

1 Untangling microfibrils: Pervasive plastic pollution in submarine
2 canyons

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17 **Ed Keavney¹, Ian A. Kane², Michael A. Clare³, David M. Hodgson¹, Veerle A.I. Huvenne³,**
18 **Esther J. Sumner⁴, Jeff Peakall¹, Furu Mienis⁵, and Jonathan Kranenburg⁵**

19 *¹ School of Earth and Environment, University of Leeds, Leeds, LS2 9JT, UK*

20 *² Department of Earth and Environmental Science, University of Manchester, Manchester, M13*
21 *9PL, UK*

22 *³ Ocean BioGeoscience, National Oceanography Centre, Southampton, SO14 3ZH, UK*

23 *⁴ School of Ocean and Earth Science, University of Southampton, Southampton, SO14 3ZH, UK*

24 *⁵ Department of Ocean Systems, Royal Netherlands Institute for Sea Research (NIOZ-Textel), Den*
25 *Burg, The Netherlands*

26 Corresponding author: Ed Keavney (ed.keavney@gmail.com)

27

28 **ABSTRACT**

29 Submarine canyons are important conduits for microplastic transport to the deep sea via
30 turbidity currents. However, other near-bed oceanographic flows and sub-seafloor processes may
31 play an important role in the transport and burial of microplastics. We use sediment push-cores for
32 microplastic and sediment grain-size analysis from two transects across the Whittard Canyon, UK,
33 to show that complex process-interactions control the transport and burial of microplastics and
34 semi-synthetic microfibrils in the thalweg and on the canyon flanks. Microplastic pollution is
35 pervasive across the canyon at both transects, from the thalweg and from 500 m higher on the
36 flanks, despite turbidity currents being confined to the canyon thalweg. Furthermore, we calculate
37 sediment accumulation rates from ²¹⁰Pb dating and show that microplastic concentrations remain
38 similar at sediment depths down to 10 cm. This reveals that the huge global-increase in plastic
39 production rates over time is not recorded, and that microplastics are present in sediments that pre-

40 date the mass-production of plastic. The interaction of turbidity currents, deep-tidally-driven
41 currents, and sub-seafloor processes shreds any potential signal that microplastics may provide as
42 indicators of historical plastic production rates, which undermines the utility of microplastics as
43 reliable markers of the onset of the Anthropocene.

44

45 **1. Introduction**

46 Plastic production increased 700%, from 50 million tonnes (Mt) in the 1970's to >400 Mt in 2022
47 (PlasticsEurope, 2023). More than 10 Mt of plastic enters the world ocean annually (Lebreton *et*
48 *al.*, 2017). Microplastics (<1 mm diameter particles) represent ~13.5% of the marine plastic budget
49 (Koelmans *et al.*, 2017), including primary (manufactured particles; Zitko and Hanlon, 1991) and
50 secondary (derived from the breakdown of macroplastics; Andrady, 2011) microplastics. Semi-
51 synthetic microfibres (*e.g.*, composed of rayon and chlorinated rubber) are equally persistent in
52 the natural environment (Athey and Erdle, 2022; Finnegan *et al.*, 2022), are observed in deep-sea
53 sediments (Woodall *et al.*, 2014), and have similar detrimental effects on biota (Jiang *et al.*, 2024)
54 as plastic microfibres. Semi-synthetic microfibres are commonly used in clothes manufacturing
55 (*e.g.*, Napper and Thompson, 2016) and cigarette filters (*e.g.*, Belzagui *et al.*, 2021). Therefore,
56 we use 'microfibre' to encompass synthetic and semi-synthetic microfibres, and 'anthropogenic
57 microparticles' to encompass both microplastic particles and microfibres.

58 Lacustrine and shallow-marine settings act as archives to calculate the rate and quantity of
59 pollutant delivery (such as anthropogenic microparticles) and allow monitoring of how stresses on
60 ecosystems have changed over time (Brandon *et al.*, 2019; Uddin *et al.*, 2021 and references
61 therein). Few studies have acquired sedimentary time-series records of anthropogenic
62 microparticles in the deep sea (*e.g.*, Chen *et al.*, 2020), despite being the ultimate sink for plastics

63 (Thompson *et al.*, 2004; Woodall *et al.*, 2014; Koelmans *et al.*, 2017; Choy *et al.*, 2019).
64 Furthermore, none exist in submarine canyons, which host important seafloor ecosystems
65 (Treigner *et al.*, 2006; Fernandez-Arcaya *et al.*, 2017), and are the main conduits for delivering
66 particulate matter (Normark, 1970; Talling *et al.*, 2023), including pollutants (Paull *et al.*, 2002;
67 Zhong and Peng, 2021; Pierdomenico *et al.*, 2023), from terrestrial and coastal settings to the deep
68 sea. Avalanches of sediment, known as turbidity currents, flow through submarine canyons and
69 are responsible for generating Earth's largest sediment accumulations (Curry and Moore, 1971).
70 These flows are thought to be the main agent for anthropogenic microparticle transfer to, and
71 sequestration on, the deep seafloor (Kane and Clare, 2019; Pohl *et al.*, 2020; Rohais *et al.*, 2024;
72 Zhang *et al.*, 2024), yet other hydrodynamic processes can affect anthropogenic microparticle
73 concentrations (Kane *et al.*, 2020). It is increasingly recognised that processes other than turbidity
74 currents control particulate transport and burial in submarine canyons (*e.g.*, Maier *et al.*, 2019;
75 Bailey *et al.*, 2024; Hage *et al.*, 2024; Palanques *et al.*, 2024), and it is possible that we have
76 underestimated the importance of other hydrodynamic and sub-seafloor processes and
77 anthropogenic activities. However, the role of hydrodynamic and sub-seafloor processes, and
78 human activities on anthropogenic microparticle dispersal in submarine canyons remains
79 unconstrained. This uncertainty results from a lack of targeted seafloor sampling and
80 sedimentological context, therefore limiting our understanding of anthropogenic microparticle
81 fluxes to the deep sea, threats to deep-seafloor ecosystems, and deep-sea anthropogenic
82 sedimentary archives.

83 Our aim is to determine the interplay of anthropogenic microparticle transport and burial
84 processes in the deep-sea using a dataset from the Whittard Canyon. We assess these processes by
85 integrating detailed seafloor observations from multibeam bathymetric mapping and video footage

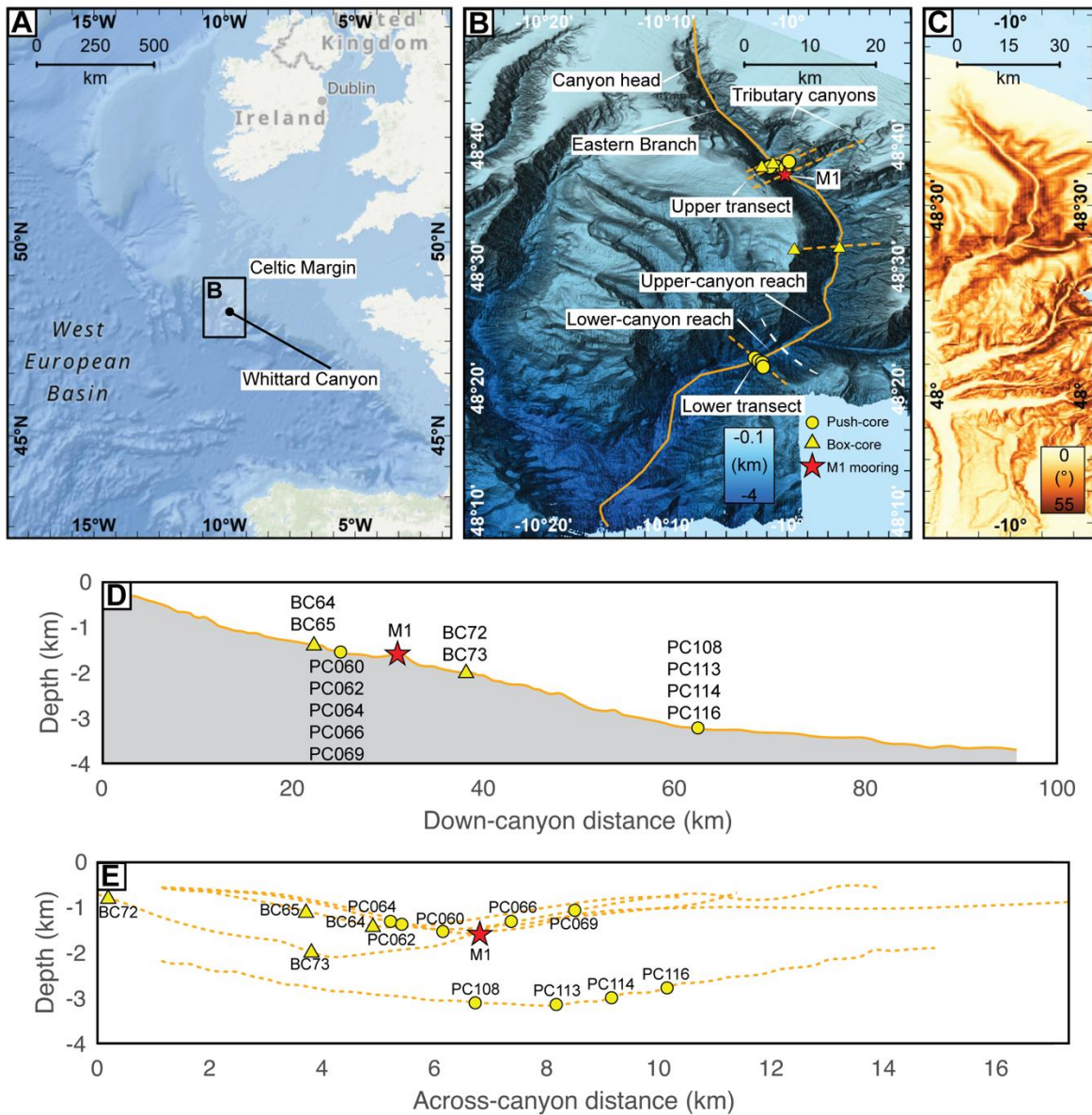
86 acquired from a Remotely Operated Vehicle (ROV), with analysis of 4 box-cores to quantify
87 sediment accumulation rates and 9 push-cores to quantify the sediment grain-size and
88 anthropogenic microparticle concentration in surficial seafloor sediments. To meet this aim, the
89 following objective are addressed: i) to map the distribution and concentration of anthropogenic
90 microparticles throughout the Whittard Canyon, ii) to document changes in anthropogenic
91 microparticle concentration with burial depth, iii) to assess sediment grain-size trends associated
92 with the anthropogenic microparticle distribution and concentration, and integrate the findings
93 with sediment accumulation rates, and iv) to discuss how anthropogenic microparticle transport
94 and burial processes controls their transfer in submarine canyons.

