.

74 A major limitation of space-borne lidar sensors is that they collect data along discrete transects,  
75 providing discrete and dislocated sample data of vegetation structure. Additionally, various factors,  
76 including atmospheric conditions, solar background photons, laser pulse energy, and topography,  
77 affect data accuracy and require filtering, resulting in significant reductions in available data (e.g.  
78 [Hayashi et al. 2013](#); [Moudrý et al. 2022](#); [2024b](#)). Although current lidar space-borne missions have  
79 substantially higher sampling densities than their predecessors, the data are still insufficient for  
80 producing high-resolution continuous CHMs, and the derived data products have limited spatial and  
81 temporal coverage. For instance, GEDI was expected to sample only 4% of the Earth's land surface  
82 over a two-year mission, enabling the production of a near-global (from -52° to 52° latitude)  
83 vegetation structure metrics, including CHM at three spatial resolutions: 1 km, 6 km, and 12 km ([Burns](#)  
84 [et al. 2024](#)). A global-scale 1 km resolution canopy model is a significant achievement but has limited  
85 utility for applications such as ecosystem mapping or species-environment relationship assessments,  
86 which typically require finer resolution ([Smith et al. 2022](#); [Anderle et al. 2023](#); [Davison et al. 2023](#);  
87 [Vogeler et al. 2023](#)).

88 A potential solution to provide global fine-resolution data on the vertical structure of vegetation for  
89 the entire world at a reasonable cost could lie in the creation of a fleet of lidar satellites that would  
90 continuously map the Earth ([Hancock et al. 2021](#); [Lowe et al. 2024](#)). [Hancock et al. \(2021\)](#) estimated  
91 that producing such continuous data at a 30 m resolution every 5 years would require a constellation  
92 of twelve satellites acting concurrently. More recently, [Lowe et al. \(2024\)](#) investigated which platform-  
93 optics-constellation design combination offers the most promising and cost-effective solution, and  
94 suggested that micro-satellites, with a mass on the order of 150 kg, may present the most attractive  
95 performance-to-cost ratio. They estimated that a constellation of eight such satellites would be  
96 sufficient to produce CHMs at a 20-meter resolution annually. In addition, NASA currently has an  
97 advanced proposal for the next-generation space-borne laser altimetry mission, known as Earth

98 Dynamics Geodetic Explorer (EDGE), which aims to significantly improve spatial and temporal  
99 coverage. At present, however, such satellite systems do not exist, and an operational mission capable  
100 of delivering higher-resolution global lidar data is unlikely before 2030. Hence, in the meantime,  
101 alternative solutions must be explored to meet immediate monitoring needs.

### 102 **3. Predicted continuous high-resolution CHMs and their limitations**

103 The lack of global high-resolution data on vertical vegetation structure has stimulated the use of  
104 space-borne lidar data in combination with other satellite products to generate predicted models of  
105 vegetation structure variables, such as canopy height, total canopy cover, and foliage height diversity  
106 at fine resolutions, such as 10 m or 30 m (e.g., [Kacic et al. 2021, 2023](#); [Schwartz et al. 2023](#); [Diaz-Kloch  
107 and Murray 2024](#)). A common approach to produce the high resolution, wall-to-wall data on  
108 vegetation structure is to train predictive models that combine direct but discrete height  
109 measurements (e.g., from space-borne lidar ICESat, GEDI, ICESat-2) with spatially continuous data  
110 (e.g., from space-borne optical and radar data). The model establishes a relationship between the  
111 discrete and the continuous data that enables the estimation of the height at locations not directly  
112 measured by lidar ([Bergen et al. 2009](#); [Lefsky 2010](#)). Predicted CHMs are among the most common  
113 and, so far, the only high-resolution vegetation structure products available at continental ([Liu et al.  
114 2023](#)) and global scales ([Potapov et al. 2021](#); [Lang et al. 2023](#)), making them suitable for illustrating  
115 the pros and cons of such data. Recently, a web application has been developed, making this approach  
116 easily accessible ([Alvites et al. 2024](#)).

