

1 Title: Towards a global deep learning model for daily soil CO₂ efflux
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Towards a global deep learning model for daily soil CO₂ efflux

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Key Points:

1. We trained a deep learning model to predict daily soil CO₂ efflux using fine temporal resolution data from 82 diverse sites across the globe;
2. The model performed best in temperate mesic ecosystems with cyclical data patterns driven by seasonal temperature variations;
3. The model performed poorly at sites with little data and noncyclical temporal patterns, and struggled to capture pulses and seasonal peaks/troughs.

Abstract. Soil CO₂ efflux, the largest flux of CO₂ to the atmosphere, is expected to rise globally under climate change. Its magnitude and temporal variability are highly uncertain, and daily-scale models capturing rapid changes to environmental drivers remain rare. We used a global database of soil CO₂ efflux (total observations = 7,797,535 from 2002-2020) to train a deep learning model (Long Short-Term Memory, LSTM) to predict daily soil CO₂ efflux in 82 sites across gradients of climate, soil type, and land cover. The model achieved a median train and test Nash Sutcliffe Efficiency (NSE) of 0.54 and 0.02, respectively, and Kling Gupta Efficiency (KGE) of 0.67 and 0.30, respectively. The model performed well (NSE > 0.5 and KGE > 0.3) at about one-third of sites, mainly in temperate mesic ecosystems (where most training sites were located) with cyclical data patterns driven by temperature. The model performed poorly at sites with little data and noncyclical temporal patterns, mostly at extreme climates including arid, Arctic/boreal, and tropical ecosystems. The model struggled to capture soil CO₂ efflux pulses and peaks/troughs, highlighting the challenges of modeling extremes in time series. Our results demonstrate that LSTM models can leverage existing data to generate synthetic daily datasets, particularly for temperate mesic regions, but also underscore the challenges of learning relationships from a spatially biased dataset. To improve model performance, future data collection should prioritize 1) historically underrepresented ecosystems with variable temperature relationships; 2) conditions under extreme weather events that may become disproportionately impactful in a warming climate.

1. INTRODUCTION

Soil CO₂ efflux, often referred to as soil respiration, is the flux of CO₂ between soil and the atmosphere driven by subsurface heterotrophic and autotrophic respiration and chemical weathering reactions (Rey, 2015). Annual soil CO₂ efflux reaches 75–100 Pg C yr⁻¹, making it the largest flux of CO₂ to the atmosphere and the second largest terrestrial carbon flux behind

83 photosynthesis (Hashimoto et al., 2023). Soil CO₂ efflux can determine net terrestrial-atmospheric
84 carbon balances from ecosystem (Desai et al., 2022) to global scales (Ballantyne et al., 2017; Metz
85 et al., 2023) and is rising under global warming (Arora et al., 2013; Hashimoto et al., 2023; Lei et
86 al., 2021). However, the magnitude and spatial-temporal variation of this rise has remained highly
87 uncertain (Bond-Lamberty, 2018; Bond-Lamberty et al., 2016; Nissan et al., 2023; Varney et al.,
88 2022). It remains unclear to what degree we will experience net soil carbon losses under climate
89 change (van Gestel et al., 2018).

90 Many global models use bottom-up approaches, in which soil CO₂ efflux is predicted from
91 empirical (e.g., Schlesinger, 1977), statistical (e.g., Bond-Lamberty & Thomson, 2010), or
92 machine learning-inferred (e.g., Warner et al., 2019) relationships based on drivers such as
93 temperature and precipitation (Raich et al., 2002). Global soil CO₂ efflux is usually estimated using
94 annual scale data (Bond-Lamberty et al., 2024; Hashimoto et al., 2023), which has greatly
95 advanced our understanding of its spatial-temporal patterns and sensitivities to environmental
96 drivers. However, annual-scale data suffers from challenges including the temporal bias of
97 measurements towards warmer seasons (Burton et al., 2004; Chang et al., 2009; Duiker & Lal,
98 2000; Janssens et al., 1998), nonlinearity in relationships with drivers (Jian et al., 2018), and spatial
99 (Luo & Zhou, 2006; Song et al., 2013; Wang et al., 2021; Zhang et al., 2023; Zhao et al., 2017)
100 and/or temporal non-stationarity (Bond-Lamberty et al., 2019; Vargas et al., 2010). Annual soil
101 CO₂ efflux data are also known to be highly spatially biased (Bond-Lamberty et al., 2024; Stell et
102 al., 2021; Xu & Shang, 2016), creating a further source of error. Perhaps unsurprisingly, estimates
103 of global annual soil CO₂ efflux between different global models can differ dramatically, making
104 this flux one of the most uncertain in global carbon cycle models (Hashimoto et al., 2023; Nissan
105 et al., 2023; Varney et al., 2022).

106 Developing models at finer temporal scales has the potential to help reduce uncertainties in the
107 global carbon cycle (Friedlingstein et al., 2024) by lowering the bias of coarser temporal
108 measurements and by yielding more accurate understandings of the relationships between soil CO₂
109 efflux and drivers such as temperature, soil moisture, and vegetation at sub-annual scales (Jian et
110 al., 2018). Jian et al. (2018) found notable differences in the magnitude of global soil CO₂ efflux
111 between models constrained with annual, monthly, and daily data, with the use of monthly data
112 reducing annual predictions by 7.43–9.46 Pg C, and daily-scale data by a further 1.82 Pg C.
113 However, statistical models at such sub-annual scales remain uncommon (Hashimoto et al., 2015;
114 Jian et al., 2018, 2022; Raich et al., 2002; Raich & Potter, 1995), and daily-scale global statistical
115 models are rare (Adachi et al., 2017; Jian et al., 2018).

116 Acquisition of soil CO₂ efflux data has substantially advanced in recent decades with the
117 technological advance of automated chambers enabling high-frequency, continuous measurements
118 (Bond-Lamberty et al., 2024; Goulden & Crill, 1997; Irvine & Law, 2002; Rayment & Jarvis,
119 1997; Vargas et al., 2011). These datasets have enabled major advances in understanding the
120 drivers of soil CO₂ efflux on seasonal, daily, and sub-daily scales (e.g., Carbone & Vargas, 2008;
121 Gaumont-Guay et al., 2014; Misson et al., 2010; Thomey et al., 2011; Vargas et al., 2011; Vargas
122 & Allen, 2008). These data have been compiled into a global database called COSORE with sites

123 covering gradients of climate, land cover, and soil types (Bond-Lamberty et al., 2020), enabling
124 large-scale syntheses (e.g., Anjileli et al., 2021).

125 In addition to advances in data collection, deep learning has emerged as a powerful tool for
126 inferring relationships between input and output variables by learning simultaneously from data in
127 many sites under diverse conditions (Fang et al., 2022; Kratzert et al., 2019; Zhi et al., 2023).
128 Although deep learning models alone cannot be used to gain an understanding of underlying
129 processes (Perry et al., 2022), they have become increasingly applied in earth sciences for large-
130 scale predictions of, for example, soil moisture (Fang et al., 2017), stream flow (Feng et al., 2020),
131 stream dissolved oxygen (Zhi et al., 2023), rainfall-runoff relationships (Kratzert et al., 2018), and
132 floods (Nearing et al., 2024). The Long Short-Term Memory (LSTM) model, a recurrent neural
133 network model that can retain temporal information, has emerged as a powerful deep learning
134 model for time series prediction (Hochreiter & Schmidhuber, 1997). It has been used to predict
135 half-hourly-scale soil CO₂ efflux in dryland ecosystems (Jiang et al., 2022) and daily-scale
136 ecosystem respiration in the Western United States (Chen et al., 2021).

137 Given these recent advances in data collection and deep learning models, we ask the following
138 questions: (1) can we train a single deep learning model for daily soil CO₂ efflux in sites across
139 gradients of climate and land cover type? (2) What are the dominant drivers of model performance?
140 Based on previous work on the drivers of soil CO₂ efflux (Jiang et al., 2022; Warner et al., 2019),
141 the impact of data bias (Bond-Lamberty et al., 2024; Stell et al., 2021; Xu & Shang, 2016), and
142 the challenges of modeling hot moments (Bernhardt et al., 2017; McClain et al., 2003; Savage et
143 al., 2014; Vargas et al., 2018; Wutzler et al., 2020), we hypothesize that (1) the model will perform
144 better in sites with many years of data and in temperate climates with strong responses to
145 temperature; and (2) the model will perform poorly at sites with a relatively shorter time series of
146 data and noncyclical, highly stochastic data patterns resulting from weak relationships with
147 temperature.

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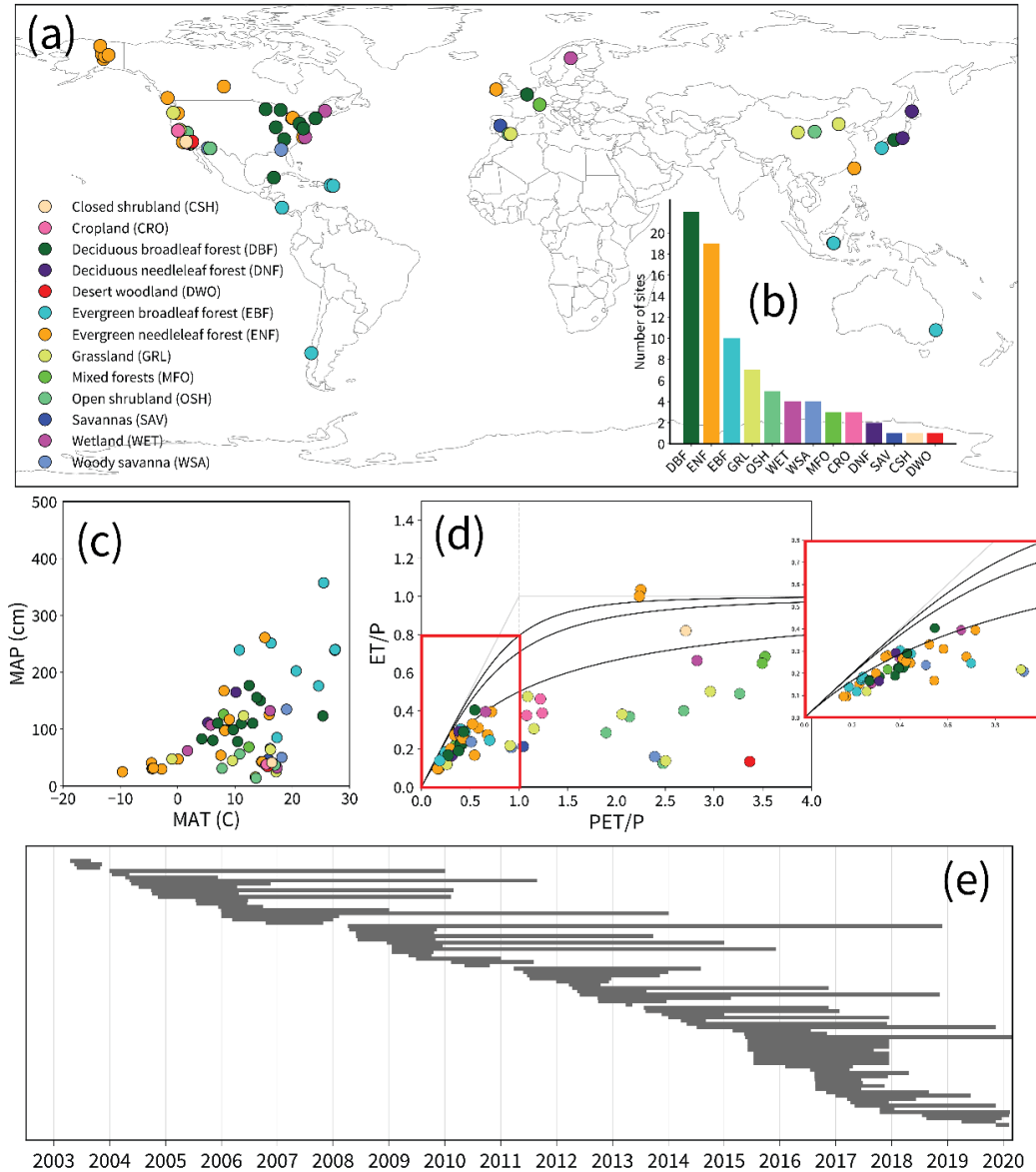
149 **2. METHODS**