95

96 **2. Setting and methods**

97 **2.1. Whittard Canyon**

98 The head of Whittard Canyon lies at ~200 m water depth in the Celtic Sea, ~300 km from the
99 nearest coast (Fig. 1A). Four main branches incise steeply into the shelf break. The canon extends
100 oceanward for ~150 km, to ~3800 m water depth (Amaro *et al.*, 2016). The upper-reach of the
101 Eastern Branch extends ~55 km, from the head to ~2960 m water depth, with steep canyon flanks
102 and a $>2^\circ$ thalweg slope, with a vertical relief from flank to thalweg of ~1000 m (Fig. 1B, C and
103 E). The lower-canyon reach extends to ~3800 m water depth, with lower gradient canyon flanks
104 and a $<2^\circ$ thalweg slope, with a vertical relief from flank to thalweg of ~1250 m (Fig. 1B, C and
105 E). Further details of the canyon's geomorphology are included in the Supplemental Material.

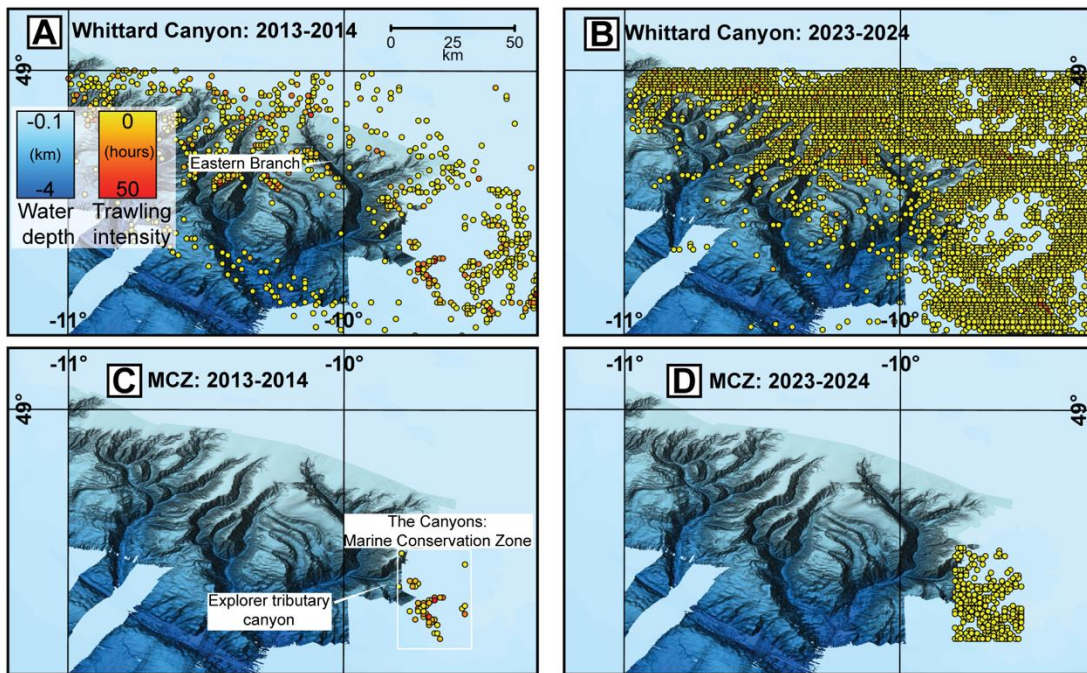


106
 107 Fig. 1. Location of data used in this study. (A) Location of Whittard Canyon. (B) Location of the
 108 cores and hydrodynamic mooring in the Eastern Branch of Whittard Canyon. (C) Slope angle map
 109 of the Eastern Branch. (D) Longitudinal profile of the canyon thalweg. (E) Cross-sections through
 110 each transect (locations on B).

111

112 **2.2. Fishing activity on the Celtic Margin**

113 Fishing activities that disturb the seafloor (*i.e.*, benthic trawling) are common around the head of
114 Whittard Canyon, and on many of its interfluves (Fig. 2). This trawling activity is a source of
115 marine pollutants (Xue *et al.*, 2020) and causes sediment resuspension (Daly *et al.*, 2018). The
116 cumulative annual benthic trawling effort for 2013-2014 and 2023-2024 was exported from
117 GlobalFishingWatch (2024) for an area of 16,650 km² (48° - 49° N to 9° - 11° W) around the
118 continental shelf, and Whittard Canyon (Fig. 2A and B). The trawling effort for the same period
119 for the 661 km² (48° 10' 2.56" - 48° 29' 59.74" N to 9° 33' 34.59" - 9° 47' 52.25" W) area covered
120 by The Canyons Marine Conservation Zone was also exported (Fig. 2C and D). The Marine
121 Conservation Zone was designated in October 2013, following the identification of vulnerable
122 ecosystems, including burrowing megafauna and cold-water corals (Duineveld *et al.*, 2001). The
123 intensity of benthic trawling on the Celtic Margin has increased fivefold in the ten-year period
124 from 2013-2014 to 2023-2024 (GlobalFishingWatch, 2024; Fig. 2).

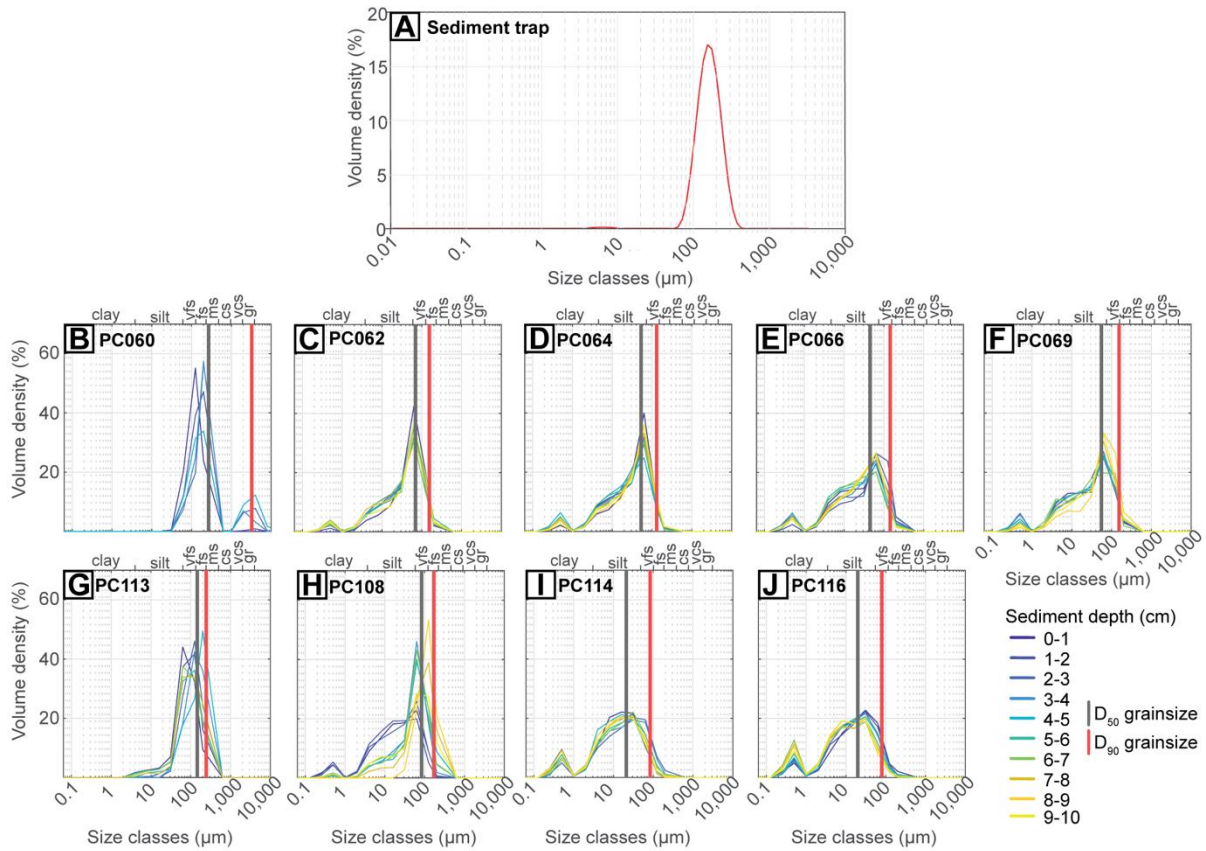


125
 126 Fig. 2. Intensity of benthic trawling as recorded by Global Fishing Watch. (A) the Whittard Canyon
 127 2013-2014. (B) the Whittard Canyon 2023-2024. (C) Marine Conservation Zone (MCZ) 2013-
 128 2014. (D) Marine Conservation Zone (2023-2024).

129 **2.3. Hydrodynamic mooring**

130 A moored, downward-looking, 600 kHz Acoustic Doppler Current Profiler (M1 mooring – Fig.
 131 1B: 30 m above seafloor; 1500 m water depth) in the Eastern Branch recorded near-bed
 132 hydrodynamic conditions from June 2019 – June 2020, including vigorous (up to 1 ms^{-1}) internal
 133 tides and 6 turbidity currents. These turbidity currents had maximum down-canyon velocities of
 134 $1.5\text{-}5.0 \text{ ms}^{-1}$, flow thicknesses $>30 \text{ m}$, and accumulated quartz-rich, fine sand in a sediment trap
 135 10 m above seafloor (Heijnen *et al.*, 2022; Fig. 3A). The frequency and velocity of the turbidity
 136 currents recorded by the ADCP during the sampling period document how the Whittard Canyon

137 experiences turbidity current activity analogous in frequency and velocity to many land-attached
 138 canyons, despite being land-detached (Heijnen *et al.*, 2022).



139
 140 Fig. 3. Grain-size distribution plots. (A) The sediment trap at the M1 mooring site of Heijnen *et*
 141 *al.* (2022). (B-J) The push-cores of the current study.