117 Several continental or global predicted CHMs have been produced. The first such dataset was  
118 developed by [Lefsky \(2010\)](#), who combined canopy heights derived from GLAS with MODIS data to  
119 produce a patch-based global CHM. Similarly, [Simard et al. \(2011\)](#) used the relationships between the  
120 GLAS-derived canopy heights and multiple environmental variables (e.g., tree cover, climate, altitude)  
121 to predict a global model of canopy heights at a 1 km spatial resolution. Recently, [Potapov et al. \(2021\)](#)  
122 and [Lang et al. \(2023\)](#) used optical satellite data (Landsat, Sentinel-2) trained on GEDI measurements  
123 to create global CHMs at 30 m and 10 m spatial resolutions, respectively. Likewise, [Liu et al. \(2023\)](#)  
124 used canopy height from airborne laser scanning (ALS) data and PlanetScope imagery to predict  
125 canopy heights in Europe at the 3 m resolution. Finally, Meta, in cooperation with the World Resources  
126 Institute, combined high-resolution data from optical satellites, ALS, and GEDI to develop a global CHM  
127 at a 1 m resolution ([Tolan et al. 2024](#)).

128 The main benefit of predicted CHMs based on space-borne lidar lies in their easy availability, especially  
129 as there are no alternatives at scales beyond the regional level. These CHMs are usually readily

130 available as open data in a raster format, allowing researchers to use them as input data in common  
131 GIS software to inform their analyses. This contrasts with the high data volume, time-consuming, and  
132 often challenging process of working with more accurate ALS point clouds, which can be difficult for  
133 many ecologists to store and handle (Moudrý et al. 2023; Kissling and Shi 2023; Wang et al. 2024).  
134 However, the easy accessibility of such predicted CHMs is both a blessing and a curse as users may be  
135 unaware of data limitations, and the reliability of predicted global datasets is questionable  
136 (Duncanson et al. 2019; Meyer and Pebesma 2022). Modeling canopy height is a complex process that  
137 involves errors and biases from multiple sources, ranging from ground detection with space-borne  
138 lidar to the saturation of optical data in closed-canopy forests. Indeed, independent validation studies  
139 showed that the accuracy of these satellite-derived global CHMs is low (e.g., Bolton et al. 2013; Pascual  
140 et al. 2022), and their use in biodiversity modeling leads to erroneous results (Moudrý et al. 2024a).

### 141 **3.1 Validation of predicted global CHMs accuracies**

142 Here we evaluate four recent predicted CHMs (Table 1). We used a 2022 ALS scan from the Giant  
143 Mountains National Park (Czechia) as a reference and compared it both to satellite-derived height  
144 models and to a second ALS scan acquired 10 years earlier, in 2012 (all ALS data were processed with  
145 standard methods, cf. Moudrý et al. 2024a). Strikingly, we found that the 2012 ALS data had a much  
146 lower error in predicting 2022 canopy height than any of the global or regional CHMs. Figure 1  
147 presents a cross-section comparison of vegetation heights extracted from four predicted space-borne  
148 data-based CHMs to reference heights extracted from ALS CHMs. Both large over- and  
149 underestimation of vegetation height can be observed in space-borne CHMs (Figure 1; see Moudrý et  
150 al. (2024a) for comparison across more sites). Moreover, the change in vegetation height over ten  
151 years is lower than the canopy height error in the four models for the selected area (Figure 2). In this  
152 specific case study, such an error hinders effective change detection analyses on the canopy height.  
153 This limitation is influenced by the magnitude of disturbance, which is relatively low in our study area  
154 (Figure 2). In cases of large-scale deforestation detectable by optical data, the signal should still be  
155 visible in predicted CHMs. However, to estimate changes in vegetation height accurately, we must  
156 first know the height before such disturbances occur.

### 157 **3.2 What to report and consider when generating and using CHMs**

158 Global datasets are indispensable for answering large-scale ecological questions, so it is imperative to  
159 improve reporting of the accuracy and uncertainty of the predicted CHMs to enable users to easily  
160 select the most appropriate map for their purposes. To select the best vegetation structure product,  
161 the overall evaluation metrics, such as mean error (ME) or root mean square error (RMSE), provided