150 **2.1. Data Acquisition**

151 The COntinuous SOil REspiration (COSORE) database version 0.7 (Bond-Lamberty et al.,
152 2020) was downloaded in June 2024 (<https://github.com/bpbond/cosore/releases>). We used 82 out
153 of the 85 total sites in the database, with two sites only reporting methane measurements and one
154 site outside of the temporal range of predictor data (Figure 1). The northern hemisphere is far more
155 represented (Figure 1a). There are many represented International Geosphere-Biosphere
156 Programme (IGBP) land cover types (Loveland et al., 2000) (Figure 1b), with the top three being
157 deciduous broadleaf forests, evergreen needleleaf forests, and evergreen broadleaf forests. Many
158 land cover types only have one or two representative sites.

159 The sites cover a wide range of the mean annual temperature and precipitation climate
160 space, although fewer sites have cold and dry or wet and warm climates (Figure 1c). A Budyko
161 plot (Budyko, 1974) (Figure 1d) shows potential evapotranspiration/precipitation (PET/P, the
162 aridity index) versus evapotranspiration/precipitation (ET/P, the evaporative index). Most sites are

163 humid ($PET/P < 1$), but there is ample representation of arid sites ($PET/P > 1$). Open shrublands,
 164 grasslands, woody savannas, savannas, desert woodlands, and croplands emerge as water-limited
 165 land cover types. The database covers 2003-2020 (Figure 1e), with the temporal coverage of
 166 individual sites ranging from 1-10 years.



167
 168 **Figure 1.** (a) Global spatial distribution of the 82 sites in the COSORE database used to train the model, color-coded
 169 by the International Geosphere-Biosphere Programme (IGBP) land cover classification of the site (Loveland et al.,
 170 2000); (b) Number of sites in each land cover type; (c) Sites plotted according to their mean annual temperature (MAT)
 171 versus mean annual precipitation (MAP); (d) Mean annual ET/P (evaporative index) versus mean annual PET/P
 172 (aridity index) (Budyko, 1974). The inset figure zooms in on sites with $PET/P < 1$, or humid sites. Sites with PET/P
 173 > 1 are water-limited or arid. Three Budyko curves with $\beta = 1, 2,$ and 3 are plotted as references. (e) Time coverage
 174 of data in all sites in the COSORE database with available CO₂ data, ordered by the date of the first data entry.
 175

176 COSORE includes static attributes such as latitude, longitude, elevation, and IGBP land
177 cover classification for each site. However, the database lacks accompanying time series data for
178 predictor variables at the same temporal scale as soil CO₂ efflux data. Previous work has shown
179 that temperature (Kätterer et al., 1998; Kirschbaum, 1995; Lloyd & Taylor, 1994; Reichstein et
180 al., 2000; Singh & Gupta, 1977), soil moisture (Carlyle & Than, 1988; Howard & Howard, 1993;
181 Huang et al., 2020; Janssens et al., 2001; Norman et al., 1992; Raich & Potter, 1995), vegetation (
182 Huang et al., 2020; Janssens et al., 2001; Norman et al., 1992; Raich & Potter, 1995), and soil
183 properties (Haaf et al., 2021) are the main drivers of soil CO₂ efflux. We extracted time series of
184 these predictor variables from remote sensing datasets from 01-01-2003 to 29-02-2020 using
185 primarily Google Earth Engine (SI Table 1). Remote sensing data were extracted from a grid cell
186 centered at each site's latitude and longitude with the finest possible scale for each remote sensing
187 product. For example, for a data product with a 250 m resolution, we extracted data over a square
188 cell with 250 m sides centered at the latitude/longitude of the site.

189 For temperature, we used MODIS daily daytime and nighttime land surface temperature
190 (LST) at a spatial scale of 1000 m (<https://doi.org/10.5067/MODIS/MOD11A1.061>) and gap-filled
191 and interpolated it to a daily temporal resolution (Shiff et al., 2021). We also used daily minimum
192 and maximum air temperature from the Global Seamless High-resolution Temperature Dataset
193 (GSHTD) at a spatial scale of 1000 m (Yao et al., 2023).

194 Hourly precipitation data were summed for each day from ERA5-Land at a spatial scale of
195 about 11 km (<https://doi.org/10.24381/cds.68d2bb30>). We also used volumetric soil water content
196 from 0-7 cm depth and 7-28 cm depth from the same mission, at the same spatial scale, and
197 averaged for each day.

198 We used MODIS 8-day evapotranspiration (ET) and potential evapotranspiration (PET) at
199 a spatial scale of 500 m (<https://doi.org/10.5067/MODIS/MOD16A2.061>) because some studies
200 have found a significant correlation between ET and soil CO₂ efflux (Raich & Schlesinger, 1992).
201 Quality flags were used for each data point to mask bad quality data, data from dead detectors,
202 significant cloud cover, and the lowest confidence data.

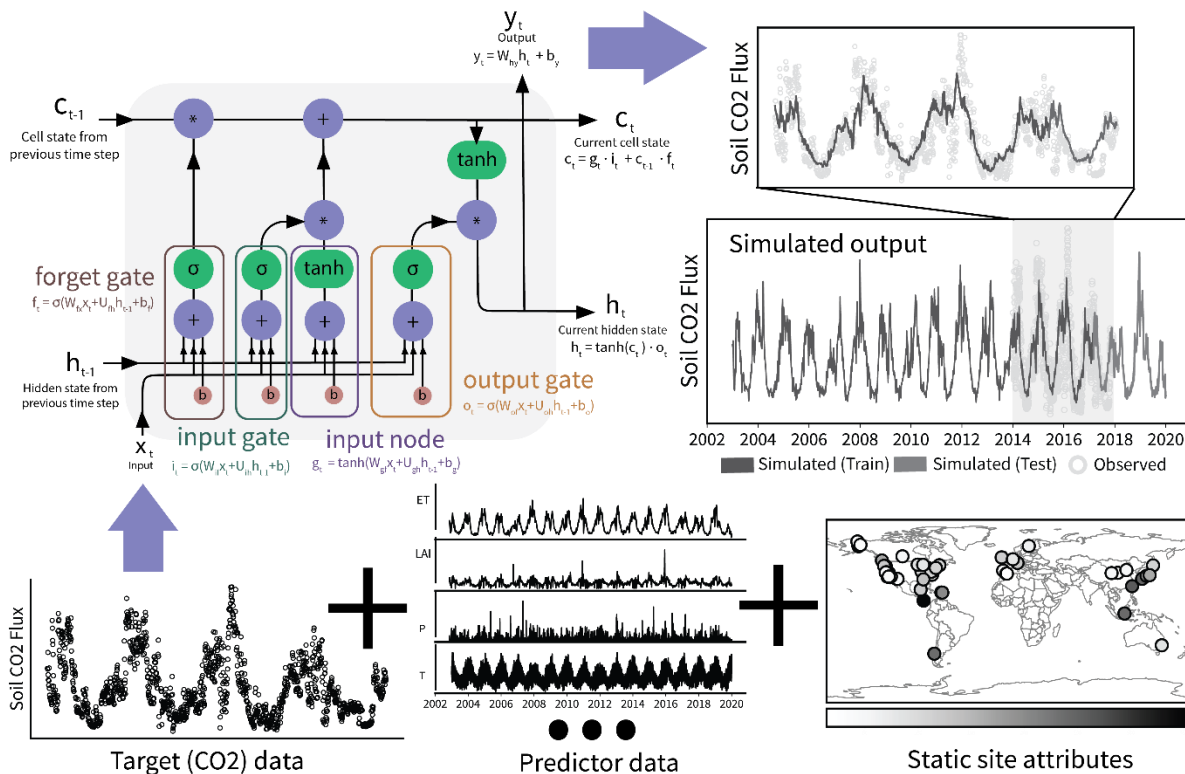
203 For vegetation, we calculated normalized difference vegetation index (NDVI) using bands
204 3 and 4 of the USGS LANDSAT 7 Level 2, Collection 2, Tier 1 dataset at a spatial scale of 30 m
205 (doi.org/10.5066/P9C7I13B). Negative NDVI values were removed, as they represent barren
206 surface or water that did not make physical sense. We also used leaf area index (LAI) and fraction
207 of photosynthetically activate radiation (FPAR) from MODIS at a spatial scale of 500 m and a
208 temporal scale of 4 days.

209 All remote sensing products were cleaned using quality flags where available as described
210 in their original data descriptors. Because the LSTM model requires all predictor data to be
211 continuous, we used a simple linear interpolation method to gap-fill missing data to a daily
212 resolution.

213 We extracted soil properties including soil organic content, sand, silt, clay content, and
214 bulk density, as static site attributes from the SoilGrids2.0 global product (Poggio et al., 2021) at
215 250 m resolution. Mean annual precipitation and temperature were extracted at 1000 m resolution

216 from 1970-2000 from WorldClim version 2 (Fick & Hijmans, 2017). We also used elevation
 217 (meters above sea level), latitude, and longitude from the COSORE database as static site
 218 attributes.