142

143 2.4. Sediment push-core recovery

144 Five precisely-located push-cores were collected using the ROV 'ISIS', along an across-canyon
 145 transect in the upper-canyon reach (24.9 km from the head, 1062-1546 m water depth) from 34
 146 metres above thalweg (m.a.t.) to 521 m.a.t. on the canyon flank. Four precisely-located push-cores

147 were also collected from an across-canyon transect in the lower-canyon reach (62.3 km from the
148 head, 2773-3204 m water depth) (Fig. 1B, D and E) from 0 m.a.t. to 431 m.a.t. on the canyon flank.
149 The push-cores were recovered from the upper-transect on the 21st August 2022, and from the
150 lower-transect on the 2nd September 2022. All 9 push-cores were subsampled at 1 cm depth-
151 intervals, down to 10 cm, depending on core recovery (subsample n=83), for anthropogenic
152 microparticle extraction, and sediment grain-size analysis (Table S1). High-resolution bathymetric
153 data enable investigation of the effects of submarine canyon geomorphology on anthropogenic
154 microparticle distribution

155

156 **2.5. Laboratory procedures**

157 ***2.5.1. Anthropogenic microparticle extraction, identification, and quantification***

158 The 1 cm sediment core horizons had variable weights and water content, so samples were dried
159 overnight in a drying oven set to 50°C. The dried samples were weighed, and for comparative
160 purposes the weight and anthropogenic microparticle content were normalised to 50 g dry sediment
161 weight. Sediment samples were then stored in glass beakers covered with aluminum foil. Samples
162 were added to a 1 L glass beaker with ~700 mL of a dense ZnCl₂ solution (1.6 g cm⁻³) and
163 disaggregated using a magnetic stirrer, and mixed until homogenized. The microplastics were
164 extracted from the sediment using a Sediment Microplastic Isolation (SMI) unit following a
165 protocol developed for microplastic extraction (Coppock *et al.*, 2017) and modified following Nel
166 *et al.* (2019). The solution was added to the SMI unit, and the beaker was rinsed with ZnCl₂
167 solution to flush any remaining sediment/anthropogenic microparticles. Prior to each use, the SMI
168 unit was disassembled and thoroughly rinsed with Class 1 Milli-Q de-ionized water. Following
169 settling overnight, the headspace supernatant was isolated by closing the ball valve of the SMI unit

170 and rinsing with extra ZnCl₂ solution to flush any remaining anthropogenic microparticles before
171 vacuum filtering over a Whatman 541, 22 μm filter paper. The prepared filter paper was then
172 placed in a labelled petri dish and covered. Throughout the extraction procedure, all individuals
173 wore white cotton laboratory coats and latex gloves. All the extraction stages were performed in a
174 clean laboratory in a fume cupboard. When the sediment samples were mixing in the 1 L beaker,
175 and settling in the SMI units they were covered with aluminum foil to limit airborne anthropogenic
176 microparticle contamination. When it was not possible during the sample preparation to cover the
177 sediment sample with aluminum foil, an opened petri dish with a blank, Whatman 541, 22 μm,
178 filter paper was placed in the fume cupboard and used as a contamination control procedural blank.
179 When the prepared filter paper was exposed during the anthropogenic microparticle identification
180 stage, a second contamination control procedural blank was also collected, again using an opened
181 petri dish with a blank, Whatman 541, 22μm, filter paper, placed in the microscopy laboratory
182 (Table S2).

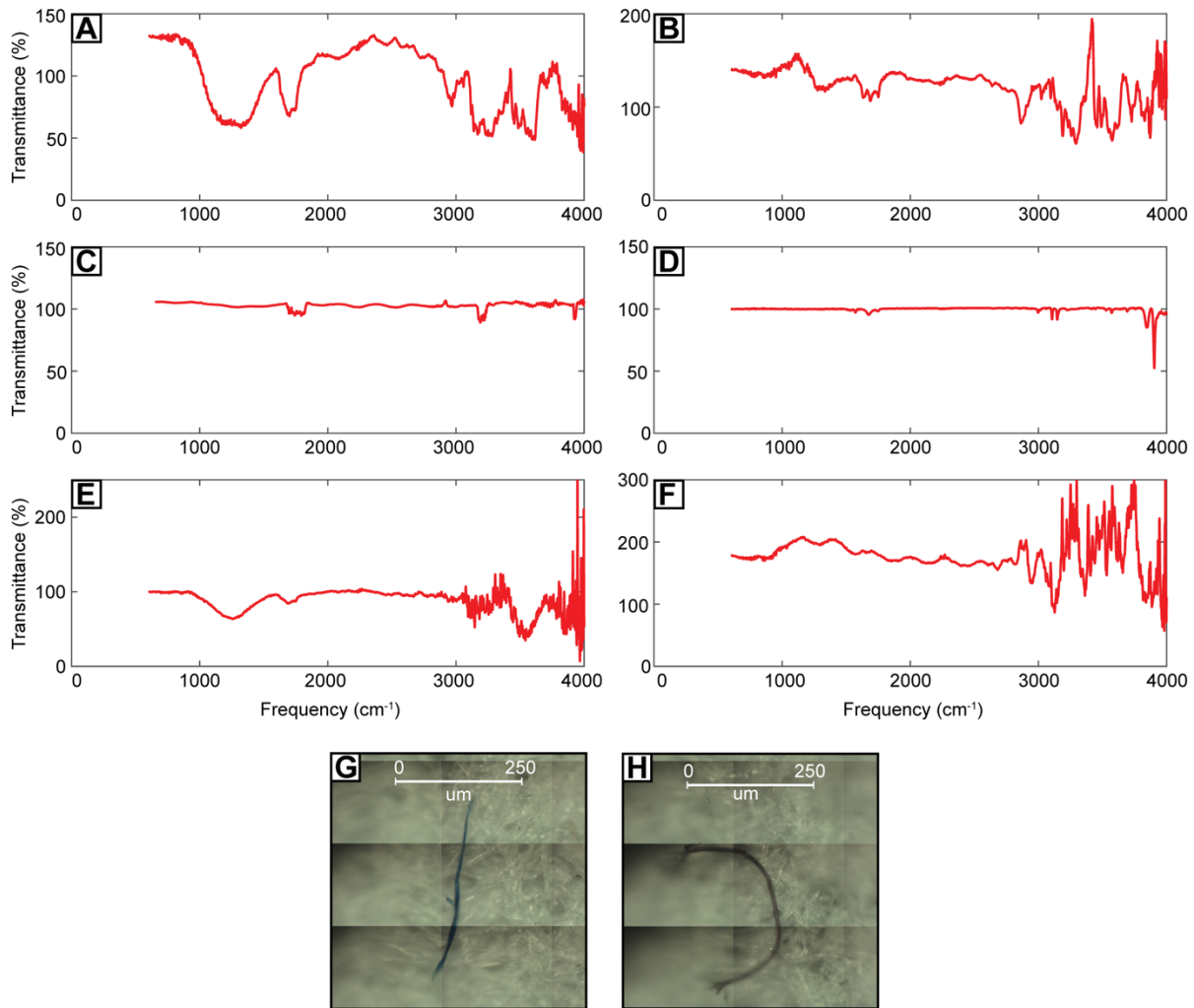
183 The prepared filter papers, both from the extraction process and the airborne contamination
184 control blanks were analysed in a clean microscopy laboratory using a Zeiss Axio Zoom, V16
185 stereomicroscope at 20-50X magnification. Here, we define anthropogenic microparticles as
186 between in 1 μm and 1 mm in size; the same size range used by prominent microplastic studies
187 (*e.g.*, Browne *et al.*, 2011; Claessens *et al.*, 2011; Van Cauwenberghe *et al.*, 2013, 2015; Vianello
188 *et al.*, 2013; Dekiff *et al.*, 2014; Kane and Clare, 2019; Kane *et al.*, 2020). Filter papers were
189 traversed systematically to identify anthropogenic microparticles based on the following criteria:
190 (i) no visible cellular or organic structures; (ii) a positive reaction to the hot needle test (de Witte
191 *et al.*, 2014); and (iii) maintenance of structural integrity when touched or moved. Anthropogenic

192 microparticles were categorised based on their color and type, including, whether they were
193 microfibrils, microplastic fragments (including films), or microbeads (Table S1).

194

195 ***2.5.2. Fourier Transform Infrared Spectroscopy***

196 Anthropogenic microparticles were visually identified using optical microscopy and a subset of
197 particles were analysed using Fourier transform infrared (FTIR) spectroscopy for polymer
198 confirmation. Identification of polymer composition was conducted on a subsample (n=13) of the
199 extracted microplastics using a PerkinElmer Spotlight 400 FTIR spectrometer using transmittance
200 mode (Fig. 4; Table S3). Further details are included in the Supplemental Material.



201
 202 Fig. 4. Fourier transform infrared (FTIR) spectroscopy spectra and microscope photographs of
 203 microfibres. (A) Rayon FTIR spectra. (B) Polyester FTIR spectra. (C) Polyethylene FTIR spectra.
 204 (D) Polystyrene FTIR spectra. (E) Chlorinated rubber FTIR spectra. (F) Polypropylene FTIR
 205 spectra. (G) Photograph of polyester microfibre. (H) Photograph of rayon microfibre.

206
 207 **2.5.3. Grain-size analysis**

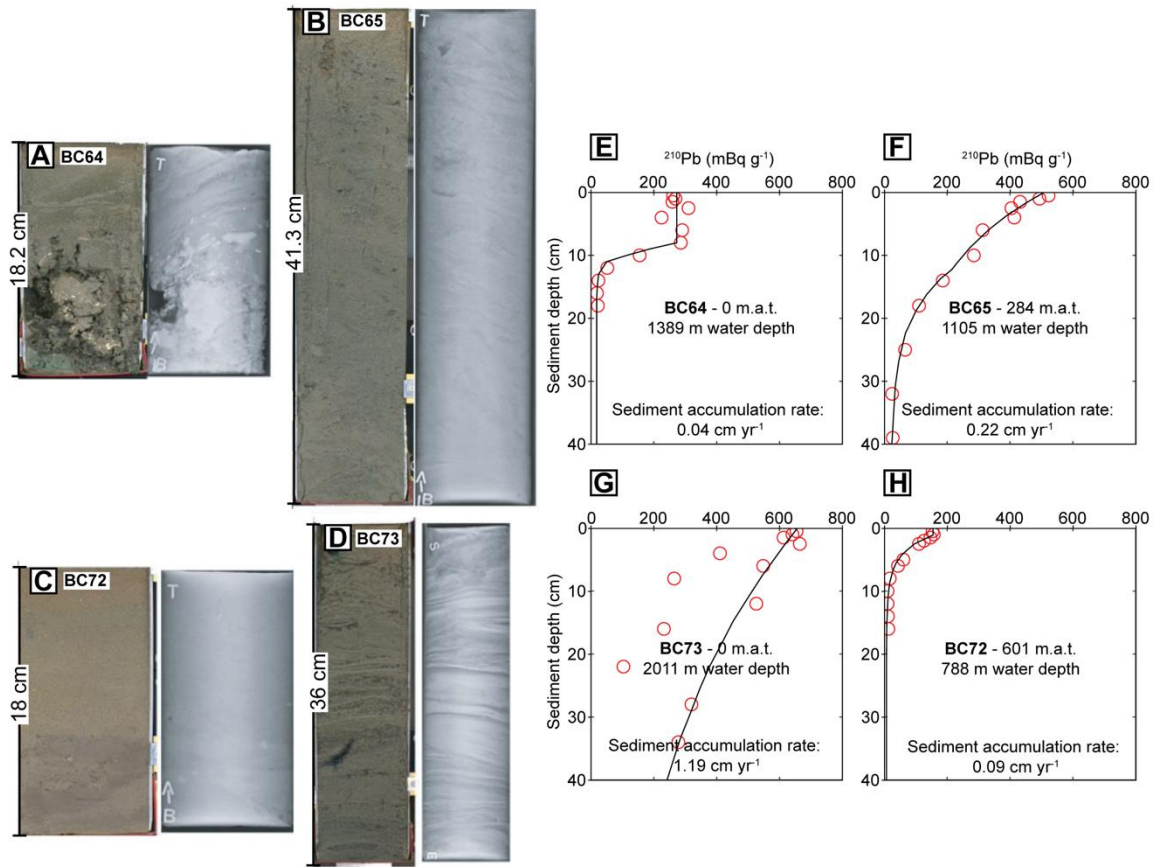
208 The grain-size of 79 of the 83 push-core samples was analysed using a Microtrac FLOWSYNC
 209 particle sizer (Microtrac MRB). The grain-size of the four remaining samples (PC060B-E) was
 210 analysed using the dry sieving method due to the FLOWSYNC particle sizer having an upper

211 particle limit of 2000 μm , and the fragmented shell material in the samples exceeded this upper
212 limit. The FLOWSYNC particle sizer uses tri-laser diffraction to measure particle size distribution
213 with a lower particle limit size of 0.01 μm . The samples were subjected to a small amount of
214 ultrasonic dispersion. Three aliquots were analysed to ensure that each sample was completely
215 dispersed. The grain-size distribution, indicating the volume percentage of grains in a certain size
216 interval, was constructed (Fig. 3B-J). The grain-size percentiles were exported from the
217 FLOWSYNC software and are documented in Table S1.