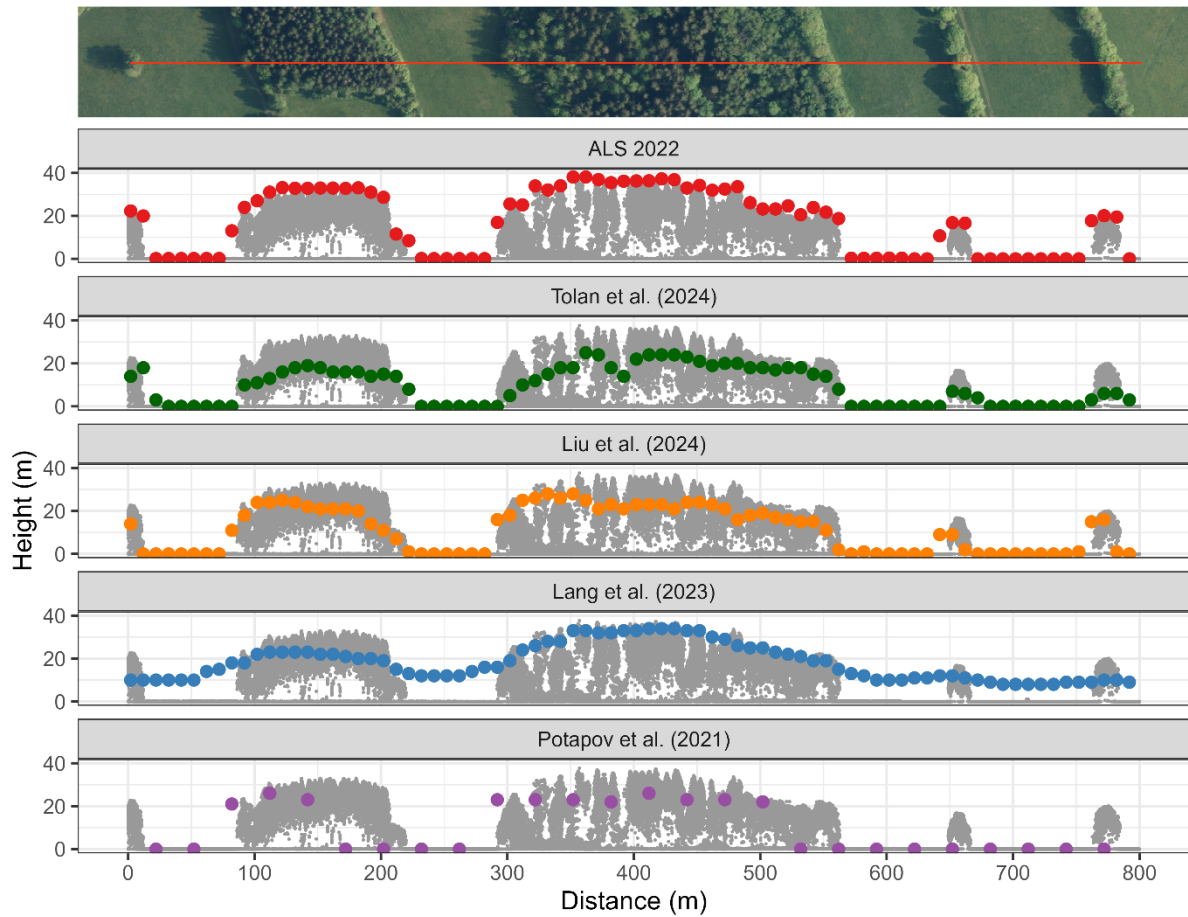
162 by existing products are fundamental, despite providing a limited insight into the local map quality.  
 163 Even if the user selects the most accurate map (i.e. that with the lowest overall RMSE), there may be  
 164 considerable biases in the subregions. However, these uncertainties are usually not quantified; if they  
 165 are, it is often unclear how this was done (but see [Lang et al. 2023](#)) or the uncertainty estimates are  
 166 inaccurate ([Moudrý et al. 2024a](#)). One way to improve the reporting of uncertainties is to assess the  
 167 area to which a prediction can be reliably applied, such as estimating the Area of Applicability ([Meyer  
 168 and Pebesma 2021](#)). Furthermore, we suggest that in addition to the common validation using ALS  
 169 data, the predicted CHM products should include representative profiles (as in Figure 1). Most authors  
 170 of the global datasets only showed product visualizations in 2D space (e.g., [Potapov et al. 2021](#); [Lang  
 171 et al. 2023](#); [Schwartz et al. 2023](#)). Even inaccurate CHM models look very plausible in a 2D visualization,  
 172 and even an experienced user is easily misled into the false impression that they accurately represent  
 173 the reality (Figure 2).