219 Before model training, we first filtered out soil CO₂ efflux values from the raw target dataset
 220 above the 99th percentile or below the 1st percentile of data in each IGBP land cover type. Data are
 221 at a sub-daily scale (2 minutes to 1.5 hours), which we aggregated to the daily scale using the
 222 arithmetic mean for each day. We pre-processed the data by applying an arcsinh transformation, a
 223 commonly used pre-processing approach for data with negative or zero values (Burbidge et al.,
 224 1988; MacKinnon & Magee, 1990) which softens extreme values to help with model training. We
 225 used the equation $D^* = \ln(D + (1+D^2)^{1/2})$, where D^* represents transformed data and D represents
 226 raw data. We then applied a minimum/maximum scaling to all data using the global minimum and
 227 maximum (D_{\min} and D_{\max}) of each predictor, target, and static data across all datasets. For this we
 228 used the equation $D^{*0} = (D^* - D^*_{\min}) / (D^*_{\max} - D^*_{\min})$, where D^{*0} represents scaled data.



229
 230 **Figure 2.** (a) Inputs to the LSTM model include the time series of target data (soil CO₂ efflux data), predictor data,
 231 and static site attributes for all sites. (b) Diagram of LSTM model structure. Inputs from the current time step x_t and
 232 the hidden state of the previous time step h_{t-1} (a “short-term” view of the state of the model) are carried through four
 233 gates: the forget gate, the input gate, the input node, and the output gate. The forget gate decides what information
 234 from the past to discard, the input gate and input node decide what information from the present to let through, and
 235 the output gate decides what information to output to future time steps. At each time step, the model outputs the hidden
 236 state of the model, the cell state of the model (a “long-term” view of the state of the model), and the desired output.
 237 (c) An example of a simulated output time series. The model can predict the entire time series of input predictor data,
 238 which in this model is from roughly 2003 to 2020.
 239

240 2.2. Model Architecture

241 We used an LSTM model, a type of recurrent neural network that excels at resolving the
 242 problem of vanishing gradients and predicting sequential time series (Greff et al., 2017; Hochreiter
 243 & Schmidhuber, 1997; Shen, 2018). We chose to develop one single general model, following the
 244 common consensus about the power of “data synergy” effects: deep learning models perform better
 245 when they use data from diverse sites (Fang et al., 2022; Nearing et al., 2024; Nearing et al., 2021).
 246 Single LSTM models trained on many sites, often in the order of hundreds to even thousands of
 247 sites with daily data across decades, have been shown to perform better than models trained for
 248 individual sites and often outperform traditional process-based models that use parameters
 249 calibrated individually for every site (Nearing et al., 2021).

250 LSTM predicts each time step based on the cumulative influence of all previous time steps.
 251 Through a series of four model “gates,” the model computes which data from the past to
 252 “remember” or carry forward for future predictive power and which data to “forget.” (Figure 2).
 253 The input x includes time series of all predictor variables (daytime land surface temperature, daily
 254 minimum and maximum temperature, precipitation, soil moisture, evapotranspiration, potential
 255 evapotranspiration, leaf area index, NDVI, fraction of photosynthetically active radiation, and
 256 wind speed) and static site attributes (latitude, longitude, elevation, MAP, MAT, soil organic
 257 carbon content, sand, silt, and clay content, and soil bulk density) for each site.

258 At each time step t , the model outputs the cell state c_t (the “long-term” memory of the model),
 259 the hidden state h_t (the “short-term” memory of the model), and the final output y_t , which is the
 260 predicted soil CO₂ efflux at daily scale for each site (Figure 2). The error between the output y and
 261 the corresponding soil CO₂ efflux data is used to inform the tuning of model weights during each
 262 step of training (more details in the Supplementary Information Equations 1.1-1.8). We adapted
 263 code used in Zhi et al. (2023) to train our model, which implemented an LSTM architecture
 264 through the open source hydroDL code (<https://github.com/mhpi/hydroDL>) in PyTorch (Feng et
 265 al., 2020, 2021; Liu et al., 2022, 2024; Shen, 2018). The model was run using the CUDA Deep
 266 Neural Network library (cuDNN) to harness the computational power of the GPU. The model was
 267 run on a cluster with ½ slice of an NVIDIA A100 GPU (~3600 CUDA cores) and took around 1
 268 hour to train.

270 2.3. Model Training and Evaluation

271 For all sites, we used the first 80% predictor and target (CO₂ efflux) time series data for training,
 272 and the last 20% for testing. Because the target data have different temporal coverage at different
 273 sites (Figure 1e), we used a flexible split scheme where data at each site was split at 80% of their
 274 total data points. We then ran a grid search to determine the optimal combination of
 275 hyperparameters. The optimal combination has a hidden state size of 128, a dropout rate of 0.2, a
 276 training-instance length of 30, and a random seed of 0. Mean squared error (MSE) was used as the
 277 objective function for model training and achieved a final loss of 0.003 after 200 training epochs.

278 The Nash Sutcliffe Efficiency (NSE) (Nash & Sutcliffe, 1970) and the Kling-Gupta
 279 Efficiency (KGE) (Gupta et al., 2009) were calculated for the train and test portion of each site to

280 evaluate model performance. NSE is a commonly used criterion for model fit in time series
 281 prediction based on the following equation:

$$NSE = 1 - \frac{\sum_{i=1}^n (R_{obs,i} - R_{sim,i})^2}{\sum_{i=1}^n (R_{obs,i} - \overline{R_{obs}})^2} \quad (1)$$

282 where $R_{obs,i}$ is the observed soil CO₂ efflux at each time step i , $R_{sim,i}$ is the simulated soil CO₂
 283 efflux at each time step i , and $\overline{R_{obs}}$ is the mean of the observed time series. An NSE value above
 284 0.0 indicates that the model fit outperformed the mean of the data, while an NSE value of 1
 285 indicates perfect adherence to observed data.

286 KGE is another commonly used criterion of model fit for time series and is calculated using
 287 the following equation:

$$KGE = 1 - \sqrt{(r - 1)^2 + \left(\frac{\sigma_{sim}}{\sigma_{obs}} - 1\right)^2 + \left(\frac{\mu_{sim}}{\mu_{obs}} - 1\right)^2} \quad (2)$$

288 where r is the linear correlation between observed and simulated soil CO₂ efflux, σ_{obs} and σ_{sim} are
 289 the standard deviation of the observed and simulated soil CO₂ efflux, and μ_{obs} and μ_{sim} are the
 290 means of the observed and simulated soil CO₂ efflux. A KGE above -0.41 indicates that model fit
 291 outperformed the mean of the data (Knoben et al., 2019), while a KGE value of 1 indicates perfect
 292 adherence to observed data.

293 We used both NSE and KGE because NSE is sensitive to extreme values and is therefore
 294 a good indicator of model prediction of pulses or seasonal peaks and troughs. In an LSTM model
 295 for stream dissolved oxygen with hundreds of sites, Zhi et al., (2023) used an NSE value of 0.7 as
 296 a threshold and $NSE < 0.4$ as thresholds for good and poor model performance, respectively. This
 297 model has much fewer sites and lower temporal coverage so we used $NSE > 0.5$ and $KGE > 0.3$
 298 as the metric for good model performance, and $NSE < 0.3$ or $KGE < 0.0$ for poor model
 299 performance. The Pearson correlation coefficient (P_{corr}) between simulation output and
 300 observation data were also used to measure how well the model captured seasonal peaks, with P_{corr}
 301 = 1 indicating perfect adherence to data.

302

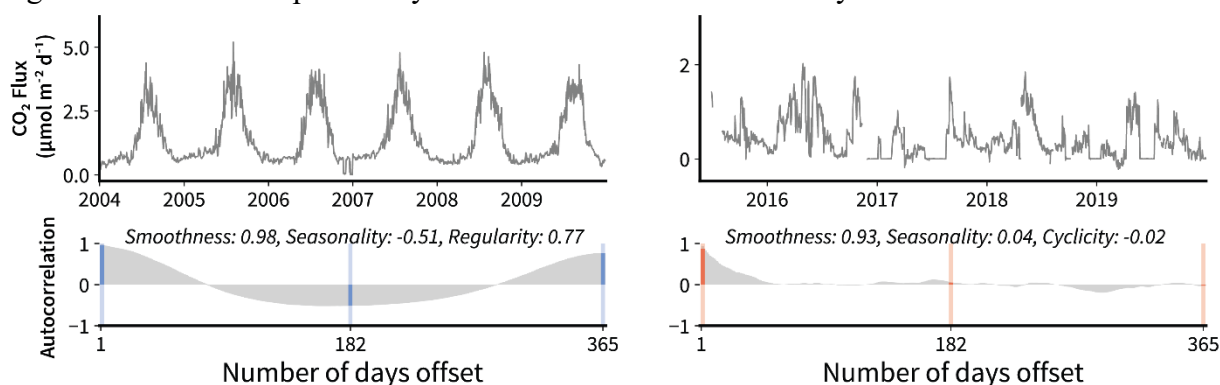
303 3.4. Data characteristics

304 We quantified target data characteristics and explored their relationships with model
 305 performance (Figure 3). Temporal patterns such as seasonality and the prevalence of spikes were
 306 quantified using autocorrelation analyses (Box & Jenkins, 1976). Data *smoothness* was defined as
 307 how much each data point differs from the preceding point, with data possessing many spikes
 308 being less smooth. Autocorrelation at a one-day offset can measure the data correlation with data
 309 from the day before and is a proxy for data smoothness, with smooth data possessing a high
 310 correlation to the preceding point.

311 Sites with highly seasonal trends tended to have a strong negative autocorrelation with their
 312 half-year (182-day) offset. Their data have diverging trends from the opposite season half a year

313 prior, leading to negative autocorrelation (close to -1) at a half-year offset. We call this metric data
 314 *seasonality*.

315 Data *cyclicality* measures how similar data patterns are from year to year, quantified by
 316 autocorrelation to a 365-day offset. More cyclical data with consistent interannual patterns have
 317 high correlation to the previous year's CO₂ efflux on the same day.



318
 319 **Figure 3** An autocorrelation analysis for a site with high smoothness, seasonality, and cyclicality (left) and vice versa
 320 (right). The top figure is a time series of observed soil CO₂ efflux, and the bottom figure is the autocorrelogram for
 321 the site (the Pearson correlation coefficient for the autocorrelation ranging from 1 day of offset to 365 days of offset).
 322 Smoothness was quantified using the extent of autocorrelation at a 1-day offset (leftmost highlighted line in each
 323 autocorrelation figure), seasonality was quantified using the extent of autocorrelation at a 182 (half-year) offset (center
 324 highlighted line in each autocorrelation figure), and cyclicality was quantified using the extent of autocorrelation at a
 325 365-day offset (rightmost highlighted line in each autocorrelation figure). Each autocorrelation figure is labeled with
 326 the smoothness, seasonality, and cyclicality. The left site is from Gaumont-Guay et al. (2014)
 327 ([10.1016/j.agrformet.2013.08.010](https://doi.org/10.1016/j.agrformet.2013.08.010)), and the right site is from Wutzler et al. (2020) ([10.5194/gi-9-239-2020](https://doi.org/10.5194/gi-9-239-2020)), with the
 328 two sites located at (53.99, -105.12) and (39.94, -5.77), respectively.