218

219 **2.5.4. ²¹⁰Pb sediment accumulation rates**

220 Sediment accumulation rates derived from ²¹⁰Pb dating of box-cores were recorded at 4 positions
221 within the upper-canyon reach; 2 in the thalweg and 2 on the canyon flanks (Fig 5; Table S4).
222 Sediment accumulation rates are calculated from the four box-cores (BC64, BC65, BC72, and
223 BC73) (Fig. 5E-H; Table S4), using ²¹⁰Pb dating. The box-cores were collected during the research
224 cruise 64PE421 conducted by NIOZ (the Royal Netherlands Institute for Sea Research) from the
225 14th May 2017 – 25th May 2017. The recovery rate of the box-cores varied by location. Further
226 details are included in the Supplemental Material.



227
 228 Fig. 5. (A-D) Core photographs and X-ray scans of the box-cores used in ^{210}Pb dating. (E-H) the
 229 sediment accumulation rate plots for the box-cores. (A and E) Box-core 64. (B and F) Box-core
 230 65. (C and G) Box-core 72. (D and H) Box-core 73. m.a.t. is meters above thalweg.

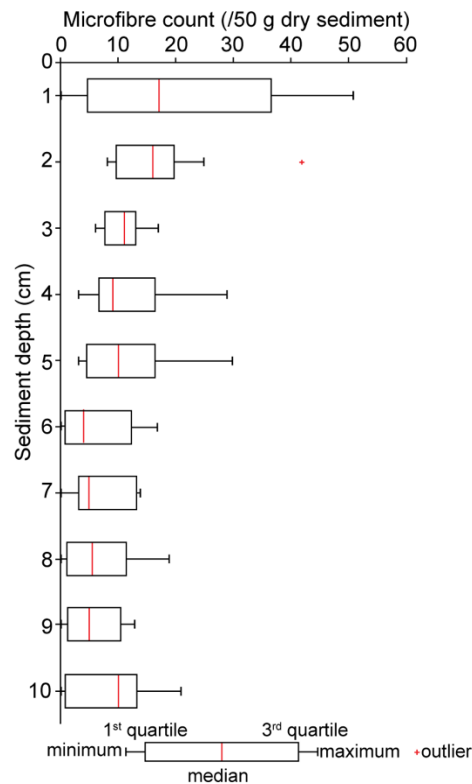
231

232 3. Results

233 3.1. Anthropogenic microparticle pollution in surficial sediments

234 Anthropogenic microparticles were present throughout all 9 push-cores (Figs. 6, 7C and 7F). A
 235 total of 1255 anthropogenic microparticles were observed with optical microscopy and a subset of
 236 the particles ($n = 13$) was verified with FTIR spectroscopy. Microfibres were the dominant
 237 anthropogenic microparticle type (microfibres = 91.3%, fragments = 5.7%, microbeads = 3.0%).

238 Herein, the anthropogenic microparticle count quantifies as the number of particles per 50 g of dry
239 sediment weight (particles/50 g). FTIR spectroscopy confirms 62% of the anthropogenic
240 microparticles are plastic, with common polymers including polyvinyl butyral, polyvinylchloride,
241 and acrylic. The remaining 38% comprise semi-synthetic polymers, including chlorinated rubber
242 and rayon (Fig. 4 ; Table S3).



243

244 Fig. 6. Box plot for microfibre concentration and sediment depth for all push-cores.

245

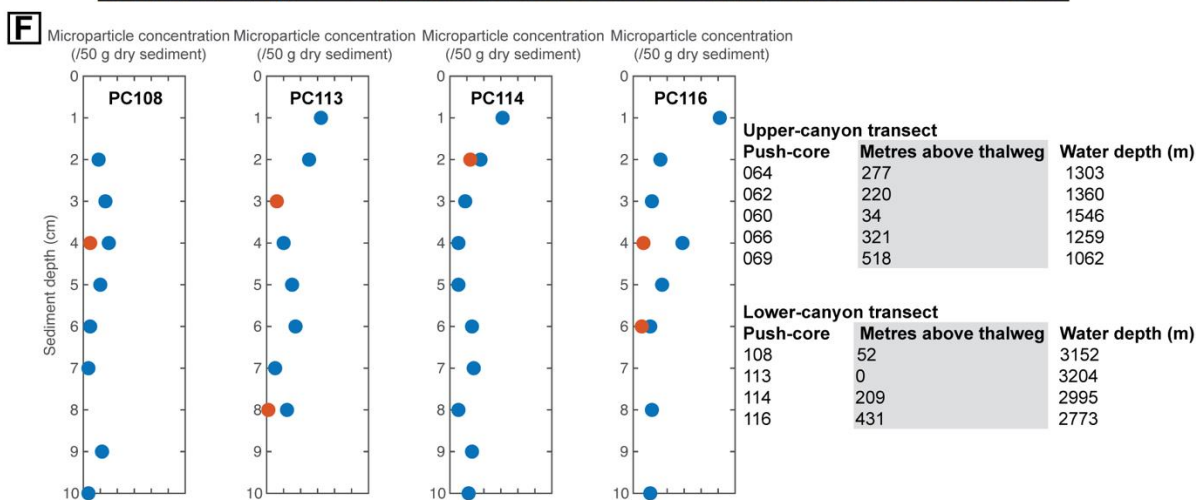
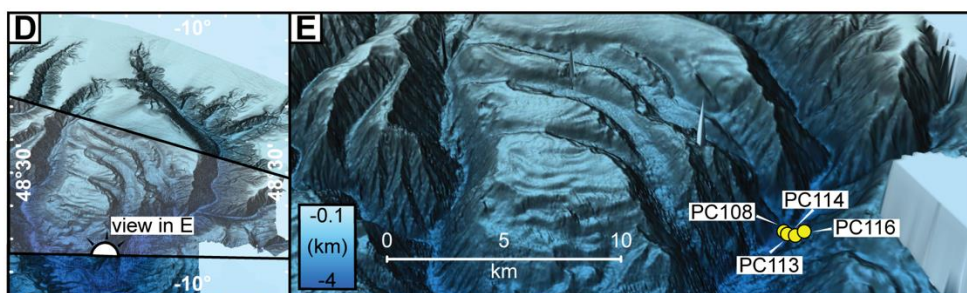
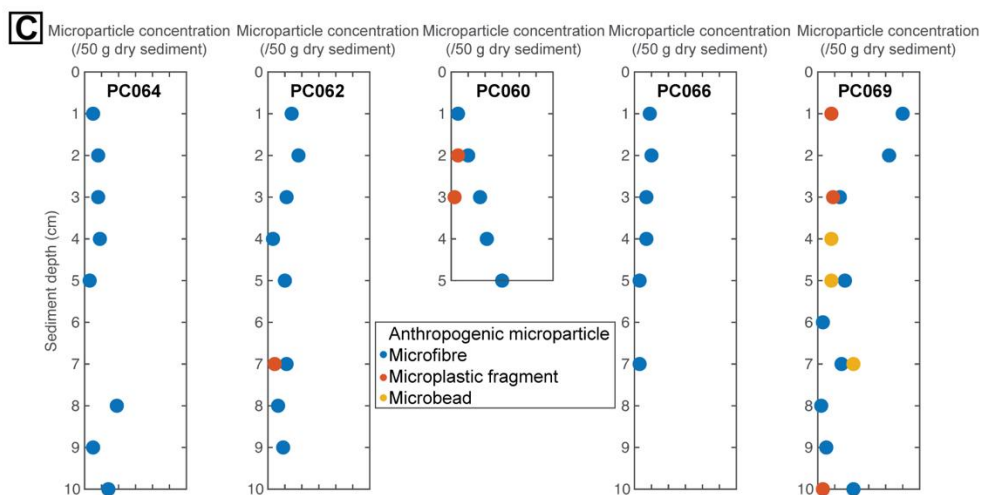
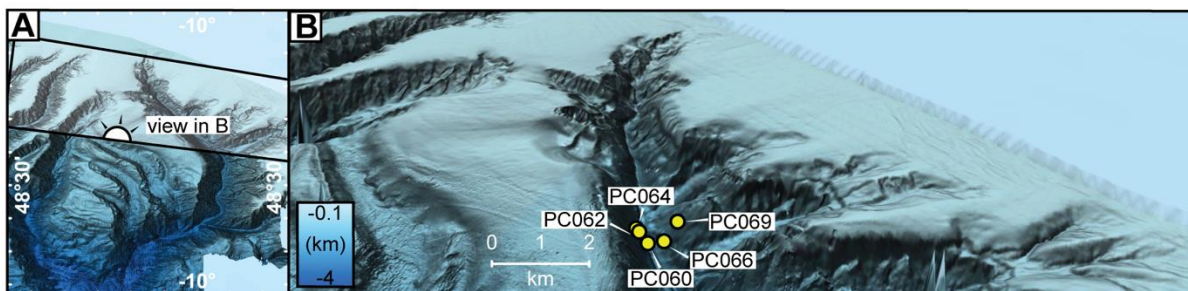
246 3.2. Microfibres in the canyon thalweg

247 In push-core 060 (PC060) (34 m.a.t., at the upper-transect), the grain-size range is 31-8000 μm ,
248 and the arithmetic mean gravel% and sand% are 9.6% and 90.3%, respectively; the granule-sized
249 particles are fragmented shells (Fig. 3B; Table S1). Microfibre count in PC060 increases with

250 sediment depth from 4 to 30 microfibrils/50 g (Fig. 7C). In PC113 (0 m.a.t. at the lower-transect),
251 the grain-size range is 2-200 μm , and the arithmetic mean sand% and silt% are 92.4% and 7.6%,
252 respectively (Fig. 3G; Table S1). Microfibre count in PC113 decreases by 62.5% with sediment
253 depth (Fig. 7F).