174

175 Table 1. Predicted global canopy height models (CHMs) evaluated in this study. The Root mean square  
 176 error (RMSE) value reported by the authors of individual CHMs in the original publications is presented  
 177 here.

Global CHM	Author	Resolution	RMSE	Valid for year
Global forest canopy height	Potapov et al. (2021)	30 m	9.1 m	2019
High-resolution canopy height model of the Earth	Lang et al. (2023)	10 m	2.8 - 9.6 m	2020
Canopy height map for Europe	Liu et al. (2023)	3 m	4.3 - 6.4 m	2019
Global map of tree canopy height	Tolan et al. (2024)	1 m	4.4 m	2018-2020

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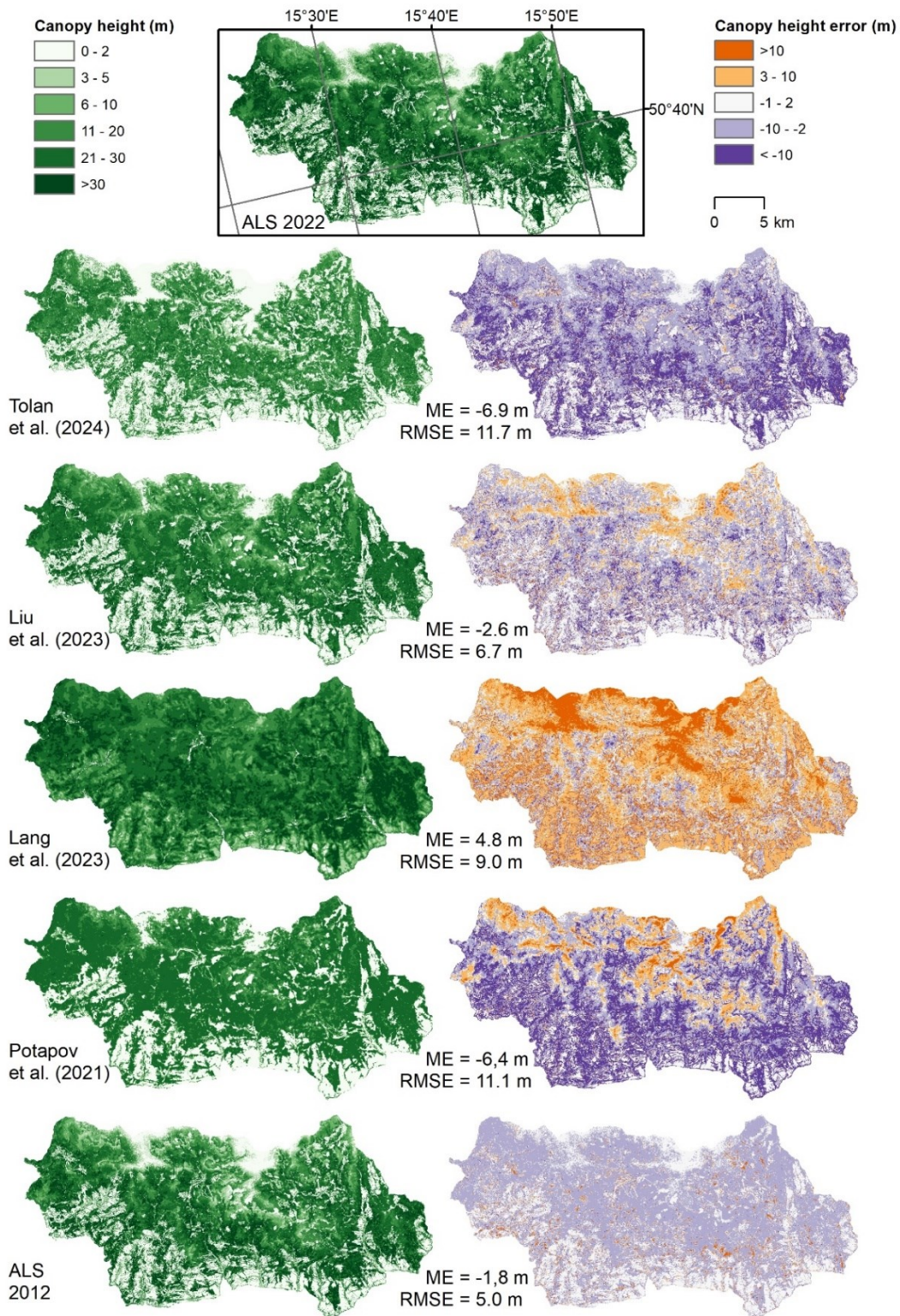
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**Figure 1.** A representative canopy height profile from the Giant Mountains National Park, Czechia. The profile spans 10 meters in width. Note the limited ability, especially in the CHM by Lang et al. (2023), to capture variations in canopy height, making the transition between forest and non-forest areas unclear. The mosaic of pastures and forests appears as a continuous forest with heights ranging from 10 to 30 meters. In contrast, the CHMs by Tolan et al. (2024) and Liu et al. (2024) more effectively differentiate between forest and non-forest areas due to the substantially higher resolution of their input data. However, both CHMs tend to underestimate the height of vegetation. This suggests that there may be room for improvement in combining multiple predicted CHMs, such as the one by Tolan et al. (2024), which accurately distinguishes forests from non-forested areas, and the model by Lang et al. (2023), which is relatively successful in predicting top canopy height. The CHM by Potapov et al. (2021) has a resolution that is too coarse to capture smaller stands.