329

330 3. RESULTS

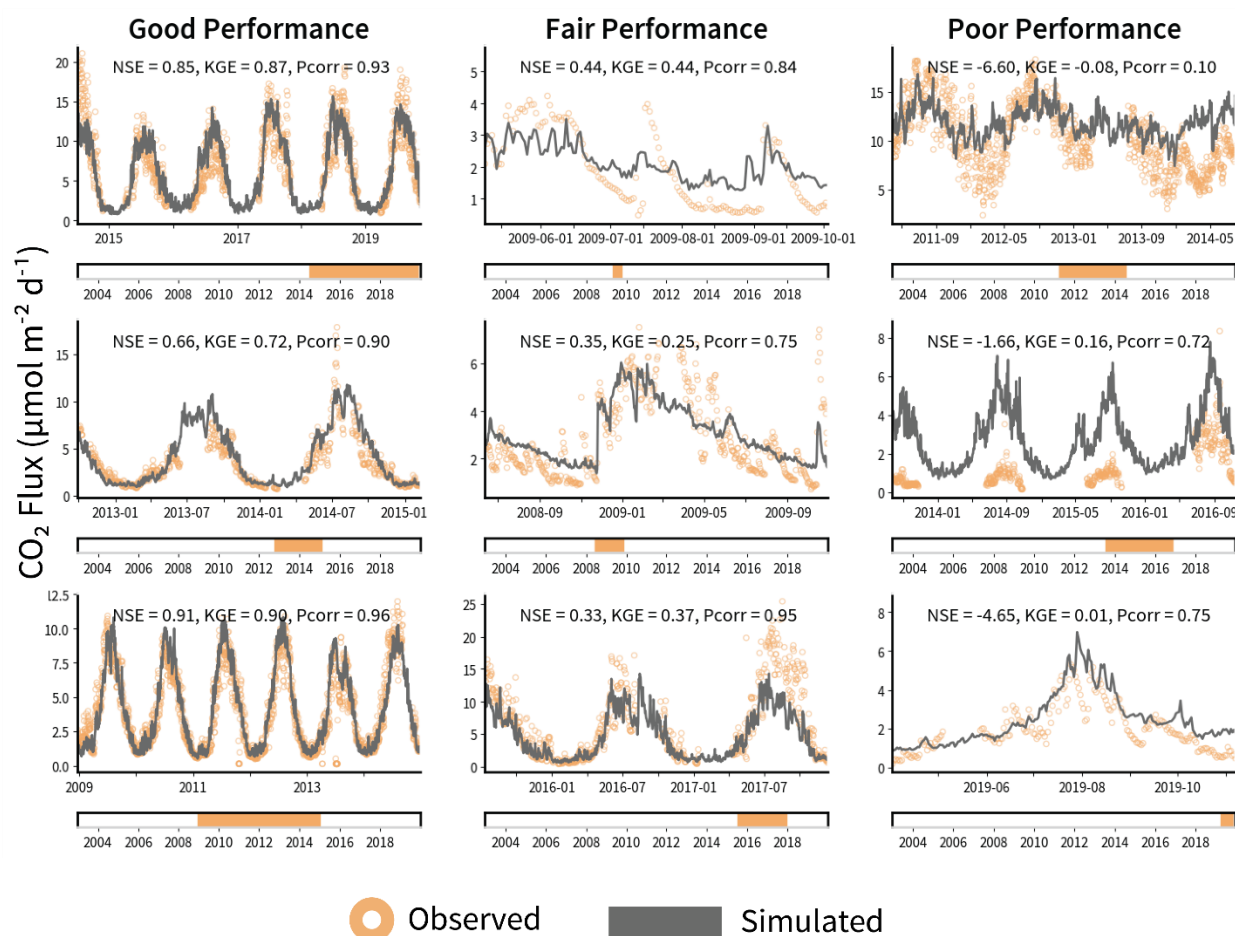
331 3.1. Model Performance Overview

332 The model had a median (mean) train NSE of 0.54 (-0.05), with 57 sites (70%) having a train
 333 NSE value above 0.3 for good or fair performance. Median (mean) train KGE was 0.67 (0.54) and
 334 76 sites (93%) had a train KGE above the threshold 0.0 for good or fair performance. The model
 335 had a median (mean) test NSE of 0.02 (-9.98), with 34 sites (41%) having a test NSE above 0.3.
 336 Model KGE was much better, with a median (mean) test KGE of 0.30 (0.06) and 58 sites (71%)
 337 having a test KGE above 0.0. Test NSE values had some notable outliers, with around 20% of sites
 338 having a test NSE < -5. Overall median (mean) train P_{corr} was 0.82 (0.74) and median (mean) test
 339 P_{corr} was 0.69 (0.55) across all sites.

340 *Well-performing sites* were defined as sites with test NSE > 0.5 and test KGE > 0.3, *fairly-*
 341 *performing sites* were defined as sites with 0.3 < test NSE < 0.5 and 0.0 > test KGE > 0.3, and
 342 *poorly-performing sites* were defined as sites with test NSE < 0.3 or test KGE < 0.0. Well-
 343 performing sites (n = 26) had a median (mean) NSE of 0.70 (0.72), median (mean) KGE of 0.79
 344 (0.78), and median (mean) P_{corr} of 0.90 (0.89), showing a robust simulation of peaks. Fairly-
 345 performing sites (n = 8) had a median (mean) NSE of 0.35 (0.37), median (mean) KGE of 0.36

346 (0.38), and median (mean) P_{corr} of 0.75 (0.77), also indicating excellent reproduction of patterns.
 347 Poorly-performing sites ($n = 48$) had a median (mean) NSE of -1.48 (-17.50), median (mean) KGE
 348 of -0.02 (-0.38), and median (mean) P_{corr} of 0.37 (0.34), showing relatively poor adherence to
 349 seasonal peaks and troughs. Though we used $\text{NSE} < 0.3$ as one measure of poor model
 350 performance, 19% of sites in the “poor performance” category had $\text{KGE} > 0.3$, which indicates
 351 exceptional model performance. 75% of sites in the “fair performance” category had $\text{KGE} > 0.3$.
 352 In other words, NSE values were overall much lower than KGE values.

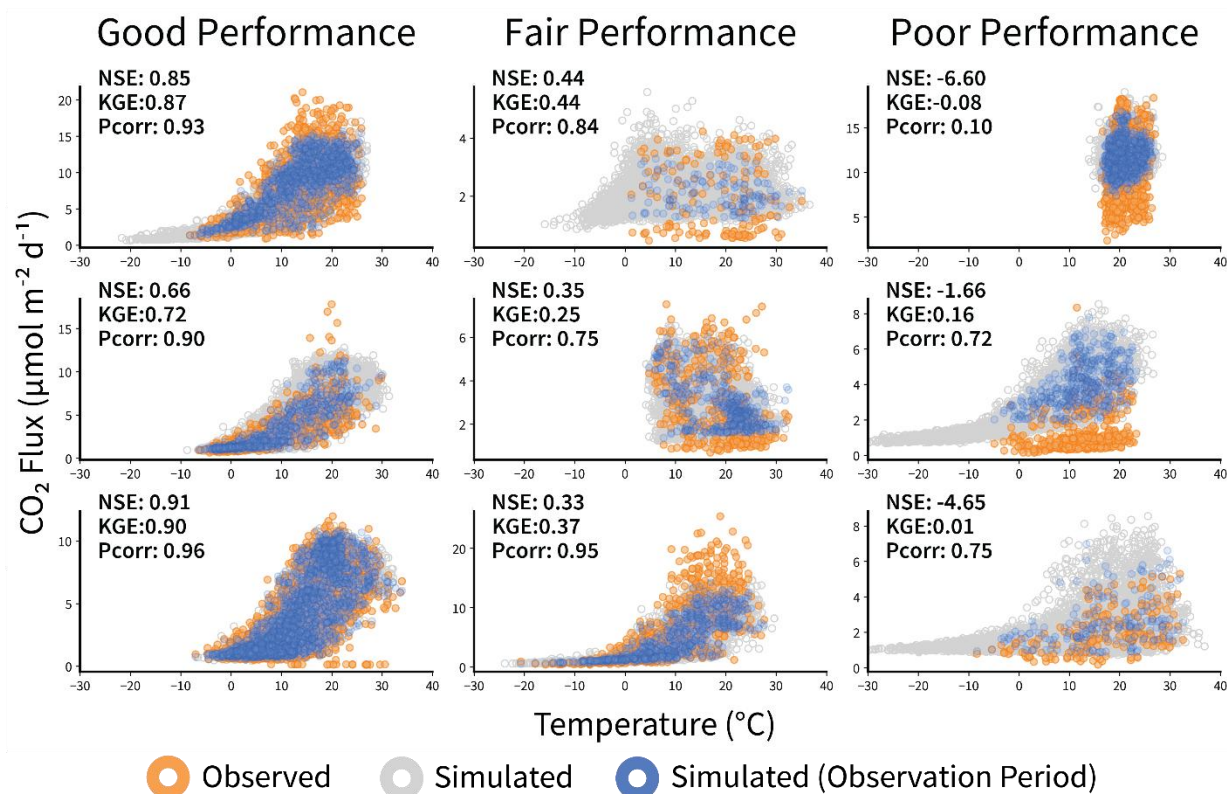
353 The model captured the generally sinusoidal dynamics at sites with good performance but often
 354 missed or underestimated spikes or seasonal maxima (Figure 4). At sites with fair performance,
 355 the model captured the general pattern in some cases while entirely missing the pattern in others.
 356 Sites with fair and poor performance often had less data coverage and less regular patterns across
 357 years. Interestingly, the model often overestimated soil CO_2 efflux at sites with poor performance
 358 (e.g. Third column, middle and last rows). In some cases, sites with the lowest number of data
 359 points had significantly different observed values in the testing portion (the last 20% of each site’s
 360 data points) than the train portion, which could also account for outlier NSE and KGE values at
 361 these sites.