254 The sediment accumulation rates in BC64 (1389 m water depth, 0 m.a.t.) and BC73 (2011
255 m water depth, 0 m.a.t.) are 0.04 cm yr^{-1} and 1.19 cm yr^{-1} , respectively (Fig. 5E and G). Therefore,
256 it could take 8.4-to-250 years to accumulate 10 cm of sediment in the canyon thalweg, meaning
257 sediments containing anthropogenic microparticles in the thalweg may pre-date the mass
258 production of plastic in the 1950's. The mobility of sediment within the thalweg can be observed
259 in a photograph captured by the ROV ISIS during the recovery of PC060; a high level of suspended
260 sediment is recorded in the water column of the thalweg following the passing of a turbidity current
261 down-canyon (Fig. 8A).

262



264 Fig. 7. Anthropogenic microparticle count with sediment depth for the push-cores located in
265 Whittard Canyon. (A, B, D, and E) Location maps and high-resolution bathymetric maps of the
266 Eastern Branch. 3X vertical exaggeration. (C and F) Anthropogenic microparticle trends for each
267 push-core.

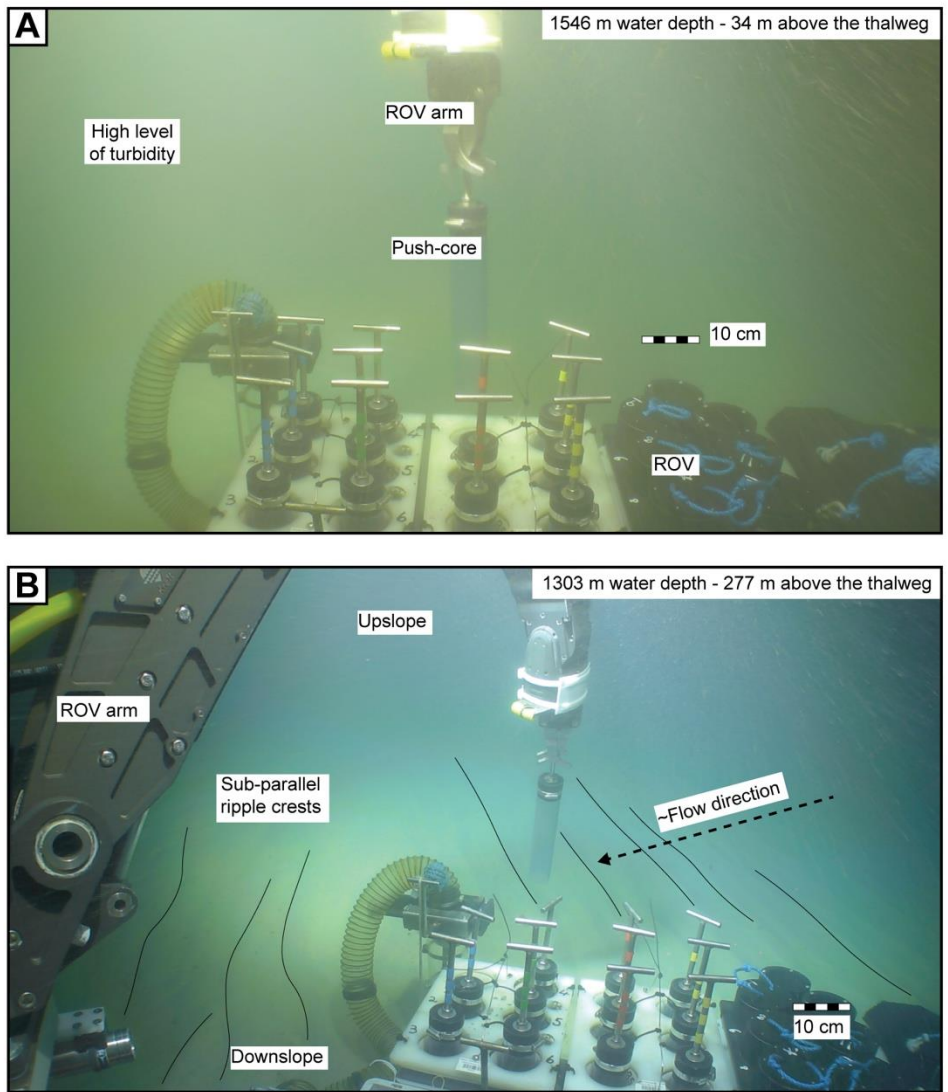
268

269 **3.3. Microfibres on the canyon flanks**

270 At the upper transect, push-cores (PC062, PC064, and PC066, located 220, 277, and 321 m.a.t.,
271 respectively) have a grain-size range of 0.25-200 μm (clay-to-fine sand) (Fig. 3C, D and E), and
272 an arithmetic mean sand% of 54.9%, 43.8%, and 39.9%, respectively (Table S1). Microfibre count
273 in these cores is low and uniform, ranging from 0-19/50 g with an arithmetic mean of 7/50 g (Fig.
274 7C). PC069 (518 m.a.t.) is located near the tributary canyons at the upper transect; the grain-size
275 range is also 0.25-200 μm , yet despite its increased height above the thalweg, the arithmetic mean
276 sand% is 47.6% (Fig. 3F; Table S1). PC069 contains the greatest range of anthropogenic
277 microparticle types, and an arithmetic mean microfibre count of 18/50 g (Fig. 7C; Table S1). At
278 the lower transect, PC114 and PC116, located 209 and 431 m.a.t., respectively, have the same
279 grain-size range as the canyon flank push-cores at the upper transect, but with an arithmetic mean
280 sand% of 17.2% and 16.5%, respectively (Fig. 3I and J; Table S1). In these push-cores, the
281 microfibre count decreases with depth by 64.5% and 80.3%, respectively (Fig. 7F and Table S1).

282 The sediment accumulation rates in BC65 (1105 m water depth, 284 m.a.t.) and BC72 (788
283 m water depth, 601 m.a.t.) are 0.22 cm yr^{-1} and 0.09 cm yr^{-1} , respectively (Fig. 5F and H).
284 Therefore, it could take 45-to-111 years to accumulate 10 cm of sediment on the canyon flanks
285 and means that sediment containing anthropogenic microparticles on the canyon flanks may pre-
286 date the mass-production of plastic.

287 On the canyon flanks at the upper transect, 277 m.a.t., and thus above the known thickness
288 of the turbidity currents recorded by Heijnen *et al.* (2022), the crest orientation of sub-parallel
289 ripples observed on the seafloor suggests a flow direction approximately perpendicular to the
290 direction of turbidity current transport (Fig. 8B). This suggests that other hydrodynamic processes
291 capable of sediment transport are also active on the canyon flanks (*e.g.*, internal tides).



292
293 Fig. 8. Photographs taken of seabed push-core sampling from the Remotely Operated Vehicle. (A)
294 Canyon thalweg at the upper-transect. (B) Canyon flanks at the upper-transect.

295

296 **4. Discussion**

297 **4.1. Microfibre transport and burial processes**

298 Microfibre pollution is pervasive throughout the Eastern Branch down to the 10 cm sediment depth
299 of the push cores. Almost all push-cores show a gradual decline in microfibre concentration with
300 depth. This gradual decline with depth is despite the marked differences in sediment accumulation
301 rates across the canyon, and the 700% increase in the background plastic production rate.
302 Microfibres are hypothesised to be transported to the canyon head via cross-continental shelf
303 currents (Fig. 9A), and via vertical settling from marine sources (Fig. 9B and F), but their
304 subsequent redistribution and burial cannot solely be explained by deposition from turbidity
305 currents.

306 From the observed grain-size trends in the canyon thalweg (notably the absence of sediment
307 $<31 \mu\text{m}$ in PC060) we infer that the frequent (sub-annual) and fast (up to 5 ms^{-1}) turbidity currents
308 serve to bypass and winnow silt-sized sediment and microfibrils further down-canyon. However,
309 microfibrils were recorded at elevations up to 518 m.a.t., over an order of magnitude above the
310 recorded thickness of measured turbidity currents. This suggests that other processes are important
311 in the Whittard Canyon, and may need be considered in other submarine canyons (Fig. 9). The
312 presence of sand in the canyon flank push-cores, and increased sand% 518 m.a.t., suggests that
313 sediment is not sourced exclusively from hemipelagic fallout. The canyon flank sands point to
314 sediment, and microfibrils and other anthropogenic microparticles, stored on the Celtic Margin
315 being transported via episodic turbidity currents in the tributary canyons or by sediment
316 resuspension by benthic trawling close to the canyon head and on the canyon interflaves (Figs. 2,
317 3F, and 9; Table S1). Internal tides break against the steep topography of the canyon flanks in the
318 upper-canyon and are focused into the canyon thalweg (Amaro *et al.*, 2016; Hall *et al.*, 2017; van

319 Haren *et al.*, 2022), thus providing another mechanism for resuspending sediment and microfibrils
320 throughout the canyon (Fig. 9D). The location of BC72 (Fig. 1D and E), high on the canyon flank
321 opposite the Celtic Margin and the tributary canyons, could help to explain the low sediment
322 accumulation rates (Fig. 5F).