190

191 **Figure 2.** Canopy height from six different sources in the Giant Mountains National Park (Czechia). This includes  
 192 airborne laser scanning (ALS) data from 2022 (reference dataset) and 2012, along with four predicted canopy  
 193 height maps: Tolan et al. (2024), Liu et al. (2023), Lang et al. (2023), and Potapov et al. (2021). The figures on the  
 194 left show canopy height, while the figures on the right show the difference in canopy height compared to the  
 195 ALS 2022 data (i.e., the error of the predicted maps). ME stands for Mean Error, and RMSE stands for Root Mean  
 196 Square Error.

#### 197 **4. Towards improved continental to global canopy height models**

198 Unlike space-borne laser altimeters, which offer broader coverage but discrete and sparse  
199 measurement (Dubayah et al. 2020), ALS offers dense continuous coverage and is commonly used for  
200 regional or state-wide mapping. However, processing of ALS point clouds and their integration into a  
201 single product is challenging for larger-scale analyses covering multiple countries (Fischer et al. 2024).  
202 As a result, large-scale studies generally rely on global predicted products (see Section 3) due to the  
203 difficulties in managing and processing ALS data at a continental scale.

204 Even if we manage to build a satellite lidar system capable of dense spatio-temporal coverage in the  
205 near future (see Section 2), it will need precise and consistent benchmark datasets over broad  
206 geographical domain for its calibration. ALS data are indispensable for such a purpose (e.g. Duncanson  
207 et al. 2019; Tang et al. 2023). Moreover, knowing the current status of vegetation structure is  
208 essential, and such data will be useful for assessing changes in vegetation structure (Guerra-  
209 Hernández and Pascual 2021; Parra and Simard 2023). Therefore, locally available data should be used  
210 to produce uniform seamless vegetation structure products.

211 In the United States, the 3D Elevation Program (3DEP), managed by U. S. Geological Survey (USGS),  
212 has been launched, aiming to collect ALS data for conterminous United States. So far, however, it only  
213 aims at providing a digital terrain model (DTM; Stoker 2020; Stoker and Miller 2022), not seamless  
214 CHMs (or other vegetation structure products) in a raster format. Digital surface models (DSMs) can,  
215 however, be created through OpenTopography (<https://opentopography.org/>) where most U.S. lidar  
216 data are also hosted. Europe is further behind, as no common data collection protocol or methodology  
217 that regulates mapping activities exists. This responsibility falls onto its Member States. As a result,  
218 ALS coverage in Europe is managed at the national (e.g., in Denmark, France, Poland, Spain) or sub-  
219 national (e.g., in Austria, Belgium, Germany, Italy) level and data are scattered among providers,  
220 leading to different characteristics across regions (D'Amico et al. 2021; Kakoulaki et al. 2021).