362

363 **Figure 4.** Three randomly selected sites from the group of well-performing sites ($n=26$, left column, test $NSE > 0.3$
 364 and $KGE > 0.0$), fairly-performing sites ($n = 17$, middle column, test $NSE > 0.0$ and < 0.3 and $KGE > -0.4$ and < 0.0),
 365 and poorly-performing sites ($n = 39$, right column, test $NSE < 0.0$ or $KGE < -0.4$). Orange points are observed data,
 366 and the grey line is the model output. Bars at the bottom of each time series plot show the fraction of the total model
 367 time (2003 to 2020) covered by the data shown. Labeled NSE and KGE values are for the test (last 20% of data points)
 368 portion of each time series. Displayed sites with good model performance are from Arain (2016)
 369 ([10.17190/AMF/1246152](https://doi.org/10.17190/AMF/1246152)), Ataka et al. (2014) ([10.1371/journal.pone.0108404](https://doi.org/10.1371/journal.pone.0108404)), and Liang et al. (2017)
 370 ([10.1038/sdata.2017.26](https://doi.org/10.1038/sdata.2017.26)), with coordinates at (42.635328, -80.557731), (34.79, 135.841), and (31.85, 131.3),
 371 respectively. Displayed sites with fair model performance are from Phillips et al. (2016) ([10.1007/s10533-016-0204-](https://doi.org/10.1007/s10533-016-0204-x)
 372 [x](https://doi.org/10.1007/s00442-011-1975)), Carbone et al. (2011) ([10.1007/s00442-011-1975](https://doi.org/10.1007/s00442-011-1975)) and the Shale Hills Critical Zone Observatory, with coordinates
 373 at (44.565, -123.293), (34.00805556, -119.8102778), and (40.66541, -77.90475), respectively. Displayed sites with
 374 poor model performance are from Gutiérrez del Arroyo & Wood (2020) ([10.1029/2019JG005353](https://doi.org/10.1029/2019JG005353)), Sihi et al. (2018)
 375 ([10.1016/j.agrformet.2018.01.026](https://doi.org/10.1016/j.agrformet.2018.01.026)) and Wu et al. (2016) ([10.1071/RJ16023](https://doi.org/10.1071/RJ16023)), with coordinates at (18.4075, -
 376 66.731389), (45.2041, -68.7402), and (40.46, 115.84), respectively.

377 All sites with good model performance had strong exponential relationships between
 378 temperature and CO_2 efflux, while at sites with fair and poor model performance, temperature and
 379 CO_2 efflux relationships were weak (Figure 5). The model captured the temperature- CO_2 efflux
 380 relationships for well-performing and fairly-performing sites but not for poorly-performing sites,
 381 often overshooting CO_2 efflux values over the same temperature range (e.g. third column, middle
 382 row in Figure 5).



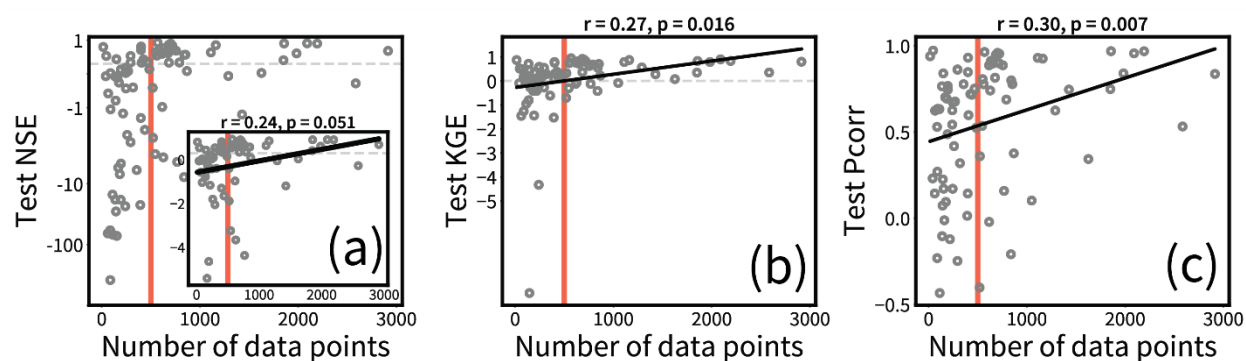
383 **Figure 5.** Temperature- CO_2 efflux relationships in three randomly sampled sites with good (left), fair (middle), and
 384 poor (right) model performance. Light grey points are model output, blue points are the model output at the observed
 385 data period, and orange points are observation data. Labeled NSE , KGE , and P_{corr} values correspond to the test portion
 386 of each dataset. In sites with good performance, the model captured the strong soil CO_2 efflux dependence on
 387

388 temperature. In sites with fair and poor model performance, soil CO₂ efflux often did not have a strong dependence
 389 on temperature. Displayed sites are the same sites in Figure 4.

390
 391 The model tended to underperform at the extremes of soil CO₂ efflux values (Figure 7c). We
 392 calculated NSE and KGE values for the top 5%, middle 90%, and bottom 5% of soil CO₂ efflux
 393 values across all sites. Notably, the model performed significantly more poorly in the top 5%
 394 (median NSE = -11.38, median KGE = -0.33) and bottom 5% (median NSE = -55.66, median KGE
 395 = -1.51) compared to the middle 90% (median NSE = 0.35, median KGE = 0.58). We also
 396 quantified model performance during spikes. Spikes were identified as any point with z-score > 3
 397 or < -3 over a centered moving window of 30 days. Model performance was much worse for spikes
 398 compared to non-spike values (Figure 7d).

400 3.2. Model performance driven by data availability

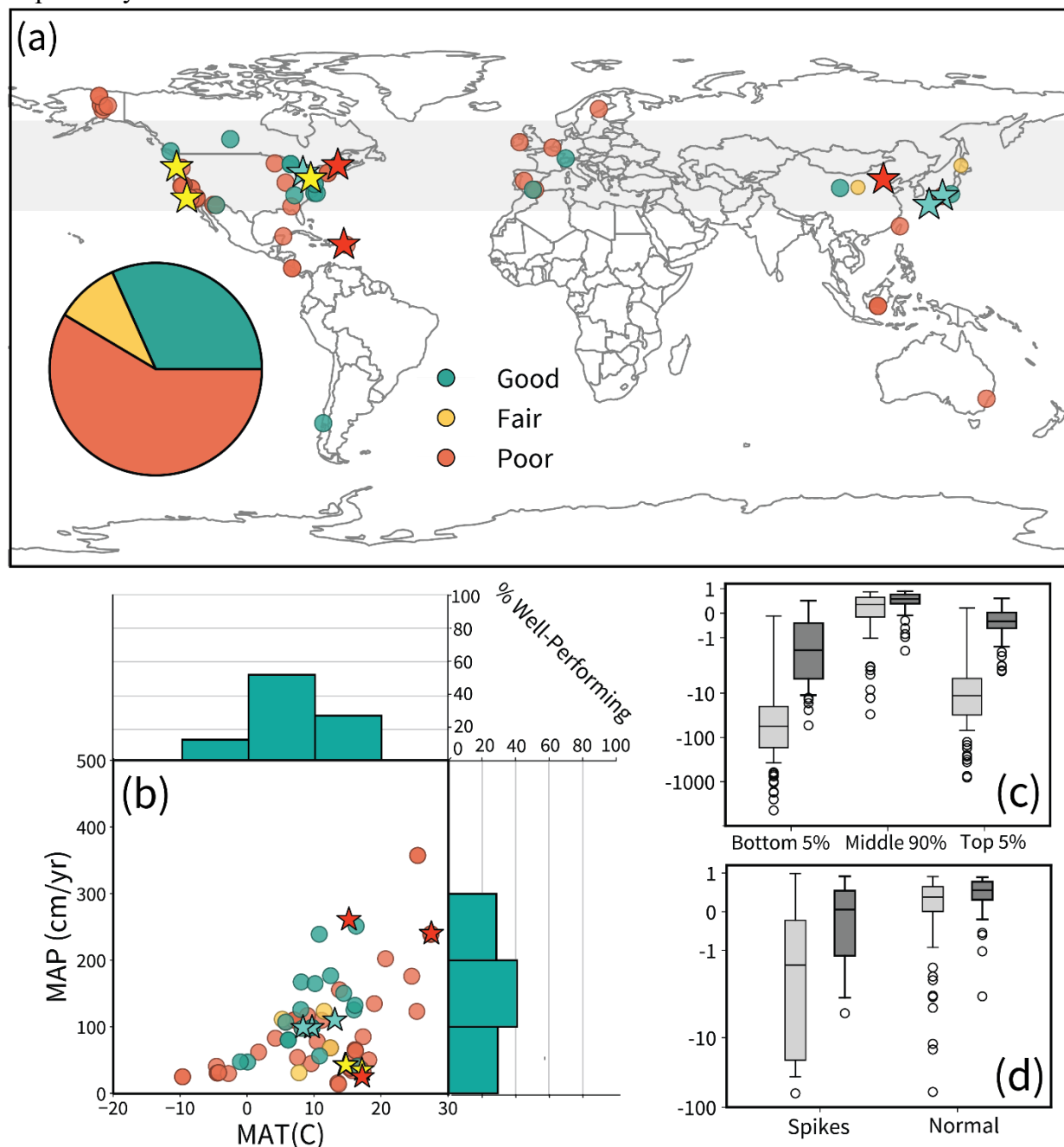
401 Model performance often depended on data attributes including the number of data points and site
 402 representation in the same land cover type. Although the model performed well in many sites with
 403 few data points, there was a threshold-type relationship between the number of data points in
 404 individual sites and model performance. Sites with less than ~500 data points were much less likely
 405 to perform well (Figure 6a, b), indicating that having at least 1.5 years of data drastically improved
 406 model performance. Test KGE values correlated positively and significantly to the number of data
 407 points ($r = 0.27$, $p = 0.016$), but test NSE values were not significantly correlated with number of
 408 data points. When NSE outliers were removed using interquartile range ($> Q3 + 1.5 \cdot IQR$ or $< Q1$
 409 $- 1.5 \cdot IQR$, where $Q3$ is the third quartile of data, $Q1$ is the first quartile of data, and IQR is the
 410 interquartile range), NSE values just barely missed the threshold for a significant correlation ($r =$
 411 0.24 , $p = 0.051$). P_{corr} values also positively correlated to the number of data points ($r = 0.30$, $p =$
 412 0.007), indicating that the number of data points is influential.



413
 414 **Figure 6** Test NSE (a), test KGE (b), and test P_{corr} (c) versus number of data points at each site. Inset plot in the left
 415 figure removes NSE outliers using the interquartile range method, revealing a significant correlation with number of
 416 data points. Correlation coefficients and p-values for relationships with significant correlations are above each plot.
 417 Sites with less than ~500 data points (Around 1.5 years of data, red vertical line) were much more likely to have bad
 418 performance ($NSE < 0.5$, $KGE < 0.0$), and $P_{corr} < 0.5$ (horizontal dashed grey line in each plot).

419 420 3.3. Model performance driven by climate

421 Climate was strongly linked to model performance. Sites with good model performance were
 422 primarily clustered in the 30-60° temperate latitudes (Figure 7a). Outside of this zone, median test
 423 NSE and test KGE were -1.21 and -0.13, respectively, compared to 0.33 and 0.38 inside the zone.
 424 The percentages of well-performing sites within and outside the temperate zone was 40% and 5%,
 425 respectively.