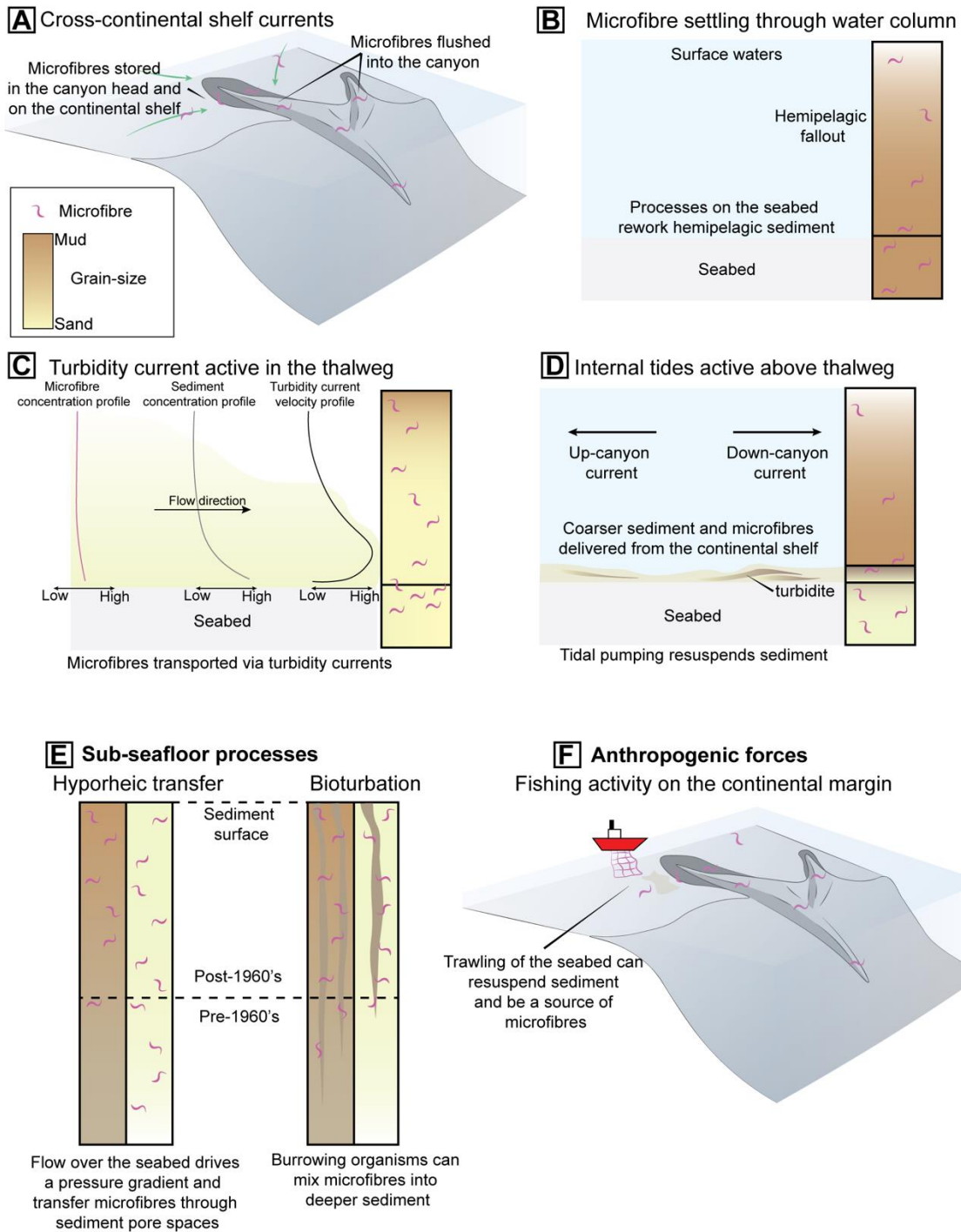
323 The observed uniformity of the gradual decline in microfibre concentration with sediment
324 depth suggests, however, that sub-seafloor processes also affect microfibre burial processes in the
325 deep sea. Hyporheic transfer of microplastics has been demonstrated in riverbeds (Frei *et al.*,
326 2019). In sub-seafloor settings, hyporheic transfer is driven by pressure gradients, as exist between
327 the base of turbidity currents and the seafloor (*e.g.*, Eggenhuisen and McCaffrey, 2012) and is
328 invoked here as a control on the stratigraphic distribution of microfibrils (Fig. 9E). Furthermore,
329 sediment resuspension via internal tidal pumping (*e.g.*, Li *et al.*, 2019; Normandeau *et al.*, 2024)
330 on the canyon flanks may generate a sufficient pressure gradient to drive hyporheic transfer of
331 microfibrils through sediment pore spaces, where turbidity currents are not active. Microplastic
332 infiltration depth increases positively with sediment grain-size (Waldschläger and Schüttrumpf,
333 2020), hence hyporheic transfer may be enhanced in the canyon thalweg compared to the canyon
334 flanks (Fig. 9E).

335 Bioturbation may also play a role in controlling the vertical distribution of microfibrils in
336 the sub-seafloor (Fig. 9E). The uppermost 10 cm of BC64 and BC65 are bioturbated (Fig. 5A and
337 B). Sediment and microplastic mixing by bioturbation has been documented experimentally
338 (Näkki *et al.*, 2017) and is hypothesised to occur in deep-sea sediments (Courteney-Jones *et al.*,
339 2020). The depth of the bioturbated layer extends to 10 cm in modern marine sediments, with
340 individual burrows extending deeper (Tarhan *et al.*, 2015) and may be enhanced on the canyon
341 flanks, due to less stressed conditions for organisms to colonize compared to the thalweg (Fig. 9E).

342 Bioturbation and hyporheic transfer are likely important in transferring anthropogenic
343 microparticles into pre-1950's deep-sea sediments; the latter supported in lakes where bioturbation
344 is absent (Dimante-Deimantovica *et al.*, 2024).

345

Transport processes



346

347 Fig. 9. Synthesis of microfiber transport and burial processes in submarine canyons. (A-D)

348 Transport processes. (E) Sub-seafloor processes. (F) Anthropogenic forces.

349

350 **4.2. Shredding of anthropogenic microparticle signals in the deep-sea**

351 We suggest that sediment transport and burial processes, and anthropogenic forcing, act as
352 nonlinear filters that shred the environmental signal of increasing plastic production rates through
353 time in submarine canyons. The efficiency of anthropogenic microparticle transfer from land-
354 based sources to the Whittard Canyon is relatively low, given the land-detached nature of the
355 canyon. This suggests that anthropogenic microparticle pollution in land-detached canyons, of
356 which there are >5000 (Harris and Whiteway, 2011), is dominantly marine-sourced, and that such
357 systems receive a buffered supply of terrestrially-sourced anthropogenic microparticles. Given the
358 dynamism of submarine canyons, the buffered supply of anthropogenic microparticles to land-
359 detached canyons, and the mobility of microfibrils and thus other anthropogenic microparticles in
360 the sub-seafloor, the efficacy of using anthropogenic microparticles as anthropogenic tracer
361 particles is questionable, along with calculations of their fluxes.

362

363 **5. Conclusions**

364 Our results show that anthropogenic microparticle pollution is pervasive in Whittard Canyon, to
365 10 cm sediment depth in both the thalweg, and on canyon flanks over 500 hundred meters above
366 the thalweg. While turbidity currents are a major agent in the transfer of anthropogenic
367 microparticles, the turbidity currents in Whittard Canyon are only 10s of meters thick (Heijnen *et*
368 *al.*, 2022), suggesting other processes and sources of anthropogenic microparticles are important.
369 Additional sources include hemipelagic settling, and sediments on the continental shelf
370 resuspended by benthic trawling and entering tributary canyons. Transport and resuspension of
371 anthropogenic microparticles by internal tidal pumping likely occurs across the entire canyon

372 water depth. Almost all the push-cores show a gradual decline in anthropogenic microparticle
373 concentrations down to 10 cm, despite the 700% increase in global plastic production since the
374 1970's. Where low sedimentation accumulation rates are recorded, much of the sediment in box-
375 cores pre-dates plastic production. This suggests mobility of anthropogenic microparticles in the
376 sub-seafloor, with likely processes including bioturbation and hyporheic transfer. The observed
377 distribution of anthropogenic microparticles in Whittard Canyon demonstrates they are not entirely
378 flushed through canyons, but may be permanently or transiently stored, and be mobile within the
379 sediment bed. These results suggest that anthropogenic microparticles incorporated in deep-sea
380 sediments may be a poor record of canyon particulate flux and that identifying the Anthropocene
381 boundary using anthropogenic microparticles in these sediments may not be straightforward.

382

383 **CRedit author contribution statement**

384 **Ed Keavney:** Conceptualisation, Methodology, Formal analysis, Investigation, Writing – Original
385 Draft, Visualisation. **Ian A. Kane:** Conceptualisation, Methodology, Resources, Writing – Review
386 & Editing, Supervision, Project administration, Funding acquisition. **Michael A. Clare:**
387 Conceptualisation, Resources, Writing – Review & Editing, Supervision, Project administration,
388 Funding acquisition. **David M. Hodgson:** Conceptualisation, Writing – Review & Editing,
389 Supervision, Project administration. **Veerle A.I. Huvenne:** Investigation, Writing – Review &
390 Editing, Project Administration, Funding acquisition. **Esther J. Sumner:** Investigation, Writing –
391 Review & Editing. **Jeff Peakall:** Conceptualisation, Writing – Review & Editing, Supervision.
392 **Furu Mienis:** Investigation, Project administration, Funding acquisition. **Jonathan Kranenburg:**
393 Methodology, Formal analysis, Visualisation.

394

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400

401 **Figure captions**

402 Fig. 1. Location of data used in this study. (A) Location of Whittard Canyon. (B) Location of the
403 cores and hydrodynamic mooring in the Eastern Branch of Whittard Canyon. (C) Slope angle map
404 of the Eastern Branch. (D) Longitudinal profile of the canyon thalweg. (E) Cross-sections through
405 each transect (locations on B).

406

407 Fig. 2. Intensity of benthic trawling as recorded by Global Fishing Watch. (A) the Whittard Canyon
408 2013-2014. (B) the Whittard Canyon 2023-2024. (C) Marine Conservation Zone (MCZ) 2013-
409 2014. (D) Marine Conservation Zone (2023-2024).

410

411 Fig. 3. Grain-size distribution plots. (A) The sediment trap at the M1 mooring site of Heijnen *et*
412 *al.* (2022). (B-J) The push-cores of the current study.

413

414 Fig. 4. Fourier transform infrared (FTIR) spectroscopy spectra and microscope photographs of
415 microfibrils. (A) Rayon FTIR spectra. (B) Polyester FTIR spectra. (C) Polyethylene FTIR spectra.
416 (D) Polystyrene FTIR spectra. (E) Chlorinated rubber FTIR spectra. (F) Polypropylene FTIR
417 spectra. (G) Photograph of polyester microfibre. (H) Photograph of rayon microfibre.

418

419 Fig. 5. (A-D) Core photographs and X-ray scans of the box-cores used in ^{210}Pb dating. (E-H) the
420 sediment accumulation rate plots for the box-cores. (A and E) Box-core 64. (B and F) Box-core
421 65. (C and G) Box-core 72. (D and H) Box-core 73. m.a.t. is meters above thalweg.

422

423 Fig. 6. Box plot for microfibre concentration and sediment depth for all push-cores.

424

425 Fig. 7. Anthropogenic microparticle count with sediment depth for the push-cores located in
426 Whittard Canyon. (A, B, D, and E) Location maps and high-resolution bathymetric maps of the
427 Eastern Branch. 3X vertical exaggeration. (C and F) Anthropogenic microparticle trends for each
428 push-core.

429

430 Fig. 8. Photographs taken of seabed push-core sampling from the Remotely Operated Vehicle. (A)
431 Canyon thalweg at the upper-transect. (B) Canyon flanks at the upper-transect.

432

433 Fig. 9. Synthesis of microfiber transport and burial processes in submarine canyons. (A-D)
434 Transport processes. (E) Sub-seafloor processes. (F) Anthropogenic forces.

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526 [30T00%3A00%3A00.000Z&end=2024-08-](https://globalfishingwatch.org/map/index?start=2024-05-30T00%3A00%3A00.000Z&end=2024-08-30T00%3A00%3A00.000Z&longitude=26&latitude=19&zoom=1.49)
527 [30T00%3A00%3A00.000Z&longitude=26&latitude=19&zoom=1.49](https://globalfishingwatch.org/map/index?start=2024-05-30T00%3A00%3A00.000Z&end=2024-08-30T00%3A00%3A00.000Z&longitude=26&latitude=19&zoom=1.49) (April 25th 2024).