221 Given the substantial investments required for country-wide ALS data acquisition, there is a clear need  
222 for continued efforts to ensure its effective use in vegetation mapping. Of the 44 countries in Europe,  
223 ALS data is collected by governmental institutions in at least 26 countries (Table 2). Although the data  
224 have different characteristics (such as point density and accuracy), it should be possible to derive  
225 vegetation structure characteristics in raster format at a relatively fine resolution. This, however,  
226 requires standardized processing pipelines that can account for differences in scanning properties  
227 (Fischer et al. 2024) and detailed documentation of the metadata, particularly regarding the  
228 acquisition time. However, metadata, if available, are documented with various degrees of depth and  
229 reliability, which significantly limits their accessibility and utility for potential users. For example, we



230 made every effort to review the characteristics of ALS data available in Europe (Table 2), but we had  
231 to limit our focus to point density and the year of data acquisition. This was due to the difficulty of  
232 narrowing down the acquisition time to the exact month, the lack of announcements regarding future  
233 acquisitions, and the absence of information on the classification categories and methods used to  
234 classify them. In addition, in many cases, accessing the data itself is difficult, as point clouds are still  
235 not freely available in several European states. Furthermore, although ALS data have been collected  
236 over multiple periods in some areas of Europe (Table 2), only the point clouds from the most recent  
237 period are easily accessible. Hence, it is important to adhere to the FAIR guiding principles for scientific  
238 data management and stewardship, which emphasize findability, accessibility, interoperability, and  
239 reusability (Wilkinson et al. 2016).

240 At the moment, only a few European countries provide ALS-derived metrics of vegetation structure in  
241 a raster format (e.g., Assmann et al. 2022; Kissling et al. 2023; Külling et al. 2024; Shi and Kissling 2024).  
242 However, the choice of vegetation metrics, the methods used to calculate them, and their resolution  
243 can vary significantly (Moudrý et al. 2023; Kissling and Shi 2023; Wang et al. 2024). Therefore, it is  
244 important to coordinate these efforts from the outset to enable their harmonization and the  
245 development of a Europe-wide product. This is particularly relevant to some of the objectives set by  
246 the European Commission in the new EU Forestry Strategy 2030, such as monitoring old-growth  
247 forests using remote sensing. Such standardization would ensure consistent interpretation and  
248 utilization of data across various studies and applications, and improve the reliability and  
249 reproducibility of results, enabling comparable assessments of vegetation characteristics.

250 Access to funding will be a crucial factor in this effort. A European funding initiative similar to 3DEP,  
251 led by the EU, would be a good approach to generate vegetation structure metrics from existing data  
252 and to collect data in European countries where ALS data is not yet available or where only limited  
253 coverage exists, such as the Balkans, Hungary, and Moldova. The first step may lie in creating a  
254 continental CHM, which would require consideration of the following aspects of lidar point clouds that  
255 present challenges to achieving consistent and reliable products: (i) inconsistencies in pulse density  
256 and classification accuracy across datasets/countries, and (ii) temporal inconsistency (e.g., scans with  
257 differences in acquisitions in the order of several years or scans conducted in leaf-on vs leaf-off  
258 periods).

259 Combining lidar data from different instruments is challenging, as vegetation structural metrics can  
260 differ due to the variations in acquisition parameters, such as point density (Roussel et al. 2017; Zhang  
261 et al. 2024, Fischer et al. 2024). Furthermore, point densities may considerably differ depending on  
262 when the data were collected (i.e., due to considerable differences in pulse repetition frequencies of

263 current and older scanning instruments) and whether the focus of data acquisition was primarily on  
264 terrain or also on vegetation (i.e., winter vs summer acquisition) and infrastructure (e.g., buildings,  
265 power lines). However, such inconsistencies can be compensated for by selecting a reasonable  
266 resolution of the final product and accounting for temporal differences through consistent metadata.  
267 The typical point densities of lidar point clouds available in Europe are around 1-5 points per square  
268 meter; still, they vary considerably across the continent (Table 2). For the low point densities, it is  
269 advisable to calculate vegetation metrics at coarser resolutions (e.g., 20 meters) to minimize potential  
270 errors in estimating the vegetation structure (Ruiz et al. 2014; Wilkes et al. 2015; D'Amico et al. 2021).  
271 On the other hand, vegetation metrics, such as upper percentiles of height, are generally less sensitive  
272 to point cloud properties (Roussel et al. 2017; LaRue et al. 2022; Fischer et al. 2024), and deriving a  
273 canopy height model at a 10-meter resolution should provide a reasonable balance between spatial  
274 resolution and vertical accuracy. The point cloud classification across countries, with differences in  
275 methods such as automated classification, visual inspection, and AI algorithms, constitute another  
276 factor. While classes like terrain, vegetation, and buildings are commonly classified, power lines,  
277 bridges, and viaducts are rarely included, potentially introducing bias in vegetation metrics (Shi and  
278 Kissling 2023). However, this may not be a major issue if the focus is primarily on forests, where even  
279 less accurate classifications can still provide better results than predicted CHMs.