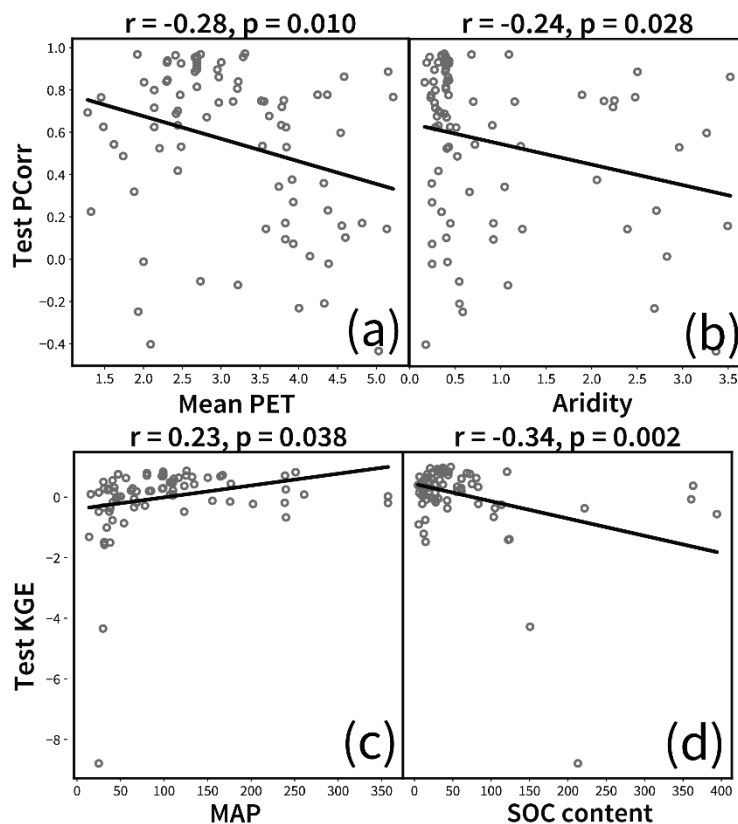


426 **Figure 7.** (a) Spatial distribution of sites with good model performance (green, test NSE > 0.3 and test KGE > 0.0),
 427 fair model performance (yellow, 0.0 < test NSE < 0.3 and -0.4 < test KGE < 0.0), and poor model performance (red,
 428 all other sites). The pie chart shows the relative proportion of sites with different performance. Notably, sites with
 429 good model performance are clustered around 30-60° latitude (light grey highlighted section) (b) Same sites in
 430 climate (MAT-MAP) space. Larger, brighter stars of each color correspond to the three randomly selected sites in
 431

432 each category of performance in Figures 4 and 5. To the right and above each MAT-MAP plot is a bar plot of the %
433 of well-performing site in each MAT or MAP interval. (c) NSE (dark grey) and KGE (light grey) distribution among
434 spikes (left) and non-spikes (right). Spikes were quantified using points with z-score > 3 or < -3 along a 30-day
435 moving window. Values are plotted on a symmetrical log scale for visual clarity. The model performed poorly for
436 spikes. (d) NSE (dark grey) and KGE (light grey) distribution among the bottom 5% of soil CO₂ efflux values (left),
437 the middle 90% of soil CO₂ efflux values (middle), and the top 5% of CO₂ flux values (right). Values are plotted on
438 a symmetrical log scale for visual clarity. The model performed much worse at the highest (spikes) and lowest
439 (troughs) soil CO₂ efflux values.

440 In general, sites at MAT and MAP's lower and upper extremes had relatively poor performance.
441 The intermediate MAT ranges (0 to 15 C) had the highest percentage of sites with good
442 performance (Figure 7b). Outside these temperate temperature ranges, the proportion of sites with
443 good performance was low or zero. Correlating the number of sites in each temperature bin with
444 median NSE and KGE in each bin or the percentage of well-performing sites in each bin did not
445 produce a significant relationship, indicating that the better model performance in temperate MAP-
446 MAT space is not explained by the fact that there are more sites at those climatic ranges.

447 A similar but weaker effect occurred in the breakdown of mean annual precipitation. Sites
448 with MAP between 50 and 300 cm yr⁻¹ generally had a higher percentage of sites with good model
449 performance. The six land cover types with the lowest mean MAP (Desert woodland, Open
450 shrubland, Cropland, Closed shrubland, Savannas, and Grassland) generally had low median test
451 NSE (-30.07, 0.14, -0.40, -0.03, 0.03, and -0.28, respectively) and median test KGE (-0.37, 0.12,
452 -0.29, -0.82, 0.01, and -0.12, respectively). Half of the arid sites (n = 13 out of 26) with MAP < 50
453 cm yr⁻¹ had poor performance.



454
455

456 **Figure 8:** (a) Test P_{corr} versus mean PET; (b) Test P_{corr} versus aridity; (c) test KGE versus MAP; (d) Test KGE versus
457 SOC content

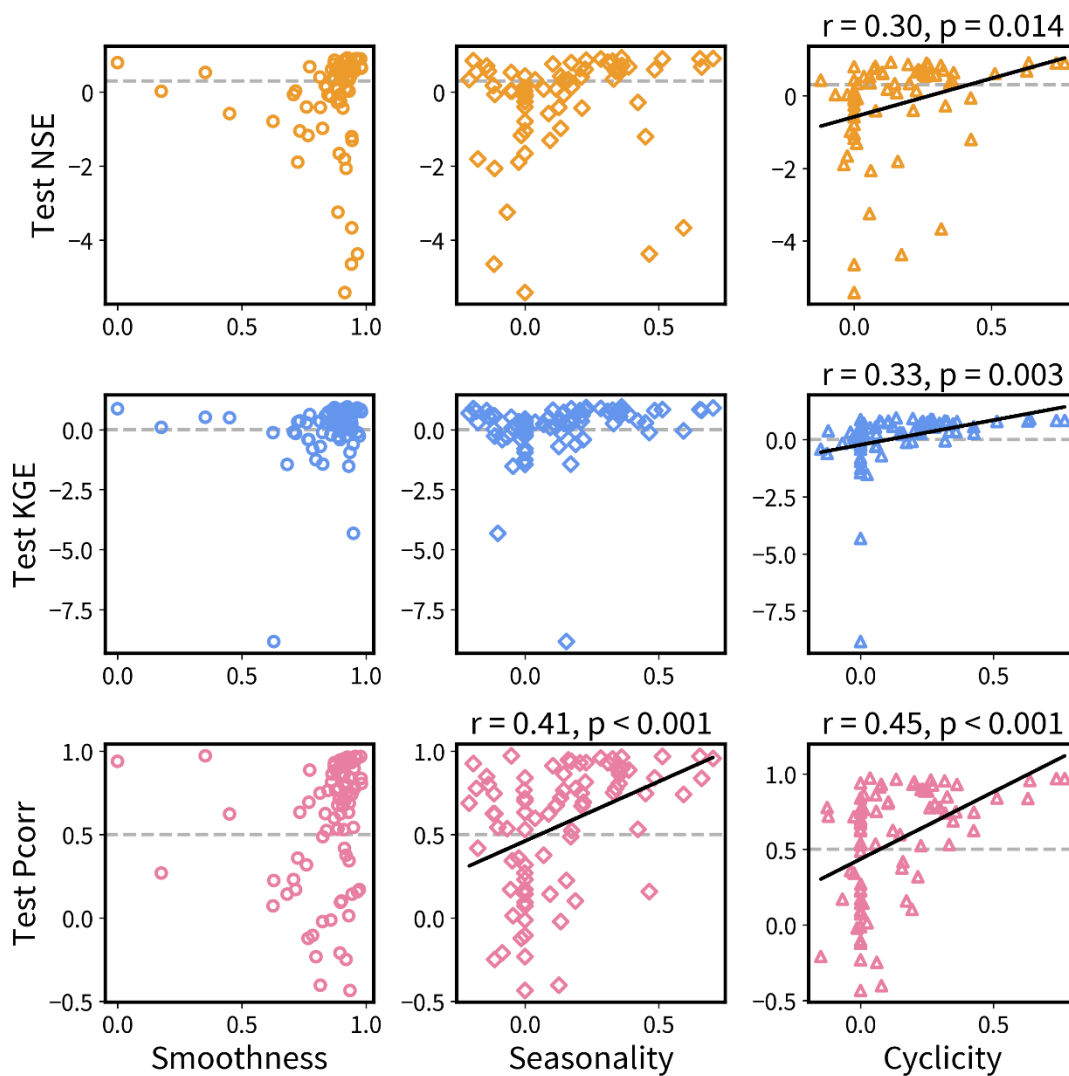
458 Test P_{corr} was negatively correlated with multiple climatic attributes (Figure 8a,b),
459 including mean PET ($r = -0.28, p = 0.010$) and aridity index (PET/P) ($r = -0.24, p = 0.028$),
460 indicating that the model could capture the dynamics of the observed data better in sites with a
461 relatively humid climate. Test KGE was also positively correlated to MAP ($r = 0.23, p = 0.038$)
462 (Figure 8c). Correlating model performance with other site attributes revealed a negative
463 correlation between test KGE and soil organic carbon (SOC) content ($r = -0.34, p = 0.002$) (Figure
464 8d), meaning that sites with higher SOC had poorer model performance.

465

466 3.4. Model performance driven by cyclic seasonal data patterns

467 Temporal patterns such as data smoothness (the prevalence of spikes), seasonality (how
468 strongly CO_2 efflux responded to seasonal temperature variations), and cyclicality (degree of
469 interannual similarity in CO_2 efflux dynamics) were quantified using autocorrelation analyses.
470 Interestingly, data smoothness was not significantly correlated with any performance metric.
471 Seasonality correlated significantly with test P_{corr} ($r = 0.41, p < 0.001$). This indicates that sites
472 where CO_2 efflux are more seasonal or more responsive to temperature performed better. Finally,
473 data cyclicality correlated strongly and significantly with test NSE ($r = 0.30, p = 0.014$), test KGE
474 ($r = 0.33, p = 0.003$), and test P_{corr} ($r = 0.45, p < 0.001$), meaning that sites with similar soil CO_2
475 efflux patterns across years performed better (Figure 9).