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684 **Supplemental Material**

685 **Setting and Methods**

686 *Bathymetric Data*

687 The bathymetry of the Northeast Atlantic Ocean is derived from the Esri Ocean Basemap
688 ([https://www.arcgis.com/apps/mapviewer/index.html?webmap=67ab7f7c535c4687b6518e6d234](https://www.arcgis.com/apps/mapviewer/index.html?webmap=67ab7f7c535c4687b6518e6d2343e8a2)
689 [3e8a2](https://www.arcgis.com/apps/mapviewer/index.html?webmap=67ab7f7c535c4687b6518e6d2343e8a2)). The Digital Terrain Model data for the Whittard Canyon is based on the 2020 EMODnet
690 digital terrain model (DTM) (<https://doi.org/10.12770/bb6a87dd-e579-4036-abe1-e649cea9881a>),
691 which has a resolution of 1/16 x 1/16 arc minute of longitude and latitude (ca. 115 x 115 meters).
692 The bathymetry for the Eastern Branch of Whittard Canyon is derived from the GEBCO_2023
693 Grid, GEBCO Compilation Group (2023) GEBO 2023 Grid ([doi:10.5285/f98b053b-0cbc-6c23-](https://doi.org/10.5285/f98b053b-0cbc-6c23-e053-6c86abc0af7b)
694 [e053-6c86abc0af7b](https://doi.org/10.5285/f98b053b-0cbc-6c23-e053-6c86abc0af7b)). All the bathymetry data are analysed using ArcGIS Pro software to mark the
695 moorings and sample locations, and to construct the longitudinal profiles and the cross-sections of
696 the canyons.

697

698 *Fishing Activity on the Celtic Margin*

699 Fishing activity data were downloaded from Global Fishing Watch (GlobalFishingWatch, 2024)
700 and formatted in estimated annual fishing effort (in hours) per 0.01 x 0.01° grids. The fishing
701 activities that disturb the seafloor were extracted.

702

703 **Laboratory Procedures**

704 *Fourier Transform Infrared Spectroscopy*

705 The analytical region was positioned over the identified particle, the particle was imaged, and then
706 scanned over a spectrum range of 4000-650 cm⁻¹, with a resolution of 4 cm⁻¹ at a rate of 16 scans

707 per analysis. The acquired spectra produced from the analysed particles were then processed and
708 compared using the PerkinElmer Spectrum IR software with a standard reference library to assign
709 polymer type.

710

711 ***Grain-size Analysis***

712 The FLOWSYNC particle sizer uses tri-laser diffraction to measure particle size distribution with
713 a lower particle limit size of 0.01 μm . The samples were subjected to a small amount of ultrasonic
714 dispersion. Three aliquots were analysed to ensure that each sample was completely dispersed.

715

716 ***^{210}Pb Sediment Accumulation Rates***

717 The sediment profiles of ^{210}Pb are determined by alpha-spectrometry from ^{210}Po . ^{210}Pb is a
718 naturally occurring radionuclide, part of the ^{238}U decay series, with a half-life of 22.3 years. ^{210}Po
719 is extracted from the sediment by leaching with concentrated HCl or by total digestion using HNO_3
720 and HF. The ^{210}Po is collected and counted with an alpha detector and the ^{210}Pb profiles are plotted
721 on a cumulative mass scale with an exponential curve. Where the ^{210}Pb profiles deviate from the
722 exponential curve, it is prudent to apply a conventional one-dimensional, two-layer vertical eddy
723 diffusion model (following Carpenter *et al.* (1982)). The model assumes: 1) a constant rate of ^{210}Pb
724 supply (Appleby and Oldfield, 1978) and 2) a constant initial sedimentation rate (Krishnaswamy
725 *et al.*, 1971). A change in the gradient of the exponential curve may be due to sediment mixing
726 processes in the sediment mixed layer, as is observed in BC64 and BC65, however this is
727 accounted for in the model (Carpenter *et al.*, 1982). The sandier intervals of the box-cores hold a
728 lower ^{210}Pb signature, so they were either avoided in the sub-sampling of the core horizons or
729 sieved below 64 μm .

Table S1. Sample information, gravel%, sand%, silt%, and clay%, and scaled-up anthropogenic microparticle counts/50 g dry sediment. The associated coordinate projection for the location of the push-cores in WGS8.

Sample number	Location		Water depth (m)	Height above thalweg (m)	Core depth horizon (cm)	Gravel (%)	Sand (%)	Silt (%)	Clay (%)	Microfibre count (/ 50 g dry sediment)	Microplastic fragment count (/ 50 g dry sediment)	Microbead count (/ 50 g dry sediment)
	Lat (°N)	Long (°W)										
PSH 060												
PSH_060_A	48.61365	-10.01576667	1546	34	0-1	0	99.62	0.38	0	4	0	0
PSH_060_B	48.61365	-10.01576667	1546	34	1-2	1.1	98.90	0	0	10	8	0
PSH_060_C	48.61365	-10.01576667	1546	34	2-3	14.07	85.93	0	0	17	3	0
PSH_060_D	48.61365	-10.01576667	1546	34	3-4	10.16	89.84	0	0	21	0	0
PSH_060_E	48.61365	-10.01576667	1546	34	4-5	22.90	77.09	0	0	30	0	0
PSH 062												
PSH_062_A	48.63688333	-10.02356667	1360	220	0-1	0	64.34	35.66	0	14	0	0
PSH_062_B	48.63688333	-10.02356667	1360	220	1-2	0	56.27	42.21	1.53	18	0	0
PSH_062_C	48.63688333	-10.02356667	1360	220	2-3	0	52.14	44.91	2.95	11	0	0
PSH_062_D	48.63688333	-10.02356667	1360	220	3-4	0	53.48	43.45	3.08	3	0	0
PSH_062_E	48.63688333	-10.02356667	1360	220	4-5	0	56.03	40.29	3.68	10	0	0
PSH_062_F	48.63688333	-10.02356667	1360	220	5-6	0	54.46	42.13	3.43	0	0	0
PSH_062_G	48.63688333	-10.02356667	1360	220	6-7	0	45.98	48.79	5.23	11	4	0
PSH_062_H	48.63688333	-10.02356667	1360	220	7-8	0	N/A	N/A	N/A	6	0	0
PSH_062_I	48.63688333	-10.02356667	1360	220	8-9	0	53.06	43.62	3.32	9	0	0
PSH_062_J	48.63688333	-10.02356667	1360	220	9-10	0	64.34	37.09	4.17	0	0	0
PSH 064												
PSH_064_A	48.63600667	-10.0248	1303	277	0-1	0	52.65	44.92	2.43	5	0	0
PSH_064_B	48.63600667	-10.0248	1303	277	1-2	0	43.78	50.71	5.51	8	0	0
PSH_064_C	48.63600667	-10.0248	1303	277	2-3	0	40.28	53.72	6	8	0	0
PSH_064_D	48.63600667	-10.0248	1303	277	3-4	0	40.65	54.73	4.62	9	0	0
PSH_064_E	48.63600667	-10.0248	1303	277	4-5	0	34.08	57.59	8.33	3	0	0
PSH_064_H	48.63600667	-10.0248	1303	277	7-8	0	41.3	52.79	5.91	19	0	0
PSH_064_I	48.63600667	-10.0248	1303	277	8-9	0	41.78	52.56	5.66	5	0	0
PSH_064_J	48.63600667	-10.0248	1303	277	9-10	0	45.21	49.46	5.33	14	0	0
PSH 066												
PSH_066_A	48.63545167	-10.00317333	1259	321	0-1	0	38.7	53.82	7.48	9	0	0
PSH_066_B	48.63545167	-10.00317333	1259	321	1-2	0	55.43	41.67	2.9	10	0	0
PSH_066_C	48.63545167	-10.00317333	1259	321	2-3	0	48.02	45.68	6.3	7	0	0
PSH_066_D	48.63545167	-10.00317333	1259	321	3-4	0	40.51	50.72	8.77	7	0	0
PSH_066_E	48.63545167	-10.00317333	1259	321	4-5	0	35.87	59.06	5.07	3	0	0
PSH_066_F	48.63545167	-10.00317333	1259	321	5-6	0	31.22	61.79	6.99	0	0	0
PSH_066_G	48.63545167	-10.00317333	1259	321	6-7	0	30.18	62.82	7	3	0	0
PSH_066_H	48.63545167	-10.00317333	1259	321	7-8	0	40.33	54.23	5.44	0	0	0
PSH_066_I	48.63545167	-10.00317333	1259	321	8-9	0	38.4	55.57	6.03	0	0	0
PSH_066_J	48.63545167	-10.00317333	1259	321	9-10	0	40.72	53.65	5.63	0	0	0
PSH 069												
PSH_069_A	48.64316667	-9.99879	1062	518	0-1	0	43.45	49.5	7.05	50	8	0
PSH_069_B	48.64316667	-9.99879	1062	518	1-2	0	41.99	51.74	6.27	43	0	0
PSH_069_C	48.64316667	-9.99879	1062	518	2-3	0	42.68	49.87	7.45	14	9	0
PSH_069_D	48.64316667	-9.99879	1062	518	3-4	0	42.25	50.59	7.16	8	0	8
PSH_069_E	48.64316667	-9.99879	1062	518	4-5	0	46.82	47.26	5.92	16	0	8
PSH_069_F	48.64316667	-9.99879	1062	518	5-6	0	44.59	51.57	3.84	3	0	0
PSH_069_G	48.64316667	-9.99879	1062	518	6-7	0	41.23	55.05	3.72	14	0	22
PSH_069_H	48.64316667	-9.99879	1062	518	7-8	0	51.5	45.41	3.09	6	0	0
PSH_069_I	48.64316667	-9.99879	1062	518	8-9	0	66.49	31.89	1.62	5	0	0
PSH_069_J	48.64316667	-9.99879	1062	518	9-10	0	55.31	42.55	2.14	22	3	0
PSH 108												
PSH_108_A	48.375593	-10.0446445	3152	52	0-1	0	32.59	61.27	6.14	0	0	0
PSH_108_B	48.375593	-10.0446445	3152	52	1-2	0	25.37	67.79	6.84	9	0	0
PSH_108_C	48.375593	-10.0446445	3152	52	2-3	0	35.47	59.21	5.32	13	0	0
PSH_108_D	48.375593	-10.0446445	3152	52	3-4	0	66.15	31.8	2.05	15	4	0
PSH_108_E	48.375593	-10.0446445	3152	52	4-5	0	71.59	27.57	0.84	10	0	0
PSH_108_F	48.375593	-10.0446445	3152	52	5-6	0	65.16	33.22	1.62	4	0	0
PSH_108_G	48.375593	-10.0446445	3152	52	6-7	0	71.01	27.91	1.08	3	0	0
PSH_108_H	48.375593	-10.0446445	3152	52	7-8	0	82.14	17.86	0	0	0	0
PSH_108_I	48.375593	-10.0446445	3152	52	8-9	0	97.33	2.67	0	11	0	0
PSH_108_J	48.375593	-10.0446445	3152	52	9-10	0	67.89	30.26	1.85	3	0	0
PSH 113												
PSH_113_A	48.371195	-10.03929667	3204	0	0-1	0	85.73	14.27	0	32	0	0
PSH_113_B	48.371195	-10.03929667	3204	0	1-2	0	95.27	4.73	0	25	0	0
PSH_113_C	48.371195	-10.03929667	3204	0	2-3	0	97.85	2.15	0	6	6	0
PSH_113_D	48.371195	-10.03929667	3204	0	3-4	0	96.8	3.2	0	10	0	0
PSH_113_E	48.371195	-10.03929667	3204	0	4-5	0	93.65	6.35	0	15	0	0
PSH_113_F	48.371195	-10.03929667	3204	0	5-6	0	88.07	11.93	0	17	0	0
PSH_113_G	48.371195	-10.03929667	3204	0	6-7	0	88.85	11.15	0	5	0	0
PSH_113_H	48.371195	-10.03929667	3204	0	7-8	0	92.87	7.13	0	12	2	0
PSH 114												
PSH_114_A	48.36871667	-10.03463333	2995	209	0-1	0	12.3	78.35	9.35	31	0	0
PSH_114_B	48.36871667	-10.03463333	2995	209	1-2	0	12.43	74.21	13.36	18	12	0
PSH_114_C	48.36871667	-10.03463333	2995	209	2-3	0	28.77	64.56	6.67	9	0	0
PSH_114_D	48.36871667	-10.03463333	2995	209	3-4	0	25.7	65.67	8.63	5	0	0
PSH_114_E	48.36871667	-10.03463333	2995	209	4-5	0	12.95	79.55	7.5	5	0	0
PSH_114_F	48.36871667	-10.03463333	2995	209	5-6	0	19.08	70.48	10.44	13	0	0
PSH_114_G	48.36871667	-10.03463333	2995	209	6-7	0	19.09	71.75	9.16	14	0	0
PSH_114_H	48.36871667	-10.03463333	2995	209	7-8	0	14	73.86	12.14	5	0	0
PSH_114_I	48.36871667	-10.03463333	2995	209	8-9	0	13.85	77.71	8.44	13	0	0
PSH_114_J	48.36871667	-10.03463333	2995	209	9-10	0	14.16	73.39	12.45	11	0	0
PSH 116												
PSH_116_A	48.36326	-10.03335333	2773	431	0-1	0	25.05	67.94	7.01	51	0	0
PSH_116_B	48.36326	-10.03335333	2773	431	1-2	0	22.27	69.15	8.58	16	0	0
PSH_116_C	48.36326	-10.03335333	2773	431	2-3	0	18.28	73.86	7.86	11	0	0
PSH_116_D	48.36326	-10.03335333	2773	431	3-4	0	19.04	73.91	7.05	29	6	0
PSH_116_E	48.36326	-10.03335333	2773	431	4-5	0	18.66	72.68	8.66	17	0	0
PSH_116_F	48.36326	-10.03335333	2773	431	5-6	0	14.94	74.67	10.39	10	5	0