280 Lastly, the temporal inconsistency of ALS data acquisition across countries is a concern, as ALS surveys  
281 remain costly and infrequent. This can limit the usability and accuracy of harmonized vegetation  
282 structure maps. Furthermore, ALS data collection often predominantly aims to provide accurate  
283 topographic modeling, so many countries carry out scans under leaf-off conditions (such as Slovenia).  
284 However, in some countries, scanning is explicitly timed to occur close to peak vegetation greenness  
285 (such as France), while in other countries, it depends on the region (e.g., Spain). Some countries may  
286 even merge point clouds across different scanning periods. If not accounted for, the resulting  
287 differences in phenology (leaf-off vs. leaf-on scanning) could introduce substantial bias into ALS-  
288 derived canopy metrics. In addition, as new advancements in scanning technology emerge (e.g., higher  
289 pulse repetition frequency, use of multiple wavelengths), older datasets can become less compatible  
290 with current data, making it challenging to ensure compatibility. If we consider European countries  
291 where data have already been collected or are in the process of being collected, the time span  
292 between the first and last scans amounts to about 15 years (2009-2024; Table 2). While this is not  
293 optimal, a 10-year difference introduces (as illustrated above) significantly less uncertainty than the  
294 predicted maps (Figure 2). It also indicates that with a coordinated effort, it should be possible to  
295 achieve a better temporal range similar to the US 3DEP (9 years) for the entire continent of Europe  
296 within this decade.

**Table 2.** Airborne laser scanning campaigns conducted by governmental institutions in Europe.

Country	Number of coverages	Density points/m <sup>2</sup>	Years of acquisition
Austria (Burgenland)	1	unknown	2019
Austria (Carinthia)	1+	4-8	2006-2015; 2022-ongoing
Austria (Lower Austria)	1	unknown	2011-2022
Austria (Upper Austria)	3	1; 4; 8	2003-2011; 2012-2020; 2021-ongoing
Austria (Salzburg)	2	1; 4-16	2006-2013; 2016-2023
Austria (Styria)	1+	2-4	2008-2014; 2022-ongoing
Austria (Tyrol)	2	0.25-4 (first scan)	2005-2010; 2017-2023
Austria (Vorarlberg)	4	1 (first scan)	2002-2004; 2011; 2017; 2023
Austria (Vienna)	1+	15-20	2007; ongoing updates
Belgium (Brussels)	2	30; 67	2012; 2021
Belgium (Flanders)	2	0.25; 16	2001-2004; 2013-2015
Belgium (Wallonia)	2	0.8; 7	2013-2015; 2021-2022
Czechia	1	1	2009-2013
Denmark	3	0.5; 4-5; 8-10	2007, 2014-15, 2018-23
England	1	2-4	2017-2023
Estonia	3+	0.5 (2008-2015); 2 (18 urban areas)	2008-2011; 2012-2015; 2017-2020; 2021-ongoing
Finland	2	0.5; 5	2008-2019; 2020-25
France	0+	10	2021-2026
Germany (Baden-Württemberg)	2+	0.8; 8; 8	2000-2005; 2016-2022; 2022-ongoing
Germany (Bavaria)	1	4	2013-2024
Germany (Berlin)	1	9.8	2021
Germany (Brandenburg)	1+	1; 5	2008-2012; 2017-ongoing
Germany (Bremen)	1	7	2012-2017
Germany (Hamburg)	3	15-30	2010; 2020; 2022
Germany (Hessen)	2+	4	2007-2014; 2015-2021; 2022-ongoing
Germany (Lower Saxony)	1	4	2014 -2022
Germany (Mecklenburg-Vorpommern)	1	5	2016-2023
Germany (North Rhine Westphalia)	1	4-10	2018-2023
Germany (Rhineland Palatinate)	1+	4	2018-2023
Germany (Saarland)	1	8	2015-2016
Germany (Saxony)	2+	1-5; 5-19; 12-18	2005-2012; 2015-2020; 2020-ongoing
Germany (Saxony-Anhalt)	0+	3-5	2015-ongoing
Germany (Schleswig Holstein)	1	3-4	2005-2007