476 Furthermore, the average absolute difference between mean summertime and mean
 477 wintertime soil efflux was 4.65 and 1.56 $\mu\text{mol m}^{-2} \text{d}^{-1}$ at sites with good and poor performance,
 478 respectively (t-statistic = 5.16, $p < 0.001$). A wider data spread indicates higher seasonality, with
 479 strong summer peaks and wintertime lows in R_s . This suggests that the model can capture the larger
 480 difference between different seasons but does not do as well at sites where CO_2 efflux is relatively
 481 consistent. This further supports the idea that the model performs better at sites with highly
 482 seasonal soil CO_2 efflux.
 483

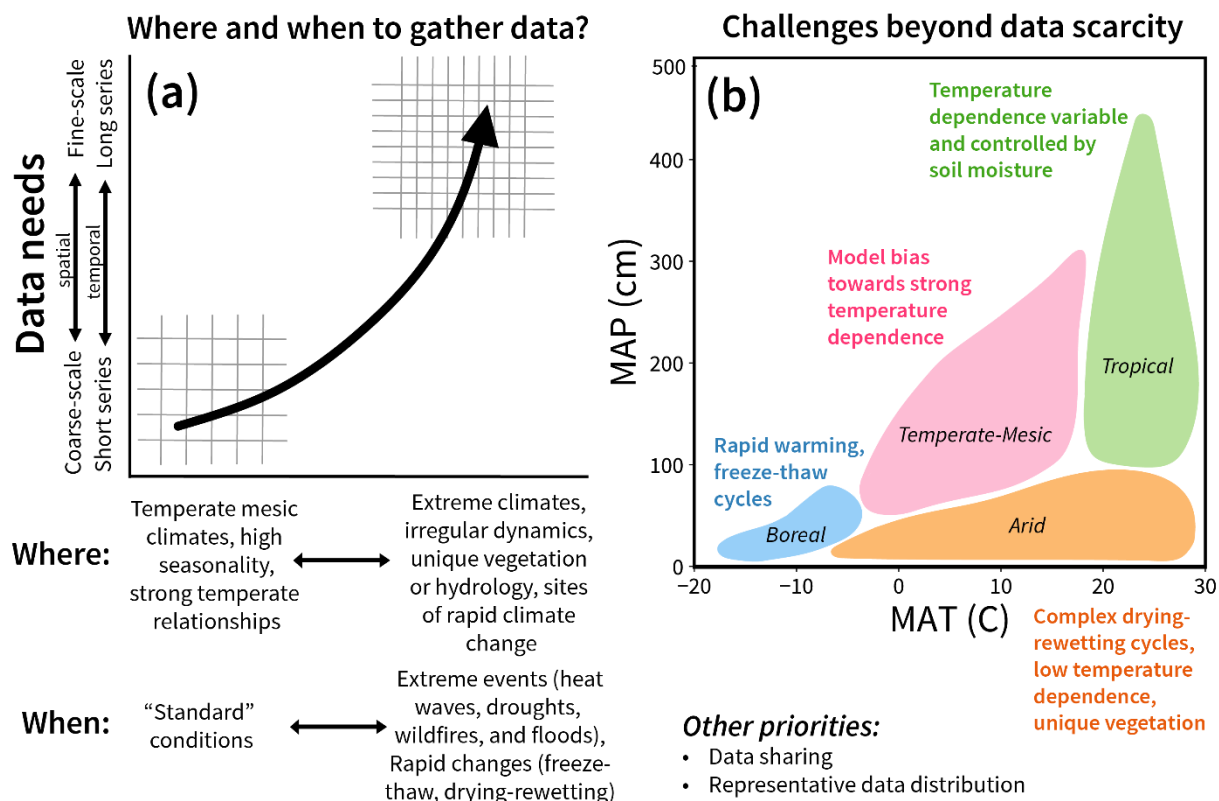


484
 485 **Figure 9.** Correlations between smoothness (circles), seasonality (diamonds), and cyclicity (triangled) of data and
 486 test NSE (orange), test KGE (blue), and test Pearson correlation coefficient (pink). Significant correlations ($p <$
 487 0.05) have labeled r and p values above the plot and plotted correlation lines.
 488

489 4. DISCUSSION

490 Global estimation of soil CO_2 efflux has remained highly uncertain, and daily-scale models
 491 have the potential to reduce estimation uncertainty by using higher temporal resolution data and

492 inferring dynamics from daily to seasonal scales. As an important step in this direction, we
 493 leverage recent advances in continuous, high-frequency data acquisition and emergent deep
 494 learning models to develop a single deep learning model to predict daily soil CO₂ efflux in sites
 495 across gradients of climate, land cover, and soil properties (Reichstein et al., 2019). Our results
 496 largely support our hypotheses regarding when and where the model will perform well. They also
 497 raise fundamental questions about the transferability of daily dynamics and relationships learned
 498 from data-rich sites to data-poor sites, data availability, equity, and priorities for further data
 499 acquisition.



500
 501 **Figure 10.** Overview of model challenges and further data needs recommended to improve future daily-scale models.
 502 (a) The most pressing data needs* are in climate at the high and low end of MAP and MAT spectrum, particularly
 503 those with unique hydrological (e.g., Wetlands and arid sites) characteristics or vegetation (e.g., Biocrusts) or and
 504 those experiencing rapid climate change, as well as during hot moments when soil CO₂ efflux is rapidly mobilized**
 505 (b) Challenges beyond data scarcity in each of four broad climate zones: temperate-mesic, boreal, arid, and tropical.
 506 Other priorities include data sharing and equitable data distribution across many climate types. *Note that more data
 507 is always better, but this represents where data collection efforts should be focused with the goal of correcting past
 508 inequalities.

509

510 4.1 How does the model compare to other models?

511 Deep learning models such as LSTM models are, by definition, data-driven, meaning they
 512 learn relationships between input and output variables based on training data without process-
 513 based knowledge. Naturally, the model will learn better with more training sites and longer data
 514 time series. Being “data greedy” is one of the major limitations of deep learning models such as

515 LSTM models (Shen et al., 2023). Compared to hundreds or even thousands of sites for streamflow
 516 and water quality data with multi-decadal coverage (Kratzert et al., 2023; Sterle et al., 2024), data
 517 here are sparse in terms of both the number of sites (82) and the temporal coverage within
 518 individual sites (0-10 years). As a result, overall model performance cannot rival those of LSTM
 519 models developed for hydrology and water quality predictions (Feng et al., 2021; Zhi et al., 2023).

520 The climate-driven model performance observed here generally aligns with insights
 521 learned from global models for annual soil CO₂ efflux (e.g., Jiang et al., 2024; Warner et al., 2019)
 522 as well as process-based Earth System Models (Guenet et al., 2024), which are most uncertain in
 523 arid/semi-arid, tropical, and cold regions, at the upper and lower ends of MAT and MAP where
 524 temperature relationships are weak and data tends to be much sparser. Our use of an
 525 unprecedentedly large (n = 7.8M) training dataset provides new information on what will be
 526 required to parameterize and assess robust models in biogeochemical and carbon-cycle domains
 527 (Reichstein et al. 2019). The struggle of reproducing hot moments has also been well-documented
 528 in previous studies (Bernhardt et al., 2017; McClain et al., 2003; Savage et al., 2014; Vargas et al.,
 529 2018; Wutzler et al., 2020). It is thus unsurprising that this model fared poorly at capturing the
 530 lowest and highest values across sites, but its challenges in this area provide interesting insights
 531 into both data and process-knowledge limitations that would be obscured using an annual-scale
 532 model.

533

534 **4.2 Can the model extrapolate relationships learned from data-rich sites to data-poor sites?**

535 The climatic distribution of COSORE sites is skewed toward countries with abundant access
 536 to scientific resources (Bond-Lamberty et al., 2024; Kim et al., 2022), reflecting a well-
 537 documented spatial bias in data acquisition (Stell et al., 2021; Xu & Shang, 2016). This is also true
 538 of ecological data more generally (Martin et al., 2012). Like the SRDB database of annual-scale
 539 soil CO₂ efflux data, COSORE overrepresents northern latitudes (30–60°) and temperate climates
 540 while underrepresenting Arctic regions, Central Asia, South America, Africa, Eastern Europe,
 541 Southeast Asia, Australia, and the Pacific islands. Data coverage is particularly sparse in climatic
 542 zones at the extremes of MAT and MAP, such as Arctic and arid ecosystems (e.g., closed
 543 shrublands, desert woodlands, and savannas), with limited site numbers and shorter data records.

544 Because of this data bias, the model mainly learned input-output relationships based on
 545 data biased toward temperate mesic sites with cyclic, seasonal patterns, which is where the model
 546 performed best. This is a common problem when training deep learning models on class-
 547 imbalanced datasets (Bauder et al., 2018; Johnson & Khoshgoftaar, 2019). In fact, all sites with
 548 good model performance had exponential, Arrhenius-type relationships between temperature and
 549 CO₂ efflux (Figure 6). The model then “extrapolated” such relationships to other sites. Such
 550 extrapolation worked well for sites with sparse data but temperate climates, where the model was
 551 able to capture dynamics well. However, when the model extrapolated such relationships to non-
 552 temperate/mesic sites where such relationships were weak or did not exist, it tended to
 553 overestimate soil CO₂ efflux, potentially because it overestimated the strength of temperature as a
 554 driver (Figure 6), leading to poor prediction.

555 This challenges the paradigm that temperature is the primary driver of soil CO₂ efflux
556 across the globe (e.g., Kirschbaum, 1995), which reflects insights from data-rich sites and has been
557 widely used to predict soil CO₂ efflux across the world. Many studies now show that Q₁₀ has high
558 spatial and temporal variability (Chen et al., 2020; Meyer et al., 2018; Zhou et al., 2009), however,
559 implying that temperature dependence is regulated by a complex interplay of factors, particularly
560 in arid and tropical regions (e.g., Le & Vargas, 2024; Wood et al., 2013).

561

562 **4.3 Where and when should we prioritize data collection?**

563 Predicting dynamics requires model generalizability—the ability to extrapolate beyond the
564 training data. Unlike process-based models, deep learning models rely solely on patterns encoded
565 in the training data and require sufficient exposure to diverse input-output relationships to capture
566 temporal trends and spatial patterns. Because we cannot measure the response of soil CO₂ efflux
567 under every condition in every location, the training data we use in such models must represent
568 global diversity in dynamics, responses, and conditions.

569 Achieving meaningful representation is, however, more complex than simply adding data
570 (Villarreal et al., 2019; Villarreal & Vargas, 2021). For example, Stell et al. (2021) demonstrated
571 that expanding spatial coverage in a newer version of SDRB increased model uncertainty; model
572 performance only improved when data distribution optimized coverage of the covariate space. In
573 other words, greater data diversity for better model performance is achieved only when the
574 diversity is relatively balanced; otherwise, models will overfit to well-sampled space, time, and
575 conditions while failing in data-scarce regions and time periods (Bauder et al., 2018; Johnson &
576 Khoshgoftaar, 2019).