PSH_116_G	48.36326	-10.03335333	2773	431	6-7	0	12.04	69.94	18.02	0	0	0
PSH_116_H	48.36326	-10.03335333	2773	431	7-8	0	12.29	71.76	15.95	11	0	0
PSH_116_I	48.36326	-10.03335333	2773	431	8-9	0	10.2	75.46	14.34	0	0	0
PSH_116_J	48.36326	-10.03335333	2773	431	9-10	0	11.76	75.44	12.8	10	0	0

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Table S2. Contamination control procedural blank data for the “sample preparation” stage and the “microparticle identification” stage. n.d. corresponds to no data.

Sample number	Sample preparation exposure time (s)	Microfibre count	Microparticle identification exposure time (s)	Microfibre count
<u>PSH_060</u>				
PSH_060_A	429	3	1150	0
PSH_060_B	540	0	920	0
PSH_060_C	515	0	879	0
PSH_060_D	457	0	681	0
PSH_060_E	500	0	534	0
<u>PSH_062</u>				
PSH_062_A	492	1	689	0
PSH_062_B	525	0	748	0
PSH_062_C	483	0	503	0
PSH_062_D	471	0	597	0
PSH_062_E	531	0	834	0
PSH_062_F	436	0	412	0
PSH_062_G	388	0	456	0
PSH_062_H	540	0	478	0
PSH_062_I	433	0	502	0
PSH_062_J	420	0	330	0
<u>PSH_064</u>				
PSH_064_A	408	0	1347	1
PSH_064_B	403	0	810	0
PSH_064_C	414	1	712	0
PSH_064_D	451	0	629	0
PSH_064_E	439	0	1422	0
PSH_064_H	381	0	878	0
PSH_064_I	430	0	720	0
PSH_064_J	490	0	934	0
<u>PSH_066</u>				
PSH_066_A	470	0	701	0
PSH_066_B	467	0	645	0
PSH_066_C	483	0	522	1
PSH_066_D	1962	0	326	0
PSH_066_E	580	0	400	0
PSH_066_F	449	0	354	0
PSH_066_G	552	0	570	0
PSH_066_H	442	0	385	1
PSH_066_I	458	0	380	0
PSH_066_J	415	0	321	0
<u>PSH_069</u>				
PSH_069_A	n.d.	0	n.d.	0
PSH_069_B	n.d.	2	n.d.	0
PSH_069_C	n.d.	0	n.d.	0
PSH_069_D	n.d.	1	n.d.	0
PSH_069_E	n.d.	0	n.d.	0
PSH_069_F	n.d.	2	n.d.	1
PSH_069_G	n.d.	2	n.d.	0
PSH_069_H	n.d.	3	n.d.	0
PSH_069_I	n.d.	3	n.d.	1
PSH_069_J	n.d.	6	n.d.	0
<u>PSH_108</u>				
PSH_108_A	421	0	410	0
PSH_108_B	410	0	329	0
PSH_108_C	407	0	717	0
PSH_108_D	467	0	682	0
PSH_108_E	403	0	724	0
PSH_108_F	367	0	520	0
PSH_108_G	366	0	489	0
PSH_108_H	406	0	226	0
PSH_108_I	354	0	460	0
PSH_108_J	362	0	357	0
<u>PSH_113</u>				
PSH_113_A	410	0	837	0
PSH_113_B	435	0	869	0
PSH_113_C	391	0	708	0
PSH_113_D	399	0	725	0
PSH_113_E	530	0	713	1
PSH_113_F	406	0	442	1
PSH_113_G	448	4	772	0
PSH_113_H	429	0	729	0
<u>PSH_114</u>				
PSH_114_A	954	0	1208	0
PSH_114_B	1065	1	1408	1
PSH_114_C	716	0	528	0
PSH_114_D	883	0	482	0
PSH_114_E	839	0	467	0
PSH_114_F	600	0	816	0
PSH_114_G	865	0	962	0
PSH_114_H	595	0	548	0
PSH_114_I	543	0	651	0
PSH_114_J	565	0	605	0
<u>PSH_116</u>				
PSH_116_A	412	0	456	0
PSH_116_B	424	1	566	0
PSH_116_C	465	0	493	0

PSH_116_D	515	1	635	0	733
PSH_116_E	451	0	607	0	
PSH_116_F	516	0	469	0	734
PSH_116_G	520	0	556	0	
PSH_116_H	443	0	433	0	
PSH_116_I	434	0	396	0	
PSH_116_J	493	0	370	0	

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Table S3. Sample number and corresponding microparticle types color obtained from optical microscopy, and composition obtained from FTIR analysis.

Sample number	Microparticle type	Microparticle color	Microparticle composition
PSH_060_A	Fibre	Black	Rayon
PSH_060_E	Fibre	Black	Chlorinated Rubber
PSH_062_A	Fibre	Black	Chlorinated Rubber
PSH_062_I	Fibre	Black	Chlorinated Rubber
PSH_064_B	Fibre	Blue	Polyester
PSH_064_C	Fibre	Black	Plastic additive
PSH_069_B	Fibre	Black	Polyvinyl chloride
PSH_108_B	Fibre	Black	Synthetic resin
PSH_113_B	Fibre	Black	Polypropylene
PSH_114_I	Fibre	Black	Acrylic
PSH_114_I	Fibre	Black	Acrylic
PSH_114_J	Fibre	Black	Chlorinated Rubber
PSH_116_B	Fibre	Black	Polyvinyl chloride

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Table S4. ^{210}Pb values used to calculate sediment accumulation rates for the four box-cores.

Box-core number	Core depth horizon (cm)	^{210}Pb total (mBq g ⁻¹)	^{210}Pb total 1s error (mBq g ⁻¹)	
<u>BC64</u>	0-0.5	261.27	12.50	
	0.5-1	268.59	11.71	
	1-1.5	259.18	11.62	
	2-2.5	310.24	12.65	
	3-4	224.54	9.82	
	5-6	290.17	12.28	
	7-8	285.75	12.30	
	9-10	154.65	7.77	
	11-12	51.92	2.21	
	13-14	23.56	1.25	
	15-16	19.02	1.10	
	17-18	21.86	1.16	
	<u>BC65</u>	0-0.5	522.64	11.90
		0.5-1	493.74	11.98
1-1.5		431.94	10.07	
2-2.5		404.47	9.58	
3-4		413.64	9.02	
5-6		312.98	8.02	
9-10		284.93	7.19	
13-14		186.10	5.17	
17-18		110.57	3.21	
24-25		66.14	2.33	
31-32		28.84	1.20	
38-39		26.82	1.22	
<u>BC72</u>		0-0.5	153.47	4.03
		0.5-1	157.50	4.14
	1-1.5	146.69	3.81	
	1.5-2	126.40	3.57	
	2-2.5	110.25	3.05	
	3-5	60.72	1.89	
	5-6	43.46	1.50	
	7-8	17.39	0.81	
	9-10	10.38	0.85	
	11-12	10.40	0.83	
	13-14	11.56	0.91	
	15-16	12.93	0.99	
	<u>BC73</u>	0-0.5	654.01	26.95
		0.5-1	640.97	24.23
1-1.5		612.71	25.47	
2-2.5		664.32	22.85	
3-4		410.35	16.89	
5-6		547.83	22.24	
7-8		264.64	11.14	
11-12		525.91	20.83	
15-16		231.96	6.95	
21-22		103.19	3.79	
27-28		319.65	9.35	
33-34		278.11	8.56	

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783 [30T00%3A00%3A00.000Z&end=2024-08-](https://globalfishingwatch.org/map/index?start=2024-05-30T00%3A00%3A00.000Z&end=2024-08-30T00%3A00%3A00.000Z&longitude=26&latitude=19&zoom=1.49)

784 [30T00%3A00%3A00.000Z&longitude=26&latitude=19&zoom=1.49](https://globalfishingwatch.org/map/index?start=2024-05-30T00%3A00%3A00.000Z&end=2024-08-30T00%3A00%3A00.000Z&longitude=26&latitude=19&zoom=1.49) (April 25th 2024).

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