Germany (Thuringia)	3+	0.05-1.8; 4 (since 2010)	1996-2006; 2010-2013; 2014-2019; 2020-ongoing
Ireland	0+	various	small areas since 2006
Italy	0+	0.4-6	2004-2017
Latvia	1	4	2013-2019
Lichtenstein	2+	minimum of 5 (15-20)	2000-2007*; 2017-2023; 2024-2029
Lithuania	2	6.5	2009-2010; 2019-2022
Luxembourg	2	15	2017; 2019
Malta	2	4; 40	2012; 2018
Netherlands	4+	at least 8-10 since second acquisition	1997-2004; 2007-2012; 2014-2019; 2020-2022; 2023-2025
Northern Ireland	0+	16	2022-ongoing
Norway	1	0.5-5	2009-2018
Poland	1	4-12	2010-2019
Portugal	0+	10	Planned between 2023-2024
Romania	0+	2-8	2004-2025
Scotland	0+	1-2	2011-2022
Slovakia	1+	≥ 15	2017-2022; 2022-2034 (ongoing)
Slovenia	1	2-10	2011, 2014-2015; 2023-ongoing
Spain	2+	0.5-2; 0.5-4; 5	2009-2015; 2015-2022; 2022-25
Sweden	1+	0.25-1; 1-2	2009 - 2017; 2018 - 2024 (ongoing)
Switzerland	2+	minimum of 5 (15-20)	2000-2007; 2017-2023; 2024-2029
Wales	1	2-4	2020-2024

298 Note that the table is incomplete both in terms of available data and their metadata, as these are documented  
299 to varying degrees and reliability. In addition, the years of acquisition may also include the preparation and  
300 processing time (i.e., +/- 1 year), as it is often difficult to distinguish whether only the acquisition years are  
301 reported or if they also include data processing. Similarly, it is difficult to distinguish whether point or pulse  
302 density is reported. Therefore, we use the single term point density.

## 303 5. Conclusions

304 The availability of remote sensing data greatly facilitates ecological research. On the other hand, the  
305 growing number of datasets of varying quality introduces challenges regarding which datasets to  
306 choose. Users typically do not have the chance (and/or expertise) to critically evaluate the available  
307 data. It is, therefore, essential to ensure that data producers clearly communicate the limitations of  
308 their datasets. Furthermore, the remote sensing community must ensure the availability of the most



309 appropriate data for ecology, forestry and climate change research. For vegetation structure,  
310 harmonized continental products derived from ALS data are the key. The main challenges in  
311 developing such products lie in limited spatial and temporal coverage, inconsistencies in point  
312 densities, and, to some degree, the accuracy of classification methods. Therefore, to ensure the  
313 effective use of ALS in vegetation mapping across Europe, establishing a common data collection  
314 protocol to regulate mapping activities (e.g., time of acquisition, pulse densities, updating period) as  
315 well as metadata reporting is needed. Besides, studies should focus on developing standardized  
316 processing pipelines to account for differences in ALS point clouds, as well as on creating methods for  
317 data fusion that leverage both space-borne and ALS data to enhance vegetation monitoring. We  
318 strongly recommend that ALS-rich regions such as Europe and the United States prioritize the  
319 production of ALS-based canopy height maps over relying solely on modeled global data. In addition,  
320 such harmonized data will provide a benchmark for calibrating space-borne laser altimetry products.  
321 The improved harmonization will result in better ecosystem monitoring, climate change modeling,  
322 and forest management on both continental and global scales.

323

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## 344 **Author Contributions Statement**

345 VM, RR, MF, MT, GVL, and VB conceived the ideas; VM and ES analysed the data; VM led the writing  
346 of the manuscript. All authors contributed critically to the drafts and gave final approval for  
347 publication.

## 348 **Data accessibility**

349 The evaluated global canopy height models, that is, the global forest canopy height (Potapov et al.,  
350 2021), the high-resolution canopy height model of the Earth (Lang et al., 2023), and the global map of  
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352 International License. The canopy height model of Europe at 3 m resolution was kindly provided by  
353 Liu et al. (2023). The canopy height models of Giant Mountains National park (2012 and 2022), derived  
354 from airborne laser scanning and used for evaluation of global canopy height models, are available  
355 from Zenodo: <https://doi.org/10.5281/zenodo.14270020>.

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