577 **Where?** To reduce data bias, we need to collect more data both by increasing the number of
578 sites and extending data duration and conditions at individual sites. Particular attention should be
579 devoted to sites at the upper and lower extremes of MAP and MAT where soil CO₂ efflux is more
580 likely to be regulated by factors other than temperature such as soil moisture. These regions are
581 important to prioritize because they are particularly susceptible to climate change (Bond-Lamberty
582 et al., 2024). Arctic or boreal regions have experienced warming at four times the rate of the rest
583 of the globe (Rantanen et al., 2022) and are highly vulnerable to shifts in vegetation and land cover
584 under climate change (Gonzalez et al., 2010), yet have extremely sparse measurements.
585 Furthermore, arid regions have been known for decades to have much less data than temperate
586 regions (Raich & Schlesinger, 1992) yet they continue to be underrepresented (Xu & Shang, 2016).
587 Arid regions are projected to expand and cover half of the globe by the end of the century (Huang
588 et al., 2016), making it crucial to monitor soil CO₂ efflux in dryland ecosystems.

589 **When?** The model consistently failed to capture the most extreme values, both low and high,
590 across many types of sites. Soil CO₂ efflux at some sites exhibits inherently complex and
591 temporally variable dynamics that are difficult for models to capture. For example, dryland
592 ecosystems experience drying-rewetting cycles that can cause hot moments in CO₂ efflux (Jarvis
593 et al., 2007); freeze-thaw cycles in cold climates can significantly increase dissolved organic
594 carbon availability and CO₂ efflux (Feng et al., 2007; Grogan et al., 2004; Herrmann & Witter,

595 2002; Song et al., 2017). These abrupt fluctuations, common in arid and cold regions, are expected
 596 to become more common in a changing climate (Blankinship & Hart, 2012; Gu et al., 2008; Henry,
 597 2008; Meehl et al., 2006; Sheffield & Wood, 2008; Sinha & Cherkauer, 2010). Similarly, it is
 598 unclear how and how much extreme events like heatwaves, floods, and wildfires, which are
 599 becoming more frequent under climate change (Arnell & Gosling, 2016; Pausas & Keeley, 2021;
 600 Trenberth, 2011), will mobilize or suppress soil CO₂ efflux globally.

601 It is thus important to prioritize measurements that capture peaks and troughs in data,
 602 especially those caused by extreme weather events that often disproportionately change long-term
 603 soil CO₂ efflux dynamics (Vargas, 2012). To capture transient events and their implications, we
 604 need sufficient data to cover the full range of conditions in individual sites. On the other hand,
 605 stochastic events such as storms or rapid thaw, can repeatedly reset ecosystem states, preventing
 606 CO₂ efflux patterns from stabilizing into learnable patterns (Post & Knapp, 2020). In such cases,
 607 longer datasets may offer limited benefits, while finer-scale data capturing the timing and impacts
 608 of stochastic events could be more informative for improving model performance.

609 Peaks and troughs during extreme events can be particularly difficult to predict due to their
 610 rare occurrence and poorly-understood dynamics. However, deep learning models inherently learn
 611 patterns when provided with sufficient data. The key question is how much data is sufficient to
 612 adequately capture these dynamics. Whereas data during extreme events are always rare in
 613 individual sites, aggregating data during extreme events from hundreds to thousands of sites has
 614 shown promises for forecasting extreme events such as flooding (Bertola et al., 2023; Nearing et
 615 al., 2024).

616

617 **4.4 What are the challenges beyond data scarcity?**

618 **Spatial and temporal scale discrepancy** Model performance depends not only on target data
 619 availability but also the quality of predictor data (e.g., temperature, precipitation). If the predictor
 620 data used to train the model does not reflect variability in soil CO₂ efflux, more data will not
 621 improve model performance. One major issue in this model is the mismatch in spatial and temporal
 622 scales between predictor (e.g., temperature, precipitation) and target data (soil CO₂ efflux). For
 623 example, although soil CO₂ efflux data is at daily scale, normalized differential vegetation index
 624 (NDVI) has a four day resolution, evapotranspiration (ET) and potential evapotranspiration (PET)
 625 have an eight day resolution, and wind speed has a monthly resolution. This temporal mismatch
 626 could potentially “smooth out” the variability in predictor data that would meaningfully affect soil
 627 CO₂ efflux dynamics.

628 In addition, soil CO₂ efflux chamber measurements are at a meter scale, whereas global
 629 predictor data are mostly derived from remote sensing products at kilometer or coarser scales. As
 630 such, local variability in soil moisture, precipitation, and vegetation cover are likely averaged out
 631 in predictor data. Such scale discrepancy in data could have substantial impacts, especially where
 632 small-scale heterogeneity in soil moisture, precipitation, vegetation cover, or soil characteristics
 633 may play a disproportionately larger role than temperature in regulating CO₂ efflux (Bloom et al.,
 634 2016; Reichstein et al., 2019). Most notably, local moisture variations are averaged over around
 635 11 square kilometers in the remote sensing data used in this model, making the predictor data used

636 here unlikely to represent local-scale precipitation/soil moisture variations. This could result in
637 model prediction errors where soil CO₂ efflux is highly sensitive to soil moisture. For example, in
638 arid regions with a pronounced dry season, soil rewetting and drying cycling can predominantly
639 control soil CO₂ release (Almagro et al., 2009; Conant et al., 2000). In tropical regions with
640 relatively steady temperatures, CO₂ efflux may be more responsive to precipitation, both positively
641 (e.g., Chambers et al., 2004; Rubio & Detto, 2017) and negatively (Cleveland et al., 2010; Wieder
642 et al., 2011). Within the same land cover or climate type, topographic position can suppress or
643 mobilize soil CO₂ efflux by regulating local topographic wetness index or “wettability” (Riveros-
644 Iregui & McGlynn, 2009).

645 **More data sharing.** In addition to new data acquisition, we also need to prioritize data
646 sharing (Wolkovich et al., 2012). While there are many more sites with soil CO₂ efflux data than
647 the 82 currently in COSORE, these data are not accessible and usable when not in a public dataset.
648 The global carbon research community would greatly benefit from broader contributions to
649 COSORE. Sharing these datasets would enable the application of advanced computational
650 techniques to extract valuable insights and facilitate robust predictions for the future. Failing to
651 upload such data is a missed opportunity to leverage them for large-scale data synthesis and
652 predictive modeling.

653

654 5. CONCLUSION

655 Global-scale estimation of soil CO₂ efflux are mostly at the annual scale and thus miss sub-annual
656 dynamics, potentially contributing to large estimation uncertainties. Here we develop a single deep
657 learning model using daily data from 82 sites in COSORE, aiming to provide a steppingstone for
658 predicting global daily soil CO₂ efflux. This model performs better in temperate and mesic sites
659 with many data points and strong responses to temperature. Our results highlight challenges
660 including temporal and spatial bias of target data and scale discrepancy between predictor and
661 target data, underscoring the need to prioritize data collection in spatially underrepresented
662 systems such as arid, Arctic/boreal, and tropical ecosystems and temporally underrepresented
663 conditions such as extreme weather events that have become increasingly frequent and impactful.
664 Nonetheless, the ability of the model to match daily-scale magnitude and dynamics across many
665 diverse sites is promising. It holds the potential to reduce global estimation uncertainty, facilitate
666 our understanding of daily to seasonal dynamics, to generate synthetic data to temporal and spatial
667 gaps in temperate, mesic regions with good model performance, and to project into the future and
668 the past. Deep learning models hold the potential to revolutionize global soil CO₂ efflux prediction
669 and to generate testable mechanistic hypotheses about how processes are changing under climate
670 change. But they will remain fundamentally limited in their predictive ability without greatly more
671 equitable data collection.

672

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675 Department of Energy DE-SC0025235.

676

677 **Open Research**

678 All predictor and target data and the code used to train the model can be found in a Zenodo
679 repository at [10.5281/zenodo.149077](https://doi.org/10.5281/zenodo.149077).

680

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SUPPLEMENTARY INFORMATION

1128 **Table 1:** Summary of remote sensing data used in this work
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Predictor	Source	Bands	Spatial resolution	Temporal resolution	Units	Access
Precipitation	ECMWF Climate Reanalysis ERA5 Land	precipitation	11132 m	Hourly, summed to daily scale	m	https://doi.org/10.24381/cds.68d2bb30
Soil Moisture	ECMWF Climate Reanalysis ERA5 Land	volumetric_soil_water_layer_1 volumetric_soil_water_layer_1			Volume fraction	https://doi.org/10.24381/cds.68d2bb30
Temperature	MOD21A1D.061 Aqua	LST_Day_1km	1000 m	Daily	K	https://doi.org/10.5067/MODIS/MYD21A1D.061

Minimum/Maximum Temperature	Global Seamless High-resolution Temperature Dataset (GSHTD)	TMIN TMAX	1000 m	Daily	K	https://doi.org/10.1016/j.rse.2022.113422
NDVI	LANDSAT 7 Level 2, Collection 2, Tier 1	SR_B3 SR_B4	30 m	Monthly	N/A	https://doi.org/10.7289/V5ZG6QH9
Evapotranspiration (ET)	MOD16A2GF.061 Terra	ET ET_QC	500 m	8-day	mm/8day	https://doi.org/10.5067/MODIS/MOD16A2GF.061
Potential Evapotranspiration (PET)	MOD16A2GF.061 Terra	PET ET_QC	500 m	8-day	mm/8day	https://doi.org/10.5067/MODIS/MOD16A2GF.061
Fraction of photosynthetically active radiation (FPAR)	MCD15A3H.061	Fpar	500 m	4-day	nm	https://doi.org/10.5067/MODIS/MCD15A3H.061
Leaf area index (LAI)	MCD15A3H.061	Lai	500 m	4-day	nm	https://doi.org/10.5067/MODIS/MCD15A3H.061
Clay, sand, and silt content; bulk density; soil organic carbon content	SoilGrids	soc sand silt clay bdod	250 m	N/A	g/kg % % % kg/dm ³	https://doi.org/10.5194/soil-7-217-2021
Mean annual temperature & precipitation (MAT &	WorldClim Version 2	N/A	1000 m	Annual	°C Cm/yr	https://doi.org/10.1002/joc.5086

MAP)						
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$$x_t = \text{ReLU}(W_I I_t + b_t) \quad (1.1)$$

$$f_t = \sigma(W_{fx} x_t + U_{fh} h_{t-1} + b_f) \quad (1.2)$$

$$i_t = \sigma(W_{ix} x_t + U_{ih} h_{t-1} + b_i) \quad (1.3)$$

$$g_t = \tanh(W_{gx} x_t + U_{gh} h_{t-1} + b_g) \quad (1.4)$$

$$o_t = \sigma(W_{ox} x_t + U_{oh} h_{t-1} + b_o) \quad (1.5)$$

$$c_t = f_t \cdot c_{t-1} + i_t \cdot g_t \quad (1.6)$$

$$h_t = \tanh(c_t) \cdot o_t \quad (1.7)$$

$$y_t = W_{hy} h_t + b_y \quad (1.8)$$

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